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A meta-analysis to compare the sensitivities of earthworms and enchytraeids to different stressors --Manuscript Draft--

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| Abstract: | Earthworms and enchytraeids are soil organisms involved in key soil functions, such as organic matter turnover and soil structure, at different scales. In natural soils, these organisms are exposed and sensitive to different abiotic factors (e.g., climate, land use and management) and are often used as bioindicators of human disturbances, particularly chemical stress. However, the sensitivity of these two groups of Oligochaeta (Annelida) to different stressors has never been compared. Using data from 49 publications and 330 observations, we performed a meta-analysis to compare the sensitivities of earthworms and enchytraeids to all kinds of stressors under similar test conditions. Earthworms and enchytraeids were found to be equally sensitive to chemical stressors (mean effect size -0.61 [-2.53; 1.30]) regardless of the studied endpoint (mortality or reproduction). Most of the observations dealt with the effects of pesticides (42%) and heavy metals (40%) on both organisms. No difference in sensitivity was revealed when these two stressors were considered separately. Regarding the two most studied species of enchytraeids and earthworms, the mean effect sizes of all the possible combinations of Eisenia fetida (41% of the studies) or Eisenia andrei (48%) or Enchytraeus crypticus (73% of the studies) or Enchytraeus albidus (27%) did not reveal any differences in sensitivity. This study also highlights the lack of studies on environmentally relevant (i.e., representative of natural soils) enchytraeid and earthworm species. We also revealed that mostly ecotoxicologists have compared the sensitivities of these two key soil organisms when they are exposed to and threatened by other important factors, such as agricultural practices and climate change. | |
| Response to Reviewers: | | |

Cécile SERBOURCE

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Avignon, July 27th 2024

Dear Editor,

Our research article entitled "A meta-analysis to compare the sensitivities of earthworms and enchytraeids to different stressors" has undergone the last requested revisions. We are again very grateful for that feedback that helped improving the quality of our manuscript.

We sincerely hope that our manuscript meets the publication standards of European Journal of Soil Biology.

Sincerely yours, Cécile Serbource, on behalf of co-authors

Jubarra

Answers to comments, in blue

Dear Dr Serbource,

Apologies for asking for one final change but there is one phrase in your new conclusion that just does not seem quite right. Apologies also for the complexity of the English language! Where you say "loosen organic materials" in line 372, could you change it to " organic rich habitats" or something similar. Then we'll get it published.

Thanks again, Bryan Griffiths

Dear Dr Griffiths, I would like to thank you for your meticulous efforts in improving my work and its clarity.

L 372 "organic rich habitats" is indeed more understandable. I changed it.

Thank you, Cécile Serbource

1 Abstract

2 Earthworms and enchytraeids are soil organisms involved in key soil functions, such as organic matter turnover and soil structure, at different scales. In natural soils, these organisms are exposed and 3 4 sensitive to different abiotic factors (e.g., climate, land use and management) and are often used as 5 bioindicators of human disturbances, particularly chemical stress. However, the sensitivity of these 6 two groups of Oligochaeta (Annelida) to different stressors has never been compared. Using data from 7 49 publications and 330 observations, we performed a meta-analysis to compare the sensitivities of 8 earthworms and enchytraeids to all kinds of stressors under similar test conditions. Earthworms and 9 enchytraeids were found to be equally sensitive to chemical stressors (mean effect size -0.61 [-2.53; 1.30) regardless of the studied endpoint (mortality or reproduction). Most of the observations dealt 10 11 with the effects of pesticides (42%) and heavy metals (40%) on both organisms. No difference in 12 sensitivity was revealed when these two stressors were considered separately. Regarding the two most 13 studied species of enchytraeids and earthworms, the mean effect sizes of all the possible combinations 14 of Eisenia fetida (41% of the studies) or Eisenia andrei (48%) or Enchytraeus crypticus (73% of the 15 studies) or Enchytraeus albidus (27%) did not reveal any differences in sensitivity. This study also 16 highlights the lack of studies on environmentally relevant (i.e., representative of natural soils) 17 enchytraeid and earthworm species. We also revealed that mostly ecotoxicologists have compared the 18 sensitivities of these two key soil organisms when they are exposed to and threatened by other 19 important factors, such as agricultural practices and climate change.

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1. Introduction

2 Earthworms and enchytraeids are macro- and meso-organisms, respectively, that perform several key 3 functions in soils. At different spatial and temporal scales, both taxa are involved in litter fragmentation and organic matter decomposition (Marinissen & Didden, 1997, Brown et al., 2000), 4 5 soil nutrient dynamics (Lavelle et al., 1998; Cole et al., 2002), soil formation (Lavelle et al., 1988, 6 Topoliantz et al., 2000), soil structure dynamics via bioporosity (van Vliet et al., 1993, Edwards & 7 Bohlen 1996), and regulation of microbial biomass (Hedlund & Augustsson, 1995, Brown et al., 8 2000). Because they play key roles in the functioning of agroecosystems and are abundant in many 9 ecosystems, earthworms and enchytraeids are both used as bioindicators of land use, agricultural practices and chemical stress (OECD, 1984, Paoletti, 1999, Didden & Römbke, 2001, Pelosi & 10 Römbke, 2017). 11

12 The ecological niches and behaviours of these two groups do not necessarily overlap (Didden et al., 1997): For example, the earthworm *Lumbricus terrestris* creates extensive burrows that improve water 13 14 infiltration. In contrast, enchytraeids, such as Enchytraeus albidus, play a crucial role in nutrient 15 cycling by decomposing finer organic materials in the soil, their activities are more concentrated in the 16 upper soil layers and leaf litter. Therefore, they can provide complementary functions in soils and 17 complementary information on soil quality (Postma-Blaauw et al., 2012). Under natural conditions, 18 both earthworms and enchytraeids face stressors of different origins. In agricultural soils, tillage or 19 chemical inputs such as pesticides may harm these soil organisms (van Capelle et al., 2012, Beketov et al., 2013). Moreover, climate change has increased the vulnerability of soils and terrestrial biodiversity 20 (IPCC 2022). Phillips et al. (2019) showed that climate variables were more important than soil 21 22 properties or habitat cover in shaping earthworm communities. Drought periods and intense 23 precipitation events now occur more frequently and cause dramatic disruptions to soil biological, physical and chemical parameters (Bellard et al., 2012). Thus, the extreme variations in soil moisture 24 25 as well as the exposure to soil tillage and high quantities of chemicals—directly applied to the soil

surface—can disturb these organisms, their metabolism and the interactions between them (Mäder et
al., 2002, Bengtsson et al., 2005).

Earthworms and enchytraeids are both soft-bodied soil invertebrates, and enchytraeids are considered the closest relatives to earthworms according to Erséus and Källersjö (2004). Enchytraeids have been compared to other soil meso-organisms, such as springtails (Urarov & Karaban, 2015, Hlavkova et al., 2020) or nematodes (Hagner et al., 2010), more often than to earthworms. Moreover, in some studies, different endpoints have been assessed for enchytraeids and earthworms to obtain additional data on the effects of stressors (Niva et al., 2021); therefore, the sensitivities of these two organismal groups cannot be compared. Among the studies that have properly compared earthworms and enchytraeids, some have shown that enchytraeid species exhibit a lower sensitivity than earthworms to certain chemicals and climatic parameters. In Bart et al. (2017), earthworms experienced a decrease in biomass, and they avoided soil contaminated with a copper fungicide at a lower concentration than enchytraeids. Holmstrup et al. (2022) noted that, in Arctic areas, the enchytraeid Enchytraeus albidus individuals had superior freezing tolerance than earthworms. Moreover, Sverdrup et al. (2002) reported that Enchytraeus crypticus exhibited a lower sensitivity to polycyclic aromatic compound exposure than the earthworm *Dendrobaena veneta*; however, their respective sensitivities can vary according to the compound itself. In contrast, enchytraeids reportedly exhibit a greater sensitivity to pesticides (e.g., Sechi et al., 2014 – insecticide alpha-cypermethrin) and pig manure (Segat et al., 2020) than earthworms. Finally, Römbke and Moser (2002), based on the results of an international ring test involving 29 institutions from 15 European countries, noted that there was no clear answer as to the sensitivity of the species tested (*Eisenia fetida/Eisenia andrei* or *E. albidus*) to two chemicals (the fungicide carbendazim and the industrial chemical 4-nitrophenol). The authors concluded that there was no "most sensitive species". Other studies comparing earthworms to enchytraeids have shown similar sensitivities between both soil organisms through a wide range of stress factors, such as uranium (Caetano et al., 2014), electrical conductivity (Gainer et al., 2019), polycyclic aromatic compounds (Kobeticova et al., 2008) and lindane, an insecticide (Lock et al., 2002). A report published in 2009 by Jarratt and Thompson showed mixed conclusions on the sensitivities of

enchytraeid and earthworm species to chemicals, although they suggested that the mortality test results
indicate that Enchytraeidae species may be less sensitive than earthworms. They reported a lower
sensitivity for enchytraeids in some cases but noted that the current data do not show consistent
differences in sensitivity between Lumbricidae and Enchytraeidae in the laboratory or under
semifield/field conditions. To summarize, the message is not clear, perhaps due to the non-quantitative
nature of all these assessments.

The aim of this study was to quantify the relative sensitivities of earthworms and enchytraeids to all kinds of stressors (i.e., any factor likely to negatively affect earthworms and enchytraeids at any level of biological organization). For that purpose, we performed a meta-analysis based on an exhaustive bibliographical review that included all the studies using both earthworms and enchytraeids under the same test conditions (e.g., same soil, tested stressor). We determined the difference in sensitivity to specific stressors according to the following assumptions: (H1) enchytraeids are generally less sensitive than earthworms to different stressors, (H2) differences in sensitivity are mainly observed when mortality is assessed compared to reproduction or avoidance, and (H3) the relative sensitivities of both groups of organisms vary according to the tested stressor and the considered species. Our findings highlight the specific gaps in the literature regarding the sensitivity of these organisms to different stressors, emphasizing the need for future research to address these gaps.

70 2. Methods

2.1. Literature search and inclusion/exclusion criteria

A systematic literature review was conducted in January 2023 to find publications dealing with
earthworms and enchytraeids exposed to different stressors. The literature search was carried out on
the basis of keywords in the ISI Web of Knowledge, using the ''All Databases'' option, with the
following scheme: '(enchytr* OR potworm* OR achaet* OR bryodril* OR buchlol* OR
cernosvitoviel* OR cognat* OR frideric* OR guaranidril* OR lumbricil* OR hemienchytr* OR
oconnorel* OR hemifrideric* OR henl* OR marionin* OR mesenchytr*) and (earthworm* OR
lumbric* OR aporrect* OR eisen* OR dendrob* OR alloloboph* OR octalas* OR dendrodril* OR

diporodril* OR eophil* OR helodril* OR kritodril* OR octodril* OR prosellodril* OR scherothec* OR satchell* OR proctodril* orodril* OR postandril* OR perelia* OR andrei* OR fetid* OR foetid*)' in Topics. From a total of 967 references, a first selection was made using titles and abstracts. The full texts were examined when the information was considered consistent for the meta-analysis. To complete the search, bibliographic reviews that included both annelids were selected, and references were carefully examined to extract any new articles that could be relevant for the meta-analysis. We considered only publications providing data on terrestrial earthworms and enchytraeids, all species included. For instance, papers dealing with aquatic and bioluminescent worm species were excluded (Rodionova et al., 2017). The data on earthworms and enchytraeids had to be from the same study to

compare the sensitivity of these groups under the same conditions (e.g., active substance, type of substrate). Moreover, we did not consider studies dealing with chemical bioaccumulation in animals, as bioaccumulation cannot be easily correlated with certain factors (Tosza et al., 2010; Santorufo et al., 2012). We selected only the studies where the tested factors were supposed to be harmful to soil organisms, specifically those detrimental to their life cycle parameters (e.g., survival, growth, fecundity, juvenile survival). We did not consider studies on biochar addition (possibly beneficial to them, e.g., Briones et al., 2020) or on tillage (Engell et al., 2021) because these factors can be beneficial to enchytraeid populations (e.g., Pelosi & Römbke, 2016). Regarding organic and inorganic waste, one paper was kept because it focused on contaminated incineration ash (Kobetičová et al., 2010). The final set was composed of 49 publications (i.e., 48 papers and one PhD thesis), which included 330 lines in the database (Table S1).

2.2. Data extraction

For each of the 49 studies, the following data were recorded (Table S1):

author, date of publication, _

stressor type (i.e., pesticides, metals, wastes, others);

stressor subcategory (i.e., for pesticides: fungicide, herbicide and insecticide; for metals: antimony, barium, beryllium, cadmium, chromium, copper, lanthanum, manganese, mercury,

| _ | 105 | molybdenum, nickel, lead, silver, uranium and zinc; for wastes: inorganic waste, organic waste, |
|----------------|-----|---|
| 1 2 3 | 106 | vinasse, pig manure waste, mineral byproducts, and dredged sediment; for others: pesticide |
| 4 5 | 107 | additive, boric acid, electrical conductivity, linear alkylbenzene sulfonate, chlorinated paraffins |
| 6 7 | 108 | and polycyclic aromatic compounds); |
| 8 9 | 109 | - earthworm species (i.e., Aporrectodea caliginosa, Aporrectodea longa, Dendrobaena veneta, |
| 10 11 12 | 110 | Eisenia andrei and Eisenia fetida) and enchytraeid species (i.e., Enchytraeus albidus, |
| 13 14 | 111 | Enchytraeus crypticus and Enchytraeus luxurious); |
| 15 16 | 112 | - the observed variable (i.e., survival, reproduction, avoidance, growth, population abundance) |
| 17 18 | 113 | and its measured endpoint (LC _x : lethal concentration; EC _x : effect concentration; NOEC: no |
| 19 20 21 | 114 | observed effect concentration; NEC: no effect concentration; and LOEC: lowest observed |
| 22 23 | 115 | effect concentration); |
| 24 25 | 116 | - the mean value of the endpoints with the standard error, standard deviation, or the respective |
| 26 27 | 117 | 95% confidence intervals; |
| 28 29 30 | 118 | - and the number of replicates. |
| 31 32 | 119 | As the choice was made to select studies that provided ecotoxicological endpoint values to ease the |
| 33 34 | 120 | comparison between earthworms and enchytraeids, all the selected studies were laboratory |
| 35 36 37 | 121 | experiments. |
| 38 39 40 | 122 | First, the data were collected from available tables; otherwise, the data were collected from the text. |
| 41 42 | 123 | Searching for data in the figures (graphs) was not ruled out, but none of the data matched our search |
| 43 44 | 124 | criteria. Because of compliance with ISO standards (ISO, 1998, OECD, 2015), some methodological |
| 45 46 | 125 | differences were tolerated when comparing data for earthworms and enchytraeids. For example, in |
| 48 49 | 126 | reproduction tests, cocoons or juveniles were counted for earthworms, while the number of juveniles |
| 50 51 | 127 | was counted for enchytraeids. The duration of the experiment was also not the same for standard tests |
| 52 53 | 128 | of earthworms and enchytraeids, i.e., 21 and 42 days, respectively. |
| 54 55 | 129 | For both earthworms and enchytraeids, when the endpoint value was not given as a precise number but |
| 50 57 58 | 130 | was bounded by > or <, the data were adjusted as follows: in the case of LC_{50} > x, the data were not |
| - | | |

considered to indicate that the sensitivity of the organism would be overestimated by considering x as the LC₅₀ value. In the case of LC₅₀ < x, the x value was used, and it is conservative to consider x as the LC_{50} value. This latter case represented 20 of the 330 observations in the database (i.e., 6% of the data).

Four comparisons involving boric acid were found in two studies (Becker et al., 2011, Niemeyer et al., 2018). Becker et al. (2011) characterized boric acid as a fertilizer and fungicide, but Niemeyer et al. (2018) did not mention any role or expected effect. Boric acid can be used as an insecticide, acaricide, herbicide, or fungicide (Büyükgüzel et al., 2013). Since it can be placed in different subcategories (insecticide, fungicide, and fertilizer), it was placed into the "others" category.

2.3. Effect size calculation

When the standard deviation (SD) was not given directly from a table, it was calculated with the t distribution for small samples (n<60) using the following formula (Higgins & Green, 2011):

$$SD = \sqrt{n} * \frac{upper \ limit \ CI - \ lower \ limit \ CI}{t_{alpha,df} * 2}$$

With t_{alpha,df}, the T.INV.2T functions in syntax with two associated arguments: probability, associated with Student's t-distribution (here, the alpha=0.05 to respect the 95% CI from extracted data), and degrees of freedom (equal to the group sample size - 1).

When neither the 95% CI nor the SD were given, we used the worst-case scenario, i.e., the highest SD calculated for the same endpoint in the database (LC₁₀, LC₅₀, EC₁₀, EC₂₀, EC₅₀). NOEC and LOEC values have no variance, so their standard deviation (SD) was set to zero. When the number of replicates was not the same between the treatments and the control, the lower number was considered to be conservative (i.e., less weight was given to the study). When the number of replicates was not given and the tests referred to ISO standards, we considered the number of replicates advised in official procedures (e.g., ISO, 1998, OECD, 2015). The confidence intervals (CIs) for each estimate of effect size were calculated using bias-corrected bootstrapping methods, which corrects for nonnormality of the data and nonconstant standard errors (Efron & Hastie, 2016). p. 6

Hedge's d is a standardized mean difference that accounts for the potential bias in estimating 2 157 population effect size (Hedges & Olkin, 1985). The magnitude of the difference between earthworm and enchytraeid sensitivities was calculated with the Hedge metric using the natural logarithm of the response ratio (LRR) as follows (van Groenigen et al., 2014):

$$LRR = ln(\frac{x_i}{x_u})$$

with x_i and x_u being the values of the enchytraeid and earthworm endpoints, respectively. Therefore, when the LRR equals zero, the two organisms have the same sensitivity.

The LRR variance (V) was calculated as follows (van Groenigen et al., 2014):

164
$$V = \frac{1}{n_i} * \left(\frac{SD_i}{x_i}\right)^2 + \frac{1}{n_u} * \left(\frac{SD_u}{x_u}\right)^2$$

where x_i and x_u are the mean values of the endpoints for earthworms and enchytraeids, respectively; SD_i and SD_u are the standard deviations of the endpoint values for enchytraeid and earthworm, respectively; and n_i and n_u are the sample sizes in the enchytraeid and earthworm tests, respectively. Finally, effect sizes were weighted by the inverse of their variance in all the models, giving more weight to well-replicated studies (Koricheva et al., 2013).

3. Meta-analysis

3.1. General modelling approach

Every model incorporated a random effect associated with the study's identity to accommodate correlated data arising from the same study (Viechtbauer, 2010). The models were selected using the Akaike information criterion (AIC). This criterion represents a compromise between bias decreasing with the number of free parameters and parsimony, the desire to describe the data with as few parameters as possible. Lower AIC values indicate a better-fit model; therefore, for each calculation, the smallest AIC was considered for the LRR calculation. Random effect models were tested for the lines (to allow the real effect to vary from one dataset to another), the authors (to allow the actual

effect to vary from one article to another), and a combination of the two (Borenstein et al., 2010). A
combination of the two modalities was associated with a lower AIC for every case, and these models
were systematically used. For each model, we used the restricted maximum likelihood (REML)
method, which aims to provide unbiased estimates of variance components by accounting for fixed
effects in the model. All analyses were performed with R Software (R Development Core Team, 2014)
using the metafor package (Viechtbauer, 2010).

3.2. Analysis

First, a meta-analysis was carried out using estimated overall effect size in a multivariate linear model, calculated using the rma.mv function from the metafor package in R (Viechtbauer, 2010). In the context of a meta-analysis, the effect size (with its 95% confidence interval) is a standardized measure that quantifies the magnitude of the treatment effect or association across all included studies. Then, the logarithm of the response ratio (LRR) was calculated to explore, throught earthworm and enchytraeid endpoints, the results of this model on all the data, all the factors and all the species combined. This main model enabled us to test our first assumption (H1).

We also compared earthworm and enchytraeid data by splitting them according to the studied endpoints (i.e., reproduction or survival rate) to test the second assumption (H2). We further investigated the effects of the different stressors (H3) on the response ratio (LRR) by stressor category, a moderator with 4 levels: pesticides, metals, wastes, and others. We finally investigated the potential "species effect" of the most represented earthworm and enchytraeid species (H3), using the response ratio of the earthworm species categorical moderator at 2 levels (i.e., E. andrei and E. fetida, which accounted for 48% and 41%, respectively; Table S1), coupled with the enchytraeid species categorical moderator at 2 levels (i.e., *E. albidus* and *E. crypticus*, which accounted for 27% and 73%, respectively; Table S1).

Publication bias was assessed by Egger's regression test for funnel plot asymmetry between the standard normal deviation of the effect sizes and precision (i.e., inverse of their standard errors). We retained the multilevel linear model rma.mv by adding a standard error moderator. This test is

powerful given the fairly large number of studies. The findings demonstrated the absence of publication bias: we detected no significant asymmetry (*z*-value = -0.62; *p* value = 0.53) in the funnel plot (Fig. S1), as none of the intercepts from Egger's regression exhibited a significant deviation from zero.

4. Results

4.1. Data description

Location Fig. 1

The 49 selected studies addressed a total of 30 different stressors (Figure 1) classified into four different categories. "Metals" accounted for 39.7% of the observations, with the most represented being molybdenum (2 studies, 36 observations) and zinc (2 studies, 19 observations). "Pesticides" represented 25.5%, with subcategories including insecticides (13 studies, 52 observations), fungicides (6 studies, 31 observations) and herbicides (1 study, 1 observation). Among all the observations, 14.2% were classified into the "Wastes" category, which is widely represented by vinasse (1 study, 26 observations). Finally, "Others" accounted for 20.6% of the observations, with polycyclic aromatic compounds (2 studies, 30 observations) and pesticide additives (1 study, 22 observations) being the most represented. In total, 98.8% of the data concerned contaminant tests (all the stressors studied except for the 4 observations for the electrical conductivity), even though no restrictions were imposed during the literature search.

The 49 selected studies reported reproduction endpoints in 71.2% of the cases (43 studies, 235 observations), survival in 20.9% (20 studies, 69 observations), avoidance in 4.54% (3 studies, 15 observations), growth in 2.4% (1 study, 8 observations), and population abundance in 1% (1 study, 3 observations), with EC₅₀ (106 observations) and NOEC (68 observations) calculations dominating (Table 1).

Location Tab.1

In total, five earthworm species and three enchytraeid species were used in the comparisons (Table 2).
 Together, *E. andrei* and *E. fetida* accounted for 89% of the total number of observations. *E. crypticus*

was the most represented enchytraeid species among the observations.

232 Location Tab.2

4.2. Meta-analysis results

When considering all the stressors together, earthworms and enchytraeids were found to be equally
sensitive, as the mean effect size was -0.61 [-2.53, 1.31] (Fig. 2a). The results were considered nonsignificant because the confidence interval crosses zero, indicating no statistically significant effect.
Similarly, exploration of the data across the four stressor types yielded nonsignificant results (Fig. 2b),
with each confidence interval intersecting zero.

Location Fig.2

Similarly, when the data on mortality (i.e., LC₁₀, LC₅₀, and some NOEC and LOEC data) and those on
reproduction (i.e., some EC₁₀, EC₂₀, EC₂₅, EC₅₀, LOEC, NEC and NOEC) were considered separately
(Fig. 3), no difference in sensitivity was found between earthworms and enchytraeids.

243 Location Fig3

The sensitivities of the different species were assessed when the dataset was statistically sufficient for
comparison. Therefore, *E. andrei* (earthworm) data were compared with *E. crypticus* (enchytraeid)
data, *E. andrei* with *E. albidus* (enchytraeid), *E. fetida* (earthworm) with *E. albidus* and *E. fetida* with *E. crypticus*. The sensitivities of these species were similar (Fig. S2).

248 5. Discussion

This study allowed us to quantitatively compare the relative sensitivities of earthworms and
enchytraeids to different types of stressors. On the basis of 49 published studies and 330 observations,
we revealed similar sensitivities to stressors in these two soil organisms (H1). The conclusion was the
same when considering the different stressors, species or endpoints separately (H2 and H3). Although

the differences were not significant, enchytraeids tended to be slightly more sensitive to metals, pesticide additives and electrical conductivity. In contrast, earthworms tended to be slightly more sensitive to pesticides regardless of their mode of action (fungicide or insecticide), but the differences were still nonsignificant. Large confidence intervals were found for all mean effect sizes. This variability was probably due to the variety of test conditions used for the observations; in particular, the tested substances revealed alternating greater sensitivities of enchytraeids or earthworms. Moreover, the high number of studies with no information on data uncertainty (e.g., standard deviation) (i.e., 58 observations) could explain the absence of significant differences since when neither the 95% CI nor the SD was given, we used a conservative approach by giving the highest SD calculated for the same endpoint in the database.

These results are not in accordance with our assumptions that enchytraeids are less sensitive than earthworms to different stressors (H1), that differences in sensitivity are mainly observed when mortality is assessed (H2), and that the relative sensitivity of both groups of organisms varies according to the tested stressor and the considered species (H3). Enchytraeids, which are smaller than earthworms and 80% microbivorous and 20% saprovorous (Didden, 1990, 1993), ingest less material and smaller particles than earthworms in soils and thus could be less exposed to chemicals (Tourinho et al., 2021, Coleman et al., 2004). However, the relationship between size and sensitivity is debatable, as earthworms are much larger than springtails but are less sensitive to all kinds of pesticides, such as formulations or active substances (Joimel et al., 2022). Moreover, in the same study (Joimel et al., 2022), differences in sensitivity were reported between springtails and enchytraeids, but both of these taxonomic groups belong to the same size class (i.e., soil mesofauna). Small animals have a greater surface area-to-volume ratio than larger organisms and are thus potentially more exposed to external stressors. Regarding chemicals, the size class would be of particular importance if the main route of exposure was dermal contact. Enchytraeids live in close contact with the soil porewater fraction, and their routes of exposure are dermal, intestinal and respiratory and earthworms one is dermal but also dietary exposure (Lock & Janssen, 2003, Römbke, 2003, Hirano & Tamae, 2011). Earthworms, which might be thought to be more sensitive to chemicals at first glance due in particular to their larger size,

can ultimately be less exposed than small enchytraeids in soils. Both earthworms and enchytraeids must maintain a moist body surface to facilitate oxygen absorption through their tegument, which is highly permeable to water. Pesticide characteristics (e.g., polarity, solubility in water) influence the available concentrations in water and in soils, leading to varying sensitivities among taxonomic groups. In accordance with our results, Römbke and Moser (2002) concluded that the overall range of sensitivities of oligochaetes to chemicals was usually quite similar. These authors linked the lack of differentiation between the sensitivities of E. albidus and E. fetida exposed to carbendazim to the high toxicity of this persistent compound. More research should be dedicated to quantifying the different routes of exposure to chemicals for soil organisms to better understand their respective sensitivities in different situations. Moreover, both earthworms and enchytraeids can be considered in risk assessments, and while earthworms are most commonly used, there are many reasons to support the use of enchytraeids in testing. Enchytraeids can thrive in soils where earthworms are scarce (Jarratt & Thompson, 2009). Moreover, Römbke et al. (1994) encouraged the use of enchytraeids when comparing laboratory results with those obtained under semifield and field conditions, as the same species (or closely related species) can be used at all test levels. Other practical and ecological points (ease of handling and rearing, short generation time, reduced soil requirements, cost-effectiveness) make them good organisms for inclusion in laboratory experiments (Römbke & Moser, 2002).

Among the tested species, the earthworm *Eisenia* sp. represented 89% of all species included in the selected studies. E. fetida and E. andrei are composting worms that are not representative of mineral soils and are less sensitive to pesticides than species found in cultivated fields (Pelosi et al., 2013). However, these species are recommended as standard test organisms in various guidelines, such as those of the OECD (OECD, 1984) and ISO (ISO, 1993, 1998, 2019), and have been widely used in published ecotoxicological studies, mainly because they are easy to rear. A few studies have used species that are found in agricultural soils. In Bart et al. (2017), the LC_{50} and EC_{50} values were systematically lower for A. caliginosa (earthworm) than for E. albidus (enchytraeid) (Table S1). In Holmstrup and Krogh (2001), the opposite occurred, with systematically higher LC_{50} and EC_{50} values for A. caliginosa and A. longa (earthworm) than for E. albidus (Table S1). Thus, the results should be

contrasted according to the tested stressor, and additional data are needed to assess the sensitivity of earthworms and enchytraeid species commonly found in the field.

The choice was made here to select studies that provided ecotoxicological endpoint values to facilitate comparisons between earthworms and enchytraeids. Potentially, interesting studies have been excluded that compared both organisms. For instance, Amossé et al. (2018) assessed the short-term effects of two fungicides (Cuprafor Micro® made of copper oxychloride and Swing Gold® made of epoxiconazole and dimoxystrobin) on enchytraeid and earthworm communities under field conditions. They found greater negative effects on the diversity and community structure of earthworms than on those of enchytraeids. Similarly, Pereira et al. (2018) studied the degree to which seawater intrusion and irrigation threaten biodiversity via a terrestrial model ecosystem (TME) approach. They found that the abundance of only enchytraeids significantly decreased, although earthworms generally suffered. These field or semifield studies allowed us to investigate the sensitivity of several species found in natural soils under realistic conditions, which offers more comprehensive insight into the ecological dynamics of the systems (Wall & Moore, 1999). TMEs, which were proposed in the 1990 s, are useful experimental systems for assessing the impact of factors such as pollution or climate change on terrestrial ecosystems.

Although the literature search was performed to include all kinds of stressors in this study, the great majority of the articles in our final selection concerned chemicals. Very few data are available on the effects of climatic parameters on soil organisms (Classen et al., 2015), and data comparing their responses to these parameters remain scarce (Otomo, 2011). However, earthworms and enchytraeids, in addition to all other soil organisms, are increasingly having to cope with climate change (Pritchard, 2011). Warming, changes in precipitation regimes, and changes in the quantity and quality of soil carbon inputs can disrupt the habitat and food resources of soil-living organisms, affecting their survival and reproductive capabilities (Maraldo & Holmstrup, 2009). Some species, such as the enchytraeid Cognettia sphagnetorum, which has strong potential to recover from severe drought stress events, appear to have adaptive potential to climate parameters (Maraldo & Holmstrup, 2009). They can vertically migrate as a strategy for evading dry surface layers and high temperatures (Springett et

al., 1970). However, these organisms were found to have no genetic adaptation to drought, as drought-exposed populations of enchytraeids had the same high sensitivity as those from unexposed control plots (Maraldo, 2009). E. albidus has a high osmotic pressure on its body fluids and can live in environments with strong fluctuations in moisture (Maraldo, 2009), which also enables Fridericia galba to survive in soil with a water retention capacity of less than 20% (corresponding to values below -9.8 bar) for more than 49 days (Dózsa-Farkas 1973). However, additional research is needed to understand the physiological responses of enchytraeids to desiccation. Earthworms are known to be sensitive to drought and high temperatures (Bayley et al., 2010; Lima et al., 2015; González-Alcaraz & van Gestel, 2016), but they have adaptive strategies such as diapause or other physiological mechanisms, such as an increase in the free amino acid alanine, to survive long periods of drought (Holmstrup et al., 2016; Jiménez & Decaëns, 2004). Thus, in the actual context of climatic disturbances and knowing the potential effects of combined stressors on soil organisms (e.g., Bindesbøl et al., 2005; Jensen et al., 2009; Singh et al., 2019), the equilibrium between species and taxa could be modified in natural ecosystems in the future. Maraldo et al. (2006) noted that the synergistic effects of climate change factors and environmental contamination on enchytraeids have been insufficiently explored. Further studies should assess the effects of climatic parameters on functionally useful soil organisms and their relative sensitivity to desiccation or heat in realistic contexts.

We also know that enchytraeids are tolerant of soil acidity (Römbke, 1991), whereas earthworm abundance and diversity are generally low in acidic soils (Chan & Mead, 2003, Moore et al., 2013, Wu and al., 2020). Enchytraeids exhibit high tolerance to other soil properties (clay and organic matter contents) (Kuperman et al., 2006; van Gestel et al., 2011, Castro-Ferreira et al., 2012) and temperature (Holmstrup et al., 2022). However, once again, there is no real comparison to date. The combined effects of stressors related to soil properties and climate were examined by Yli-Olli and Huhta (2000), who assessed the co-occurrence of soil pH, moisture and resource addition on *Dendrobaena octaedra* (Lumbricidae) and *Cognettia sphagnetorum* (Enchytraeidae). They found that *D. octaedra* exhibited

the highest biomass in spruce stand humus at higher pH values, while *C. sphagnetorum* thrived best athigher moisture levels.

362 6. Conclusion

For the first time, this exhaustive synthesis provides a quantitative comparison of the sensitivity of earthworms versus enchytraeids to extrinsic factors (such as metals and pesticides) potentially affecting soil organisms. Under laboratory conditions, comparable sensitivities were found for earthworms and enchytraeids regardless of the studied species, or endpoint. Therefore, to test the effects of metals or pesticides, we could recommend testing on one of these groups in order to limit costs and labour. However, for other stressors such as climate parameters and for a thorough understanding of soil functioning, it remains crucial to consider the physiological responses of both groups due to their distinct ecological roles and potential sensitivity difference. Moreover, to fully capture the complexity of natural habitats, it would be relevant to use species inhabiting mineral soils instead of model species commonly found in organic rich habitats. For decades, researchers have been deepening their understanding of the impacts of natural and anthropic stressors on soil fauna and their related functions, which has allowed to reveal the importance of maintaining the presence and function of biodiversity in our soils. Let's keep digging in that direction.

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1 Figure captions – <u>EVERY FIGURES IN COLOR PLEASE</u>

Figure 1. Relative coverage of different stressors tested for the earthworms and enchytraeids
sensitivity comparisons, in 330 observations from 49 studies. Sunburst diagram representing
proportions of observations for each category.

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Figure 2. Log ratio of the mean effect size (in bold) and its 95% confidence interval (in square
brackets) for (a) all stressors, and (b) the different stressor types separately: pesticides, metals, wastes
and others stressors. The number of observations and papers are into round brackets. Error bars
correspond to 95% confidence intervals computed either from the standard deviations reported by the
authors when available (292 data for earthworm; 293 data for enchytraeids) or from the standard
deviation using the worst-case scenario (38 data for earthworms; 37 data for enchytraeids). On the left
of the 0: higher sensitivity of enchytraeids; on the right: higher sensitivity of earthworms.

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14 Figure 3. Log ratio of the mean effect size (in bold) and its 95% confidence interval (in square 15 brackets) for (a) survival endpoints, and (b) reproduction endpoints. The number of observations and 16 papers are into round brackets. Error bars correspond to 95% confidence intervals computed either 17 from the standard deviations reported by the authors when available (314 data for earthworm; 309 data 18 for enchytraeids concerning reproduction data; 316 data for earthworm; 318 data for enchytraeids 19 concerning survival data) or from the standard deviation using the worst-case scenario (16 data for 20 earthworms; 21 data for enchytraeids concerning reproduction data; 14 data for earthworms; 14 data 21 for enchytraeids concerning survival data). On the left of the 0: higher sensitivity of enchytraeids; on 22 the right: higher sensitivity of earthworms.

23 Figure 1



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26 Figure 2

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29 Figure 3



1 **TABLES**

- 2 Tab.1. Summary table of the number of observations relating to the different endpoints used for the
- 3 comparisons of the sensitivity of earthworms and enchytraeids in 330 observations from 49 studies.
- 4 ECx Effect Concentration; LCx Lethal Concentration; NOEC No Observed Effect Concentration;
- 5 LOEC Lowest Observed Effect and NEC No Effect Concentration.

| ECx | Number of observations | LCx | Number of observations | Other | Number of observations |
|------------------|------------------------|-----------|------------------------|-------|------------------------|
| EC ₅₀ | 106 | LC_{50} | 37 | NOEC | 68 |
| EC_{10} | 44 | LC_{10} | 10 | LOEC | 36 |
| EC_{20} | 23 | | | NEC | 2 |
| EC ₂₅ | 4 | | | | |

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- 8 Tab.2. Summary table of the repartition of the species (in percent) used in for the comparisons of the
- 9 sensitivity of earthworms and enchytraeids in 330 observations from 49 studies.

| Earthworm species | % | Enchytraeid species | % |
|-------------------------|----|-----------------------|----|
| Eisenia andrei | 48 | Enchytraeus crypticus | 73 |
| Eisenia fetida | 41 | Enchytraeus albidus | 27 |
| Dendrobaena veneta | 5 | Enchytraeus luxurious | <1 |
| Aporrectodea caliginosa | 4 | | |
| Aporrectodea longa | 2 | | |

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Declaration of interests

⊠The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

□The authors declare the following financial interests/personal relationships which may be considered as potential competing interests:

Supplementary Material

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1. Introduction

10 2 Earthworms and enchytraeids are macro- and meso-organisms, respectively, that perform several key 12 ³ functions in soils. At different spatial and temporal scales, both taxa are involved in litter 13 4 fragmentation and organic matter decomposition (Marinissen & Didden, 1997, Brown et al., 2000), 15 5 soil nutrient dynamics (Lavelle et al., 1998; Cole et al., 2002), soil formation (Lavelle et al., 1988, 17 6 Topoliantz et al., 2000), soil structure dynamics via bioporosity (van Vliet et al., 1993, Edwards & 7 Bohlen 1996), and regulation of microbial biomass (Hedlund & Augustsson, 1995, Brown et al., 20 8 2000). Because they play key roles in the functioning of agroecosystems and are abundant in many 22 9 ecosystems, earthworms and enchytraeids are both used as bioindicators of land use, agricultural 23 24¹⁰ practices and chemical stress (OECD, 1984, Paoletti, 1999, Didden & Römbke, 2001, Pelosi & 25₁₁ Römbke, 2017). 27 28¹² The ecological niches and behaviours of these two groups do not necessarily overlap (Didden et al., 2913 1997): For example, the earthworm Lumbricus terrestris creates extensive burrows that improve water 3114 infiltration. In contrast, enchytraeids, such as Enchytraeus albidus, play a crucial role in nutrient 32 33¹⁵ cycling by decomposing finer organic materials in the soil, their activities are more concentrated in the 3416 upper soil layers and leaf litter. Therefore, they can provide complementary functions in soils and 3617 complementary information on soil quality (Postma-Blaauw et al., 2012). Under natural conditions, 37 38¹⁸ both earthworms and enchytraeids face stressors of different origins. In agricultural soils, tillage or 39₁₉ chemical inputs such as pesticides may harm these soil organisms (van Capelle et al., 2012, Beketov et 4120 al., 2013). Moreover, climate change has increased the vulnerability of soils and terrestrial biodiversity 42 43²¹ (IPCC 2022). Phillips et al. (2019) showed that climate variables were more important than soil 44 22 properties or habitat cover in shaping earthworm communities. Drought periods and intense 4623 precipitation events now occur more frequently and cause dramatic disruptions to soil biological, 47 48²⁴ physical and chemical parameters (Bellard et al., 2012). Thus, the extreme variations in soil moisture 49₂₅ as well as the exposure to soil tillage and high quantities of chemicals-directly applied to the soil

65

surface—can disturb these organisms, their metabolism and the interactions between them (Mäder et al., 2002, Bengtsson et al., 2005).

Earthworms and enchytraeids are both soft-bodied soil invertebrates, and enchytraeids are considered the closest relatives to earthworms according to Erséus and Källersjö (2004). Enchytraeids have been compared to other soil meso-organisms, such as springtails (Urarov & Karaban, 2015, Hlavkova et al., 2020) or nematodes (Hagner et al., 2010), more often than to earthworms. Moreover, in some studies, different endpoints have been assessed for enchytraeids and earthworms to obtain additional data on the effects of stressors (Niva et al., 2021); therefore, the sensitivities of these two organismal groups cannot be compared. Among the studies that have properly compared earthworms and enchytraeids, some have shown that enchytraeid species exhibit a lower sensitivity than earthworms to certain chemicals and climatic parameters. In Bart et al. (2017), earthworms experienced a decrease in biomass, and they avoided soil contaminated with a copper fungicide at a lower concentration than enchytraeids. Holmstrup et al. (2022) noted that, in Arctic areas, the enchytraeid Enchytraeus albidus individuals had superior freezing tolerance than earthworms. Moreover, Sverdrup et al. (2002) reported that Enchytraeus crypticus exhibited a lower sensitivity to polycyclic aromatic compound exposure than the earthworm Dendrobaena veneta; however, their respective sensitivities can vary according to the compound itself. In contrast, enchytraeids reportedly exhibit a greater sensitivity to pesticides (e.g., Sechi et al., 2014 - insecticide alpha-cypermethrin) and pig manure (Segat et al., 2020) than earthworms. Finally, Römbke and Moser (2002), based on the results of an international ring test involving 29 institutions from 15 European countries, noted that there was no clear answer as to the sensitivity of the species tested (Eisenia fetida/Eisenia andrei or E. albidus) to two chemicals (the fungicide carbendazim and the industrial chemical 4-nitrophenol). The authors concluded that there was no "most sensitive species". Other studies comparing earthworms to enchytraeids have shown similar sensitivities between both soil organisms through a wide range of stress factors, such as uranium (Caetano et al., 2014), electrical conductivity (Gainer et al., 2019), polycyclic aromatic compounds (Kobeticova et al., 2008) and lindane, an insecticide (Lock et al., 2002). A report published in 2009 by Jarratt and Thompson showed mixed conclusions on the sensitivities of

enchytraeid and earthworm species to chemicals, although they suggested that the mortality test results
indicate that Enchytraeidae species may be less sensitive than earthworms. They reported a lower
sensitivity for enchytraeids in some cases but noted that the current data do not show consistent
differences in sensitivity between Lumbricidae and Enchytraeidae in the laboratory or under
semifield/field conditions. To summarize, the message is not clear, perhaps due to the non-quantitative
nature of all these assessments.

The aim of this study was to quantify the relative sensitivities of earthworms and enchytraeids to all kinds of stressors (i.e., any factor likely to negatively affect earthworms and enchytraeids at any level of biological organization). For that purpose, we performed a meta-analysis based on an exhaustive bibliographical review that included all the studies using both earthworms and enchytraeids under the same test conditions (e.g., same soil, tested stressor). We determined the difference in sensitivity to specific stressors according to the following assumptions: (H1) enchytraeids are generally less sensitive than earthworms to different stressors, (H2) differences in sensitivity are mainly observed when mortality is assessed compared to reproduction or avoidance, and (H3) the relative sensitivities of both groups of organisms vary according to the tested stressor and the considered species. Our findings highlight the specific gaps in the literature regarding the sensitivity of these organisms to different stressors, emphasizing the need for future research to address these gaps.

2. Methods

2.1. Literature search and inclusion/exclusion criteria

A systematic literature review was conducted in January 2023 to find publications dealing with
earthworms and enchytraeids exposed to different stressors. The literature search was carried out on
the basis of keywords in the ISI Web of Knowledge, using the "All Databases" option, with the
following scheme: '(enchytr* OR potworm* OR achaet* OR bryodril* OR buchlol* OR
cernosvitoviel* OR cognat* OR frideric* OR guaranidril* OR lumbricil* OR hemienchytr* OR
oconnorel* OR hemifrideric* OR henl* OR marionin* OR mesenchytr*) and (earthworm* OR
lumbric* OR aporrect* OR eisen* OR dendrob* OR alloloboph* OR octalas* OR dendrodril* OR

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diporodril* OR eophil* OR helodril* OR kritodril* OR octodril* OR prosellodril* OR scherothec* OR satchell* OR proctodril* orodril* OR postandril* OR perelia* OR andrei* OR fetid* OR foetid*)' in Topics. From a total of 967 references, a first selection was made using titles and abstracts. The full texts were examined when the information was considered consistent for the meta-analysis. To complete the search, bibliographic reviews that included both annelids were selected, and references were carefully examined to extract any new articles that could be relevant for the meta-analysis. We considered only publications providing data on terrestrial earthworms and enchytraeids, all species

included. For instance, papers dealing with aquatic and bioluminescent worm species were excluded (Rodionova et al., 2017). The data on earthworms and enchytraeids had to be from the same study to compare the sensitivity of these groups under the same conditions (e.g., active substance, type of substrate). Moreover, we did not consider studies dealing with chemical bioaccumulation in animals, as bioaccumulation cannot be easily correlated with certain factors (Tosza et al., 2010; Santorufo et al., 2012). We selected only the studies where the tested factors were supposed to be harmful to soil organisms, specifically those detrimental to their life cycle parameters (e.g., survival, growth, fecundity, juvenile survival). We did not consider studies on biochar addition (possibly beneficial to them, e.g., Briones et al., 2020) or on tillage (Engell et al., 2021) because these factors can be beneficial to enchytraeid populations (e.g., Pelosi & Römbke, 2016). Regarding organic and inorganic waste, one paper was kept because it focused on contaminated incineration ash (Kobetičová et al., 2010). The final set was composed of 49 publications (i.e., 48 papers and one PhD thesis), which included 330 lines in the database (Table S1).

2.2. Data extraction

For each of the 49 studies, the following data were recorded (Table S1):

- author, date of publication,

- stressor type (i.e., pesticides, metals, wastes, others);

- stressor subcategory (i.e., for pesticides: fungicide, herbicide and insecticide; for metals: antimony, barium, beryllium, cadmium, chromium, copper, lanthanum, manganese, mercury,

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| 7 <u>105</u> | molybdenum, nickel, lead, silver, uranium and zinc; for wastes: inorganic waste, organic waste, |
| 9106 | vinasse, pig manure waste, mineral byproducts, and dredged sediment; for others: pesticide |
| 10 11 ¹⁰⁷ | additive, boric acid, electrical conductivity, linear alkylbenzene sulfonate, chlorinated paraffins |
| 12 ₁₀₈ 13 | and polycyclic aromatic compounds); |
| 14109 | - earthworm species (i.e., Aporrectodea caliginosa, Aporrectodea longa, Dendrobaena veneta, |
| 15 16 ¹¹⁰ | Eisenia andrei and Eisenia fetida) and enchytraeid species (i.e., Enchytraeus albidus, |
| 17 ₁₁₁ | Enchytraeus crypticus and Enchytraeus luxurious); |
| 19112 | - the observed variable (i.e., survival, reproduction, avoidance, growth, population abundance) |
| 20 21 ¹¹³ | and its measured endpoint (LCx: lethal concentration; ECx: effect concentration; NOEC: no |
| 22 114 23 | observed effect concentration; NEC: no effect concentration; and LOEC: lowest observed |
| 24 <u>1</u> 15 | effect concentration); |
| 25 26 ¹¹⁶ | - the mean value of the endpoints with the standard error, standard deviation, or the respective |
| 27 117 28 | 95% confidence intervals; |
| 29118 | - and the number of replicates. |
| 30 31 ₁₁₉ | As the choice was made to select studies that provided ecotoxicological endpoint values to ease the |
| 32 | |
| 33120 34 | comparison between earthworms and enchytraeids, all the selected studies were laboratory |
| 35 ¹²¹ | experiments. |
| 36 37122 | First, the data were collected from available tables; otherwise, the data were collected from the text. |
| 38 30 ¹²³ | Searching for data in the figures (graphs) was not ruled out, but none of the data matched our search |
| 40 ₁₂₄ | criteria. Because of compliance with ISO standards (ISO, 1998, OECD, 2015), some methodological |
| 4⊥ 42µ25 | differences were tolerated when comparing data for earthworms and enchytraeids. For example, in |
| 43 44 ¹²⁶ | reproduction tests, cocoons or juveniles were counted for earthworms, while the number of juveniles |
| 45 ₁₂₇ | was counted for enchytraeids. The duration of the experiment was also not the same for standard tests |
| 47128 | of earthworms and enchytraeids, i.e., 21 and 42 days, respectively. |
| 48 49 ₁₂₉ | For both earthworms and enchytraeids, when the endpoint value was not given as a precise number but |
| 50 51 130 | was bounded by > or <, the data were adjusted as follows: in the case of LC_{50} > x, the data were not |
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considered to indicate that the sensitivity of the organism would be overestimated by considering x as the LC_{50} value. In the case of $LC_{50} < x$, the x value was used, and it is conservative to consider x as the LC_{50} value. This latter case represented 20 of the 330 observations in the database (i.e., 6% of the data).

Four comparisons involving boric acid were found in two studies (Becker et al., 2011, Niemeyer et al., 2018). Becker et al. (2011) characterized boric acid as a fertilizer and fungicide, but Niemeyer et al. (2018) did not mention any role or expected effect. Boric acid can be used as an insecticide, acaricide, herbicide, or fungicide (Büyükgüzel et al., 2013). Since it can be placed in different subcategories (insecticide, fungicide, and fertilizer), it was placed into the "others" category.

2.3. Effect size calculation

When the standard deviation (SD) was not given directly from a table, it was calculated with the t distribution for small samples (n<60) using the following formula (Higgins & Green, 2011):

 $SD = \sqrt{n} * \frac{upper \ limit \ CI - \ lower \ limit \ CI}{t_{alpha,df} * 2}$

With $t_{alpha,df}$, the T.INV.2T functions in syntax with two associated arguments: probability, associated with Student's t-distribution (here, the alpha=0.05 to respect the 95% CI from extracted data), and degrees of freedom (equal to the group sample size - 1).

When neither the 95% CI nor the SD were given, we used the worst-case scenario, i.e., the highest SD calculated for the same endpoint in the database (LC₁₀, LC₅₀, EC₁₀, EC₂₀, EC₅₀). NOEC and LOEC values have no variance, so their standard deviation (SD) was set to zero. When the number of replicates was not the same between the treatments and the control, the lower number was considered to be conservative (i.e., less weight was given to the study). When the number of replicates was not given and the tests referred to ISO standards, we considered the number of replicates advised in official procedures (e.g., ISO, 1998, OECD, 2015). The confidence intervals (CIs) for each estimate of effect size were calculated using bias-corrected bootstrapping methods, which corrects for nonnormality of the data and nonconstant standard errors (Efron & Hastie, 2016).

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Hedge's d is a standardized mean difference that accounts for the potential bias in estimating population effect size (Hedges & Olkin, 1985). The magnitude of the difference between earthworm and enchytraeid sensitivities was calculated with the Hedge metric using the natural logarithm of the response ratio (LRR) as follows (van Groenigen et al., 2014):

$$LRR = ln(\frac{x_i}{x_u})$$

with x_i and x_u being the values of the enchytraeid and earthworm endpoints, respectively. Therefore, when the LRR equals zero, the two organisms have the same sensitivity.

The LRR variance (V) was calculated as follows (van Groenigen et al., 2014):

$$V = \frac{1}{n_i} * \left(\frac{SD_i}{x_i}\right)^2 + \frac{1}{n_u} * \left(\frac{SD_u}{x_u}\right)^2$$

where x_i and x_u are the mean values of the endpoints for earthworms and enchytraeids, respectively; SD_i and SD_u are the standard deviations of the endpoint values for enchytraeid and earthworm, respectively; and n_i and n_u are the sample sizes in the enchytraeid and earthworm tests, respectively. Finally, effect sizes were weighted by the inverse of their variance in all the models, giving more weight to well-replicated studies (Koricheva et al., 2013).

3. Meta-analysis

3.1. General modelling approach

Every model incorporated a random effect associated with the study's identity to accommodate correlated data arising from the same study (Viechtbauer, 2010). The models were selected using the Akaike information criterion (AIC). This criterion represents a compromise between bias decreasing with the number of free parameters and parsimony, the desire to describe the data with as few parameters as possible. Lower AIC values indicate a better-fit model; therefore, for each calculation, the smallest AIC was considered for the LRR calculation. Random effect models were tested for the lines (to allow the real effect to vary from one dataset to another), the authors (to allow the actual

effect to vary from one article to another), and a combination of the two (Borenstein et al., 2010). A combination of the two modalities was associated with a lower AIC for every case, and these models were systematically used. For each model, we used the restricted maximum likelihood (REML) method, which aims to provide unbiased estimates of variance components by accounting for fixed effects in the model. All analyses were performed with R Software (R Development Core Team, 2014) using the metafor package (Viechtbauer, 2010).

3.2. Analysis

First, a meta-analysis was carried out using estimated overall effect size in a multivariate linear model, calculated using the rma.mv function from the metafor package in R (Viechtbauer, 2010). In the context of a meta-analysis, the effect size (with its 95% confidence interval) is a standardized measure that quantifies the magnitude of the treatment effect or association across all included studies. Then, the logarithm of the response ratio (LRR) was calculated to explore, throught earthworm and enchytraeid endpoints, the results of this model on all the data, all the factors and all the species combined. This main model enabled us to test our first assumption (H1). We also compared earthworm and enchytraeid data by splitting them according to the studied

endpoints (i.e., reproduction or survival rate) to test the second assumption (H2). We further investigated the effects of the different stressors (H3) on the response ratio (LRR) by stressor category, a moderator with 4 levels: pesticides, metals, wastes, and others. We finally investigated the potential "species effect" of the most represented earthworm and enchytraeid species (H3), using the response ratio of the earthworm species categorical moderator at 2 levels (i.e., *E. andrei* and *E. fetida*, which accounted for 48% and 41%, respectively; Table S1), coupled with the enchytraeid species categorical moderator at 2 levels (i.e., *E. albidus* and *E. crypticus*, which accounted for 27% and 73%, respectively; Table S1).

Publication bias was assessed by Egger's regression test for funnel plot asymmetry between the standard normal deviation of the effect sizes and precision (i.e., inverse of their standard errors). We retained the multilevel linear model rma.mv by adding a standard error moderator. This test is

powerful given the fairly large number of studies. The findings demonstrated the absence of publication bias: we detected no significant asymmetry (*z*-*value* = -0.62; *p value* = 0.53) in the funnel plot (Fig. S1), as none of the intercepts from Egger's regression exhibited a significant deviation from zero.

4. Results

4.1. Data description

Location Fig. 1

The 49 selected studies addressed a total of 30 different stressors (Figure 1) classified into four different categories. "Metals" accounted for 39.7% of the observations, with the most represented being molybdenum (2 studies, 36 observations) and zinc (2 studies, 19 observations). "Pesticides" represented 25.5%, with subcategories including insecticides (13 studies, 52 observations), fungicides (6 studies, 31 observations) and herbicides (1 study, 1 observation). Among all the observations, 14.2% were classified into the "Wastes" category, which is widely represented by vinasse (1 study, 26 observations). Finally, "Others" accounted for 20.6% of the observations, with polycyclic aromatic compounds (2 studies, 30 observations) and pesticide additives (1 study, 22 observations) being the most represented. In total, 98.8% of the data concerned contaminant tests (all the stressors studied except for the 4 observations for the electrical conductivity), even though no restrictions were imposed during the literature search.

The 49 selected studies reported reproduction endpoints in 71.2% of the cases (43 studies, 235 observations), survival in 20.9% (20 studies, 69 observations), avoidance in 4.54% (3 studies, 15 observations), growth in 2.4% (1 study, 8 observations), and population abundance in 1% (1 study, 3 observations), with EC₅₀ (106 observations) and NOEC (68 observations) calculations dominating (Table 1).

Location Tab.1

In total, five earthworm species and three enchytraeid species were used in the comparisons (Table 2). Together, *E. andrei* and *E. fetida* accounted for 89% of the total number of observations. *E. crypticus* was the most represented enchytraeid species among the observations.

Location Tab.2

4.2. Meta-analysis results

When considering all the stressors together, earthworms and enchytraeids were found to be equally sensitive, as the mean effect size was -0.61 [-2.53, 1.31] (Fig. 2a). The results were considered non-significant because the confidence interval crosses zero, indicating no statistically significant effect. Similarly, exploration of the data across the four stressor types yielded nonsignificant results (Fig. 2b), with each confidence interval intersecting zero.

Location Fig.2

Similarly, when the data on mortality (i.e., LC₁₀, LC₅₀, and some NOEC and LOEC data) and those on reproduction (i.e., some EC₁₀, EC₂₀, EC₂₅, EC₅₀, LOEC, NEC and NOEC) were considered separately (Fig. 3), no difference in sensitivity was found between earthworms and enchytraeids.

Location Fig3

The sensitivities of the different species were assessed when the dataset was statistically sufficient for comparison. Therefore, *E. andrei* (earthworm) data were compared with *E. crypticus* (enchytraeid) data, *E. andrei* with *E. albidus* (enchytraeid), *E. fetida* (earthworm) with *E. albidus* and *E. fetida* with *E. crypticus*. The sensitivities of these species were similar (Fig. S2).

5. Discussion

This study allowed us to quantitatively compare the relative sensitivities of earthworms and enchytraeids to different types of stressors. On the basis of 49 published studies and 330 observations, we revealed similar sensitivities to stressors in these two soil organisms (H1). The conclusion was the same when considering the different stressors, species or endpoints separately (H2 and H3). Although

the differences were not significant, enchytraeids tended to be slightly more sensitive to metals,
pesticide additives and electrical conductivity. In contrast, earthworms tended to be slightly more
sensitive to pesticides regardless of their mode of action (fungicide or insecticide), but the differences
were still nonsignificant. Large confidence intervals were found for all mean effect sizes. This
variability was probably due to the variety of test conditions used for the observations; in particular,
the tested substances revealed alternating greater sensitivities of enchytraeids or earthworms.
Moreover, the high number of studies with no information on data uncertainty (e.g., standard
deviation) (i.e., 58 observations) could explain the absence of significant differences since when
neither the 95% CI nor the SD was given, we used a conservative approach by giving the highest SD
calculated for the same endpoint in the database.

These results are not in accordance with our assumptions that enchytraeids are less sensitive than earthworms to different stressors (H1), that differences in sensitivity are mainly observed when mortality is assessed (H2), and that the relative sensitivity of both groups of organisms varies according to the tested stressor and the considered species (H3). Enchytraeids, which are smaller than earthworms and 80% microbivorous and 20% saprovorous (Didden, 1990, 1993), ingest less material and smaller particles than earthworms in soils and thus could be less exposed to chemicals (Tourinho et al., 2021, Coleman et al., 2004). However, the relationship between size and sensitivity is debatable, as earthworms are much larger than springtails but are less sensitive to all kinds of pesticides, such as formulations or active substances (Joimel et al., 2022). Moreover, in the same study (Joimel et al., 2022), differences in sensitivity were reported between springtails and enchytraeids, but both of these taxonomic groups belong to the same size class (i.e., soil mesofauna). Small animals have a greater surface area-to-volume ratio than larger organisms and are thus potentially more exposed to external stressors. Regarding chemicals, the size class would be of particular importance if the main route of exposure was dermal contact. Enchytraeids live in close contact with the soil porewater fraction, and their routes of exposure are dermal, intestinal and respiratory and earthworms one is dermal but also dietary exposure (Lock & Janssen, 2003, Römbke, 2003, Hirano & Tamae, 2011). Earthworms, which might be thought to be more sensitive to chemicals at first glance due in particular to their larger size, p. 11

can ultimately be less exposed than small enchytraeids in soils. Both earthworms and enchytraeids must maintain a moist body surface to facilitate oxygen absorption through their tegument, which is highly permeable to water. Pesticide characteristics (e.g., polarity, solubility in water) influence the available concentrations in water and in soils, leading to varying sensitivities among taxonomic groups. In accordance with our results, Römbke and Moser (2002) concluded that the overall range of sensitivities of oligochaetes to chemicals was usually quite similar. These authors linked the lack of differentiation between the sensitivities of E. albidus and E. fetida exposed to carbendazim to the high toxicity of this persistent compound. More research should be dedicated to quantifying the different routes of exposure to chemicals for soil organisms to better understand their respective sensitivities in different situations. Moreover, both earthworms and enchytraeids can be considered in risk assessments, and while earthworms are most commonly used, there are many reasons to support the use of enchytraeids in testing. Enchytraeids can thrive in soils where earthworms are scarce (Jarratt & Thompson, 2009). Moreover, Römbke et al. (1994) encouraged the use of enchytraeids when comparing laboratory results with those obtained under semifield and field conditions, as the same species (or closely related species) can be used at all test levels. Other practical and ecological points (ease of handling and rearing, short generation time, reduced soil requirements, cost-effectiveness) make them good organisms for inclusion in laboratory experiments (Römbke & Moser, 2002). Among the tested species, the earthworm Eisenia sp. represented 89% of all species included in the selected studies. E. fetida and E. andrei are composting worms that are not representative of mineral soils and are less sensitive to pesticides than species found in cultivated fields (Pelosi et al., 2013).

solis and are less sensitive to pesticides than species round in cultivated fields (Pelosi et al., 2013). However, these species are recommended as standard test organisms in various guidelines, such as those of the OECD (OECD, 1984) and ISO (ISO, 1993, 1998, 2019), and have been widely used in published ecotoxicological studies, mainly because they are easy to rear. A few studies have used species that are found in agricultural soils. In Bart et al. (2017), the LC₅₀ and EC₅₀ values were systematically lower for *A. caliginosa* (earthworm) than for *E. albidus* (enchytraeid) (Table S1). In Holmstrup and Krogh (2001), the opposite occurred, with systematically higher LC₅₀ and EC₅₀ values for *A. caliginosa* and *A. longa* (earthworm) than for *E. albidus* (Table S1). Thus, the results should be

contrasted according to the tested stressor, and additional data are needed to assess the sensitivity of earthworms and enchytraeid species commonly found in the field. The choice was made here to select studies that provided ecotoxicological endpoint values to facilitate comparisons between earthworms and enchytraeids. Potentially, interesting studies have been excluded that compared both organisms. For instance, Amossé et al. (2018) assessed the short-term effects of two fungicides (Cuprafor Micro® made of copper oxychloride and Swing Gold® made of epoxiconazole and dimoxystrobin) on enchytraeid and earthworm communities under field conditions. They found greater negative effects on the diversity and community structure of earthworms than on those of enchytraeids. Similarly, Pereira et al. (2018) studied the degree to which seawater intrusion and irrigation threaten biodiversity via a terrestrial model ecosystem (TME) approach. They found that the abundance of only enchytraeids significantly decreased, although earthworms generally suffered. These field or semifield studies allowed us to investigate the sensitivity of several species found in natural soils under realistic conditions, which offers more comprehensive insight into the ecological dynamics of the systems (Wall & Moore, 1999). TMEs, which were proposed in the 1990 s, are useful experimental systems for assessing the impact of factors such as pollution or climate change on terrestrial ecosystems.

Although the literature search was performed to include all kinds of stressors in this study, the great majority of the articles in our final selection concerned chemicals. Very few data are available on the effects of climatic parameters on soil organisms (Classen et al., 2015), and data comparing their responses to these parameters remain scarce (Otomo, 2011). However, earthworms and enchytraeids, in addition to all other soil organisms, are increasingly having to cope with climate change (Pritchard, 2011). Warming, changes in precipitation regimes, and changes in the quantity and quality of soil carbon inputs can disrupt the habitat and food resources of soil-living organisms, affecting their survival and reproductive capabilities (Maraldo & Holmstrup, 2009). Some species, such as the enchytraeid *Cognettia sphagnetorum*, which has strong potential to recover from severe drought stress events, appear to have adaptive potential to climate parameters (Maraldo & Holmstrup, 2009). They can vertically migrate as a strategy for evading dry surface layers and high temperatures (Springett et p. 13

al., 1970). However, these organisms were found to have no genetic adaptation to drought, as droughtexposed populations of enchytraeids had the same high sensitivity as those from unexposed control plots (Maraldo, 2009). E. albidus has a high osmotic pressure on its body fluids and can live in environments with strong fluctuations in moisture (Maraldo, 2009), which also enables Fridericia galba to survive in soil with a water retention capacity of less than 20% (corresponding to values below -9.8 bar) for more than 49 days (Dózsa-Farkas 1973). However, additional research is needed to understand the physiological responses of enchytraeids to desiccation. Earthworms are known to be sensitive to drought and high temperatures (Bayley et al., 2010; Lima et al., 2015; González-Alcaraz & van Gestel, 2016), but they have adaptive strategies such as diapause or other physiological mechanisms, such as an increase in the free amino acid alanine, to survive long periods of drought (Holmstrup et al., 2016; Jiménez & Decaëns, 2004). Thus, in the actual context of climatic disturbances and knowing the potential effects of combined stressors on soil organisms (e.g., Bindesbøl et al., 2005; Jensen et al., 2009; Singh et al., 2019), the equilibrium between species and taxa could be modified in natural ecosystems in the future. Maraldo et al. (2006) noted that the synergistic effects of climate change factors and environmental contamination on enchytraeids have been insufficiently explored. Further studies should assess the effects of climatic parameters on functionally useful soil organisms and their relative sensitivity to desiccation or heat in realistic contexts. We also know that enchytraeids are tolerant of soil acidity (Römbke, 1991), whereas earthworm

abundance and diversity are generally low in acidic soils (Chan & Mead, 2003, Moore et al., 2013, Wu and al., 2020). Enchytraeids exhibit high tolerance to other soil properties (clay and organic matter contents) (Kuperman et al., 2006; van Gestel et al., 2011, Castro-Ferreira et al., 2012) and temperature (Holmstrup et al., 2022). However, once again, there is no real comparison to date. The combined effects of stressors related to soil properties and climate were examined by Yli-Olli and Huhta (2000), who assessed the co-occurrence of soil pH, moisture and resource addition on *Dendrobaena octaedra* (Lumbricidae) and *Cognettia sphagnetorum* (Enchytraeidae). They found that *D. octaedra* exhibited

the highest biomass in spruce stand humus at higher pH values, while *C. sphagnetorum* thrived best at higher moisture levels.

6. Conclusion

For the first time, this exhaustive synthesis provides a quantitative comparison of the sensitivity of earthworms versus enchytraeids to extrinsic factors (such as metals and pesticides) potentially affecting soil organisms. Under laboratory conditions, comparable sensitivities were found for earthworms and enchytraeids regardless of the studied species, or endpoint. Therefore, to test the effects of metals or pesticides, we could recommend testing on one of these groups in order to limit costs and labour. However, for other stressors such as climate parameters and for a thorough understanding of soil functioning, it remains crucial to consider the physiological responses of both groups due to their distinct ecological roles and potential sensitivity difference. Moreover, to fully capture the complexity of natural habitats, it would be relevant to use species inhabiting mineral soils instead of model species commonly found in organic rich habitatsloosen organic materials. For decades, researchers have been deepening their understanding of the impacts of natural and anthropic stressors on soil fauna and their related functions, which has allowed to reveal the importance of maintaining the presence and function of biodiversity in our soils. Let's keep digging in that direction.

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