

Associations between human internal chemical exposure to Persistent Organic Pollutants (POPs) and In Vitro Fertilization (IVF) outcomes: Systematic review and evidence map of human epidemiological evidence

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Associations between human internal chemical exposure to Persistent Organic Pollutants (POPs) 1 2 and In Vitro Fertilization (IVF) outcomes: Systematic review and evidence map of human 3 epidemiological evidence 4 5 **Authors** Tiphaine Lefebvre a,b,c; Thomas Fréourb,c,d; Stéphane Ploteauc,e; Bruno Le Bizeca; Jean-Philippe 6 7 Antignaca; German Cano Sancho a,*. 8 **Affiliations:** 9 10 ^a Oniris, INRAe, UMR1329, LABERCA 44037 Nantes, France 11 ^b Department of Biology and Reproductive Medicine, University Hospital of Nantes, Nantes, France 12 ^c Faculty of Medicine, University of Nantes, Nantes, France 13 ^d Center for Research in Transplantation and Immunology UMR 1064, INSERM, University of 14 Nantes, Nantes, France 15 ^e Department of Gynecology and Obstetrics, University Hospital of Nantes, Nantes, France 16 17 * Corresponding author: German Cano-Sancho, LABERCA, Oniris, INRAe, Université Bretagne-18 19 Loire, 44037 Nantes, France 20 Email: german.cano-sancho@oniris-nantes.fr 21 Tel: +33 02 40 68 78 80 22 23

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- 25 **Keywords**
- Persistent organic pollutants; biomarkers; follicular fluid; In Vitro Fertilization; infertility; systematic
- 27 review

1. Introduction

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The constant advances in chemical and technological industry have provided unquestionable advantages for humans. Nonetheless, as a side effect, a large list of synthetic compounds, whose safety is increasingly being called into question by the scientific community and public health authorities, are released to the environment. Special concern has been growing around synthetic substances structurally analogous to endogenous hormones, capable to induce mimetic or antagonistic biological effects in hormone systems, thus known as endocrine disrupting chemicals [1]. As a matter of concern, modern synthetic substances merge with mixtures of historical pollutants, banned decades ago but still present in the environment and the food chain due to their stable and bioaccumulative nature [2]. These historical chemicals known as persistent organic pollutants (POPs), include substances used in agriculture like organochlorine pesticides (OCP) or industrial applications including polychlorinated biphenyls (PCBs), which can be also released unintentionally through thermal processes like the family of dioxins [3]. Other classes of more emerging concern includes polybrominated diphenyl ethers (PBDE) used as flame retardants, and perfluoroalkylated substances (PFAS) with large industrial and consumer applications [4]. They have received international attention because of their hydrophobic or amphiphilic nature, and thus accumulated in fatty tissues of living organisms, including humans [5]. In addition, some of these congeners have been associated with notorious adverse health effects in animals and humans [1], and thus banned or strongly regulated. Despite the regulatory efforts, substantial amounts of POPs can still be detected in fatty tissues and fluids from human worldwide, including blood lipids, breast milk or adipose tissue [6,7]. The health impacts of POPs include, among others, developmental defects, metabolic dysregulation and cancers [1], but there is also an increasing concern on their effects on human reproductive health [8,9]. Due to the ability of some POPs to interfere with the hypothalomo-hypophyseal-gonadal axis, the highly hormone-sensitive female reproductive system becomes especially vulnerable to endocrine disrupting substances. For that reason, a list of POPs have been proposed to contribute to gynaecological diseases including uterine fibroids [10], endometriosis [11,12], polycystic ovary syndrome [13] longer time-to-pregnancy [14] or infertility [15,16]. Infertility is a pathology defined by the World Health Organization (WHO) as the absence of pregnancy after 12 months or more with regular unprotected sexual intercourse [17]. Affecting 8-12% of reproductive aged women worldwide, it has become a public health priority in many countries [18]. As a consequence, a growing number of couples requires Assisted Reproductive Technologies (ART) including *in vitro* fertilization (IVF), whose low success rates has been related to a number of female factors including the body burden of environmental pollutants [19]. Studies on the associations between POPs and IVF outcomes provide highly valuable information on the specific population attending ART, often with more severe forms of infertility, however it may not be generalizable to general population presenting infertility.

To date, the impact of POPs on couple fecundability has been recently reviewed [13,20] and more specifically, the impact of POPs on IVF has been narratively discussed elsewhere with special focus

specifically, the impact of POPs on IVF has been narratively discussed elsewhere with special focus on follicular fluid [21]. However, no formal systematic review on the specific between associations POPs and IVF outcomes has been conducted. In this context, the aim of the present study was to evaluate the published evidence on the association between human internal chemical exposure levels of POPs and IVF outcomes in infertile women in order to identify knowledge gaps and guiding future studies.

2. Materials and methods

We applied a systematic review approach following the National Toxicology Program Office of Health Assessment and Translation's (NTP/OHAT) Handbook for Conducting a Literature-Based Health Assessment guidelines, which provides a standardized methodology to implement the Grading of Recommendations Assessment, Development and Evaluation (GRADE) approach to environmental health assessment [22,23]. A systematic review (SR) protocol was developed a priori, calibrated and registered at PROSPERO (Register number CRD42020159267). Search and selection were performed in duplicate by two reviewers (TL and GCS). Assessment of risk of bias was also performed in duplicate by two independent members (TL and CGS) and discrepancies were discussed after proofreading the articles with a third reviewer TF. The results were structured and presented in

accordance with the Preferred Reporting Items for Systematic Reviews and Meta-Analyses (PRISMA) and Meta-Analyses of Observational Studies in Epidemiology (MOOSE) guidelines for Meta-analyses and Systematic reviews of Observational Studies [24]. The entire systematic review process was calibrated beforehand in order to refine the different steps. Specifically, a test set of articles were used to pilot and refine the search strategy iteratively developed to ensure the eligible retrievals with minor burden of non-eligible items.

2.1 Study question

The research question was: "Is exposure to persistent organic pollutants associated with in vitro fertility outcomes in women?" (See supplemental Material Section 1.2)

2.2 Search

The initial search was performed on the 8th of November 2019 using the electronic literature databases MEDLINE (http://www.ncbi.nlm.nih.gov/pubmed) and WEB of SCIENCE (WoS, https://apps.webofknowledge.com), using a comprehensive search string with the main PECO (Population; Exposure; Comparators and Outcomes) elements (See Supplemental Material Section 1.3 and Figure S1). No filters were implemented during the search, including all publication years. A follow-up search was conducted on the 20th of September 2020 for updated results based on the identical previously used search string.

2.3 Eligibility criteria and selection

Eligibility criteria for the key PECO elements were defined and summarized in the PECO statement (See Supplemental Material Section 1.3.1). Were excluded conference and review papers, articles not written in English, studies without quantitative exposures and studies overlapping information with another publication.

Participants/ population: Only women undergoing in vitro fertilization treatment were studied without restrictions on country, race and religion.

Exposure(s): Exposure to one or multiple POPs listed in the Stockholm Convention (including derivatives or isoforms) [25], based on internal concentrations and surrogate biomarkers. Environmental measures or indirect measures were not considered. The exposure must be measured individually using direct validated based on biomonitoring methods, irrespectively of the biological matrix used (i.e. serum, FF).

Comparator(s). Reference groups of population exposed at lower levels of POPs than the rest of population groups were considered.

Outcome(s): Primary outcome: pregnancy rate (PR). Secondary outcome 1: number of oocytes retrieved (NOR). Secondary outcome 2: fertilization rate (FR). Secondary outcome 3: embryo quality (EQ). Secondary outcome 4: live birth rate (LBR).

Only peer-reviewed reports containing original data were included. Conference and review papers were excluded from the analysis. Studies with overlapping information with another publication were also excluded (the most complete publication originating from the same cohort was the only one considered). Study selection was carried out using a two-stage sequential process, with step-1 based on title/abstract agreement to the eligibility criteria and step-2 based on full-text examination. All references were screened and assessed by two independent reviewers (TL and GCS). Three additional reviewers (SP, TF and JPA) checked the final reference list and made the final decision in case of disagreement (See Supplemental Material Section 1.3).

2.4 Data extraction

Data from included records were extracted using a predefined form (Supplemental Material Section 1.3.4) and included: year, study design, recruitment strategy, inclusion/exclusion criteria, dates of study and sampling time frame, geography, number of participants, age, race/ethnicity, type of POP and compartments (serum, follicular fluid), exposure assessment with methodological details and units, exposure levels, type of infertility, number of oocyte retrieved, embryo quality, clinical outcome measure (pregnancy rate, live birth rate), adjustment with patients' characteristics, statistical approach and main conclusion.

2.5 Risk of bias assessment

The 'NTP/OHAT Risk of Bias Rating Tool for Human and Animal Studies' was used to classify each individual study in a tier (1 to 3) of risk of bias (See Supplemental Material Section 1.3.5). The NTP/OHAT's risk of bias tiered approach considered some key questions or risk of bias domains of higher relevance to establish the classification criteria for each individual study. The risk of bias domains and related questions were: **A)** Confounding Bias: (Key question 1) Did the study design or analysis account for important confounding and modifying variables? **B)** Attrition/Exclusion Bias: Were outcome data incomplete due to attrition or exclusion from analysis? **C)** Detection Bias: Can we be confident in the exposure characterization? (Key question 2) Can we be confident in the outcome assessment (Key question 3)? **D)** Selective Reporting Bias: Were all measured outcomes reported? **E)** Other sources of bias: Were there other threats to internal validity? The risk of bias evaluation was entirely conducted and managed in Health Assessment Workplace Collaborative (HAWC) to ensure transparency, publicly available at https://hawcproject.org/assessment/1021/.

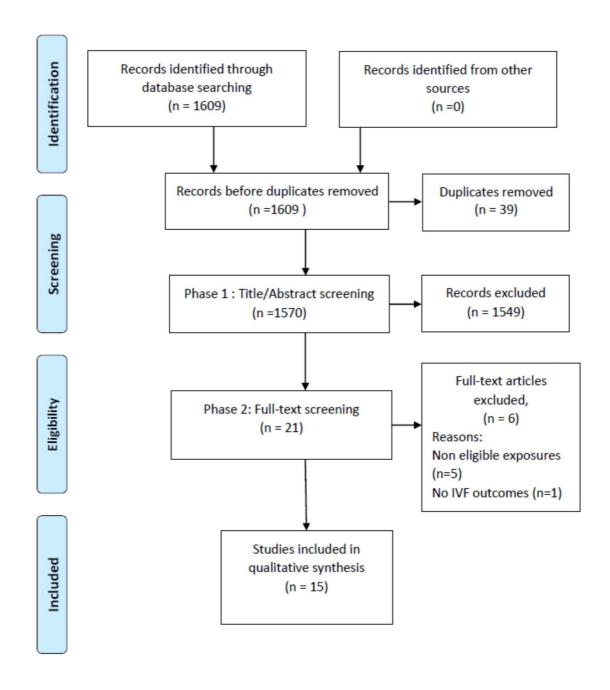
2.6 Data synthesis

For facilitated interpretation and summarized result purposes, chemicals were grouped to four main families of POPs (PFAs, PCBs, OCPs and PBDEs) and five main IVF outcomes (Number of oocyte retrieved, Fertilization rate, Embryo quality, Pregnancy rate and live birth rate). A quantitative synthesis with meta-analysis was unfeasible due to the low number of studies reporting quantitative effect estimates in common metrics for the exposure-outcome pairs. Thus, we conducted qualitative synthesis presented in summary tables with the main overall direction of results for each exposure-outcome pair. In order to harmonize the outcome language, we considered "Pregnancy rate" a consolidated outcome combining the results from outcomes reported as "Pregnancy rate", "Clinical pregnancy rate" and "Failed implantation". Indeed, "Pregnancy rates" and "Failed implantation rates" refer to similar event in reverse metrics, we therefore inversed the direction of the "Failed implantation" effect to ease the visualization of overall effects. The overview of gathered evidence across exposures and outcomes was displayed in an evidence map, an emerging feature that facilitates the visualization of heterogeneous evidence and the identification of research gaps [26].

3. Results

3.1 Source and selection of key studies

The process applied for sourcing and selection key studies is summarized in the PRISMA flow diagram given in Figure. 1. The initial search resulted in 1609 items, reduced to 1570 after duplicated removal of pooled retrievals from Pubmed and WoS. Among the 21 studies retained for the full-text screening, 15 studies were finally eligible for data extraction and synthesis. Manual searches did not contribute additional manuscripts to be included.



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3.2 Study characteristics

The characteristics of included studies are summarized in Table 1. The records included in the main body of evidence comprised 12 prospective cohort studies and 3 case control studies. Taking into account that 5 records focused on the same 2 populations, this systematic review covered a total of 2033 individuals. Study populations were generally modest in size, with 11 studies including <100 samples (range 16-99) [27-37], while others ranged from 215 [38] to 765 participants [39-41]. Six out of the 15 studies were conducted in the United States [30,34,35,38-40], while others were conducted in various countries. Concerning the exposures, OCPs (n=8 studies) and PCBs (n=7 studies) were the most evaluated POPs among the selected studies, followed by PFAs (n=4) and PBDEs (n=2). The number and type of particular exposure markers varied across studies for all POP families, for instance, the number of monitored PCB congeners varied from 2 [27] to 57 [40], but most of studies ranged between 2 and 8 PCBs. The most commonly analyzed markers among the OCPs substance class dichlorodiphenyltrichloroethane metabolite were (DDT) and the dichlorodiphenyldichloroethylene (DDE), followed by Lindane, Dieldrin, Mirex, Endosulfan, hexachlorobenzene (HCB), hexachloroethane (HCE), 1,2,4-trichlorobenzene (TCB) and betahexachlorocyclohexane (β-HCH). In case of PBDEs, both selected studies included at least the most abundant exposure markers commonly reported for this substance class (i.e. BDE47, 99, 100, 153, 154) [42], whereas Ingle et al. 2020 also analyzed the hydroxylated forms [38].

Most studies reported the simultaneous determination of exposure biomarkers in FF and serum. Serum was collected at different phases of IVF stimulation: prior, during stimulation [35,39] or on the day of oocyte retrieval [27,28,32,33]. Concerning the IVF outcomes, 9 studies evaluated the association between POP and the number of oocytes retrieved [27–29,31,32,34–36,38], 11 studies analyzed fertilization rate [27–29,31–35,37,38,41], 6 evaluated embryo quality [27,28,31–34], 12 analyzed pregnancy rate or failed implantation [27–32,34,36,38–41], and 5 analyzed live birth rate [32,34,35,38,40].

| Reference | | Popu | Population | | Outcomes | Confounders | Exposure as | sessment | | Statistics |
|------------------------------------|------------|--------------------------|----------------------------|---------------------|---|---|--|--|-----------------------|---|
| | Dates | Country | Age ^a | n | | | Chemicals ^b | Matrix (Units) | Analytical Method | |
| Al Hussaini et al 2018[36] | 2010- 2013 | Egypt | 31.5 ± 6.2 (20-38) | 94 | N° oocytes Pregnancy rate | NR | PCB 28,52,138,180 OCP: Lindane, DDT | FF (µg/L) | GC/MS SIM | Linear regression |
| Al-Saleh et al 2009 [41] | 2002-2003 | Saudi Arabia | 31.8 ± 5.1 (19-50) | 619 | Fertilization rate Pregnancy rate | Age, residence the former province, location of former, school, fish consumption | 3 OCPs (DDE) | Serum FF (µg/L) | GC-MS-ESI | Logistic Regression (OR) |
| Bloom et al 2017 [34] | 2007- 2008 | California US | 35.6 (28-42) | 32 | N° oocytes Fertilization rate Quality embryo Pregnancy rate Live birth rate | BMI, smoking, race, age | 43 PCB (grouped in oestrogenic and antioestrogenic), OCP: DDT, DDE | FF (ng/ml) | GC/MS-SIM | MLR (β/RR) Negative binomic Poisson regression GEE |
| Governini et al 2011 [29] | NR | Italy | 33.6 ±2.6 | 16 | N° oocytes Fertilization rate Pregnancy rate | None | PFAs | FF | HPLC-MS ESI | t-test, Chi square Correlation analysis |
| Ingle et al 2020 [38] | 2005-2016 | Massachuss ets US | Median:35 | 215 (330 cycles) | N° oocytes Fertilization rate Pregnancy rate Live birth rate | Lipids, age, BMI, race, year of collection, IVF protocol, FSH (previous treatment, education, E2 level, infertility diagnosis, ICSI, endometrial thickness) | PBDE 47,99,100,153 and 154 4 OH-BDE metabolites | Serum (ng/g) | GC-MS LC-MS/MS | Multivariate mixed models Poisson/logbin GEE |
| Jirsová et al 2010 [33] | 2003-2004 | Czech Republic | 30.9 ± 3.7 (25.3-41.3) | 99 | N° oocytes Fertilization rate Embryo quality Pregnancy rate | NR | 36 PCB (PCB 44,81,101,105,118,123,158) 3 OCP (DDT, DDE) | FF (ng/g of lipid) | GC-MS | Wilcoxon/Krusk 1 Wallis tests MLogR /Poisson (OR) |
| Johnson et al 2012 [30] | 1994- 2003 | Boston, US (3sites) | 38 ±3.8 (27-44) | 65 | Failed implantation | Age, BMI (ethnicity, smoking status, serum levels of PBCs) | PBDE (28, 47, 99, 100, 153, 154, 183, 209) | Serum (ng/g) FF | GC-MS NCI | MLogR (OR) |
| Kim et al 2020 [37] | 2006-2011 | Australia | 35 ± 4 (23-42) | 97 | Fertilization rate | Age | 32 PFAs (8 detected) | FF | HLPC- MS/MS ESI | ANOVA |
| Mahalingaiah et al 2012 [39] | 1994- 2003 | Boston, US (3sites) | 35.9±4.21 | 720 (774 cycles) | Failed implantation | TSL, site, study phase, ethnicity, previous live birth, age, BMI, smoking, gonadotropins, protocol, ICSI, N° embryos, diagnosis | HCB, p-p'DDE, Total DDT (p,p'-DDT, o,p'DDT, p- p'DDE and o,p'DDE) | Serum (ng/g of serum and ng/g of lipid) | GC-MS ESI | MLR/MLogR Random effects Q4-2 vs Q1 (OR) |
| McCoy et al 2017 [35] | 2013- 2014 | South Carolina, US | 33.7 ± 4.5 | 26–31 | N° oocytes Fertilization rate Live birth rate | Age (BMI) | 15 PFAs (6 detected: PFHxA, PFOA, PFNA, PFDA, PFUnA, PFOS) | FF Plasma (ng/g) | LC-MS/MS ESI- | Partial correlatio analysis (age-corrected) |
| Meeker et al 2011 [40] | 1994- 2003 | Boston, US (3sites) | 35.9 ±4.2 | 720 (774 cycles) | Failed implantation Live birth rate | TSL, site, study phase, ethnicity, previous live birth, age, BMI, smoking, gonadotropins, protocol, ICSI, N° embryos | 57 PCB (PCB 11, 138 and 153; on 3 bioactivity-based groups; and on sum of PCB) | Serum (ng/g serum and lipid) | GC-ECD | MLR/MLogR Random effects Q4-2 vs Q1 (OR) |

| | 2008-2009 | Belgium | 34.5 ± 4.4 | 20 | N° oocytes | Age, estradiol, BMI, male | 26 PCB (PCB 118 138 153 170 180) | FF | GC-MS | PCA-MLR |
|----------------|-----------|---------|----------------|------------|--------------------|---------------------------|------------------------------------|--------------|-----------|-----------------|
| Petro et al | | | (25.2-42.8) | | Fertilization rate | subfertility, ICSI | 7 PBDE (Not detected >50%) | (pg/ml) | ECNI | (OR) |
| 2012 [32] | | | | | Embryo quality | | OCPs (HCB, β-HCH, DDE) | Serum | | |
| | | | | | Pregnancy rate | | | (pg/ml and | | |
| | | | | | Live birth rate | | | ng/g lipid) | | |
| Petro et al | 2008-2009 | Belgium | 34.6 ± 4.4 | 20 | Fertilization rate | Age, estradiol, BMI, male | 14 PFAs (PFOA, PFOS, PFNA, | FF | UPLC | PCA- MLR |
| 2014 [31] | | | (25.2-42.8) | | Embryo quality | subfertility, ICSI | PFHxS) | Serum | ESM-MS/MS | (β/OR) |
| | | | | | | | | (ng/ml) | | |
| Weiss et al | NR | Germany | NR | 21 Germany | N° oocytes | None | 5 PCB (PCB 138, 153) | FF | GC | ANOVA |
| 2006 [27] | | | | | Fertilization rate | | 8 OCPs (DDT, DDE, Dieldrin) | Serum | | t-test |
| | | | | | Embryo quality | | | $(\mu g/kg)$ | | |
| | | | | | Pregnancy rate | | | | | |
| Younglai et al | NR | Canada | 28-38 | 21 | N° oocytes | None | 35 PCB (PCB 49, 99, 138, 153, 180) | Serum | GC-MS | t-test |
| 2002 [28] | | | | | Fertilization rate | | 36 OCPs (Endosulfan, DDE, HCE, | FF | | Linear/Logistic |
| | | | | | Embryo quality | | TCB, Mirex) | (pg/ml) | | Regression (OR) |
| | | | | | Pregnancy rate | | | | | |

^a Mean ± standard deviation (range)

Abbreviations: β, beta; BMI, body mass index; DDE, dichlorodiphenyldichloroethylene; DDT, dichlorodiphenyltrichloroethane; ESI, Electro-Spray Ionization; ECNI, electron-capture negative ionization; ESM, electrospray interface operating in negative ion mode; FF, follicular fluid; GC Gas Chromatography; GC-ECD, Gas Chromatography with dual micro-Electron Capture Detection; GEE, generalized estimating equation; HCB, hexachlorobenzene; HCE, hexachloroethane; HCH, hexachlorocyclohexane; HPLC, High Performance Liquid Chromatography; IVF, In Vitro fertilization; LC-MS/MS, Liquid Chromatography tandem Mass Spectrometry; Logbin, logbinomial; MS Mass Spectrometry; MLR/MLogR, multivariate linear/logistic regression; N°, number; NCI negative chemical ionization; NR, Not reported; OR Odds Ratio; PCA Principal Component Analysis; PCB Polychlorinated Biphenyls; PFA Perfluoroalkyl Acids; PFDA, perfluorodecanoic acid; PFHpA, perfluorohexane sulfonate; PFNA, perfluorononanoic acid; PFOA, perfluorooctanoic acid; PFOS, perfluorooctane sulfonic acid; PFUnDA, perfluoroundecanoic acid; Q Quartile; SPE Solid Phase Extraction system; SIM Selective Ion Monitoring Mode; TCB, 1,2,4-trichlorobenzene; TSL, total serum lipids; UPLC Ultra Performance Liquid Chromatography

^b Analyzed chemicals (detected chemicals included in statistical analysis)

3.3 Association between PFAs exposure and IVF outcomes

The evidence on the associations between exposure to PFAs and IVF outcomes is summarized in Table 2, either for sums of congeners or selected ones like perfluorooctanoic acid (PFOA), perfluorooctane sulfonic acid (PFOS) or perfluorononanoic acid (PFNA). The four studies reported PFAs concentration levels in FF while only 2 reported PFAs in serum [33,35]. No significant association was found between PFAs and number of oocyte retrieved [29,35]. Concerning fertilization rate, results were inconsistent across the different studies, with reported negative [29], positive [33], or no association [35,37]. The one available study on embryo quality reported a positive association between embryo quality and follicular level of PFAs [33]. No associations among PFAs and pregnancy rate nor live birth rate were reported [29,35].

Table 2. Summary of main associations between internal concentrations of PFAS and IVF outcomes.

| Reference (n) | Detected congeners | |] | Level FF | | | Leve | ls Serum/F | Plasma | Main effects on IVF outcomes | | | | | |
|--|--------------------|-----------|-----------|----------|--|-----------|-----------|------------|---------------|------------------------------|------------|------|-----------|--------------------------|--|
| | O | Units | DF (%) | Mean | SD (range) | Unit s | DF (%) | Mean | SD (range) | NOR | FR | EQ | PR | LBR | |
| Governini et al 2011 [29] (n=16) | ∑PFAs | ng/m L | NR | 6.65 | 3.54 | | | | | 0 | 111 | NA | \otimes | NA | |
| Kim et al 2020 [37] | PFOS | ng/ml | 98 | 4.8 | 3.1 | | | | | | | | | | |
| (n=97) | PFHxS | _ | 98 | 1.7 | 2.7 | | | | | | | | | | |
| | PFHpS | | 96 | 0.1 | 0.1 | | | | | | | | | | |
| | PFOA | | 98 | 2.4 | 1.7 | | | | | NA | \Diamond | NA | NA | NA | |
| | PFNA | | 97 | 0.5 | 0.3 | | | | | INA | G | INA | INA | INA | |
| | PFDA | | 98 | 0.2 | <lod< td=""><td></td><td></td><td></td><td></td><td></td><td></td><td></td><td></td><td></td></lod<> | | | | | | | | | | |
| | PFHpA | | 90 | 0.01 | 0.008 | | | | | | | | | | |
| | PFUnDA | | 97 | 0.1 | 0.007 | | | | | | | | | | |
| Mc Coy et al 2017 [35] (n=36) | ∑PFAs | ng/g | 100 | 10.1 | 0.8 | ng/g | 100 | 12.6 | 1.1 | 0 | 0 | NA | NA | \(\rightarrow\) * | |
| Petro et al 2014 [33] | PFOA | ng/m | 92 | 1.8 | (0.3-3.3) | | | | | | | | | | |
| (n=20) | PFOS | Ĺ | 97 | 7.5 | (0.1- | | | | | | 1 | 1 | | | |
| | PFNA | | 76 | 0.4 | 30.4) | | | | | NA | (PC1 | (PC1 | NA | NA | |
| | PFHxS | | 76 | 0.3 | (0.2-2.1) (0.1-1.4) | | | | | |) |) | | | |

Main effect annotation:

No statistical associations; *Data not shown; ↓↓↓ statistically significant decrease (p<0.001); ↑ statistically significant increase (p<0.05). Abbreviations: DF, detection frequency; EQ Embryo quality; FR Fertilization Rate; FF Follicular Fluid; LBR, live birth rate; NA, not assessed; NOR number of oocytes retrieved; NR, not reported; P plasma; PC1, principal component 1; PFAs, perfluoroalkyl substances; PFDA, perfluorodecanoic acid; PFHpA, perfluoroheptanoic acid; PFHpS, perfluoroheptanesulfonic acid; PFHxS, perfluorohexane sulfonate; PFNA, perfluorononanoic acid; PFOA, perfluorooctanoic acid; PFOS, perfluorooctane sulfonic acid; PFUnDA, perfluoroundecanoic acid PR pregnancy rate; SD standard deviation

3.4 Associations between PCBs exposure and IVF outcomes

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analyzed PCB in FF [31,33,36], 1 in serum [40] and 3 studies in paired serum and FF samples 231 232 [27,28,32]. Most studies evaluated the associations with number of oocytes retrieved showing mainly no statistical associations and only two studies found statistically significant associations in opposite 233 directions for different exposure markers. Bloom et al. [34] noted that higher PCB151 was associated 234 235 with fewer oocyte retrieved [RR -0.54 (-0.93; -0.13) p<0.05] but not for the rest of congeners. 236 Conversely, Al-Hussaini et al. [36] found that PCB28 was positively associated (adjusted ρ = 0.07, p = 237 0.0001). Similarly, for the five studies that evaluated fertilization rate, three of them found null results (detailed data not shown in the original publications) and two studies significant yet divergent results. 238 Whereas Bloom et al. [34] found positive association between PCB105 and oocyte fertilization, Petro 239 240 et al. [32] associated it negatively (p < 0.0001) with the first principal component scores representing 241 the overall contamination of FF (including PCBs and OCPs). This discrepancy must be analyzed with care because the analysis of the impact of one pollutant is different from the analysis of a mixture of 242 pollutants Two out of four studies that analysed the associations between embryo quality and PCBs, 243 reported significant inverse associations [32,34] and the other two reported no association [28,31]. The 244 245 7 included studies evaluated impact of PCBs on pregnancy rate or failed implantation rate. Meeker et al. [40] found that serum PCB 153 and sum of PCBs were significantly correlated with failed 246 implantation. In accordance, Bloom et al reported that higher FF PCB 28 and 66 and sum of PCB 247 248 (total and oestrogenic) were associated with lower embryo implantation. But PCB146 were associated 249 with higher implantation rate, but the authors remarked a very low detection frequency for this 250 compound (6.2%). In turn, Younglai et al. [28] reported a positive correlation between PCB 49 level in 251 FF and pregnancy rate; while Petro et al., Weiss et al., and Al-Hussaini et al., reported no associations 252 (data not shown) [27,32,36]. Jirsova et al., did not find statistically significant associations, however 253 the authors highlighted that associations with PCB47 were marginally significant (OR=0.73, p<0.1) 254 [31]. Only 3 studies evaluated impact of PCBs on live birth rate reporting no correlation in one study

The association between PCBs exposure and IVF outcomes is summarized in Table 3. Three studies

255 [32] and negative correlation in the 2 others [34,40] for part of the PCBs analysed (PCB 153, 66, 74, 256 101 and sum of PCB).

Table 3. Summary of main associations between internal concentrations of PCBs and IVF outcomes.

| Reference (n) | Detected congeners | | Le | vels FF | | | Levels | s Serum/Plasn | na | | | Main effe | | |
|---------------------|--------------------|----------|-----------|-----------------|--------------|----------|-----------|---------------|---------------|---------------|----------------------|------------|--------------------|-------------|
| ` ' | S | Units | DF (%) | Mean | SD (range | Units | DF (%) | Mean | SD (range) | NOR | FR | EQ | PR | LBR |
| Al Hussaini | PCB 28 | ng/mL¤ | NR | 0.05 | 0.09 | | | | | | | | | |
| et al. 2018 | PCB52 | ng/mb | 1111 | 0.4 | 0.05 | | | | | 111 | | | | |
| [36] | PCB 138 | | | 0.15 | 0.03 | | | | | (PCB28) | NA | NA | \bigcirc * (CPR) | NA |
| (n=94) | PCB 180 | | | 0.10 | 0.02 | | | | | (I CB20) | | | | |
| Bloom et al. | ΣPCBs | ng/mL¤ | 1- | ΣPCBs | 0.45 | | | | | 1 | 1 | 111 | 1 | 1 |
| 2017 [34] | $\sum e-PCBs$ | 116,1112 | 100 | 1.22 | 0.22 | | | | | (PCB15 | (PCB | (PCB8 | (PCB | (PCB66 |
| (n=32) | Σae-PCBs* | | 100 | Σ e-PCBs | 0.06 | | | | | 1) | 105) | 7, 149, | 28,66, | 74, |
| (02) | <u>_</u> | | | 0.63 | 0.00 | | | | | -/ | / | ∑PCB | Σ PCBs, | 101, ∑e |
| | | | | ∑ae- | | | | | | \Diamond | \Diamond | s, | $\sum e-PCBs$) | PCBs) |
| | | | | PCBs | | | | | | $(\sum PCBs)$ | $(\sum PC$ | ∑ae- | 1 | , |
| | | | | 0.08 | | | | | | • | Bs) | PCBs) | (PCB146) | ⊘ (∑PCBs |
| Jirsova et al. | PCB 44 | ng/g of | NR | 1.8 | 2.5 | | | | | | | | | |
| 2010 [31] | PCB 47 | lipid | | 13.1 | 34 | | | | | ⊗* | \(\right)\) * | ○ * | ⊗* | NA |
| (n=99) | PCB 101 | | | 8.1 | 9.1 | | | | | . | G . | O . | Q . | INA |
| | PCB 158 | | | 10.2 | 10.9 | | | | | | | | | |
| Meeker et al. | PCB 153 | | | | | ng/g of | NR | 0.23 / 45.4 | NR | | | | | |
| 2011 [40] | ∑PCB | | | | | serum | | 1.32 / 257 | | NA | NA | NA | ↓ (I-FIR) | 1 |
| (n=827) | | | | | | - ng/g | | | | IVA | IVA | IVA | + (1-11K) | * |
| | | | | | | of lipid | | | | | | | | |
| Petro et al. | PCB118 | ng/mL | 68 | 0.06 | 0.01 | ng/mL | 100 | 0.04 / 8.1 | 0.03 / 0.9 | | | | | |
| 2012 [32] | PCB 138 | | 98 | 0.05 | 0.03 | - | 100 | 0.12 / 23.5 | 0.06 / 9.2 | | 111 | | | |
| (n=20) | PCB 153 | | 98 | 0.07 | 0.04 | ng/g | 100 | 0.20 / 39.2 | 0.11 / 17.6 | \Diamond | (PC1 | ļ | ○ (CPR) | \Diamond |
| | PCB 170 | | 85 | 0.02 | 0.01 | of lipid | 100 | 0.04 / 8.7 | 0.02 / 4.4 | O | (1 (1 | (PC1) | G (CIR) | Ŭ |
| | PCB 180 | | 93 | 0.05 | 0.03 | | 100 | 0.11 / 21.9 | 0.06 / 10.9 | | , | | | |
| | ∑PCBs | | | 0.21 | 0.14 | | | 0.56 / 110 | 0.29 / 48.9 | | | | | |
| Weiss et al. | ∑PCB138 | ng/g¤ | >90 | 0.26 | 0.02 | ng/g | NR | 0.25 | 0.05 | _ | _ | | _ | |
| 2006 [27] (n=21) | +153 | | | | | | | | | ⊘ * | ⊘ * | NA | ⊗* | NA |
| Younglai et | PCB 49 | ng/mL¤ | >50 | 0.62 | 0.07 | ng/mL | <50 | NR | NR | ○ * | 0* | ○ * | 1 | NA |
| al. 2002 [28] | PCB 99 | | < 50 | NR | NR | | >50 | 0.77 | 0.02 | G | G | 0 | $\Diamond *$ | 11/1 |

| (n=21) | PCB 138 | <50 | NR | NR | >50 | 0.24 | 0.03 | \(\rightarrow\) * |
|--------|---------|-----|-------|------|-----|------|------|--------------------------|
| | PCB 153 | >50 | 0.073 | 0.07 | >50 | 0.24 | 0.03 | ⊗* |
| | PCB 180 | >50 | 0.062 | 0.05 | >50 | 0.14 | 0.02 | $\otimes *$ |

a Statistical models build between the first principal component, including the mixture of PCBs and the outcomes using multivariate regression models. Main effect annotation: ○ No statistical associations *Data not shown; ↑ Statistically significant increase (p<0.05) ↑↑↑ (p<0.001) ↓ Statistically significant decrease (p<0.05) ↓↓↓ Statistically significant decrease (p<0.001). [¤]No adjustement on lipid

Abbreviations: ae-PCB antiestrogenic PCBs; CPR clinical pregnancy rate; DF, detection frequency; e-PCB estrogenic PCBs; ER embryo quality; FR Fertilization Rate; FF Follicular Fluid; I-FIR inverse of failed implantation rate; LBR live birth rate; NA Not assessed; NOR number of oocytes retrieved; PR pregnancy rate; S serum; SD standard deviation.

* Σ PCBs (PCB 28,44,49, 52,66,74,87,99,101,105,110,118,138,146,149,151,153,170, 180,183,187); Σ e-PCBs (PCB 44,52,49, 99,101,110,15) Σ ae-PCBs (PCB 105,118)

3.5 Associations between OCPs and IVF outcomes

Main findings from the 8 studies evaluating the association between OCPs exposure and IVF outcomes are summarized in Table 4. All 6 studies evaluating the impact of OCPs on oocyte number retrieved no statistically significant association. Six studies evaluated the association between OCPs and fertilization rate. Among them, 2 found a statistically significant decrease [28,32]. Concerning embryo quality, only one study out of 6 reported a negative association with the overall exposure estimate [32]. The levels of DDT or the main metabolite DDE were not found to be associated with pregnancy rate in any of the 8 studies considered. A statistically significant negative association was found for HCB and pregnancy rate [39], and for lindane and lower implantation rate [36]. Only 2 studies evaluated the impact of DDT (or DDE) on live birth rate leading to discordant conclusions. Bloom et al., reported a positive association between DDE and live birth rate using the exposure contrast in quartiles, yet this association was not confirmed with continuous log-transformed DDE [34]. Petro et al, reported no association [32].

Table 4. Summary of main associations between internal concentrations of OCPs and IVF outcomes.

| Reference (n) | Detected congeners | | L | evels FF | | | Levels S | erum/Plasn | na | | Main effects on IVF outcomes | | | | |
|--|--|----------------|---------------------------------|--------------------------------|----------------------------------|----------------|---------------------------------|-------------------------------|------------------------------|-------------------|---------------------------------|------------|------------|--------|--|
| | | Units | DF (%) | Mean | SD (range) | Units | DF (%) | Mean | SD (range) | NOR | FR | EQ | PR | LBR | |
| Al Hussaini et al. 2018 [36] (n=94) | Lindane DDT | ng/mL | NR | 0.41 0.02 | 0.17 0.04 | | | | | 0 | NA | NA | (IR) | NA | |
| Al-Saleh et al. 2009 [41] (n=619) | DDE | ng/mL | 50.6 | 0.41 | 1.35 | ng/mL | 77.7 | 1.65 | 1.63 | NA | 0 | NA | (CFR) | NA | |
| Bloom et al. 2017 [34] (n=32) | DDT DDE | ng/mL | 21 100 | 0.01 0.68 | 0.03 0.92 | | | | | 0 | 0 | 0 | 0 | †(DDE) | |
| Jirsova et al. 2010 [31] (n=99) | DDE DDT | ng/g fat | NR | 3303.3/7 5.9 | 4205.2/75.9 | | | | | ○ * | ○ * | ○ * | ○ * | NA | |
| Mahalingaiah et al. 2012 [39] (n=774) | HCB DDE ∑DDT | | | | | ng/g | 100 | 0.088 1.09 1.22 | | NA | NA | NA | (HCB | NA | |
| Petro et al. 2012 [32] (n=20) | HCB DDE β-HCH | ng/mL | 93 98 78 | 0.03 0.39 0.03 | 0.02 0.35 0.04 | ng/mL | 100 100 95 | 0.06 0.62 0.04 | 0.03 0.41 0.02 | 0 | III (PC1) | (PC1) | ○ (CPR) | 0 | |
| Weiss et al. 2006 [27] (n=21) | DDT+DDE Dieldrin | ng/g | NR | 0.58 0.03 | 0.10 0.01 | ng/g | NR | 0.91 0.02 | 0.13 0.01 | \(\right)* | 0 | 0 | 0 | NA | |
| Younglai et al. 2002 [28] (n=21) | DDE Endosulfan Mirex HCE TCB | ng/g, ng/mL | >50 <50 >50 <50 <50 | 2.7 NR 3.6 0.2 0.2 | 1.5 NR 0.6 0.03 0.02 | ng/g, ng/mL | >50 <50 <50 <50 <50 | 7.9 0. 8 NR NR NR | 5.6 0.2 NR NR NR | ○ * | (DDE) | ○ * | ○ * | NA | |

Abbreviations: CPR clinical pregnancy rate; DDE, dichlorodiphenyldichloroethylene; DDT, dichlorodiphenyltrichloroethane; DF, detection frequency; ER embryo quality; FR Fertilization Rate; FF Follicular Fluid; HCB, hexachlorobenzene; HCE, hexachloroethane; HCH, hexachlorocyclohexane; IR implantation rate; LBR live birth rate; NA

Not assessed; NOR number of oocytes retrieved; TCB, 1,2,4-trichlorobenzene; PC1, principal component 1; PR pregnancy rate; S serum; SD standard deviation. Main effect annotation: So No statistical associations *Data not shown; Statistically significant increase (p<0.05) 11 (p<0.001) Statistically significant decrease (p<0.05) 11 (p<0.001) Statistically significant decrease (p<0.05) 11

3.6 Associations between PBDE exposure and IVF outcomes

The association between PBDE exposure and IVF outcomes is summarized in Table 5. Only 2 studies reported the association between PBDEs and IVF outcomes [30,38]. A similar study attempted to quantify PBDEs by finally excluded them from statistical analysis because the detection frequency was below 50% [32]. Johnson et al., analysed BDE28, 47, 99, 100, 153, 154 and 183 in serum and FF, finding detection frequencies over 50% for BDE47, 99, 100 and 154 in serum and BDE47 in FF [30]. The authors reported significantly higher failed implantation risk with high levels of BDE 53 in FF [Age and body mass index (BMI) adjusted OR 95%CI: 10 (1.9-51); *p*=0.0006]. They did not analyse the number of oocytes retrieved, fertilization rate, embryo quality, neither live birth rate. Ingle et al., reported an unexpected positive association between implantation, clinical pregnancy and live birth rates and BDE 153 (Relative Risk [RR] = 1.16; 95% CI: 1.16, 1.36, RR = 1.25; 95% CI: 1.16, 1.35, RR = 1.26; 95% CI: 1.13, 1.41, respectively) and 5-OH-BDE47 (RR = 1.56, 95% CI: 1.14, 2.14, RR = 1.56; 95% CI: 1.12, 2.18, and RR = 1.84, 95% CI: 1.26, 2.68, respectively) [38]. Nonetheless, no associations were observed for PBDEs and OH-BDEs with IVF outcomes prior to implantation.

Table 5. Summary of main associations between internal concentrations of PBDEs and IVF outcomes.

| Reference (n) | Detected congeners | |] | Levels FF | | L | asma | Main effects on IVF outcomes | | | | | | |
|---------------------------------------|---|----------------|----------------------------------|---|--|----------------|----------------------------------|---|---|-----|----|----|---|-----------------------------------|
| | | Units | DF (%) | Mean | SD (range) | Units | DF (%) | Mean | SD (range) | NOR | FR | EQ | PR | LBR |
| Johnson et al. 2012 (n=62) [30] | BDE 28 BDE47 BDE 100 BDE 99 BDE 154 BDE 153 ∑BDEs | ng/g- ng/mL | 39 70 38 47 44 39 | 0.001 0.026 0.006 0.014 0.003 0.007 0.058 | ND-0.001 ND-0.012 ND-0.002 ND-0.003 ND-0.002 ND-0.002 0.002-0.03 | ng/g- ng/mL | 43 90 83 70 86 44 | 0.3 7.1 1.4 2.1 2.1 1.9 15.8 | ND-0.9 2.3-9.7 0.3-2.0 ND-3.3 0.8-2.9 ND-1.2 5.5-20.8 | NA | NA | NA | ↓↓ (1/FIR) (BDE53) | NA |
| Ingle et al. 2020 [38] (n=215) | BDE 47 BDE 99 BDE 100 BDE 153 BDE 154 ∑BDEs ∑OH-BDE | | | | | ng/g | 83 59 71 94 43 | 28.1 5.5 5.1 26.3 7.3 86.3 0.33 | 13.4-57.8 2.5-10.6 2.3-10.8 13.0-47.7 4.8-12.2 44.2-157.8 0.16-0.74 | 0 | 0 | NA | ↑(BDE53, ∑ OH-BDE) (IR and CPR) | † (BDE 153, Σ OH BDE) |

Main effect annotation: ○ No statistical associations *Data not shown; ↑ Statistically significant increase (not reported p-value) ↓↓ Statistically significant decrease (p<0.01). Abbreviations: CPR clinical pregnancy rate; DF, detection frequency; ER embryo quality; FR Fertilization Rate; FF Follicular Fluid; IR implantation rate; LBR live birth rate; NA Not assessed; ND, not detected; NOR number of oocytes retrieved; p, percentile; PR pregnancy rate; S serum; SD standard deviation.

3.7 Correlation between blood and follicular levels of POPs

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Most PFAs detected in FF appeared to be highly correlated to those levels in plasma (ρ =0.81-0.99) or serum (ρ =0.51-0.65) and remained stable over the course of ovarian stimulation [33,35]. A positive correlation was reported between follicular and serum levels of PCB [28,32] and OCPs [27,28,32,41]. Johnson et al. reported moderate correlations between serum and FF concentrations of BDE 28,47, 100 and 154 (Kendall's tau correlation = 0.29-0.38, p < 0.005) but BDE 99 and 153 were not correlated between the 2 matrices (Kendall's tau correlation<0.2, p> 0.05) [30].

3.8 Risk of bias assessment and evidence gap mapping

Summary of risk of bias evaluations for each individual study can be found in Figure 2. Specific evaluations for each chemical family can be found in the Supplemental Figures S2-S4 and HAWC (https://hawcproject.org/summary/assessment/1021/visuals/). Globally, the most concerning domains affected by the evaluation were the confounding bias and the low statistical power for a substantial number of studies. Considering the consistent evaluations across outcomes in the same study we judged minor differences to support stratified risk of bias assessment for each individual IVF outcome. In the case of PFAs, the overall body of evidence was classified as "Very serious" risk of bias considering that 2 studies were classified in Tier 2 and 2 in Tier 3 according the NTP/OHAT guidelines. Small sample size (n<100), lack of adjustment for confounding variables and underreporting of outcome methodology and results were the major threats of confidence. The body of evidence on PCBs compiled with 7 studies was judged to be at "Serious" risk of bias. Lack of adjustment for confounding variables, low sample size for 4 out of 7 studies (n<35) and underreporting of the analytical methodology for exposure biomarker assessment were the main issues. The group of 8 studies on OCPs, mainly focusing on DDT or its metabolite DDE, were also classified at "Serious" risk of bias. The 2 studies on PBDEs were classified in Tier 1 and 2, penalized by the underreporting of exposure assessment methods. The overall evidence is displayed in Figure 3 for each IVF and chemical family. The aim of the evidence map is to visualize the size and certainty of published evidence in a broad area and identify understudied areas with special research needs. Each bubble represents a single study with size

proportional to the population included and colour related to the respective risk of bias Tier classification according the NTP/OHAT approach (e.g. Tier 1 "unlikely"; Tier 2 "serious" and Tier 3 "very serious").

3.9 Confidence ratings and level of evidence

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The confidence rating has been conducted using the NTP/OHAT approach, one of the few available guidelines to conduct Systematic Reviews and Evidence Integration tailored for environmental health and based on the GRADE Working Group guidelines [23]. According the NTP/OHAT approach, the initial rating is established on the basis of study design features providing confidence that exposures are causally associated with the outcome. On this regard, the guidelines establish that for human observational studies the initial confidence must be considered at "moderate" due to the lack of experimental control of exposures. The rest of features are ensured by a proper prospective design including that "exposures precede the outcome", "individual outcome data" or presence of a "comparison group" [24]. Among the different factors considered to upgrade (i.e. magnitude, consistence, dose-response, residual confounding) or downgrade the confidence (i.e. risk of bias, unexplained inconsistency, indirectness, impression or publication bias), only risk of bias was identified to be compelling to justify a confidence downgrading of the overall body of body of literature based on "Serious" and "Very serious" classification of studies on PCBs, OCPs and PFAs (Section 3.8). Thus we judged to downgrade the confidence to "low". The step of translating the confidence into a level of evidence seeks to integrate the confidence with the nature of the effect ("health effect") or "no health effect"). On this regard, we conservatively judged the direction of the effect towards the presence of adverse health effects, despite the lack of consistency across studies. Thus, considering the NTP/OHAT framework, we established the level of evidence being "low".

Figure 2. Risk of bias heatmap summarizing the evaluations for each key question adapted from the NTP/OHAT tool. Instructions can be found in the Supplementary Materials Appendix 1. Further details on specific evaluations and justifications in HAWC.

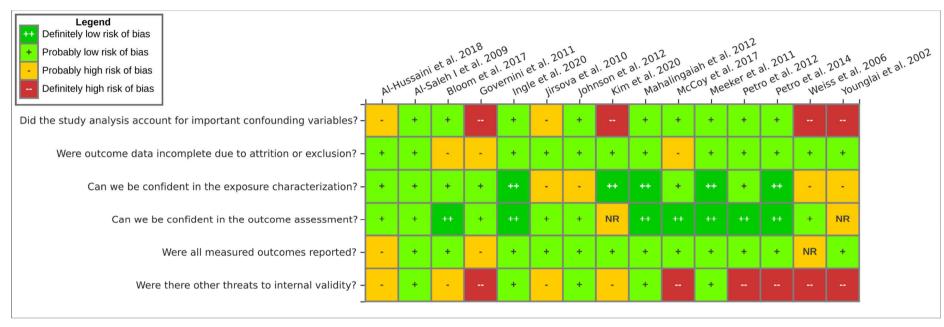
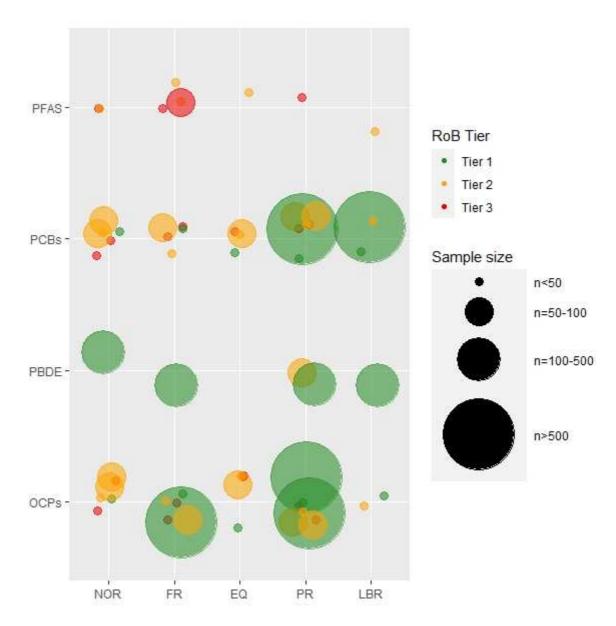


Figure 3. Evidence gap map summarizing the size and risk bias tier classification for included studies assessing the relationships between the main POP families and *In vitro* fertility outcomes. Each bubble represents a single study with size proportional to the population included in the analysis and colour related to the respective risk of bias Tier classification according the NTP/OHAT approach. Tier 1 refers to those studies classified as "unlikely", Tier 2 as "serious" and Tier 3 "very serious" risk of bias.



Abbreviations: EQ, embryo quality; FR, fertilization rate; LBR, live birth rate; NOR, number of oocytes retrieved; OCPs, organochlorine pesticides; PBDE, polybrominated diphenyl ethers; PCBs, polychlorinated biphenyls; PFAS, perfluoroalkyl substances; PR, pregnancy rate; RoB, risk of bias.

4. Discussion

To the best of our knowledge this is the first systematic review on the association between internal exposure levels of POPs in women and IVF outcomes. This review shows that most studies have been performed in small cohorts and focalized on PCBs and OCPs, whereas major research gaps remain for PFAs and PBDEs. Providing an accurate and quantitative synthesis of results was unfeasible due to the heterogeneity of metrics and exposure-outcomes dyads analyzed across the studies. Thus, we focused the synthesis on the identification of research gaps and main direction and magnitude of effects across POPs exposure-IVF outcomes pairs. In this respect and considering the broad topic, we introduced the evidence map as a novel tool to ease the visualization of a broad body of evidence and identification of main research needs. The overall results of studies included in this review were mainly inclusive, but there is evidence that some POPs could impair the chance of successful IVF treatments in infertile women, especially for the PCBs family which exhibited the most consistent deleterious effects. Most studies were classified in Tier 2 or 3, as "Serious" or "Very serious" risk of bias according the NTP-OHAT risk of bias tool due to lack of adjustment for confounding variables, low sample size and/or underreporting of outcome/exposure methods.

4.1. Study design and inter-individual variables

The global inconsistency of results across studies are likely to be due in part to temporal and geographic differences in POPs exposure and infertility treatment practices, as well as differences in congeners measured and study designs. A major methodological issue lies on the small sample size of most included studies (n<50 women). Thus, considering the modest effect size of POPs, it may be likely that many of studies may be underpowered to identify such associations [43]. Geographical differences of studies conducted in 9 different countries may be associated with a high variability on dietary habits and other lifestyles factors affecting IVF outcomes. Indeed, racial and ethnic disparities in exposure to endocrine disrupting chemicals [44] and in the chances of live birth after IVF have been reported [45,46]. Whereas the mean or median age of female was quite consistent across studies (32-35 years), the upper age bounds varied substantially (between 40 and 50 years old) which may

also substantially influence the results if not accounted for in the analysis. Indeed, female age has a major impact on the IVF outcomes [47].

4.2. Exposures

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The chemicals considered in the present review are mainly hydrophobic, but in a different extent. For instance, the octanol-water partition coefficient (log Kow), a commonly used measure the tendency of chemicals to bioconcentrate in the lipid compartments, vary across congeners from high (e.g. DDE ~ 6.5) to moderate or weak lipophilicity (e.g. PFOS ~4.5). In this respect, the lipid and protein content of biological matrices intended for chemical analysis may influence the analytical performance of lipophilic and amphiphilic chemicals, respectively, especially for low abundant congeners. Interestingly, many of included studies reported the paired determinations of biomarkers in FF and serum, mostly finding high correlations between both matrices for PCBs, OCPs and PFAs (\$\rho > 0.5\$) but weaker or even no correlation for PBDES (p<0.4). This could be explained by different physicochemical and metabolic properties of congeners. It should be noted that fasting status was not standardized for blood samples. Moreover, blood samples could have been collected at different time points of ovarian stimulation (i.e. prior, throughout stimulation or on the day of oocyte retrieval), and thus the circulating levels of lipophilic POPs (PCBs, OCPs and PBDEs), being affected by lipid mobilization seconding the hormonal status. For instance, endogenous estrogens may modulate the action of lipoprotein lipases and the profiles of free fatty acids and lipoproteins in blood [48] and thus modifying the release lipophilic compounds from lipid droplets [49]. On this regard, Mc Coy et al. reported that levels of PFAS, with both lipophilic and hydrophilic properties, were mostly stable in serum during the stimulation, except for PFOA [35]. The authors actually hypothesized that this change might reflect physiological and hormonal changes, consequently impacting the clearance of PFAs. Whereas we acknowledge that FF may be a convenient and representative matrix of the target site (i.e. ovarian environment), we may anticipate that its lower lipid concentration as compared to serum (e.g. ~ 10 folds) and lipid dynamics during ovarian stimulation might contribute to measurement error and variability [50]. In this regard, the lipid normalization of POPs biomarkers has been widely embraced in biomonitoring. Nonetheless, there is still an active debate about the best modeling practices for POPs biomarkers in epidemiology, especially when circulating lipids may be altered by the exposures or the outcomes [46]. Most included studies used the default approach, and few studies compared results between lipid-normalized and raw biomarkers [32,39,41]. The partition of lipophilic POPs and lipids between blood and FF has not been characterized, precluding related conclusions about the impact of lipid-normalization on statistical models.

Another major issue associated to observational research on POPs hampering the comparison across studies is the distribution of chemicals within study population, including the variable average concentration levels, ranges of distributions (e.g. high vs low exposures) or the baseline concentrations in the lower bound of distributions that would determine the exposure contrast. For instance, the mean level and standard deviations of PCB180 in FF ranged between 0.04±0.02 ng/mL [34] and 0.1±0.02 ng/mL [36].

4.3. IVF outcomes

Numerous clinical factors could also contribute to the divergent associations reported between IVF outcomes and POPs. First, populations were heterogeneous and not always described in details. For instance, the infertility etiologies and duration of infertility may impact the consistency of the findings. Although the impact of the type of infertility (primary versus secondary infertility) on IVF outcomes is controversial [51], some etiologies such as adenomyosis, ovulatory dysfunction [52,53] or some individual factors such as age, obesity or long duration of infertility [47,51,54] have been shown to have a clear detrimental impact on IVF outcomes. Furthermore, some studies reported increased levels of POPs in certain comorbidities, such as endometriosis [12], polycystic ovarian syndrome (PCOS) [13] or other ovulatory disorders [20] that could result on infertility through different signalling pathways. From a mechanistic point of view, some studies suggested that POPs acted on reproductive outcomes via disruption of modification of endocrine pathways. Indeed, antisteroidogenic or steroidogenic properties of PCBs, as well as interaction with progesterone receptor pathway have been well documented [55]. The antiestrogenic effect of PBDE, impacting the

uterine expression of estrogen and progesterone receptor [30], and the ability of PFAs to bind to peroxisome proliferator-activated receptors have also been described [33]. POPs like dioxins can have pro-inflammatory effects on the endometrium [56], potentially impairing uterine receptivity [57]. Concerning progesterone pathway, assisted reproduction generally implies the administration of exogenous progesterone. However, this might not overcome the potential deleterious role of a progesterone disruptor, albeit it might be probably less harmful than in spontaneous cycles. In any case, the evaluation of putative associations between exposure (i.e. endocrine disruptors) and pathophysiological effects should be cautious in the context of a clinical intervention which can bias the analysis. Furthermore, animal studies reported an impact of estrogen levels on the clearance of POPs, such as PFOA [35], that may prone to reverse causation in cross-sectional study designs or modifying the estimates in FF following the stimulation. Second, most studies did not provide information on male partner / factors, although sperm parameters play a critical role in the chances of birth following ART [58]. In the same vein, protocols for controlled ovarian hyperstimulation are rarely described in the available studies, even though this parameter does not seem to be significantly associated with IVF outcomes [59]. Third, the main heterogeneity across literature lies within the various endpoints used to report IVF outcomes. This is illustrated by the dissimilar assessment of more advanced events, for instance by the use of "failed implantation rate" in some studies [29,39,41], while others used more conventional "pregnancy rates". We considered both as similar events with inverse metrics in this study in order to ease the synthesis. Additionally, whereas early and intermediate IVF outcomes (prior to implantation) provide interesting information on oocyte and embryo quality, most studies failed to assess LBR which is considered the most relevant outcome for infertile couples and medical staff [60]. Indeed, statistical differences found in terms of embryo quality and/or early pregnancy events might be of little or even null clinical relevance if no impact on LBR is found at the end. Thus, future studies should consider including LBR as main outcome in order to provide an overall evaluation of the reproductive process and help improve relevant information and counselling for patients and care providers.

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It should be noted that findings of this review on the impact of POPs on female reproductive function might not be easily generalizable to fertile women and/or to infertile women not undergoing IVF. Indeed, although one study reported an association between exposure to POPs and increased time to achieve pregnancy [61], further analysis in properly designed prospective cohort studies is required to assess clinical relevance. More specifically, the heterogeneity in IVF population and practices across regions calls for caution when drawing conclusions from such a review and advocates for external validation studies.

4.4. Statistical methods

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We have noticed a variety of criteria regarding the inclusion of chemicals in statistical analysis based on detection rates and methods to manage non-detected samples. For instance, some studies included biomarkers detected in at least one sample [34] whereas other studies included biomarkers detected in all samples [39]. Low detection rates may introduce exposure misclassification and generate artefacts in regression models, that could help to explain the counterintuitive positive association between PCB 146 (detection frequency of 6.2%) and implantation rate (RR = 4.09; 95% CI 2.09, 8.02) reported by Bloom et al. [34]. Most studies used standard univariate or multivariate regression approaches (either linear, negative binomial, logistic binomial or Poisson models) to study the associations between POPs and IVF outcomes, either considering continuous, counts or binary outcomes. We have noticed that a substantial number of studies have not considered relevant confounding variables in the statistical analysis. For instance, female BMI and age are well-established factors that can affect both the exposure levels of POPs and the IVF outcomes [47,54]. Conversely some few studies considered a comprehensive list of additional confounders, including study phase, ethnicity, previous live birth, smoking, gonadotropins or estradiol levels [30,32,39]. In turn, the lack of proper consideration of negative or positive confounding, interactions or effect modification due to diet or co-occurrence of protective micronutrients (e.g. vitamin D or ω-3 polyunsaturated fatty acids) [62,63], may result in attenuation bias or nullifying the associations [64]. Other common statistical pitfalls in IVF studies have been highlighted elsewhere [60]. For instance, little attention has been generally paid to adjustment for women cycles in order to account for 'unit of analysis' errors or the misuse and misinterpretation of odds ratios with binary outcomes [60]. Another source of bias that has not commonly reported in the selected studies involve the exclusion of cycles where embryos were not viable in the analysis of implantation. As Messerlian and Gaskins argued [60], the exclusion of these cycles (about 10-20% of cycles before embryo transfer) will likely lead to bias if the exposure of interest is associated with early ART failure (e.g. no oocytes fertilized or poor embryo). On this regard, some studies have applied generalized estimating equations to accommodate clustering of oocytes and embryo outcomes within women [34,38] or woman-specific random-effect terms to account for the correlation among outcomes from different cycles [39, 40]

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Most studies considered the biomarkers as independent exposure variables, however the multidetection methods have shown that most POPs are embedded in complex mixtures of highly correlated chemicals with similar physicochemical properties. Statistical analysis of highly correlated mixtures were not conducted in most included studies, however in some cases the sum of congeners was considered to elaborate a combined variable. Additionally, 2 studies within the same population considered a two-step approach to, first, identify the Principal Component-based variables representing the different chemical mixtures, and second, to estimate the associations of these composed variables in the standard regression models [32,32]. Whereas this principal component analysis approach is a step-forward beyond the analysis of single pollutants, caution should be taken in terms of interpretation of resulting associations because the factor loadings are estimated without taking into account the underlying association with the outcome [65]. The methodological development and computational apparel for multipollutant modeling is quickly evolving to answer multiple questions related to chemical mixtures in epidemiological research, so future studies should consider some of the available models to gain insight on the mixture effects [66,67]. Last but not least, it is noteworthy to consider the potential presence of non-monotonic dose-responses in endocrine-related health outcomes, thus the application of statistical approaches able to identify nonlinear effects should be considered as a complementary approach to standard methods [68]. Among included studies few studies have considered for instance the chemical exposures as quartiles, in addition of continuous variables, that may be helpful to elucidate non-monotonic trends [39,40].

5. Conclusion

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To the best of our knowledge this is the first attempt to systematically gather and evaluate the evidence about the associations between POPs and IVF outcomes to deliver formal conclusions. Applying a robust evidence-based approach, the present systematic review shows that all main families of POPs are still pervasive in follicular fluid and serum of women attending modern IVF treatments. Complex mixtures of legacy POPs are likely to occur together with more emerging chemicals such as PFAs or PBDEs. The formal synthesis of effects across studies was not feasible due to the heterogeneous designs and metrics favouring a quantitative analysis. Globally, we found inconsistent findings across studies for specific exposure-outcome dyads, suggesting that adverse effects of POPs on IVF outcomes cannot be ruled out. Specifically, our results suggest that POPs, notably some PCBs and OCPs, may impair embryo quality and pregnancy rates, even though the biological mechanisms remain to be deciphered. Most published studies have 'serious' or 'very serious' risk of bias according the NTP/OHAT framework, mainly due to lack of adjustment for confounding variables or small sample size, that may help to explain the divergent results. Globally, we judged the level of evidence being "low". We emphasized some gaps and directions for future studies including: 1) the use of individual exposure biomarkers with a large proportion of detected observations for statistical analysis; 2) blood and FF may provide complementary information on POP exposure, but caution should be paid due to the little knowledge available about the respective partition coefficients, the impacts of hormonal stimulation or the biomarker lipid normalization; 3) LBR should be considered the main outcome recapitulating the entire fertilization process, and thus more biologically relevant than other intermediary outcomes whose statistical findings should be considered with caution; 4) statistical models should accommodate the different women stimulation cycles and non-transferred embryos in the analysis, as well as relevant confounding variables

- 559 including age and BMI. Given the high economical and societal costs associated with infertility and
- 560 IVF, further well-designed research is urged to fill the highlighted gaps.

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Conflict of interest

The authors disclose no competing conflict of interest.

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