

Rare earth element organotropism in European eel (Anguilla anguilla)

Marjorie Lortholarie, Laurence Poirier, Abderrahmane Kamari, Christine Herrenknecht, Aurore Zalouk-Vergnoux

► To cite this version:

Marjorie Lortholarie, Laurence Poirier, Abderrahmane Kamari, Christine Herrenknecht, Aurore Zalouk-Vergnoux. Rare earth element organotropism in European eel (Anguilla anguilla). Science of the Total Environment, 2021, 766, pp.142513. 10.1016/j.scitotenv.2020.142513 . hal-04654530

HAL Id: hal-04654530 https://hal.science/hal-04654530

Submitted on 22 Jul 2024

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers. L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.



Distributed under a Creative Commons Attribution - NonCommercial 4.0 International License

Version of Record: https://www.sciencedirect.com/science/article/pii/S0048969720360423 Manuscript_006e69539d745d646f81868d820eb526

1	Rare earth element organotropism in European eel (Anguilla anguilla)
2	
3	Marjorie Lortholarie ¹ , Laurence Poirier ¹ , Abderrahmane Kamari ¹ , Christine Herrenknecht ¹ , Aurore
4	Zalouk-Vergnoux ¹
5	
6	¹ Laboratory Mer, Molécules, Santé (MMS, EA 2160), University of Nantes, Nantes F-44322,
7	France
8	
9 10	Email: m.iortholarie@gmail.com
10	
11 12	Abstract
12	Rare earth elements (REEs) are metallic elements with electronic magnetic optical and
15	Kare earli clements (KLLs) are metanic clements with electronic, magnetic, optical and
14	catalytic properties which make them essential in many industrial and medical fields. REEs are
15	therefore becoming emerging pollutants and it is important to understand their implications for
16	ecosystem health. However, little knowledge of REE bioaccumulation in aquatic organisms is
17	available and especially on their internal distribution in fish. In the present study, REE
18	organotropism was determined in Anguilla anguilla from the Loire estuary (France) by determining

19 burdens in a wide set of tissues, organs and biological fluids. Differences have been observed 20 between life stages and genders. For yellow eels, the most accumulating organ was the gills 21 $(126.90 \pm 50.78 \,\mu\text{g/kg dw})$ and for silver eels, it was the liver $(181.78 \pm 62.04 \,\mu\text{g/kg dw})$ for males; 22 $203.79 \pm 111.86 \,\mu$ g/kg dw for females). The comparison between female silver and yellow eels 23 shown that female silver individuals accumulated significantly more REEs in the urinary system 24 (US), muscles, gonads, spleen and liver, while yellow individuals accumulated more in gills. The 25 comparison between male and female silver eels also highlighted differences, indeed the females 26 accumulated significantly more REEs in the US, gonads, skin and spleen, compared to males which 27 accumulated significantly more in muscles and gills. REEs abundances are also different between organs, life stages and genders. The gonads of female silver eels exhibited a particular profile with 28 29 the dominance of gadolinium (Gd) (up to 74.2% of $\Sigma REEs$). Moreover, the presence of Anguillicola 30 crassus in the swim bladder of organisms seemed to have an impact on REE bioaccumulation: 31 parasitized yellow eels present higher concentrations of REEs in muscles, gills, gonads and liver than

32	non-parasitized individuals. Regarding glass eels, REE contribution profiles in the whole body were
33	close to those of yellow and silver eel skin.

34 Keywords: Fish organs, Rare earth elements, Life stage, Gender, European eel

35

36 Highlight:

37 38

- Organotropism determined in muscles, skin, gills, spleen, US, liver and gonads
- REE accumulation and organotropism were influenced by gender and life stages
- Yellow eels accumulated more in gills, female and male silver eels in livers
- Parasitized yellow eels accumulated more REEs than non-parasitized
- 43
- 44 45

1. Introduction

46

47 Anguilla anguilla is an euryhaline and catadromous species with a really special life cycle. After a transatlantic migration of nearly 6,000 km from east to west traveled from 4 to 6 months, 48 49 silver eels arrive in the Sargasso Sea for the reproduction where they die after. Laid eggs turn into 50 leptocephalus larvae which migrate using the Gulf Stream to European coasts. Arriving at the level 51 of the continental slope about 100 km from the coast, leptocephalus larvae become glass eels. In 52 freshwater, they begin to feed and after some time, glass eels gradually pigment in black then yellow 53 to become yellow eel. Yellow eels live in freshwater. They are benthic and sedentary for several 54 years waiting for the silvering. This metamorphosis changes yellow eels into silver eels upon sexual 55 maturity (Adam et al. 2008; Corolla et al. 2016). This transition phase leads to many morphological 56 and physiological changes (Adam et al. 2008; Fontaine 1994; Durif et al.2015). During the 57 downstream migrating period, the silver eels are sexually differentiated but they are not yet 58 completely mature.

59 Long considered to be a harmful species as a result of its snake-like appearance and the untrue idea that it is the cause of the noble fish species decline (salmon, trout or char), the European 60 61 eel is now recognized as a species with an economic, ecological and scientific interest. However, the 62 populations are declining since the 1980s. This decrease is not a single factor consequence but a 63 combination of overfishing at all life stages, habitat reduction, water quality degradation, obstacle 64 during the migration, diseases such as the Anguillicosis causes by the parasites Anguillicola crassus. Consequently, European eel was listed in CITES (Convention International Trade Endangered 65 66 Species) Appendix II in 2009 and ranked in the red list of threatened species as a critically 67 endangered species in 2014 by IUCN (International Union for Conservation of Nature).

68 Among the threats to eels, the aquatic ecosystems contamination is very worrying. In fact, the European eel are able to accumulate a wide variety of environmental contaminants and some studies 69 70 have highlighted the accumulation of polycyclic aromatic hydrocarbons (Ribeiro et al. 2005), 71 pesticides and persistent organic pollutants (Couderc et al., 2015; van der Oost et al., 1996). They are 72 also able to accumulate essential metals (Ni, Zn, Cu, Mn, Co and Fe), non-essential metals and 73 metalloids (Hg, Al, Cr, Pb, Cd and As) (Barak & Mason, 1990; Bordajandi et al., 2003; Durrieu et 74 al., 2005; Mansouri et al., 2018; Neto et al., 2011; Ribeiro et al., 2005; Storelli et al., 2007; Tabouret 75 et al., 2011). The metal bioaccumulation can cause sublethal biological effects. For example, yellow 76 eels exposed during one month to Cd (5 µg/L) presented an increase of fat consumption in their reserve necessary for migration and could compromise successful reproduction (Pierron et al., 2007). 77 78 Another study at the same concentration of Cd showed structural or functional perturbations in the 79 gills of glass eels after 2 h exposure (Gony, 1990). A one-month exposure to 300 µg/L of Pb, 80 increased the number of lymphocytes and the plasma lactate levels in yellow eels (Santos & Hall, 81 1990). After a 7 day exposure of eels to Cu 0.2 µmol/L, significant decreases of free triiodothyronine 82 and cortisol as well as a significant increase of glucose were observed in the plasma. In addition, a 83 lactate decrease and an increase of nuclear abnormality frequency were measured in blood, gills, 84 liver and urinary system (Oliveira et al., 2008). The ecotoxicology of essential and non-essential 85 metals is now relatively well documented for eels, but other types of metallic elements, such as
86 REEs, are becoming emerging contaminants and should be better considered.

87 REEs are a set of 17 metallic elements including scandium, yttrium and 15 lanthanides from 88 lanthanum to lutetium. They have close chemical and physical properties due to their main identical 89 charge +III (except for some REEs may have two oxidation numbers, +III and +IV for Ce, Pr, Tb or 90 +II and +III for Eu, Sm and Yb) and their neighboring ionic radius. REEs are usually classified 91 according to their increasing atomic number into three groups: light (from lanthanum to 92 promethium), medium (from samarium to gadolinium) and heavy (from terbium to lutetium, 93 including yttrium). Due to its physico-chemical properties and its presence in the same deposits, 94 yttrium should be among the light REEs but it is classified with the heavy REEs due to the similarity 95 of its property and its deposits. Sc is not classified in any of these REE groups because of its absence 96 in the same deposits (Bru et al., 2015).

97 REEs possess highly electronic, magnetic, optical and catalytic properties, leading to a 98 prevalent use in several industrial fields such as the production of a super-magnet and luminophore, 99 petroleum catalysis, pigmentation of glasses and plastics, metallurgical alloys, nuclear and medical 100 imaging (Bru et al., 2015). In 2016, their global production was 126 Kt of rare earth oxides (Zhou et 101 al., 2017). This production is set to increase for several years with the development of clean energy 102 such as wind turbines, electric vehicles, and energy-efficient lighting (Alonso et al. 2012; Zhou et al. 103 2017).

In aquatic environments, REEs are ubiquitous contaminants, their presence has already been measured in waters, suspended particles and sediments in different environmental compartments in the worldwide such as rivers (Keasler & Loveland, 1982; Négrel et al., 2000), estuaries (Elbaz-Poulichet and Dupuy, 1999; Vrel, 2012) and oceans (Akagi & Edanami, 2017; Dahlqvist et al., 2005; Hirata et al., 2002; Johannesson et al., 2017; Li et al., 2015). Off the Loire estuary, the area concerned by the present study, sediments of the Bay of Biscay showed mean concentrations ranging from 0.11 (Lu) to 47.79 (Ce) mg/kg (Chaillou et al., 2006). With the proximity of the Loire estuary,

111 the mean concentrations in sediments reached 41 and 92 mg/kg for La and Ce respectively. In the 112 suspended particulate matter of the Loire estuary, mean values of total REEs (Σ REEs) reached 113 137 mg/kg upstream of Nantes, 164 to 489 mg/kg downstream of Nantes and 89 mg/kg near the 114 coast (De Chanvalon et al., 2016). Levels between 195 and 199 mg/kg were also measured in 115 sediments downstream of Nantes. In the Garonne River (La Réole), Lerat-Hardy et al. (2019) 116 measured concentrations in waters from 0.469 ng/L (Tm) to 37 ng/L (Ce) in 2017. Environmental 117 concentrations normalized to reference reservoirs makes possible to highlight anomalies. When these 118 anomalies are positive, that could point out an anthropic enrichment. The Sm and Gd often appeared 119 as positive anomalies, as shown in the Garonne estuary (France) by Lerat-Hardy et al. (2019). These 120 aquatic environmental contaminations are worrying because REEs were shown to be able to accumulate in aquatic organisms. In the European eel, REE accumulation has already been 121 122 highlighted in different life stages, from glass eels (Figueiredo et al., 2018) to yellow and silver eels 123 (Lortholarie et al., 2019).

124 The aim of this work was to explore REE internalization and distribution in a key species of 125 the estuarine ecosystem. The European eel represent an interesting model organism due to its benthic 126 and partial sedentary life, and several metamorphosis processes that can influence metal 127 organotropism. The plan of the present study was i) to investigate the internal distribution of REEs in 128 wild European eels, by determining REE concentration in muscles, skin, gills, blood, spleen, urinary 129 system, liver and gonads, ii) to study life stage and gender influences by sampling glass, yellow and 130 silver eels, iii) to explore the presence of parasites A. crassus in the swim bladder and their potential 131 impact on bioaccumulation and organotropism.

- 132
- *133* 2. Materials and Methods
- 134
- 135

2.1.

- Chemicals
- 136

137 Stock solutions of 15 REEs (yttrium: Y; lanthanum: La; cerium: Ce; praseodymium: Pr, 138 neodymium: Nd; samarium: Sm; europium: Eu; gadolinium: Gd; terbium: Tb, dysprosium: Dy; 139 holmium: Ho; erbium: Er, thulium: Tm; ytterbium: Yb and lutetium: Lu) and Rhenium (internal 140 standard) at 1000 mg/L in 2% HNO₃, were provided by Roth (single-element ICP standard solution). 141 REEs were oxide forms in solution $(Y_2O_3, La_2O_3, Ce (NO_3)_3, Pr_6O_{11}, Nd_2O_3, Sm_2O_3, Eu_2O_3, Gd_2O_3, Ce (NO_3)_3, Pr_6O_{11}, Nd_2O_3, Sm_2O_3, Eu_2O_3, Cd_2O_3, Ce (NO_3)_3, Pr_6O_{11}, Nd_2O_3, Sm_2O_3, Cd_2O_3, Cd_2O_3, Ce (NO_3)_3, Pr_6O_{11}, Nd_2O_3, Sm_2O_3, Ce (NO_3)_3, Pr_6O_{11}, Nd_2O_3, Ce (NO_3)_3, Pr_6O_{11}, Nd_2O_3, Ce (NO_3)_3, Pr_6O_{11}, Nd_2O_3, Ce (NO_3)_3, Pr_6O_{11}, Nd_2O_3, Ce (NO_3)_3, Pr_6O_{11}, Nd_2O_{12}, Ce (NO_3)_3, Pr_6O_{11}, Nd_2O_{12}, Ce (NO_3)_3, Pr_6O_{11}, Nd_2O_{12}, Ce (NO_3)_3, Pr_6O_{11}, Pr_6O_{12}, Pr_6O$ 142 Tb_4O_7 , Dy_2O_3 , Ho_2O_3 , Er_2O_3 , Tm_2O_3 , Yb_2O_3 , Lu_2O_3). In this work, 15 out of the 17 REEs were 143 studied. Pm was not included because it is not naturally present in the environment and neither Sc 144 because of many interferences during the ICP-MS analysis, which led to overestimation of 145 concentrations. Mineralization and solutions were made with HNO₃ 65% from Fisher Scientific. 146 Mineralization protocol was validated using the international certified material BCR-668 consisted 147 of muscle of mussels from LGC standards.

148 149

2.2. Sampling

150

151 Anguilla anguilla were fished in the Loire estuary (Figure 1) at three life stages: glass, yellow and silver. Glass eels were fished in December 2019 at Cordemais (47°16'58.7''N -1°53'55.1''O) 152 153 located 30 km downstream of Nantes. Yellow eels were caught in June 2019 at Haute-Indre (47°11'39''N – 1°40'4''O) located 12 km downstream of Nantes. Silver eels were caught during the 154 155 downstream period in December 2018 at Varades (47°21'58''N - 1°1'31''O). About silver eels, 156 since they were downstream migrating, they cannot be considered as from this specific site. Eels were captured by professional fishermen by means of fish traps. For each study site and sexual 157 158 maturity, 10 to 15 individuals were collected. After fish capture, eels were conducted to the 159 laboratory in 200 L tanks with water from the sampling site and aeration. At the laboratory, fish were 160 directly euthanized by lethal dose of eugenol just before their dissection.

- 161
- 162 163

2.3. Preparation of Fish Tissue and Parasites for REE Analyses

6

164 After euthanasia, eels were weighed (g) and measured (cm) to get respectively the body 165 weight (BW) and the body length (BL). During the dissection, the observation of the gonads allowed 166 to visually identify the gender of individuals. Interest organs for the organotropism study were 167 removed and weighed: muscles, skin, gills, spleen, urinary system, liver, and gonads. Two different 168 biological fluids were also collected: blood and bile. When present, A. crassus parasites were also 169 collected from the swim bladder of eels. A. crassus is a nematode feeding on blood and causing a 170 disease called Anguillicosis (Adam et al., 2008). The parasites were counted, weighed and analyzed 171 as fish tissue. The parasite index (PI) was determined using PI = (weight Anguillicola crassus/weight 172 eel body) *100. The otoliths were recovered to determine the age of each individual according to 173 ICES 2009 (International Council for the Exploration of the Sea) methodology. Glass eels were 174 observed with a binocular and trilled into three groups VI A0, V A1 and VI A2, depending on their 175 pigmentation stage (Grellier et al., 1991). The gonads were recovered and weighed to determine the 176 gonadosomatic index (GSI). The GSI was calculated as follows: GSI = gonad weight (g) / whole 177 body weight (g) * 100. The Fulton condition index (K) was determined as following: K = (whole body weight $(g)^{*}10^{5}$ / whole body length³ (cm), K allows evaluating the overall global eel health in 178 179 front of environmental stresses such as pollution or feeding conditions.

180 At the beginning of the experiments, all the material was cleaned by immersion in HNO₃ 10% 181 for one night and rinsed with Milli-Q water before use. The low REE concentrations expected in the 182 muscles required a preconcentration step by calcination allowed due to sufficient muscle masses. 183 About 1 g of muscles were calcined in a muffle furnace (Nobertherm) at 300°C during 1h followed 184 by 500°C during 2h. The muscle samples were therefore concentrated about 10 times. For the 185 mineralization, 250 mg of ashes of muscles, of the other organs or of A. crassus were collected and 186 weighted. Then, the mineralization relied on a predigestion with a mix of HNO₃/HCl (2:1, v/v) for 187 one night at room temperature and a following step using microwaves (Anton Paar, MW GO 50Hz) 188 (Lortholarie et al., 2019). The solutions were then transferred into Pyrex tubes and evaporated to 189 dryness in hot-plate at 130°C. Samples were resuspended with 1.8 mL of HNO₃ 2% and then *vortexed, sonicated (20 min) and centrifuged (5 min, 2.0 rpm, 15°C) to eliminate particles in suspension. Only the supernatants were analyzed.*

192 193

194

2.4. Rare Earth Element Analyses

195 Concentrations of REEs were determined by Inductively Coupled Plasma coupled to Mass Spectrometry (ICP-MS, Nexion 350x PerkinElmer). All samples, blanks and standards were diluted 196 using HNO₃ 2%. Rhenium was used as internal standard (¹⁸⁷Re isotope), with a final concentration of 197 198 1 µg/L in the solutions. Matrix effect determination was performed using a matrix matched 199 calibration. Sample supernatant solutions were used as the matrix to prepare calibration solutions 200 with increasing concentrations of REEs. REE isotopes were selected according to a compromise 201 between the maximum of natural abundance and the minimum of interferences. They corresponded to: ⁸⁹Y, ¹³⁹La, ¹⁴⁰Ce, ¹⁴¹Pr, ¹⁴⁶Nd, ¹⁴⁷Sm, ¹⁵³Eu, ¹⁵⁷Gd, ¹⁵⁹Tb, ¹⁶³Dy, ¹⁶⁵Ho, ¹⁶⁶Er, ¹⁶⁹Tm, ¹⁷¹Yb and 202 203 ¹⁷⁵Lu. Limits of detection (LOD) and quantification (LOQ) were determined for each REE. 204 Analytical and procedural blanks were prepared and analyzed using the same analytical procedure as 205 for samples. Finally, the blank values were subtracted from the values of the samples. The accuracy 206 of the whole analytical protocol was validated using an international certified material containing 207 REEs (BCR-668, muscle of mussels).

- 208
- 209

211

210 2.5. Data Processing

In this work the bioaccumulation results were given in micrograms per kilogram of dry weight tissue (μ g/kg dw). The results were presented using 3 groups of REEs: light (LREEs: La to Nd; Pm was not studied in this work because it is a radioactive element which is not naturally present in the environment)(*Bru et al.*, 2015), medium (MREEs: Sm to Gd) and heavy (HREEs: Tb to Lu and Y).

217 The potential life stage influence on bioaccumulation was studied by the comparison of 218 concentrations measured in different organs of the yellow eels and the female silver eels which 219 represented sexually undifferentiated and mature sexual stages respectively. Even if yellow eels were 220 not sexually differentiated, their size was already greater than the maximum size expected for silver 221 males (45 cm). Indeed in this species, there is a sexual dimorphism with females larger than males 222 (Adam et al., 2008). It was thus possible to suppose that yellow eels, sampled for this study, would 223 become females, allowing their data comparison with those of female silver eels to study the life 224 stage influence.

About parasitism, the aim was to determine if influences exist with REE concentrations in all organs of eels according to life stage or gender. Moreover, the REE contents in parasites were determined to find a potential link between REE concentrations in fish tissue and in *A. crassus*.

228 The data were statistically treated using the software Statistica software (v7, StatSoft, Inc.). 229 The results of Kolmogorov-Smirnov test showed the need to use non-parametric tests to work on the 230 data. The correlation of biological parameters such as: body length, body weight, gonadosomatic 231 index, Fulton index, parasite index and age with REE concentrations in each organ was investigated 232 using a correlation test with a casewise deletion. The casewise deletion of missing data while 233 calculating a correlation matrix is to exclude all cases that have missing data in at least one of the 234 selected variables. The correlations were significant when p-value<0.05. The Kruskal-Wallis test was 235 used for studying: (i) the influence of both sexual maturity and the gender of adult eels, by 236 comparing respectively organ REE concentrations of yellow eels to female silver eels and male silver 237 to female silver eels; (ii) the bioaccumulation differences between organs within a life stage or a 238 gender; (iii) the bioaccumulation differences between the three different stages of pigmentation 239 (VI A0, VI A1 and VI A2) of glass eels. The REE bioaccumulation differences between parasitized 240 individuals and non-parasitized ones were studied with a Mann-Whitney U test. The results of both 241 Mann-Whitney and Kruskal-Wallis tests were considered as significant when p-value was lower than 242 0.05.

244 3. Results

245

247

246 3.1. Biometric Parameters

The correlation study between biological parameters (BL, BW, GSI, K, PI, age) was performed on 15 yellow eels, 14 female silver eels and 15 male silver eels.

In yellow eels, BL and BW were positively correlated. The linear regression equation was BW=22.192*BL-896.2 (R^2 =0.947). Moreover, BL and BW were also significantly correlated with age. The linear regressions were respectively: age=0.226*BL-5.115 (R^2 =0.565) and age=0.015*BW+3.889 (R^2 =0.697).

In female silver eels, BL and BW were also positively correlated: BW=15.256*BL-543.5 (R²=0.978). However, BW and GSI were negatively correlated, with a linear regression of GSI=-0.002*BL+0.282 (R²=0.848).

In male silver eels, BL and BW were positively correlated: BW=6.078*BL-139.1 (R²=0.749). Moreover, for male silver eels, BL and GSI were also positively correlated with a linear regression GSI=0.112*BL-3.701 (R²=0.749). In male silver eels as in females, no correlation of BL or BW with age was observed.

261 The BL and BW were significantly different (p-value<0.01) between female and male silver 262 eels. Indeed, the males were smaller and lighter than the females, with respectively a median of 263 38.2±2.73 cm for BL (vs 61.75±8.99 cm) and of 92.0±20.13 g for BW (vs 387.5±179.03 g). These 264 differences were not observed between yellow eels and female silver eels. There was no significant 265 difference of age between yellow eels and female silver eels or between male and female silver eels 266 (table 3 supplementary material). The same trend was observed for the PI, despite some differences 267 observed in the rate of parasitized individuals: 66.6% of the yellow eels, 53.8% of the female and 268 86.6% of the males.

270 3.2. REE Levels

271

- 272 3.2.1. In Yellow Eels
- 273

All the yellow eels caught bioaccumulated REEs and their presence were observed and quantified in the urinary system, muscles, gills, gonads, skin, spleen, liver and bile (Figure 2A). A similar pattern among all the organs was observed, with higher concentration of LREEs followed by HREEs and then MREEs. The \sum REE median concentrations (µg/kg dw) were 46.61±16.56 for urinary system, 2.93±1.02 for muscles, 126.90±50.78 for gill, 5.42±3.96 for gonads, 26.92±6.36 for skin, 11.21±5.53 for spleen, 99.32±61.16 for liver and 4.58±4.23 for bile. All REE concentrations are presented in Table 1 in the supplementary material.

281

282 3.2.2. According to the Life Stage

283

284 Figure 2B shows the Σ REE, LREE, MREE and HREE concentrations in the different organs 285 of female silver eels. Σ REEs were significantly higher in the urinary system, muscles, spleen, gonads 286 and liver of female silver eels compared to those of yellow eels. These significant differences are 287 displayed with # between parts A and B of the Figure 2. Σ REE enrichment factors reached 2.1, 2.3, 288 3.2, 7.0 and 9.6 in liver, muscles, urinary system, gonads and spleen respectively. The significant 289 higher \sum REE concentrations in female silver eels were noticed among all the different groups, *i.e.* 290 LREEs, MREEs and HREEs. Only \sum REE concentrations in gills of yellow eels were significantly 291 higher than those of female silver eels with a factor of 2.2, due to increases of LREE and MREE 292 concentrations.

293

- *294 3.2.3. According to the Gender*
- 295

296 Gender influence between male and female silver eels was highlighted by some differences in 297 REE concentrations. Figures 2B and 2C show the Σ REE, LREE, MREE and HREE concentrations 298 in the different organs of female and male silver eels respectively. The significant differences are 299 displayed with * between parts B and C of the Figure 2. Females accumulated significantly more 300 Σ REEs in their urinary system and skin than males, due to increases of LREE, MREE and HREE 301 concentrations. Furthermore, females accumulated significantly more $\Sigma REEs$ in their gonads and 302 spleen than males because of significantly higher MREE levels. In female tissues, ΣREE levels were 303 more than 1.7, 2.0, 3.1 and 6.7 times higher in spleen, urinary system, skin and gonads, respectively. 304 On the contrary, male silver eels accumulated significantly more MREEs in gills as well as ΣREE in 305 muscles due to increases of LREEs, MREEs, and HREEs.

- 306
- 307 308

3.3. REE Organotropism

309

3.3.1. Yellow Eels

310

Table 1 presents the classification of yellow eel organs, from the more to the less concentrated in \sum REEs, LREEs, MREEs and HREEs. For these organisms, the organotropisms were similar considering \sum REEs or the categorized groups, LREEs, MREEs and HREEs. It existed only 2 inversions between urinary system and liver as well as between gonads and bile for HREEs. The liver and gills accumulated significantly more than the spleen, gonads, bile and muscles. \sum REEs were less accumulated in gonads, bile and muscles.

317

318 3.3.2. Female Silver Eels

For female silver eels (Table 2), the organotropisms were very different considering \sum REEs, LREEs, MREEs and HREEs. The \sum REE highest concentrations were found in liver, then in the urinary system. The liver accumulated significantly more \sum REEs than the gills, gonads, skins, muscles and bile. The urinary system accumulated significantly more \sum REEs than the skin, muscles and bile. For the other REE groups, muscles and bile were systematically the less concentrated. Liver, urinary system and spleen (purple and blue colored) were the three more concentrated organs 325 but not in the same order depending on the group of REEs considered. Gills, skin and gonads (green 326 and yellow colored) were organs intermediately concentrated. Particular attention should be paid to 327 the gonads, and their highest particular MREE concentrations leading to the gonads as the most 328 concentrated organ for this REE category.

329 330

3.3.3. Male Silver Eels

331 For male silver eels (Table 3), as for female silver eels, the classification of the organs 332 depended on the considered REEs, *i.e.* the accumulation of Σ REEs, LREEs, MREEs and HREEs 333 was different. Considering Σ REEs, the liver, then the gills and urinary system were the strongest 334 accumulator organs. The liver accumulated significantly more $\sum REEs$ than muscles, skin, gonads 335 and bile with respective enrichment factors reaching 15, 29, 33 and 192. The gills accumulated 336 significantly more Σ REEs than muscle, skin, gonads and bile with respective factors of 7, 15, 16 and 337 95. The lowest Σ REE concentrations measured in males were in the gonads and bile. Bile was the 338 less concentrated for all the groups of REEs. Then, the order of the organs depending on their REE 339 concentrations were less homogenous according to the different REE categories. For example, the muscles were in the 5th position for Σ REEs and LREEs while they were in the 7th position for 340 341 MREEs and HREEs. In addition, for HREEs, the gills were in the first position whereas it was the 342 liver for the other groups of REEs.

343

344

3.3.4. Life Stage and Gender Influences

The \sum REE organotropism comparison between yellow eels, female and male silver eels 345 346 revealed some differences (Tables 1, 2 and 3). The order of the organs according to the different 347 categories of REEs was homogenous for yellow eels, which was not the case for the silver 348 individuals whatever the gender. Comparing the two genders of the silver eels, some differences 349 existed. This observation indicated that the life stage and gender influenced the processes by which 350 REEs were preferentially directed and stored by certain organs. In addition, it was possible to notice 351 that there were variations in organotropism depending on the considered groups of REEs: Σ REEs, 352 LREEs, MREEs and HREEs.

355

3.4. REE Pattern in Organs

Figure 3 shows the REE distribution, in each organ of yellow eels, female and male silver eels, like patterns. This figure highlighted the predominance of some REEs according to the organs and the eels considered. It also allowed to characterize which REEs were responsible for the concentrations of \sum REEs, LREEs, MREEs and HREEs studied in the latter parts of the present study.

The urinary system, gills, skin, liver and spleen presented similar profiles whatever the life
stage or the gender of eels. Indeed, Ce had the highest contribution, reaching 25.5 to 39.6 of ∑REEs.
The following elements were La (12.6-26.1%), Nd (15.3-22.2%) and Y (4.9-24.6%).

364 The gonads of female silver eels displayed a very particular profile which was completely 365 different from the gonads of other individuals (yellow eels and males silver eels) and even from the 366 trends of all the other organs. This profile was dominated by Gd which represented up to 74.2 % of 367 Σ REEs with a median concentration of 27.94±38.45 µg/kg dw. For comparison purpose, this value 368 was greater than Gd concentration in the liver $(9.76\pm7.38 \text{ µg/kg dw})$ which was the most REE 369 accumulating organ for female silver eels. However, it remained lower than Ce concentration 370 $(69.98 \pm 36.69 \,\mu\text{g/kg}\,\text{dw})$ which was the predominant element in the liver. In female silver eel gonads, 371 Nd (9.3%) and Ce (4.7%) represented the second and third most abundant elements, after Gd.

The male silver eel muscles showed similar patterns compared to female silver eels. These patterns were different from yellow eel muscles due to a greater contribution of Pr (13.9-16.9% in silver eels *vs* 3.0% in yellow ones). This great contribution of Pr in silver eel muscles, leading it to the 2^{nd} position among REEs after Ce, made the pattern also different from the other organs of all other eels. Regarding the yellow eel muscles, the largest contribution of Ce, reaching 72.7%, was also characteristic, compared to the other organs. 378 Regarding the bile, the REE profiles of female silver eels presented differences from those of
379 yellow eels and male silver eels without Y and Dy, and consequently higher contributions of Ce
380 (38.8%) and Nd (35.5%).

Whatever the organ, the life stage or the gender of eels considered, Tm and Lu were the least
 represented REEs, with <1.2% and < 0.4% respectively.

- 383
- 384 385

3.5. Correlations between Biological Parameters and REE Concentrations in Organs

Correlations between biological parameters and REE concentrations were analyzed to investigate the influence of some living characteristics on REE bioaccumulation and internal distribution. It was also interesting to look at the correlations between REE elements to evaluate some similar bioaccumulation patterns or dissimilarity.

- 390
- *391 3.5.1. Yellow Eels*

In yellow eels, BW and BL were negatively correlated to REE concentrations in gonads for La, Ce, Pr, Nd, Sm and Gd. These growth parameters were positively correlated to REE concentrations in spleen for all REEs (except Tm, Yb and Lu). There were also negative correlations of BL and BW with Pr, Nd, Dy and Er concentrations in muscles. Moreover, REE concentrations in the spleen and urinary system were positively correlated for all REE (except Tm, Yb and Lu). The concentrations in gills were positively correlated with La, Ce, Pr, Nd and Tb concentrations in gonads.

399 3.5.2. Silver Eels

400 In female silver eels, REE concentrations in the urinary system were negatively correlated
401 with the Fulton index for all REEs. Moreover, significant positive correlations were observed
402 between concentrations of liver for all REEs and age.

In male silver eels, the concentrations in gonads were positively correlated with concentrations measured in liver for all REEs (except Tm, Yb and Lu) and with concentrations measured in spleen for Sm, Gd, Tb, Dy, Ho and Er. Moreover, REE concentrations in gills were

15

406 positively correlated with the age for La, Ce and Nd. There were also positive correlations between 407 spleen and liver concentrations for Pr. Sm and Tb.

408

409 3.6. Parasitism influence

410 *3.6.1. Yellow eels*

411 A. crassus appeared to influence the REE contents in yellow eel organs. In fact, the levels of 412 La, Ce and Pr were significantly higher in muscles of parasitized individuals compared to non-413 parasitized ones. The concentrations in parasitized vs non-parasitized yellow eels were: 0.30 ± 0.08 vs 414 $0.150\pm0.04 \ \mu g/kg \ dw$ for La, $2.28\pm0.19 \ vs \ 1.31\pm0.38 \ \mu g/kg \ dw$ for Ce and $0.09\pm0.03 \ vs \ 0.07\pm$ 415 $0.02 \mu g/kg dw$ for Pr.

416 The same trend was observed in the gills, since the concentrations of all REEs, except Tm 417 and Lu, were significantly higher in parasitized individuals compared to non-parasitized ones. For 418 example, SREE concentrations in gills reached 152.36±63.69 and 78.44±39.92 µg/kg dw, for 419 parasitized and non-parasitized individuals respectively.

420 Regarding the gonads, parasitized individuals displayed concentrations of Y, La, Ce, Pr, Nd 421 and Sm significantly higher than those of non-parasitized ones. To illustrate this observation, 422 concentrations in parasitized vs non-parasitized vellow eels were: 1.25±0.97 vs 0.55±0.24 µg/kg dw, 423 2.68±2.49 vs 0.53±0.2 µg/kg dw, 5.80±4.92 vs 1.22±0.16 µg/kg dw, 0.73±0.54 vs 0.18±0.07 µg/kg 424 dw, 2.32±2.06 vs 0.45±0.27 µg/kg dw and 0.46±0.38 vs 0.09±0.04 µg/kg dw for Y, La, Ce, Pr, Nd 425 and Sm respectively.

426 The same result could be noticed for liver, where La concentrations were significantly higher 427 in parasitized yellow eels (31.44±11.73 µg/kg dw) compared to non-parasitized ones (16.15±7.4 µg/kg dw). REE concentrations in the urinary system, skin, spleen and bile were not 428 429 significantly affected by the A. crassus presence.

430 The \sum REE concentrations measured in the A. crassus collected in the swim bladder of yellow 431 eels were below the LOQ.

433 3.6.2. Silver eels

The female silver eels showed no significant difference of REE concentrations between parasitized and non-parasitized individuals for all organs. *A. crassus* collected in the swim bladder of female silver eels presented \sum REE concentrations of 12.56±13.18 µg/kg dw. The effect of the parasitism on REE concentrations on organs of male silver eels was not achieved because only two males among the 15 sampled individuals were not parasitized by *A. crassus*. The \sum REE concentration measured in *A.crassus* in the swim bladder of male silver eels was 19.78±24.83 µg/kg.

- 440
- 441 442

3.7. REE Bioaccumulation in Glass Eels

443 Figure 4A presents the bioaccumulation of Σ REEs and the different groups (LREEs, MREEs 444 and HREEs) in the whole body of glass eels at different pigmentation stages. The ΣREE median 445 concentrations were 8.68±0.8, 6.74±1.95 and 9.63±3.92 µg/kg dw for VIA0, VIA1 and VIA2 446 stages respectively. All REE concentrations are presented in Table 2 in the supplementary material. 447 No significant difference of concentrations between the 3 stages was highlighted. Moreover, there 448 was no significant difference between the three stages for LREEs, MREEs and HREEs as well. The 449 same accumulation trends were observed in glass eels than for the other life stages (yellow and 450 silver) presented in Figure 2. LREEs were accumulated in higher concentrations, followed by 451 HREEs and then by MREEs. The profiles of REE abundances in glass eels are shown in Figure 4B. 452 Whatever the stage of pigmentation, the predominant REEs in the whole body of glass eels were Ce 453 (27-32%) followed by Y (23%) and then La (14-16%).

454

456

455 4. Discussion

In this study, significant positive correlations between BL and BW were observed regardless of the life stage or the gender of the fish. In addition, the males were significantly smaller and lighter than the females. This sexual dimorphism was expected when the European eel is at the silver stage (Amérand et al., 2010; Durif, 2003). The ages between yellow eels and female silver eels or between female and male silver eels were not significantly different. This could be attributed to intraspecific variations (the yellow phase can last between 3 and 15 years) (Adam et al., 2008) as well as environmental effects which could lead to sedentary lifestyle of organisms for example (Feunteun et
al., 2000). This allowed the comparison of REE bioaccumulation between the different groups and to
investigate the influence of other variables such as physiological and morphological differences
related to life stage or gender.

467 The metal uptake and internal distribution in aquatic organisms are influenced by many 468 parameters linked to their living environment (compartment, pH and temperature), to the nature of 469 the element (physico-chemical characteristics, concentration, speciation and bioavailability) and to 470 the species (absorption, detoxification, feeding, trophic level, gender and life stage). It had already 471 been shown that the REE bioaccumulation in European eel muscles changed with the life stage and 472 the gender (Lortholarie et al., 2019). This observation was deepened in this work by investigating the 473 REE concentrations in skin, gills, spleen, urinary system, liver, bile and gonads in the yellow eels, 474 female and male silver eels.

475 Regarding the life stage influence, the comparison between yellow and female silver eels 476 revealed that the silver eels accumulated significantly more in the urinary system, muscles, gonads, 477 spleen and liver, while yellow eels accumulated more in gills. These two life stages presented REE 478 level differences but also differences in the element distribution throughout the organism. Indeed, the 479 strongest REE bioaccumulation was observed in gills for yellow eels and in liver for female silver 480 eels. To go further, the ΣREE organotropism in yellow eels was gills > liver > kidney > skin > 481 spleen > gonad > bile > muscles and in females silver eels was liver > kidney > spleen > gills >482 gonads > skin > muscles > bile. Moreover, by observing REEs in each individual organ, differences 483 had been observed in the REE contribution within certain organs: i.e. muscles, gonads, gills and bile. 484 Only urinary system, skin, liver and spleen presented similar profiles each other between yellow eels 485 and female silver eels. The yellow eel muscles had a specific profile too, relatively different from the 486 female silver eel muscles. The female silver eel gonads displayed a special profile, with a very 487 important contribution of Gd, followed by Nd, then by Ce which was usually the REE predominant 488 in the other gonads and organs in general. In other estuarine field studies (Amorim et al., 2007;

489 Moermond et al., 2001; Nozaki et al., 2000), Gd was not the predominant REE measured in sediment 490 and water. High concentrations of Gd in female silver eel gonads suggests a particular affinity of 491 this element for this organ. An in-depth study of this phenomenom is still needed, this observation 492 being not previously reported in the relatively rare studies about REE bioaccumulation in the gonads and particularly in fish gonads. Thus, further research must be performed to explain this specific Gd 493 494 accumulation and to evaluate if it could be related to Gd anthropogenic uses which could therefore 495 have a greater impact in estuarine and coastal areas than expected. Indeed, the Gd is a toxic ion with 496 very interesting paramagnetic properties, so it is often used as a contrast agent for MRI (Magnetic 497 Resonance Imaging) after being made inert by chelating it with a ligand. Nowadays, 25% of MRI use 498 complexed Gd, this is worrying because the Gd is not removed in wastewater treatment plants and 499 leads to aquatic environmental contamination (Parant et al., 2019). The REE bioaccumulation 500 differences observed between the two life stages, could be caused by the silvering. Indeed, this 501 metamorphosis leads to many external changes such as: coloration, thickening of the skin, 502 lengthening of the pectoral fins, increased mucus secretion and increased size of the sensory organs. 503 There is also an increase of gonad weight, a change of swim bladder wall, lipid accumulation, liver 504 structural and metabolic changes, discontinuation and digestive tract retraction (Adam et al. 2008; 505 Fontaine 1994; Durif et al.2015). The digestive tract retraction which takes place during this phase 506 can be an explanation of the higher REE concentrations in female silver eels (urinary system, 507 muscles, gonads, spleen and liver) than in yellow eels. Indeed, the intestines are an important 508 excretion route. The suppression of this excretion pathway may increase the pollutant accumulation 509 in silver eels. Moreover, during the metamorphosis into silver eel, the liver undergoes morphological 510 and physiological changes, these liver changes can cause higher accumulation of REEs. Finally, in 511 order to successfully migrate to the Sargasso Sea, the silver eels undergo a change in their muscles. 512 This may explain that the silver females accumulated more REEs than yellow females. Since yellow 513 eels were only future silver female eels, this conclusion cannot be drawn for males.

514 Regarding the gender influence, the comparison between male and female silver eels 515 highlighted differences. The females accumulated significantly more $\Sigma REEs$ in kidneys, gonads, 516 skin and spleen, whereas the males accumulated significantly more $\Sigma REEs$ in muscles and gills. 517 Thus, it is possible to conclude that, depending on gender, the preferential REE accumulation did not 518 occur in the same organs. The \sum REE organotropism in female was liver > kidney > spleen > gills > 519 gonads > skin > muscles > bile and in male, liver > gills > kidney > spleen > muscles > skin > 520 gonads > bile. As shown by the results, the gender impacted the levels of REEs but also the REE 521 abundances. These results could be explained by differences in physiology, behavior, detoxification 522 mechanisms and parasitism. A metabolism difference linked to sexual dimorphism could be another 523 explanation. Because of their smaller size, males spend more energy than females on their migration 524 (Amérand et al., 2010). The increased metabolism of males can lead to faster pollutant elimination, 525 which perhaps explains why males accumulated less REEs in the urinary system, the gonads, the 526 skin and the spleen. Another explanation may be due to male/female differences during the sexual 527 maturation phase, this hypothesis will be developed later in this discussion. The study of MacMillan 528 et al. (2017) conducted on brook trout (Salvelinus fontinalis) caught in Quebec lakes, showed a 529 significant effect of gender on Σ REE concentrations in liver. The values found for liver of males 530 were 2.5 times higher than those for females. This bioaccumulation difference between male and 531 female livers observed for the brook trout was not observed in the present study of the European eel.

REE organotropism study, i.e. the determination of the organs where REEs were 532 533 accumulated, is important in the ecotoxicity understanding of their fate and potentially their mode of 534 toxic action. Indeed, depending on the preferential accumulation organ, REEs will not have the same 535 toxicity or detoxification way. In this work, by ordering the organs from the most to the least 536 contaminated, some divergences were underscored between yellow and female silver eels as well as 537 between female and male silver eels. These results can be compared with other studies carried out on 538 fish caught in the wild or exposed in the laboratory. After a laboratory exposure of carps, Cyprinus carpio for 43 days to a mixture of Ce (0.27 mg/L), La (0.30 mg/L), Nd (0.29 mg/L), Pr (0.06 mg/L) 539

540 and Sm (0.25 mg/L),the organotropism determined for **LREEs** was: organs >> gills > skeleton > muscles (Hao et al., 1996). The results were less detailed but similar to 541 542 the results of the present study since the LREE organotropism (from La to Nd) was: liver > gills > 543 spleen > kidney > muscles > skin > gonads > bile for male silver eels. Other publications exist on 544 REE organotropism. For example, considering carps (Cyprinus carpio) exposed during 45 days at 545 concentrations of 0.50 mg/L of La (as a model of LREEs), Gd (for MREEs) and Y (for HREEs), the 546 organotropism was: internal organs > gills > skeleton > muscles (Qiang et al., 1994). The study of 547 Cardon et al., in 2020, conducted on rainbow trout exposed to different Y concentrations by 548 waterborne or by food obtained a completely different organotropism: 549 intestine >gills > liver >muscles. The lionfish from Cuba accumulated more \sum REEs in kidney with a 550 concentration of 0.053 mg/kg ww, higher than liver (0.018 mg/kg ww) or muscles (0.016 mg/kg ww) 551 (Squadrone et al., 2019). The comparison of results from literature and from the present study 552 remained complex because there was a lack of precision on the life stage or gender of fish. However, 553 the main observations made through this present work were the differences of results between yellow 554 eels and female silver eels as well as female and male silver eels. Moreover, the organs investigated 555 in this study were more numerous and individually studied which is not the case of some studies in 556 which the samples were pooled. The more detailed organotropism of the present work allowed to 557 highlight that liver, urinary system and spleen were the most accumulating organs of REEs. In fish, 558 the liver and the kidneys represent the main sites of metal sequestration and elimination (Pannetier et 559 al., 2016). These organs have an important REE accumulation which can be explained by their role 560 in the production of metallothionein and the metal detoxification. Finally, the results from the 561 literature mentioned above concerned in vivo controlled experiments with one up to five elements in 562 mixture. These conditions were greatly different to those of the present in situ study. In natural 563 environment, REEs are more or less all present in the medium. An interaction can occur each other 564 and with other contaminants or natural constituents. REE speciation and distribution in natural 565 environments can also be different from those found in controlled experiments.

566 In the present work, gills were often positioned among the organs with the highest REE 567 bioaccumulation. On one hand, this could be explained by the fact that they are in direct contact with 568 water therefore with the metal dissolved forms. They represent the organ leading to metal penetration 569 into the fish organism during the water filtration by the gill epithelium during the respiration. On the 570 other hand, the gills for freshwater, marine and euryhaline fish species are the main route of calcium absorption before the intestines (Baldisserotto, 2019; Flik et al., 1995). REEs are antagonists of 571 calcium absorption. In fact they are able to mimic Ca^{2+} due to their similar ionic radius (Figueiredo 572 573 et al., 2018; Martin & Richardson, 1979)). Indeed, ionic radius for the calcium is 0.09 nm and range 574 between 0.085 and 0.106 nm for Lu and La, or 0.089 nm for yttrium. In addition, REEs have a higher valence (3+) than calcium, so they are able to take the Ca²⁺ place (Cui et al., 2012). In fact, their +III 575 576 charge allows them to form stronger complexes (Switzer, 1978) and gives them a higher charge / volume ratio so they generally present higher affinities than Ca²⁺ for given binding sites (Evans, 577 578 1983). In addition, their charge can also promote a higher coordination number, this is particularly 579 true for smaller REEs which, by increasing their coordination number, can reach an ion radius closer to Ca^{2+} (Evans, 1983). However, the REE size is more important than their charge in the calcium 580 replacement (Switzer, 1978). The similarities between REEs and calcium can cause: higher REE 581 582 assimilation by the organisms, an alteration of the calcium cycle which can lead to a calcium 583 deficiency, a decrease in the essential element assimilation depending on calcium channels, an 584 inhibition of skeleton muscles, smooth muscle and heart muscle contractions. Moreover, REEs are 585 able to replace the calcium in proteins leading to a change of their function. In addition, the gills play 586 a role in the calcium excretion also. These potential characteristics of REEs could explain their high 587 concentrations particularly found in gills of yellow eels and male silver eels. The high REE 588 concentrations determined in the eel urinary system can also be linked to their similarity with 589 calcium. In fact, kidneys are the main site for divalent ion excretion like calcium (Flik et al., 1995). 590 Finally, during sexual maturation of female teleosts as eels, the vitellogenin production lead to an 591 increase of the calcium demand. This demand is satisfied by an increase of environmental

592 contributions or by internal remobilization (Persson et al., 1998). It had also been shown that during 593 this maturation period there is the incorporation of calcium in the oocytes (Persson et al., 1998) for female organisms. Considering REE and Ca²⁺ similarities, this could also explain why silver female 594 595 gonads accumulated more REEs, and especially Gd, than those of yellow females and silver males. 596 Concerning parasitism, in female yellow eels, non-parasitized individuals displayed significantly 597 lower REE concentrations compared to parasitized individuals in muscles (La, Ce and Pr), in gills 598 (all REEs, except Tm and Lu), in gonads (Y, La, Ce, Pr, Nd, and Sm) and in liver (La). On one hand, 599 the presence of A. crassus in the swim bladder is able to lead to numerous pathologies at 600 macroscopic (change in thickness and opacity of the swim bladder), immunological (induction of 601 non-specific and humoral responses) and physiological (alteration of gas excretion in the swim bladder, decrease in blood pressure, swimming ability, increased stress) levels (Habbechi et al., 602 603 2012; Kirk, 2003). Yellow parasitized eels could be more sensitive than yellow non-parasitized ones 604 to their environment and to the contaminants, less able to detoxification, which could lead to a 605 greater REE bioaccumulation. On the other hand, another hypothesis could be that more sensitive 606 individuals, for other reasons than parasitism like REE contamination, could be more parasitized. 607 Finally, it is impossible to conclude in favor of one of these two hypothesis. The reality was perhaps 608 a mix of these explanations but both could explain the link between parasitism and higher REE 609 concentrations in several studied tissues of female yellow eels. The results and corresponding 610 hypothesis for female vellow eels are not appropriated for silver ones because no significant 611 difference of REE contents was found between parasitized and non-parasitized individuals for all 612 organs. This result highlighted that the life stage of eels could have an influence on the link between 613 parasitism and REE concentration. In literature, some results exist about parasitism and its influence 614 on metal bioaccumulation but without considering different life stages and gender. The publication 615 of Palíková and Baruš (2003) did not show link between A. crassus and mercury bioaccumulation in 616 the European eel. Another study including different metals (Cd, Cr, Cu, Fe, Hg, Mn, Pb and Zn) did 617 not show significant differences in metal accumulations in European eel tissues according to the

618 parasitism factor (Genc et al., 2008). In this latter study, Genc et al. (2008) found Fe accumulation in 619 A. crassus up to 25.52 times than the muscle of A. anguilla caught in the Asi River (Hatay-Turkey). 620 In the present study, the same trend was observed for female silver eels with REE concentrations in 621 the parasites $(12.56\pm13.18 \,\mu\text{g/kg dw})$ about 2 times higher than in muscles $(6.66\pm2.13 \,\mu\text{g/kg dw})$ 622 and more than 9 times than in bile (1.29±0.88 µg/kg dw). Parasites could accumulate REEs in favor 623 of female silver eels. Nevertheless, the greater concentration of REEs in parasites compared to 624 muscle and bile of female silver eels was not observed for female yellow eels, as REE were not 625 detected in the A. crassus of yellow individuals. The life stage seemed to have an influence on the 626 parasite REE contents also, probably by affecting mechanisms in parasites or in parasitized 627 organisms. In the study of Schneebauer et al., . (2017), infected yellow eels presented in the swim bladder 1675 genes differentially transcribed compared to uninfected yellow eels, but silver eels only 628 629 exhibited 291 differentially transcribed genes compared to uninfected silver eels. In 2016, 630 Schneebauer *et al.*, have shown that silver eels have poorer detoxification of ROS when they are 631 parasitized, while yellow eels do not show significant differences. It seems that the two life stages do 632 not react in the same way to parasitism, but the result obtained in this study remained difficult to 633 explain since knowledge on the sensitivity of life stage to parasitism are limited.

634 Regarding the glass eels, the pigmentation stage (VI A0, VI A1 and VI A2) did not seem to 635 influence the REE bioaccumulation. Glass eels had median REE concentrations ranging from 6.74 to 636 9.63 µg/kg dw respectively for VIA1 and VIA2 stages. After the metamorphosis of the 637 leptocephalus into the glass eels, individuals stop feeding and start eating again at the VI A3 stage. 638 The glass eels studied in this work were caught during this fasting period. Their REE concentrations 639 could be the result of external contamination reflecting the environmental concentrations. In another 640 study, glass eels exposed during 3 days at 120 ng/L of lanthanum chloride in water accumulated 1.6-641 4.7, 1.1-6.7 and 4.8-13.0 ng/g dw in head, skinless body and viscera respectively (so with an 642 organotropism of viscera > skinless body > head) (Figueiredo et al. 2018). These results and the 643 concentrations found in the present work were in the same order of magnitude. The REE 644 predominance in the whole body of glass eels showed a profile closer to the skin of yellow and silver
645 eels than female gonads dominated by Gd. The high abundance of Gd observed in female silver eel
646 gonads were not present in glass eels, suggesting no or weak potential maternal transfer.

647

648 5. Conclusion

649 This study reported data on the REE organ contents in wild organisms of European eels from 650 the Loire estuary. The bioaccumulation examination in glass eels, yellow eels, male and female 651 silver eels allowed to highlight the life stage and the gender influence on the REE concentrations, REE abundance in organs and hence their organotropism. This work, through a detailed 652 653 organotropism, made it possible to underscore the liver, kidneys and spleen importance in the REE 654 bioaccumulation. This can be explained by the role in the sequestration and elimination metals from 655 these organs. In addition, the gills also appeared as organs with a high REE accumulation. This may 656 be linked to the REE ability to mimic calcium; indeed, the gills are the main route of calcium entry in 657 fish. To go further, the REE contribution study have shown that the REEs present in higher 658 concentration can be different between some organs, life stages and gender. The silver female gonads 659 attracted attention because they presented a very special profile dominated by the Gd presence. 660 Moreover, the Anguillicola crassus presence seemed impact REE bioaccumulation in yellow eels 661 because, parasitized yellow eels accumulated more in muscles, gills, gonads and liver than non-662 parasitized individuals. Regarding glass eels, the stage of pigmentation did not seem to have an 663 influence on the REE bioaccumulation. In addition, the REE contribution profile in the whole body 664 was close to the profile of the yellow and silver eel skin. In order to go deeper into the REE 665 organotropism determination in European eels and better understand the influence of parameters 666 such as gender or life stage it would be necessary to investigate the kinetic processes of 667 internalization underlying this organotropism. It would also be interesting to explore reasons why Gd 668 accumulated so much in female silver eel gonads and if it is related to Gd anthropogenic uses.

6. Acknowledgments

671	The authors acknowledge the French ministry for the funding of doctoral position of M.
672	Lortholarie. Furthermore, we would also thank the fishermen in the Loire estuary and Y. François for
673	providing technical assistance during the field sampling and the ICP-MS analysis respectively.
674 675 676	7. Bibliography
677	Adam, G., Feunteun, E., Prouzet, P., & Rigaud, C. (2008). L'anguille européenne : Indicateurs
678	d'abondance et de colonisation. Éditions Quæ.
679	http://public.eblib.com/choice/publicfullrecord.aspx?p=3399257
680	Akagi, T., & Edanami, K. (2017). Sources of rare earth elements in shells and soft-tissues of bivalves
681	from Tokyo Bay. Marine Chemistry, 194, 55- 62.
682	https://doi.org/10.1016/j.marchem.2017.02.009
683	Amérand, A., Vettier, A., Moisan, C., Belhomme, M., & Sébert, P. (2010). Sex-related differences in
684	aerobic capacities and reactive oxygen species metabolism in the silver eel. Fish Physiology
685	and Biochemistry, 36(3), 741 - 747. https://doi.org/10.1007/s10695-009-9348-0
686	Amorim, E. P., Fávaro, D. I. T., & Berbel, G. B. B. (2007). METAL AND TRACE ELEMENT
687	ASSESSMENT OF ESTUARY SEDIMENTS FROM SANTOS, BRAZIL, BY NEUTRON
688	ACTIVATION AND ATOMIC ABSORPTION TECHNIQUES. 9.
689	Baldisserotto, B. (2019). Fish Osmoregulation. CRC Press.
690	Barak, N. AE., & Mason, C. F. (1990). Mercury, cadmium and lead in eels and roach : The effects
691	of size, season and locality on metal concentrations in flesh and liver. Science of The Total
692	Environment, 92, 249 - 256. https://doi.org/10.1016/0048-9697(90)90334-Q
693	Bordajandi, L. R., Gómez, G., Fernández, M. A., Abad, E., Rivera, J., & González, M. J. (2003).
694	Study on PCBs, PCDD/Fs, organochlorine pesticides, heavy metals and arsenic content in

- 695 freshwater fish species from the River Turia (Spain). Chemosphere, 53(2), 163 171.
- 696 https://doi.org/10.1016/S0045-6535(03)00417-X
- 697 Bru, k., Christmann, P., Labbé, J. F., & Lefebvre, G. (2015). Panorama 2014 du marché des Terres
 698 Rares.
- 699 http://www.mineralinfo.fr/sites/default/files/upload/documents/Panoramas_Metaux_Strateg/r
 700 p-65330-fr-terresrarespublic.pdf
- Cardon, P.-Y., Roques, O., Caron, A., Rosabal, M., Fortin, C., & Amyot, M. (2020). Role of prey
 subcellular distribution on the bioaccumulation of yttrium (Y) in the rainbow trout.
- 703 Environmental Pollution, 258, 113804. https://doi.org/10.1016/j.envpol.2019.113804
- 704 Chaillou, G., Anschutz, P., Lavaux, G., & Blanc, G. (2006). Rare earth elements in the modern
- sediments of the Bay of Biscay (France). Marine Chemistry, 100(1), 39-52.
- 706 https://doi.org/10.1016/j.marchem.2005.09.007
- 707 Cui, J., Zhang, Z., Bai, W., Zhang, L., He, X., Ma, Y., Liu, Y., & Chai, Z. (2012). Effects of rare earth
- *elements La and Yb on the morphological and functional development of zebrafish embryos.*
- Journal of Environmental Sciences, 24(2), 209-213. https://doi.org/10.1016/S1001-
- 710 0742(11)60755-9
- 711 Dahlqvist, R., Andersson, P. S., & Ingri, J. (2005). The concentration and isotopic composition of
- 712 *diffusible Nd in fresh and marine waters. Earth and Planetary Science Letters, 233(1), 9- 16.*
- 713 https://doi.org/10.1016/j.epsl.2005.02.021
- 714 Durif, C. (2003). La migration d'avalaison de l'anguille européenne. 350.
- 715 Durrieu, G., Maury-Brachet, R., Girardin, M., Rochard, E., & Boudou, A. (2005). Contamination by
- *heavy metals (Cd, Zn, Cu, and Hg) of eight fish species in the Gironde estuary (France).*
- 717 Estuaries, 28(4), 581 591. https://doi.org/10.1007/BF02696069
- 718 Evans, C. H. (1983). Interesting and useful biochemical properties of lanthanides. Trends in
- 719 Biochemical Sciences, 8(12), 445 449. https://doi.org/10.1016/0968-0004(83)90032-4

- 720 Feunteun, E., Acou, A., Laffaille, P., & Legault, A. (2000). European eel (Anguilla anguilla) :
- 721 Prediction of spawner escapement from continental population parameters. Canadian
- Journal of Fisheries and Aquatic Sciences, 57(8), 1627 1635. https://doi.org/10.1139/f00 096
- Figueiredo, C., Griloa, T., Lopes, C., Brito, P., Diniz, M., Caetano, M., Rosa, R., & Raimundo, J.
- (2018, septembre). Accumulation, elimination and neuro-oxidative damage under lanthanum
 exposure in glass eels (Anguilla anguilla)—ScienceDirect.
- 727 https://www.sciencedirect.com/science/article/pii/S0045653518308713
- 728 Flik, G., Verbost, P. M., & Bonga, S. E. W. (1995). 12 Calcium Transport Processes in Fishes. In C.
- M. Wood & T. J. Shuttleworth (Éds.), Fish Physiology (Vol. 14, p. 317- 342). Academic
 Press. https://doi.org/10.1016/S1546-5098(08)60251-4
- 731 Genc, E., Sangun, M. K., Dural, M., Can, M. F., & Altunhan, C. (2008). Element concentrations in
- *the swimbladder parasite Anguillicola crassus (nematoda) and its host the European eel,*
- 733 Anguilla anguilla from Asi River (Hatay-Turkey). Environmental Monitoring and
- 734 Assessment, 141(1), 59- 65. https://doi.org/10.1007/s10661-007-9878-9
- 735 Gony, S. (1990). Short Note on the Effects of Cadmium on the Gill of the Glass Eel (Anguilla
- 736 anguilla). Internationale Revue Der Gesamten Hydrobiologie Und Hydrographie, 75(6),
- 737 835-836. https://doi.org/10.1002/iroh.19900750625
- 738 Grellier, P., Huet, J., & DESAUNAY, Y. (1991). Stades pigmentaires de la civelle Anguilla anguilla
- 739 (L.) dans les estuaires de la Loire et de la Vilaine. 19.
- 740 Habbechi, B. H., Kraiem, M. M., & Élie, P. (2012). Étude de la contamination de l'anguille
- 741 européenne (Anguilla anguilla L., 1758) par Anguillicoloides crassus dans quelques
- 742 hydrosystèmes de la Tunisie septentrionale : Analyse de son impact sur les paramètres de
- *croissance.* 18.

- 744 Hao, S., Xiaorong, W., Zhaozhe, H., Chonghua, W., Liansheng, W., Lemei, D., Zhong, L., & Yijun, C.
- 745 (1996). Bioconcentration and elimination of five light rare earth elements in carp (Cyprinus
- 746 *carpio L.*). *Chemosphere*, 33(8), 1475 1483. *https://doi.org/10.1016/0045-6535(96)00286-X*
- 747 Hirata, S., Kajiya, T., Aihara, M., Honda, K., & Shikino, O. (2002). Determination of rare earth
- *elements in seawater by on-line column preconcentration inductively coupled plasma mass*
- 749 spectrometry. Talanta, 58(6), 1185 1194. https://doi.org/10.1016/S0039-9140(02)00418-6
- 750 Johannesson, K. H., Palmore, C. D., Fackrell, J., Prouty, N. G., Swarzenski, P. W., Chevis, D. A.,
- 751 Telfeyan, K., White, C. D., & Burdige, D. J. (2017). Rare earth element behavior during
- 752 groundwater-seawater mixing along the Kona Coast of Hawaii. Geochimica et
- 753 *Cosmochimica Acta*, 198, 229- 258. https://doi.org/10.1016/j.gca.2016.11.009
- *Keasler, K. M., & Loveland, W. D. (1982). Rare earth elemental concentrations in some Pacific Northwest rivers. Earth and Planetary Science Letters, 61(1), 68-72.*
- 756 https://doi.org/10.1016/0012-821X(82)90039-5
- 757 Kirk, R. S. (2003). The impact of Anguillicola crassus on European eels. Fisheries Management and
 758 Ecology, 10(6), 385- 394. https://doi.org/10.1111/j.1365-2400.2003.00355.x
- 759 Lerat-Hardy, A., Coynel, A., Dutruch, L., Pereto, C., Bossy, C., Gil-Diaz, T., Capdeville, M.-J.,
- 760 Blanc, G., & Schäfer, J. (2019a). Rare Earth Element fluxes over 15 years into a major
- 761 European Estuary (Garonne-Gironde, SW France) : Hospital effluents as a source of
- *increasing gadolinium anomalies. Science of The Total Environment, 656, 409 420.*
- 763 https://doi.org/10.1016/j.scitotenv.2018.11.343
- 764 Lerat-Hardy, A., Coynel, A., Dutruch, L., Pereto, C., Bossy, C., Gil-Diaz, T., Capdeville, M.-J.,
- 765 Blanc, G., & Schäfer, J. (2019b). Rare Earth Element fluxes over 15 years into a major
- 766 European Estuary (Garonne-Gironde, SW France) : Hospital effluents as a source of
- *increasing gadolinium anomalies. Science of The Total Environment, 656, 409 420.*
- 768 https://doi.org/10.1016/j.scitotenv.2018.11.343

- 769 Li, J.-X., Zhu, Z.-W., Yin, X.-F., Han, B., Zheng, L., Wang, J.-T., & Wang, X.-R. (2015). Analysis of
- 770 Contents and Distribution Patterns of Rare Earth Elements in the Surface Sediments of the
- 771 South Mid-Atlantic Ridge. Chinese Journal of Analytical Chemistry, 43(1), 21 26.
- 772 https://doi.org/10.1016/S1872-2040(15)60796-4
- 773 Lortholarie, M., Zalouk-Vergnoux, A., Couderc, M., Kamari, A., François, Y., Herrenknecht, C., &
- 774 Poirier, L. (2019). Rare earth element bioaccumulation in the yellow and silver European eel
- 775 (Anguilla anguilla) : A case study in the Loire estuary (France). Science of The Total

776 Environment, 134938. https://doi.org/10.1016/j.scitotenv.2019.134938

- 777 MacMillan, G. A., Chételat, J., Heath, J. P., Mickpegak, R., & Amyot, M. (2017). Rare earth
- *elements in freshwater, marine, and terrestrial ecosystems in the eastern Canadian Arctic.*
- *Environmental Science: Processes & Impacts, 19(10), 1336-1345.*
- 780 https://doi.org/10.1039/C7EM00082K
- Mansouri, D., Fegrouche, R., Allami, H., & Fadli, M. (2018). Bioaccumulation of some metallic
 elements in the gills, the muscle and the liver of the eel (Anguilla anguilla L.) and the cabot
 mullet (Mugil cephalus L.) of the estuary of Oued Sebou (Morocco).
- 784 *Martin, B., & Richardson, F. S. (1979). Lanthanides as probes for calcium in biological systems.*
- 785 Quarterly Reviews of Biophysics, 12(2), 181 209.
- 786 https://doi.org/10.1017/S0033583500002754
- 787 Moermond, C. T. A., Tijink, J., Wezel, A. P. van, & Koelmans, A. A. (2001). Distribution, speciation,
- and bioavailability of lanthanides in the Rhine-Meuse estuary, The Netherlands.
- 789 Environmental Toxicology and Chemistry, 20(9), 1916-1926.
- 790 https://doi.org/10.1002/etc.5620200909
- 791 Négrel, P., Guerrot, C., Cocherie, A., Azaroual, M., Brach, M., & Fouillac, C. (2000). Rare earth
- 792 elements, neodymium and strontium isotopic systematics in mineral waters : Evidence from
- *the Massif Central, France. Applied Geochemistry, 15(9), 1345-1367.*
- 794 https://doi.org/10.1016/S0883-2927(00)00008-1

- Neto, A. F., Costa, J. L., Costa, M. J., Pereira, M. E., Duarte, A., Caçador, I., & Domingos, I.
- 796 (2011). Accumulation of metals in Anguilla anguilla from the Tagus estuary and relationship
- *to environmental contamination. Journal of Applied Ichthyology, 27(5), 1265 1271.*
- 798 https://doi.org/10.1111/j.1439-0426.2011.01814.x

- *Nozaki, Y., Lerche, D., Alibo, D. S., & Snidvongs, A. (2000). The estuarine geochemistry of rare*
- 801 Cosmochimica Acta, 64(23), 3983 3994. https://doi.org/10.1016/S0016-7037(00)00473-7

earth elements and indium in the Chao Phrava River, Thailand. Geochimica et

- 802 Oliveira, M., Serafim, A., Bebianno, M. J., Pacheco, M., & Santos, M. A. (2008). European eel
- 803 (Anguilla anguilla L.) metallothionein, endocrine, metabolic and genotoxic responses to
- 804 *copper exposure. Ecotoxicology and Environmental Safety, 70(1), 20- 26.*
- 805 https://doi.org/10.1016/j.ecoenv.2007.10.034
- 806 Parant, M., Sohm, B., Flayac, J., Perrat, E., Chuburu, F., Cadiou, C., Rosin, C., & Cossu-Leguille,
- 807 *C.* (2019). Impact of gadolinium-based contrast agents on the growth of fish cells lines.
- 808 *Ecotoxicology and Environmental Safety, 182, 109385.*
- 809 https://doi.org/10.1016/j.ecoenv.2019.109385
- 810 Persson, P., Sundell, K., Björnsson, B. T., & Lundqvist, H. (1998). Calcium metabolism and
- 811 osmoregulation during sexual maturation of river running Atlantic salmon. Journal of Fish
- 812 Biology, 52(2), 334- 349. https://doi.org/10.1111/j.1095-8649.1998.tb00801.x
- 813 Pierron, F., Baudrimont, M., Bossy, A., Bourdineaud, J.-P., Brèthes, D., Elie, P., & Massabuau, J.-
- 814 C. (2007). Impairment of lipid storage by cadmium in the European eel (Anguilla anguilla).
- 815 Aquatic Toxicology, 81(3), 304- 311. https://doi.org/10.1016/j.aquatox.2006.12.014
- 816 Qiang, T., Xiao-rong, W., Li-qing, T., & Le-mei, D. (1994). Bioaccumulation of the rare earth
- 817 elements lanthanum, gadolinium and yttrium in carp (Cyprinus carpio). Environmental
- 818 Pollution, 85(3), 345 350. https://doi.org/10.1016/0269-7491(94)90057-4
- 819 Ribeiro, C. A. O., Vollaire, Y., Sanchez-Chardi, A., & Roche, H. (2005). Bioaccumulation and the
- 820 effects of organochlorine pesticides, PAH and heavy metals in the Eel (Anguilla anguilla) at

- 821 the Camargue Nature Reserve, France. Aquatic Toxicology, 74(1), 53-69.
- 822 https://doi.org/10.1016/j.aquatox.2005.04.008
- 823 Santos, M. A., & Hall, A. (1990). Influence of inorganic lead on the biochemical blood composition
- 824 of the eel, Anguilla anguilla L. Ecotoxicology and Environmental Safety, 20(1), 7-9.
- 825 https://doi.org/10.1016/0147-6513(90)90040-C
- 826 Schneebauer, G., Dirks, R. P., & Pelster, B. (2017). Anguillicola crassus infection affects mRNA
- 827 *expression levels in gas gland tissue of European yellow and silver eel. PLOS ONE, 12(8),*

828 e0183128. https://doi.org/10.1371/journal.pone.0183128

- 829 Schneebauer, G., Hanel, R., & Pelster, B. (2016). Anguillicola crassus impairs the silvering-related
- 830 enhancements of the ROS defense capacity in swimbladder tissue of the European eel
- 831 (Anguilla anguilla). Journal of Comparative Physiology B, 186(7), 867-877.
- 832 https://doi.org/10.1007/s00360-016-0994-0
- 833 Squadrone, S., Brizio, P., Stella, C., Mantia, M., Favaro, L., Biancani, B., Gridelli, S., Da Rugna, C.,
- & Abete, M. C. (2019). Differential Bioaccumulation of Trace Elements and Rare Earth
- 835 Elements in the Muscle, Kidneys, and Liver of the Invasive Indo-Pacific Lionfish (Pterois
- 836 spp.) from Cuba. Biological Trace Element Research. https://doi.org/10.1007/s12011-019-
- 837 01918-w
- 838 Storelli, M. M., Barone, G., Garofalo, R., & Marcotrigiano, G. O. (2007). Metals and
- 839 organochlorine compounds in eel (Anguilla anguilla) from the Lesina lagoon, Adriatic Sea
- 840 (Italy). Food Chemistry, 100(4), 1337-1341.
- 841 https://doi.org/10.1016/j.foodchem.2005.10.071
- Switzer, M. E. (1978). The lanthanide ions as probes of calcium ion binding sites in biological
 systems. Science Progress (1933-), 65(257), 19- 30. JSTOR.
- 844 Tabouret, H., Bareille, G., Mestrot, A., Caill-Milly, N., Budzinski, H., Peluhet, L., Prouzet, P., &
- 845 Donard, O. F. X. (2011). Heavy metals and organochlorinated compounds in the European

846	eel (Anguilla anguilla) from the	Adour estuary and associated wetlands	(France). Journal of
-----	----------------------------------	---------------------------------------	----------------------

- 847 Environmental Monitoring, 13(5), 1446- 1456. https://doi.org/10.1039/C0EM00684J
- 848 Zhou, B., Li, Z., & Chen, C. (2017). Global Potential of Rare Earth Resources and Rare Earth
- 849 Demand from Clean Technologies. Minerals, 7(11), 203. https://doi.org/10.3390/min7110203



Figure 1: Sampling sites of glass eels (blue triangle), yellow (yellow ring) and silver (gray square) eels captured in the Loire estuary (1: Cordemais, 2: Haute-Indre and 3: Varades).



Figure 2: Median and Median Absolute Deviation for $\sum REE$ (black bars), LREE (from La to Nd, black striped bars), MREE (from Sm to Gd, blue bars) and HREE (from Tb to Lu and Y, blue striped bars) concentrations (μ g/kg dw) in organs of (A) yellow eels (n=15), (B) female (n=14) and (C) male (n=15) silver eels.

#: significant differences between values of yellow eels and female silver eels with Kruskal Wallis test (p-value≤0.05)

*: significant differences between values of female and male silver eels with Kruskal Wallis test (p-value≤0.05)



Figure 3: REE contribution (%) in each organ for yellow eels, female and male silver eels. US: Urinary system



Figure 4: (A) Median and Median Absolute Deviation (error bars) for REE (back bars), LREE (from La to Nd, black striped bars), MREE (from Sm to Gd, blue bars) and HREE (from Tb to Lu and Y, blue striped bars) concentrations (ng/kg dw) in different stages of glass eels: VI A0 (n=15), VI A1 (n=15) and VI A2 (n=15). (B) REE contribution (%) in the whole body of glass eels.

Table 1: Classification of yellow eel organs from the more to the less concentrated in \sum REEs, LREEs, MREEs and HREEs

	More							Less	
	concentr	concentrated concentrated							
∑REEs	Gills	Liver	Urinary	Skin	Spleen	Gonads	Bile	Muscle	
			system						
LREEs	Gills	Liver	Urinary	Skin	Spleen	Gonads	Bile	Muscle	
			system						
MREEs	Gills	Liver	Urinary	Skin	Spleen	Gonads	Bile	Muscle	
			system						
HREEs	Gills	Urinary	Liver	Skin	Spleen	Bile	Gonads	Muscle	
		system							

Table 2: Classification of female silver eel organs from the more to the less concentrated in \sum REEs, LREEs, MREEs and HREEs

	More Les concentrated concentrated							
∑REEs	Liver	Urinary	Spleen	Gills	Gonads	Skin	Muscle	Bile
		system						
LREEs	Liver	Urinary	Spleen	Gills	Skin	Gonads	Muscle	Bile
		system						
MREEs	Gonads	Liver	Urinary	Spleen	Gills	Skin	Muscle	Bile
			system					
HREEs	Urinary	Spleen	Liver	Gills	Skin	Gonads	Muscle	Bile
	system							

Table 3: Classification of male silver eel organs from the more to the less concentrated in \sum REEs, LREEs, MREEs and HREEs

	More							Less	
	concentrated concentrate								
∑REEs	Liver	Gills	Urinary	Spleen	Muscle	Skin	Gonads	Bile	
			system						
LREEs	Liver	Gills	Spleen	Urinary	Muscle	Skin	Gonads	Bile	
				system					
MREEs	Liver	Gills	Urinary	Skin	Spleen	Gonads	Muscle	Bile	
			system						
HREEs	Gills	Urinary	Liver	Spleen	Gonads	Skin	Muscle	Bile	
		system							

Graphical abstract:

