Nursery function rehabilitation projects in port areas can support fish populations but they remain less effective than ensuring compliance to fisheries management

Supporting information

## S1 Model description

## S1.1. Population structure

In the model, the $D$. sargus population is divided into 15 age classes ( $C_{0}$ to $C_{14}$ ), each one year apart except for $C_{0}$ and $C_{1}$, in order to cover the different life stages of $D$. sargus (Belharet et al., 2020). The individuals in $\mathrm{C}_{0}$ are between 1 and 6 months old. They are the young of the year (YOY) arriving in $\mathrm{C}_{0}$ at $D$. sargus' cycle starting month: April (Lloret and Planes, 2003). Individuals in $C_{1}$ are between 7 months and one and a half years old, and those in $\mathrm{C}_{14}$ are 14 years old or older (Fig. 1). Each year, the transition from one age class to the next occurs in late September, aligning with the departure of


Figure 1: Age classes timetable in the model
juvenile nurseries from the study area. (Harmelin-Vivien et al., 1995; Planes and Romans, 2004).

## S1.1.1. Growth

During the first year of $D$. sargus' development, the growth in length is considered linear (equation 1a, Planes et al., (1998). Older individuals' growth follows a Von Bertalanffy equation (equation 1b) (Belharet et al., 2020; Hussein et al., 2011):

$$
\begin{aligned}
& L(a)=(a+m) * \frac{e q-1}{e q}+1 \\
& L(a)=l_{\infty}\left(1-\mathrm{e}^{-k\left(\frac{a+m}{12}-t_{0}\right)}\right)
\end{aligned} \quad(\text { eqn } 1 a)
$$

with $l_{\infty}$ being $D$. sargus maximum length ( cm ), $t_{0}$ the hypothetical time at which the size of the individuals is zero, $k$ the Brody's growth coefficient, $a$ the age in months of the individuals, $m$ the month in the cycle at which the calculation is performed and eq the succession point between the two equations (Table 1).

## S1.1.2. Weight and maturity rate

The average weight $W(L)$ and maturity rate of individuals $M(L)$ are computed as a function of their size ( $L$ ) according to the following equations (equation 2) (Le Cren, 1951) and (equation 3) (Gonçalves and Erzini, 2000):

$$
\begin{gathered}
W(L)=a_{w} * L^{b_{w}} \\
M(L)=\frac{1}{1+e^{-m_{50}\left(L-c_{m}\right)}}
\end{gathered}
$$

With $a_{p}$ and $b_{p}$ being two coefficients, $m_{50}$ the size of $D$. sargus for which the maturity rate is $50 \%$ and $c_{m}$ the slope of the maturity curve (Table 1).

## S1.1.3. Breeding and recruitment

The reproduction of $D$. sargus in our study area occurs between February and April, and the peak of juvenile settlement between April and June (Harmelin-Vivien et al., 1995; Lloret and Planes, 2003; Planes et al., 1998; Vigliola et al., 1998). D. sargus is a non-binding protandrous hermaphrodite fish. The proportion of females able to breed $\left(s_{r}\right)$ in each class size $(L)$ is determined by the following equation (Mouine et al., 2007):

$$
s_{r}=1-2 e^{-0.15 L} \quad(\text { eqn } 4)
$$

The reproduction rate $\left(r_{r}\right)$ corresponds to the number of eggs that can be laid by individuals within each class according to their sex ratio $\left(s_{r}\right)$, and their size $L$ (equation 5). The coefficients af and $b f$ were estimated by Belharet et al. (2020):

$$
r_{r}=s_{r} \cdot a f . L^{b f}
$$

The breeding algorithm scans all classes living in the areas defined as "breeding area" during the breeding period (March). This algorithm takes in the reproductive rate $r_{r}$, maturity rate $M(L)$ and number of individuals in each group $N b_{\text {ind }}$, and multiplies them (equation 6) by the estimated egg and larval survival rate $\left(l_{s r}\right)$ (Belharet et al., 2020). The breeding matrix is browsed and filled with the number of eggs or future larvae within each zone, denoted $N_{L}$.

$$
\begin{equation*}
N_{L}=r_{r} * M(L) * N b_{i n d} * l_{s r} \tag{eqn6}
\end{equation*}
$$

## S1.1.4. Recruitment

Recruitment is limited by the nursery area capacity to host juveniles, which is different in natural areas ( $R_{\text {max_nat }}$ ) and in port area ( $R_{\text {max_port }}$ ) (Planes et al., 1998 ). This capacity represents the number of individuals able to settle in the recruitment area and to join $\mathrm{C}_{0}$ equations 7 a and 7 b :

$$
R_{\text {max_nat }}=L_{n} * d_{m} * S_{w}(\operatorname{eqn} 7 a) \quad R_{\max ^{-}}=L_{p} * d_{m} * S_{w}(\text { eqn } 7 b)
$$

Where $L_{n}$ is the coastline length (meters) of the nursery area, $d_{m}$ the maximum number of juveniles that can settle per meter of coastline, $S_{w}$ the number of settlement waves and $L_{p}$ the length of docks were $D$. sargus juveniles can settle (Table 1).

The number of recruits $N_{r}$ in each area ( $\mathrm{C}_{0}$ in nurseries and ports) is defined by a Beverton-Holt equation (equation 8) (Belharet et al., 2020):

$$
N_{r}=N_{L}\left(\frac{R_{\max }}{N_{L}+R_{\max } * g}\right)
$$

With $N_{L}$ being the initial number of larvae arriving in the area and $R_{\max } * g$ the intensity of the density dependency (Table 1).

## Determination of juvenile habitat coastal length

The length of the "natural" and porta' benthic settlement zones ( $L_{n}$ and $L_{p}$ ) were measured on the basis of a map overlay according to criteria defined by previous studies (Cheminée et al., 2021, 2011; García-Rubies and Macpherson, 1995; Harmelin-Vivien et al., 1995; Ventura et al., 2014)) such as the bathymetry of the studied area (© Shom-IGN, 2021), benthic biocenosis (Andromède Océanologie, 2021) and exposure to winds and currents (Ifremer, 2021).

The coastal length of natural nurseries covers all sheltered areas within a depth of less than 3 m , a slope of less than $23 \%$ and a biocenosis corresponding to either infralittoral rocks or loose infralittoral bottoms such as gravel and sand (García-Rubies and Macpherson, 1995; Harmelin-Vivien et al., 1995). This resulted in an estimation of $55,820 \mathrm{~m}$ of coastline (at a scale of $1: 500$ ). Using GIS we measured the linear length of port docks able to support the benthic settlement of juvenile fish from aerial photos (IGN and BRGM, 2022), resulting in an estimation of $24,751 \mathrm{~m}$ (on a scale of $1: 500$ ). It corresponds to all the quays in the area with a depth of less than 2 m . The total length of port docks was measured from aerial photos (IGN and BRGM, 2022) and estimated at 60060 m (comprising 24751m of docks already welcoming juvenile fish and 35309 m welcoming none, on a scale of 1:500) (Table 1).

## S1.1.4. Natural mortality

The natural mortality rate $\left(M_{r}\right)$ of $D$. sargus depends on the life stage (Belharet et al., 2020; Bouchoucha et al., 2016; Hussein et al., 2011; Planes et al., 1998) as well as the benthic settlement zone (natural nursery vs. ports). For natural nurseries, Planes and Romans, (2004) estimated 80.8\% mortality of juveniles in the first four months after benthic settlement. In the port areas, the juvenile mortality rate during the first 4 months is very high due to the lack of refuge areas (Bouchoucha et al., 2016). In absence of data for this value, we set it to $99 \%$ (assumption not verified to date but on which most current nursery rehabilitation projects are based). We made the strong assumption that, for fish older than $C_{0}$, there is no excess natural mortality of individuals having lived their first months in port areas compared to those from natural nurseries. For classes $C_{1}$ to $C_{14}$, the annual natural mortality rates were set to $26 \%$ (Belharet et al., 2020).

These mortality rates are first transformed into survival rates $\left(t_{s}\right)$ with ( $t_{s}=1-M_{r}$ ), expressed between 0 and 1, and then transformed in instantaneous mortalities (Table 2) according to the following equation (equation 9).

$$
M_{i}=-\ln \left(t_{s}\right) \quad(\text { eqn } 9)
$$

|  | Parameter | Value | Reference |
| :---: | :---: | :---: | :---: |
| Growth |  |  |  |
| $t_{\infty}$ | Asymptotic body length (cm) | 45 cm | (Froese and Pauly, 2022) |
| $t_{0}$ | Age at length zero (yr) | -0.99 | (Belharet et al., 2020) |
| $k$ | Brody growth coefficient ( $\mathrm{yr}^{-1}$ ) | 0.17 | (Belharet et al., 2020) |
| $e q$ | Continuity factor | 11.93 |  |
| Weight |  |  |  |
| $a_{w}$ | Scale parameter (g cm-b ) | 0.016 | (Belharet et al., 2020) |
| $b_{w}$ | Exponent | 3.05 | (Belharet et al., 2020) |
| Maturity |  |  |  |
| $m_{50}$ | Length of 50\% mature | 20.06 cm | (Mouine et al., 2012) |
| $c_{m}$ | Slope of the maturity curve | 1.26 | (Mouine et al., 2012) |
| Breeding |  |  |  |
| $a f$ | Coefficient | 0.084 | (Belharet et al., 2020) |
| bf | Coefficient | 4.51 | (Belharet et al., 2020) |
| $l_{s r}$ | Larval survival rate | 0.008 | (Belharet et al., 2020) |
|  | Breeding month | March | (Lloret and Planes, 2003) |
| Recruitment |  |  |  |
| $L_{n}$ | Coastal length of nursery area (m) | 55820 | Mesured on maps |
| $L_{p}$ | Length of quay were juveniles fish settle (m) | 24751 | Mesured on maps |
| $d_{m}$ | Max number of settlers (ind. ${ }^{-1}$ ) | 10 | (Cuadros et al., 2018) |
| $S_{w}$ | Number of settlement waves | 3 | (Faillettaz et al., 2020) |
|  | Recruitment month | April (4) | (Lloret and Planes, 2003) |
| Mortality |  |  |  |


| $M_{i . C 0 p}$ | Natural of class $\mathrm{C}_{0}$ in ports areas (see table <br> 1) | 11.11263 | Expert estimation |
| :---: | :--- | :---: | :--- |
| $M_{i . C 0 n}$ | Natural mortality of class $\mathrm{C}_{0}$ in nursery <br> areas (see table 1) | 4.020845 | (Planes and Romans, 2004) |
| $M_{i . f}$ | Natural mortality of class $\mathrm{C}_{1}$ to $\mathrm{C}_{14}$ (see <br> table 1) | 0.3011051 | (Belharet et al., 2020) |

Annual mortalities for each group were therefore computed according to (equation 9) to both follow the literature data and fit our model (Table 2).

Table 1 : Summary table of the parameters governing the population dynamics sub-model in ISIS-Fish

Table 2: Mortality rates according to D. sargus life stages, calculation of natural mortality rates for ISIS-Fish

| Class | Residence <br> time | Survival | Formulae | Value in ISIS-Fish |
| :--- | :--- | :--- | :--- | :--- |
| $\mathrm{C}_{0}$ port <br> nursery | 5 month | $t_{\text {s. } p}=0,01$ <br> $t_{\text {s.f }}=0,74$ | $t_{\text {sa.Cop }}=\left(\sqrt[5]{t_{\text {s.p }} \times \sqrt[12]{t_{\text {s.f }}}}\right)^{12}$ | $M_{\text {i. } 00 p}=11.11263$ |
| $\mathrm{C}_{0}$ natural <br> nursery | 5 month | $t_{\text {s.n }}=0,192$ <br> $t_{\text {s.f }}=0,74$ | $t_{\text {sa.con }}=\left(\sqrt[5]{t_{\text {s.n }} \times \sqrt[12]{t_{\text {s.f }}}}\right)^{12}$ | $M_{\text {i. } 0 n}=4.020845$ |
| $\mathrm{C}_{1}-\mathrm{C}_{14}$ | 12 month | $t_{\text {s.f }}=0,74$ | $t_{\text {s.f }}$ | $M_{\text {i.f }}=0,3011051$ |

## S1.1.5. Abundance at equilibrium

The parameterisation of $D$. sargus population in ISIS-Fish is based on the parameter values of mortality, carrying capacity, number of benthic settlement waves and the length of natural nurseries (Table 2). Based on these parameters, ISIS-Fish was run to compute the equilibrium abundance per age class of $D$. sargus population. In this work, the assessment of the effectiveness of management measures involving rehabilitation or conservation methods is based on this equilibrium abundance.

## S1.1.6 Accessibility

The accessibility is a biological parameter accounting for biological behaviour (e.g. hidden behaviour) affecting $D$. sargus technical catch probability according to different fishing gear (usually called catchability, as in (Mahévas and Pelletier, 2004). Because of a lack of information in the literature, we calibrated this value with fishing parameters to reproduce the total annual catches.

## S1.2. Fishing activities

## Available data:

The fishing activities $D$. sargus populations were parameterised on the basis of information provided by Ifremer's Service d'Information Halieutique (SIH), corresponding to declarative data from commercial fisheries in 2019 and 2020 (Système d'Information Halieutique, 2022).

The declarative data from professional fisheries indicate an average annual catches of $D$. sargus in the fishing district of Toulon of 17.3 tons, spread over five fishing gears (Table 3). Recreational fishing catches of D. sargus represent $5 \%$ of total recreational catches in France (BVA, 2009). Thus, recreational fishers catches are estimated at 12.5 tons, according to an extrapolation of the data from Cadiou et al. (2009).

In addition to the value of total catches by gear, observations made during field inspections by agents of the Blue Coast Marine Park MPA (Parc Marin de la Côte Bleue) west of Marseilles, report that 19\% and $81 \%$ of $D$. sargus catches by professional and recreational fishing respectively were individuals smaller than 23 cm (Font and Lloret, 2014). However, current regulations prohibit the fishing of $D$.

Table 3: Gear implemented in D. sargus fishing, associated with their catches.

| Gear | Fishing activity | Mean annual catches 2019-2020 (kg.year ${ }^{-1}$ ) | Fishing effort (hour.year ${ }^{-1}$ ) |
| :---: | :---: | :---: | :---: |
| Long lining | Professional | 14280 | 757 |
| Trap | Professional | 25 | 4 |
| Gillnet | Professional | 780 | 494 |
| Trammel net | Professional | 1920 | 1060 |
| Gangui (Trawl) | Professional | 300 | 88 |
| Boat or shore line | Recreational | 6000 | 1216 |
| Spearfishing | Recreational | 6400 | 1216 |

sargus individuals below this size.

## Fishing parameters in ISIS-Fish:

ISIS-Fish describes the fishing activity using gears characterized by a gear and a fleet (i.e. group of vessels sharing same technical characteristics and gears). We divided fishing activities into two categories based on available knowledge and data: (1) professional fishing, composed with 5 gears (longline, gillnet, trammel net, gangui and trap) and grouped into 3 fleets, and (2) recreational fishing, composed of 2 gears (angling and spearfishing) and grouped into one fleet.

Each gear is characterized by (i) a standardisation factor that allows to compare fishing time between the different gears, and (ii) a selectivity curve of $D$. sargus depending on fish size that allows, in this case study, to distinguish a probability of capture below and above the minimum legal capture size of 23 cm .

For each fleet, the proportion of time dedicated to the practise of a metier is set for each month (called strategy in ISIS-fish) using SIH data from 2019 and 2020. We defined the annual fishing effort per gear as a function of the fishing profile and the average annual effort. The fishing profile per fleet was identical for every month of the year. Annual fishing time was broken down into 8-hour fishing days, with the number of days per month varying according to the season. Vessel activity increases between winter and summer and then decreases between summer and winter. The number of fishing days per year was defined such that the sum of the hours corresponded to the total effort per trade reported by the SIH data.

## S1.2.5 Calibration of the fishing parameters

The standardisation factor, the probability of selectivity for both professional and recreational fishing activity and the accessibility coefficient were estimated using a calibration procedure aiming at reproducing catches observations and expert knowledge described above (section available data and Table 4).

Calibration consisted in setting initial values for 12 unknown parameters (standardisation of each gear, selectivity for professional and recreational fishing and accessibility), running ISIS-Fish, comparing simulated catches to observations, and changing the initial values of the unknown parameters to reduce the discrepancy between observations and simulated values. To identify a good combination of parameters values, we select randomly 5000 combinations using a random LHS (Latin Hypercube sampling).

The initial value of standardisation factor for the different gears was calculated by fitting a generalised linear lognormal model to the annual catches per unit of effort associated with fishing gear and by years (2019 and 2020) (Lehuta et al., 2010). The coefficients generated by the model and attributed for each gear were then exponentially transformed (Table 4). For recreational fishing, the coefficients were estimated in relation to the values calculated for professional fishing.

The calibration process was led following the ODDO method (Mahévas et al., 2019), (see table 5). The objective function (OF) used to select the best combinations of values minimising the difference between observed catches and simulated catches (Table 4). This function computes the sum of the comparisons between catches of $D$. sargus under 23 cm from commercial and recreational fisheries from observations ( $D_{N M P_{-} o b s} \& D_{N M L_{-} o b s}$ ) and simulations ( $\left.D_{N M P_{-} s i m} \& D_{N M L_{-} s i m}\right)$ as well as total
observed ( $D_{P M m_{-} o b s}$ ) and simulated ( $D_{P M m_{-}} \operatorname{sim}$ ) catch by métiers, taken one by one. The aim is to minimise $O F$ to get closer to observations.

$$
O F=\left|D_{N M P_{-} \text {sim }}-D_{N M P_{-} o b s}\right|+\left|D_{N M L_{-} \text {sim }}-D_{N M L_{-} o b s}\right|+\sum_{m}\left|D_{P M m_{-} s i m}-D_{P M m_{-} o b s}\right|
$$

Three successive calibrations were carried out with a fishing parameter variation capacity of respectively $\pm 20 \%, \pm 2 \%$ and $\pm 1 \%$ around the established value in order to get closer to an optimal combination of parameter values. The values of the fishing parameters in the simulation minimising OF at the end of the last calibration process were selected and kept fixed for the rest of the study (Table 4).

Table 4: Initial and calibrated value of fishing parameters estimated in the model

|  | Gear | Standardization factor |  | Selectivity |  | Accessibility |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Initial value | Calibrated value | Initial value | Calibrated value | Initial <br> value | Calibrated value |
| Recreational fishing | Spearfishing Boat or Shore line | 1.3 1,22 | 1.265361803 1.185281584 | For Groups $\mathrm{C}_{1}$ to $\mathrm{C}_{4}$ : $S=0.8$ <br> For Groups $\mathrm{C}_{5}$ to $\mathrm{C}_{14}$ : $S=0.2$ | For Groups $\mathrm{C}_{1}$ to $\mathrm{C}_{4}$ : $S=0.770614002$ <br> For Groups C5 to $\mathrm{C}_{14}$ : $S=0.228198927$ |  |  |
| Professional fishing | Long lining <br> Trap <br> Gillnet <br> Trammel net <br> Gangui (Trawl) | 10,7 1.0 0,86 0,74 0,24 | 10.54983412 0.998235716 0.879663216 0.756877503 0.240777458 | For Groups $\mathrm{C}_{1}$ to $\mathrm{C}_{4}$ : $S=0.8$ <br> For Groups $\mathrm{C}_{5}$ to $\mathrm{C}_{14}$ : $S=0.2$ | For Groups $\mathrm{C}_{1}$ to $\mathrm{C}_{4}$ : $S=0.15697549$ <br> For Groups C5 to $\mathrm{C}_{14}$ : $S=0.824428461$ | $1 \times 10^{-4}$ | $9.8336 \times 10^{-5}$ |

Table 5 : ODDO filling with the ISIS-Fish model, fishing parameters calibration (see Mahévas et al., 2019)

| Model | Performance | Time per run | 45 sec |
| :---: | :---: | :---: | :---: |
|  |  | Parallelisation | 10 runs in parallel |
|  | Development | Language | Java |
|  |  | Implementation of the optimisation algorithm | Recoding of the algorithm in Java |
| Preprocessing | Problem formulation | Model | ISIS-fish model of D. sargus fisheries |
|  |  | Question | Estimating fishing parameters for D. sargus |
|  |  | Data | Two years catches in kg associated with gear and effort |
|  |  | Parameters bounds | Continuous values around 20\% from the initial value |
|  |  | Uncertainty (process and data) | Declarative and projected data for professional fisheries, estimation for recreational fisheries |
|  |  | Initial objective function | Absolute value of differences: $O F=\left\|D_{N M P_{-} s i m}-D_{N M P_{-} o b s}\right\|+\left\|D_{N M L_{-} s i m}-D_{N M L_{-} o b s}\right\|+\sum_{m}\left\|D_{P M m_{-} s i m}-D_{P M m_{-} o b s}\right\|$ |
|  | Objective Function | Building | Other OF were tested: <br> The standardised version: gives too much importance to the catches of gears that don't fish a lot of $D$. sargus and creates an imbalance. <br> The squared version: Create an imbalance with large catches over-represented |
|  |  | Final | Absolute value of differences: $O F=\left\|D_{N M P_{\_} s i m}-D_{N M P_{-} o b s}\right\|+\left\|D_{N M L_{\_} s i m}-D_{N M L_{-} o b s}\right\|+\sum_{m}\left\|D_{P M m_{-} s i m}-D_{P M m_{-} o b s}\right\|$ |
|  | Exploratory Analysis | data | No change |
|  |  | reduction dimension | No |
| Algorithm | Family | Random LHS around the parameter values seed fixed |  |
|  | Description Justification | The aim is to explore parameter values within a broad spectrum, as the values established initially are uncertain. |  |
|  | Changes in the algorithm | No |  |


|  | Settings | Test strategy: <br> Number year of simulation $=15$ <br> Number of simulation $=5000$ <br> Number of parameters tested 12 comprising 4 groups (2 of 2 parameters and 2 of 5 parameters) |
| :---: | :---: | :---: |
| Postprocessing | Convergence | Visual examination of FO, extraction of catches and parameters of the best simulation (minimising FO). Visual comparison with catch targets. Progressive targeting of parameters in 3 rounds of simulations: <br> First run : 20\% around the parameter values -> select parameters of the best simulation <br> Second run : $2 \%$ around the new parameter values -> select parameters of the best simulation <br> Third run : 1\% around the new parameter values -> select parameters of the best simulation, visual validation with observations and set of the values for the rest of the study |
|  | Optimum properties including identifiability | FO |
|  | Residual analysis | No |
|  | Multicriteria | No |
| Optimisation process | Number of simulation required | 15000 |
|  | Duration | 21h |
|  | Reached stopping criteria | Yes |
| Comments and failures | Initially, the aim was to carry out a single analysis to select the simulation that minimised FO. However, the $20 \%$ variation around the values of the 13 (continuous) parameters was too large to find an ideal value for the parameters with this single design. 3 analyses were therefore carried out, progressively targeting the optimised values. This targeting concerns a single combination of parameters. It is very likely that there are other combinations of parameters that meet our objectives. Ideally, the simplex or bootstrap method should be applied to select the best of these combinations. However, given the uncertainty in our fishing data, the combination selected at the end of this calibration analysis is more than sufficient for our purposes. |  |

## S2: Management scenarios

Each scenario was established by modifying the value of specific parameters from the reference simulation in relation to the measure tested by the scenario. The values of these new parameters are described in Table 6.

|  |  | Parameter | New Value |
| :--- | :--- | :--- | :---: |
| Scenario 1 | Recruitment | Coastal length of nursery area (m) | 61826 |
|  |  | Length of quay were juveniles fish settle (m) | 22276 |
| Scenario 2 | Recruitment | Coastal length of nursery area (m) | 115880 |
|  |  | Length of quay were juveniles fish settle (m) | 0 |
|  | Accessibility | $\mathrm{C}_{0}-\mathrm{C}_{4}$ | 0 |
|  |  | $\mathrm{C}_{5}-\mathrm{C}_{14}$ | $9.8336 \times 10^{-5}$ |
| Scenario 4 | Recruitment | Coastal length of nursery area (m) | 61826 |
|  |  | Length of quay were juveniles fish settle (m) | 22276 |
|  | Accessibility | $\mathrm{C}_{0}-\mathrm{C}_{4}$ | 0 |
|  |  | $9.8336 \mathrm{E}-5$ |  |

Table 6 : New parameter values introduced to test measures for the various scenarios

## S3: Uncertainty analysis

In the literature, there is no consensus on the values attributed to these parameters (Cuadros et al., 2018; Planes et al., 1998). To account for this uncertainty, 5000 simulations were carried out for each scenario in the model, using a random LHS (Latin Hypercube scaling) analysis. Each simulation therefore corresponds to a combination of values, sampled within the range of variation of the ecological parameter uncertain values mentioned above (Table 7). This gives 5000 equilibrium numbers of $D$. sargus for each scenario. All these equilibria were considered to constitute the population associated with each scenario. When the simulations were launched, the random LHS seed was kept fixed and identical for both the initial population and the scenarios. That way, each simulation has its identifiable counterpart in terms of the combination of ecological parameter values in the other scenarios. The comparison of the scenarios with each other consists of comparing all the 5000 simulations with their counterpart.

Table 7 : Interval of variation for the parameters identified as sensitive in the reference model, associated with their initial value and the range of variation applied during the sensitivity analysis.

|  | Parameters | Initial value | Variation range |
| :---: | :---: | :---: | :---: |
| ӨMortality | Class $\mathrm{C}_{0}$ in nurseries (5 month) | $80,80 \%$ | $[64,06 ; 96,9]$ |
|  | Class $\mathrm{C}_{0}$ in harbours (5 month) | $99 \%$ | $[79,2 ; 99]$ |
|  | Class $\mathrm{C}_{1}-\mathrm{C}_{14}(1$ year) | $26 \%$ | $[20,8 ; 31,2]$ |
| ӨRecruitment | linear nurseries $(\mathrm{m})$ | 55820 | $[44656 ; 66984]$ |
|  | Carrying capacity (individuals.m ${ }^{-1}$ ) | 10 | $[6 ; 13]$ |
|  | Number of settling waves | 3 | $[1 ; 4]$ |

## S4: Additional Figures



Figure 2 : Yearly gains in abundance (in \%) of adult D. sargus individuals. Boxplots present the median at their centre surrounded by the first and third quantiles. The outliers are represented by black dots.


Figure 3 : Bar plot showing the proportion (in \%) of catches by fishing gear in relation to total catches for each scenario and the mean gain per fishing type and per gear.

Differences in catches between gears can be explained by differences in catchability and effort:
$\mathrm{C}_{\mathrm{s}} \simeq \mathrm{q}(\mathrm{s}, \mathrm{g}) * \mathrm{E}(\mathrm{g}) * \mathrm{~N}_{\mathrm{s}}$
$Q(s, g)$ is the product of the parameters in table 4. Scenarios S1 and S2 affect $N_{s}$ while scenario S3 affects the parameter of selectivity of $q(s, g)$. This impact is different for the recreational and pro fleets but similar for the gears within each of these fleets. This assumption in the model is due to the fact that we did not have sufficient knowledge and data to parameterize differences. The only available data are catches (in weight) and the proportion of these catches under and over 23 cm for professional and recreational fisheries. The consequences of this assumption are that the impacts of the scenarios on catches within each fleet do not differ between gears.

## S5-References

© Shom-IGN, 2021. Limite terre-mer. http://dx.doi.org/10.17183/LIMTM
Andromède Océanologie, 2021. Donia Expert - Carthographie détaillée des habitats marins.
Belharet, M., Franco, A.D., Calò, A., Mari, L., Claudet, J., Casagrandi, R., Gatto, M., Lloret, J., Sève, C., Guidetti, P., Melià, P., 2020. Extending full protection inside existing marine protected areas, or reducing fishing effort outside, can reconcile conservation and fisheries goals. Journal of Applied Ecology 57, 1948-1957. https://doi.org/10.1111/1365-2664.13688
Bouchoucha, M., Darnaude, A.M., Gudefin, A., Neveu, R., Verdoit-Jarraya, M., Boissery, P., Lenfant, P., 2016. Potential use of marinas as nursery grounds by rocky fishes: insights from four Diplodus species in the Mediterranean. Marine Ecology Progress Series 547, 193-209. https://doi.org/10.3354/meps11641
BVA, 2009. Enquête relative à la pêche de loisir (récréative et sportive) en mer en Métropole et dans les DOM. Synthèse des résultats finaux 1-13.
Cadiou, G., Boudouresque, C.F., Bonhomme, P., Le Diréach, L., 2009. The management of artisanal fishing within the Marine Protected Area of the Port-Cros National Park (northwest

Mediterranean Sea): a success story? ICES Journal of Marine Science 66, 41-49. https://doi.org/10.1093/icesjms/fsn188
Cheminée, A., Francour, P., Harmelin-Vivien, M., 2011. Assessment of Diplodus spp. (Sparidae) nursery grounds along the rocky shore of Marseilles (France, NW Mediterranean). Scientia Marina 75, 181-188. https://doi.org/10.3989/scimar.2011.75n1181
Cheminée, A., Le Direach, L., Rouanet, E., Astruch, P., Goujard, A., Blanfuné, A., Bonhomme, D., Chassaing, L., Jouvenel, J.-Y., Ruitton, S., Thibaut, T., Harmelin-Vivien, M., 2021. All shallow coastal habitats matter as nurseries for Mediterranean juvenile fish. Sci Rep 11, 14631. https://doi.org/10.1038/s41598-021-93557-2.
Cuadros, A., Basterretxea, G., Cardona, L., Cheminée, A., Hidalgo, M., Moranta, J., 2018. Settlement and post-settlement survival rates of the white seabream (Diplodus sargus) in the western Mediterranean Sea. PLOS ONE 13, e0190278. https://doi.org/10.1371/journal.pone. 0190278
Faillettaz, R., Voué, R., Crec'hriou, R., Garsi, L.H., Lecaillon, G., Agostini, S., Lenfant, P., Irisson, J.O., 2020. Spatio-temporal patterns of larval fish settlement in the northwestern Mediterranean Sea. Marine Ecology Progress Series 650, 153-173. https://doi.org/10.3354/meps13191
Font, T., Lloret, J., 2014. Biological and Ecological Impacts Derived from Recreational Fishing in Mediterranean Coastal Areas. Reviews in Fisheries Science \& Aquaculture 22, 73-85. https://doi.org/10.1080/10641262.2013.823907
Froese, R., Pauly, D., 2022. FishBase [WWW Document]. World Wide Web electronic publication. URL www.fishbase.org (accessed 2.6.23).
García-Rubies, A., Macpherson, E., 1995. Substrate use and temporal pattern of recruitment in juvenile fishes of the Mediterranean littoral. Marine Biology 124, 35-42.
Gonçalves, J.M.S., Erzini, K., 2000. The reproductive biology of the two-banded sea bream (Diplodus vulgaris) from the southwest coast of Portugal. Journal of Applied Ichthyology 16, 110-116.
Harmelin-Vivien, M.L., Harmelin, J.G., Leboulleux, V., 1995. Microhabitat requirements for settlement of juvenile sparid fishes on Mediterranean rocky shores. Hydrobiologia 300, 309-320. https://doi.org/10.1007/BF00024471
Hussein, C., Verdoit-Jarraya, M., Pastor, J., Ibrahim, A., Saragoni, G., Pelletier, D., Mahévas, S., Lenfant, P., 2011. Assessing the impact of artisanal and recreational fishing and protection on a white seabream (Diplodus sargus sargus) population in the north-western Mediterranean Sea using a simulation model. Part 1: Parameterization and simulations. Fisheries Research 108, 163-173.
Ifremer, 2021. MARS3D - Vagues.
IGN, BRGM, 2022. Photographies aériennes.
Le Cren, E.D., 1951. The Length-Weight Relationship and Seasonal Cycle in Gonad Weight and Condition in the Perch (Perca fluviatilis). Journal of Animal Ecology 20, 201-219.
Lehuta, S., Mahévas, S., Petitgas, P., Pelletier, D., 2010. Combining sensitivity and uncertainty analysis to evaluate the impact of management measures with ISIS-Fish: marine protected areas for the Bay of Biscay anchovy (Engraulis encrasicolus) fishery. ICES Journal of Marine Science 67, 1063-1075. https://doi.org/10.1093/icesjms/fsq002
Lloret, J., Planes, S., 2003. Condition, feeding and reproductive potential of white seabream Diplodus sargus as indicators of habitat quality and the effect of reserve protection in the northwestern Mediterranean. Marine Ecology Progress Series 248, 197-208. https://doi.org/10.3354/meps248197
Mahévas, S., Pelletier, D., 2004. ISIS-Fish, a generic and spatially explicit simulation tool for evaluating the impact of management measures on fisheries dynamics. Ecological Modelling 171, 6584.

Mahévas, S., Picheny, V., Lambert, P., Dumoulin, N., Rouan, L., Soulié, J.-C., Brockhoff, D., Lehuta, S., Riche, R.L., Faivre, R., Drouineau, H., 2019. A Practical Guide for Conducting Calibration and Decision-Making Optimisation with Complex Ecological Models. https://doi.org/10.20944/preprints201912.0249.v1

Mouine, N., Francour, P., Ktari, M.H., Chakroun-Marzouk, N., 2012. Reproductive biology of four Diplodus species Diplodus vulgaris, D. annularis, D. sargus sargus and D. puntazzo (Sparidae) in the Gulf of Tunis (central Mediterranean). Marine Biological Association of the United Kingdom. Journal of the Marine Biological Association of the United Kingdom 92, 623-631. http://dx.doi.org/10.1017/S0025315411000798
Mouine, N., Francour, P., Ktari, M.H., Chakroun-Marzouk, N., 2007. The reproductive biology of Diplodus sargus sargus in the Gulf of Tunis (central Mediterranean). Scientia Marina 71, 461469.

Planes, S., Jouvenel, J.-Y., Lenfant, P., 1998. Density dependence in post-recruitment processes of juvenile sparids in the littoral of the Mediterranean sea. Oikos 83, 293-300. https://doi.org/10.2307/3546840
Planes, S., Romans, P., 2004. Evidence of genetic selection for growth in new recruits of a marine fish. Molecular Ecology 13, 2049-2060.
Système d'Information Halieutique, 2022. Données de production et d'effort de pêche SACROIS (2012 à 2020). Ifremer SIH. https://doi.org/10.12770/8eac4cd1-a546-445c-b3fa7ed580333403
Ventura, D., Giovanna, J.L., Giandomenico, A., 2014. Temporal partitioning of microhabitat use among four juvenile fish species of the genus Diplodus (Pisces: Perciformes, Sparidae). Marine Ecology 37, 1013-1032.
Vigliola, L., Harmelin-Vivien, M., Biagi, F., Galzin, R., Garcia-Rubies, A., Harmelin, J., Jouvenel, J., Le Direach-Boursier, L., Macpherson, E., Tunesi, L., 1998. Spatial and temporal patterns of settlement among sparid fishes of the genus Diplodus in the northwestern Mediterranean. Mar. Ecol. Prog. Ser. 168, 45-56.

