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1 Integration of animal movement into vehicle-wildlife collision models

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10 Abstract

11

12 Road networks have major ecological impacts on living organisms consequent to habitat loss and fragmentation, chemical and acoustic pollution, and direct mortality when wildlife-vehicle 13 collisions are involved (WVC). The many past empirical studies revealed major variables 14 15 accounting for WVC incidence (e.g., population density). Similarly, spatial locations of WVC hotspots are associated to landscape features at large spatial scales, and to road characteristics at small 16 17 spatial scale. Yet, we currently lack a comprehensive theoretical framework for WVC. Animal 18 movement in relation to habitats is likely an essential driver of encounters with roads, but this 19 remains largely ignored in most studies. Movement ecology now provides the necessary tools to 20 investigate the impact of animal movement on WVC. We built a general individual-based model 21 incorporating recent knowledge in movement ecology (movement typology: roaming, migratory 22 route crossing a road, active attraction and active repulsion of roads) to estimate WVC risks. We explored the relative effects of animal and vehicle movement parameters (speed, abundance, road 23 24 sinuosity and animal movement pattern) on collision probability. We show that animal behaviour toward roads has major impacts on the number and risks of WVC, but also modulate the effects of 25 other factors (animal traveling speed, species local abundance, road traffic volume) on WVC. 26 27 Sensitivity analyses show that the movement and behaviour of the animal has more influence on WVC risks than any of the characteristics of roads and vehicles we tested. Our results suggest that 28 29 empirical studies of WVC should incorporate knowledge about the behavioural habits of the focal 30 species in relation to roads.

31 **1. Introduction**

The last century has witnessed a continuous urbanization of almost all ecosystems, with profound 32 influences on living organisms (Corlett, 2015). A particularly important anthropogenic factor 33 34 generating major ecosystem modification and disturbance is the increasing densification and 35 encroachment of the earth-bound transportation systems, and particularly of the road network (Davenport & Davenport, 2006). The presence of roads modify the local and large-scale 36 37 environment of species as a consequence of chemical and noise pollution, changes in vegetation 38 diversity, habitat loss and fragmentation (see Forman & Alexander, 1998; Coffin, 2007, for 39 reviews). Roads can also have profound effects on ecosystem functioning and the biology of living 40 organisms (Holderegger & Di Giulio, 2010). For instance, road mortality is a key driver of 41 biodiversity loss in amphibians (Hels & Buchwald, 2001), birds and mammals (Grilo et al., 2020) 42 and has been shown to affect animal behaviour (Andrews & Gibbons, 2005; Legagneux & Ducatez, 43 2013). In the context of the ongoing biodiversity crisis (Ceballos & Ehrlich, 2002), knowledge 44 about the mechanisms that lead to road mortality in wildlife is therefore critical.

45 For decades, empirical studies have been accumulating, reporting on collision hot-spots (Teixeira et 46 al., 2013; Silveira Miranda et al., 2020), animal foraging habits near roadsides (Carvalho-Roel et 47 al., 2019), or the seasonal variation in the number of reported cases (Medinas et al., 2021). We now 48 have a clearer picture of what factors mainly account for collision risks (Table 1). Species with life 49 history traits such as extended home ranges or slow travel speeds, and species occurring at high 50 abundances, are more often subjected to roadkill. Road characteristics such as high traffic volume 51 and landscape features (e.g. habitats promoting high levels of animal activity and mobility) account 52 for WVC hot-spots locations at small spatial scales, but collision distribution can also be predicted at large spatial scales from variables such as road density, hunting bags (Saint-Andrieux et al., 53 54 2020) or landscape connectivity (Girardet et al., 2015). However, the relative effects of risk factors 55 are difficult to disentangle. For example, as traffic mortality depletes local populations, roadkill 56 numbers will decrease and simply counting collisions in high-risk areas could lead to an 57 underestimation of the importance of animal abundance as a key determining factor of roadkill hot-58 spots (Teixeira et al., 2017). So far few attempts at building a theoretical model for wildlife-vehicle 59 collisions (WVC) have been made, and our current knowledge on the quantitative effects of traffic 60 on encounter and mortality risks for wildlife remains essentially descriptive, with little predictive abilities. 61

The very first model dealing with collision probabilities is the ideal gas model, first coined by physicist J. C. Maxwell in the context of molecule collisions in a gas (Maxwell, 1860). The ideal gas model posits that encounter rate within a defined volume or plane is a function of molecule 65 speed, density and collision zone (Hutchinson & Waser, 2007). Despite its heuristic value in ecology, to estimate encounter rates between males and females of a given species for example 66 (Harcourt & Greenberg, 2001), it has seldom been used to study WVC as the general application in 67 68 this context is not straightforward. For example, in most wildlife-vehicle encounters, the animal is 69 likely to die and will no longer be at risk of collision again, which is not reflected in colliding 70 molecules. Others proposed WVC-specific mechanistic models: Hels & Buchwald (2001) modelled 71 the probability of death for an amphibian crossing a road (see Gibbs & Shriver, 2002; van 72 Langevelde & Jaarsma, 2005, for an application to other species), allowing the integration of road 73 mortality into population persistence models (Gibbs & Shriver, 2002).

74 A shortcoming shared by all current WVC models is the lack of consideration for animal movement 75 and habitat use behaviour, often limiting when one aims at describing the complexity of animal 76 movement in a heterogeneous landscape (Siniff & Jessen, 1969). For example, not only do animals 77 cross roads, but they can roam along roadsides to forage, hence putting them at much higher risk of 78 being hit by a car (scavenging: Schwartz et al. 2018; Ratton et al. 2014, hunting: Gomes et al. 2009; 79 D'Amico et al. 2013, salt licking: Fraser & Thomas 1982). More importantly, collision probabilities 80 in a landscape are intrinsically linked to the number of encounters an animal has with roads, and 81 peaks in activity. Mate searching, breeding, and juvenile dispersal have repeatedly been cited to explain seasonal patterns in WVC numbers (Steiner et al., 2014; Ryu & Kim, 2021; Raymond et al., 82 83 2021). Despite its expected effects, the influence of patterns of space use by animals on WVC has 84 yet to be adequately addressed in mechanistic models.

85 Movement patterns of animals are now well described within the frameworks of movement ecology (Demšar et al., 2015). The advances of the last decades in bio-logging (e.g., animal-borne GPS 86 87 collars, tags or transponders) have made it increasingly easy to acquire position data characterizing 88 these individual trajectories for a growing number of species, therefore providing a solid empirical 89 basis to develop mathematical models of animal movement. Large-scale patterns emerge from 90 movement at the individual level: for example, oriented movement, based on environmental cues or 91 memory, is the basis for migratory routes and sedentary ranges (Mueller & Fagan, 2008; Van 92 Moorter et al., 2009), which in turn affect WVC occurrences (Bonnet et al., 1999; Kantola et al., 93 2019). To our knowledge however, WVC studies seldom incorporate theoretical advances of 94 movement ecology, although space use is the main driver of animal encounters with roads. Therefore, we argue that modelling wildlife-vehicle collisions from a movement ecology 95 96 perspective would improve our ability to understand the main factors driving WVC (Hels & 97 Buchwald, 2001; van Langevelde & Jaarsma, 2005).

We here propose an individual-based model to simulate animal trajectories using biased and unbiased correlated random walks, as a mean of exploring the effects of road characteristics. species traits and patterns of space use on the number and probability of collisions between animals and vehicles. Our two main goals are i) to quantify the relative effects of several road characteristics (road sinuosity, vehicle speed and traffic volume) and species traits (animal speed, local abundance, movement patterns, foraging and dispersal behaviours) on WVC occurrences and ii) for a set of biologically realistic scenarios of animal movement, to identify what factors have the largest impact on wildlife-vehicle collision risks. Understanding which of the biological, human and environmental factors modulate the risks of wildlife-vehicle collisions the most is key to the selection and implementation of efficient mitigation policies such as fences, over or underpasses, and traffic signage that raises drivers' attention (Davenport & Davenport, 2006). Our model should help to build qualitative predictions either for inter-specific studies or WVC mitigation.

Table 1: Overview of the different factors correlated to the incidence of wildlife-vehicle collisionsin the current literature.

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125 **2. Material & Methods**

Following previous work on simulations for the ideal gas model (Hutchinson & Waser, 2007), we 126 127 constructed a discrete-time simulation of particles movement in which all trajectories were approximated by a series of straight-line steps of constant duration. Two classes of particles were 128 129 considered: vehicles and animals, each with their own speed, movement patterns and density. We 130 conducted preliminary testing of the model by replicating the conditions under which the ideal gas 131 model applies, and subsequently ensured that simulation results matched the model predictions. 132 From there, we used biased and unbiased correlated random walk models to represent animal 133 movement. Random walk parameter values were chosen to match real-life scenarios. The goal of 134 the simulations was to keep track of the number of encounters between vehicles and animals per 135 unit of time (fig. 1). We set a duration of 12 hours per simulation run, as a broad representation of 136 the amount of time during which most diurnal and nocturnal terrestrial species are active (Refinetti, 137 1999). Two types of experiments were conducted: (1) to test the quantitative and qualitative effects 138 of the different focal parameters (animal and vehicle speed, movement and densities) independently, 139 we replicated the simulations for different values of the parameter under study, holding everything 140 else constant; (2) to evaluate the relative importance of each focal parameters on the number of 141 WVC, we conducted one-at-a-time sensitivity analyses.

Basic simulation layout

We first assumed two separate entities, thereafter designated 'vehicles' and 'animals', moving on a 143 $30x30km^2$ plane mapped by x and y axes. We treated the plane like a torus, meaning that an animal 144 145 or a vehicle crossing the edge reappeared at the same location, but on the opposite edge of the plane 146 (see fig. 2). The vehicle movement followed a road defined as $0 \le x \le 30$ and $y = \cos(\alpha x)$, where α 147 controlled road sinuosity. We computed an index of sinuosity for each simulated road as the ratio of 148 the road's length to 30 (i.e. the x axis' length). A sinuosity index of 1 indicated a straight road, and 149 the index increased as the road became more and more sinuous. The number of vehicles on the road 150 at any given time followed a Poisson distribution (Breiman, 1963) and all vehicles moved at 151 constant speed. For animal movement, we used biased correlated random walks (BCRW; Turchin, 152 1998), a combination of correlated random walks where animals have a directional persistence and 153 therefore avoid backtracking, and of biased random walks, where animals are attracted to one or 154 several centers of attraction spatially defined on the simulation plane. BCRW are highly flexible 155 models proven to fit animal trajectories at different spatial scales for a range of terrestrial species in 156 heterogeneous landscapes: individuals can be biased toward a home range (Red deer Cervus 157 elaphus, Berthelot et al., 2020), or resources such as larval hosts plants (Butterfly Plebejus 158 icarioides fender, Bailey et al., 2018) and food (Elk Cervus canadensis, Fortin et al., 2005).

160 Animals moved on the plane from time steps t = 0 to T according to a BCRW:

 $x_a(t) = x_a(t-1) + \cos(\theta_t) \cdot d$ $y_a(t) = y_a(t-1) + \sin(\theta_t) \cdot d$ $x_a(0) \sim U(0,30)$ $y_a(0) \sim U(0,30)$

with (x_a, y_a) the coordinates of the animal on the simulation plane, step length d drawn from a 161 zero-truncated normal distribution $N(\mu, 0.5)$, μ the mean animal speed (km.h⁻¹), and U(0, 30) the 162 uniform distribution between 0 and 30. We computed the direction of the walk $0 \le \theta_t \le 2\pi$ as the 163 mean between the *correlated* direction $\theta_t^c \sim N(\theta_{t-1}^c, \sigma^2 = 5)$ with θ_0^c randomly drawn between 0 164 and 2π , and the *biased* direction θ^b , the direction of the nearest attraction center. It resulted that the 165 direction of the walk was the outcome of two competing processes: the persistence, the animal's 166 internal propensity to walk without backtracking in a given direction θ_0^c , and the *attraction* toward 167 the nearest attraction center. The relative importance of the *attraction* with respect to the *persistence* 168 169 was defined by the associated weight of attraction w (Duchesne et al., 2015). Accordingly, w 170 controlled the strength with which animals are attracted to the attraction center: w = 0 produces 171 unbiased correlated random walks (no attraction), and $w \rightarrow \infty$ tends towards biased random walks. 172 In practice, for high values of w the animal made a beeline towards the nearest center of attraction because the attraction to the nearest center takes over the persistence of the correlated walk in 173 174 orienting the walk. Once the animal reached the attraction center, it orbited around it, as a result of 175 the *persistence* pulling away in its own direction while the attraction component brought the animal 176 back towards the center. In consequence, high values of w would imply less spread-out orbiting 177 behaviours, mimicking animals staying in proximity of the attraction center (see fig. S1 in 178 supplementary material).

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times to produce a mean collision number and estimate an individual probability of collision. The
simulation used R package *Rcpp* (Eddelbuettel & Balamuta, 2018).

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190 **2.2 Behavioural scenarios of animal movement and space use**

We considered 5 different scenarios designed to mimic real-life animal behaviours described in theliterature (fig. 2).

193 2.2.1 Scenario 0: no influence of the road

In scenario 0, the weight of attraction w was set at 0 resulting in animals performing (unbiased) 194 195 correlated random walks (CRW). In CRW, successive steps directions are correlated, resulting in a 196 directional bias: the animal walked forward and rarely backtracked. CRW have been used to 197 describe the movement of several species (Mouse Apodemus sylvaticus: Benhamou 1991, Reindeer 198 Rangifer tarandus: Marell, Ball & Hofgaard 2002, Red fox Vulpes vulpes, Snowshoe hare Lepus 199 americanus: Siniff & Jessen 1969). CRW imply that animals did not factor in the presence of a road 200 or moving vehicles in their movement, which could be especially prevalent on smaller or unpaved 201 roads (Brehme et al., 2013). This simple movement model also served as a null model, *i.e.* a 202 scenario without attractive or repellent effects of road presence on animal movement.

203 2.2.2 Scenario 1: habitat fragmentation

All animals started the simulation run in the top-left corner of the simulation plane, always above the road, while the single attraction center (w = 1.5) was randomly located below the road. The resulting BCRW simulated a migration of individuals between patches during which they had no choice but to cross the road. This scenario could match many migrating species such as amphibians converging towards open waters for reproduction (Hels & Buchwald, 2001; Sillero, 2008), large mammals migrating to their summer or winter ranges (Avgar et al., 2014) or crab species moving back to sea during breeding migrations (Ryu & Kim, 2020).

211 2.2.3 Scenarios 2 & 3: attraction to the road

Scenario 2 emerged from the frequently reported attraction to roads in many species: salt licking on roadsides by large mammalian herbivores (Fraser & Thomas, 1982), use of roads and roadsides for foraging or hunting (Diurnal raptors: Meunier et al., 2000; Passerines: D'Amico et al., 2013) or carnivores scavenging on carrion from animals hit by vehicles (Schwartz et al., 2018; Ratton et al., 2014). Arguably, this scenario could encompass the case of reptiles that bask on warm road surfaces (Meek, 2009). Attraction centers (w = 3) were placed every 300m on the road, such that animals converged towards the closest attraction center. Once an animal had reached an attraction center, it would orbit around until another attraction center in the vicinity became the new closest center, at which point the animal would leave its current attraction center to walk towards the new one. Little by little, an animal would visit several attraction centers during a simulation run, and would often walk on or in close proximity to the road while moving in-between attraction centers.

223 Scenario 3 is a slightly modified version of scenario 2, intended for animals with flying abilities 224 (birds, bats...) and using roads as corridors for long-range displacements; or for hunting because 225 roads "drive" preys in the landscape (Kerth & Melber, 2009; Schwartz et al., 2018). Flying animals 226 often make low-altitude flights to catch a prey or land on the road to feed (Grilo et al., 2014; see also the 'leap and strike' hunting behaviour of owls: Norberg & Norberg, 1970; Southern, 1954). To 227 228 this end, we added a third dimension z_a describing the altitude of the animals to their existing (x_a, y_a) 229 position on the plane. The altitude was a function of the distance to the nearest attraction center: $z_a(t) = \exp(0.3 * ||d||) - 1$, with ||d|| the Euclidean distance between the animal and the closest 230 231 attraction center. This distribution of altitudes entailed that animals got progressively closer to the 232 ground as they approached the attraction center. In consequence, they were only susceptible to 233 collision when within a radius of 5m around a center of attraction.

234 2.2.4 Scenario 4: road avoidance

The last scenario accommodated animal movement for road avoidance behaviours. Road avoidance has been reported in several species (Grizzly bear *Ursus arctos*: Mace et al., 1996; Wild boar *Sus scrofa* and Red deer *Cervus elaphus*: D'Amico et al., 2016), possibly resulting from chemical, noise pollution, or road surface avoidance (Andrews & Gibbons, 2005; Schaub et al., 2008; Jaeger et al., 2005). Building upon scenario 2, we only modified the attraction angle to be computed as θ^b + 180°, meaning that the animal moved away from the center instead of towards it.

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244 **2.3 Qualitative predictions of the model**

For each of the scenarios described above, we derived qualitative predictions for 6 parameters of interest, by running the simulation at different values of this parameter while holding everything else constant. The parameters of interest were: animal mean speed, vehicle speed, animal density, traffic volume, weight of attraction centers w (scenarios 1 through 4 only) and road sinuosity (Table 2). 250 Mean animal travel speed values covered the observed speed range of most terrestrial species (from 251 slowest to fastest, migrating amphibians: Hels & Buchwald 2001, foraging badger Meles meles: San 252 et al. 2007, migrating ungulates: Berger et al. 2006; Singh & Ericsson 2014). Considering that this 253 range is wide, we extracted mean travel speed estimations for several species from the literature (Rowcliffe et al., 2016; Farley et al., 1993) and set our baseline at the mean value of 5 km.h⁻¹. 254 Vehicle speed encompassed slow vehicles up to high-speed trains, and the baseline was set at the 255 current speed limit on main roads in France (80 km.h⁻¹). Local animal densities, traffic volume 256 ranges and baselines were mostly dictated by computational limitations, although the baseline of 257 250 vehicles.hour⁻¹ is representative of average sized roads in France (French Ministry of 258 259 Ecological Transition, 2022). Finally, road sinuosity was investigated following observed 260 sinuosities which are publicly available for the road network of Ireland (Transport Infrastructure 261 Ireland, 2022).

The weight of attraction w in BCRW scenarios was a dimensionless unit ranging from 1 (attraction to centers is equal to the correlated direction of the walk, spread-out orbiting behaviour around attraction center) to 20 (when attraction was 20 times more important in walk orientation, animals walked in nearly straight paths toward the attraction center and stayed mostly in place once it was reached). Baseline values for w were scenario-dependent as w = 5 for scenario 1 ensured that over 90% of animals cross the road in less than 12 hours, and w = 3 for scenarios 2 and 3 ensured that animals would effectively visit several attraction centers during the simulation run.

Two types of results were extracted from simulation runs: the mean number of collisions recorded over the course of a 12 hours run, and the individual probability of collision for animals. Considering that new individuals replaced dead animals, not all animals were present from the start of the simulation run and exposure to vehicles was not constant across all animals. Consequently, we estimated the individual probability of collision from a logistic regression (0: alive, 1: hit by a vehicle) as a function of the parameter under study, and where the time of arrival in the simulation was entered as an offset predictive variable (Agresti & Coull, 2002).

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Table 2: Model parameters of interest, relating to animal and vehicle movement and that appear to be pivotal factors driving collision numbers according to most studies (Table 1). While the effects of one parameter on collisions are being explored, all other parameters assume their baseline values. To produce quantitative and qualitative predictions, we explored a range of values contained within biologically relevant intervals for each parameter.

Parameter of interest	Baseline value	Explored range
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ANIMAL	Mean speed (km.h ⁻¹)	5	0.1 - 30
	Density (ind.km ⁻²)	0.016	0.001 - 0.03
	W	5 (scenario 1) ; 3 (scenarios 2-4)	1 - 20
VEHICLE	Speed (km.h ⁻¹)	80	30 - 300
	Traffic volume (vehicle.h ⁻¹)	30	10 - 250
	Road sinuosity index	1 (straight road)	1 - 3.5

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279 **2.4 Sensitivity analyses of collision number**

280 Sensitivity analyses were conducted on each scenario, following a one-at-a-time approach, meaning 281 that we measured the variation in collision numbers while keeping all input factors fixed except the 282 one that is being perturbed (Saltelli et al., 2019). In practice, we computed the mean number of 283 collisions y_{xl} for factor of interest x at a reference value x_l , and after 10% increase in value (denoted 284 x_2). To explore possible non-linearity, we repeated this for several reference values x_1 within the 285 relevant range of x (see section 2.3 for parameter ranges). We then defined the sensitivity measure 286 as $OR_{x1} = y_{x2} / y_{x1}$, *i.e.*, the odd ratio of the collision probabilities that results from a 10% increase of parameter x at reference value x_1 . For example, an OR_{x1} of 1 indicated no change in the odds of 287 288 collisions between parameter values x_1 and $x_1+10\%$, while $OR_{x1} > 1$ indicated greater odds of collision after a 10% increase in parameter x. The overall strength of the influence of parameter x on 289 290 the collision number in its relevant range was defined by how far the mean of the ratios (OR_{mean}) for 291 all tested values of x deviated from 1 (e.g., $OR_{mean} = 1$ indicated no overall influence of factor x on 292 collision numbers).

293

3. Results

295 **3.1 Qualitative model predictions**

296 Applying baseline values to all parameters of interest, mean collision rate ranged from 0.16 to 17.2 297 animal-vehicle collisions over 12 hours (individual probabilities of collision for animals: 0.4% -298 45.6%). As expected, the movement of animals had major effects on the probability for an animal to 299 get hit by a vehicle (fig. 3). With baseline parameter values, animals attracted to the road (scenarios 300 2 and 3) were hit more often than in any other scenario, with an increase in the odds of collision for 301 terrestrial animals (scenario 2) compared to flying animals such as birds and bats (scenario 3). 302 Compared to the null model (scenario 0, roaming with no attraction/avoidance of the road), 303 migrating animals crossing the road exactly once (scenario 1) were hit more often, while the odds of 304 collisions for active road avoidance (scenario 4) were reduced. This ranking between scenarios was305 not always maintained for other values of the tested parameters.

Overall, response functions of parameters had similar shapes whether the outcome was measured as the mean number of collisions or as the individual probability of collision, except for animal density. Vehicle speed had no effect on collision numbers and probabilities in any of the scenarios (fig. 3b and 4b). Animal density was positively and linearly associated with collision numbers, which translated into a constant individual probability of collision no matter the local abundance in animals (fig. 3d and 4d). The influence of the remaining parameters (animal speed, weight of attraction, traffic volume and road sinuosity) was non-linear and/or scenario-dependent.

313 Traveling faster reduced collision numbers and probabilities for migrating animals (scenario 1) but 314 had the opposite effect on roaming and road-avoiding animals (scenarios 0 and 4). In scenarios with 315 road-attracted animals, collision occurrences increased with mean animal speed up to moderate 316 speed values, after which the relation was reversed (fig. 3A and 4a). Higher volumes of traffic on the road increased collision numbers and probabilities, although scenarios 0, 2, 3 and 4 presented 317 318 inflection points after which the increase in collision occurrences slowed down, which was not 319 observed within the explored range of traffic volume for scenario 1 (fig. 3C and 4c). Similarly, 320 increasingly sinuous roads had moderate to low positive effects on collision number and probability 321 in all but one scenario: scenario 3 yielded no apparent effect of road sinuosity (fig. 3f and 4f). 322 Finally, increasing weights of attraction (or repulsion for scenario 4) led to more collisions (and higher individual probabilities) in scenarios 2 and 3 with inflection points around w = 5, but 323 324 increasing w values were associated to less collisions and lower probabilities of WVC in scenarios 1 325 and 4 (fig. 3e and 4e).

326

327 **3.2 Sensitivity analyses**

328 For roaming animals (scenario 0), the most influential parameter was the animal density (OR_{mean}= 329 1.091) followed by animal speed ($OR_{mean} = 1.032$). In all other scenarios (scenarios 1, 2, 3 and 4), 330 the weight of attraction w had the largest influence (respectively $OR_{mean} = 0.888$, 2.019, 1.155 and 331 0.843) followed by animal density (respect. OR_{mean}= 1.098, 1.095, 1.101 and 1.108). The least 332 influential parameters were traffic volume for scenario 0 (OR_{mean}= 0.996), vehicle speed for 333 scenarios 1 and 3 (respect. OR_{mean}= 1.01, 1.006) and road sinuosity for scenarios 2 and 4 (respect. 334 OR_{mean}= 0.998, 0.993). Using OR_{mean} is, however, a simplified view of the non-linear effect of 335 parameters, as it was sometimes pulled in one direction by one or a few extreme OR values. For 336 example, some parameters such as the weight of attraction/repulsion generally had a substantial

- influence on collision numbers at lower values (w = 1) while higher w values yielded OR values
- 338 closer to 1 (fig. 5).





344 **4. Discussion**

Road mortality of wildlife is a pervasive and widespread cause of biodiversity loss across the world that should be seriously considered in wildlife management and conservation programs. In spite of its ecological importance and relevance, we are still lacking a strong formalization of WVC to grasp what are its main biological, ecological and anthropogenic determinants. By combining earlier encounter models with the current knowledge about movement ecology, our work should help formulating qualitative predictions for reducing the number of WVC or predict which species should be at greater risks of being hit by vehicles.

Our simulations show how collision numbers and probabilities differ markedly with the movement and behaviour of animals, such as migration or home-ranging, or with their locomotion (terrestrial *vs.* flight). We suggest that this contrast in mortality risk between individuals with different movement types could lead to contrasting selection forces on animal habitat use with regard to road avoidance or attraction. Knowledge about animal movement of species should be an essential part of WVC studies as it seems to be the most influential variable on WVC occurrence in the absence of reaction from vehicle drivers and animals.

359

360 **4.1 Effects of animal movement on WVC risks**

WVC broadly results from two competing mechanisms: the number of times an individual will 361 362 cross or stay close to a road, and the time spent at risk of collision when on the road while crossing, 363 foraging or basking. Both mechanisms are modulated by patterns of space use (in our simulation, 364 the different scenarios of movement behaviour), and their modulations through the weight of 365 attraction/repulsion w and the speed at which animals travel. Obviously, individuals actively 366 avoiding roads (scenario 4) should be the least at risk of collision with a vehicle and, conversely, 367 terrestrial individuals attracted to the road for foraging or basking (scenario 2) should be the most at 368 risk. This is in line, for example, with Western green lizards (Lacerta bilineata) being road-killed in 369 higher proportions than the other sympatric lizard species that do not bask on roads (Meek, 2009). 370 Flying carrion feeders (mostly diurnal and nocturnal raptors) are less likely to be killed than their 371 terrestrial counterparts, a difference that may be further widened by some species' cognition and 372 ability to successfully evade oncoming vehicles, as is the case with corvids (Mukherjee et al., 2013; 373 Lima et al., 2015). Nevertheless, we show here that other parameters of both animal and vehicle 374 movement can modulate these conclusions. For example, slow moving animals (less than 15km.h⁻¹) 375 were more often hit when migrating (scenario 1) than roaming freely (scenario 0), but the reverse 376 was true for faster animals (fig. 3a).

377 By mimicking the presence of attractive or repulsive areas for animals in the landscape, the 378 attraction/repulsion weight (w) ultimately controls the time spent in the vicinity of the road, which 379 is positively correlated to collision numbers and probabilities. The relationship between w and the 380 time spent on the road is easily interpreted for scenarios of road attraction (scenarios 2 and 3, w is 381 positively correlated to time on road surface) and avoidance (scenario 4, w value is negatively 382 correlated to time on road surface). When animals travel from a departure point to an attraction 383 center (scenario 1), the time spent crossing is negatively correlated to w: high w values means 384 straight trajectories between departure and attraction center, and thus less time on the road.

385 Collision risks in animals roaming freely in a landscape (scenario 0, no w variable) are dictated by 386 their speed: the faster the animal, the more it explores its environment during a simulation run and 387 thus encounter roads more often, even though each crossing becomes concurrently less dangerous 388 as fast animals cross the road in less time (see fig. S2 in supplementary material). For the same 389 reason, when animals cross the road only once (scenario 1), we would also expect the animal travel 390 speed to be a key variable in decreasing collision numbers. However, this was not reflected by the sensitivity analyses, the most likely explanation being that the mean sensitivity OR_{mean} for travel 391 392 speed in scenario 1 is pulled up by the odds ratio for very low travel speeds (odds ratio $OR_{0.5}$ comparing collisions at 0.5 and 0.55 km. h^{-1}) where more animals would cover the distance between 393 their departure point and the road during the simulation when traveling at 0.55 km.h^{-1} compared to 394 395 0.5 km.h⁻¹ (fig. 5). Nevertheless, these results demonstrate that the cumulative time spent on or near roads by animals, through repeated road encounters and/or a long time spent on the road at each 396 397 encounter (modulated here by patterns of space use, animal speed, and biased random walk 398 parameter w), is the most influential parameter driving collisions probabilities for an animal, and 399 has more importance than road characteristics such as traffic volume or speed.

400 These findings are consistent with studies exploring the relationship between landscape features and 401 roadkill hot-spots: factors such as the presence of attractive habitats on roadsides as well as a high 402 connectivity landscapes conducive to wide ranges of movement are often linked to high road 403 mortality (Table 1, see also de Freitas et al. 2013; de Freitas et al. 2015). Similarly, the temporal 404 distribution of roadkill is dictated by the level of activity, and periods of high mobility such as rut, 405 mating or migrating seasons are especially risky for animals (Table 1). Larger home ranges and 406 exploratory behaviours contribute to higher fitness for individuals by increasing foraging 407 opportunities (Andersson, 1978), but the presence of roads (also applicable to aircrafts, ships or 408 trains) in a landscape lead to costs in the form of road mortality, and selection pressure may lead to 409 road avoidance (Meisingset et al., 2013; Jaeger et al., 2005; Husby & Husby, 2014) or collision 410 avoidance strategies (Lima et al., 2015; Brown and Bomberger Brown, 2013).

411 **4.2 Effect of animal density on WVC risks**

412 The impact of spatial and temporal variations in animals density has been repeatedly reported as a significant driver of WVC (Joyce & Mahoney, 2001; Taylor et al., 2010; Saint-Andrieux et al., 413 2020). Some authors suggested that the number of carcasses on roads could serve as an index of 414 415 animal abundance (Seiler & Helldin, 2006; George et al., 2011). We show that this relationship is 416 linear but strongly modulated by animal movement (fig. 3): compared to animals with no attraction 417 or avoidance towards roads, the slope of the relation between local density and roadkill number is 8 418 times greater for terrestrial animals attracted to roads, and about 70 times smaller when there is road 419 avoidance. When dealing with different species, or with large temporal or spatial scales in which 420 animal movement might be modified by biological seasons and/or landscape features, WVC counts 421 are meaningless, unless the average movement of animals is precisely known. Furthermore, all slopes deviates from isometry ($\beta < 1$), suggesting that the number of casualties on the roads 422 423 underestimates high population abundances.

424 If animal density has an impact on collision numbers, it has no effect on the probability of collision 425 for the individual: WVC are coincidental and have no upper threshold for the number of roadkill, 426 contrary to, for example, a predator looking to feed on a set number of prey. Because of this, there 427 is no mechanism of dilution, or "safety in numbers", in which we might expect a lower collision 428 risk for the individual in a large group of conspecifics (Lehtonen & Jaatinen, 2016). We need to 429 mention that changes in local density have documented effects on individual behaviour which were 430 not explored in the model, and that coult in turn affect movement and hence, WVC indirectly 431 (Kjellander et al., 2004; Dahle & Swenson, 2003; Sword, 2005).

432

433 **4.3 Effects of traffic volume on WVC risks**

434 A general consensus arises in the literature on the number of WVCs increases with road traffic 435 (Fahrig et al., 1995; Rosen & Lowe, 1994; Inbar & Mayer, 1999; Joyce & Mahoney, 2001). This 436 correlation is, however, much discussed with some authors proposing a linear relationship (Fahrig et 437 al., 1995); and others a barrier effect (high-traffic roads having fewer WVCs than less frequented 438 roads because animals avoid the noise and other disturbances they generate: Seiler & Helldin, 2006; 439 Clevenger et al., 2002; van Langevelde et al., 2009). Others explained the lower-than-expected 440 number of collisions reported on some high-traffic roads by a depletion of local animal populations 441 due to roadkill (Ascensao et al., 2019). Those arguments may be unnecessary given our results: the 442 theoretical relationship between traffic and WVCs can be non-linear even in the absence of trafficdependent road avoidance or source-sink dynamics, and roads with heavy traffic would notnecessarily have higher collision rates than intermediate roads.

Whatever assumption we make on animal movement, the number and the individual probability of 445 446 WVC grow asymptotically within the explored range of traffic (fig. 3c and 4c). This observation 447 agrees with the mechanistic model of WVCs by Hels & Buchwald (2001) that too predicts a 448 logarithmic relationship between traffic volume and WVC occurrence. A noticeable difference 449 between the two models is that while the probability of dying animals asymptotically approaches 1 450 for Hels & Buchwald's model, the asymptotic number and probability of WVC lie below a 100% in 451 all our simulations. Note that no asymptote is visible in scenario 1 (habitat fragmentation), which 452 possibly results from the limited range of traffic volume we considered due to computational 453 limitations (major highways worldwide can see up to 15 000 vehicles per hour).

454

455 **4.4 Effects of vehicle speed on WVC risks**

456 In our model, the speed of the moving vehicles is unrelated to the number and individual risk of WVC over the simulated range of speed (0–300 km.h⁻¹, see fig. 3b and 4b). At first sight, a lack of 457 458 effect of vehicle speed may appear at odds with the previous theoretical models suitable for WVC, 459 but it is not: in the ideal gas model (Hutchinson & Waser, 2007), the equations for encounters 460 feature the length of the path of the molecule (speed \times time spent moving) rather than the speed 461 itself. In our model, the distance travelled by a vehicle is constant: they move from one end of the 462 road to the other, and a higher speed does not equate with a longer distance travelled, therefore speed does not increase WVC risks. In contrast, on the subject of the animal's speed, when higher 463 speed means longer travel distances (for scenarios 0 and 4; scenarios 1, 2 and 3 are more complex 464 465 as the movement is also directed), it leads to higher WVC occurrences and risks. In addition, the 466 distance travelled by vehicles was increased by simulating more tortuous roads (high sinuosity 467 index), and the number and probabilities of collision for animals increased as well (fig. 3f and 4f), 468 as expected under the ideal gas framework. Therefore, vehicle speed alone did not impact collision 469 risks in the model.

Empirically, speed is regarded as a significant component of WVCs: the faster the vehicles, the more likely is a collision with an animal (Table 1, see also Farmer & Brooks 2012; Hobday & Minstrell 2008). Vehicles moving at high speeds are most of the time traveling on multi-lane highways with heavy traffic, and the reported association between WVCs and vehicle speed may therefore be confounded with traffic density that we find to be a stronger driver of WVC, or road width which relates to the time required to cross the road. 476 For flying carrion feeders (scenario 3) only, sinuosity is not or slightly negatively correlated to 477 WVC numbers and individual probabilities. We do not have a satisfactory answer to this, but 478 suggest that the Euclidean distance between two attraction centers placed on the road changes with 479 road sinuosity and could impact (with our method of simulating flight) the altitude and therefore the 480 collision risk, which is why terrestrial animals (scenario 2) are not impacted. Road sinuosity can 481 also be treated as a measure of the density of roads, i.e. the surface of road per unit of area, where 482 sinuous roads equate to a higher density than straight roads. In that sense, we show that the density 483 of roads is positively correlated to WVC number and risks for most scenarios of movement.

484

485 **4.5 Model limits**

486 We find that the time spent on the road is the critical factor driving WVC number and probabilities 487 independently of space use patterns, and consequently behaviours we did not implement but that 488 modify the time the animal spends on the road surface (road avoidance, pausing on the road, 489 speeding away from vehicles, evading oncoming vehicles, see Jacobson et al. 2016; Lima et al. 490 2015) are likely to be key components of WVC occurrences. As previously discussed, some of these 491 behaviours, such as intentional collision avoidance, are complex and may be difficult to simulate 492 realistically (Broadhurst & Kanade, 2005; Kuge et al., 2000). Although this was not the focus of 493 this work, we briefly investigated the effects of some of these mechanisms on WVC risks based on 494 the different scenarios of movement already implemented in the model (Table S4).

495 For example, road avoidance dependant on characteristics such as traffic volume, or conversely, 496 habituation to traffic would translate into modifications of the probability of coming into contact 497 with the road, controlled in the model by the strength of avoidance of the road w under scenario of 498 animal movement 4 (road avoidance). The relationships between w and the number and individual 499 probability of collision with a vehicle have been explored in the model: increasing the strength of 500 road avoidance w (e.g., with increasing intensities of road traffic) decreases the risks and numbers 501 of WVC, and low values of w (e.g., habituation to traffic) leads to increased numbers and risks of 502 WVC (fig. 3e and 4e). Therefore, the effect of these mechanisms can be understood in terms of 503 modifications in parameters that have already been explored in the model. In particular, increasing 504 values of road avoidance w with road traffic may lead to relation between WVC risks and traffic to 505 be an inverted U-shaped curve where roads of intermediate traffic volume are more dangerous than 506 high-traffic and low-traffic roads, as conceptualized by Seiler (2006) and Jacobson et al. (2016).

507 We also investigated the possibility for vehicle drivers and animals to evade oncoming collisions 508 (Sup. Material). We show that adding driver or animal reactions to oncoming vehicles – reduced 509 here to a probability of successful collision avoidance depending on vehicle speed - does not 510 change the qualitative results regarding the effects of animal movement on collision risks although 511 the absolute risks did (fig. S4b and S4c). However, behaviours not implemented in the model, such 512 the possibility of evading oncoming collisions, may account for discrepancies between empirical 513 findings and model predictions. For example, while the model did not predict that vehicle speed 514 affects collision risks, reducing speed limits on road has proven efficient in reducing WVC (Hobday 515 & Minstrell, 2008; Jones, 2000), possibly because decreased speeds may afford sufficient time for 516 both animals and drivers to evade potential collisions.

517 The main focus of this work was the movement of the animals, and we reduced "collisions" to the 518 number of times an animal and a vehicle were located less than 5 meters away from each other. This 519 will likely overestimate collision probabilities, especially for smaller-bodied species that can easily 520 fit in-between car tires (Hels & Buchwald, 2001; van Langevelde & Jaarsma, 2005). This is a 521 classical simplification from the ideal gas model that we expect will not alter the general patterns 522 we report, but simply move all the curves down. Finally, we designed the simulation to have 523 constant animal density, which does not allow for population depletion dynamics (justified by the 524 limited temporal scale of 12 hours per simulation run). Based on our conclusions, the total number 525 of roadkill would decrease linearly with population decline but have no impact on individual 526 probabilities of collision (fig. 3d and 4d).

527

528 **4.6 Conclusion**

529 Empirical studies focusing on animal behaviour and movement in addition to other biological traits, 530 road-side habitat type and road characteristics can give valuable insights as to why some species are 531 more often killed on roads than others. We show here that animal behaviour in the presence of roads is a key component of roadkill occurrences, and that understanding space use patterns can give 532 533 insights on the vulnerability of a species to roadkill. Although this model is not intended to provide 534 quantitative predictions of roadkill for particular species, we show that future predictive models should incorporate components of movement ecology to properly address road mortality. Animal 535 536 behaviours in the proximity of transport infrastructures could also be of interest in evolutionary 537 studies, an aspect of road ecology that is lacking so far in the literature (Brady & Richardson, 2017).

538

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543 Annaëlle Bénard: coding work and analysis of the data generated, lead writing of the manuscript, 544 equal contribution to the methodology. Thierry Lengagne and Christophe Bonenfant: equal 545 contribution to the methodology, writing, reviewing and editing of the manuscript. All authors 546 contributed critically to the drafts and gave final approval for publication.

547

548 **Open research statement**

- 549 Code is provided via the following link: https://gitlab.in2p3.fr/annaelle.benard/a-simulation-of-
- animal-vehicle-collisions. No data were collected for this study.

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Section 4 - Regarding drivers' reaction to oncoming collisions and animals' reactions to traffic and
 oncoming vehicles.

The model presented in this paper focuses on animal and vehicle movement in landscapes. Therefore, it does not consider the possibility of vehicle drivers reacting to the presence of an animal on the road or roadside, such as braking or swerving. Likewise, we do not consider potential behaviours from animals: fleeing the road surface or taking off before impact (Brown & Bomberger Brown, 2013; Lima et al., 2015), road avoidance depending on the intensity of traffic (Seiler & Helldin 2006) or habituation to traffic (Legagneux & Ducatez, 2013).

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593 Although the investigation of these behaviours and choices made in the face of a danger was not the 594 focus of this work, some of these behavioural responses by the driver or the animals were somewhat 595 already embedded in the different scenarios of movement already implemented in the model (Table 596 S4). For example, traffic-dependent road avoidance and habituation to road traffic would translate 597 into modifications of the probability with which animals come into contact with the road's surface 598 (Jaeger et al., 2005). In our model, this probability is best modelled by changes in the strength of 599 avoidance of the road w under scenario of animal movement 4 (road avoidance): high values of 600 avoidance result in lower probabilities that animal will approach the road. The relationships 601 between w, and the number and individual probability of collision with a vehicle, have been 602 explored in the model: increasing the strength of road avoidance w (e.g., with increasing intensities 603 of road traffic) decreases the risks and numbers of WVC, and low values of w (e.g., habituation to 604 road traffic where animal no longer avoid roads and roadsides) leads to increased numbers and risks of WVC (fig. 3e and 4e). Therefore, introducing new mechanisms of road avoidance and 605 606 habituation in the model would have predictable impacts on the simulation results.

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608 Other mechanisms, such as reactions to oncoming collisions, may be more difficult to incorporate 609 into the model. Indeed, animal reaction to oncoming collision may differ depending on the species' 610 cognitive abilities or adaptive behaviours (Lima et al., 2015), morphological traits (Brown & 611 Bomberger Brown, 2013), and vehicle speed (Legagneux & Ducatez, 2013). Collision avoidance on the driver's part is modulated by driver decisions (Kuge et al., 2000; Broadhurst & Kanade, 2005), 612 613 vehicle characteristics (e.g., weight, tyres) and speed (Samson et al., 2022), and road characteristics (Samuelraj et al., 2018). Because of this complexity, incorporating realistic animal or driver 614 615 reactions into a mechanistic model is complex and would take our study far from our original idea 616 of exploring the effect of animal movement only on WVC.

We nevertheless introduced here the outcome of of these mechanisms into our modelling to test whether the main conclusions from the model regarding the impact of animal and vehicle trajectories on WVC would differ or not.

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Table S4: Animal and driver behaviours not currently implemented in the wildlife-vehicle collision model. The model presented in this work focuses on realistic animal and vehicle movement, but does not incorporate elements related to animal and vehicle drivers reactions to oncoming collisions or road traffic.

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	Mechanism	Possibility of inclusion in the model	Impact
Animal	Road avoidance dependent on traffic, or other road characteristics	Road avoidance (parameter <i>w</i> , scenario 4 – road avoidance) increasing with intensity of a pre- determined characteristic.	Similar to scenario 4: number
	Habituation	Increasing levels of habituation to traffic translated into decreasing levels of avoidance of the road (parameter <i>w</i> , scenario 4 – road avoidance)	WVC decrease with increasing (fig. 3e and fig. 4e, scenario 4
		Increasing levels of habituation to vehicles lead to a reduction in collision avoidance strategies, such as fleeing from oncoming traffic.	Difficult to predict. In general
	Adaptive reactions to oncoming collisions (<i>e.g.</i> , fleeing the road)	Complex mechanisms that depend on species, vehicles and road	
Driver	Evasive manoeuvres (<i>e.g.</i> , braking, swerving)	implemented as independent mechanisms in the model.	probability of WVC.

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629 Methods

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631 We present here an additional model of WVC that incorporates the possibility for animal or drivers 632 to avoid an oncoming collision. We reduced the mechanism of collision avoidance to a probability 633 of successfully evading an oncoming collision that depended on the speed of the vehicle. Faster 634 vehicles lead to a decreased window of time to flee the road for animals (Legagneux & Ducatez, 635 2013), or to brake/swerve for vehicle drivers (braking distance is also increased, Samson et al., 636 2022), and therefore the probability of being able to avoid the collision is decreased (fig. S4a). In this model, the probability of evading collisions was high (> 98%) for vehicles moving at speeds of 637 10 km.h⁻¹ or less, and decreased with vehicle speed to < 5% for vehicles moving at or over 100 638 km.h⁻¹. For each oncoming collision (animal and vehicle located < 5 meters apart), we drew from a 639

Bernoulli distribution with parameter p = probability of evading collision depending on vehicle speed, to determine whether the collision happened or was avoided. The rest of the model and simulation parameters (parameters baseline values, scenarios of movement, number of runs) were kept identical to the main body of this manuscript. We ran the model for two parameters of interest: a p that is not involved in the implemented mechanism of collision avoidance (mean anii) and a parameter that is involved in the mechanism (vehicle speed).

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648 **Results and Discussion**

649 We show the mean number of collisions for varying mean speeds of animals (fig. S4b) and vehicles 650 (fig. S4c). When testing the impact of mean animal speed on WVC, the baseline value for vehicle speed is 80 km.h⁻¹ (Table 2), and therefore the probability of avoiding an oncoming collision is 651 652 constant at around 83%. Therefore, the mean number of collisions for each scenario of movement is 653 decreased in the model implementing the possibility of evading collisions (fig. S4b) compared to 654 the results showed on the original model (fig. 3a). However, the shape of the curves for each 655 scenario and the ranking of each scenario were not changed compared to the original model, and 656 thus the main conclusions of our work regarding the importance of animal speed and the impact of 657 animal trajectories on WVC remain unchanged, with only the quantitative results shifted downward.

658 The speed of the vehicle had little to no impact on WVC numbers in the original model (fig. 3b). 659 However, in the new model with probabilities of evading collisions, we see an increase in mean 660 WVC numbers with increasing vehicle speeds (fig. S4c). Results interpretation is straightforward, 661 as the probability of a collision happening increases with the speed of the vehicle: the observed impact of vehicle speed on the number of collisions reflects the probability of collision (*i.e.*, 1 -662 663 probability of evading collision, fig. S4a). Indeed, as vehicle speed increases and probability of 664 collision nears 1 (probability of evading collision nears 0, fig. S4a), the mean number of collisions 665 for each scenario of movement tends toward the values we found in the original model (fig. 3b). 666 Therefore, the ranking between each scenario of movement was unchanged.

In conclusion, we find that this new model yielded results similar to the original model presented in the main body of this work when testing the effect of a parameter not involved in oncoming collision avoidance (mean animal speed), with the curves for mean number of WVC per scenario of movement simply shifted downwards. This conclusion should hold for all other parameters of interest not implicated into the mechanism (road sinuosity, animal density, strength of attraction/avoidance *w*, volume of traffic) where vehicle speed is similarly held constant at 80 km.h⁻

- ¹. In regard to the speed of vehicles, which directly influenced the probability of avoiding collisions,
 we found that the mean number of WVC represented the distribution of collision avoidance
 probabilities we defined in the model.
- 676 This mechanism is an oversimplified representation of reality, and effectively modelling animal or
- 677 driver behaviour for avoiding collisions with oncoming vehicles is a complex task well beyond the
- 678 scope of this study. The inclusion of this new collision avoidance mechanism does not alter the key
- 679 findings and conclusions drawn in the model regarding the effects of animal movement on wildlife-
- 680 vehicle collisions.

