

# Automated identification of invasive rabbitfishes in underwater images from the Mediterranean Sea

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1	Automated identification of invasive rabbitfishes in underwater images
2	from the Mediterranean Sea
3	
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#### 17 Abstract

Coastal ecosystems of the Mediterranean Sea are among the richest in non-indigenous
 species, mostly due to the establishment of species coming from the Red Sea through the
 Suez Canal. Two herbivorous rabbitfishes, *Siganus rivulatus* and *Siganus luridus*, are
 already invasive in the South-Eastern part of the Mediterranean Sea where they cause
 ecological damage by overgrazing algae.

23 2. The early detection and the counting of these non-indigenous species in the rest of the
24 Mediterranean Sea is thus a major challenge for scientists and ecosystem managers.
25 However, analyzing images from divers or remote cameras is a demanding task.

3. Here, a dataset of 31,285 images of *Siganus* spp and of six common native fishes to the Mediterranean Sea was built from 40 underwater videos recorded in three reef habitats. A deep learning algorithm was then trained to identify *Siganus spp* on images mixing the eight Mediterranean species. Finally, the algorithm and a post-processing filtering were tested with an independent dataset of 2,024 images.

31 4. The model had a recall of 0.92 for the Siganus genus (i.e. two Siganus species combined). 32 After a confidence-based post-processing the recall increased to 0.98 with only 4 out of 272 33 images of Siganus spp being misclassified. Accuracy reached a score of 0.61 meaning that 34 experts would have to discard false positives. Images of five native species not present in the 35 training dataset yielded similar false positive rates than species present in the training dataset. 36 Overall, the automatic processing of images by our model then the checking of putative 37 Siganus images by experts require up to five times lower effort than a full processing by 38 experts.

5. Our algorithm can efficiently help to detect these two invasive fishes in underwater images
to evaluate progress towards conservation objectives and accelerate citizen-based monitoring
of coastal ecosystems.

Keywords : coastal ecosystems,deep learning, image analysis, Mediterranean sea, non indigenous species, Siganus spp

#### 44 Introduction

45 Biodiversity has been increasingly impacted by global change in all ecosystems 46 (Jaureguiberry et al., 2022). Among those impacts, climate warming and introduction of non-47 indigenous species are acting synergistically in many regions (La Sorte & Jetz, 2010; Bennett 48 et al., 2021). Mediterranean coastal ecosystems are an emblematic case of such marked 49 changes in biodiversity (Albano et al., 2021). Indeed, the opening of the Suez Canal in 1869 50 and its recent expansion have permitted the entry of many fish species from the Red Sea into 51 the Mediterranean Sea. To date, the total number of these non-indigenous species (NIS 52 hereafter) is above 150 (Azzurro, Smeraldo & D'Amen, 2022). The northward and westward 53 geographic range expansion and the increase in abundance of those tropical NIS has been 54 favored by the rapid warming of Mediterranean Sea surface waters (>1°C since 1980) 55 (Shaltout & Omstedt, 2014).

56 Rabbitfishes Siganus rivulatus and Siganus luridus have become the most abundant 57 herbivorous fishes along the South-Eastern coast of the Mediterranean Sea since their 58 establishment in 1927 and 1956, respectively (Ben-Tuvia, 1964; Bariche, Letourneur & 59 Harmelin-Vivien, 2004; Sala et al., 2011). They now contribute to more than 50% of the 60 herbivory activity (Magneville et al., 2023). Importantly, in many of these coastal habitats, 61 Siganus spp have overgrazed macroalgae to the point that the diversity of other organisms 62 has decreased and ecosystem functioning has been disrupted (Vergés et al., 2014; Peleg et 63 al., 2019). Siganus spp have been colonizing northward up to the Aegean Sea and westward 64 up to Sicily for the last two decades (Azzurro & Andaloro, 2005). Some individuals were even 65 observed along the coast of France (Daniel, Piro & FranCour, 2009) and Croatia (Dulčić et al., 66 2011). Yet, their long-term establishment and new colonization pathways are still under 67 scrutiny.

In this context of accelerating NIS expansion, it is thus of utmost importance for scientists and ecosystem managers to early detect the arrival of such potentially invasive species and then to monitor their settlement and abundance dynamics. Monitoring non-indigenous marine fishes has been mostly carried out using standardized surveys by scuba divers (Brock, 1954;

Thresher & Gunn, 1986; Holland et al., 2020) and opportunistic records from fisheries landings (e.g. Kleitou et al., 2022). Yet, underwater visual census of fish requires trained divers who may still miss elusive species, such as *Siganus*, especially when they are rare. As an alternative, monitoring NIS based on underwater remote cameras and underwater pictures from divers have been increasingly applied (Peters, Sink & Robinson, 2019). However, the identification of target NIS on such images is a demanding task in terms of time and expertise in fish taxonomy.

79 Owing to the recent development of artificial intelligence and especially Deep Learning 80 algorithms (DLa) there is a growing effort to automate the identification of fish species in 81 underwater images (Ditria, Jinks & Connolly, 2021). DLa have been demonstrated to efficiently 82 identify indigenous reef fishes in several coastal ecosystems (Villon et al., 2018; Lopez-83 Marcano et al., 2021) but none addressed the key issue of detecting NIS in the Mediterranean 84 Sea. As Siganus spp can change color rapidly from olive green to dark brown with a mottled 85 pattern, training an algorithm able to identify them remains challenging. In this study, images 86 of the two Siganus spp and six common native fish species were collected across the Mediterranean Sea. Then, a DLa was trained to identify polymorphic Siganus spp in 87 88 underwater images, and its performance was tested in various conditions.

89

#### 90 Material and Methods

#### 91 Image datasets

92 DLa require two independent image datasets: the first one used to train the model, *i.e.* to find 93 the optimal parameters discriminating the species of interest, and the second one used to test 94 the trained model reliability in the various conditions of application. To build those datasets, 95 underwater remote videos were recorded during three field campaigns in Northern Crete 96 Island (Greece) and the Gulf of Lion (France) in 2017, 2019, and 2020 (Table 1).

98	Table 1. Summary of u	derwater videos and fish images use	ed to train and test the deep
		0	

99

learning classifier.

Use	Territory	Site	Date (year- month)	Habitat	Number of videos	Number of annotated images of fishes	Presence of <i>Siganus</i> spp
Train	France	Frontignan	2019-09	macro-algae / bare rocks / turf	1	134	
Train	Crete	Hersonissos	2019-10	macro-algae / bare rocks / sand	26	5247	Х
Train	Crete	Kokkini Hani	2019-10	macro-algae / turf / seagrass	13	4636	Х
Test	France	Banyuls	2017-07	macro-algae / bare rocks / turf / seagrass	2	114	
Test	Crete	Hersonissos	2019-06	macro-algae / bare rocks / sand	3	292	Х
Test	Crete	Kokkini Hani	2019-06	macro-algae / turf / seagrass	1	831	Х
Test	Crete	Bali	2019-06	macro-algae / turf / seagrass	1	139	Х
Test	France	Carry le Rouet	2020-05	macro-algae / bare rocks / turf / seagrass	2	783	

#### 100

All videos were recorded with GoPro HERO5 Black (GoPro Inc., USA) set with "linear" angleof-view, full HD resolution (1920\*1080 pixels) and 25 frames per second. The cameras were
set up on the seafloor in shallow (1-5m deep) coastal habitats dominated by seagrass,
macroalgae, bare rocks, or sand (Table 1).

To ensure that the test of the algorithm was performed independently from the training, the two datasets were built using videos recorded in different locations or seasons (Table 1). The training dataset was built with images from 40 videos for a total of 140h recorded between the 108 2<sup>nd</sup> and the 10<sup>th</sup> of October 2019 in Crete (Greece). For the testing dataset, nine videos
109 recorded in France (July 2017 and May 2020) and in Crete (June 2019) were used.

110 Videos of different lengths were extracted from the raw recordings based on the presence of 111 fish and images were extracted at the rate of one or three images per second yielding 11 to 112 1,064 frames from a video. On each of these frames, fish individuals which body was entirely 113 visible (i.e. not hidden by another fish, rocks or algae and not facing camera and not recorded 114 from the face or behind), and which could be identified to the species level were annotated 115 using a bounding box encompassing the body and labeled with species name. Because of 116 these rules, most of the annotated fish individuals were not present on temporally successive 117 frames. Furthermore, as fish individuals were of various sizes and were at various distances 118 from the camera, the annotated images had variable sizes and length/depth ratios. Given the 119 diversity of seafloors among videos (rocks, algae, seagrass), varying hydrological conditions 120 (turbidity, light), as well as the movement of filmed components (algae, other fishes), the 121 images displayed different backgrounds even within a single video (Figure S1). Additionally, 122 considering the mobility of the target fish species, the images exhibited a range of fish body 123 orientations. Special attention was given to ensuring that images of Siganus spp represented 124 the diversity of colorations encountered in the wild (Figure 1).



126Figure 1. Examples of the different body shapes and colors of Siganus luridus (a, b, c) and Siganus rivulatus (d,127e, f, g) in the test dataset

For this study, images of the two non-indigenous *Siganus luridus* and *Siganus rivulatus*, were used along with images of the six most common fish species in our videos (*Chromis chromis*, *Diplodus sargus*, *Diplodus vulgaris*, *Sarpa salpa*, *Sparisoma cretense*, and *Thalassoma pavo*), which are also prevalent in the Mediterranean Sea. Each species represents a class learned by the deep learning classifier.

A total of 31,285 images were annotated from videos of the training dataset. For each class, 1,250 to 1,256 unique images were randomly selected among those available and horizontally flipped to augment the representativity of the dataset (Villon et al., 2018). For the test, 2,159 images (Table S1) of the eight species common with the training dataset (two *Siganus* spp. and six common fishes) were used, along with 100 images of five species not present in the 138 training dataset (Diplodus puntazzo, Oblada melanura, Mullus surmuletus, Lithognathus

139 mormyrus, Trachurus mediterraneus, N = 20 individuals for each species).

#### 140 Training of the classifier

141 The classifier involved fine-tuning a ResNet50 model that had been previously trained on the 142 ImageNet dataset using PyTorch (torch 1.8.1). At the beginning of the training, all images were 143 resized to 224x224 pixels. The batch size was set at 16. The Adam algorithm was employed 144 as the optimizer, and the negative log-likelihood loss function was used. To prevent overfitting, 145 it was ensured that the loss was decreasing after each epoch. If this was not the case for 20 146 consecutive epochs, the training was stopped. The algorithm was trained using an Nvidia 147 Quadro RTX 3000 GPU card on a Dell Precision 7550 laptop with an Intel® Core™ i7 148 processor and 62GiB RAM.

#### 149 Testing the classifier

For each class, the number of true positive (TP) images, representing correctly identified images, was computed. Additionally, the number of false negative (FN) images, indicating images belonging to a given class but missed by the classifier, and the number of false positive (FP) images, representing misidentified images in a given class, were also calculated.

The model efficiency was assessed using the recall and precision metrics for each class. The recall is the proportion of correctly labeled individuals among all individuals of this class and is computed as:  $\frac{TP}{TP+FN}$ . It is maximal (and equals 1) when all individuals of the class are correctly identified, and it decreases when some individuals of the analyzed class are misidentified as fish species from other classes.

The precision is the proportion of correctly labeled individuals among the items identified as this class and is computed as:  $\frac{TP}{TP+FP}$ . It is maximal (and equals 1) when there are no individuals belonging to another class, and decreases when some individuals of other classes are misidentified as the analyzed class. Metrics were also computed for the two *Siganus* spp combined, considering that in most applications, the goal is to detect the presence of this non-indigenous genus. For this purpose, a new class "siganus" was created, that bundles the classes "siganus\_luridus" and "siganus rivulatus".

167 In the context of automated identification of NIS, the priority is to ensure that NIS individuals 168 are not missed, hence to maximize recall (i.e. few FN). However, the time of experts required 169 for checking the images classified as NIS has also to be as low as possible, hence precision 170 should remain as high as possible (i.e. few FP).

For each image processed by the Dla, the most likely class returned has an associated confidence score ranging from 0 to 1, with values close to 1 indicating the most reliable classes. Thus, images with a confidence score lower than 0.8 were transferred to an "unsure" class. The recall was computed after removing these "unsure" images.

Image size may play an important role in fish identification. Indeed, a smaller image can be indicative of an individual which is far away from the camera and difficult to identify. Thus, the influence of image area (height x width in pixels) on TP and FP rates was analyzed.

178

#### 179 **Results**

180 The training lasted 31 minutes and ran 36 epochs when the early stopping procedure ceased181 the training as the loss did not decrease for the last 20 epochs.

The average recall was 0.62 among the eight classes, the confusion matrix being shown in Figure 2. The recall for the classes *S. luridus* and *S. rivulatus* were 0.74 and 0.47, respectively. When merging the two *Siganus* species in a single class, the recall reached 0.92. The precision was 0.42 for the class *siganus\_luridus*, 0.18 for the class *siganus\_rivulatus*, and 0.44 when considering the merged class "siganus".

a)										b)	-									
_	chromis chromis -	202	31	14	28	34	25	7	3		chromis chromis -	171	13	3	10	20	3	1		
	diplodus sargus -	37	80	46	70	20	79	2	0		diplodus sargus -	25	56	22	27	45	0	0	NI	a ima
	diplodus vulgaris -	60	35	255	42	99	95	1	0	-	diplodus vulgaris -	23	12	157	8	55	1	0		300
labe	sarpa salpa -	6	3	13	61	11	40	1	0	labe	a sarpa salpa - an	0	0		26	15	0	0		- 200
True I	siganus luridus -	2	4	4	5	202	46	3	1	ne		0	0	4	30	15	0	0		
	siganus rivulatus -	1	2	5	4	58	64	0	1		siganus -	0	0	2	2	268	0	0		100
	sparisoma cretense -	0	0	0	0	3	6	7	0		sparisoma cretense -	0	0	0	0	5	4	0		0
	thalassoma pavo -	24	3	8	7	54	10	7	228		thalassoma pavo -	15	0	2	0	29	0	184		
	l.	1		1			-	se -	_	1	L	1		1	1	1	se-			
		chromis chromis	diplodus sargus	diplodus vulgaris	sarpa salpa	siganus luridus	siganus rivulatus	sparisoma creten:	thalassoma pavo			chromis chromis	diplodus sargus	diplodus vulgaris	sarpa salpa	siganus	sparisoma creten:	thalassoma pavo		
Predicted label														Predi	cied	lanel				

187

Figure 2. Performance of the Deep-Learning algorithm in identifying fish species in underwater images using a confusion matrix where correct classifications are on the diagonal. Largest numbers of individuals are highlighted.
 Comparison between the initial confusion matrix (a) and the confusion matrix after merging the two *Siganus* classes and removing the images for which the confidence score was lower than 0.8 (b). Each row of the matrix represents an actual class while each column represents a predicted class.

After the post-processing removal of unsure predictions, the recall for the class *Siganus* increased to 0.99 and the precision increased to 0.61. On the four videos with at least one *Siganus* individual, the *Siganus* genus was detected in at least one frame (84% up to 100% of *Siganus spp* were detected per video).

197 Among the 100 images of species not present in the training database, only 17 (5 Diplodus

198 puntazzo, 5 Lithognathus mormyrus, 3 Mullus surmuletus, 3 Oblada melanura, 1 Trachurus

199 mediterraneus) were misidentified as Siganus spp among which seven remained misidentified

200 as *Siganus spp* after the post processing removing unsure predictions.





208 None of the correctly identified Siganus images had an area < 4,900 pixels while all the 13 209 images with area <4,900 pixels, which were labeled as Siganus, were in fact of one of the six 210 native species (Figure 3). Choosing a cut-off at 10,000 pixels removes 62 FP (36%) and 9 TP 211 (3%) increasing the precision to 0.71 while marginally decreasing the recall for the class 212 Siganus to 0.98.

213

#### 214 Discussion

215 The recall of 0.92 for the class "siganus" demonstrates the reliability of the trained DLa since 216 only 32 out of 402 images of Siganus spp were misclassified as images of indigenous fish 217 species (Figure 2a). Among those 32 incorrectly labeled images (FN), 28 had a confidence 218 score lower than 0.8 and only 4 (Figure 2b) had a confidence score greater than 0.8 so were 219 retained in the final model outputs after post-processing. Those 4 false negatives were images 220 of Siganus individuals shot in turbid water, exposed to direct sunlight, and of an individual 221 masked by a Sarpa salpa (Figure 4). If an expert checks the 28 unsure images (7% of the 222 images analyzed) and identifies Siganus on all of them, only 1% (4 out of 272) of Siganus spp 223 images would be missed (false negatives). Besides, ignoring the 28 unsure images had no 224 impact on the detection of *Siganus* at the scale of the video sequence as each of our videos 225 contains at least 10 images of Siganus spp. So, our classifier coupled to long running videos 226 would most likely detect rabbitfish presence even when elusive or rare.

227 The precision of 0.44 for the class "siganus" means that 56% of images labeled as "siganus" 228 were actually indigenous species. The confidence-based post-processing removal of unsure 229 predictions increased the precision for the "siganus" class to 0.61, meaning that among the 230 437 images labeled as Siganus 169 were actually images of native species. This model's 231 performance level is within the average performance level of existing models. The recall for 232 Siganus obtained by this model is lower to other studies on a single fish species (Ditria et al., 233 2020; Lopez-Marcano et al., 2021). The overall performance on the 8 species is higher to the 234 classification algorithm by Catalán et al., (2023) for 9 Mediterranean indigenous fishes (recall: 235 0.76, precision: 0.37) and to the one by Jenrette et al., (2022) on 47 species of tropical sharks 236 and rays, which proposes a combination of three models (a detection model, a binary sorting 237 model and a classification model), achieving an overall accuracy of 70%. The two species 238 most commonly misidentified as Siganus spp were the two seabream Diplodus vulgaris 239 (194FP) and *Diplodus sargus* (99FP) (Figure 2a). Yet, these two species have a shape and 240 color pattern different from Siganus spp. Visual inspection of these confidently misclassified 241 images revealed they mostly depict individuals in the background or with their body partially 242 hidden (because of their orientation relative to the camera, or movement of the tail) making 243 them more difficult to identify even by humans (Figure 4). Meanwhile, the *Diplodus* individuals

- 244 close to the camera and not masked were correctly classified (i.e. among the 921 Diplodus
- 245 vulgaris and D. sargus, 44% were classified as a Diplodus). Hence, the confusion between
- 246 Diplodus and Siganus is context dependent.
- 247



As a consequence, not accounting for *Siganus* identified in images of less than 10,000 pixels discarded more than one third of false positive *Siganus* identifications while keeping recall above 98% (i.e. only 9 images of *Siganus* would be misidentified as a native fish species). Such small images correspond to individuals filling less than 1/20th of a FullHD video frame (1920×1080, > 2 millions pixels). Hence, as in all our remote underwater videos, it is likely that when present *Siganus* individuals will be recorded with an apparent size > 10,000 pixels.

The processing of the 2,024 images belonging to the test dataset took 10s on a laptop computer and the optional checking of the 28 predictions of *Siganus* with low confidence threshold would take less than 3 minutes for an expert (assuming 5s per image). Checking only the 437 images identified as *Siganus* would take 36 minutes and would allow discarding false positives. Hence, as the analysis of the 2,024 images by an expert would take 168 minutes (assuming 5s per image with no break), the DLa thus provides at least a 5-fold gain of time.

270 One limitation of our DLa classifier is the identification of Siganus only in images containing a 271 single individual at its center. Thus, in practice, raw videos or pictures from underwater 272 cameras must be cropped to focus only on fish individuals, which requires no expertise in 273 taxonomy and is fast using any software able to save a screenshot (e.g. VLC). A future priority 274 will be to couple this identification model with a general fish detection model able to segment 275 all fish individuals from a video frame or a picture (e.g. Knausgård et al., 2022). Then our 276 model will process those extracted images. Even if current detection models are only able to 277 identify a single species (Lopez-Marcano et al., 2021), recent advances in other taxa show encouraging results such as DeepFaune (Rigoudy et al., 2023) for French terrestrial fauna, or 278 279 for insects (Teixeira et al., 2023). However, these detection models require images with all 280 individuals annotated, which is time-consuming as underwater images could contain >100 fish 281 individuals. Another limitation is that classifiers are only able to identify species in the training 282 dataset, which requires collecting more than 1,200 images per species or applying few-shot 283 learning techniques that usually decrease recall and precision (e.g. Villon et al., 2021). In our

case, among images from the five indigenous species not present in the learned classes, only
7% were confidently classified as *Siganus spp*, which was lower than the false positive rate
for the six indigenous species present in the training dataset. Hence, our model would likely
not misidentify as *Siganus* species that are not present in the training dataset.

288

#### 289 **Conservation perspectives**

290 Developing rapid and robust knowledge on the distribution and density of NIS is crucial to 291 evaluate progress against international agreements, including the UN Sustainable 292 Development Goals, Convention on Biological Diversity Targets, EU Biodiversity Strategy, and 293 the Green Deal. Towards this goal, improving our capacity to automate data collection and 294 processing is needed to dramatically increase the amount and quality of information and 295 knowledge available to scientists and decision-makers.

296 Remote or diver-operated underwater videos have been increasingly used by scientists and 297 ecosystem managers to monitor fishes for the last decade across the Mediterranean Sea 298 (Nalmpanti et al., 2023). We believe that our DLAs will be useful for ecosystem managers to 299 accelerate the analysis of videos collected during those monitoring programs of NIS, ensuring 300 detection of Siganus spp. Towards this aim we release it in a public repository 301 (https://github.com/valentine-fleure/Siganus identifier). Our classifier applied to long 302 underwater videos will pave the way to high-frequency monitoring of rabbitfishes throughout 303 the Mediterranean basin. Such massive data are needed to describe the geographic 304 expansion of rabbitfishes through time to ultimately unravel its driver (climate, connectivity), 305 and test the potential benefit of MPAs on lowering their abundance through increased level of 306 predation by large fishes.

In addition, as the Mediterranean coastal ecosystems are a popular place for underwater activities (snorkeling, freediving, SCUBA diving) and as many of these recreational users own underwater cameras, there is a high potential to collect a large number of images in many places. As most recreational users have no to little ability to identify all fish species, our publicly available algorithm will favor current citizen science monitoring of NIS. Indeed, our algorithm

will favor the participation of new users to those programs as well as help them improving their identification skills so that less individuals of rabbitfish will be missed. Ultimately, these recorded and validated new occurrences would be integrated in the Global Biodiversity Information Facility (GBIF) and the Ocean Biodiversity Information System (OBIS) to actualize rabbitfish geographic range size and estimate of invasion rate.

317 Given the increasing number of NIS entering the Mediterranean Sea as well as the expansion 318 of those already present in the South-Eastern part (Azzurro & Andaloro, 2005; Azzurro, 319 Smeraldo & D'Amen, 2022), video-algorithm based monitoring programs should be able to 320 monitor more species. Toward this aim, it would be relevant to upgrade the model by 321 identifying more indigenous species as well as more NIS. Among the NIS that should be 322 prioritized to be added to the training database, lionfish (*Pterois miles*), silver-cheeked toadfish 323 (Lagocephalus sceleratus), yellowspotted puffer (Torquigener flavimaculosus), and sweepers 324 (Pempheris spp) are of the highest priority given their abundance in the Eastern Mediterranean 325 and their rapid northward and westward expansion (Zenetos et al., 2022). These Lessepsian 326 species are morphologically distinct from one another and from indigenous species and should 327 thus be easily identified by algorithms.

328 One step further, AI algorithms may unlock the potential of video-based monitoring beyond 329 the case of non-indigenous species detection through automated identification of all fish 330 species, including endangered one, as well as estimating their size and behaviors (Ditria, Jinks 331 & Connolly, 2021) to assess a large set of indicators as envisioned by the Essential 332 Biodiversity Variable framework (Kissling et al., 2018). These variables would deliver multiple 333 indicators of coastal ecosystem status, climate change effects, and spatio-temporal trends in 334 fish biodiversity, as required to meet the monitoring requirements of the EU Water Framework 335 Directive (WFD) and the Marine Strategy Framework Directive (MSFD).

336

337

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349

### 350 Conflict of interest

351 The authors have no conflict of interest to declare.

352

### 353 Author contributions

- 354 Valentine Fleuré, David Mouillot and Sébastien Villéger conceived and designed the study;
- 355 Camille Magneville and Sébastien Villéger collected the data. Valentine Fleuré wrote the first
- draft of the manuscript. All authors contributed to subsequent revisions and approved the
- 357 final version.

- Albano, P.G., Steger, J., Bošnjak, M., Dunne, B., Guifarro, Z., Turapova, E., et al. (2021).
  Native biodiversity collapse in the eastern Mediterranean. *Proceedings of the Royal Society B: Biological Sciences*, 288(1942), 20202469. https://doi.org/10.1098/rspb.2020.2469
- Azzurro, E. & Andaloro, F. (2005). A new settled population of the Lessepsian migrant Siganus luridus (Pisces: Siganidae) in Linosa Island - Sicily Strait. *Journal of the Marine Biological Association of the UK*, 84. https://doi.org/10.1017/S0025315404009993h
- Azzurro, E., Smeraldo, S. & D'Amen, M. (2022). Spatio-temporal dynamics of exotic fish species in the Mediterranean Sea: Over a century of invasion reconstructed. *Global Change Biology*, 28(21), 6268–6279. https://doi.org/10.1111/gcb.16362
- Bariche, M., Letourneur, Y. & Harmelin-Vivien, M. (2004). Temporal Fluctuations and Settlement Patterns of Native and Lessepsian Herbivorous Fishes on the Lebanese Coast (Eastern Mediterranean). *Environmental Biology of Fishes*, 70(1), 81–90. https://doi.org/10.1023/B:EBFI.0000022928.15148.75
- Bennett, S., Santana-Garcon, J., Marbà, N., Jorda, G., Anton, A., Apostolaki, E.T., et al. (2021). Climate-driven impacts of exotic species on marine ecosystems. *Global Ecology and Biogeography*, 30(5), 1043–1055. https://doi.org/10.1111/geb.13283
- Ben-Tuvia, A. (1964). Two siganid fishes of Red Sea origin in the eastern Mediterranean. Ministry of Agriculture, Department of Fisheries, Sea Fisheries Research Station.
- Brock, V.E. (1954). A Preliminary Report on a Method of Estimating Reef Fish Populations. *The Journal of Wildlife Management*, 18(3), 297–308. https://doi.org/10.2307/3797016
- Catalán, I.A., Álvarez-Ellacuría, A., Lisani, J.-L., Sánchez, J., Vizoso, G., Heinrichs-Maquilón,
   A.E., et al. (2023). Automatic detection and classification of coastal Mediterranean fish from underwater images: Good practices for robust training. *Frontiers in Marine Science*, 10.
- Daniel, B., Piro, S. & FranCour, P. (2009). Lessepsian rabbitish Siganus luridus reached the French Mediterranean coasts. *Société Française d'Ichtyologie*, 33(2), 163–164. https://doi.org/10.26028/cybium/2009-332-007
- Ditria, E.M., Jinks, E.L. & Connolly, R.M. (2021). Automating the analysis of fish grazing behaviour from videos using image classification and optical flow. *Animal Behaviour*, 177, 31–37. https://doi.org/10.1016/j.anbehav.2021.04.018
- Ditria, E.M., Sievers, M., Lopez-Marcano, S., Jinks, E.L. & Connolly, R.M. (2020). Deep learning for automated analysis of fish abundance: the benefits of training across multiple habitats. *Environmental Monitoring and Assessment*, 192(11), 698. https://doi.org/10.1007/s10661-020-08653-z

- Dulčić, J., Dragičević, B., Grgičević, R. & Lipej, L. (2011). First substantiated record of a Lessepsian migrant—the dusky spinefoot, <I>Siganus luridus</I> (Actinopterygii: Perciformes: Siganidae), in the Adriatic Sea. *Acta Ichthyologica Et Piscatoria*, 41(2), 141–143. https://doi.org/10.3750/AIP2011.41.2.12
- Holland, M.M., Smith, J.A., Everett, J.D., Vergés, A. & Suthers, I.M. (2020). Latitudinal patterns in trophic structure of temperate reef-associated fishes and predicted consequences of climate change. *Fish and Fisheries*, 21(6), 1092–1108. https://doi.org/10.1111/faf.12488
- Jaureguiberry, P., Titeux, N., Wiemers, M., Bowler, D.E., Coscieme, L., Golden, A.S., et al. (2022). The direct drivers of recent global anthropogenic biodiversity loss. *Science Advances*, 8(45), eabm9982. https://doi.org/10.1126/sciadv.abm9982
- Jenrette, J., Liu, Z.Y.-C., Chimote, P., Hastie, T., Fox, E. & Ferretti, F. (2022). Shark detection and classification with machine learning. *Ecological Informatics*, 69, 101673. https://doi.org/10.1016/j.ecoinf.2022.101673
- Kissling, W.D., Walls, R., Bowser, A., Jones, M.O., Kattge, J., Agosti, D., et al. (2018). Towards global data products of Essential Biodiversity Variables on species traits. *Nature Ecology & Evolution*, 2(10), 1531–1540. https://doi.org/10.1038/s41559-018-0667-3
- Kleitou, P., Moutopoulos, D.K., Giovos, I., Kletou, D., Savva, I., Cai, L.L., et al. (2022). Conflicting interests and growing importance of non-indigenous species in commercial and recreational fisheries of the Mediterranean Sea. *Fisheries Management and Ecology*, 29(2), 169–182. https://doi.org/10.1111/fme.12531
- Knausgård, K.M., Wiklund, A., Sørdalen, T.K., Halvorsen, K.T., Kleiven, A.R., Jiao, L., et al. (2022). Temperate fish detection and classification: a deep learning based approach. *Applied Intelligence*, 52(6), 6988–7001. https://doi.org/10.1007/s10489-020-02154-9
- La Sorte, F.A. & Jetz, W. (2010). Projected range contractions of montane biodiversity under global warming. *Proceedings of the Royal Society B: Biological Sciences*, 277(1699), 3401– 3410. https://doi.org/10.1098/rspb.2010.0612
- Lopez-Marcano, S., L. Jinks, E., Buelow, C.A., Brown, C.J., Wang, D., Kusy, B., et al. (2021). Automatic detection of fish and tracking of movement for ecology. *Ecology and Evolution*, 11(12), 8254–8263. https://doi.org/10.1002/ece3.7656
- Magneville, C., Leréec Le Bricquir, M.-L., Dailianis, T., Skouradakis, G., Claverie, T. & Villéger, S. (2023). Long-duration remote underwater videos reveal that grazing by fishes is highly variable through time and dominated by non-indigenous species. *Remote Sensing in Ecology and Conservation*, 9(3), 311–322. https://doi.org/10.1002/rse2.311

- Nalmpanti, M., Chrysafi, A., Meeuwig, J.J. & Tsikliras, A.C. (2023). Monitoring marine fishes using underwater video techniques in the Mediterranean Sea. *Reviews in Fish Biology and Fisheries*, 33(4), 1291–1310. https://doi.org/10.1007/s11160-023-09799-y
- Peleg, N., Molnar, P., Burlando, P. & Fatichi, S. (2019). Exploring stochastic climate uncertainty in space and time using a gridded hourly weather generator. *Journal of Hydrology*, 571, 627–641. https://doi.org/10.1016/j.jhydrol.2019.02.010
- Peters, K., Sink, K.J. & Robinson, T.B. (2019). Sampling methods and approaches to inform standardized detection of marine alien fouling species on recreational vessels. *Journal of Environmental Management*, 230, 159–167. https://doi.org/10.1016/j.jenvman.2018.09.063
- Rigoudy, N., Dussert, G., Benyoub, A., Besnard, A., Birck, C., Boyer, J., et al. (2023). The DeepFaune initiative: a collaborative effort towards the automatic identification of French fauna in camera-trap images. 2022.03.15.484324.
- Sala, E., Kizilkaya, Z., Yildirim, D. & Ballesteros, E. (2011). Alien Marine Fishes Deplete Algal Biomass in the Eastern Mediterranean. *PLOS ONE*, 6(2), e17356. https://doi.org/10.1371/journal.pone.0017356
- Shaltout, M. & Omstedt, A. (2014). Recent sea surface temperature trends and future scenarios for the Mediterranean Sea. *Oceanologia*, 56(3), 411–443. https://doi.org/10.5697/oc.56-3.411
- Teixeira, A.C., Ribeiro, J., Morais, R., Sousa, J.J. & Cunha, A. (2023). A Systematic Review on Automatic Insect Detection Using Deep Learning. *Agriculture*, 13(713), 713. https://doi.org/10.3390/agriculture13030713
- Thresher, R.E. & Gunn, J.S. (1986). Comparative analysis of visual census techniques for highly mobile, reef-associated piscivores (Carangidae). *Environmental Biology of Fishes*, 17(2), 93–116. https://doi.org/10.1007/BF00001740
- Vergés, A., Tomas, F., Cebrian, E., Ballesteros, E., Kizilkaya, Z., Dendrinos, P., et al. (2014). Tropical rabbitfish and the deforestation of a warming temperate sea. *Journal of Ecology*, 102(6), 1518–1527. https://doi.org/10.1111/1365-2745.12324
- Villon, S., Iovan, C., Mangeas, M., Claverie, T., Mouillot, D., Villéger, S., et al. (2021). Automatic underwater fish species classification with limited data using few-shot learning. *Ecological Informatics*, 63, 101320. https://doi.org/10.1016/j.ecoinf.2021.101320
- Villon, S., Mouillot, D., Chaumont, M., Darling, E., Subsol, G., Claverie, T., et al. (2018). A deep learning algorithm for automatic identification of coral reef fish species on images.
- Zenetos, A., Albano, P.G., Garcia, E.L., Stern, N., Tsiamis, K. & Galanidi, M. (2022). Established non-indigenous species increased by 40% in 11 years in the Mediterranean Sea. *Mediterranean Marine Science*, 23(1). https://doi.org/10.12681/mms.29106