

Rapid turnover of a pea aphid superclone mediated by thermal endurance in central Chile

Sebastián I Martel, Cristián A Zamora, Natalia Ricote, Daniela A Sepúlveda, Frédérique Mahéo, Jean-Christophe Simon, Christian C. Figueroa, Enrico L. Rezende, Francisco Bozinovic

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25 ABSTRACT

Global change drivers are imposing novel conditions on Earth's ecosystems at an 26 unprecedented rate. Among them, biological invasions and climate change are of critical 27 concern. It is generally thought that strictly asexual populations will be more susceptible to 28 rapid environmental alterations due to their lack of genetic variability and, thus, of adaptive 29 responses. In this study, we evaluated the persistence of a widely distributed asexual 30 lineage of the alfalfa race of the pea aphid, Acyrthosiphon pisum, along a latitudinal transect 31 of approximately 600 Km in central Chile after facing environmental change for a decade. 32 Based on microsatellite markers, we found an almost total replacement of the original aphid 33 superclone by a new variant. Considering the unprecedented warming that this region has 34 experienced in recent years, we experimentally evaluated the reproductive performance of 35 these two A. pisum lineages at different thermal regimes. The new variant exhibits higher 36 rates of population increase at warmer temperatures, and computer simulations employing 37 a representative temperature dataset suggest that it might competitively displace the original 38 superclone. These results support the idea of a superclone turnover mediated by differential 39 reproductive performance under changing temperatures. 40

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Keywords: aphid superclone, heat tolerance, intrinsic rates of population increase,
 microsatellite

44 **1. INTRODUCTION**

As global environmental conditions change rapidly, they threaten biodiversity significantly, 45 leading to an uncertain future for living systems [1, 2]. While adaptive responses to rapid 46 changes might be anticipated, identifying which species and populations would be capable 47 of adapting and the speed at which they could do so remain challenging in the current 48 context of global change [1 - 5]. Among global change drivers, biological invasions and 49 climate change represent two of the most critical threats to biodiversity [1, 2, 6]. Indeed, 50 biological invasions are now considered one of the "pillars" of biological extinctions [7], being 51 facilitated to some extent by climate change via the rise in environmental temperatures and 52 increased frequency of extreme climatic events [2]. 53

Under the current scenario of accelerated environmental changes, we could 54 anticipate problematic effects of global change on asexual populations due to their lack of 55 intrinsic variability and presumed scarce potential to buffer environmental challenges [8]. At 56 an ecological timescale, migrational, behavioral, or physiological adjustments may facilitate 57 the exposed individuals and populations to cope with the environmental change for a while 58 [9, 10]. In an evolutionary context, there is increasing evidence of evolutionary change 59 without standing genetic variation by epigenetic factors in asexual populations, including 60 asexual aphids [11, 12]. From a classical point of view, though, an increased rate of 61 environmental change would be difficult, at least in principle, for asexual populations to 62 evolve adaptive responses quickly enough [2, 8]. Thus, rapid adaptation would be unlikely 63 or at least unexpected, leaving the extinction of affected populations as a possible 64 evolutionary outcome (but see [3]). 65

⁶⁶ On the one hand, asexual reproduction has large immediate rewards. One of the ⁶⁷ most conspicuous advantages is that asexual individuals can allocate more energy and time ⁶⁸ to reproduction [13 - 15]. Consequently, asexual populations can show a clear demographic ⁶⁹ profit compared to their sexual counterparts [13 - 15]. When it comes to biological invasions,

asexuality can confer advantages to efficiently invade and monopolize new territories
 because only one individual is enough to multiply and start a new population [16]. Indeed,
 asexual lineages arriving and dominating extensive novel areas have been repeatedly
 reported [17 – 20].

On the other hand, in contrast, it has been suggested that the energetic and 74 demographic advantages conferred by asexuality may be offset by the low resilience and 75 evolvability of asexual populations [21 - 23]. Thus, many species that have moved away 76 from strict sexual reproduction, adopting complex life cycles that include asexuality or strict 77 clonality, could lose much of their ability to cope with environmental change, making them 78 highly vulnerable to ecological challenges [14, 24, 25]. Consequently, the ecological (and 79 evolutionary) persistence of an asexual population (or lineage) facing environmental change 80 can be jeopardized by this lack of genetic diversity. 81

An extreme example of asexual populations with a high degree of geographic 82 expansion is the case of aphids (Hemiptera: Aphididae). Aphids are sap-sucking insects 83 characterized by a complex life cycle that alternates sexual and asexual reproduction [26, 84 27]. However, they frequently lose the sexual phase, reproducing by obligate 85 parthenogenesis (*i.e.*, clonal reproduction), especially in novel temperate environments [28, 86 29, 30]. Thus, aphids are excellent invaders if environmental conditions are suitable due to 87 their rapid demographic growth. Interestingly, some aphid clonal lineages may prevail over 88 others in wide invaded ranges due to intraspecific exclusion and are commonly referred to 89 as "superclones" [31]. Superclones can expand their geographical ranges significantly by 90 means of phenotypic plasticity or, in some cases, with the assistance of bacterial secondary 91 endosymbionts (see [32]). However, while superclones are common and their arrival, 92 establishment, and expansion are well-documented (e.g., [23, 33]), the factors that promote 93 their persistence or exclusion in a given territory are poorly understood (but see Peccoud et 94 al. [23]). 95

Aphid populations inhabiting central Chile represent an ideal system to study the 96 response and persistence of an asexual population facing environmental change after local 97 aphid populations are generally characterized by the predominance of a single or few 98 superclones, regardless of the aphid species [23, 33 - 36]. On the one hand, despite the 99 natural geographic barriers of the country [37], several aphid invasion processes have been 100 reported in this range, which has increased in frequency given the growth in trade practices 101 [38]. On the other hand, this extended territory is experiencing an extreme change in its 102 abiotic environmental conditions, which would enable the study of population changes in 103 real time. Indeed, temperatures have shown an increased tendency over the last decade, 104 with summer heat wave events more common due to the pervasive effects of climate change 105 [39 – 44]. 106

In this context, the case of the pea aphid, Acyrthosiphon pisum (Harris), is 107 remarkably interesting: (i) A genetic survey performed between 2005-2007 in central Chile 108 [23] showed the predominance of an obligate asexual multilocus genotype (MLG) in alfalfa 109 (Medicago sativa) crops, namely Ap2, inhabiting a ~600 Km latitudinal range (from 35 to 110 40°S). (ii) For the same species, Sepúlveda et al. [45] found that, within this geographic 111 range, A. pisum specimens collected on the same crop showed differences in the secondary 112 symbionts they carried. These findings suggest the coexistence of different clonal lineages, 113 considering that the main transmission mechanism for the secondary symbionts in aphids is 114 maternal inheritance and that the pea aphid population of central Chile has shown to be 115 strictly asexual [23, 46]. (iii) In a more recent work, Martel et al. [47] found not only one but 116 two different MLGs in the same latitudinal range (APG2 and APG3) but in guite different 117 proportions (~4% and ~96%, respectively). However, aphid MLGs were not comparable with 118 the ones found by Peccoud et al. [23] because of some distinctive microsatellite markers 119 and a different capillary sequencer used for this work (see Materials and Methods section in 120 Martel et al. [47]). 121

Considering this background and the current process of environmental change due 122 to global change across central Chile, the main objective of the present work was to evaluate 123 the response of the clonal population of A. pisum inhabiting central Chile after a decade of 124 facing extreme thermal conditions and the risk of intraspecific invasive events. We compared 125 the current two lineages of A. pisum populations on alfalfa with the dominant lineage found 126 by Peccoud et al. [23] using the same microsatellite markers and protocol. Considering that 127 among the main environmental changes in central Chile over the last decade are the rise in 128 mean summer temperatures and the increased prevalence and magnitude of summer heat 129 waves [39 - 44], we hypothesized that the unbalanced proportions of the two MLG recorded 130 by Martel et al. [47] could be attributable to differences in thermal response between 131 lineages. Additionally, we assessed the presence/absence and the identity of the secondary 132 endosymbionts in these two lineages, considering that some secondary endosymbionts can 133 protect their host aphids from acute heat events [48, 49]. 134

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137 2. MATERIALS AND METHODS

¹³⁸ Sampling and laboratory rearing

In a previous study [47], A. pisum individuals were sampled from alfalfa fields in the spring 139 and summer of 2017-2018 in central Chile, the same region previously surveyed by Peccoud 140 et al. [23] in 2006-2007 (latitude 35 to 40 °S, see [47]). This study showed that there were 141 only two MLGs in this area (named APG2 and APG3 in [47]), with 96% of the genotyped 142 individuals belonging to APG3 and the remaining 4% to APG2. Clonal lineages were 143 established in the laboratory from these two MLG by rearing them separately on broad bean 144 plants (Vicia faba L.) inside 4 L transparent plastic buckets, kept in a climatic chamber at 145 $20^{\circ}C (\pm 0.5^{\circ}C)$ with a long photoperiod (LD = 16:8) to ensure their sustained parthenogenetic 146

reproduction [46, 50]. Five aphids from each lineage were taken before performing the
 experimental procedures on the thermal response (see below), and their DNA was extracted
 to confirm their genetic profile.

150

Genetic profiles and secondary symbionts assessment

Following Peccoud et al. [23], we used fifteen highly polymorphic genetic markers 152 (microsatellites) to identify MLGs, including the seven original markers from [23] and eight 153 additional markers from Nouhaud et al. [51] (Table 1). We extracted total genomic aphid 154 DNA by the "salting-out" method, using proteinase-K digestion and precipitation by ethanol 155 [52]. Then, the DNA was resuspended in ultrapure water and set to a concentration of 30 156 ng/mL. Each locus was amplified separately with fluorescent-labeled primers following the 157 same methodology performed by Peccoud et al. [23]. The electrophoresis of amplified 158 fragments was completed using the same capillary sequencer ABI PRISM 3130xl (Applied 159 Biosystems, U.S.A.), and the same laboratory set-up. MLG characterization was made after 160 the automatic allele calls given by the software GeneMapper (version 3.5 and 3.7, Applera 161 Corp., U.S.A.) and the visual checking of the allele sizes. 162

Besides carrying the obligate endosymbiont bacteria Buchnera aphidicola, aphids 163 can host a range of facultative endosymbionts that can produce diverse phenotypic effects 164 on the aphid host, including heat tolerance [32, 49]. Hence, we screened five facultative 165 bacterial symbionts commonly reported in A. pisum by randomly selecting pea aphid 166 samples (n = 5) from each clonal lineage. Aliquots of DNA extracted from live samples of 167 APG2 and APG3 lineages were used to amplify diagnostic PCR fragments for Spiroplasma 168 (1500 bp), Serratia symbiotica (1140 bp), Regiella insecticola (840 bp), Rickettsia sp. (600 169 bp), Hamiltonella defensa (480 bp), and B. aphidicola (270 bp), the obligate aphid symbiont, 170 used as control [42, 43]. The amplification steps were an initial DNA denaturation at 94 °C 171 for 5 min, 30 cycles of 94 °C for 30 sec, 58 °C for 30 sec, and 72 °C for 1 min, and a final 172

extension step at 72 °C for 5 min. Amplicons were separated through electrophoresis in
1.5% agarose gels, using a GeneRuler 100 bp Plus DNA Ladder (Thermo Fisher Scientific,
Waltham, Massachusetts, USA). The presence of each facultative endosymbiont was
visually identified according to the size of the diagnostic amplicon in base pairs (bp) obtained
from the DNA of each aphid [46].

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179 Temperature and heat-shock effects on life-history traits

To evaluate the performance of individuals from APG2 and APG3 lineages following 180 exposure to short and extreme heat events, we used an already-established protocol used 181 by Montlor et al. [48] and Russell and Moran [53] on the pea aphid using two extreme 182 temperatures registered in central Chile (i.e., 38 and 42 °C; [36-41]). In brief, 40 to 60 age-183 synchronized aphid nymphs of each lineage were separately kept on V. faba seedlings in 184 three separated climatized chambers at 20°C (± 0.5°C, LD = 16:8). After 30 min of 185 acclimation, trials started by increasing the temperature of two chambers to 38°C and 42°C 186 (heat-shock treatments, heating rate = 0.15 and 0.18 °C/min respectively) in the two first 187 chambers respectively, while the third chamber was kept with no temperature change (20°C 188 constant, control). The temperature rose for 2 h for each treatment until the trial temperature 189 was reached. Once reached, this temperature was maintained for 4 h and then lowered to 190 20° C in the next 2 h (total heat challenge time = 8.5 h). After the trials, individuals were 191 returned to their original acclimation temperature (20°C ± 0.5°C, LD = 16:8) for 48 additional 192 hours to allow for recovery and avoid over-manipulation. Subsequently, after checking that 193 all the plants were well, surviving aphids were counted to obtain the proportion of alive/dead 194 individuals and randomly transferred to new V. faba seedlings until they were reproductive 195 adults. Furthermore, we replicated the control treatment at constant rearing temperatures of 196 15 and 25°C ($n \ge 6$ individuals per temperature). 197

Comparisons of reproductive performance were made by using intrinsic growth rates 198 (r_m) , calculated as a derivation from the Euler-Lotka equation, where $r_m = ln (R_0)/t$ (see Birch 199 [54]). Fecundity (R_0) of the survivor nymphs was measured as the number of nymphs 200 produced by each focal individual aphid in a lapse of ten days since the birth of its first nymph 201 since this period generally comprises 95% of the total progeny of an aphid [55]. The 202 developmental time (t), in turn, was measured as the number of days elapsed from the focal 203 individual's birth to the birth of its first nymph. R_0 and t were recorded individually for the 204 maximum number of survivors after the trials for each MLG and heat challenge combination. 205 To summarize, five experimental treatments were used: three constant acclimation 206 temperatures (15, 20, and 25 °C) and two heat-shock temperatures (38 and 42 °C), totaling 207 154 individual measures of r_m , a consistent number given the clonal nature of these aphid 208 lineages (see [56]). 209

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211 Data analysis

Since the MLGs used in this study came from the same region previously surveyed by Peccoud *et al.* [23], a Neighbor-Joining (NJ) tree of MLGs was built with the MLGs found in this study and those found by Peccoud *et al.* [23] using the seven original microsatellite markers to make both studies comparable. NJ tree was constructed based on a distance matrix of allele shared distances (D_{AS}) with bootstrapping of 1000 replicates and computed by the software Populations v1.2.31 [57].

To analyze and compare the intrinsic growth rate of each lineage within their thermal niche, we fitted a linear model after a visual diagnostic of the data using constant acclimation temperature and MLG as independent variables and intrinsic growth rate as dependent variable. Then, predicted values of r_m were used to show how detected differences in r_m between genotypes could be translated into an ecologically relevant scenario. We assessed how daily population growth and cumulative population size are comparatively predicted to

change as a function of mean diurnal temperatures (08:00 - 18:00 h). This time range was 224 selected to capture better the influence of heat-temperature episodes recorded in the field. 225 First, we obtained temperature data at hourly intervals for the Quinta Normal weather station 226 (33°25'S, 70°42'W) in Santiago for the years 2018, 2019, and 2020, sourced from the 227 Chilean Agricultural Research Institute (https://agrometeorologia.cl/, data downloaded on 12 228 May 2022). We chose this locality because the effects of thermal change are representative 229 of the predominant climate in central Chile [44]. This location shows a marked seasonality, 230 and it is a centered spot considering the wide latitudinal distribution of the pea aphid along 231 central Chile. Temperature was converted into Kelvin degrees (adding 273°) and then 232 returned to the Celsius scale to avoid possible confounding effects of below-zero 233 temperatures. Then, predicted r_m based on these thermal conditions were separately 234 converted into daily population growth () for APG2 and APG3, and subsequently into 235 cumulative population size as $N(t) = \prod \Delta N$ over a three-year period (1096 days in total, 236 considering one leap year). For simplicity, calculations assume that initial population sizes 237 were identical for both genotypes and negative r_m predicted from the linear regressions were 238 set to zero (*i.e.*, aphids are assumed not to die due to cold, which is likely realistic during the 239 growth period of aphid populations). On the other hand, to estimate how different MLG 240 responded to the heat-shock treatment, we performed a multiple regression including r_m as 241 the dependent variable, genotype as a categorical factor, and heat-shock temperature as a 242 covariable. 243

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246 **3. RESULTS**

The results of the genetic analysis showed that MLGs Ap2 reported by Peccoud *et al.* [23] and APG2 reported by Martel *et al.* [45] shared the same microsatellite profile at the fifteen evaluated loci (Table 1). On the other hand, the genetic profile of APG3 showed allele

differences at fourteen out of fifteen loci compared to Ap2 and APG2 (Table 1). Thus, 250 considering (i) the genetic identity between Ap2 and APG2 at the fifteen loci evaluated and 251 (ii) that both MLGs are obligate parthenogenetic lineages (see Peccoud et al. [23] and Martel 252 et al. [47]), our results suggest that Ap2 and APG2 belong to the same clonal lineage (Ap2 253 = APG2, Fig. 1A, in blue), hereafter APG2, which was widespread in central Chile before 254 2007 (Fig. 1B, in blue). On the other hand, the NJ tree representing shared allele distances 255 suggests that APG3 is a novel lineage genetically distant from those previously reported by 256 Peccoud et al. [21] (Fig. 1A, in green). Interestingly, the frequency of APG2 has dramatically 257 changed between the two studies in central Chile. While Peccoud et al. [21] showed that 258 98% of the individuals sampled on alfalfa belonged to APG2 (n = 100 out of 102), the 259 proportion of APG2 dropped to ~4% a decade later. The APG3 variant constituted ~96% of 260 the sampled populations (n = 6 and 137 respectively, out of 143), appearing as a newly 261 emerged genotype of A. pisum in central Chile (Martel et al. [47]). Additionally, none of the 262 five facultative symbionts studied were detected in APG2 (n = 5), as opposed to APG3 (n = 263 5) specimens, all of them showing a stable coinfection with *H. defensa* and *Rickettsia sp.* 264 along the experimental procedures (Table 1). 265

Aphid's reproductive performance at different acclimation temperatures differed 266 between genotypes, with APG2 and APG3 displaying higher r_m at cold and warm 267 temperatures, respectively. This impact was significantly different between MLGs (Fig. 2A) 268 based on the interaction between genotype and acclimation temperature in the linear model 269 (F_{1.68} = 28,34, $p = 1.23 \times 10$ -6). The predicted r_m values based on the linear regressions for 270 each MLG were used to test for a clonal lineage turnover in the field using a real temperature 271 dataset (Fig. 3A). In this regard, Fig. 3B depicts the difference in daily population growth of 272 APG3 (in green) relative to APG2 (in blue) when temperatures are warm. While the opposite 273 trend is observed when temperatures are cool, the simulation shows that the population size 274 of APG3 is expected to increase disproportionally compared to APG2 under warm periods 275

(Fig. 3C). Moreover, these differences are expected to become more pronounced under heat stress conditions. Following the 38°C and 42°C heat shock laboratory assays, both MLG lineages exhibited a notorious detrimental effect on relative survival, developmental time (*t*), offspring number (R_0), and, consequently, on intrinsic growth rate (r_m) (Table 2). However, heat tolerance varied significantly between MLGs as APG3 better copes with heat stress (Fig. 2B), as evidence by the significant interaction between MGL and heat-shock temperature in our multiple regression ($F_{1,118} = 6.72$, p = 0.011).

283

4. DISCUSSION

In this study, we report a dramatic turnover between two strictly asexual genotypes of the pea aphid in central Chile over a decade, presumably following a new and recent invasion event (Fig. 1A and 1B). Our results suggest that this turnover might result from the contrasting reproductive performance displayed by both APG2 (original) and APG3 (putatively invasive) lineages experiencing the changing thermal conditions of this region, characterized in part by increasingly common heat-shock events in the last decade (Fig. 2 and Fig. 3).

High genetic diversity is commonly considered advantageous for natural populations 292 because of a higher adaptative capacity than populations with low genetic variation [14, 25, 293 58]. Indeed, hereditary variability combined with sex can allow for rapid evolutionary change 294 and adaptation to changing ecological conditions [59]. It is thought that this variability could 295 promote invasions, as genetically diverse invaders with an enhanced evolutionary potential 296 may successfully establish and spread in the novel range [60, 61]. In contrast, variability in 297 invaded populations could favor resilience to invasions, reducing their extinction risk [22]. 298 Consequently, local asexual populations can be severely threatened by invaders because 299 of low phenotypic variation in some traits (both at the intra- and inter-specific level), 300 potentially leading to competitive exclusion [62]. Thus, the lack of genetic diversity of the 301

pea aphid in central Chile due to its strictly asexual reproductive mode in this area [23, 47] 302 could explain the rapid change in the genetic structure of this population in a decade by the 303 replacement of the original superclone (APG2) by a genetically distant one (APG3; Fig. 1). 304 Importantly, APG3 is quite divergent from the MLGs found by Peccoud et al. [23] on any 305 host plant including alfalfa, and even from any genotype reported by Martel et al. [47] 306 inhabiting the southern region of Chile which could have served as a source of genetic 307 variation (Fig. 1A and [47]). This suggests either the introduction of a new lineage between 308 2007 and 2017 or the clonal amplification of a previously present MLG but at such a low 309 frequency that it was not detectable in previous samples. 310

Despite the differences in reproductive performance between both lineages following 311 the heat challenge (Table 2 and Fig. 2), other changes than thermal conditions can also be 312 involved in the superclone turnover reported here. Multiple factors, such as differences 313 between the two clones in their ability to fend off natural enemies, cope with inter-specific 314 competition, or resist insecticide applications, can be relevant to explain this shift in clonal 315 frequency [63]. However, we have no evidence for rapid changes in the guilds of natural 316 enemies in central Chile [64, 65] nor the type of insecticide treatment [66]. In contrast, there 317 is well-supported evidence for climate change during the period of turn, including extreme 318 summer temperatures in central Chile in recent years, reaching unprecedented records 319 above 42°C in 2017 and 2019 [39 – 44]. Thus, we propose that the pattern observed in the 320 field (Fig. 1B) is more related to different physiological responses to the thermal environment 321 between APG2 and APG3 (Table 2 and Fig. 2). For instance, APG2 has a greater 322 reproductive output at low temperature (*i.e.*, 15 °C) compared to APG3. In contrast, its 323 survival and reproductive output facing acute warming episodes is lower (Figure 2). Thus, 324 APG2 would start reproducing earlier at low temperatures but would be replaced by APG3 325 after heat stress (Fig. 3). In this sense, the heat challenge approach remains to some extent 326 realistic as it estimates the impact of high temperatures in fitness, which mirrors adequately 327

the conditions encountered by aphid populations in nature (Fig. 3). Importantly, these differences in 'fitness' and population growth between MGL are expected to be exacerbated at higher environmental temperatures (*i.e.*, through time with global warming or across warmer localities). Having said that, specific mechanisms giving thermal advantage to APG3 over APG2 after acute extreme heat events remain unknown.

Interestingly, it has been reported that facultative bacterial endosymbionts can confer 333 ecological advantages to the aphid host, such as thermal tolerance by mechanisms aiding 334 the recovery of the obligate symbiont, B. aphidicola, after heat stress [32, 48, 67], leading, 335 for example, to seasonal dynamics of secondary symbionts prevalence [67 - 70]. For 336 instance, in populations of the grain aphid Sitobion avenae, the presence of the 337 endosymbiont S. symbiotica might partly explain recent changes in the composition of 338 superclones observed in Chile (Sepúlveda et al., submitted). However, S. symbiotica was 339 not detected in APG3, and neither H. defensa nor Rickettsia sp., the bacterial 340 endosymbionts detected in APG3 (Table 1), have been unequivocally demonstrated to 341 confer protection to heat stress in previous studies (but see [53]), although this is 342 undoubtedly something worth exploring in future research. In this sense, one limitation of 343 our study was the impossibility of effectively disinfecting facultative symbionts from APG3 344 individuals and, conversely, infecting APG2 with them, something that is not always 345 successful due to biological or technical limitations [see 71]. Disentangling the respective 346 effects of symbiont and aphid genotypes would have provided valuable insights into the 347 mechanism responsible for the observed turnover. 348

Overall, our study suggests that a rapid turnover of *A. pisum* superclone composition occurred in central Chile, likely driven by warmer temperatures, emphasizing that extreme environmental changes can trigger different responses across genotypes that could ultimately end up in competitive exclusion [72]. Notably, the lack of detectable genetic variability in the previous and more recent populations of *A. pisum* strongly supports some

form of competitive displacement, either triggered by environmental changes alone or by the 354 introduction of a new genotype. Short but extreme thermal events, concurrently with 355 increased mean temperatures, could have contributed to the emergence of APG3 due to its 356 increased thermal endurance (Table 2 and Fig. 2), leading to a competitive advantage 357 against APG2 for the same ecological niche (Fig. 3). This phenomenon is comparable to 358 some extent with that reported by Brévault et al. [73] for the cotton aphid A. gossypii, in 359 which only one of the two overrepresented genotypes was prevalent (>90%) in insecticide 360 sprayed cotton producing regions of the west and central Africa, despite the equivalent or 361 even higher reproductive performance of a second clone on non-sprayed fields. Overall, our 362 results show that asexual populations can thrive over large areas under favorable conditions 363 on ecological timescales and may even be able to adapt to environmental changes [11, 12, 364 74, 75]. However, asexuality and clonal reproduction suffer from a lack of adaptability on an 365 evolutionary scale, potentially leading to the extinction of entire asexual populations and the 366 collapse of their genetic heritage. 367

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382 Author's contributions SIM, JCS, CCF, ELR, and FB conceived the ideas and designed

methodology; SIM, CAZ, JCS, and FM collected the data; SIM, CAZ, NR, DAS, JCS, CCF

- and ELR analyzed the data; SIM wrote the first draft and led the writing of the manuscript.
- All authors contributed to subsequent drafts and gave their final approval for publication.

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Table 1. Allele combinations at fifteen microsatellite loci, including seven reported by Peccoud *et al.* [21] (in bold) and eight reported by
 Nouhaud et al. [49], for the clonal lineages of the pea aphid found in central Chile. The composition of facultative bacterial symbionts
 detected in alive individuals from APG2 and APG3 are also shown.

Lineage (MLG)	Study	Locus ID															
		AIA09M	AIB07M	AIB08M	AIB12M	ApF08M	ApH08M	ApH10M	111775	114782	117294	118775	118777	126104	127526	128001	Facultative symbionts
Ap2	Peccoud et al. 2008	308/310	117/153	276/284	312/318	163/175	246/248	196/198	275/302	301/303	372/376	219/219	345/355	282/287	406/416	203/207	-
APG2	Martel <i>et</i> <i>al.</i> 2020	308/310	117/153	276/284	312/318	163/175	246/248	196/198	275/302	301/303	372/376	219/219	345/355	282/287	406/416	203/207	None
APG3	Martel <i>et</i> <i>al.</i> 2020	306/308	117/133	258/272	304/330	165/171	248/266	184/194	275/281	295/301	378/382	219/219	343/351	281/292	408/410	198/200	H. defensa/ Rickettsia

Table 2. Raw mean data for relative survival, generational time in days (*t*), fecundity to the 10th day (R_0), and intrinsic growth rate (r_m) for aphids of the APG2 and APG3 lineages following the temperature shock assays. Values are shown as mean ± SE.

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MLG	Replicates	Temperature (⁰C)	Relative survival	<i>t</i> (days)	<i>R₀</i> (# nymphs)	r _m (days⁻¹)
APG2	6	15	1	14.8 ± 0.2	104.3 ± 1.7	0.314 ± 0.004
	20	20	1	9.7 ± 0.1	96.4 ± 1.2	0.474 ± 0.006
	10	25	1	$\textbf{6.5} \pm \textbf{0.2}$	69.5 ± 3.6	0.655 ± 0.021
	22	38	0.77	10.5 ± 0.4	45.0 ± 8.6	0.283 ± 0.042
	16	42	0.38	11.8 ± 0.5	$\textbf{4.2}\pm\textbf{3.2}$	0.043 ± 0.024
APG3	6	15	1	15.7 ± 0.8	52.0 ± 4.5	0.255 ± 0.018
	20	20	1	9.2 ± 0.1	93.0 ± 2.1	0.491 ± 0.006
	10	25	1	6.1 ± 0.1	90.1 ± 4.1	$\textbf{0.738} \pm \textbf{0.013}$
	25	38	1	10.5 ± 0.2	66.1 ± 5.0	0.399 ± 0.015
	19	42	0.84	11.8 ± 0.6	$\textbf{22.0} \pm \textbf{5.0}$	0.223 ± 0.034

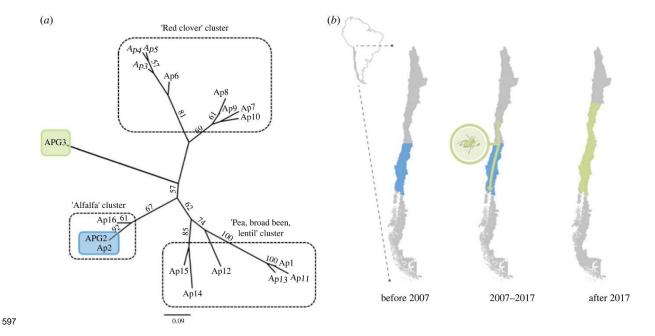


Figure 1. (A) Neighbor-Joining tree using allele shared distances for seven microsatellite loci [23] combining the MLGs studied in this work (APG2 and APG3) and the 16 MLGs (Ap1 – Ap16) for all host species found in central Chile by Peccoud *et al.* [23]. Dashed boxes indicate MLG clusters relying on genetic differences and related to host species [23]. Colors indicate the original (APG2, blue) and the putative invasive (APG3, green) alfalfa lineages. Numbers along branches represent the bootstrap percentage of the supporting nodes (shown > 50 %). (B) Hypothetical model for the invasion of APG3 in central Chile.

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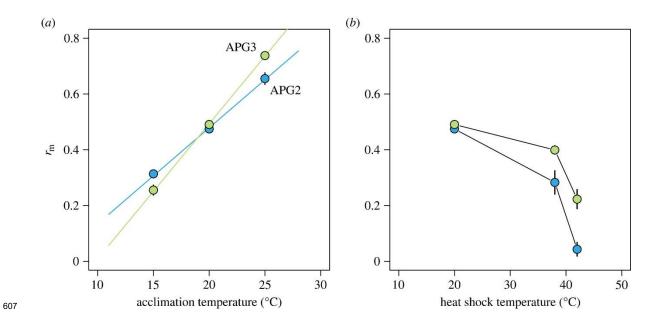
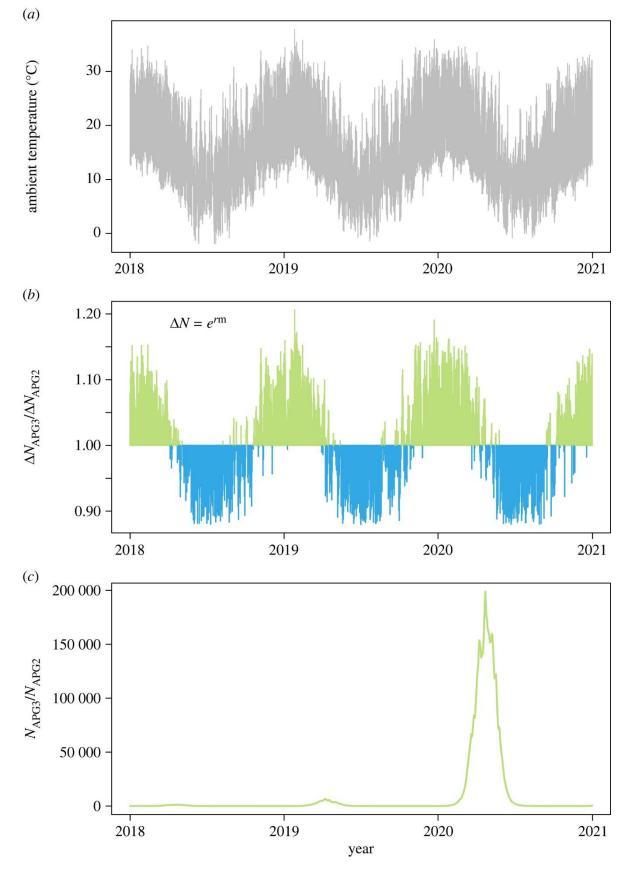


Figure 2. Mean intrinsic growth rate (r_m) for both MLG lineages used in this study (APG3 in green and APG2 in blue) by experimental settings: (A) constant acclimation (15, 20, and 25 °C), and (B) heat-shock (38 and 42 °C). Values are mean ± SE. In panel A, lines represent linear regressions performed separately for each genotype ($R^2 > 0.89$), used to simulate comparatively the population growth of each MLG under field conditions (see Figure 3). Panel B shows the decline in r_m after experimental heat-shock procedures at 38 and 42° (see Materials and Methods section).

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- ⁶¹⁶ **Figure 3.** Comparative population responses to ambient temperature between APG3 (in green) and
- APG2 (in blue) using r_m predictions from linear regressions (Fig. 2) and real field temperatures from
- central Chile (see Materials and Methods section). (A) Hourly ambient temperatures of Quinta Normal
- (33°25'S, 70°42'W) for a three-year period. (B) Population growth ratio (daily) between APG3 and
- APG2. (C) Cumulative population size ratio between APG3 and APG2.