

# **Interplay of competition and facilitation in grazing succession by migrant Serengeti herbivores**

T. Michael Anderson, Staci Hepler, Ricardo Holdo, Jason Donaldson, Robert Erhardt, J. Grant C. Hopcraft, Matthew Hutchinson, Sarah Huebner, Thomas Morrison, Jeffry Muday, et al.

## **To cite this version:**

T. Michael Anderson, Staci Hepler, Ricardo Holdo, Jason Donaldson, Robert Erhardt, et al.. Interplay of competition and facilitation in grazing succession by migrant Serengeti herbivores. Science, 2024, 383 (6684), pp.782-788. 10.1126/science.adg0744. hal-04461495

# **HAL Id: hal-04461495 <https://hal.science/hal-04461495v1>**

Submitted on 23 Feb 2024

**HAL** is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

## **Title: Interplay of competition and facilitation in grazing succession by migrant Serengeti herbivores**



**Abstract:** Competition, facilitation, and predation offer alternative explanations for successional patterns of migratory herbivores. However, these interactions are difficult to measure, leaving uncertainty about the mechanisms underlying body-size dependent grazing—and even whether 30 succession occurs at all. We used data from an eight-year camera-trap survey, GPS-collared herbivores, and fecal DNA metabarcoding to analyze the timing, arrival order, and interactions among migratory grazers in Serengeti National Park. Temporal grazing succession is characterized by a "push-pull" dynamic: competitive grazing nudges zebra ahead of co-migrating wildebeest, while grass consumption by these large-bodied migrants attracts trailing, small-35 bodied gazelles that benefit from facilitation. "Natural experiments" involving intense wildfires and rainfall respectively disrupted and strengthened these effects. Our results highlight a balance between facilitative and competitive forces in co-regulating large-scale ungulate migrations.

**One-Sentence Summary:** Movement, occupancy, and diet data show how biotic interactions 40 and abiotic factors influence multi-species mass-migration patterns.

**Main Text:** Seasonal migrations, defining features of many terrestrial and marine ecosystems worldwide (*1, 2*), are threatened by habitat destruction, overhunting, and climate change (*3*). Although migration is common in large mammalian herbivore species (*4*), the mechanisms underlying multi-species migration dynamics remain poorly understood (*2*). The annual 5 ungulate migration in Serengeti National Park is the archetypal example of body-size dependent "grazing succession," (*5*), where zebra (*Equus quagga*; ~230 kg), wildebeest (*Connochaetes taurinus*; ~180 kg), and Thomson's gazelle (*Eudorcas thomsonii*; ~20 kg; 'gazelle' hereafter) sequentially follow the same migratory routes.

- 10 Early descriptions of grazing succession invoked facilitation to explain this pattern (6). In this scenario, bulk consumption by larger grazers keeps grasses short and regrowing, creating nutritious forage for smaller grazers that require less food (*7*). Removal of grass biomass may also increase accessibility and abundance of low-growing, high-quality herbaceous eudicots favored by smaller species (*8, 9*). In Serengeti, evidence for facilitation is based largely on (i) 15 staggered arrival times of zebra, wildebeest, and gazelles, as measured at a single location over three years (*5*), (ii) partitioning of grass stem, leaf, and sheath observed in a small sample of stomach contents (*9*), and (iii) gazelle preference for habitat patches previously grazed by wildebeest (*10*).
- 20 Other lines of evidence suggest that grazing succession is instead driven by competition (*11*). An indirect competitive process (exploitation; Table S1) may operate if multiple species use the same resource patches, but smaller grazers delay their arrival due to resource depletion by larger grazers. A direct competitive process (interference; Table S1) may operate if species comingle and smaller herbivores propel larger herbivores onwards into taller vegetation by grazing swards 25 down to heights unprofitable for the larger species (*12, 13*). Citing "invasion" of wildebeest during their northernmost path, Sinclair (*14*) proposed that interspecific competition forced zebra to stay ahead of migrating wildebeest. A third possibility is that predators regulate migration dynamics. Mixed-species migration could reduce individuals' risk through increased vigilance and predator satiation (*14, 15*), especially in areas of high vulnerability to ambush predators (*16*).

30

Disentangling the potential mechanisms that drive the movement patterns, arrival times, and foraging behaviors of the three migrating species has been hampered by a lack of long-term, detailed data at the appropriate scale (*2*). We conducted a landscape-scale analysis of animal movements in Serengeti by integrating three spatially and temporally overlapping datasets: (i) 35 the citizen-science-supported Snapshot Serengeti camera trap survey ( $N_{\text{cameras}} = 48 - 130$ ; coverage =  $1000 \text{ km}^2$ ; Fig. S1) that operated continuously between  $2011 - 2018$  (17; Fig. 1), (ii) GPS collar data collected in 2017 and 2018 for migratory zebra and wildebeest (*18*; Fig. 1), and (iii) dietary analysis of fecal DNA for each migrant species collected in 2017 and 2018 (*19*). Our study area, in the fire-prone grassland-to-woodland transition north of the Serengeti plains, lies 40 in the center of the Serengeti migration (Fig. 1). We focused on the 'transition season' (April – August), when migrants move off the plains to the north-west and plants undergo dry season senescence (*20*).

The biotic interactions influencing grazing succession are modulated by disturbance. Rainfall 45 and fire may decouple associations between migratory species by controlling food availability through biomass reduction (fire) or stimulating plant growth (both) (*21*). The extent of fire within the study area varied 33-fold across years (range:  $21.3 - 694.4 \text{ km}^2$ ; Fig. S2), rainfall in the four months prior to the transition period varied three-fold  $(217 - 621$  mm; Fig. S2), and

vegetation greenness (normalized difference vegetation index; *NDVI*) varied over ten-fold (1.9 – 29.0; Supplementary Materials). We exploited these extremes to probe the mechanistic underpinnings of co-migration and test three alternative hypotheses: facilitation (H1), competition (H2), and predation (H3) (Table S1) that make contrasting predictions about the 5 order, timing, and strength of species associations.

The facilitation hypothesis predicts consistent separation in migrant arrival time: zebra followed by wildebeest then gazelles (*5*). Grazing by larger herbivores is expected to attract smaller herbivores, but not the reverse. Digestible energy per unit mass decreases with grass biomass 10 (*22*); by creating patches of high-quality regrowth (*23*), fire should disrupt grazing succession and uncouple any positive associations between species because regrowth will not be restricted to previously grazed areas. In contrast, high pre-migration rainfall promotes accumulation of stem-dominated grass biomass and should strengthen species associations by enhancing facilitation (otherwise, small herbivores would constantly encounter tall, stem-dominated swards 15 with low digestible energy). Finally, diets should be dominated by grasses with low species partitioning between each migratory herbivore (*5, 9 – 10*). As hind-gut fermenters (*24*), zebra can digest low-quality food and maximize intake from tall grass, while smaller ruminant species select nutritious short grasses.

20 The competition hypothesis predicts the same species arrival order and high dietary overlap as facilitation, but timing and species associations should depend on the form of competition. Under indirect competition (H2*i*), species should arrive several weeks apart (to allow for grass regrowth), and the lag should be greater after longer occupancy by the first-arriving species (Table S1). Under direct competition (H2*d*) we expected little to no separation in arrival time, 25 with larger migrants avoiding dense herds of smaller migrants that deplete food availability (Table S1). In both scenarios, higher pre-migration rainfall should weaken negative associations between migrants of similar size by reducing feeding competition, whereas pre-migration fire should intensify negative associations.

30 The predation hypothesis predicts closer spatial proximity especially for small-bodied species, as they are vulnerable to a wider range of carnivore species (*25*). Positive spatial associations should be stronger in open grasslands in wet years, when taller grasses provide cover for ambush predators, and in areas with high-risk landscape features such as river courses (based on a 50 year spatial dataset of lion predation risk; *26*) regardless of rainfall or fire. We rely on lion 35 predation as an overall proxy for ambush risk as cheetahs select the same small-scale areas as lions (*27*) and leopards' habitat preferences are highly congruent with lions' (*28*). In contrast, spotted hyenas and African wild dogs make no attempt at concealment and first disturb the herds to identify vulnerable individuals (*29*, *30*). Hence, we consider it unlikely that risks from coursing predators would confer strong advantages either from forming mixed-species herds or 40 from migrating separately.

## **RESULTS**

## *Arrival order and time lags*

The classic formulation of grazing succession  $(5)$  predicts consistent  $\sim$ 2-month separations in 45 peak abundance of each species (Fig. 1C). However, dates of migration and peak abundance for zebra and wildebeest varied across years and did not exhibit separation consistent with hypotheses based on either facilitation (H1) or indirect competition (H2*i*). In most years, zebra and wildebeest entered the study area simultaneously. Wildebeest numbers accumulated rapidly

and reached peak abundance  $7.3 \pm 9.3$  d (range: 22 to -6 d) before zebra, although this difference was inconsistent (Wilcoxon signed-rank test:  $v = 3.5$ ,  $P = 0.09$ , df = 6). In contrast, wildebeest and zebra abundance both consistently peaked  $27.6 \pm 15.4$  d before gazelle (Wilcoxon sign rank test:  $v = 105$ ,  $P < 0.001$ ,  $df = 12$ ). These time lags are smaller than those reported by Bell (5), and 5 only the separation between the larger migrants and gazelle is qualitatively consistent with classic predictions (Fig. 1). Furthermore, the observed lags are inconsistent with direct competition (H2*d*) or predation (H3), which predict that smaller herbivores should maintain adjacency to larger herbivores either because their foraging drives large herbivores to sites with more food (*12–14*) or because they benefit most from increased predator vigilance when 10 foraging in mixed-species herds (*14, 15*).

#### *Directionality and reciprocity of interactions*

Facilitation (H1) predicts positive effects of larger herbivores on the probability of occupancy by smaller herbivores, with negligible reciprocal effects of small on large herbivores (Fig. 1). Using 15 two-step, conditional Bayesian occupancy models (*31*, *32*), we analyzed four-day and sixteenday temporal windows to quantify concurrent and lagged associations between each pair of species for each transition season. Each model included static (i.e., landscape) and dynamic (i.e., rainfall and vegetation) environmental predictors as covariates to control for potentially confounding factors (*33*) and included a unique species-specific conditional term 20  $(y_{species 2|species 1})$  that measured the effect of one species on the occupancy probability of another.

Over four-day windows, we found reciprocal positive effects of zebra and wildebeest occupancy throughout the migration. Zebra increased wildebeest occupancy in 6 of 8 years (with marginally 25 positive effects in the other 2), and wildebeest increased zebra occupancy in 3 of 8 years (and marginally in another) (Fig. 1G; Fig. S3; Table S2). On this four-day scale, gazelles were not firmly associated with zebra or wildebeest occupancy in any year (Table S2), exhibiting only marginal negative overall associations reflecting the temporal separation between the arrival time of gazelles and larger herbivores (Fig. 1F, 1G).

30

At sixteen-day intervals, and consistent with the facilitation hypothesis, we detected positive effects of larger grazers on gazelles in 5 of 8 years, and gazelles were more strongly associated with locations previously occupied by wildebeest than by zebra (Fig. 1H; Fig. S3; Table S3). Zebra–wildebeest associations on this longer timescale were consistently negative, reflecting 35 their positive associations on shorter time scales (Fig. 1F, 1H).

## *Influences of fire and rainfall on species' associations*

Exceptions to the prevailing patterns were related to environmental extremes. The two years without four-day conditional effects of zebra and wildebeest also showed the most extensive 40 fires (2013 and 2016). Further, 2 of the 3 years without sixteen-day lagged effects of larger species on gazelles had the lowest rainfall in the entire study period (2014 and 2017; Fig. S3; Table S2), suggesting that species associations were resource-dependent.

To test the effect of environmental conditions, we analyzed the relative strength of association 45 between species (*RSAi*) as functions of fire and rainfall from January–June each year (Fig. S3). We predicted that *RSA<sub>i</sub>* should be negatively associated with fire and positively associated with pre-migration rainfall. Indeed, *RSAi* in four-day intervals declined with area burned (Fig. 2;

Table S4). This correlation was directionally consistent but strongest for zebra on wildebeest and for wildebeest on gazelles—size-adjacent species that precede each other in the migration likely because fire mimics large herbivores by reducing grass height and promoting regrowth in ways that benefit small herbivores (*23*). *RSAi* in sixteen-day intervals increased with rainfall for 5 gazelles trailing wildebeest and zebra (Fig. 2). Thus, gazelles more closely tracked the larger migrants in years with higher herbaceous production, suggesting greater reliance on the facilitative effects of biomass removal (H1).

*RSAi* for zebra and wildebeest occupancies did not vary in response to rainfall at either timescale, 10 despite their reciprocal spatial associations (Fig. 2). For additional insight, we examined paths of GPS-collared individuals for one dry (2017) and one wet (2018) year to test whether movements measured at four-hour intervals revealed subtler responses to heterospecifics. Wildebeest habitat selection was independent of proximity to dense herds of zebra (Fig. 3A). In contrast, zebra selected habitat distant from dense wildebeest herds in both years, and farther in the dry (1.4 km) 15 than the wet (0.8 km) year (Fig. 3B, Table S5). Wildebeest and zebra step lengths and turn angles also depended on rainfall and distance to heterospecifics. During the leaner dry year, wildebeest used larger step lengths of ~400 m as they searched for food, especially when near dense zebra herds; in the more plentiful wet year, wildebeest used smaller step lengths indicative of active foraging (Fig. 3; Table S5). Zebra step lengths in the dry year were larger than the null 20 expectation. However, in the wet year, zebra migrating close to dense wildebeest herds had smaller step lengths and more circuitous turn angles, indicating greater local food availability (Fig. 3; Table S5). This last inference is also supported by evidence that both species avoided senescing vegetation more strongly in the wet than the dry year (scaled change in NDVI from Table S5). Altogether, these fine-scale movements are most consistent with asymmetric direct 25 competition (H2*d*) between the two species: wildebeest habitat selection is governed by resource availability independent of zebra, while zebra are subject to interference competition and locally repelled from massive wildebeest herds, especially in dry years (*14*).

#### *No evidence for a role of predation risk*

30 We tested whether spatial associations were stronger in risky areas (H3) by analyzing conditional occupancy probabilities relative to the distribution of ambush-predation risk inferred from lion densities and kills (Fig. S4; *34*). Conditional species associations were indistinguishable between high- and low-risk areas at both four-day and sixteen-day timescales (figs. S5–S12). The results are also incongruent with the predictions from H3 that associations in high-risk areas would be 35 stronger in years with higher rainfall and more pronounced for associations involving gazelle.

#### *Diet selection and overlap*

We used fecal DNA metabarcoding in 2017 and 2018 to identify plant taxa and estimate their relative abundance in herbivore diets (*19, 35*; Supplemental Materials). Zebra diets consisted 40 almost exclusively of grasses (96.9% of sequence reads on average); wildebeest diets were grassdominated (73.9%) but included substantial proportions of herbaceous eudicots, mostly legumes (Fig. 4A, B; Table S6). In contrast, grasses were a minority of gazelle diets (26.2%), especially in earlier months (May–July) before the larger species departed from the study area (Fig. 4A, B). In both years, zebra and wildebeest ate highly overlapping sets of plant taxa (mean niche overlap 45 0.82), while overlap was lower for wildebeest and gazelles (0.53) and lowest for zebra and gazelles (0.34). Wildebeest and gazelles avoided common tall grasses of middling to low quality, whereas zebra showed higher selectivity for these taxa (Fig. 4C). In later months, however, gazelles increased consumption of high-quality grasses such as *Digitaria macroblephara* that

were likely regrowing after being grazed earlier in the season (figs. S14, S15; Supplemental Materials).

Accounts that grazers partition grass stems and leaves without reference to plant identity (*9*) 5 were integral to the classic grazing-succession model (*5*) and later elaborations (*7*, *11–13*, *36*). The generally high overlap in plant taxa eaten by zebra and wildebeest is compatible with this mechanism, and with both facilitative and competitive interpretations, although the differential selectivity for grass species (Fig. 4C) aligns with the fine-scale competitive displacement inferred from movement data (Fig. 3B). However, we also find new support for a facilitative 10 mechanism first proposed in 1968 (*9*) whereby larger migrants benefit gazelles by opening the grass canopy and providing access to nutrient-rich prostrate eudicots (Fig. S15), at least until the late dry season when gazelles switch to high-quality grasses grazed by earlier migrants (Fig. 4B; Fig. S14). Coupled with the observed lags in occupancy, these dietary differences are incompatible with the idea that gazelles competitively displace larger migrants (*13*) and may also 15 help explain the gazelles' stable population size following the rapid increase in wildebeest numbers following rinderpest eradication in the 1960s (*36*).

*Discussion* – Our findings provide insight into the drivers of the iconic Serengeti migration. We found no evidence that the grazing succession is driven by predation risk (H3). Risk from lions, 20 leopards, and cheetahs is especially high in the northwestern region of our study area (Fig. S4; *38*), and if risk mitigation were a key reason for co-migration, strong species associations should have persisted near high-risk landscape features (*26*) even after fire. Previous analyses had suggested that high herbaceous biomass promotes mixed species group formation to reduce predation risk (*16*); however, the low prevalence of multi-species associations in camera-trap 25 photos (only 1.1% of gazelle detections, 5.7% of wildebeest detections, and 4.6% of zebra detections included any of the other species), does not support a general anti-predator response.

Instead, our analyses of occupancy, animal movement, and diet provide strong evidence that foraging processes underlie grazing succession, occurring at short intervals (< four-day) between 30 zebra and wildebeest and longer intervals (sixteen- to thirty-two-day) between larger migrants and gazelles. Multiple lines of evidence point to a "push-and-pull" dynamic in which wildebeest, the dominant grazer, push zebra ahead by reducing grass biomass (H2*d*) while pulling along gazelles via facilitation (H1). Our study helps to reconcile decades of conflicting results (*5, 6, 9- 14*) by demonstrating that competition and facilitation operate concurrently during animal 35 migration with effect sizes that depend on resource availability (*7*).

A similar seasonal shift between competition and facilitation was observed in Kenyan rangeland, where interactions between cattle and wild grazers shifted from facilitation during the wet season to competition in the dry season (*37*). In our study, smaller migratory herbivores benefitted both 40 during both dry and wet periods (*20*), when they selected regrowing high-quality grasses previously consumed by the preceding larger-bodied migrants (Fig. 4C; Fig. S13) and/or gained improved access to protein- and mineral-rich legumes and other forbs (Fig. S14, Fig. S15). In a typical year, habitat use by gazelles tracks co-migrating zebra and wildebeest migration, and this association intensifies with increasing rainfall. In high-fire years, short-term movements among 45 grazers are largely uncoupled, indicating that gazelles benefit from grass removal and regrowth, regardless of whether it results from migratory herbivores or fire. We do not claim that effects of zebra on wildebeest or of both species on gazelles are purely facilitative, but rather that facilitation is an important process during herbivore migration, likely helping to stabilize

populations of smaller herbivores (*22*). Herbivore population dynamics are not strongly coupled at the ecosystem level (*11*), which implies that extrinsic factors are important, but facilitation may still be a key stabilizing force and driver of individual behavior during critical periods of the annual migration.

5

Multi-species migratory systems were once common across tropical savannas (*3, 39*) and probably involved similarly nuanced interplay of positive and negative interactions (*40*). Large herbivores have suffered disproportionate extinction since the Quaternary (*41*), especially among migratory species (*4*), resulting in concomitant degradation of ecosystem functions (*42*). Our 10 findings identify a previously unappreciated role of plant functional diversity in sustaining megafaunal migratory systems and indicate that fire management can help to maintain species persistence. Any disruption of the vegetative phenology that underlies herbivore movements would alter the strong systemwide consequences of migration (*43*) and portends additional challenges associated with extreme climate and weather events in coming years (*44*).

#### **References and Notes:**

- 1. N. Putman, Marine migrations. *Current Biology* **28**, R972-R976 (2018).
- 2. M. J. Kauffman, E. O. Aikens, S. Esmaeili, P. Kaczensky, A. Middleton, K. L. Monteith, T. A. Morrison, T. Mueller, H. Sawyer, J. R. Goheen, Causes, consequences, and conservation 20 of ungulate migration. *Annual Review of Ecology, Evolution and Systematics* **52**, 453-478 (2021).
	- 3. D. S. Wilcove, M. Wikelski, Going, going, gone: Is animal migration disappearing. *PLoS Biology* **6**, e188 (2008).
- 4. J. O. Abraham, N. S. Upham, A. Damian-Serrano, B. R. Jesmer, Evolutionary causes and 25 consequences of ungulate migration. *Nature Ecology & Evolution* **6**, 998-1006 (2022).
	- 5. R. H. Bell, A grazing ecosystem in the Serengeti. *Scientific American* **225**, 86-93 (1971).
	- 6. D. F. Vesey-Fitzgerald, Grazing succession among East African game animals. *Journal of Mammalogy* **41**, 161-172 (1960).
- 7. R. Arsenault, N. Owen‐Smith, Facilitation versus competition in grazing herbivore 30 assemblages. *Oikos* **97,** 313-318 (2002).
	- 8. A. M. Louthan, D. F. Doak, J. R. Goheen, T. M. Palmer, R. M. Pringle, Mechanisms of plant–plant interactions: concealment from herbivores is more important than abiotic-stress mediation in an African savannah. *Proceedings of the Royal Society B: Biological Sciences* **281**, 20132647 (2014).
- 35 9. M. D. Gwynne, R. H. V. Bell, Selection of vegetation components by grazing ungulates in the Serengeti National Park. *Nature* **220**, 390-393 (1968).
	- 10. S. J. McNaughton, Serengeti migratory wildebeest: facilitation of energy flow by grazing. *Science* **191**, 92-94 (1976).
- 11. A. R. E. Sinclair, M. Norton-Griffiths, Does competition or facilitation regulate migrant 40 ungulate populations in the Serengeti? A test of hypotheses. *Oecologia* **53**, 364-369 (1982).
	- 12. A. W. Illius, I. J. Gordon, The allometry of food intake in grazing ruminants. *The Journal of Animal Ecology* **56**, 989-999 (1987).
	- 13. J. T. du Toit, H. Olff, Generalities in grazing and browsing ecology: using across-guild comparisons to control contingencies. *Oecologia* **174**, 1075-1083 (2014).
- 45 14. A. R. E. Sinclair, Does interspecific competition or predation shape the African ungulate community? *The Journal of Animal Ecology* **54**, 899-918 (1985).
	- 15. K. Stears, M. H. Schmitt, C. C. Wilmers, A. M. Shrader, Mixed-species herding levels the landscape of fear. *Proceedings of the Royal Society B* **287**, 20192555 (2020).

7

- 16. L. Beaudrot, L., M. S. Palmer, T. M. Anderson, C. Packer, C., Mixed‐species groups of Serengeti grazers: a test of the stress gradient hypothesis. *Ecology* **101**, e03163 (2020).
- 17. A. Swanson, M. Kosmala, C. Lintott, R. Simpson, A. Smith, C. Packer, Snapshot Serengeti, high-frequency annotated camera trap images of 40 mammalian species in an African 5 savanna. *Scientific Data* **2**, 1-14 (2015).
	- 18. J. G. C. Hopcraft, J. M. Morales, H. L. Beyer, M. Borner, E. Mwangomo, A. R. E. Sinclair, A. R. E., H. Olff, D. T. Haydon, Competition, predation, and migration: individual choice patterns of Serengeti migrants captured by hierarchical models. *Ecological Monographs* **84**, 355-372 (2014).
- 10 19. T. R. Kartzinel, P. A. Chen, T. C. Coverdale, D. L. Erickson, W. J. Kress, M. L. Kuzmina, D. I. Rubenstein, W. Wang, R. M. Pringle, DNA metabarcoding illuminates dietary niche partitioning by African large herbivores. *Proceedings of the National Academy of Sciences* **112**, 8019-8024 (2015).
- 20. S. J. McNaughton, Grazing as an optimization process: grass-ungulate relationships in the 15 Serengeti. *The American Naturalist* **113**, 691-703 (1979).
	- 21. W. J. Bond, J. E. Keeley, Fire as a global 'herbivore': the ecology and evolution of flammable ecosystems. *Trends in Ecology & Evolution* **20**, 387-394 (2005).
- 22. J. M. Fryxell, J. F. Wilmshurst, A. R. Sinclair, D. T. Haydon, R. D. Holt, P. A. Abrams, Landscape scale, heterogeneity, and the viability of Serengeti grazers. *Ecology Letters* **8**, 20 328-335 (2005).
	- 23. J. E. Donaldson, S. Archibald, N. Govender, D. Pollard, Z. Luhdo, C. L. Parr, Ecological engineering through fire-herbivory feedbacks drives the formation of savanna grazing lawns. *Journal of Applied Ecology* **55**, 225-235 (2018).
- 24. P. J. van Soest, Allometry and ecology of feeding behavior and digestive capacity in 25 herbivores: a review. *Zoo Biology* **5**, 455-479 (1996).
	- 25. A. R. Sinclair, S. Mduma, J. S. Brashares, Patterns of predation in a diverse predator–prey system. *Nature* **425**, 288-290 (2003).
	- 26. J. G. C. Hopcraft, A. R. Sinclair, C. Packer, Planning for success: Serengeti lions seek prey accessibility rather than abundance. *Journal of Animal Ecology* **74**, 559-566 (2005).
- 30 27. A. Swanson, T. Caro, H. Davies‐Mostert, M. G. Mills, D. W. Macdonald, M. Borner, E. Masenga, C. Packer, Cheetahs and wild dogs show contrasting patterns of suppression by lions. *Journal of Animal Ecology* **83**, 1418-1427 (2014).
- 28. P. Strampelli, P. Henschel, C. E. Searle, D. W. Macdonald, A. J. Dickman, Spatial cooccurrence patterns of sympatric large carnivores in a multi-use African system. *PloS One* 35 **18**, p.e0280420 (2023).
	- 29. H. Kruuk, Surplus killing by carnivores. *Journal of Zoology* **166**, 233-244 (1972).
	- 30. C. D. Fitzgibbon, J. H. Fanshawe, Stotting in Thomson's gazelles: an honest signal of condition. *Behavioral Ecology and Sociobiology* **23**, 69-74 (1988).
- 31. L. Sun, M. K. Clayton, Bayesian analysis of cross classified spatial data with autocorrelation. 40 *Biometrics* **64**, 74-84 (2008).
	- 32. S. A. Hepler, R. Erhardt, T. M. Anderson, Identifying drivers of spatial variation in occupancy with limited replication camera trap data. *Ecology* **99**, 2152-2158 (2018).
- 33. P. C. Caragea, E. Berg, A centered bivariate spatial regression model for binary data with an application to presettlement vegetation data in the Midwestern United States. *Journal of*  45 *Agricultural, Biological, and Environmental Statistics* **19**, 451-469 (2014).
	- 34. M. S. Palmer, J. Fieberg, A. Swanson, M. Kosmala, C. Packer, A 'dynamic' landscape of fear: prey responses to spatiotemporal variations in predation risk across the lunar cycle. *Ecology Letters* **20**, 1364-1373 (2017).

35. J. Pansu, M. C. Hutchinson, T. M. Anderson, M. te Beest, C. M. Begg, K. S. Begg, A. Bonin, L. Chama, S. Chamaillé-Jammes, E. Coissac, J. P. G. M. Cromsigt, M. Y. Demmel, J. E. Donaldson, J. A. Guyton, C. B. Hansen, C. I. Imakando, A. Iqbal, D. F. Kalima, G. I. H. Kerley, S. Kurukura, M. Landman, R. A. Long, I. N. Munuo, C. M. Nutter, C. L. Parr, A. B. 5 Potter, S. Siachoono, P. Taberlet, E. Waiti, T. R. Kartzinel, R. M. Pringle, The generality of cryptic dietary niche differences in diverse large-herbivore assemblages. *Proceedings of the National Academy of Sciences*. **119**, e2204400119 (2022).

36. J. M. Fryxell, P. Abrams, R. D. Holt, J. F. Wilmshurst, A. R. E. Sinclair, R. Hilborn, "Spatial dynamics and coexistence of the Serengeti grazer community" in *Serengeti III: Human*  10 *Impacts on Ecosystem Dynamics* (University of Chicago Press, Chicago, 2008), pp. 277-300.

- 37. W. O. Odadi, M. K. Karachi, S. A. Abdulrazak, T. P. Young. African wild ungulates compete with or facilitate cattle depending on season. *Science* **333**, 1753-1755 (2011).
- 38. A. Mosser, J. M. Fryxell, L. Eberly, C. Packer, Serengeti real estate: density vs. fitness‐based indicators of lion habitat quality. *Ecology Letters* **12**, 1050-1060 (2009).
- 15 39. N. Owen‐Smith, G. Hopcraft, T. Morrison, S. Chamaillé‐Jammes, R. Hetem, E. Bennitt, F. van Langevelde, Movement ecology of large herbivores in African savannas: current knowledge and gaps. *Mammal Review* **50**, 252–266 (2020).

40. G. Harris, S. Thirgood, J. G. C. Hopcraft, J. P. Cromsigt, J. Berger, Global decline in aggregated migrations of large terrestrial mammals*. Endangered Species Research* **7**, 55-76 20 (2009).

- 41. F. A. Smith, R. E. E. Smith, S. K. Lyons, J. L. Payne, Body size downgrading of mammals over the late Quaternary. *Science* **360**, 310-313 (2018).
- 42. E. J. Lundgren, D. Ramp, J. Rowan, O. Middleton, S. D. Schowanek, O. Sanisidro, S. P. Carroll, M. Davis, C. J. Sandom, J. C. Svenning, A. D. Wallach, Introduced herbivores 25 restore Late Pleistocene ecological functions. *Proceedings of the National Academy of Sciences* **117**, 7871-7878 (2020).
	- 43. E. O. Aikens, K. L. Monteith, J. A. Merkle, S. P. H. Dwinnell, G. L. Fralick, M. J. Kauffman, Drought reshuffles plant phenology and reduces the foraging benefit of green-wave surfing for a migratory ungulate*. Global Change Biology* **26**, 4215-4225 (2020).
- 30 44. IPCC, 2022: Climate Change 2022: Impacts, Adaptation, and Vulnerability. Contribution of Working Group II to the Sixth Assessment Report of the Intergovernmental Panel on Climate Change [H.-O. Pörtner, D.C. Roberts, M. Tignor, E.S. Poloczanska, K. Mintenbeck, A. Alegría, M. Craig, S. Langsdorf, S. Löschke, V. Möller, A. Okem, B. Rama (eds.)]. Cambridge University Press. Cambridge University Press, Cambridge, UK and New York, 35 NY, USA, 3056 pp., doi:10.1017/9781009325844.
	- 45. F. White, The vegetation of Africa, *Natural Resources Research, UNESCO* **20**, 1-356 (1983).
	- 46. H. A. de Wit, "*Soils and grassland types of the Serengeti Plain (Tanzania): their distribution and interrelations*,", thesis, Agricultural University Wageningen, The Netherlands (1978).
- 47. T. Jaeger, "*Soils of the Serengeti woodlands, Tanzania*," thesis, Agricultural University 40 Wageningen, The Netherlands (1982).
	- 48. A. R. E. Sinclair, J. G. C. Hopcraft, H. Olff, S. A. Mduma, K. A. Galvin, G. J. Sharam, "Historical and future changes to the Serengeti ecosystem" in *Serengeti III: Human Impacts on Ecosystem Dynamics* (University of Chicago Press, Chicago, 2008), pp. 7-46.
- 49. A. Swanson, M. Kosmala, C. Lintott, C. Packer, A generalized approach for producing, 45 quantifying, and validating citizen science data from wildlife images. *Conservation Biology* **30**, 520-531 (2016).
	- 50. M. S. Norouzzadeh, A. Nguyen, M. Kosmala, A. Swanson, M. S. Palmer, C. Packer, J. Clune, Automatically identifying, counting, and describing wild animals in camera-trap
		- 9

images with deep learning. *Proceedings of the National Academy of Sciences* **115**, E5716- E5725 (2018).

- 51. M. Eickenberg, A. Gramfort, G. Varoquaux, B. Thirion, Seeing it all: Convolutional network layers map the function of the human visual system. *NeuroImage* **152**, 184-194 (2017).
- 5 52. G. Chen, T. X. Han, Z. He, R. Kays, T. Forrester, Deep convolutional neural network based species recognition for wild animal monitoring. *IEEE International Conference on Image Processing (ICIP)* 858-862. (2014). doi: 10.1109/ICIP.2014.7025172.
- 53. A. G. Villa, A. Salazar, F. Vargas, Towards automatic wild animal monitoring: Identification of animal species in camera-trap images using very deep convolutional neural networks. 10 *Ecological Informatics* **41**, 24-32 (2017).
	- 54. M. A. Tabak, M. S. Norouzzadeh, D. W. Wolfson, S. J. Sweeney, K. C. VerCauteren, N. P. Snow, J. M. Halseth, P. A. Di Salvo, J. S. Lewis, M. D. White, B. Teton, Machine learning to classify animal species in camera trap images: Applications in ecology. *Methods in Ecology and Evolution* **10**, 585-590 (2019).
- 15 55. M. Willi, R. T. Pitman, A. W. Cardoso, C. Locke, A. Swanson, A. Boyer, M. Veldthuis, L. Fortson, Identifying animal species in camera trap images using deep learning and citizen science. *Methods in Ecology and Evolution* **10**, 80-91 (2019).
	- 56. R Core Team, R: A language and environment for statistical computing. R Foundation for Statistical Computing, Vienna, Austria. URL https://www.R-project.org/ (2019).
- 20 57. T. M. Anderson, S. White, B. Davis, R. Erhardt, M. Palmer, A. Swanson, M. Kosmala, C. Packer, The spatial distribution of African savannah herbivores: species associations and habitat occupancy in a landscape context. *Philosophical Transactions of the Royal Society B: Biological Sciences* **371**, 20150314 (2016).
- 58. N. S. Novella, W. M. Thiaw, African rainfall climatology version 2 for famine early warning 25 systems. *Journal of Applied meteorology and Climatology* **52**, 588-606 (2013).
	- 59. T. Dinku, C. Funk, P. Peterson, R. Maidment, T. Tadesse, H. Gadain, P. Ceccato, Validation of the CHIRPS satellite rainfall estimates over eastern Africa. *Quarterly Journal of the Royal Meteorological Society* **144**, 292-312 (2018).
- 60. T. M. Anderson, P. M. Ngoti, M. L. Nzunda, D. M. Griffith, J. D. Speed, F. Fossøy, E. 30 Røskaft, B. J. Graae, The burning question: does fire affect habitat selection and forage preference of the black rhinoceros Diceros bicornis in East African savannahs? *Oryx* **54**, 234- 243 (2020).
	- 61. J. M. Paruelo, H. E. Epstein, W. K. Lauenroth, I.C. Burke, ANPP estimates from NDVI for the central grassland region of the United States. *Ecology* **78**, 953-958 (1997).
- 35 62. J. Dempewolf, S. Trigg, R. S. DeFries, S. Eby, Burned-area mapping of the Serengeti–Mara region using MODIS reflectance data. *IEEE Geoscience and Remote Sensing Letters* **4**, 312- 316 (2007).
	- 63. R. J. Hijmans, raster: Geographic Data Analysis and Modeling. R package version 3.4-5. https://CRAN.R-project.org/package=raster (2020).
- 40 64. H. W. Borchers, pracma: Practical Numerical Math Functions. R package version 2.2.9, https://CRAN.R-project.org/package=pracma (2019).
	- 65. O. M. Richmond, J. E. Hines, S. R. Beissinger, Two‐species occupancy models: a new parameterization applied to co‐occurrence of secretive rails. *Ecological Applications* **20**, 2036-2046 (2010).
- 45 66. P. C. Caragea, M. S. Kaiser, Autologistic models with interpretable parameters. *Journal of Agricultural, Biological, and Environmental Statistics* **14**, 281-300 (2009).
	- 67. S. Sturtz, U. Ligges, A. Gelman, R2WinBUGS: A Package for Running WinBUGS from R. *Journal of Statistical Software* **12**, 1-16 (2005).
- 68. C. Packer, R. Hilborn, A. Mosser, B. Kissui, M. Borner, G. Hopcraft, J. Wilmshurst, S. Mduma, A. R. Sinclair, Ecological change, group territoriality, and population dynamics in Serengeti lions. *Science* **307**, 390-393 (2005).
- 69. C. Calenge, The package "adehabitat" for the R software: a tool for the analysis of space and 5 habitat use by animals. Ecological modelling **197**, 516-519 (2006).
	- 70. A. E. Gelfand, S. Shirota, Clarifying species dependence under joint species distribution modeling. https://arxiv.org/abs/1908.09410v1 (2019).
- 71. M. P. Veldhuis, M. E. Ritchie, J. O. Ogutu, T. A. Morrison, C. M. Beale, A. B. Estes, W. Mwakilema, G. O. Ojwang, C. L. Parr, J. Probert, P. W. Wargute, J. G. C. Hopcraft, H. Olff, 10 The Serengeti squeeze: Cross-boundary human impacts compromise an iconic protected ecosystem. *Science* **363**, 1424-1428 (2019).
	- 72. T. Avgar, J. R. Potts, M. A. Lewis, M. S. Boyce, Integrated step selection analysis: bridging the gap between resource selection and animal movement. *Methods in Ecology and Evolution* **7**, 619–630 (2016).
- 15 73. Signer, J., J. Fieberg, and T. Avgar, Animal movement tools (amt): R package for managing tracking data and conducting habitat selection analyses. *Ecology and Evolution* **9**, 880-890 (2019).
	- 74. Therneau, T, A Package for Survival Analysis in R. R package version 3.2-13 https://CRAN.R-project.org/package=survival (2021).
- 20 75. Muff, S., J. Signer, J. Fieberg, Accounting for individual-specific variation in habitatselection studies: Efficient estimation of mixed-effects models using Bayesian or frequentist computation. *Journal of Animal Ecology* **89**, 80–92 (2020).
- 76. Magnusson, A., H. J. Skaug, A. Nielsen, C. W. Berg, K. Kristensen, M. Maechler, K. van Bentham, B. Bolker, M. E. Brooks, glmmTMB: generalized linear mixed models using 25 template model builder (2017).
	- 77. Merkle, J. A., K. L. Monteith, E. O. Aikens, M. M. Hayes, K. R. Hersey, A. D. Middleton, B. A. Oates, H. Sawyer, B. M. Scurlock, and M. J. Kauffman, Large herbivores surf waves of green-up during spring. *Proceedings of the Royal Society B: Biological Sciences* **283**, 20160456 (2016).
- 30 78. R. M. Pringle, M. C. Hutchinson, Resolving food-web structure. *Annual Review of Ecology, Evolution, and Systematics* **51**, 55-80 (2020).
	- 79. J. Pansu, J. A. Guyton, A. B. Potter, J. L. Atkins, J. H. Daskin, B. Wursten, T. R. Kartzinel, R. M. Pringle, Trophic ecology of large herbivores in a reassembling African ecosystem. *Journal of Ecology* 107, 1355-1376 (2019).
- 35 80. T. R. Kartzinel, R. M. Pringle, Multiple dimensions of dietary diversity in large mammalian herbivores. *Journal of Animal Ecology* **89**, 1482-1496 (2020).
	- 81. P. Taberlet, E. Coissac, F. Pompanon, L. Gielly, C. Miquel, A. Valentini, T. Vermat, G. Corthier, C. Brochmann, E. Willerslev, Power and limitations of the chloroplast trn L (UAA) intron for plant DNA barcoding. *Nucleic Acids Research* **35**, e14-e14 (2007).
- 40 82. A. Valentini, C. Miquel, M. A. Nawaz, E. V. A. Bellemain, E. Coissac, F. Pompanon, L. Gielly, C. Cruaud, G. Nascetti, P. Wincker, J. E. Swenson, New perspectives in diet analysis based on DNA barcoding and parallel pyrosequencing: the trnL approach. *Molecular Ecology Resources* **9**, 51-60 (2009).
- 83. J. Pansu, C. Giguet‐Covex, G. F. Ficetola, L. Gielly, F. Boyer, L. Zinger, F. Arnaud, J. 45 Poulenard, P. Taberlet, P. Choler, Reconstructing long‐term human impacts on plant communities: An ecological approach based on lake sediment DNA. *Molecular Ecology* **24**, 1485-1498 (2015).

- 84. J. Binladen, M. T. P. Gilbert, J. P. Bollback, F. Panitz, C. Bendixen, R. Nielsen, E. Willerslev, The use of coded PCR primers enables high-throughput sequencing of multiple homolog amplification products by 454 parallel sequencing. *PLoS One* **2**, e197 (2007).
- 85. G. F. Ficetola, J. Pansu, A. Bonin, E. Coissac, C. Giguet‐Covex, M. De Barba, L. Gielly, C. 5 M. Lopes, F. Boyer, F. Pompanon, G. Rayé, Replication levels, false presences and the estimation of the presence/absence from eDNA metabarcoding data. *Molecular Ecology Resources* **15**, 543-556 (2015).

86. F. Boyer, C. Mercier, A. Bonin, Y. Le Bras, P. Taberlet, E. Coissac, Obitools: a unix-inspired software package for DNA metabarcoding. *Molecular Ecology Resources* **16**, 176–182 10 (2016).

- 87. B. A. Gill, P. M. Musili, S. Kurukura, A. A. Hassan, J. R. Goheen, W. J. Kress, M. Kuzmina, R. M. Pringle, T. R. Kartzinel, Plant DNA‐barcode library and community phylogeny for a semi‐arid East African savanna. *Molecular Ecology Resources* **19**, 838-846 (2019).
- 88. G. F. Ficetola, E. Coissac, S. Zundel, T. Riaz, W. Shehzad, J. Bessière, P. Taberlet, F. 15 Pompanon, An in silico approach for the evaluation of DNA barcodes. *BMC genomics* **11**, 1- 10 (2010).
	- 89. J. A. Guyton, J. Pansu, M. C. Hutchinson, T. R. Kartzinel, A. B. Potter, T. C. Coverdale, J. H. Daskin, A. G. da Conceição, M. J. Peel, M. E. Stalmans, R. M. Pringle, Trophic rewilding revives biotic resistance to shrub invasion. *Nature Ecology & Evolution* **4**, 712-724 (2020).
- 20 90. T. Carlsen, A. B. Aas, D. Lindner, T. Vrålstad, T. Schumacher, H. Kauserud, Don't make a mista(g)ke: is tag switching an overlooked source of error in amplicon pyrosequencing studies? *Fungal Ecology* **5**, 747–749 (2012).
- 91. I. B. Schnell, K. Bohmann, M. T. P. Gilbert, Tag jumps illuminated reducing sequence-tosample misidentifications in metabarcoding studies. *Molecular Ecology Resources* **15**, 1289– 25 1303 (2015).
	- 92. E. Willerslev, J. Davison, M. Moora, M. Zobel, E. Coissac, M. E. Edwards, E. D. Lorenzen, M. Vestergård, G. Gussarova, J. Haile, J Craine, Fifty thousand years of Arctic vegetation and megafaunal diet. *Nature* **506**, 47-51 (2014).
- 93. B. Gebremedhin, Ø. Flagstad, A. Bekele, D. Chala, V. Bakkestuen, S. Boessenkool, M. 30 Popp, G. Gussarova, A. Schrøder-Nielsen, S. Nemomissa, C. Brochmann, DNA metabarcoding reveals diet overlap between the endangered Walia ibex and domestic goatsimplications for conservation. *PLoS One* **11**, p.e0159133 (2016).
- 94. B. E. Deagle, A. C. Thomas, J. C. McInnes, L. J. Clarke, E. J. Vesterinen, E. L. Clare, T. R. Kartzinel, J. P. Eveson, Counting with DNA in metabarcoding studies: How should we 35 convert sequence reads to dietary data? Molecular Ecology **28**, 391-406 (2019).
	- 95. T. M. Anderson, J. Shaw, H. Olff, Ecology's cruel dilemma, phylogenetic trait evolution and the assembly of Serengeti plant communities. *Journal of Ecology* **99**, 797-806 (2011).
	- 96. P. Taberlet, L. Gielly, G. Pautou, J. Bouvet, Universal primers for amplification of three noncoding regions of chloroplast DNA. *Plant Molecular Biology* **17**, 1105–1109 (1991).
- 40 97. M. Kearse, R. Moir, A. Wilson, S. Stones-Havas, M. Cheung, S. Sturrock, S. Buxton, A. Cooper, S. Markowitz, C. Duran, T. Thierer, Geneious Basic: an integrated and extendable desktop software platform for the organization and analysis of sequence data. *Bioinformatics* **28**, 1647-1649 (2012).
- 98. J. Oksanen, F. Guillaume Blanchet, M. Friendly, R. Kindt, P. Legendre, D. McGlinn, P. R. 45 Minchin, R. B. O'Hara, G. L. Simpson, P. Solymos, M. H. H. Stevens, E. Szoecs, H. Wagner, vegan: Community Ecology Package. R package version 2.5-6. https://CRAN.Rproject.org/package=vegan (2019).

- 99. E. R. Pianka, The structure of lizard communities. *Annual Review of Ecology and Systematics* **4**, 53-74 (1973).
- 100. N. J. Gotelli, E. M. Hart, A. M. Ellison, EcoSimR: Null model analysis for ecological data. R package version 0.1.0. http://github.com/gotellilab/EcoSimR DOI: 5 10.5281/zenodo.16522 (2015).
	- 101. T. Hastie, gam: Generalized Additive Models. R package version 1.20. https://CRAN.Rproject.org/package=gam (2020).
- 102. S. N. Wood, N. Pya, B. Saefken, Smoothing parameter and model selection for general smooth models (with discussion). *Journal of the American Statistical Association* **111**, 1548- 10 1575 (2016).
	- 103. M. Smithson, J. Verkuilen, A better lemon squeezer? Maximum-likelihood regression with beta-distributed dependent variables. *Psychological Methods* **11**, 54-71 (2006).
- 104. S. J. McNaughton, Serengeti grassland ecology: the role of composite environmental factors and contingency in community organization. *Ecological Monographs* **53**, 291-320 15 (1983).
	- 105. T. M. Anderson, K. L. Metzger, S. J. McNaughton, Multi-scale analysis of plant species richness in Serengeti grasslands. *Journal of Biogeography* **34**, 313-323 (2007).
	- 106. T. M. Anderson, Plant compositional change over time increases with rainfall in Serengeti grasslands. *Oikos* **117**, 675-682 (2008).
- 20 107. V. S. Ivlev, *Experimental Ecology of the Feeding of Fishes* (Yale University Press, New Haven, 1961).
- 108. M. E. Brooks, K. Kristensen, K. J. van Benthem, A. Magnusson, C. W. Berg, A. Nielsen, H. J. Skaug, M. Maechler, B. M. Bolker, glmmTMB Balances Speed and Flexibility Among Packages for Zero-inflated Generalized Linear Mixed Modeling. *The R Journal* **9**, 378-400 25 (2017).
	- 109. F. Hartig, DHARMa: Residual Diagnostics for Hierarchical (Multi-Level / Mixed) Regression Models. R package version 0.3.3.0. https://CRAN.Rproject.org/package=DHARMa (2020).
- 110. R. V. Lenth, emmeans: Estimated Marginal Means, aka Least-Squares Means. R package 30 version 1.5.4. https://CRAN.R-project.org/package=emmeans (2021)
	- 111. T.M. Anderson, J. Bukombe, K. L. Metzger, "Spatial and temporal drivers of plant structure and diversity in Serengeti savannas" in *Serengeti IV: Sustaining Biodiversity in a Coupled Human-Natural System*, A. R. E. Sinclair, K. L. Metzger, S. A. R. Mduma, J. M. Fryxell, Eds. (University of Chicago Press, Chicago, 2015), pp. 105-124.
- 35 112. G. P. Hempson, S. Archibald, W. J. Bond, A continent-wide assessment of the form and intensity of large mammal herbivory in Africa. *Science* **350**, 1056-1061 (2015).
	- 113. Cain, M.L., Bowman, W.D. and Hacker, S.D., 2011. Ecology, Second edition, Sunderland, Mass.: Sinauer.
- 114. T. M. Anderson, D. M. Griffith, J. B. Grace, E. M. Lind, P. B. Adler, L. A. Biederman, 40 D. M. Blumenthal, P. Daleo, J. Firn, N. Hagenah, and W. S. Harpole, Herbivory and eutrophication mediate grassland plant nutrient responses across a global climatic gradient. *Ecology* **99**, 822-831 (2018).

**Acknowledgments:** We thank the >130,000 volunteers from 77 countries who classified images on Snapshot Serengeti as well as the team at Zooniverse for their support in hosting imagery.

45 The Minnesota Supercomputing Institute and Wake Forest University High Performance Computing Cluster provided resources that contributed to data storage and processing. Snapshot Serengeti was publicized by the Association of Zoos and Aquariums' 'Saving Animals from Extinction' (SAFE) program and Disney Conservation Fund's 'Protect the Pride' campaign. We

acknowledge Ciara Nutter for her laboratory work that facilitated the construction of the local DNA reference library for Serengeti grasses. Research clearance was provided by the Tanzania Wildlife Research Institute and Tanzania National Parks. **Funding:** 5 National Geographic Society grant WW-025R-17 (TMA) National Geographic Society grant NGS-52921R-18 (TMA) National Science Foundation grant BCS-1461728 (TMA, RMH) European Union's Horizon 2020 research and innovation program grant 641918 (JGCH, TAM) National Science Foundation grant PRFB-1810586 (CP) 10 National Science Foundation grant IIS-1619177 (CP) National Science Foundation grant OAC-1835530 (CP) National Science Foundation grant IOS-1656527 (RMP) National Science Foundation grant DEB-1457697 (RMP) Frankfurt Zoological Society (JGCH, TAM) 15 British Ecological Society (JGCH, TAM) **Author contributions:** Conceptualization: TMA, SAH, RMH, JED, RJE, JGCH, MCH, TAM, JP, RMP, CP Data curation: TMA, MCH, SHE, JM, INM, JP Formal analysis: TMA, SAH, RMH, RJE, MCH, TAM, JP 20 Funding acquisition: TMA, JGCH, RMP, CP Investigation: TMA, SAH, RMH, RJE, MCH, TAM, JP, RS Methodology and data collection: TMA, SAH, RMH, JED, RJE, JGCH, MCH, SEH, TAM, JM, INM, MSP, JP, RMP, RS, CP Project administration: TMA, SAH, JED, RJE, JGCH, MCH, SEH, TAM, JM, JP, RMP, CP 25 Writing – original draft: TMA Writing – review & editing: TMA, SAH, RMH, JED, RJE, JGCH, MCH, SEH, TAM, JM, MSP, JP, RMP, CP **Competing interests:** Authors declare that they have no competing interests. **Data and materials availability:** 30 Movement data collected with GPS collars for migratory Serengeti wildebeest and zebra presented in Figure 1 are made available through MoveBank with unique identifiers: "Wildebeest (Western White-Bearded) Hopcraft Serengeti-Mara" and "Plains zebra Hopcraft Serengeti-Mara". Raw and filtered dietary sequencing data are deposited in DRYAD with the unique doi: 10.5061/dryad.brv15dvcj. Data, code, and materials used for the Bayesian occupancy 35 models, integrated step selection models of migratory herbivores, and R scripts to produce the dietary results presented here are made available through DRYAD with the unique doi: 10.5061/dryad.7sqv9s4zk. **Supplementary Materials:** Materials and Methods 40 Supplementary Text Tables S1 – S6 Figs.  $S1 - S15$ References (*45*–*115*) Data  $S1 - S5$ 



**Fig. 1. Movements (1999-2020), temporal dynamics, and species interactions of migratory grazers in the Serengeti ecosystem.** Our study area (blue polygons in **A**, **B**) is a migratory transition zone linking the Serengeti plains, Western corridor, and Masai Mara Game Reserve.

Based on GPS collaring data collected between 1999-2020 (*18*), 96% of collared wildebeest (*n* = 54, grey lines in **A**) and 100% of zebra ( $n = 28$ , black lines in **B**) passed through the study area. Grey boxes show monthly mean locations for wildebeest (**A**) and zebra (**B**) across the entire dataset with orange arrows indicating the direction of movements. Camera traps (points in **A**, **B**) 5 are separated by ~2.2 km and operated almost continuously between July 2010 and August 2018. Bell's (*5*) foundational study was from a single plot in the Western corridor (asterisk in **A**, **B**) at a time when the wildebeest population was approximately 25% of their current size (*18*). Thomson's gazelles have not been collared in the Serengeti National Park; consequently, we focused our analysis of individual movements on zebra and wildebeest. **(C–E)** Classically 10 hypothesized expectations from grazing succession (*5, 13*), including migratory species' abundance over the dry season (**C**) and species-specific effects on the probability of other species' occurrence at four-day (**D**) and sixteen- to thirty-two-day intervals (**C**). The classic model predicts unidirectional spatial associations between migrants at substantial time lags (**E**) driven by activities of large herbivores (foraging, trampling, nutrient redistribution) that facilitate 15 foraging by small herbivores. Negative interactions, not explicitly considered in early work (*5, 9*; cf. *13*), would appear on short timescales (**D**) because of lagged positive associations. (**F–H**) Associations observed in this study. Average abundance during the transition season (**F**) supported the classic model for gazelles, but not for zebra and wildebeest, which co-migrated through the study area (Fig. S3); standardized species abundances were plotted against satellite-20 derived vegetation greenness (an index of plant senescence during the dry season) to control for variation in migration timing across years. Within four-day windows, zebra and wildebeest had strong reciprocal effects on each other's occupancy throughout migration, while lagged arrival times led to negative associations with gazelles (**G**). Within sixteen-thirty-two-day windows, zebra and wildebeest had positive effects on occupancy of gazelles, and negative effects on each 25 other due to their concurrent migration (**H**). Each arrow represents the average positive (black) or negative (grey) conditional association strength ( $\gamma_{species}$  2 | species 1) from Bayesian occupancy models fit at four-day (**G**) or sixteen-day (**H**) time intervals across the entire study period (Supplemental Materials). Values are mean  $\pm$  95% confidence intervals (CI; tables S2 and S3) with arrow widths proportional to the magnitude of the effect size. Dashed lines represent weak 30 effects that were below included zero in the 95% CI.

#### Submitted Manuscript: Confidential Template revised October 2022



**Fig. 2. Fire weakens but rainfall strengthens species associations during grazing succession.**  The relative strength of species association (*RSA*) is the difference between species associations from Bayesian occupancy models fit at four-day or sixteen-day time interval for a given year 5 minus the mean across all years  $(RSA_i = \hat{\gamma}_i - \bar{\gamma})$ ; Supplementary Materials).  $RSA_i$  values are plotted against (**A**) the natural log of the burned area and (**B**) cumulative rainfall from January– June of year *i*. Fire was negatively related to variation in  $RSA_i$  calculated from four-day spatial associations (**A**; Table S4), with the strongest correlations between smaller migrants conditioned on the presence of larger migrants (WB|ZB and TG|WB); rainfall was positively related to 10 variation in  $RSA_i$  calculated from sixteen-day spatial associations  $(B)$ , but only for Thomson's gazelles tracking larger migrants (TG|ZB and TG|WB). Pearson correlation coefficients (*r*) are shown in each graph. Active camera density during the 2015 transition season was lower than for other years (Fig. S1) and was therefore excluded from these analyses.



**Fig. 3. Wildebeest and zebra movements during one dry and one wet year.** 

(**A–B**) Movement models of collared individuals were analyzed in relation to distances from 5 large heterospecific herds as determined by camera trap data (see Supplementary Materials). In both dry (2017) and wet (2018) years, wildebeest selected habitat independent of distance to large, dense zebra herds (**A**). In contrast, zebra selected sites at moderate distances from dense wildebeest herds  $(B)$ , with greater distances in the dry (light grey line, mode = 1383 m) compared to wet year (dark grey line; mode = 800 m). (**C–J**) Foraging behaviors inferred from 10 movement data. In the dry year, wildebeest moved with large step lengths of ~400 m between GPS fixes (4-hour fixed rate), especially individuals towards the front of the migration nearest to zebra (**C**), indicating directed movement (**D**). Zebra step lengths during the dry year were marginally larger than the expected from the tentative distribution (**E**). In the wet year (2018) wildebeest used smaller step lengths than expected from the tentative distribution (**G**), consistent 15 with greater local resource availability and foraging behavior. Zebra movement changed dramatically in the wet year, with smaller step lengths and wide, circuitous turn angles, especially for individuals within 1000 m from dense wildebeest herds (**I**, **J**), suggesting decreased competitive effects of heterospecifics. Black dashed lines show 'tentative' distributions derived from the average movement of animals during a migration season. Color 20 gradients represent binned zebra (**C**, **D**) or wildebeest (**E**, **F**) distances to large heterospecific herds as determined by camera trap data (see Supplementary Materials).

#### Submitted Manuscript: Confidential Template revised October 2022



#### **Fig. 4. Differential consumption of plant taxa of varying nutritional quality.**

(**A**) Mean diet composition by plant family (colors) during the migration period in dry (2017) 5 and wet years (2018), showing unexpectedly high consumption of high-quality (Fig. S15) forbs by wildebeest and gazelles, notably legumes (Fabaceae). (**B**) Proportional grass consumption by gazelles shifted throughout the migration, increasing late in the season after zebra and wildebeest had moved on (beta regression, effect of date:  $\beta = 0.02$ ,  $Z = 6.15$ ,  $P < 0.001$ ,  $R^2_{adj} = 0.35$ ). (C) Ivlev's selectivity (proportion consumed/proportion of available forage) for the six most heavily 10 eaten grass taxa; taxa are identified to the lowest possible level based on fecal DNA sequences, including species, tribe (Andropogoneae, Cynodonteae) and clade (PACMAD). Bars below plot

show C:N ratio (colors, lower values indicate higher quality) and leaf height (length) of each grass taxon. Dashed line at selectivity 1 indicates consumption in proportion to availability; positive values, selection; negative values, avoidance. Data underlying (**A**) are in Table S6.