

Molecular-level investigation into UV-induced transformation of hydrophobic aquatic dissolved organic matter

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2	Molecular-level Investigation into UV-induced Transformation of
3 4	Hydrophobic Aquatic Dissolved Organic Matter
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Abstract

The ubiquitously present dissolved organic matter (DOM) greatly influence the efficiency of 25 26 UV-based technologies due to its reactivity to UV irradiation. In this work, UV-induced changes within three hydrophobic DOM fractions isolated from different surface waters were 27 investigated. Analysis on UV absorbance at 254 nm, electron donating capacity, fluorescence 28 intensity and carbon content revealed small changes in DOM bulk properties associated with 29 the UV-induced photochemical reactions. Fourier transform ion cyclotron resonance mass 30 31 spectrometry (FT-ICR MS) was further used to explore the modification of the molecular distribution based on H/C and O/C ratios, m/z and DBE. The molecular-level investigation 32 revealed that an average of 296 aromatic and lignin-like molecules were degraded, leading to 33 34 the production of around 306 new molecules. The UV-reactive community were identified as CHO molecules with higher DBE (>10) and carbon number (>25) which could be readily 35 transformed into smaller saturated molecules. Molecules containing nitrogen(N) or sulfur(S) 36 atom, independent of aromaticity and molecular weight (m/z), were also highly UV 37 susceptible and transformed into molecules with larger DBE and m/z. Possible reaction 38 39 pathways responsible for the observations were discussed. The results indicated that UV-40 reactivity and subsequent transformation of DOM are remarkably correlated with its molecular composition and characteristics. Though the changes in bulk properties of DOM 41 42 following UV irradiation were observed to be very small, the significant alteration in its molecular structures would have a profound impact on the UV-based treatment processes. 43

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Keywords: UV irradiation; aquatic DOM; FT-ICR MS; reactive molecules; transformation.

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46 **1. Introduction**

The strong efficiency of UV-based technologies, UV or UV-based advanced oxidation 47 processes (UV-AOPs), for disinfection or removal of emerging contaminants is well 48 documented (Ji et al. 2018, Li and Blatchley III 2009, Qin et al. 2019, Sharpless and Linden 49 2003, Zhang et al. 2017). It is also known that dissolved organic matter (DOM) greatly 50 51 influence the efficiency of the UV-based processes. Due to the prevalence of chromophoric structures, DOM can readily absorb and interact with UV light, leading to a complex 52 sequence of direct and indirect photolysis reactions. This may hinder the efficiency of UV-53 based technologies due to light attenuation, but contrariwise, generate reactive species 54 through DOM photosensitization which can contribute to the elimination of organic 55 contaminants (Berg et al. 2019, Serna-Galvis et al. 2017). UV-induced DOM reactivity would 56 directly define the efficacy of a UV-based process and impact the downstream water quality, 57 e.g., through formation of disinfection by-products, microbial regrowth nutrients, etc 58 (Imoberdorf and Mohseni 2011, Magnuson et al. 2002, Thomson et al. 2002a). For instance, 59 UV-irradiated DOM was reported to be more reactive to chlorine, leading to an increase in 60 the formation of disinfection byproducts (Magnuson et al. 2002, Meng et al. 2016). UV 61 62 exposure can also alter the bioavailability of DOM to microorganisms and consequently the efficiency of the biological treatment (Sulzberger and Durisch-Kaiser 2009, Thomson et al. 63 64 2002b). UV irradiation may also induce toxicity following DOM transformation (Frimmel 1998, He et al. 2015). Therefore, it is crucial to elucidate the UV-induced alteration of DOM 65 to provide fundamental knowledge and better understanding on the reactivity of irradiated 66 DOM. 67

Previous study has documented insignificant changes of DOM in optical properties and size
distribution under various UV doses (i.e., up 140 mJ·cm⁻²) typical for disinfection (Magnuson
et al. 2002). Interestingly, another study reported an increase in UV absorbance of DOM with

similar UV doses (~60 mJ·cm⁻²) (Meng et al. 2016). Elevated UV dosages (10⁴-10⁵ mJ·cm⁻²)
can lead to the depletion of chromophoric structures, transformation of higher molecular
weight (MW) hydrophobic (HPO) fractions to hydrophilic counterparts with lower MW, and
even carbon removal (Buchanan et al. 2005, Corin et al. 1996). However, the ultimate
molecular-level information shaping the reactivity and influencing the transformation of
DOM bulk properties during UV irradiation was not yet provided.

Electrospray ionization Fourier transform ion cyclotron resonance mass spectrometry (FT-77 ICR MS) has been widely used for the molecular-level investigation of DOM transformation 78 in various UV-based (e.g., UV/H₂O₂, UV/persulfate and UV/Chlorine) as well as other (e.g., 79 ozonation and chlorination) engineered processes (Andersson et al. 2020, Phungsai et al. 80 2016, 2018, 2019, Ruan et al. 2021, Zhang et al. 2021). Due to the ultrahigh resolution and 81 82 accuracy of this novel technique, molecules for individual mass peak could be unambiguously assigned, and the corresponding molecular characteristics and the changes 83 could be easily obtained. For instance in DOM matrix, molecules with higher aromaticity and 84 MW were found more reactive in UV/H2O2 or UV/PDS process, and transformed into 85 molecules with lower aromatic character and MW (Zhang et al. 2021). The analytic tool has 86 87 remarkable potential in providing detailed information on the UV-induced transformation of DOM at a molecular-level. The application of this advanced technique in photochemical 88 89 transformation of DOM in natural aquatic systems has also been widely documented 90 (Gonsior et al. 2009, Mesfioui et al. 2015, Miranda et al. 2020, Stubbins et al. 2010, Wilske et al. 2020). Numerous molecular-level observations have been obtained in respective system 91 92 with the use of natural sunlight or light sources simulating sunlight. However, the use of 254 93 nm light as irradiation source, widely used in water treatment processes, remains missing among these previous work. Considering the highly heterogeneous character of DOM, the 94 shift to shorter wavelength/higher energy irradiation would cause the response/reaction of 95

96 different compositions and the formation of different products. The information plays a
97 crucial role in better application of UV-based water treatment strategies and improvement of
98 water quality.

To fill these knowledge gaps, 254 nm UV-induced changes in DOM were investigated with 99 the characterization of both bulk properties using traditional analytic tools (e.g., UV-vis 100 101 spectrophotometer and fluorometer) and molecular characteristics using FT-ICR MS. The pool of molecules reactive to and produced from UV (i.e., 254 nm) irradiation was identified. 102 The possible UV-triggered reaction pathways were putatively presented based on the 103 comparison between UV-reactive and UV-produced molecules. Humic substances (i.e., HPO 104 DOM fractions) isolated from three surface waters and presenting different aromatic 105 character (i.e., SUVA value) were used in this investigation. 106

107 2. Experimental Section

108 **2.1 Chemicals and DOM samples.**

High-performance liquid chromatography (HPLC) grade methanol (MeOH) and ethyl acetate 109 (EtOAc) were supplied by Honeywell Burdick & Jackson. Sulfuric acid (H₂SO₄, 98%) and 110 111 hydrochloric acid (HCl, 32%) were purchased from UNIVAR. (3-ethylbenzothiazoline-6sulfonic acid) diammonium salt (ABTS; ≥98%) was provided by Sigma-Aldrich. Three 112 surface water DOM fractions previously isolated were selected for investigation: 113 Hydrophobic acids isolated from Suwannee River water (S-HPOA, USA), hydrophobic DOM 114 from Ribou Reservoir (R-HPO, France), and hydrophobic DOM from Colorado River (C-115 HPO, USA). The basic characteristics of the DOM samples have been reported in previous 116 117 studies (Niu et al. 2018, Zhang et al. 2021). Briefly, S-HPOA exhibited a highest SUVA value (aromaticity) of 4.93, a medium SUVA (3.16) recorded with R-HPO, and the lowest 118 SUVA (2.01) for C-HPO. 119

120 **2.2 Experimental procedures.**

A low pressure (LP)-UV collimated beam device (CBD) was used for UV irradiation 121 experiments. Briefly, the CBD is equipped with a total of three 15 W low-pressure Hg UV-122 lamps. A xy-cross slide table with a magnetic stirrer was employed to accommodate the UV 123 reactor. An incident UV-C intensity of ~0.9 mW·cm⁻² determined using a certified UV-C 124 radiometer was applied to irradiate a 250 mL DOM solution (~ 4 mgC·L⁻¹) in a petri-dish 125 (i.e., UV reactor, 84 mm in diameter). The initial UV fluence was further corrected with 126 radiometer, divergence, reflection, petri and water factor (Bolton and Linden 2003). UV 127 doses of 633, 760, and 886 mJ·cm⁻² were calculated for S-HPOA-, R-HPO-, and C-HPO-128 containing solutions, respectively. Experiments were performed in absence of buffer to avoid 129 the potential formation of buffer-derived radicals. System pH dropped from ~4.74 to ~3.21 130 131 after 30 min irradiation. Samples for FT-ICR MS analysis were acidified to ~ pH 2 and then extracted by solid phase extraction (SPE) using Bond Elut PPL cartridges (6 mL, 1 g, Agilent 132 Technologies). The extraction procedure was previously reported for FT-ICR MS analysis 133 (Zhang et al. 2021). 134

135 **2.3 Analytic methods**

136 **2.3.1** Analysis of bulk properties

Bulk DOM properties were analysed using a Cary 60 spectrophotometer for UV₂₅₄ 137 measurement (Agilent, USA), a Shimadzu TOC-L analyser (SHIDMAZU, Japan) for TOC 138 139 (Total Organic Carbon) determination, and a fluorescence spectrometer (Cary Eclipse, Varian) for FEEM (Fluorescence Excitation Emission Matrix) recording (Zhang et al. 2019). The 140 analysis of EDC (Electron Donating Capacity) of DOM samples using size exclusion 141 chromatography (SEC) based on the method developed by Chon et al. (2015). Briefly, DOM 142 sample was injected into A TOYOPEARL HW-50S column (8 mm \times 30 cm) using 50 mM 143 borate eluent at a flow rate of 0.2 mL/min. Following separation, post-column reaction of 144

DOM with the oxidant ABTS^{*+} took place in a reaction coil. ABTS^{*+} was generated through oxidation of ABTS by sodium persulfate under acidic conditions and injected to the reaction coil at 0.05 mL/min using a helium pressurised generator. Finally, two UV detectors (Agilent 148 1100 series, USA) positioned in series were used for recording UV absorbance of DOM at 254 nm and the signal for ABTS/ABTS^{*+} at 405 nm.

150 2.3.2 FT-ICR MS analysis

Procedures of sample pre-treatment by SPE using Bond Elut PPL cartridges (6 mL, 1 g, Agilent Technologies) (Dittmar et al. 2008) and sample analysis using 15.0 T Bruker SolariX FT-ICR MS were detailed in our previous publication (Zhang et al. 2021). Information on molecular assignment, classification of molecules based on H/C and O/C, and calculation of molecular characteristics can also be found in this article (Zhang et al. 2021).

156 **3. Results and discussion**

157 **3.1 UV-induced changes in bulk properties of DOM**

Upon 30 min UV irradiation with an incident intensity of ~0.9 mW·cm⁻², a slight decrease in 158 UV_{254} of the DOM samples was observed. Specifically, the decrease was recorded as 3%, 3%, 159 and 11% for S-HPOA, R-HPO, and C-HPO (Fig. 1a), respectively. It was more pronounced 160 with C-HPO, finding attributed to its lower initial UV₂₅₄ value, leading to a lower calculated 161 water factor and greater UV fluence. The change in EDC with UV irradiation was also 162 observed and found comparable with that of UV_{254} except for S-HPOA (Fig. 1a). The 163 nominally higher decrease in EDC (i.e., 6%) than UV₂₅₄ (i.e., 3%) for S-HPOA was probably 164 related with its initial higher content of phenolic structures (i.e., major EDC contributors) 165 (Zhang et al. 2019). A greater UV₂₅₄ decrease compared to EDC was previously interpreted 166 as a reaction initiated with hydroxyl radical addition to non-phenolic aromatic rings, the 167 opposite was considered to refer to an initiated depletion of phenolic moieties with the 168 preservation of aromatic structures (Chon et al. 2015, Zhang et al. 2019). The findings in the 169

170 current study implied that a UV-specific reaction mechanism was involved in the171 transformation of DOM.

172

Fig. 1.

Changes of fluorescence properties following the irradiation of DOM solutions was also 173 investigated based on the modifications recorded in the EEM spectra, i.e., excitation (Ex) and 174 175 emission (Em) maxima. Two fluorescent peaks denoted as Peak I and Peak II (Fig.1b and Fig. S1) typical of hydrophobic fractions were detected. A decrease in peak intensity was 176 determined for R-HPO (e.g., approximately 13% for Peak II) and C-HPO (e.g., 177 approximately 43% for Peak II), indicating the transformation of fluorescent structures to less 178 or non-fluorescent components. Interestingly, an increase in the intensity of both peaks (~ 179 45%) were observed for S-HPOA (Table S1). The enhancement in fluorescence signal after 180 UV irradiation or radical-based oxidation processes has also been reported (Meng et al. 2016, 181 Zhang et al. 2019). Moreover, not only was a change in the fluorescence intensity observed 182 before and after UV irradiation, but also a shift of the EEM spectra peak. For instance, the 183 emission wavelength of peak II shifted from 456 nm to 450 nm, 438 to 430 nm and 448 to 184 440 nm (Table S1) for S-HPOA, R-HOP, and C-HPO, respectively. The blue shift of the 185 emission wavelength implied the formation of smaller fluorescent molecules (Korshin et al. 186 1999) probably due to UV-induced bond cleavage. The more significant variation in 187 188 fluorescence as compared to that of UV254 indicated a more pronounced UV-induced transformation of fluorescent chromophores than non-fluorescent components. In addition, no 189 significant change in DOC concentration was observed for all DOM samples subjected to UV 190 irradiation. The results reflected that UV-induced changes in molecular compositions and 191 characteristics occurred without carbon mineralization under current experimental conditions. 192

193 **3.2 UV-induced molecular changes of DOM**

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The numbers of assigned molecules before and after UV irradiation were 1974/1990, 194 2886/2838, and 2513/2578 for S-HPOA, R-HPO, and C-HPO, respectively. Parameters such 195 as DBE, modified aromaticity index (AImod), m/z, H/C, O/C, and nominal oxidation state of 196 carbon (NOSC) (Koch and Dittmar 2006, LaRowe and Van Cappellen 2011, Remucal et al. 197 2020, Yuan et al. 2017) widely used to reflect specific DOM properties (i.e., degree of 198 unsaturation, molecular size, and degree of oxygenation), were calculated. Based on the 199 200 molecular-level statistics of the assigned peaks (Table S2), some differences in the molecular characteristics of the three DOM isolates were observed. Specifically, S-HPOA showed 201 202 highest DBE and m/z (i.e., averaged value for all the assigned molecules), but the lowest level of nitrogen (N) or sulfur (S) incorporation (0.16%) and NOSC (-0.18). R-HPO exhibited 203 medium DBE but the highest N or S content (0.45%) and NOSC (0.05). The smallest DBE 204 205 was determined for C-HPO which also exhibited medium level of N or S elements (0.19%) and NOSC (-0.10). A van Krevelen diagram was plotted to visualize the distribution of the 206 detected molecules based on their H/C and O/C as well as relative peak intensity (i.e., as 207 208 represented by the bubble size, Fig. 2). Generally, no pronounced difference in the molecular distribution (i.e., due to the minimal change in H/C or O/C) was found between the original 209 and the irradiated samples for the three DOM isolates (Fig. 2). Interestingly, the appearance 210 of N- or S- containing molecules with even lower H/C and O/C after UV irradiation were 211 found in all three DOM samples. This finding is different from what was previously detected 212 in radical-based oxidation processes, where an increase in saturation and oxidation state was 213 observed (Zhang et al. 2021). Furthermore, the change in the calculated H/C or O/C of the 214 DOM samples before and after irradiation was also negligible (Table S2). The observed 215 216 molecular-level changes in H/C or O/C was probably due to the incomplete UV-specific reactions as discussed in the following sections. 217

218

Fig. 2.

9 / 23

219 **3.3 UV-reactive, stable and produced molecules**

Based on previous studies (Remucal et al. 2020, Yuan et al. 2017), the disappeared molecules 220 (i.e., the molecular formulas detected in the original samples but not in the irradiated samples) 221 were considered as UV-reactive, while the appeared molecules (i.e., molecular formulas 222 detected only in the irradiated samples but not in the original samples) were denoted as UV-223 224 produced. Approximately 200-350 molecules disappeared upon UV irradiation, accounting for only 10%–14% of the molecules in original samples (Table 1). Around 200–400 new 225 molecules appeared after UV exposure, contributing to 10-16% of the pool of molecules in 226 the treated samples (Table 1). The cumulative intensity of disappeared or appeared molecules 227 represented only few per cents of the overall peak intensity detected for the initial or treated 228 samples (Table 1), indicating that these molecule species were less abundant or efficient in 229 generating signal. The van Krevelen plot (Fig. 3) showed the distributions of three sets of 230 molecules, i.e., disappeared, remained and appeared, based on their H/C and O/C. The 231 disappeared molecules primarily showed H/C and O/C indicative of aromatic or phenolic 232 structures (i.e., H/C = 0.7-1.5, O/C = 0.1-0.67, Fig. 3). Considering the semi-quantitative 233 nature of peak intensity, intensity-weighted average values of molecular characteristics such 234 235 as H/C_{wa}, O/C_{wa}, and DBE_{wa} for the disappeared, appeared and remained molecules were calculated and summarized in Table 1. Interestingly, lower H/C_{wa}% (~5%) and O/C_{wa} (~6%) 236 237 were detected for disappeared molecules as compared to the remained structures (Table 1). Furthermore, the disappeared molecules showed remarkably larger m/z_{wa} (~17%) and DBE_{wa} 238 (~23%) than the remained ones (Table 1), indicating the preferential UV targeting of 239 molecules with larger size and greater unsaturation. In addition, molecules incorporating N or 240 241 S were more UV-liable than the counterparts containing only CHO, as indicated by the higher relative abundance of these molecules within the pool of disappeared molecules (Fig. 3 and 242 Table 1). Interestingly, exposure to UV led to the formation of molecules with lower H or O 243

content (Table 1), as described by the decreased H/C_{wa} (~7%) and O/C_{wa} (~15%). For the 244 remained pool, molecules with change in peak intensity before and after irradiation, may also 245 contribute to the UV-reactive and UV-produced pool. Previous studies (Jennings et al. 2022, 246 Phungsai et al. 2016, 2019) have applied 30% or 50% intensity change as cut-off line for the 247 identification of these molecules from the remained pool. We also calculated the relative 248 intensity change of each molecule, based on which a van Krevelen diagram for the remained 249 250 molecules was plotted (Fig. S2). As shown in Fig. S2, the intensity variation for the majority of the molecules is within 20%. Molecules showing 30% decrease or increase in intensity 251 252 only accounted for less than 6% of the remained pool for the three DOM fractions, and the percentages for those with 50% variation is even lower. Due to the small number of these 253 molecules and the comparison uncertainty, the contribution of the remained molecules to the 254 255 reactive or the produced pool was not quantitatively discussed in drawing conclusions. However, molecules with 30% change in intensity was taken into account to support some of 256 our conclusions. 257

258

Fig. 3.

259

Table 1.

260 **3.3.1** UV-reactive molecules

To examine the role of unsaturation, molecular size as well as non-oxygen heteroatoms in 261 UV susceptibility, the disappeared and the remained molecules were plotted based on their 262 DBE and C number. For the disappeared pool, significantly different distribution in the DBE 263 versus C number map was observed between sole CHO and non-oxygen heteroatom 264 containing molecules (i.e., CHON, CHOS, and CHONS) (Fig. S3). For instance, the majority 265 of the disappeared CHO molecules occupied a region with larger DBE as well as C number, 266 while CHON, CHOS, and CHONS molecules mainly distributed in a region with smaller 267 DBE and C number. Based on this observation, the following discussion was conducted by 268

categorizing the molecules into two subgroups, i.e., molecules containing only CHO and also 269 N or S (denoted as CHO+N/S). The UV-reactive CHO molecules showed larger DBE as well 270 as C number (Fig. 4), for instance, the disappeared molecules were primary those with a 271 DBE >10 and a C number >25. Interestingly, the reactive CHO+N/S molecules spanned the 272 same range of DBE and C carbon as the preserved counterparts (Fig. 4). The potentially UV-273 reactive molecules in the remained pool (i.e., with intensity decrease >30%) was also 274 estimated regarding their aromaticity and molecular size, and identical observations with the 275 disappeared molecules were obtained (Fig. S4). For instance, 108 out of 110 such molecules 276 277 in S-HOPA contained only CHO and exhibited larger DBE and C number (Fig. S4a), and 38 out of 53 such molecules belonging to CHO+N/S subgroup in R-HPO primarily distributed in 278 the region with smaller DBE and C number (Fig. S4b). The results indicated that CHO 279 280 molecules with higher degree of unsaturation as well as complexity were more UV sensitive. However, the reactivity of the molecules bearing non-oxygen heteroatoms was independent 281 of molecular unsaturation and size. Though larger MW molecules were reported more 282 reactive to UV exposure, this observation provided complementary information for the 283 identification of UV reactive molecules. Moreover, the different observations regarding UV 284 reactivity for the CHO and CHO+N/S molecules implied disparate UV-induced reactions 285 between the two subgroups. 286

287

Fig. 4.

288 3.3.2 UV-produced molecules

289 3.3.2.1 Shift in carbon number

Exposure to UV irradiation caused the appearance of 208, 307 and 405 molecules following the disappearance of 192, 355 and 340 molecules for S-HPOA, R-HPO, and C-HPO, respectively. To characterize the UV-induced molecular-level transformation of DOM, the number of produced molecules and reactive (i.e., disappeared) molecules were plotted as a 12/23 294 function of carbon number (Fig. 5) and DBE (Fig. 6) and presented separately as two subgroups (i.e., CHO and CHO+N/S molecules). Different trends in the shift of carbon 295 number were observed between CHO and CHO+N/S molecules for the three DOM isolates 296 (Fig. 5). For CHO subgroup, the respective number of removed/produced molecules for S-297 HPOA, R-HPO, and C-HPO was 170/145, 122/134, and 122/148. Generally, a shift in C 298 number from larger to smaller upon UV irradiation could be detected, with the specific range 299 300 of C number varying for different DOM samples (Fig. 5a, c, and e). For instance, a remarkable decrease in number (i.e., from 115 to 50) was observed with molecules bearing 301 302 30-38 carbons for the produced molecules as compared to the reactive ones, primarily followed by a number increase (\sim 62%) of molecules with <30 C atoms in S-HPOA (Fig. 5a). 303 Interestingly, an increase of compounds with larger C number (~40) was also observed for 304 305 the three DOM samples (Fig. 5a, c, and e), which will be explained in Section 3.4. For the 306 CHO+N/S subgroup, the removed/produced number pair was 22/63, 233/173, and 218/257 for S-HPOA, R-HPO, and C-HPO, respectively. Specifically, the increased CHO+N/S 307 molecules in S-HPOA were found incorporating more carbon atoms (14–30), with the 308 disappearance of the smaller molecules (i.e., with a carbon number lower than 11) upon UV 309 irradiation. The lost CHO+N/S molecules observed with R-HPO were mainly those with a 310 carbon number ranging from 15-30. Interestingly, the loss of C-HPO CHO+N/S molecules 311 with a carbon number between 17 and 23 was followed by the formation of molecules with 312 313 even higher carbon number (24–32).

314

Fig. 5.

315 3.3.2.2 Shift in DBE

For the CHO subgroup, a significant decrease in the number was found with molecules showing a DBE >12 for all the DOM isolates, mainly accompanied by an increase of molecules with lower DBE (Fig. 6). A different trend in the change of DBE was observed 13/23

with CHO+N/S subgroup. For S-HPOA, the newly formed molecules were found with larger 319 DBE (i.e., >7) than the disappeared molecules. A remarkable loss of molecules with DBE 320 ranging from 7 to 14 was observed with R-HPO, followed by the formation of molecules with 321 DBE <7 or >14. Interestingly, the major depletion of molecules for C-HPO occurred with 322 those showing DBE between 9 and 13, also leading to the formation of molecules with 323 greater DBE. The observed converse trend between CHO and CHO+N/S subgroups clearly 324 325 indicated that the characterization on bulk properties could not fully reflect the changes at a molecular level and the nature of the transformation. 326

As reflected by the yellow bar in Fig. 5 and 6, the potentially UV-reactive molecules (i.e., 327 with intensity decrease greater than 30% in the remained pool) also exhibited the same trend 328 in the shift of carbon number. For instance, the vast majority of such molecules in S-HPOA 329 belong to the CHO subgroup, and they were featured with larger C number and DBE as also 330 observed with the disappeared molecules. For R-HPO, the potentially UV-reactive molecules 331 were primarily in the CHO+N/S subgroup, and their C number and DBE also fell into the 332 same range observed with the disappeared molecules. The consistency in the molecule 333 aromaticity and size between the UV-reactive and potentially UV-reactive molecules, further 334 335 confirms the conclusions drawn in this work by only comparing disappeared and appeared molecules. However, limitation of the current conclusions remains due to the uncertainty in 336 337 identifying reactive and produced molecules.

338

Fig. 6.

339 **3.4 Putative UV-induced reaction pathways responsible for the observed changes**

The reactivity of DOM to UV irradiation is highly dependent on its molecular composition and characteristics. The presence of non-oxygen heteroatoms, unsaturation level, and molecular size played an important role in shaping DOM reactivity. CHO molecules with

higher molecular weight (i.e., C number >25) and degree of unsaturation (i.e., DBE >10) 343 were more UV reactive. The reaction would readily be initiated through photon absorption by 344 these electron-rich complex structures, leading to the degradation of the molecules (Helms et 345 al. 2014, Sarathy and Mohseni 2007, Schmitt-Kopplin et al. 1998, Sulzberger and Durisch-346 Kaiser 2009). These electron-rich complex moieties of CHO molecules seem to undergo only 347 partial depletion as indicated by the insignificant shift in C number and DBE (Table S2, Fig. 348 5 and 6) as well as H/Cwa and O/Cwa (Table S2, Fig S5 and S6). The minor changes could 349 probably be attributed to the weak photoreactions mainly inducing cleavage between 350 351 aromatic structures (Scheme 1a). It should be noted that Photo-Fries rearrangement (Scheme 1b) could also occur, which however could not be detected due to the formation of 352 isomerization products (Coppinger and Bell 1966, Griesser et al. 2009). The larger molecules 353 354 (i.e., as indicated by the emergence of molecules with larger C numbers in Fig. 5a, c and e) could probably be the products of UV-induced breakdown of specific molecules unable to be 355 detected in the original samples. There is also a possibility that they were produced through 356 the UV-induced covalent linking of benzophenone to target molecules (Scheme 1c) 357 (Hermanson 2013, Ross et al. 2003). This distinct UV-induced photo-process involving 358 primary photoproducts (e.g., radical cations, electrons and triplets) and reactive intermediates 359 (e.g., ¹O₂, O₂⁻⁻, and [•]OH) is different from that of a radical-dominant (e.g., [•]OH) system. This 360 hypothesis is supported by the different UV_{254} and EDC decrease relationship (Fig. 1) as the 361 362 one reported in radical-based oxidation system (Zhang et al. 2021). As a consequence, distinct reaction pathways as well as products would be expected in a UV-induced DOM 363 transformation. 364

The reactivity of CHO+N/S molecules showed no dependency on unsaturation level and molecular size. Remarkably, molecules with even more carbons and DBE were found upon UV irradiation (Fig. 5 and 6), probably through either breakdown of originally undetected

molecules or covalent linkage between molecules (Zhong et al. 2001). For the latter, larger 368 molecules could probably be produced through the transformation of triazole-containing or 369 alike structures (Eicher et al. 2013, Kahle et al. 2008) under 254 nm UV irradiation. For 370 instance, benzotriazole has been widely used and detected in aquatic systems due to its 371 resistance to biodegradation (Giger et al. 2006, Osburn et al. 2016). The structures thus 372 present in aquatic DOM could undergo alkylation (Scheme 1d) or polymerisation (Scheme 1e) 373 374 upon UV treatment, leading to formation of bigger molecules and loss of N elements (Liu et al. 2011). The release of N₂ due to bond scission of N-N in these reactions (Scheme 1d and e) 375 376 could be supported by the decrease of nitrogen content after UV irradiation (10–30%, Table S2). The co-functioning of the chromophoric CHO and CHO+N/S molecules may explain the 377 small changes in its optical properties. The UV-induced transformation may have profound 378 379 impact on the water quality (e.g., toxicity) (Frimmel 1998, He et al. 2015) and subsequent water treatment process (e.g., disinfection, membrane filtration and biological treatment) 380 (Sulzberger and Durisch-Kaiser 2009, Thomson et al. 2002b). 381

382

Scheme 1.

383 **4.** Conclusions

Small but significant 254 nm UV-induced changes were observed with three HPO DOM 384 isolates exhibiting different level of aromaticity and non-oxygen heteroatom incorporation. 385 Though minor changes in bulk properties such as EDC, UV254, fluorescence intensity or 386 carbon content were measured, remarkable transformation of UV-reactive components 387 388 occurred at a molecular-level. The UV-reactive CHO molecules were characterized with larger C number (i.e, >25) and DBE (i.e, >10), consequently transformed into smaller 389 saturated components. Independent of aromaticity and complexity, molecules containing N or 390 391 S were also UV-reactive. Formation of larger molecules were found with these molecules probably through UV-induced covalent linkage. 392

The results of the current study also revealed that the initial composition and characteristics 393 of DOM played a significant role in shaping its reactivity toward UV irradiation. The 394 difference in the reactivity would lead to distinct transformation products and consequently 395 influence on the UV-based treatment process or water quality. The findings provided new 396 insights into the application of UV-based technology especially the direct UV treatment 397 technique. However, the results found in this study may not be expected in a typical UV 398 disinfection process as the dosages applied (600–900 mJ \cdot cm⁻²) were higher than that for UV 399 disinfection (40–100 mJ \cdot cm⁻²). 400

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404 **Reference**

Andersson, A., Lavonen, E., Harir, M., Gonsior, M., Hertkorn, N., Schmitt-Kopplin, P., Kylin,
H. and Bastviken, D. (2020) Selective removal of natural organic matter during drinking
water production changes the composition of disinfection by-products. Environmental
Science: Water Research & Technology 6(3), 779-794.

Berg, S.M., Whiting, Q.T., Herrli, J.A., Winkels, R., Wammer, K.H. and Remucal, C.K.
(2019) The role of dissolved organic matter composition in determining photochemical
reactivity at the molecular level. Environmental Science & Technology 53(20), 11725-11734.
Bolton, J.R. and Linden, K.G. (2003) Standardization of Methods for Fluence (UV Dose)
Determination in Bench-Scale UV Experiments. Journal of Environmental Engineering
129(3), 209-215.

- 415 Buchanan, W., Roddick, F., Porter, N. and Drikas, M. (2005) Fractionation of UV and VUV
- 416 Pretreated Natural Organic Matter from Drinking Water. Environmental Science &
- 417 Technology 39(12), 4647-4654.

- 418 Chon, K., Salhi, E. and von Gunten, U. (2015) Combination of UV absorbance and electron
- donating capacity to assess degradation of micropollutants and formation of bromate duringozonation of wastewater effluents. Water Research 81, 388-397.
- 421 Coppinger, G.M. and Bell, E.R. (1966) Photo-Fries Rearrangement of Aromatic Esters. Role
- 422 of Steric and Electronic Factors. The Journal of Physical Chemistry 70(11), 3479-3489.
- 423 Corin, N., Backlund, P. and Kulovaara, M. (1996) Degradation products formed during UV-
- 424 irradiation of humic waters. Chemosphere 33(2), 245-255.
- 425 Dittmar, T., Koch, B., Hertkorn, N. and Kattner, G. (2008) A simple and efficient method for
- 426 the solid-phase extraction of dissolved organic matter (SPE-DOM) from seawater.
- 427 Limnology and Oceanography: Methods 6(6), 230-235.
- 428 Eicher, T., Hauptmann, S. and Speicher, A. (2013) The chemistry of heterocycles: structures,
- 429 reactions, synthesis, and applications, John Wiley & Sons.
- 430 Frimmel, F.H. (1998) Impact of light on the properties of aquatic natural organic matter.
- 431 Environment International 24(5), 559-571.
- 432 Giger, W., Schaffner, C. and Kohler, H.-P.E. (2006) Benzotriazole and Tolyltriazole as
- 433 Aquatic Contaminants. 1. Input and Occurrence in Rivers and Lakes. Environmental Science
- 434 & Technology 40(23), 7186-7192.
- 435 Gonsior, M., Peake, B.M., Cooper, W.T., Podgorski, D., D' Andrilli, J. and Cooper, W.J.
- 436 (2009) Photochemically Induced Changes in Dissolved Organic Matter Identified by
- 437 Ultrahigh Resolution Fourier Transform Ion Cyclotron Resonance Mass Spectrometry.
- 438 Environmental Science & Technology 43(3), 698-703.
- Griesser, T., Höfler, T., Jakopic, G., Belzik, M., Kern, W. and Trimmel, G. (2009) Refractive
 index modulation in polymers bearing photoreactive phenyl and naphthyl ester units using
 different UV wavelengths. Journal of Materials Chemistry 19(26), 4557-4565.
- 442 He, X., de la Cruz, A.A., Hiskia, A., Kaloudis, T., O'Shea, K. and Dionysiou, D.D. (2015)
- 443 Destruction of microcystins (cyanotoxins) by UV-254 nm-based direct photolysis and
- 444 advanced oxidation processes (AOPs): Influence of variable amino acids on the degradation
- kinetics and reaction mechanisms. Water Research 74, 227-238.
- 446 Helms, J.R., Mao, J., Stubbins, A., Schmidt-Rohr, K., Spencer, R.G.M., Hernes, P.J. and
- 447 Mopper, K. (2014) Loss of optical and molecular indicators of terrigenous dissolved organic
- 448 matter during long-term photobleaching. Aquatic Sciences 76(3), 353-373.
- Hermanson, G.T. (2013) Bioconjugate Techniques (Third Edition). Hermanson, G.T. (ed), pp.
- 450 229-258, Academic Press, Boston.

- 451 Imoberdorf, G. and Mohseni, M. (2011) Degradation of natural organic matter in surface
 452 water using vacuum-UV irradiation. Journal of Hazardous Materials 186(1), 240-246.
- 453 Jennings, E., Kremser, A., Han, L., Reemtsma, T. and Lechtenfeld, O.J. (2022) Discovery of
- 454 Polar Ozonation Byproducts via Direct Injection of Effluent Organic Matter with Online LC-
- 455 FT-ICR-MS. Environmental Science & Technology 56(3), 1894-1904.
- Ji, Y., Yang, Y., Zhou, L., Wang, L., Lu, J., Ferronato, C. and Chovelon, J.-M. (2018)
 Photodegradation of sulfasalazine and its human metabolites in water by UV and
 UV/peroxydisulfate processes. Water Research 133, 299-309.
- Kahle, M., Buerge, I.J., Hauser, A., Müller, M.D. and Poiger, T. (2008) Azole Fungicides:
 Occurrence and Fate in Wastewater and Surface Waters. Environmental Science &
- 461 Technology 42(19), 7193-7200.
- 462 Koch, B.P. and Dittmar, T. (2006) From mass to structure: an aromaticity index for high-
- resolution mass data of natural organic matter. Rapid Communications in Mass Spectrometry
 20(5), 926-932.
- Korshin, G.V., Kumke, M.U., Li, C.-W. and Frimmel, F.H. (1999) Influence of chlorination
 on chromophores and fluorophores in humic substances. Environmental Science &
 Technology 33(8), 1207-1212.
- LaRowe, D.E. and Van Cappellen, P. (2011) Degradation of natural organic matter: A
 thermodynamic analysis. Geochimica et Cosmochimica Acta 75(8), 2030-2042.
- 470 Li, J. and Blatchley III, E.R. (2009) UV Photodegradation of Inorganic Chloramines.
 471 Environmental Science & Technology 43(1), 60-65.
- 472 Liu, Y.-S., Ying, G.-G., Shareef, A. and Kookana, R.S. (2011) Photolysis of benzotriazole
- and formation of its polymerised photoproducts in aqueous solutions under UV irradiation.
- 474 Environmental Chemistry 8(2), 174-181.
- 475 Magnuson, M.L., Kelty, C.A., Sharpless, C.M., Linden, K.G., Fromme, W., Metz, D.H. and
- 476 Kashinkunti, R. (2002) Effect of UV Irradiation on Organic Matter Extracted from Treated
- 477 Ohio River Water Studied through the Use of Electrospray Mass Spectrometry.
- 478 Environmental Science & Technology 36(23), 5252-5260.
- 479 Meng, Y., Wang, Y., Han, Q., Xue, N., Sun, Y., Gao, B. and Li, Q. (2016) Trihalomethane
- 480 (THM) formation from synergic disinfection of biologically treated municipal wastewater:
- 481 Effect of ultraviolet (UV) irradiation and titanium dioxide photocatalysis on dissolve organic
- 482 matter fractions. Chemical Engineering Journal 303, 252-260.

- 483 Mesfioui, R., Abdulla, H.A.N. and Hatcher, P.G. (2015) Photochemical Alterations of
- 484 Natural and Anthropogenic Dissolved Organic Nitrogen in the York River. Environmental
- 485 Science & Technology 49(1), 159-167.
- 486 Miranda, M.L., Osterholz, H., Giebel, H.A., Bruhnke, P., Dittmar, T. and Zielinski, O. (2020)
- 487 Impact of UV radiation on DOM transformation on molecular level using FT-ICR-MS and
- 488 PARAFAC. Spectrochimica Acta Part A: Molecular and Biomolecular Spectroscopy 230,489 118027.
- 490 Niu, X.-Z., Harir, M., Schmitt-Kopplin, P. and Croué, J.-P. (2018) Characterisation of
- 491 dissolved organic matter using Fourier-transform ion cyclotron resonance mass spectrometry:
- 492 Type-specific unique signatures and implications for reactivity. Science of The Total493 Environment 644, 68-76.
- 494 Osburn, C.L., Handsel, L.T., Peierls, B.L. and Paerl, H.W. (2016) Predicting Sources of
- 495 Dissolved Organic Nitrogen to an Estuary from an Agro-Urban Coastal Watershed.
 496 Environmental Science & Technology 50(16), 8473-8484.
- Phungsai, P., Kurisu, F., Kasuga, I. and Furumai, H. (2016) Molecular characterization of
 low molecular weight dissolved organic matter in water reclamation processes using Orbitrap
 mass spectrometry. Water Research 100, 526-536.
- Phungsai, P., Kurisu, F., Kasuga, I. and Furumai, H. (2018) Changes in dissolved organic
 matter composition and disinfection byproduct precursors in advanced drinking water
 treatment processes. Environmental Science & Technology 52(6), 3392-3401.
- 503 Phungsai, P., Kurisu, F., Kasuga, I. and Furumai, H. (2019) Molecular characteristics of 504 dissolved organic matter transformed by O_3 and O_3/H_2O_2 treatments and the effects on 505 formation of unknown disinfection by-products. Water Research 159, 214-222.
- Qin, R., How, Z.T. and Gamal El-Din, M. (2019) Photodegradation of naphthenic acids
 induced by natural photosensitizer in oil sands process water. Water Research 164, 114913.
- Remucal, C.K., Salhi, E., Walpen, N. and von Gunten, U. (2020) Molecular-Level
 Transformation of Dissolved Organic Matter during Oxidation by Ozone and Hydroxyl
 Radical. Environmental Science & Technology 54(16), 10351-10360.
- 511 Ross, G.F., Smith, P.M., McGregor, A., Turnbull, D.M. and Lightowlers, R.N. (2003)
- 512 Synthesis of Trifunctional PNA-Benzophenone Derivatives for Mitochondrial Targeting,
- 513 Selective DNA Binding, and Photo-Cross-Linking. Bioconjugate Chemistry 14(5), 962-966.
- Ruan, X., Xiang, Y., Shang, C., Cheng, S., Liu, J., Hao, Z. and Yang, X. (2021) Molecular
- 515 characterization of transformation and halogenation of natural organic matter during the

- 516 UV/chlorine AOP using FT-ICR mass spectrometry. Journal of Environmental Sciences 102,517 24-36.
- 518 Sarathy, S.R. and Mohseni, M. (2007) The Impact of UV/H₂O₂ Advanced Oxidation on
- 519 Molecular Size Distribution of Chromophoric Natural Organic Matter. Environmental
- 520 Science & Technology 41(24), 8315-8320.
- 521 Schmitt-Kopplin, P., Hertkorn, N., Schulten, H.-R. and Kettrup, A. (1998) Structural changes
- in a dissolved soil humic acid during photochemical degradation processes under O_2 and N_2
- atmosphere. Environmental Science & Technology 32(17), 2531-2541.
- 524 Serna-Galvis, E.A., Ferraro, F., Silva-Agredo, J. and Torres-Palma, R.A. (2017) Degradation
- 525 of highly consumed fluoroquinolones, penicillins and cephalosporins in distilled water and
- simulated hospital wastewater by UV_{254} and UV_{254} /persulfate processes. Water Research 122,
- 527 128-138.
- 528 Sharpless, C.M. and Linden, K.G. (2003) Experimental and Model Comparisons of Low- and
- 529 Medium-Pressure Hg Lamps for the Direct and H₂O₂ Assisted UV Photodegradation of N-
- Nitrosodimethylamine in Simulated Drinking Water. Environmental Science & Technology
 37(9), 1933-1940.
- 532 Stubbins, A., Spencer, R.G.M., Chen, H., Hatcher, P.G., Mopper, K., Hernes, P.J., Mwamba,
- V.L., Mangangu, A.M., Wabakanghanzi, J.N. and Six, J. (2010) Illuminated darkness:
 Molecular signatures of Congo River dissolved organic matter and its photochemical
 alteration as revealed by ultrahigh precision mass spectrometry. Limnology and
 Oceanography 55(4), 1467-1477.
- Sulzberger, B. and Durisch-Kaiser, E. (2009) Chemical characterization of dissolved organic
 matter (DOM): a prerequisite for understanding UV-induced changes of DOM absorption
 properties and bioavailability. Aquatic Sciences 71(2), 104-126.
- Thomson, J., Roddick, F. and Drikas, M. (2002a) Natural organic matter removal by
 enhanced photo-oxidation using low pressure mercury vapour lamps. Water Supply 2(5-6),
 435-443.
- Thomson, J., Roddick, F. and Drikas, M. (2002b) UV photooxidation facilitating biological
 treatment for the removal of NOM from drinking water. Journal of Water Supply: Research
- 545 and Technology—AQUA 51(6), 297-306.
- 546 Wilske, C., Herzsprung, P., Lechtenfeld, O.J., Kamjunke, N. and von Tuempling, W. (2020)
- 547 Photochemically induced changes of dissolved organic matter in a humic-rich and forested
- 548 stream. Water 12(2), 331.

- Yuan, Z., He, C., Shi, Q., Xu, C., Li, Z., Wang, C., Zhao, H. and Ni, J. (2017) Molecular
 Insights into the Transformation of Dissolved Organic Matter in Landfill Leachate
 Concentrate during Biodegradation and Coagulation Processes Using ESI FT-ICR MS.
 Environmental Science & Technology 51(14), 8110-8118.
- 553 Zhang, S., Hao, Z., Liu, J., Gutierrez, L. and Croué, J.-P. (2021) Molecular insights into the
- reactivity of aquatic natural organic matter towards hydroxyl ('OH) and sulfate (SO₄⁻⁻)
- radicals using FT-ICR MS. Chemical Engineering Journal 425, 130622.
- 556 Zhang, S., Rouge, V., Gutierrez, L. and Croue, J.-P. (2019) Reactivity of chromophoric
- dissolved organic matter (CDOM) to sulfate radicals: Reaction kinetics and structuraltransformation. Water Research 163, 114846.
- 559 Zhang, Y., Zhang, J., Xiao, Y., Chang, V.W.C. and Lim, T.-T. (2017) Direct and indirect
- 560 photodegradation pathways of cytostatic drugs under UV germicidal irradiation: Process
- 561 kinetics and influences of water matrix species and oxidant dosing. Journal of Hazardous
- 562 Materials 324, 481-488.
- 563 Zhong, H., Wang, J., Jia, X., Li, Y., Qin, Y., Chen, J., Zhao, X.S., Cao, W., Li, M. and Wei,
- 564 Y. (2001) Fabrication of Covalently Attached Ultrathin Films Based on Dendrimers via H-
- 565 Bonding Attraction and Subsequent UV Irradiation. Macromolecular Rapid Communications
- 566 22(8), 583-586.
- 567

568 **Figure Captions**

- 569 Fig. 1. Changes in (a) UV_{254} and EDC of the three DOM isolates and (b) FEEM of R-HPO
- 570 arising from 30 min UV irradiation. Conditions: Incident UV intensity = 0.9 mW·cm⁻², (20.02)
- 571 $[TOC]_0 = 3.89 \pm 0.18 \text{ mgC} \cdot \text{L}^{-1}$, at unadjusted pH (4.74 ± 0.19) and room temperature (20 °C).
- Fig. 2. van Krevelen diagrams of the DOM isolates before and after UV irradiation (30 min).
 Bubble size reflects the relative peak intensity of the assigned molecules.
- Fig. 3. van Krevelen diagrams (left) and relative abundance of molecules with different
 compositions (right) for the disappeared, remained and produced molecules upon 30 min UV
 irradiation.
- **Fig. 4.** DBE versus C number map showing distribution of disappeared and remained molecules containing (a,c,e) only CHO and (b,d,f) also N or S. Bubble size reflects the relative abundance of the molecules (i.e., the number of the molecules).
- Fig. 5. Number of molecules associated with Carbon number within disappeared, remained
 and produced (a,c,e) CHO and (b,d,f) CHO+N/S community upon 30 min UV irradiation.
 Note: The remained molecules are only those with intensity decrease greater than 30%.
- **Fig. 6.** Number of molecules associated with DBE within disappeared, remained and produced (a,c,e) CHO and (b,d,f) CHO+N/S community upon 30 min UV irradiation. Note: The remained molecules are only those with intensity decrease greater than 30%.
- Scheme 1. Representative illustrations on UV-induced reactions responsible for the observed
 DOM transformation. (a). Cleavage of large DOM molecules (b). Photo-Fries rearrangement
 of phenyl ester (c) Covalent linkages between benzophenone and target molecules. (d)
 Addition of alkyl to benzotriazole structures (e) Polymerization of benzotriazole structures.
- 590 R-H represents active H-containing target molecules.
- 591

592



Fig. 1. Changes in (a) UV₂₅₄ and EDC of the three DOM isolates and (b) FEEM of R-HPO arising from 30 min UV irradiation. Conditions: Incident UV intensity = 0.9 mW·cm⁻², [TOC]₀ = 3.89 ± 0.18 mgC·L⁻¹, at unadjusted pH (4.74 ± 0.19) and room temperature (20 °C).



Fig. 2. van Krevelen diagrams of the DOM isolates before and after UV irradiation (30 min). Bubble size reflects the relative peak intensity of the assigned molecules.



Fig. 3. van Krevelen diagrams (left) and relative abundance of molecules with different compositions (right) for the disappeared, remained and produced molecules upon 30 min UV irradiation.



Fig. 4. DBE versus C number map showing distribution of disappeared and remained molecules containing (a,c,e) only CHO and (b,d,f) also N or S. Bubble size reflects the relative abundance of the molecules (i.e., the number of the molecules).



Fig. 5. Number of molecules associated with Carbon number within disappeared, remained and produced (a,c,e) CHO and (b,d,f) CHO+N/S community upon 30 min UV irradiation. Note: The remained molecules are only those with intensity decrease greater than 30%.



Fig. 6. Number of molecules associated with DBE within disappeared, remained and produced (a,c,e) CHO and (b,d,f) CHO+N/S community upon 30 min UV irradiation. Note: The remained molecules are only those with intensity decrease greater than 30%.

(a) Cleavage of linkages between aromatic structures



(b) Photo-Fries rearrangement



(c) Covalent linking of benzophenone



(d) Alkyl addition to benzotriazole structures



(e) Polymerisation of benzotriazole strutures



Scheme 1. Representative illustrations on UV-induced reactions responsible for the observed DOM transformation. (a). Cleavage of large DOM molecules (b). Photo-Fries rearrangement of phenyl ester (c) Covalent linkages between benzophenone and target molecules. (d) Addition of alkyl to benzotriazole structures (e) Polymerization of benzotriazole structures. R-H represents active H-containing target molecules.

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Sample		Number (%)	Intensity (%)	H/Cwa	O/Cwa	<i>m/z</i> wa	Cwa	DBEwa	#CHO	#CHON	#CHOS	#CHONS
S-HPOA	disappeared	192 (9.7)	7.35E+08 (3.2)	1.01	0.51	610.16	28.74	15.30	170	9	8	5
	remained	1782 (90.3)	2.21E+10 (96.8)	1.05	0.53	429.58	20.01	10.61	1719	47	9	7
	appeared	208 (10.5)	8.06E+08 (3.5)	0.93	0.50	535.88	25.47	14.76	145	49	11	3
R-HPO	disappeared	355 (12.3)	1.59E+09 (4.1)	1.06	0.49	486.18	22.50	11.94	122	184	47	2
	remained	2531 (87.7)	3.74E+10 (95.9)	1.12	0.53	422.09	19.48	9.63	1713	720	96	2
	appeared	307 (10.8)	1.18E+09 (3.1)	1.03	0.47	486.38	23.26	12.43	134	105	66	2
C-HPO	disappeared	340 (13.5)	1.71E+09 (3.9)	1.11	0.48	485.49	22.97	11.51	122	180	34	4
	remained	2173 (86.5)	4.20E+10 (96.1)	1.19	0.52	435.87	20.36	9.21	1709	322	141	1
	appeared	405 (15.7)	2.34E+09 (5.3)	0.98	0.30	465.18	25.32	14.22	148	80	170	7

Table 1. Statistics and characteristics of molecules disappeared, remained, and appeared with 30 min exposure to UV.

Graphical Abstract

