

202 years of changes in Mediterranean fire regime in Pinus nigra forest, Corsica

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1	202 years of changes in Mediterranean fire regime in <i>Pinus nigra</i> forest,
2	Corsica
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18 Abstract

Wildfires are critical socio-ecological features in the Mediterranean basin. In a context of global 19 changes (climate, land use), we questioned whether the wildfire regime was altered in the 20 mountains of Corsica, France. Using tree-ring analysis of fire-scarred trees, we tested for 21 changes in frequency, seasonality, area. We hypothesized that the fire regime changed during 22 the middle-20th century due to human activities, as observed elsewhere in the Mediterranean. 23 We sampled fire-scarred trees, geolocated for mapping fire areas, in a forest of black pine (Pinus 24 *nigra laricio*). The oldest fire was in 1684 but the fire chronology with adequate sample depth 25 for analysis covered 202 years [1820-1921]. Between 1820-2012, 15 fires were recorded, 8 of 26 27 which scarred at least 25% of the sampled trees. The mean fire interval was 14 years, corresponding to a high fire frequency with 4 major fires per century. Most fires occurred 28 between 1931-1970. On average, about 50% of trees were scarred by fires before 1931, but this 29 30 percentage decreased thereafter. The exception was the 2000 fire that impacted 100% of living trees. Mapping showed spatially heterogenous fire areas. These results match other 31 Mediterranean studies showing longer fire intervals since the late-20th century, and wildfires 32 generally occurred during the period of late-earlywood or latewood formation, i.e., summer or 33 early autumn, which is the season of contemporary fires. Although fires were recurrent for more 34 35 than 200 years with no change in the fire season, the regime changed twice in frequency. These changes likely result from a combination of land use and warmer summer conditions. While 36 pines survived most past fires of low intensity through the last large fire in the 1970's, the severe 37 fire of 2000, following ca 30 years of fire suppression, killed large patches of mature trees. Such 38 chronology provides rational arguments for black pine ecosystem management, notably to 39 consider seeking to recover the sustainable fire regime of the 19th century, for instance by using 40 prescribed fires to manage surface fuel. 41

- 42 Key words: climate; dendrochronology; fire frequency; fire season; fire area; fire scars; land
- 43 use; management; mountain; Pinus nigra laricio; tree-ring

44 **1. Introduction**

Wildfire is a key component of ecosystem functioning at global scale (Bond and Keeley, 2005; 45 Sayedi et al., 2023). The ecological roles of wildfire depend on fire regime parameters such as 46 fire intervals, intensity, and seasonality, which are related to the frequency, extent, and severity 47 of fires (Johnson, 1992; Keeley, 2009; Pausas and Ribeiro, 2013). Globally, fire regime shifts 48 are driven by complex interplays between climate and land uses as reported in Mediterranean-49 climate areas of Australia, Europe and North America (Moritz et al., 2014) and at a global scale 50 as well (Andela et al., 2017). Because fire regimes are under control of both climate and land 51 uses over millennia, they are thus a socio-ecosystems component (sensu Liu et al., 2007), 52 resulting from coupled natural and social drivers, including legacies of past conditions, 53 producing multivariate responses of fire behaviour (e.g., Niklasson and Granström, 2000; 54 Guiterman et al., 2019). 55

56 The Mediterranean Basin is fire prone due to dry and warm summers, as well as fire-adapted native species (Pausas et al., 2008). Climate models predict that the late-21st century will be 57 warmer and drier, including in winter, leading to more fire years and a broader fire season in 58 the region (Ozturk et al., 2015; Turco et al., 2018). Mediterranean ecosystems of southeastern 59 60 France have already experienced changes in fire regimes since the 1970's through the mid-1990's, with more burned areas and higher yearly and seasonal fire occurrences (Fréjaville and 61 Curt, 2015). However, since 1986 (Ruffault and Mouillot, 2015) or the mid-1990's (Fréjaville 62 and Curt, 2017), the trend has changed suggesting a partial disconnection between fire regime 63 and climate, or the effects of other drivers such as policies of land management. Similar trends 64 have been recorded in other Mediterranean countries such as Spain (Pausas and Fernandez-65 Munoz, 2012), although not in north Africa (Chergui et al., 2018). Reconstructing and 66 decomposing long-term fire regime characteristics (frequency, season, areas) remains helpful 67

for forest managers to constitute baselines for analyzing current trends in light of past recentregimes.

70 In the Mediterranean basin, black pine (Pinus nigra Arnold) is common in many countries where it is a dominant tree, considering that its large biomass drives most ecosystem processes 71 (the mass ratio hypothesis sensu Grime, 1998). Black pine has been also abundantly planted 72 since late 19th century (hereafter c.) for timber production and soil protection, resulting globally 73 in 3.5 million hectares of new forests (Isajev et al., 2004) sometimes for assisted migration 74 75 (Michalet et al., 2023). Its management in the face of climate and fire changes would benefit from more precise baselines on historical fire regime of this region over time. Detailed baselines 76 on historical fire regimes in P. nigra forests have started to emerge (Fulé et al., 2008; 77 78 Christopoulou et al., 2013; Sahan et al., 2022), providing site- and time-specific reconstructions of past fires. All these studies showed that fire intervals were shorter during the late-19th or the 79 early-20thc., and that the fire intervals became longer in the late-20thc. Generally, in Europe and 80 81 the Mediterranean basin, the 19thc. was a period of intense rural and forest land uses for domestic herding and wood production (Grove and Rackham 2001). Such intense usages could 82 explain the more frequent fires during the 19thc., despite a cooler climate (Büntgen et al., 2013; 83 Luterbacher et al., 2016). However, the generalized land-use abandonment in the 84 Mediterranean, varying somewhat among countries and regions (Chauchard et al., 2007; San 85 Roman Sanz et al., 2013; Fortuny et al., 2014), altered risks of ignition. At the turn of the 21stc., 86 global warming enhanced a trend for more severe fires, indirectly fostered by land-use 87 abandonment that triggers fuel build-up at stand scale and forest connectivity at landscape scale 88 89 (Ganteaume and Jappiot, 2013, Mouillot et al., 2003).

A growing body of dendrochronological literature reported extensive surface fire in *P. nigra* massifs around the Mediterranean (Fulé et al., 2008; Christopoulou et al., 2013; Camarero et al., 2018; Sahan et al., 2022). However, Corsica is still an area of gap of knowledge on fire

regime, except for a multimillennial study based on sedimentary charcoal (Leys et al., 2014) 93 reporting that fire was closely associated to black pine for more than 12,000 years. Szymczak 94 et al. (2020) carried out a fire study based on tree-ring analysis of 10 fire-scarred cross-sections 95 supplemented by increment cores and an administrative-based study in lowland forest of oaks 96 and maquia was published by Mouillot et al. (2005). Therefore, the present study aims to test 97 the hypothesis that the fire frequency has changed since the 19thc in forests of *Pinus n*. subsp. 98 laricio (Poir.) Maire (hereafter P. nigra). The study is based on dendrochronological analysis 99 of fire scars (Arno and Sneck, 1977; Zackrisson, 1977; Huffman et al., 2020). 100

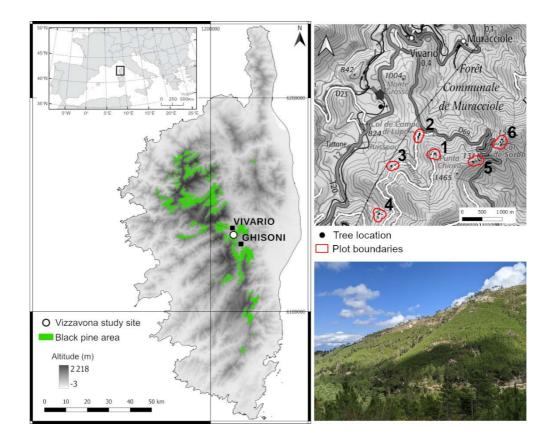
On the 27th August 2000, a substantial area of lethal crown fire occurred near Vizzavona 101 (central Corsica), reported to have burned a total of 3900 ha, i.e., 10% of black pine cover in 102 103 Corsica. We selected this site to determine if fire regime has changed significantly during past 104 centuries, and if there is evidence of drivers supporting the eventual change in fire regime in an area that experienced changes in its human demography. The main objective of the present 105 106 study is to reconstruct the longest fire chronology, the temporal, seasonal and spatial variability in this central Corsican forest and, to a lesser extent, explore possible links between human 107 demography and fire activity. We hypothesized that changes in fire regime were linked to local 108 trends in human demography, and fire season was stable, occurring during the warm/dry 109 summer season (climate driver). All these fire parameters will be helpful to establish the 110 historical range of variability of fire regime useful to delineate baselines of Corsican black pine 111 forest management. 112

113

114 **2.** Materials and methods

115 **2.1 Study site**

Corsica is a French Mediterranean island, located west of the Italian peninsula and southeast 116 of southeastern mainland France (Figure 1a). Corsica is mountainous with several summits 117 above 2000 m above sea level (a.s.l.). The study site, hereafter the "Vizzavona forest", is 118 situated in central Corsica (42.14 N, 9.19°E). It corresponds to the state forest of Rospa-Sorba, 119 and the two communal forests of Vivario and Muracciole, all mostly used for timber production, 120 but not where the sampling was carried out. There are few semi-feral cattle extensively grazing 121 122 in the study forest. One shepherd sometimes uses the forest for his goats. Vizzavona forest is also an emblematic site with high touristic and patrimonial value, located at the junction 123 between north and south parts of the island. The bedrock is composed of igneous rocks or 124 granites. Vizzavona forest soils are 35-65 cm deep brunisols, slightly acidic (pH: 5.5 to 6.5) 125 sandy-silty, respectively 60% and 40% of sands and silts (IFN report 2004), resulting a low 126 water storage. 127



129

Figure 1. (a) Map of Corsica displaying black pine distribution (green) according to the French forest inventory (IFN data, 2017); the white dot represents the location of the study site or Vizzavona forest; Vivario and Ghisoni are the two nearest villages of the study site, situated on both Sorba pass sides; coordinates in meters are according to the Lambert 93 French grid. (b) Local map of sampling area, and plots in red. (c) Scenery illustration. *(figure in color)*

135

At Vizzavona, the climate is Mediterranean with hot summers and cool winters, but the orography results in cold seasonal conditions, notably in winter with 4 to 5 months of intermittent snow cover. During the 1990-2020 period (nearest weather station at 772 m a.s.l.; 42.152608°N, 9.149188°E), mean annual precipitation was 1265 ± 291 mm with a mean annual temperature of 9 ± 0.6 °C. Mean monthly minimum precipitation and maximum temperatures were both recorded in July (32 ± 3.6 mm; 17.8 ± 2.9 °C) and August (33 ± 3.2 mm; 17.7 ± 2.8 °C). Mean monthly precipitation peaked in November ($208 \pm 16.4 \text{ mm}$), including snowfall, and monthly minimal average temperatures occurred in February ($1.9 \pm 3.5^{\circ}$ C).

Regarding wildfires in Corsica, the Prométhée fire database between 1973-2022 (https://bdiff.agriculture.gouv.fr/) indicates that 55% of fires occurred between July and September, including 39% in July and August (Supplementary Information Figure S1); 95% of total area burned between July and September, especially between July and August (86%). In black pine forests, the main fuel is composed of shrubs, i.e., *Erica arborea* below 1400 m a.sl. (max. 3-m tall) and *Juniperus sibirica* Lodd. Ex-Burgsd from 1400 to 1600 m a.s.l. (< 1-m), saplings of pines, a thick litter layer of pine needles, and woody debris (Figure S2).

151 **2.2 Material**

In Corsica, black pine covers 372 km² (Figure 1a) mostly above 1000 m up to 1800 m a.s.l. Its lifespan is potentially over 800 years (Szymczak et al. 2014). *P. nigra* needs a strong annual precipitation (800 to 1200 mm.yr⁻¹) but supports dry summers and colonize soils with low water storage capacity. Trees can reach 40 m in height. They have a thick bark and a high crown base protecting adult trees from fires (Fréjaville et al., 2018b). Because it does not have serotinous cones, seedling regeneration is limited after severe fires (Pausas et al., 2008).

158 2.3 Field sampling

We selected six plots within the study area (Figure 1b) according to three main parameters: (i) presence of large black pines (assumed old trees) to sample multi-scarred trees and maximize the probability to obtain the longest fire chronology; (ii) areas characterized by a dense treecover of *Pinus pinaster* or *Fagus sylvatica* were excluded because they both present different flammability traits (Fréjaville et al., 2018b) and niche (Fréjaville et al., 2018a); (iii) due to severe relief, plots needed to be close to paths for transport of sampling equipment and wood samples. Mean plot slopes ranged from 45 to 80%, and aspect varied from north to west, following an altitudinal gradient ranging from 1000 m to 1400 m. (Table 1). Plots were
relatively homogeneous with respect to topography and vegetation between and within them.
Depending on the plot, there could be regeneration of *P. nigra* only, or both *P. nigra* and *P. nigra* and *P. pinaster* (VZ02); in more open areas, there was dense cover of tree heather (*Erica arborea* L.;
VZ03).

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Table 1: Plot characteristics (mean values) of the study site (Vizzavona). Plot areas were
calculated by creating a 100 m buffer around sampled tree points. Coordinates correspond to
the World Geodetic System 84. Understory vegetation: PN: *Pinus nigra*; PP: *Pinus pinaster*(Aiton); PA: *Pteridium aquifolium* (L.) Kuhn; EA: *Erica arborea* L.; FS: *Fagus sylvatica* L.;
CS: *Castanea sativa* Mill. (cf. Figure S2).

Plots	Latitude	Longitude	Elevation	Slope	Aspect	Area	Understory
(VZ21)	(°)	(°)	(m a.s.l.)	(%)	•	(ha)	
VZ2101	42.1453800	9.1744550	1250	45	Ν	6.6	PN and PP saplings,
							EA, PA
VZ2102	42.1497071	9.1695536	1040	43	W	6.3	Abundant saplings
							of PP; EA
VZ2103	42.1434320	9.1606060	1100	43	Ν	5.5	Abundant saplings
							of PP; EA
VZ2104	42.1327850	9.1556100	1050	46	NW	7.8	PP, PN, PA, EA,
							Scattered FS, CS
VZ2105	42.1430427	9.1859809	1400	54	N-NW	7.6	PN, EA
VZ2106	42.1468027	9.1941118	1370	79	NW	8.7	Scattered EA

177

We selected trees presenting the highest number of visible scars on living trees, stumps, logs 178 or snags to establish a thorough fire history (Farris et al., 2013). We used chainsaws to remove 179 small samples of wood from standing trees (living or dead), partial cross-sections that 180 represented <10% of a tree's cross-sectional area, a sampling approach that presents minimal 181 risks of tree breakage (Heyerdahl and McKay, 2008). Samples were labeled and wrapped with 182 183 plastic film for transport to the lab. For every sample, we measured diameter at breast height $(\sim 1.3 \text{ m})$, latitude, longitude, elevation, slope, slope aspect, aspect of the scar, and the number 184 of suspected scars. Each sample was photographed. 185

Further, we sampled ten unscarred old trees in the VZ05 plot using increment borers to develop tree-ring chronologies on trees that were obviously not scarred, helpful for cross-dating and for comparison in addition to a reference chronology of *P. nigra*, extracted from the International Tree-Ring Data Bank (www.ncei.noaa.gov/access/paleo-search/study/4389; author: Schweingruber, 2002). For this purpose, two cores were sampled per tree at 1.3 m height, one on the uphill side of the trunk and one at a 90° angle."

192 **2.4** Tree-ring crossdating and fire scar dating

In the lab, wood samples were air-dried for at least one month, then sanded with progressively 193 194 finer grits from P40 to P400, to obtain a smooth surface with cells visible under microscopes (×100) (Stokes and Smiley, 1968). After surfacing, we digitized samples using an A3-format 195 scanner EPSON expression 12000XL at 1200 dpi. We crossdated series with the reference 196 197 chronology visually and with graphical assistance from CDendro© (Cybis.se) software (Maxwell and Larsson, 2021). Ring-widths were measured of scanned images in Coorecorder© 198 software (Cybis.se) and we used Cofecha software (Holmes, 1983) to carry out quality control 199 200 of crossdating.

The Schweingruber ring-width chronology (FRAN027) covered the period AD 1518 to 1980 (Figure S3). We measured and crossdated cores from living unscarred pines then we added them to Schweingruber's reference chronology in order to update the chronology to 2021 (Figure S3). The new chronology was used to crossdate scarred trees. The R package dplR (Bunn 2008) was used to calculate dendrochronological statistics of this chronology (master series length, series intercorrelation, mean sensitivityetc.).

We identified fire scars to the year and seasonal position within tree-rings using the designations of Baisan and Swetnam (1990): dormant season (D), earlywood (EE), midearlywood (ME), or late-earlywood (LE) and latewood (LW). The code U for undetermined was used to specify when we could not locate the scar position in the ring, usually in the case

of very narrow rings, decayed or fragmented samples (when scars are located at cracks). Pinus 211 *nigra* growth starts in mid-April when temperatures are increasing coupled with available water 212 (rain or melted snow). Growth slows in June and can be interrupted in early July when 213 temperatures become too high and dryness too severe (Haüsser et al., 2021). Thus, we defined 214 the dormant season as the period between autumn and the following spring. Based on the 215 overwhelming prevalence of summer to early-autumn fires reported in black pine forest since 216 217 1973 when the French national fire database, Promethee, began (www.promethee.com), we assumed that dormant-season scars primarily formed in the late summer (August) or early 218 autumn (September). However, we cannot rule out wildfires in spring or in winter prior to 219 220 earlywood formation if conditions of ignition (lightning or people) and spread (dryness) were 221 appropriate in winter; occurrences of winter fires have increased since the 1990's (Fréjaville et al., 2015). 222

We plotted composite fire chronologies for each plot beginning with the first fire event with a minimum of 2 recording samples. We also inventoried fire dates according to two different thresholds (25 and 50%) resulting in three classes of affected trees or severity: fire year affecting less than 25% of scarred samples, fire years affecting between 25 and 50% of scarred samples and fire years affecting 50% or more scarred samples.

228 **2.5 Mapping**

For mapping scarred trees, coordinates of each sample were measured with a Garmin GPS through the World Geodetic System 84 (WGS84) providing a positioning accuracy of 3-5 m. The plot areas were retrospectively estimated from GIS locations of sampled trees by defining a 100 m buffer zone around each geolocated tree. It is common to map locations of trees scarred by specific fires as an estimate of the spatial extent of fire (Farris et al., 2010). When samples have been collected over a systematic grid, maps based on Thiessen polygon tessellations have been shown to provide accurate reconstructions (Farris et al., 2010). In the case of irregularly distributed samples, however, Thiessen polygons may create unusual shapes that are artifacts
of the technique. Therefore, we estimated burned areas using a neutral kriging approach (Fulé
et al., 2023), i.e., quadratic Kernel densities, called the "heatmap" tool in QGIS. We developed
a map for each fire year. Densities of probabilities were calculated based on a spreadsheet
containing binary values for each dated sample: 1 for fire year, and 0 without fire. Kriging
parameters are explained in SI, section 4 and figure S4.

We merged all probability density functions for all fire years to provide a general probability of past fire occurrences corresponding to the reconstructed fire history. Probability densities were weighted by the total number of scars for each dated sample. This map makes it possible to estimate the wildfire pattern at the landscape level.

Note that the interpretations of burning probability are only accurate between plots and are contingent on the selection of plot locations. Thus, the kriged maps are useful for relative comparisons between fire years, but do not provide landscape-scale fire probability information.

250 **2.6 Human demography data**

We evaluated human demography as a proxy of land uses. Data were extracted from the INSEE 251 252 (Institut national de la statistique et des études économiques) and EHESS (Ecole des Hautes *Études en Sciences Sociales*) databases that provided population census figures from 1800 to 253 2019 (Figure S5). Censuses of human populations were made every 5 years between 1831 and 254 2006 and each year since 2006. Before 1831, census intervals were not constant: 1800, 1806 255 and 1820. For the present study, we extracted data for all of Corsica for the general trend, as 256 well as the local trends reported for the village of Vivario (located on the northwest slope of 257 Sorba pass), and for the village of Ghisoni (located on the southeast slope of Sorba pass). We 258 also compared the dates of fires with historical events of World War 2 (WW2). 259

260 2.7 Data analysis

We used the R package burnr (Malevich et al., 2018) to calculate fire regime metrics for each plot such as composite fire return intervals (FI), their arithmetic mean (MFI), median (median FI), and the Weibull median probability interval.

Regarding FIs, we tested based on the two-samples Wilcoxon rank sum test whether they were significant differences of mean burned areas and fire frequencies between periods characterized by change point assumed tipping point. This method was used as fire intervals per period were small (n < 10) and non-parametric methods better suit this type of dataset than t-tests requiring normal distribution and larger sample size.

The Weibull calculation (eq. 1) provides the *c* parameter, which characterizes the shape of the curve characterizing the flammability risk trend (McCarthy et al., 2001). A *c* value < 0 indicates a decrease of the flammability over time (*t*), *c* value > 0 indicates that the flammability (*h*) increases over time (*t*), while c = 0 indicates that the flammability is constant through time, such as:

$$h(t) = ht^c \qquad \text{eqn. 1}$$

275

276 **3. Results**

On average, 11 trees were sampled per plot, ranging from 8 to 14 (Table 2). A total of 64 scarred
samples was collected in the Vizzavona forest, totaling 196 dated fire-scars, in a total survey
area of 43 ha (Table 1).

The ring-width chronology we constructed by updating Schweingruber's original FRAN027 chronology contained 30 series and covered the period 1518 to 1980. The overall correlation (rbar) was 0.57, mean sensitivity0.21, and EPS 0.96.

		Crossd	ated samples	Chronology (year AD)		
Plot	Sample number (<i>n</i>)	<i>(n)</i>	(%)	Starting	Finishing	
1	8	7	88	1819	2021	
2	14	9	64	1855	2021	
3	10	8	80	1824	2021	
4	10	8	80	1796	2021	
5	11	9	82	1663	2021	
6	11	7	64	1631	2021	

Table 2: Number of samples (tree, stump, snag, log), total of cross-dated samples and time

284 boundaries of chronologies

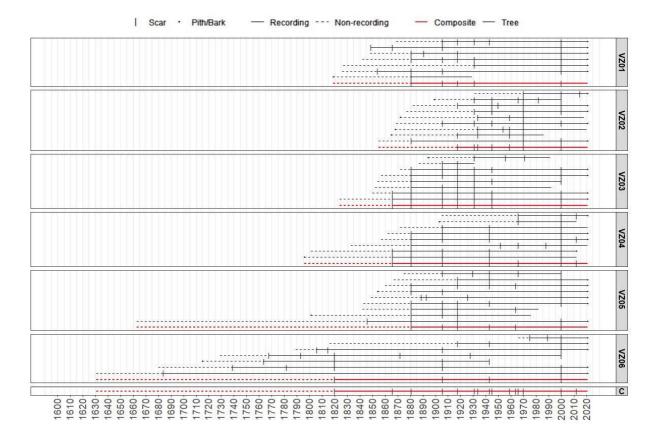
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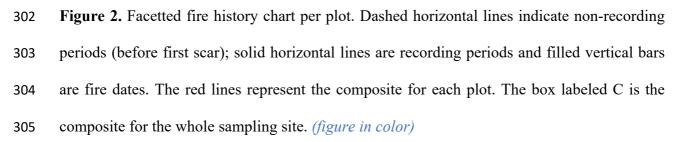
287 **3.1 Fire records**

Out of the 64 samples collected, 48 were successfully crossdated (75%), resulting in a mean 288 of 8 crossdated samples per plot (Table 2). The samples that could not be crossdated were 289 generally due to too much decayed wood (due to fungi or ants) or samples composed of too 290 many broken sections leading to unusable series containing not enough number of rings. Our 291 whole chronology covered a 390-year period, from 1631 to 2021. The oldest fire was dated to 292 1684 (VZ06) and the first fire present on more than one tree occurred in 1820. From 1820 293 294 onward, the 196 fire-scars recorded corresponded to 15 fire years (Figure 2). Fire interval analyses were calculated beginning from the first fire year with a minimum of 2 samples 295 recording fire events. The common period for analysis was thus 1820–2021 (202 years). 296

Our chronology and associated frequency changes partly matched administrative records (Promethee, ONF, OEC databases, period 1950-2019). Two fire events were found in both fire records and the fire scars (1970 and 2000) while others, such as the 2012 fire in VZ04, were identified by dendrochronology only (Figure 2).







306 3.2 Fire interval analysis

Between 1820–2021, a total of 15 fires was reported among the different plots covering 43 ha 307 in total (Figure 2). The minimum fire interval per plot was 3 years (VZ02) and the maximum 308 was 86 years (VZ06; Table 3). At the scale of the study site (43 ha), the minimum interval 309 between two occurrences was 2 years, and the maximum was 46 years (Table 3). The mean fire 310 interval (MFI) of 14 (±13) years is close to the median interval of 12.0 years and the Weibull 311 median probability interval of 10.7 years as well. Eight fires affected at least 25% of sampled 312 trees in 1820, 1866, 1881, 1906, 1918, 1931, 1943 and 2000. Considering this filter, minimum 313 and maximum FI were 12 and 57 years respectively, MFI of 26 ± 19 years and Weibull median 314

probability interval of 23.2 years (Table 2). Five fires affected at least 50% of recorded samples

316 (1820, 1881, 1906, 1918, 2000) with a minimal FI of 12 years, a maximum of 82 years, a mean

 $(\pm sd)$ of 45 ± 32 years and a Weibull median probability interval of 40 years. Regarding the 8

major fires (>25% of samples with fire-scars), all occurred before 1943, except the one of 2000.

320	Table 3. Fire return intervals (FI) per plot and for the whole study site (Vizzavona): total
321	number of fire intervals, MFI, median FI and Weibull median, minimum and maximum values
322	per plot and for the whole study area. The calculation of FI is bounded at 2012 (last fire date
323	inventoried) because the period between 2012 and the sampling date (2021) is not a fire interval.
324	"Major fires" refer to fire events scarring at least 25% of samples. Rule for brackets:]1931-
325	1970], means 1931 is not included in the period, and 1970 is included.

	Fire-return interva				I, years)	
Plots and site	Intervals	ervals Mean		Weibull	Min	Max
r iots and site	no (<i>n</i>)	(±sd)	Median	median	interval	interval
VZ01	4	29.8 ±26.8	19.0	25.4	12	69
VZ02	6	$13.7 \pm \! 8.9$	12.0	12.5	3	30
VZ03	6	$22.3 \pm \! 16.7$	14.5	20.2	12	55
VZ04	6	$24.3 \pm\! 10.0$	24.0	24.2	12	37
VZ05	5	$23.8 \pm \! 8.6$	25.0	23.8	12	36
VZ06	3	$60.0\pm\!\!24.6$	57.0	60.2	37	86
Vizzavona (whole site, all fires)	14	13.7 ± 12.5	12.0	10.7	2	46
[1820-1931] (112 yrs) <i>c</i> =1.89	5	$22.2\pm\!\!14.3$	15.0	20.8	12	46
]1931-1970] (39 yrs) <i>c</i> =1.48	7	5.6 ±4.4	4.0	4.9	2	14
]1970-2012] (32 yrs) <i>c</i> =n/a	2	n/a	n/a	n/a	12	30
[1820-2012] (major fires) <i>c</i> =1.64	7	$25.7\pm\!\!18.5$	15.0	23.2	12	57

The FI trend shows that three significant fire periods occurred over the past 202 years (figure 326 3a): [1820-1931],]1931-1970] and]1970-2012]. Two of these periods: [1820-1931] (n = 5)327 intervals), [1931-1970] (n = 7) were statistically different (Wilcoxon test, p = 0.015). The MFI 328 for the period [1820-1931] was 22 ± 14 years (Weibull median = 21 years, min = 12, max = 46 329 years), MFI for the period 1931-1970 was 6 ± 4 years (Weibull median of 5 years, min interval 330 = 2 years, max interval = 14 years). For the last period 1970-2012, FI values could not be 331 calculated as it did not contain enough fire intervals (n = 2). The c parameter inferred from 332 Weibull calculus indicates that it was 1.89 ± 0.63 (unitless) for [1820-1931], and 1.48 ± 0.42 for 333 1931-1970. Like the MFI for [1970-2021], the c parameter could not be calculated. Over the 334 335 whole period, [1820-2021], the c parameter was 1.18 ± 0.24 . Both c values for first and second period indicated that the flammability increased over time, but this increase was more important 336 for the first period (1820-1931) characterized by longer FI (Table 3). We also tested if the end 337 338 of the second world war in 1945 could be a tipping point. We found that fire frequencies were not different before and after this year in the Vizzavona forest, i.e., MFI 16 \pm 14 yrs versus 11 339 ± 10 yrs (Wilcoxon test: p = 0.65). 340

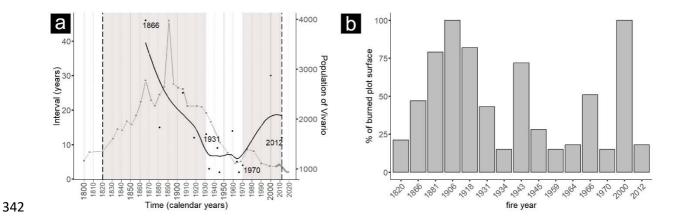


Figure 3. (a) Fire interval duration plotted against fire years. Red line is the local regression
(also known as Loess method), averaging interval duration variability. It is a non-parametric
regression approach performing least squares regression based on a span of 0.75. Background

shading distinguishes the period defined by a type of duration interval variation (Wilcoxon test, p = 0.015): tipping points of FI are 1931 and 1970. Vertical black dashed lines represent the study temporal boundaries [1820-2012]. (b) Percentage of burned areas per fire year. *(figure in grey scale)*

350

351 3.3 Fire season

Fire season could be identified on 72% of the fire scars, i.e., 142 out of 196. We found that 71% of scars were formed during latewood (55%), late earlywood (2%) or the dormant season (14%), i.e., during the summer to autumn seasons (Figure 4). We identified 2% of scars in midearlywood and no scars in earlywood (dataset available in Table S1).

356

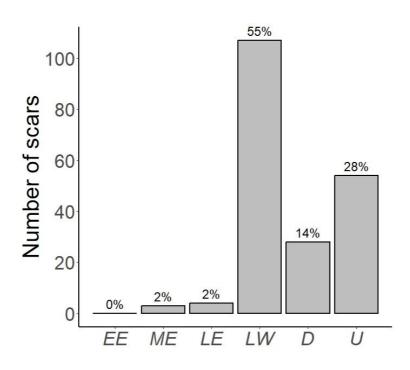


Figure 4. Fire seasonality summary showing the positions of scars in the tree-rings. EE: early
earlywood; ME: middle earlywood; LE: late earlywood; LW: latewood; D: dormant period; U:
undetermined. *(figure in grey scale)*

361

362 3.4 Fire extent

Kriging-based reconstructed areas for each of the 15 fire years recorded (Figure 5a) shows 363 that in the case of major fires, probability densities of buffers were connected because a large 364 part of the area experienced a common fire year, whereas in the case of minor fires, one or a 365 few plots were affected. We also calculated the percentage of burned surface over total sampled 366 area burned by fire year (Figure 3b; values in Table S2). However, the oldest fire of 1820 was 367 only encountered in plot VZ06; because trees were not yet recruited in other plots, it is not 368 369 possible to determine if the 1820-fire was large enough to affect other plots. Therefore, the area of the 1820-fire is a minimum area and must be considered with caution. 370

When we consider burned surfaces, we observe a difference between the [1866-1931] and [1931-1970] periods. All fires between 1866 and 1931 burned at least 50% of the whole area, but after 1931, only three out of nine fires burned more than 50% of the sampled area. The difference is significant between [1866-1931] (n = 5, 70 ±24% of the sampled area) and]1931-1970] (n = 7, 31±22%) periods (Wilcoxon test, p = 0.03) but not between [1866-1931] and]1931-2012[(n = 9, 38 ±29%) periods (p = 0.07).

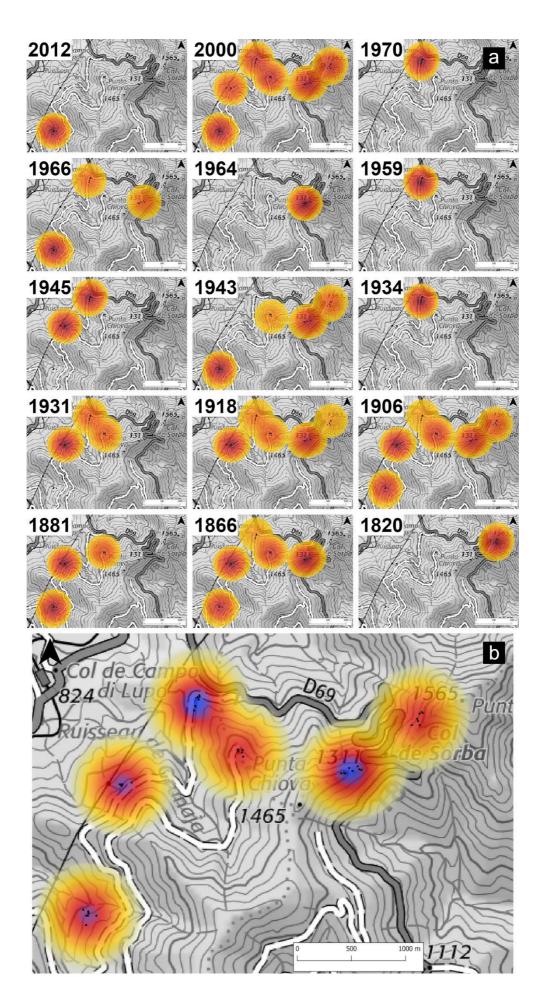


Figure 5. (a) Spatial pattern of all 15 recorded fires. Black dots represent cross dated samples and red-to-yellow chart indicates kriged fire areas. Densities of probabilities of fire occurrence in the past fire regime based on the number of scarred trees are displayed according to a red gradient to yellow chart (the deeper the red, the higher the probability); (b) probability density function integrated for all 15 fires over the past 202 years (kriging), from yellow (lower probabilities) to blue (higher probability). Black dots represent dated samples. *(figure in color)*

385

Based on the merging of all the 15 fire years and the probability density assessment, it appears 386 387 that two main locations burned more frequently in the past (Figure 5b): first the VZ01 area corresponding to a slope exposed to northwest and second, in the east, the VZ05 area, in the 388 northwest of the Sorba Pass, on the top of a talweg. The southwestern part of the study area 389 390 (plots VZ03 and VZ04) is an area where fewer past fires burned, even though the 2000 fire spread there, but at lower severity. From personal observations, this southwestern part is 391 intermingled by small stream (grey dashed line on Figure 5b), and a few beeches (F. sylvatica, 392 less flammable species). 393

394

395

4. Discussion

Like other *P. nigra* forests in the Mediterranean basin, the Corsican black pine forest of Vizzavona was frequently subject to fire events since the 17th century (Figure 2). In Greece, oldest evidence of fire scars was dated of 14thc. (Touchan et al., 2012), whereas in Anatolia they were from 16thc. (Sahan et al., 2022). However, the established variability in frequency and severity at Vizzavona covers the last 202 years, from 1820 to 2012, i.e., first and last events (Table 4). Wildfire regimes have changed in frequency at least twice since 1820. Finally, the most recent fire in Vizzavona, in 2000, was much more severe than all the previously recorded fires that were characterized by numerous surviving scarred trees. This knowledge of fire regimes is useful for land and forest managers, whether it is to plan fire control policies (Alcasena et al., 2019) or in ecosystem management that bases sustainable forest management on natural disturbance regimes through Nature-Based Solution processes (Gauthier et al., 2009).

Table 4. Synthesis of fire regimes in natural black pine forests in the Mediterranean basin, inferred from fire-scars and tree-ring analyses; studies are presented from west to east; fire frequency was calculated only on major fires, which are those that scarred 25% of trees, over the 19th, 20th and 21st centuries that is the comparable covered period of all studies.

Country	Fire year Fire years number			Frequency of major fires			Reference	
	First	Last	All	Major	Period for frequency	Number of years	Fire.years ⁻¹	
Spain	1834	2003	11	n/a	1834-2005	171	0.064	Fulé et al., 2008
Corsica	1684	2012	15	8	1820-2012	193	0.041	Present study
Corsica	1800	2000	6	n/a	-	-	-	Szymczak et al., 2020
Greece	1309	1891	16	n/a	-	-	-	Touchan et al., 2012
Greece	1801	2007	38	11	1845-2007	165	0.067	Cristopoulou et al., 2013
Anatolia (Turkey)	1726	2005	88	12	1800-2005	205	0.059	Sahan et al., 2022

413

414 **4.1 A 202-year chronology with changes in fire frequency**

The Corsican chronology cover 202 years, similar in coverage with the studies in southern Greece (Christopoulou et al. 2013) and eastern Spain (Fulé et al., 2008). The only longer chronology is the one in Anatolia with 553 years (Sahan et al., 2022). In total, we found 15 fire years in 193 years, including 8 major fires that result a frequency of 4.1 major fires per 100 years (Table 4). These values are little lower than those found elsewhere in the Mediterranean basin since the 19th century: 5.9 fires.100 yrs⁻¹ in Anatolia, 6.7 fires.100 yrs⁻¹ in Greece; in Spain, 6.4 fires.100 yrs⁻¹ is also close to the general trend if, in this case, we consider all fire events and not major fires only (Table 4). The comparison of fire intervals with the other studies
in Corsica (Szymczak et al., 2020) or Greece (Touchan et al., 2012) was not possible as they
did not explore this aspect in depth, since they had a limited number of sampled trees, or only
one scar per fire. The Mediterranean-wide results correspond to an average frequency of one
major fire every 18 years in black pine forests since the beginning of 19th c., meaning that the
overall mean frequency has been high during the past centuries.

However, the fire frequency changed twice at Vizzavona, first during the 1930's and second 428 429 after 1970 (Figure 3a). Between 1931 and 1970, the frequency was high, characterized by a short MFI of 6 \pm 4 years [2-14 years], while it was 22 \pm 14 from 1820 to 1930 (Table 3). In 430 Anatolia, the fire frequency has also changed twice over the last 200 years, first around 1852 431 432 corresponding to an increase in frequency and then in 1934 with a drop in frequency (Sahan et 433 al., 2022). In Spain, the frequency changed four times (Fulé et al., 2008). Finally, in southern Greece, the trend is less obvious; the 100-year long period before 1945 had lower frequency 434 435 and the period since 1945 had an increase in frequency. The years after 1853 seemed notable both in Anatolia and Spain, corresponding to a brief period of increase in frequency. It partially 436 matched increasing frequencies in Corsica (Figure 3a) between 1820 and 1931. This year of 437 1931 characterized a tipping point in our study as frequency increased from 1820 to this date 438 (Figure 3a). During the 1930's, this trend of frequency was similar in Spain (Fulé et al., 2008) 439 440 but is the reverse in Anatolia where it corresponded to a drop in frequency (Sahan et al., 2022). These substantial differences in timing of increasing/decreasing frequencies between studies 441 mean that local trends cannot be upscaled to the entire Mediterranean basin; each trajectory has 442 local drivers, because of site-specific variability (altitude, average precipitation, pastoralism, 443 etc.) and local fire suppression policies (timing of initiation of fire suppression during the 19th 444 c. and level of restrictions of laws). 445

446 **4.2 Potential drivers**

A common pattern corresponds with the Second World War (WW2): 1943 in Corsica corresponded to severe fights the 9th September and the 4th October between the German forces and the Allied forces that led to the bombing of forests by German army to cover their retreat and as an act of retaliation against Corsican resistors (Ministry of Armed Forces, France: www.cheminsdememoire.gouv.fr/fr/la-liberation-de-la-corse-9-septembre-4-octobre-1943).

452 Similarly, in southern Greece, fire events occurred at the time of multiple battles in 1944 and
453 1945, burning respectively 38% of samples during the first year and 49% of samples during the
454 second year (Christopoulou et al., 2013).

After the WW2, the population began to drop in Corsica (Figure S5). The village of Ghisoni 455 (figure 1a) followed this demographic trend, while the village of Vivario (figure 1a) followed 456 457 a different trajectory as population started to fall earlier, during the 1900's (Figure S5). The timing of these two demographic decreases was different and did not coincide with change in 458 fire frequency at Vizzavona. Furthermore, the fire frequency was not different before and after 459 460 1945 in the Vizzavona forest. This indicates that the WW2 did not change the fire frequency, and 1945 was not a tipping point. However, out of the 8 major fires, 7 happened before 1945, 461 while only one major fire occurred after, i.e., 2000. According to Gilbert (1989), abandonment 462 of farming lands started around 1920's-1930's and ended around the 1950's. First, 463 abandonment of farming lands led to an increase in slash-and-burn practice to increase pastoral 464 465 resources. Second, once abandonment of farming was definitely established, fire numbers decreased followed by a biomass accumulation. Thus, the weight of drivers changed between 466 the first and second part of the 20thc. 467

The Vizzavona area is a key socio-economic area in Corsica. The Sorba pass connects the center to the southeast of Corsica, whereas the Vizzavona pass connect the northeast to the southwest of the island (Figure 1). Although local population decreased, the usage of the Vizzavona area for socio-economic purposes has remains significant, and provides a source of

ignition. A significant example of risk linked to the traffic is the 1970 fire which is well 472 documented: the fire started from the roadside and spread toward the top of the slope through 473 a talweg. Because Sorba and Vizzavona passes are important traffic points, this makes this area 474 potentially sensitive to social-made fires. 475

476

4.3 A high fire frequency regime

Our results and those from Greece, Anatolia and Spain (Sahan et al., 2022, Christopoulou et 477 al., 2013, Fulé et al., 2008) reveal a regime of high frequency in black pine forests whatever the 478 country, their geographical location and cultural habits. In Vizzavona, considering all fires, the 479 480 mean fire interval of 14 years corresponds to a high fire frequency (sensu Sugihara et al., 2006), with mainly surface fires. This MFI was lower than the ones from other studies: 23 years (Sahan 481 et al., 2022), and 30 years (Christopoulou et al., 2013). However, considering only major fire 482 483 years at Vizzavona, we obtained a MFI of 26 years, which remains in the category of high fire frequency. Plot number VZ06 differs from the others as the MFI is 60 years which correspond 484 to a middle frequency. 485

At Vizzavona, the frequency changed twice around 1931 (increase) and 1970 (decrease), 486 indicating that within a short period of two centuries, there was substantial variability. 487 Interestingly, the data of human demography (Figure S5), show that 1931 corresponded to the 488 start of a second drop of the Vivario population (first was in 1906), ending in 1962. 489 Demographic censuses of 1962 and 1968 constituted the lowest inhabitant numbers since 1800 490 and before the third drop in population around 2010. 491

Despite the different nature of samples and data compared to dendrochronology, palynological 492 study can be complementary to understanding fire history (Clark, 1990). Paleoecological study 493 on the temporal variability of fire in Corsica using sedimentary charcoal and historical 494 495 documents (Camarero et al., 2018) was at millennial scale with a 20-year resolution (Leys et

al., 2014). Results showed two maxima of fire frequency during the last 4500 years, attaining 496 0.012 fires yr⁻¹ (1.2 fires 100 yrs⁻¹; MFI of about 80 yrs). In total, the reconstructed MFI over 497 the Holocene (12,000 years) varied between 30 and 490 years (Leys et al., 2014). These larger 498 values might correspond to major fires knowing that charcoal peaks corresponded mainly to 499 large and severe fires and did not probably capture low severity and small fires, which were 500 well recorded in the present study. Further, the Leys et al study (2014) was based on a 501 climatologically wetter area in western Corsica, which could explain the longer FI. The current 502 Vizzavona study site is on the eastern side of the Corsican mountain range, and thus drier 503 (leeward slopes, rain-shadow effect). 504

505 Flammability increased over time in both analysis periods before and after 1931 as shown by 506 the c parameter Weibull values of 1.89 (1820-1931) and 1.48 (1931-1970). This result means 507 that burning hazard is dependent on processes that happened between fires and not only drivers during the fire year. The reduced fire regime during the 20th century and the long gap since the 508 509 previous fire could have contributed to the severity of the 2000 fire through allowing fuel to accumulate. Such a pattern happened in Greece (Christopoulou et al., 2013), where the longest 510 interval was reported between 1973 and 2007 (34 years), resulting in fuel accumulation 511 followed in 2007 by a severe crown fire. The policy of suppression of traditional fires in France 512 since the king Louis XIV (AD 1669), including in Corsica since the French Revolution, was 513 514 strengthened first during the 1920's (Chalvet, 2016) and later during the 1980's (Alexandrian, 2008). This latest fire suppression policy has probably contributed to the large fire of 2000 that 515 was expected by increasing the time since last fire combined with a severe drought (Piñol et al., 516 517 2007; Evin et al., 2018).

518 4.4 No change in fire season for 202 years

519 The majority of scars were formed in latewood and adjacent ring positions corresponding to 520 summer and the first weeks of autumn (Häusser et al., 2021), which is the driest period of the year for Mediterranean climate (Ozturk et al., 2015). These results are similar to other studies (Fulé et al., 2008; Touchan et al., 2012; Christoupoulou et al., 2013; Sahan et al., 2022), which revealed that the majority of fire scars occurred mostly within latewood and during dormant season. Further, ring position of scars were similar through the whole studied period, indicating that the fire season was steady through fire years. Notably, this was also the season of the latest major fire in 2000 which happened on 27th of August.

527 5. Fire management and conclusion

To sum-up, our 202 year-long history of wildfire regime in a black pine forest of central Corsica provides rational elements for improving forest management. Fire frequency changed twice since 1820, first around 1931 and then after 1970, and peaked significantly between 1931 and 1970 with short intervals. The fire season was steady, but most larger fires (7 out of 8) occurred before 1945. However, the largest and likely most severe ever recorded has occurred very recently in August 2000.

The forest was subject to frequent fires with ignition sources sometimes independent from 534 535 drought; the relationships with human activities remain to explain, excepted the 1943 fire that 536 is iconic of an anthropogenic event, the Second World War. Regarding the importance of fire frequency before 1970, and the highest number of major fires before 1945, human pressures 537 could partly explain these changes in fire size and frequency. Indeed, forest uses decreased 538 progressively throughout the 20thc. However, Vizzavona is a key socio-economic crossroad in 539 Corsica since centuries until today. Generally speaking, the land use is linked with ignition: the 540 higher the land use, the higher the ignition risk (Bowman et al., 2011; Moritz et al., 2014). 541

The synthetic mapping of fires (Figure 5b) suggests that topography and landforms probably control the fire spread at landscape scale in the black pine forest. A more fragmented landscape decreases the risk of severe fires. To define a management strategy, it is necessary to keep in mind that even though the main ignition source are land uses, drought episodes are currently 546 more frequent and longer fostering more flammability. Rare ignition can result in large fires, 547 like in 2000. Further, the increasing time since last fire (Figure 3a) increases the fire hazard (*c* 548 parameter), and the landscape homogeneity favors the spread of wildfires.

The fire regime of the 19th c. until 1930's that was about 80 years must be recover. It was 549 characterized by low severity fires, probably surface fires, preventing fuel accumulation on the 550 ground and improving seedbed for black pines by consuming litter and humus that contain 551 abundant inhibiting components like polyphenols. By consuming fuel bed, such strategy should 552 decrease the risk of large and severe fires. The hot and dry weather has only contributed to 553 improve the fuel quality for burning. Sustainable forest management must consider low severity 554 surface fire in the management of black pine forest, for instance based on prescribed burning. 555 In such case, medium-sized pines (i.e. neither mature overstory trees nor regeneration) should 556 be thinned to break up the fuel ladder and, a fuel break should be established around the area to 557 be burned. Larger trees should mostly be unaffected by surface fires due to their insulating 558 characteristics, such as trunk diameter and bark thickness, as well as crown height, which 559 protects the black pine from damage (Fréjaville et al., 2018b; Ferrat et al., 2021). 560

561

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569

570 Authors' contributions

Justin Badeau: methodology, validation, formal analysis, investigation, writing - original draft, visualization. Frédéric Guibal: methodology, validation, investigation, resources, writing – review & editing. Peter Z. Fulé: conceptualization, methodology, validation, investigation, resources, writing – review & editing, supervision, funding acquisition. Sandrine Chauchard: validation, investigation, writing – review & editing, visualization. Pasquale Moneglia: writing – review & editing, project administration, funding acquisition. Christopher Carcaillet: conceptualization, methodology, investigation, writing – review & editing, supervision, project administration, funding acquisition.

578 Declaration of Competing Interest

579 The authors declare that none of them has an interest financial or personal which may be 580 perceived as competing interest.

581

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