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- 1 Soil properties and climate affect arbuscular mycorrhizal fungi and soil
- 2 microbial communities in Mediterranean rainfed cereal cropping systems
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20 ABSTRACT

21 Soil microorganisms are generally considered sensitive indicators of soil health and fertility. Understanding the role of environmental factors in driving soil microorganisms' 22 23 distributions is vital for sustainable crop production, particularly in vulnerable agricultural systems. This study aims to investigate the effect of environmental factors on the variation of 24 soil microbial biomass and arbuscular mycorrhizal (AM) fungi in Mediterranean rainfed 25 cereal systems. To do so, plant roots and soil were sampled from three cereal crops: Avena 26 27 sativa L., Hordeum vulgare L. and Triticum turgidum L. growing in 24 rainfed sites from three different bioclimatic zones (sub-humid, superior semi-arid and middle semi-arid) in 28

northern Tunisia. Soil microbial phospholipid fatty acids (PLFA) fingerprints, percentage of 29 30 root length colonized by AM fungi as well as AM fungal spore density were then assessed. The redundancy analysis (RDA) revealed that soil organic matter (SOM), temperature, 31 32 altitude and soil sand content accounted for 22.7% of the variation of the microbial PLFA profiles used to estimate both soil microbial biomass and their distribution across different 33 microbial groups. These environmental factors were the most important drivers affecting 34 35 bacterial PLFAs. However, no significant relationships were observed between fungal PLFAs and environmental factors. Mycorrhizal colonization differed significantly among the 36 different host species and regions. Our RDA results indicated that environmental factors, in 37 38 particular mean annual precipitation (MAP), pH, available phosphorus (P) and carbon/ nitrogen ratio (C/N) explained 44.8% of AM fungal colonization and spore number variation. 39 Mycorrhizal colonization was positively correlated with pH but negatively with P, C/N and 40 41 MAP. Concerning AM fungal spore number, it was positively linked to P and MAP. In contrast, it was inversely related to soil pH. Highlighting the main edaphic and climatic 42 factors disturbing soil microorganisms, especially AM fungi, helps to better understand how 43 environmental change will influence soil microbial communities in Mediterranean rainfed 44 cereal systems. 45

46 Keywords:

Soil properties, climatic variables, cereal crops, arbuscular mycorrhizal fungi, soil microbialbiomass.

49 1. Introduction

Soil microorganisms represent a major part of the earth's biota (Whitman et al., 1998) and their number is estimated to be around $\sim 10^{29}$ in terrestrial ecosystems (Cavicchioli et al., 2019). Soil microbial communities have key roles in ecosystem functioning as soil respiration

and denitrification (Bodelier, 2011; Philippot et al., 2013), C and N cycling processes and 53 54 other biogeochemical reactions (Delgado-Baquerizo et al., 2016; Falkowski et al., 2008; Schimel and Schaeffer, 2012), decomposition and stability of organic matter (Cotrufo et al., 55 2013; Nemergut et al., 2014), and plant nutrient acquisition (van der Heijden et al., 2008). 56 Hence, soil microbial communities are contributing to mitigate the impacts of the greenhouse 57 gases and climate change (Singh et al., 2010). Amongst the microbial communities, 58 59 arbuscular mycorrhizal (AM) fungi establish a beneficial association with the roots of more than 85% of vascular plant species (Brundrett and Tedersoo, 2018) and are also known to 60 improve mineral nutrition and growth of the host plant (Smith and Read, 2008). 61

62 In microbial ecology, understanding the factors affecting the composition and the abundance of bacterial and fungal communities represents an important issue. Many 63 environmental factors such as soil properties and climate characteristics were identified as 64 65 abiotic predictors of soil microbial communities (Graham et al., 2014). According to Brockett et al. (2012) and Zheng et al. (2019), environmental factors can be divided into two 66 categories: (i) direct factors represented by edaphic variables which have a dominant 67 influence on microbial community structure and function, and (ii) indirect factors that are in 68 relation with the geographic site characteristics such as climatic conditions. It was proved that 69 70 soil microbial communities could be influenced by many edaphic variables such as pH (Rousk et al., 2010; Wakelin et al., 2008; Zhalnina et al., 2015), soil organic carbon (Fierer et al., 71 2009; Franklin, and Mills, 2009), nutrient elements including nitrogen and phosphorus (Dong 72 et al., 2015; Li et al., 2018; Ling et al., 2017; Wang et al., 2019), soil moisture (Brockett et al., 73 74 2012) and texture (Hansel et al., 2008). In addition, several climatic conditions may determine and/or alter the composition of fungal and bacterial communities such as temperature 75 (Bárcenas-Moreno et al., 2009; Braker et al., 2010), precipitation (Lau and Lennon, 2012; 76 Martiny et al., 2017) and altitude (Frindte et al., 2019; Lanzén et al., 2016). 77

In addition, climate change represents a major threat for biodiversity (Bálint et al., 78 79 2011; Bellard et al., 2012), and thus for the functioning of the natural ecosystem (Cardinale et al., 2012). Based on global scenarios of climate change, the Mediterranean area is expected to 80 be largely exposed to biodiversity (including microbial biodiversity) change and loss (Sala et 81 al., 2000). These changes also affect the microbial communities in agroecosystems, already 82 altered by anthropogenic activities (Cheng et al., 2011). Some of the most represented 83 84 agricultural systems across the Mediterranean area are grain systems (Gibbon, 1981). They occupied about 46 million hectares in 2018 (FAOSTAT, 2018). Previous studies have 85 demonstrated that agricultural practices affect soil microorganism composition (Carbonetto et 86 87 al., 2014; García-Orenes et al., 2013; Wallis et al., 2010). Crop species also have an important effect on the variation of soil microbial biomass. Host plant identity can have a highly 88 variable effect depending on the species, genotype and phenological stage (Maul and 89 90 Drinkwater, 2010; Qiao et al., 2017) and also depending on soil type and climate (Classen et al., 2015). According to Innes et al. (2004), soil microbial biomass was constrained by host 91 92 plant species and varied significantly between three grasses (Lolium perenne L., 93 Anthoxanthum odoratum L. and Holcus lanatus L.), and the three dicots (Achillea millefolium L., Trifolium pratense L. and Plantago lanceolata L.). In addition, by studying the dynamics 94 of the microbial community in the rhizosphere of maize (Zea mays L.), oat (Avena sativa L.), 95 barley (Hordeum vulgare L.) and commercial grass mix, Garbeva et al. (2008) also revealed 96 that plant species had a strong effect on microbial community composition. 97

98 The interaction between environmental factors, host plants and the soil microbial 99 community have been investigated by several authors (Singh et al., 2009; Thakur et al., 2015). 100 However, there is still a lack of conclusive evidence of the main biotic or abiotic factors that 101 drive soil microbial communities. Previous studies have reported that soil microbial 102 communities, estimated by their specific phospholipid fatty acid (PLFA) profiles, was related to environmental factors (e.g., altitude, temperature, rainfall and soil properties) in grasslands
on the Mongolian plateau in China (Chen et al., 2014) and in agricultural fields (wheat) and
natural ecosystems (forest and grasslands) in Australia (Xue et al., 2018). Other authors have
found that the microbial community biomass was more associated with the cropping system
(corn, switchgrass and prairies soils) compared to soil physico-chemical properties in
Wisconsin USA (Liang et al., 2012).

109 Understanding microbial interactions with soil and climate characteristics may help to mitigate the negative effect of climate change through the development of sustainable crop 110 production in agroecosystems. Tunisia is considered among the most vulnerable countries to 111 112 climate change (IPCC, 2013). Most cereal cultivation lands are conducted under rainfed conditions in the northern area. Soils in this region are low in soil organic matter due to the 113 harsh environmental conditions. It's necessary to determine the most important environmental 114 115 factors affecting soil biological fertility in cereals cropping systems which represent the most important strategic crops in the country. Thus, the current work aims at (i) investigating the 116 117 influence of three important cereal crop species: Avena sativa L., Hordeum vulgare L. and 118 Triticum turgidum L., sampled from 24 sites located in different regions in sub-humid and semi-arid areas of northern Tunisia, on soil microbial biomass (using PLFA fingerprints), and 119 120 AM fungi root colonization and spore number, and (ii) studying the relationships between these soil microbial variables and environmental factors, in particular soil physico-chemical 121 properties and climatic characteristics. 122

- 123 **2.** Material and methods
- 124 2.1. Field sites description and samples collection

The study was conducted in rainfed cereal fields located in northern Tunisia (Fig. 1).
Twenty-four fields were selected randomly from 8 different regions (from 3 bioclimatic

zones: sub-humid, superior semi-arid and middle semi-arid), consisting of 7, 13 and 4 sites 127 128 where oats (Avena sativa L.), barley (Hordeum vulgare L.) and durum wheat (Triticum turgidum L.) were respectively, cultivated. For each site, latitude, longitude and altitude were 129 recorded. The sampling campaign was carried out during two cereal growth stages according 130 to Zadoks scale: tillering (Z21) in December 2015-January 2016 and heading (Z58) in April-131 May 2016. During the 2015-2016 agricultural campaign, the maximum temperature of the 132 warmest month (M), the minimum temperature of the coldest month (m), and the mean annual 133 precipitation (MAP) were obtained from MRRA-2 application V5.12.4 (GMAO, 2015). 134

From each site and for each growth stage, three samples were taken following a 135 136 diagonal sampling pattern, each neighbouring sample separated by 50 m. Each sample was composed of four plants, so twelve plants per site were collected. As for the soil, samples 137 were taken from the same place of sampled plants with an auger at a depth of 20 cm. The 138 139 three soil samples taken from each site were then sieved to 2 mm and air-dried for physicochemical analyses. Roots were collected from each of the three samples, washed with water, 140 141 kept in ethanol (50%) and conserved at 4 °C for the evaluation of arbuscular mycorrhizal 142 fungi colonization. Only at the heading growth stage, the three soil samples taken from each site were divided into two subsamples. The first one was stored at -20°C for PLFA analysis 143 and the second subsample was kept in a refrigerator at 4°C for the counting of the indigenous 144 AM fungal spores. 145

146

2.2. Physico-chemical soil analyses

Twelve physico-chemical soil properties (supplementary Table S1, Table S2) were assessed in this study. Soil moisture content (SMO) was determined by oven-drying (at 105 °C) the soil sample until the weight remained constant. The moisture content (%) was calculated using the difference between soil fresh weight and dried one. Soil texture was determined using the Robinson's pipette method (Naanaa and Susini, 1988). Soil pH on soil:

water (1: 2.5, v:v) suspension and soluble salt were determined by measuring the electrical 152 conductivity (EC) (dS/m) on soil: water (1: 5, v:v) suspension at 25 °C with a pH meter 153 (CrisonS.A, Barcelona, Spain) (Pauwels et al., 1992). Total calcium carbonates CaCO₃ 154 (TotalCaCO₃) were measured using the Bernard calcimeter method (Petard, 1993) and active 155 calcium carbonates CaCO₃ (ActiveCaCO₃) were determined according to the method 156 described by Drouineau (1942). Soil organic carbon was determined with the Walkley and 157 Black method using potassium dichromate oxidation (Pauwels et al., 1992) and soil organic 158 159 matter in the soil (SOM) was calculated by multiplying the organic carbon amount by 1.72. The available (Olsen) phosphorus (ppm) (P) was extracted with NaHCO₃ (0.5 M) and then 160 161 measured by a spectrophotometer (OPTIZEN 3220UV, Daejeon, South Korea) (Olsen et al., 1954). Soil organic nitrogen (mg/g) (ON) was measured according to the Kjeldhal method 162 (Jones, 1987). 163

164 *2.3. Microbial analyses*

165

2.3.1. Soil phospholipid fatty acid (PLFA) analysis

166 Soil microbial composition was determined using a hybrid procedure of PLFA analysis according to Frostegård et al. (1991) and Labidi et al. (2015). For each replicate, 167 lipids were extracted from 3 g of freeze lyophilized soils with methanol: chloroform: citrate 168 169 buffer (2:1:0.8, v:v:v) solution. The extracted lipids were fractionated into neutral lipids, glycolipids and polar lipids on a silica acid column by successively eluting with chloroform, 170 acetone and methanol. The PLFA and the NLFA were transformed into free fatty acid methyl 171 esters by transesterification. The final extracts were then analyzed using a Perkin Elmer 172 Autosystem gas chromatograph (GC) equipped with a flame-ionization detector (Norwalk, 173 CT) with hydrogen as carrier gas (40 ml/ min). Fatty acids were characterized by gas 174 chromatography-mass spectrometry (GC-MS) with the use of a SHIMADZU 2010 Plus 175 system gas chromatograph equipped with an automatic injector AOC-20i+s and coupled to a 176

mass spectrometry SHIMADZU QP 2010 Ultra. The separation was performed with Zebron Phenomenex ZB-Wax capillary column (length 10 m, diameter 0.1 mm, 0.1 μ m film thickness). The GC instrument was operated in the split mode (200 as split ratio) and the linear velocity (37 cm/s) using helium as carrier gas.

Fatty acid quantification was made by using nonadecanoic acid methyl ester (C19:0) 181 as an internal standard. Their identification relied on the retention times of fatty acids standard 182 183 (Bacterial Acid Methyl Ester CP Mix in methyl coproate SUPELCO) and its mass analysis. Fatty acids were identified by comparing their mass spectra with the standard mass spectra in 184 the NIST11s library. The amounts of the PLFA C16:1w5 and NLFA C16:1w5 representing 185 186 respectively viable fungal hyphal biomass and storage lipids of the AM fungi were used as indicators of the AM fungal biomass (Frostegård et al., 1991). Since the PLFA C16:1ω5 is not 187 only specific for AM fungi but also for some Gram-negative bacteria, the ratio between 188 189 NLFA and PLFA C16:1w5 amounts was also calculated. This ratio is high in AM fungi (1-200) and low in bacteria (< 1) (Olsson, 1999). The PLFA C18:206.9 was used as indicator for 190 saprotrophic fungi biomass (Frostegård et al., 1993). The fatty acids i15:0, a15:0, i16:0, i17:0, 191 192 a17:0 were used as indicators of Gram-positive bacteria PLFAs and cy17:0, C18:107 and cy19:0 as indicators of Gram-negative bacteria PLFAs in the soil (Frostegård et al., 1993; 193 Tunlid et al., 1989). 194

195 2.3.2. Isolation and quantification of AM fungal spores

Indigenous AM fungal spores were extracted from 50 g of soil using the method of
wet-sieving (sieves of size 250, 125 and 40µm) and decanting described by Gerdemann and
Nicolson (1963), followed by the sucrose centrifugation method of Daniels and Skipper
(1982). The spores were collected from the different sieves and washed with distilled water.
Only intact and healthy spores were counted using a dissecting microscope. The AM fungal
spore number was evaluated by counting the number of spores per gram of soil.

202

2.3.3. AM fungi root colonization

203 In order to quantify arbuscular and vesicular AM fungal colonization, fresh and fine roots were bleached in KOH (2.5%), then acidified in HCl (1%) and stained with trypan blue 204 205 (0.05%) (Phillips and Hayman, 1970). The percentage of mycorrhizal root length was determined by microscopic examination (× 40 magnification) of root fragments of 1 cm using 206 the method of McGonigle et al. (1990). From each sample, 225 observations (75 root 207 208 fragments with 3 intersections per root fragment) were realized in order to count specific 209 mycorrhizal structures (arbuscules and vesicles). In total, 675 observations per site were analyzed. 210

211 *2.4. Data analysis*

All data analyses were conducted using R v.3.6.0 (R Core Team, 2019). The Shapiro 212 and Bartlett tests were used to assess for the normality and homogeneity assumptions of the 213 214 data. To satisfy the assumptions, data of available phosphorus (P) and soil organic nitrogen (ON) were converted to natural logarithm and square root values, respectively. A one-way 215 216 ANOVA with Tukey's HSD (honestly significant difference) test (*p*-values < 0.05) was used 217 to compare the mean values of the soil microbial variables: soil microbial PLFA profiles (PLFA C16:1ω5 for AM fungi, PLFA C18:2ω6.9 for saprotrophic fungi, Gram-positive and 218 Gram-negative PLFAs); percentage of total root colonization at the tillering and heading 219 220 stages; and AM fungal spores' number, among the different host plant species (A. sativa, H. *vulgare* and *T. turgidum*). In addition, ANOVA was performed by host species as well as by 221 region of origin. Furthermore, a simple linear regression analysis was performed between AM 222 223 fungal spores' number and mycorrhizal root colonization in order to determine their specific relationships. A PCA was also conducted to examine the variation between the 72 samples 224 225 based on the environmental variables (climatic and soil physico-chemical properties). Canonical redundancy analysis (RDA) based on Hellinger-transformed data was used to 226

examine the relationships between the soil microbial and the environmental variables. Two 227 RDA were conducted, the first RDA was used to analyse the effect of the environmental 228 factors on the soil microbial biomass (using PLFAs profiles) and the second RDA, to analyse 229 the effect of the environmental factors on the AM fungal spores' number and root 230 colonization. This analysis was performed using the 'vegan' library in the R software package 231 (Oksanen et al., 2019). The proportion of explained variation was calculated using the 232 adjusted R^2 values (Borcard et al., 2011). The significance of the results was tested with a 233 Monte Carlo test (using 999 permutations). Backward selection of the explanatory variables 234 that were significantly correlated with these soil microbial variables was performed using the 235 236 stepwise regression and the Monte Carlo permutation test.

237 **3. Results**

238 *3.1. Soil microbial biomass*

In order to study the influence of host plant and the pedo-climatic conditions of the 239 cultivation sites, microbial biomasses were assessed using specific lipid biomarkers (PLFAs). 240 241 The amounts of PLFA C16:1ω5, used here to quantify the biomass of AM fungi in soils, were not significantly different between the three studied host plant species (p > 0.05), while the 242 NLFA C16:105 soil contents, which represent mainly storage lipid in spores, were 243 significantly different (p < 0.05) (Table 1, supplementary Table S4). However, whereas the 244 ratio NLFA /PLFA C16:1w5 was inferior to 1 for A. sativa, indicating a predominant bacterial 245 contribution, it was superior to 1 for H. vulgare and T. turgidum, indicating AM fungi 246 predominance (Table 1). 247

There were no significant differences in saprotrophic fungi biomass, estimated by the PLFA C18:2 ω 6.9 (p = 0.99). Gram-positive bacteria biomass (quantified by the sum of the PLFAs i15:0, a15:0, i16:0, i17:0, a17:0) as well as Gram-negative bacteria (estimated by the

sum of the PLFAs cy17:0, C18:1 ω 7 and cy19:0) biomass did not significantly vary between the different host plant species (p > 0.05) (Table 1, supplementary Table S4).

At a regional scale, no significant differences were observed between the studied regions for the amounts of PLFA C16:1 ω 5, PLFA C18:2 ω 6.9, Gram-positive bacteria PLFAs and Gram-negative bacteria PLFAs (p > 0.05) (Table S5, supplementary Table S6).

256 *3.2. Arbuscular mycorrhizal fungi*

There were no significant differences in the AM fungal spore number between the three host plant species (p > 0.05, supplementary Table S4) (Table 2a). However, the AM fungal spore number significantly varied among the different studied regions (p < 0.0001, supplementary Table S6) (Table 2b).

Total AM fungi root colonization rates were significantly different between the 261 different host plant species (*p*-values < 0.01, supplementary Table S4) (Table 2a). Compared 262 263 to A. sativa, total mycorrhizal rates for H. vulgare and T. turgidum were 1.9 and 2.3-fold higher, respectively at tillering and 1.5 and 1.9-fold higher, respectively at heading (Table 2a). 264 Arbuscules were observed at both growth stages and they were significantly different among 265 266 the three studied host plant species (p < 0.0001 at tillering and p < 0.01 at heading, supplementary Table S3a and Table S4). However, vesicle formations were observed only at 267 the heading stage for *H. vulgare* and *T. turgidum* and not for *A. sativa*. Vesicular colonization, 268 was two-times higher for T. turgidum as compared with H. vulgare (supplementary Table 269 270 S3a).

At the regional scale, total mycorrhizal colonization as well as arbuscular and vesicular colonization rates differed significantly (p < 0.001, supplementary Table S6) (Table 273 2b, supplementary Table S3b). Arbuscular and vesicular colonization rates were 3 times 274 higher in the site of Tunis compared with Jendouba (Table S3b).

Overall, negative correlations between mycorrhizal colonization and AM fungal spore number were observed in particular at the heading growth stage ($R^2 = 0.55$, p < 0.001) (Fig. 277 2).

278 *3.3. Variation of the environmental variables*

The PCA analysis based on the 16 environmental variables showed differentiation 279 between the 72 studied samples. The first two dimensions of the PCA accounted for 47.5% of 280 281 the total variation in the environmental data (Fig. 3, supplementary Figure S1). The first dimension which explained 29.7% of the environmental data variation, was strongly and 282 negatively correlated with the minimum temperature of the coldest month (m) ($r^2 = -0.85$, p < -0.85283 284 0.0001), maximum temperature of the warmest month (M) ($r^2 = -0.44$, p < 0.0001), soil sand content ($r^2 = -0.49$, p < 0.0001) and carbon/ nitrogen ratio (C/N) ($r^2 = -0.47$, p < 0.0001). 285 However, soil pH ($r^2 = 0.78$, p < 0.0001), soil silt content ($r^2 = 0.83$, p < 0.0001), total CaCO₃ 286 $(r^2 = 0.74, p < 0.0001)$ and active CaCO₃ $(r^2 = 0.77, p < 0.0001)$ contents showed a high and 287 positive effect on the first dimension. The second dimension accounted for 17.8% of the 288 environmental data variation, to which soil moisture content (SMO) ($r^2 = 0.65$, p < 0.0001) 289 and soil sand content ($r^2 = -0.67$, p < 0.0001) contributed the most (Fig. 3, supplementary 290 Table S7). According to the PCA results, the region of Jendouba, Kef and Siliana were 291 292 separated from Tunis, Bizerte and Nabeul along the first dimension. However, the variability within each region was mostly displayed along the second dimension of the PCA (Fig. 3). 293

294

3.4. Correlations between soil microbial and environmental variables

In order to determine which environmental factors affected soil microbial variables, two redundancy analysis (RDA) were conducted. The first RDA was used to check the effect of the environmental factors on the soil microbial PLFA profiles and the second RDA, to analyse the effect of the environmental factors on AM fungal root colonization and spore density. For soil microbial PLFA, 19.7 and 2.6% of the variation were explained by the first two canonical axes (Fig. 4). The environmental factors (SOM, Sand, M, m and altitude) had a significant effect on the variation of these microbial PLFA profiles (F= 3.88, p < 0.001, supplementary Table S8) and they accounted for (R^2 = 22.3%) of the total variation. In addition, the RDA revealed that Sand (score=-0.5, p < 0.01), M (score=-0.26, p < 0.05) and altitude (score=-0.20, p < 0.01) were highly and positively correlated with BG-, while SOM (score=0.51, p < 0.01) and m (score=0.05, p < 0.01) were correlated negatively with BG+. (Fig.4, supplementary Table S9).

Similarly, there was a significant effect of the environmental variables on the 307 mycorrhizal colonization and AM fungal spore number (F= 10.73, p < 0.01) and they 308 309 explained 44.8% of the total variation (Fig. 5). According to the RDA analysis, only soil pH, soil available phosphorus (P), C/N and mean annual precipitation (MAP) contributed 310 significantly to the variation of mycorrhizal colonization at both tillering and heading stages 311 and spore number of AM fungi (Table S9). The RDA1 axis which explained most of the 312 variation in mycorrhizal colonization and AM fungal spore number (44.35%, p < 0.001, Table 313 S9), was positively correlated with C/N (score= 0.3, p < 0.01), MAP (score=0.7, p < 0.01, 314 Table S8) and P (score= 0.53, p < 0.01) but negatively correlated with pH (score= -0.55, p < 0.01) 315 0.01). Among these environmental factors, MC.Tillering and MC.Heading were positively 316 317 correlated with pH. However, the AM fungal spore number was positively correlated with MAP, P and C/N but negatively with soil pH (Fig. 5, Table S8). 318

319 **4. Discussion**

The present study aims to determine which environmental factors influence soil microbial PLFA profiles, AM fungal root colonization and spore density associated with the rhizospheres of the main cereal crop species (*A. sativa, H. vulgare* and *T. turgidum*) located in 24 sites from 8 regions in sub-humid and semi-arid areas in northern Tunisia. Relationship between cereal species, soil microbial biomass, AM fungal root colonization
and spore density

Using specific lipid biomarkers, we showed that there was no significant difference in 326 327 the microbial biomasses (AM fungi, saprotrophic fungi and bacteria) between the three studied cereal crops (A. sativa, H. vulgare and T. turgidum) indicating that the host plant 328 species had no effect on soil microbial biomass. The influence of host plant species on 329 330 microbial biomass is still controversial. Some studies have suggested that the soil microbial biomass did not differ significantly among several Poaceae species (Briza media, Festuca 331 ovina and Bromus erectus) (Bezemer et al., 2006). However, in other studies, the plant species 332 333 (Arrhenatherum elatius, Holcus lanatus and Dactylis glomerata) had a significant effect on the variation of soil microbial PLFA (Patra et al., 2008). 334

In the present study, AM fungi root colonization rates were higher in T. turgidum and 335 H. vulgare than in A. sativa. This is in accordance with the work of Castillo et al. (2012), who 336 337 found that the AM fungi root colonization rate was lower in oats compared to barley and 338 wheat. The difference between the three cereal species could be linked to their carbohydrate production. In fact, the quantity and quality of carbon resources produced by the plant vary 339 according to the plant species (Ngosong et al., 2012), resulting in a different response by the 340 341 mycorrhizal symbiosis (Jones et al., 2004). Moreover, root colonization percentages differed between tillering and heading stages for the three cereal species. It has been reported in 342 previous studies that AM fungal colonization also changes across different phenological 343 344 stages of the host plant species (Mohammad et al., 1998; Schalamuk et al., 2006). AM fungal root colonization increased with advancing plant growth (Mohammad et al., 1998; Panwar et 345 346 al., 2011). This may be explained by the fact that a low temperature during winter at the tillering stage inhibited the germination of AM fungal spores and thus mycorrhizal root 347 colonization (Grey, 1991; Koske, 1981). Concerning AM fungal spore density, no significant 348

differences were observed between the studied cereal species. This could be explained by the
fact that sporulation depends more on AM fungi species physiology. There are AM fungi that
sporulate more (e.g., *Acaulospora* and *Glomus* species), while others sporulate less
(*Gigaspora* and *Scutellospora* species) (Jefwaa et al., 2012; Wang and Jiang, 2015).
Furthermore, the AM fungal spore number usually varies according to edaphic factors and
cropping systems.

Soil characteristics, temperature and site altitude influence soil microbial biomass in rainfed conditions

According to the current data, the redundancy analysis (RDA) revealed a significant and positive correlation between soil organic matter (SOM) and Gram-positive bacteria PLFAs. This result was consistent with those of Liang et al. (2012) who found that Grampositive bacteria tend to correlate with SOM. Gram-positive and Gram-negative bacteria depend on different carbon sources (Fanin et al., 2019). Compared with Gram-negative bacteria, Gram-positive bacteria use carbon resources derived from SOM that are more complex and recalcitrant (Bird et al., 2011; Kramer and Gleixner, 2006).

It has been reported in previous studies that soil texture had a strong influence on soil 364 microbial biomass (Mellado-Vázquez et al., 2019). In the present study, soil sand content was 365 366 positively correlated with Gram-negative bacteria but negatively with Gram-positive bacteria. These results were in line with those of Hemkemeyer et al. (2018). Gram-positive bacteria 367 were more abundant in silt and clay soils, while the Gram-negative bacteria were found in the 368 larger fractions (> 200 mm) in sand particles (Zhang et al., 2015). These differences could be 369 related to the nutrient preference of Gram-positive and Gram-negative bacteria. In fact, the 370 371 nutrient concentration in the soil differs with soil particle size fractions, which have an important influence on bacterial community structure (Carson et al., 2009; Nishiyama et al., 372 373 2012).

For climatic characteristics, Gram-positive and Gram-negative bacteria showed a 374 375 different response to temperature. According to the current data, the maximum temperature of the warmest month had a positive and significant effect on Gram-negative bacteria but a 376 377 negative effect on Gram-positive bacteria. These results are in contradiction with those of Zhou et al. (2017) who worked on the effect of temperature on soil microbial PLFAs 378 signatures. These inconsistent results could be related to the differences in substrate 379 preferences and metabolic abilities between Gram-positive and Gram-negative bacteria (Ho et 380 al., 2017). Gram-positive and Gram-negative bacteria contain the characteristics of 381 oligotrophic (nutrient-poor) and copiotrophic (nutrient-rich) microorganisms, respectively 382 383 (Fierer et al., 2007; Naylor et al., 2017). Increasing temperature enhanced fast growers (copiotrophic) bacteria that prefer high nutrient availability (Männisto et al., 2016). 384

In the present study and in line with previous studies, the altitude was positively correlated with Gram-negative bacteria but negatively with Gram-positive (Margesin et al., 2009; Siles and Margesin, 2016). Altitude had a significant effect on soil properties such as pH, and nutrient availability (Wu et al., 2015). The low availability of nutrients such as carbon and nitrogen at high altitude could negatively affect the growth of some bacterial groups but select other adapted ones (Kumar et al., 2019).

Soil properties and precipitation influence AM fungal root colonization and spore density inrainfed conditions

Based on the current data, chemical soil characteristics and climate had a significantinfluence on both AM fungal root colonization and spore density.

Concerning the soil available phosphorus content, our results showed that this parameter was negatively correlated with mycorrhizal colonization but positively with AM fungal spore number. The inverse relationship between AM fungi root colonization and soil phosphorus levels has been already described in several studies (Kluber et al., 2012; Liu et al.,

2016; Treseder, 2004). Increasing soil available phosphorus can cause depressed growth of 399 400 the extra-radical hyphae of the AM fungi, which reduce mycorrhizal root colonization (Kahiluoto et al., 2001). From our data, the AM fungal spore number and available 401 402 phosphorus were positively correlated. This result is consistent with previous studies (Welemariam et al., 2018; Wu et al., 2011) but contradicts the findings of Soka and Ritchie 403 (2018) who observed a significant decrease in AM fungal spore density with increasing soil 404 phosphorus. According to García et al. (2017), the increase in spore density with increasing 405 soil available phosphorus could be explained by a change in pattern of resource allocation 406 between the hyphal networks as functional structures that provide nutrients to host plants and 407 408 the spores as survival resting structures (Sheng et al., 2013).

Furthermore, the soil pH ranging between 6.6 and 8.6 had a positive influence on mycorrhizal root colonization but a negative one on AM fungal spore density. It was reported that AM fungi had a preference for a near neutral or alkaline soil pH (Marschner et al., 2005). The negative effect on spore number as observed by Panwar et al. (2011), could be linked to the presence of different AM fungi taxa in the rhizosphere that cannot tolerate neutral or strongly alkaline soil pH.

Among the different soil properties, a negative correlation was found between soil C/N ratio and mycorrhizal colonization could be due to the availability of nitrogen in the soil. It has been reported by Atul-Nayyar et al. (2009) that C/N ratio decreased in the presence of AM fungi (from 17.8 to less than 13), indicating that AM fungi enhance organic matter decomposition.

For climate parameters, the negative correlation between precipitation and AM fungi root colonization contrasted with previous finding showing a positive correlation between mycorrhizal variables and precipitation (Lingfei et al., 2005; Oliveira and Oliveira, 2010; Silvana et al., 2018) but was consistent with other studies (He et al., 2016; Muthukumar and

Udaiyan, 2002). The AM fungi are aerobic (Sylvia and Williams, 1992), soil water and 424 425 aeration have an important impact on their development and effectiveness. Increase in soil moisture is associated with a decrease in oxygen level in soil, which could negatively affect 426 427 the spore germination and the hyphal growth of AM fungi (Hattori et al., 2014). For the AM fungal spore number, our data showed a positive correlation between this parameter and 428 precipitation. The high sporulation is a stress response of AM fungi to adverse weather 429 conditions (Escudero and Mendoza, 2005). It has been suggested that when soil conditions are 430 suitable for spore germination, AM fungal root colonization increases and spore density 431 decreases and vice versa (Oliveira and Oliveira, 2005). The relationships between 432 433 precipitation and soil moisture are complex and irregular (Li et al., 2016) and depends on others environmental conditions such as soil texture, infiltration and evapotranspiration (Yang 434 et al., 2018). 435

436 **5.** Conclusion

This research represents the first investigation on soil microbial communities and AM 437 438 fungi symbiosis of cereal crops in sub-humid and semi-arid regions in northern Tunisia. The results obtained herein revealed that the microbial distribution in soils did not differ according 439 to the three studied crop species (A. sativa, H. vulgare and T. turgidum). However, there is a 440 complex interaction between the distribution of the soil microorganisms and the 441 environmental factors (climatic conditions and soil properties). In addition, Gram-positive and 442 Gram-negative bacterial biomass were mainly linked to SOM, soil texture, temperature and 443 altitude. This study shows that mycorrhizal root colonization varied significantly according to 444 the host species. It also highlighted that soil properties, particularly pH, soil available 445 446 phosphorus and precipitation, were all correlated with mycorrhizal root colonization status. All these results demonstrate the effect of environmental factors on soil microorganisms' 447 448 distribution in rainfed Mediterranean agricultural systems. Further research should be

- conducted to select suitable native microorganisms playing a crucial role in enhancing cerealcrop productivity and reducing the impact of conventional cultural practices in rainfed
- 451 agricultural systems.

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870 **Figures**

Figure 1. Map showing the location of the 24 studied sites in northern Tunisia. *Avena sativa*, *Hordeum vulgare* and *Triticum turgidum* sites are represented in blue, yellow and red,
respectively.

Figure 2. Simple regression analysis between AM fungal spore number and mycorrhizal root

colonization at tillering and heading stages for all the sampled plants (from all studied sites).

876 Figure 3. Principal component analysis of the 72 samples collected from the 24 studied sites, based on the environmental variables. Samples were color-coded according to the region of 877 origin. Ellipses are drawn at 95% confidence intervals for each region of origin and they 878 describe one standard deviation around the centroid position of each region. Mean annual 879 precipitation (MAP); maximum temperature of the warmest month (M); minimum 880 881 temperature of the coldest month (m); altitude; soil pH; soil electrical conductivity (EC); total calcium carbonates content (TotalCaCO₃); active calcium carbonates content (ActiveCaCO₃); 882 soil organic matter (SOM); soil organic nitrogen (ON); carbon/ nitrogen ratio (C/N); available 883 phosphorus (P); soil moisture (SMO), soil clay content (Clay); soil silt content (Silt) and soil 884 sand content (Sand). 885

Figure 4. Redundancy Analysis (RDA) triplots showing the significant relationship between the environmental factors and the soil microbial composition based on the PLFA profiles using the Hellinger distance. Scaling type = 2. RDA1 (19.66%) and RDA2 (2.63%). The PLFA profiles are shown by filled orange triangles " \blacktriangle ": PLFA C16:1 ω 5 for AM fungi (AMF); ratio between NLFA and PLFA C16:1 ω 5 (NL/PL C16:1 ω 5); PLFA C18:2 ω 6.9 for saprotrophic fungi (SF); Gram-positive bacteria PLFAs (BG⁺); Gram-negative bacteria PLFAs (BG⁻) and the significant environmental variables by blue arrows: maximum temperature of the warmest month (M); minimum temperature of the coldest month (m); soil organic matter (SOM); soil sand content (Sand) and altitude. "+" represent the different studied samples.

896 Figure 5. Redundancy Analysis (RDA) triplots showing the significant relationship between the environmental factors and the root colonization and the spore number of AM fungi using 897 the Hellinger distance. Scaling type = 2. RDA1 (44.35%) and RDA2 (0.46%). The 898 mycorrhizal colonization and spore number are shown by filled orange triangles "^{\[]}." 899 mycorrhizal colonization at tillering (MC.Tillering); mycorrhizal colonization at heading 900 901 (MC. Heading); AM fungal spore number (Spore) and the significant environmental variables by blue arrows: mean annual precipitation (MAP); pH; carbon/ nitrogen ratio (C/N); available 902 903 phosphorus (P) and altitude. "+" represent the different studied samples.

904 **Tables**

Table 1. Fungal and bacterial biomass estimated by PLFA ($\mu g/g dry soil$) analysis of 72 soil samples collected from the different host plant species Data were reported as means and standard deviations.

Host plant	PLFA C16:1w5	NLFA C16:1w5	NLFA /PLFA C16:1ω5	PLFA C18:2ω6.9	Gram-positive bacteria PLFAs	Gram-negative bacteria PLFAs	Total microbial PLFA
<i>A. sativa</i> (<i>n</i> = 21)	0.97 ± 0.12	0.94 ± 0.524 ab	0.90 ± 0.49	0.90 ± 0.15	2.42 ± 0.80	1.27 ± 0.22	5.56 ± 0.99
<i>H. vulgare (n=39)</i>	0.93 ± 0.10	1.49 ± 1.04 a	1.37 ± 0.90	0.90 ± 0.24	2.21 ± 0.81	1.22 ± 0.22	5.27 ± 1.02
<i>T. turgidum</i> (<i>n</i> =12)	1.01 ± 0.11	1.01 ± 0.54 a	1.01 ± 0.55	0.89 ± 0.18	2.43 ± 1.10	1.32 ± 0.28	5.66 ± 1.27
ANOVA							
<i>P</i> -value	0.103	0.0407	0.0615	0.997	0.569	0.363	0.412
Significance	Ns	*	Ns	ns	ns	ns	ns

907 Different letters in the same row indicate significant differences at 0.05 by Tukey's HSD test. Significance: * p < 0.05, ** p < 0.01, *** p < 0.001, **** p < 0.0001.

Table 2. Arbuscular Mycorrhizal (AM) fungi colonization across the 72 samples collected from the different studied sites. Means and standard
 deviations were reported by (a) host plant species and by (b) sites region.

911	(a)
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Host plant	Total mycorrhizal colonization (%) at tillering	Total mycorrhizal colonization (%) at heading	AM fungal spore number (per 50 g of soil)
<i>A. sativa</i> (<i>n</i> = 21)	5.14 ± 3.18 b	21.11 ± 14.41 b	124.71 ± 56.50
H. vulgare (n=39)	9.68 ± 4.03 a	31.11 ± 14.82 a	104.23 ± 55.23
T. turgidum (n=12)	12.04 ± 3.33 a	39.77 ± 11.52 a	86.00 ± 36.13
ANOVA			
<i>P</i> -value	1.96e ⁻⁶	0.00166	0.123
Significance	****	**	ns

912 (b)

Region	Total mycorrhizal	Total mycorrhizal	AM fungal spore number
	colonization (%) at tillering	colonization (%) at heading	(per 50 g of soil)
Bizerte $(n=9)$	8.0 ± 2.5 bc	26.4 ± 11.6 abc	109.0 ± 43.6 bc
Tunis (<i>n</i> = 9)	13.5 ± 5.8 a	44.2 ± 19.0 a	77.3 ± 44.8 c
Nabeul (n=12)	8.4 ± 3.8 abc	31.7 ± 13.9 ab	70.7 ± 45.9 c
Zaghouan ($n=6$)	9.3 ± 2.7 abc	30.0 ± 10.3 abc	173.5 ± 57.5 ab
Beja (<i>n</i> = 9)	10.4 ± 1.3 ab	27.7±13.7 abc	110.0 ± 29.3 bc
Jendouba (n= 9)	3.5 ± 2.6 c	11.7 ± 8.9 c	175.7 ± 22.3 a
Kef (<i>n</i> = 12)	7.8 ± 2.9 bc	25.9 ± 9.4 bc	94.8 ± 37.7 c
Siliana (<i>n</i> = 9)	9.6 ± 5.5 ab	40.9 ± 13.1 ab	72.7 ± 41.6 c
ANOVA			
<i>P</i> -value	8.00 e ⁻⁵	6.22 e ⁻⁵	$1.49 e^{-7}$
Significance	****	****	****

913 Different letters in the same row indicate significant differences at 0.05 by Tukey's HSD test. Significance: * p < 0.05, ** p < 0.01, *** p <

914 0.001, **** *p* < 0.0001.



Figure 1.



Figure 2.



Figure 3.



Figure 4.



Triplot RDA - scaling 2

Figure 5.