

Sunlight Induces the Production of Atmospheric Volatile Organic Compounds (VOCs) from Thermokarst Ponds

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¹ Sunlight Induces the Production of

- ² Atmospheric Volatile Organic Compounds
- 3 (VOCs) from Thermokarst Ponds
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- 15 KEYWORDS: permafrost, thermokarst, volatile organic compounds (VOCs), Arctic,
- 16 photochemistry
- 17

SYNOPSIS: Photoproduction of VOCs from thermokarst ponds water is not directly
correlated to the amount of dissolved organic matter but rather to the availability of
organic matters at the top layers of the liquid phase.

21 **ABSTRACT:** Ground subsidence caused by permafrost thaw causes the formation of 22 thermokarst ponds, where organic compounds from eroding permafrost accumulate. 23 We photolyzed water samples from two such ponds in Northern Quebec and discovered 24 the emission of volatile organic compounds (VOCs) using mass spectrometry. One 25 pond near peat-covered permafrost mounds was organic-rich, while the other near 26 sandy mounds was organic-poor. Compounds up to C10 were detected, comprising O, N, and S atoms. The main compounds were methanol, acetaldehyde, and acetone. 27 28 Hourly VOC fluxes under actinic fluxes similar to local solar fluxes might reach up to 29 1.7 nmol C m⁻² s⁻¹. Unexpectedly, the fluxes of VOCs from the organic-poor pond were 30 greater than those from the organic-rich pond. We suggest that different segregation of 31 organics at the air/water interface may partly explain this observation. This study 32 indicates that sunlit thermokarst ponds are a significant source of atmospheric VOCs, 33 which may affect the environment and climate, via ozone and aerosol formation. 34 Further work is required for understanding the relationship between pond organic 35 composition and VOC emission fluxes.

36 INTRODUCTION

The top 3 m of terrestrial permafrost contains ~ 1000 Pg of carbon in the form of 37 frozen vegetal debris that has accumulated over the past millennia^{1,2}. This is 38 significantly more than the atmospheric reservoir, 750 Pg³, so that the mineralization 39 40 of permafrost carbon induced by global warming has the potential to dramatically 41 accelerate climate change. Projections of CO2 and CH4 emissions by thawing permafrost show a very wide range of values, most of which are between 7 and 100 Pg 42 C by the year 2100 under the emission scenario RCP4.5⁴⁻⁷. However, these estimates 43 44 are based on gradual permafrost thaw and ignore numerous processes, including rapid erosion and abrupt permafrost thaw⁸, increased lightning and fires, and the rapid 45 northward migration of trees⁹. Much greater emissions are therefore likely^{8,10} and 46 47 additional process studies are required for more reliable projections.

Many model parameterizations of the permafrost contribution to the atmospheric CO₂ 48 budget rely on field or laboratory warming experiments that observe static 49 processes^{11,12}. Recent experiments however stress that lateral hydrological export could 50 explain unexpectedly high losses of permafrost carbon¹³. Indeed, ice melting in 51 52 permafrost leads to ground subsidence and geomorphological changes with the formation of erosion gullies and thermokarst ponds¹⁴. Erosion processes lead to the 53 54 accumulation of dissolved and particulate organic carbon (DOC and POC) in thermokarst ponds^{15,16}, where it can be mineralized by microbial respiration and 55

photochemistry¹⁷⁻¹⁹. These latter processes are seldom included in models of permafrost
carbon and in particular photolytic processes are ignored.

Photolysis of DOC and POC leads to the formation of CO2 and its release to the 58 59 atmosphere. However, products of incomplete oxidation are also formed and have been 60 observed in the liquid phase²⁰⁻²². The more volatile of these products can be released 61 into the atmosphere and affect climate through several processes. The first one is their subsequent oxidation to CO2. Since gaseous emission measurements from lakes focus 62 on CO₂ and CH4^{19,23-25}, neglecting this delayed CO₂ source may underestimate the 63 64 greenhouse gas (GHG) emissions by lakes. Both field and laboratory studies indicate 65 that incomplete photo-oxidation may represent up to half the photolysis products in lakes^{22,26}. The second one is through the formation of ozone (O₃), a potent GHG²⁷, as 66 67 many of the reactive organic compounds emitted are likely to have a strong O₃ formation potential²⁸. Lastly, the further atmospheric oxidation of emitted compounds 68 69 may lead to the formation of less volatile compounds that contribute to aerosol 70 formation²⁹. These aerosols affect climate through light scattering and cloud droplet formation^{30,31}. 71

Here we performed laboratory experiments to explore the release of volatile organic compounds (VOCs) from the photolysis of waters from thermokarst ponds in palsa and lithalsa environments. Palsas are peat-covered permafrost mounds whose collapse feeds large amounts of organic matter into thermokarst ponds. Lithalsas are also permafrost mounds, but without peat, and the amount and nature of organic matter released to thermokarst ponds are therefore expected to be different. Investigating the contrasted environments informs us about the variability of VOC emissions from thermokarst ponds. Below we will demonstrate that both sampled ponds are photochemically active. We will focus our attention on the different behaviors, and we discuss some aspects of the mechanisms involved and their underlying atmospheric implications at the sampling period and sites. The aim of this study is to encourage follow-up investigations to cover a wider range of conditions and spatiotemporal coverage.

84

MATERIALS AND METHODS

Sample Collection. The water samples used in this study were sampled in September
2019 from two contrasted thermokarst ponds located within the Canadian subarctic
region at Kuujjuarapik (55°13'11'' N, 77°42'52'' W) and Umiujaq (56°33'31'' N,
76°28'57'' W), as shown in Figure 1.

89 The Kuujjuarapik pond was located at the base of a collapsing palsa, and its water 90 was visibly colored. The Umiujaq pond was surrounded by collapsing lithalsas with 91 sandy soil, with water less colored than that from the Kuujjuarapik pond. The vegetation 92 near Umiujag was mostly lichen (Cladonia sp., mostly C. stellaris) and dwarf birch (Betula glandulosa) as detailed by Gagnon et al.³². The soil organic content is low in 93 this area, around 4 kg m⁻²³². In the sporadic permafrost zone, i.e., where we collected 94 the water samples, numerous palsas and lithalsas have formed³³, as carbon reservoirs³⁴. 95 To collect the liquid samples, amber glass containers, initially rinsed with 96 97 concentrated HNO₃ and then several times with ultrapure water, were then prefilled

98	with ultrapure water (18.2 M Ω .cm) and emptied at each site, and then filled with pond
99	water. The triplicate samples were taken from the top 10 cm of the ponds. Blanks were
100	also collected by opening and closing a bottle filled with ultrapure water at each site.
101	All samples were frozen after collection and kept frozen until the actual photochemical
102	experiments. The water samples were used for experiments as received i.e., unfiltered.
103	All results given below are blank-corrected, where the signals arising from blanks
104	were subtracted from those of actual experiments. The VOC production from blanks
105	was always significantly lower than those from collected samples (see Figure S1).
106	Experimental Setup. The approach used here was similar to that in our previous
107	studies ^{35,36} . Briefly, a water- and Pyrex-filtered Xenon arc lamp delivered light
108	irradiation in the actinic region (i.e., at wavelengths longer than ~ 280 nm) to a
109	cylindrical quartz cell (2 cm diameter and 10 cm length) half-filled with 6 mL of the
110	unfrozen pond water sample (Figure S2). The headspace of this reactor was then
111	continuously purged with purified air at a flow rate of 200 mL min ⁻¹ , and its content of
112	VOCs can be monitored by mass spectrometry (see below). After an initial stabilization
113	period, the lamp was switched on, simulating solar irradiation for 180 min. The
114	temperature fluctuations of the reactor were small, i.e., within $\pm 2^{\circ}$ C (Figure S3). The
115	experimental temperature (~ 25°C) is comparable to the maximum air temperature
116	during the warm seasons in northern Quebec ³⁷ . All experiments were performed in
117	triplicates.

118 Figure S4 gives the spectral profiles of experimental and natural irradiation. Here, 119 irradiance is calculated for the wavelength range of 280 - 420 nm because only radiation 120 therein initiates photochemistry, since shorter wavelength may cause photolysis, while 121 the energy of longer wavelength is not sufficient to induce photochemical reactions. 122 Laboratory works commonly used this wavelength range to infer the photochemical production and fluxes of VOCs³⁸⁻⁴⁰. The experimental irradiance was 8.0 mW cm⁻² as 123 previously set^{41,42}, with an actinic flux of 1.20×10^{16} photon cm⁻² s⁻¹. The surface of 124 125 irradiated water sample in this laboratory study was several orders of magnitude lower 126 than in nature, which reduces the flux of products. In order to meet our detection limits, 127 the experimental irradiance was set to be higher than the natural levels at the sampling sites, but it still compares to the midday conditions elsewhere and should thus be 128 129 regarded as realistic on Earth^{43,44}.

Measurement of VOCs. The VOCs produced upon irradiation were monitored using a high-resolution proton transfer reaction time-of-flight mass spectrometer (HR-PTR-ToF-MS, 8000, Ionicon Analytic GmbH, Austria). This technique was previously described by Müller et al.⁴⁵ and uses H_3O^+ as a soft ionization agent. This MS continuously sampled 200 mL min⁻¹ gas flow at the reactor's outlet through a heated silicon steel tube. Spectra were acquired with a time resolution of 10 s at a mass resolution of 4500 m/ Δ m.

137 A calibration standard (TO-14A Aromatic Mix, Restek Corporation, USA)
138 containing 14 VOCs (100 ppb for each diluted in N₂) was utilized to calibrate and test

139 the instrument performance. The sensitivity of standard compounds ranged between 15 and 70 cps/ppb, depending on the actual mass. As it was not possible to calibrate the 140 141 instruments for each compound detected, the protonation rate constant was assumed to be 2×10^{-9} cm³ s⁻¹, and therefore the mean sensitivity was set to be 30 cps/ppb. The 142 143 mass calibration was performed by internal standards: hydronium ion isotope ($H_3^{18}O^+$, m/z = 21.022) and protonated acetone (C₃H₇O⁺, m/z = 59.049) as commonly selected. 144 The PTR-MS viewer software (version 3.2.8) was used for the raw spectral analysis, 145 mass calibration, peak fitting, and area extraction. The mixing ratios of VOCs were 146 finally determined based on the existing protocols⁴⁶. More details are provided in 147 Supporting Information (Text S1). 148

149 Aqueous Phase Analysis. For the liquid analysis, the solution before and after the 150 VOC photochemical measurements were filtered, in contrast to the actual experiments, 151 by a 0.22 µm polytetrafluoroethylene (PTFE) membrane filter (Pall Corporation, USA). 152 Each treated sample was analyzed using a Q-Exactive Hybrid Quadrupole-Orbitrap 153 mass spectrometer (Thermo Scientific, USA) to characterize the liquid molecules 154 included. In addition, each sample's pH, optical properties, and concentration of nonpurgeable organic carbon (NPOC) were measured. More details are provided in 155 156 Supporting Information (Text S2).

Data Processing. The detected molecules in both phases can be characterized by a variety of indices, including oxygen-to-carbon ratio (O/C), hydrogen-to-carbon ratio (H/C), ring and double-bound equivalence (RDBE), aromaticity equivalent (Xc),

160 carbon oxidation state (OSc), and modified aromaticity index (AI_{mod}). Kendrick mass 161 defect (KMD) analysis was performed to identify an array of compounds containing 162 similar repeated functional groups⁴⁷. Compounds whose elemental compositions only 163 differ by the number of subunits would possess the same KMD level and line up 164 horizontally in the KMD plot⁴⁸. In this study, two subunits (CH₂, COO) were considered 165 as key to describe the observed photochemistry. The characterization approaches are 166 detailed in Text S3.

167 Correlation analysis and principal component analysis (PCA) were applied to the 168 temporal VOC concentration profiles during the second half of light exposure to gain 169 more insights into the relevant chemistry. Independent samples t-test was used to 170 compare the experimental results of two different samples, whereas paired samples t-171 test compared the scenarios of pre and post irradiation for a specific sample. The 172 outcome can be characterized by the *p* value, which indicates significant discrepancy 173 or variation, when lower than 0.05. The measurement results from mass spectrometers are reserved for at least two decimals. 174

175 **Carbon Flux Calculation.** Following our previous studies on the photochemistry 176 and VOC production at the air/sea interface^{38,39,41,49,50}, we calculated the carbon flux 177 associated with the emission of a given VOC normalized to the surface of water 178 available under the simulated sunlight, denoted as SF_{VOCs} (nmol C m⁻² s⁻¹), according

180
$$SF_{VOCs} = \sum_{i=1}^{m} \frac{L \times C_i \times N_i \times 10^{-3}}{V_T \times A}$$
(1)

181 where *L* is the flow rate of air (~ 3.33 mL s⁻¹), C_i is the concentration of VOC *i* (ppb),

182 N_i is the number of carbon atoms in VOC *i* (dimensionless), V_T is the molar volume of

183 gas at temperature T (24.4 L mol⁻¹ at 24.3 °C), A is the area of surface water exposed to

184 air $(17.85 \times 10^{-4} \text{ m}^2)$, *m* is the number of organic species produced over the interface

185 (38 for Kuujjuarapik and 55 for Umiujaq).

Given that the irradiance (*I*) of the natural study site differs from the experimental setting, we estimated the natural VOC fluxes, denoted as N*Fvocs* (nmol C m⁻² s⁻¹), by using a linear relationship between the photoinduced VOC production and irradiance as previously reported in other cases⁴¹⁻⁴³, as shown by equation (2). This correction method facilitates the comparison between the VOC emission fluxes derived from the different laboratory experiments and/or in-field measurements.

192
$$NF_{VOCS} = SF_{VOCS-dark} + (SF_{VOCS-light} - SF_{VOCS-dark}) \times \frac{I_{natural}}{I_{experimental}}$$
(2)

193 where SFvocs-dark and SFvocs-light are respectively the carbon fluxes attributed to the 194 VOCs emitted before and during irradiation, *Iexperimental* is the experimental irradiance (8.0 mW cm⁻²), Inatural is the natural irradiance. Hourly Inatural during the warm season 195 of 2019 (1st June - 30th September) was measured at a meteorological site near the 196 Umiujaq sampling site^{51,52}. Temporal variations of *Inatural* are given in Figure S5. Solar 197 198 irradiance at Kuujjuarapik is very similar to that at Umiujaq (see Figure S4), therefore 199 we used the Umiujaq data for both sites. Because the experimental irradiance was 22.7 200 folds of the natural daily mean level, the 180 min of irradiation in the laboratory 201 simulates ~ 3 days' photochemistry in the field. Similar photoinduced VOC production was previously observed across other water samples under strong and weak irradiation in a smog chamber system³⁸, and therefore no irradiance threshold exists in such type of investigations. Applying equation 2 (as described above) aims only to underline some differences in the pattern of emitted VOCs at both sampling sites, without capturing the full environmental variability of this Arctic region.

Implication Evaluation. To estimate whether the photochemically produced VOCs could locally (i.e., at the sampling sites) affect the O₃ budget under a given NO_x concentration, O₃ formation potential, *OFP* (μ g m⁻³), was calculated according to²⁸:

$$210 \quad OFP = \sum_{j=1}^{n} [X]_j \times MIR_j \tag{3}$$

211 where $[X]_j$ is the concentration of VOC j (µg m⁻³), MIR_j is the maximum incremental

212 reactivity (MIR) coefficient of VOC *j*, *n* is the number of VOC with documented *MIR*

213 (22 for Kuujjuarapik and 33 for Umiujaq). *MIR* can be taken from the updated research

214 results (https://intra.engr.ucr.edu/~carter/SAPRC/, last access: 10th, November 2022).

The calculated OFP can be normalized by the available air/water interfacial area to the simulated efflux of O₃, denoted as SFo₃ (nmol m⁻² s⁻¹), which is further corrected by the outdoor irradiance as that done for carbon flux to the natural Fo₃, denoted as NFo₃ (nmol m⁻² s⁻¹), according to:

219
$$SF_{O_3} = \frac{L \times OFP \times 10^{-3}}{M_{O_3} \times A}$$
 (4)

220
$$NF_{O_3} = SF_{O_3} \times \frac{I_{natural}}{I_{experimental}}$$
 (5)

where M_{O3} is the molar mass of O₃ (48.0 g mol⁻¹).

222 **RESULTS**

As soon as the pond water was illuminated in the actinic region, a strong net 223 224 photochemical production of a variety of VOCs can be observed (Figure 2a). The 225 production is sustained and increasing, demonstrating that products are building up as 226 long as the light was presented. Such production stopped when the light was switched 227 off, with the concentration decreasing over time. While Figure 2a shows the integrated 228 signal over all VOCs, more product details are presented in Table S1. The nonmonotonic increases suggest that 2nd generation products are formed in both the liquid 229 230 and gas phases. The net concentrations of total VOCs produced by photochemistry, 231 determined by subtracting the background signal of each VOC (i.e., the averaged dark 232 level before irradiation) by the corresponding maximum signal observed during the 180 233 min of irradiation, are 13.83 ± 2.01 ppb (mean \pm standard error [SE]) and 20.14 ± 1.93 234 ppb for the samples from Kuujjuarapik and Umiujaq, contributed by 38 and 55 species, 235 respectively. That is, the Umiujaq water exceeds the Kuujjuarapik water in the VOC photoproduction (p < 0.05, Figure S6). Thirty volatile products are common to both 236 237 sites (Figure 2b).

Half of the VOCs emitted from the sunlit Kuujjuarapik sample contain \leq 3 carbons, and these compounds are responsible for ~ 87% of the concentration of the Kuujjuarapik VOCs (Figure 2c, top half). By contrast, for the Umiujaq VOCs, while the total concentration is mainly contributed by C1-C3 species, most compounds contain \geq 4 carbons. The concentration-weighted mean molecular mass of the Umiujaq

243	VOCs (53.26) is slightly greater than that of the Kuujjuarapik VOCs (50.98).
244	Furthermore, the VOCs are classified into CHO, CH, CHON, CHN, and CHS
245	subgroups according to their elemental compositions (Figure 2c, bottom half). The
246	CHO subgroup generally dominates the structure of VOCs observed, and the water
247	sample from Umiujaq is more efficient than that from Kuujjuarapik in producing all
248	these subgroups, except for the concentration of CHON and the number of CHS. The
249	average molecular formulas can be given as C2.370H4.839O1.026N0.064S0.011 for the
250	Kuujjuarapik VOCs and C2.651H5.189O0.948N0.065S0.003 for the Umiujaq VOCs.
251	Speciation of the observed photoproduced VOCs is given in Figure 2d. As the most
252	abundant product from the sunlit Kuujjuarapik pond surface water, CH4O accounts for
253	28.3% of the total VOC concentration, followed by C_2H_4O (20.8%), C_3H_6O (16.1%),
254	C ₃ H ₄ O (4.6%), C ₂ H ₂ O (3.2%), C ₂ H ₄ O ₂ (2.2%), CH ₃ NO (2.2%), C ₄ H ₈ O (2.1%), C ₄ H ₆ O
255	(1.4%), C ₂ H ₅ NO (1.3%) and other 28 compounds (accounting up to 17.8%). For the
256	pond water from Umiujaq, the ten most abundant compounds are C2H4O (30.9%),
257	CH4O (15.5%), C3H6O (14.8%), C3H4O (5.0%), C2H4O2 (2.4%), C4H8O (2.2%), C3H5N
258	(2.0%), C ₂ H ₂ O $(2.0%)$, C ₂ H ₆ O $(1.8%)$, C ₄ H ₈ $(1.7%)$, followed by the other 45 products
259	(accounting up to 21.8%). Also, for a given VOC, its relative production is negatively
260	correlated with its molecular weight (Figure S7).
261	Overall, three dominant species, CH4O, C2H4O, and C3H6O that were normally
262	identified as methanol, acetaldehyde, and acetone (or possibly propanal or methyl vinyl

263 ether), are responsible for > 60% of the total VOC concentrations for both samples. A

264 statistical treatment (PCA in Figure S8) further reveals that these three compounds are 265 not correlated and do originate from distinct processes, depending on the type of 266 thermokarst pond surface water. Furthermore, positive correlations are common for the 267 Umiujaq VOCs, whereas more negative correlations are observed among the Kuujjuarapik VOCs (Figure S9). In general, the Kuujjuarapik VOCs are more 268 269 oxygenated, unsaturated, and aromatic than the Umiujaq ones, as reflected by the higher 270 O/C and OSc, lower H/C and higher RDBE, and higher Xc and AImod, respectively 271 (Table S2). Such molecular differences are certainly associated with the compositions 272 of permafrost organic carbon in the two contrasted ponds studied here.

273 The photochemistry of both samples also differs significantly in the filtered liquid 274 phase. The NPOC levels were 1320 ± 16 and $1070 \pm 28 \mu$ M for the waters from 275 Kuujjuarapik and Umiujaq, respectively, before irradiation. After irradiation, as shown 276 in Figure S10, the NPOC level for Kuujjuarapik significantly increased to 1823 ± 170 277 μ M (p < 0.05), whereas that of Umiujaq generally showed no significant change, with 278 a value of $1003 \pm 123 \mu M$ (p > 0.05). Moreover, as shown in Figure S11, the pH 279 increased from 6.9 ± 0.0 to 7.4 ± 0.1 for the Umiujaq water (p < 0.05), while it decreased 280 from 7.1 \pm 0.0 to 6.9 \pm 0.0 for the Kuujjuarapik water (p < 0.05). Lastly, the Umiujaq 281 sample was observed to photobleach with irradiation, while for Kuujjuarapik the 282 absorbance increased over time for wavelengths below 300 nm (Figure S12). 283 Our observations indicate that the organics contained in the thermokarst ponds are

284 photochemically active and produce a variety of VOCs. Based on equation (1), under

285	the experimental setting, the integrated SF_{VOCs} for the Kuujjuarapik and Umiujaq ponds
286	are 2.83 and 4.61 nmol C m ⁻² s ⁻¹ , respectively. We further estimated the associated
287	NF_{VOCs} according to equation (2), and compared the two surface waters. The diel cycle
288	of hourly NFvocs is shown in Figure 3a, and both surface waters produced the highest
289	carbon fluxes at noon. The average NF_{VOCs} of the sampling day is estimated to be 0.21
290	nmol C m ⁻² s ⁻¹ across the Kuujjuarapik pond and 0.51 nmol C m ⁻² s ⁻¹ across the Umiujaq
291	pond. The photoinduced carbon flux across Kuujjuarapik pond primarily consists of
292	C ₃ H ₆ O (20.3%), C ₂ H ₄ O (17.5%), and CH ₄ O (12.0%), and that across Umiujaq pond is
293	mostly made of C ₂ H ₄ O (23.4%), C ₃ H ₆ O (16.7%), and CH ₄ O (5.8%), as shown by
294	Figure 3b. That is, acetaldehyde, acetone (or possibly propanal or methyl vinyl ether)
295	and methanol, which account for $> 60\%$ of the concentration of photoproduced VOCs
296	(see Figure 2d), also explain nearly half of the photoinduced VOC-associated carbon
297	flux over sunlit thermokarst ponds.
298	In order to further reveal the potential importance of the discovered photochemistry,

299 we speculate that the photoinduced VOC production for a longer time scale can be

300 estimated by extrapolating the measurement results to the natural irradiance of the warm

301 season (Figure S13). The highest daily NFvocs over the thermokarst ponds would occur

302 around the summer solstice. As irradiance decreased since then, the NFvocs over both

303 ponds generally decreased. Since NFvocs is influenced by temperature, irradiance, and

304 the properties of water, the atmospheric relevance of this photochemistry should be

305 accurately evaluated according to the actual situation.

306 **DISCUSSION**

To the best of our knowledge, this is the first detection and quantification of 307 photochemically produced non-methane VOCs from thermokarst ponds. The weak 308 309 VOC emissions before irradiation are probably explained by the diffusion to the gas 310 phase of volatile species photoproduced under natural conditions before sampling. This 311 is supported by the positive correlation between the individual VOC concentrations 312 before and during irradiation (Figure S14). Thermokarst ponds are therefore a 313 photochemical reactor that is able to transform permafrost carbon into atmospheric 314 VOCs. VOCs have been reported to be also emitted directly by thawing permafrost, without the intervention of liquid phase photochemistry⁵³⁻⁵⁵. Current understanding is 315 316 that these VOCs were produced by microbial metabolism, mostly anaerobic, and 317 accumulated over the years before being released by thawing⁵³⁻⁵⁵. Here, however, we 318 observed rapid VOC emission which is well correlated to irradiation conditions, so that 319 production can be ascribed to photochemistry with reasonable certainty, rather than 320 microbial production. This is in line with the observation of photoproducts over the liquid phase by others⁴³. Given that many photoinduced reactions at the air/water 321 322 interface are significantly enhanced in comparison to the corresponding gas- or liquidphase bulk processes⁵⁶, and that our experiments were performed in the absence of any 323 324 gaseous oxidants (and over very short residence time of the gas phase), the observed photochemical production of VOCs is mainly related to processes occurring at the 325 air/water interface, with a possible contribution of aqueous bulk photochemistry. 326

327 Figure 4 illustrates the generic mechanism summarizing the photochemical production of VOCs from the surface water in these ponds. Thawing permafrost 328 329 releases soil carbon into thermokarst ponds as POC and DOC, which both undergo an 330 array of photochemical reaction pathways producing secondary DOC (DOCs), 331 followed by the release of reacted DOC (DOCr) into the atmosphere as VOCs. 332 Depending on the actual chemistry taking place, partial oxidation or decarboxylation leads to the formation of DOCr, which is more volatile and therefore more easily 333 diffuses to the gas phase^{20,22,57}. Given that the DOC loading of the Kuujjuarapik pond 334 335 water significantly increased after irradiation, photolysis of POC to DOC may be 336 responsible for the photochemistry in this palsa-affected surface water. In contrast, the 337 decrease in absorbance of the Umiujaq water after irradiation indicates a decrease in 338 photoreactive compounds, implying a lower transformation of POC to DOC. 339 Interestingly, the Kuujjuarapik pond water, which is organic-rich because it is 340 affected by peat, is less efficient in the photochemical production of VOCs than the Umiujaq one which is organic-poor. Based on our previous research on the 341 342 photochemistry of humic acids in multiphase systems similar to the one explored

here^{38,41}, we are considering that the photoproduction observed for the thermokarst pond surface water is strongly impacted by the processes occurring at the air/water interface, and in particular by the availability of organic matters at the top layers of the liquid phase. In other words, if more appropriate precursors and oxidants are located at the sunlit air/water interface, more VOCs should be produced. Such availability is

highly pH-dependent, as for example more long-chain organic acids are increasingly
soluble with increasing pH. Herein, the photo-decreased pH of the Kuujjuarapik water
indicates the generation of organic acids derived from the photolysis of POC, during
which the availability of surface acids decreases, whereas the photo-increased pH of
the Umiujaq water implies the higher VOC production efficiency relying on photodecarboxylation that can be described as the breakdown of carboxylic acids (R-COOH)
to hydrocarbons (R-H) and CO₂ by sunlight^{20,58}:

$$355 \quad R - COOH \xrightarrow{n\nu} R - H + CO_2 \tag{6}$$

h ...

356 Another indication can be found in the evolution of the optical property of surface 357 water. The photobleaching of the Umiujaq water implies that large compounds may 358 photodegrade into smaller fragments with increased volatility. By contrast, the photo-359 enhanced absorbance of the Kuujjuarapik water indicates that photochemistry induces 360 the increase of chemical functionalities, as evidenced by the photo-increased liquid 361 aromaticity (Table S3). This variation may be attributed to the significant photolysis of 362 POC, followed by the conversion of non-condensed aromatics to monoaromatics with higher AI_{mod}⁵⁹. Overall, the photo-productivity of thermokarst pond water may thus be 363 better predicted by considering the segregation and processes at the air/water interface 364 365 than bulk organic concentration. This is in line with previous research in other aqueous systems, where the prevalence of interface processes was demonstrated^{38,41}. 366

367 As additionally suggested by the KMD plots for CH₂ (Figure S15) and COO (Figure
368 S16), most photo-enhanced VOCs would "theoretically" form after the removal of one

369 or more target subunits (e.g., CH₂ or COO) from the photodegraded liquid compounds. 370 As an example, for both samples, C4H4O2 may generate via the photo-dealkylation of 371 C5H6O2 and its homologous series, and the photo-decarboxylation of C4H6O3 or 372 C5H6O5 would produce C3H6O. More suggested "precursor-product" pairs are listed 373 alongside the KMD plots in the Supporting Information. This points to the fact that, as 374 expected, we do observe multigeneration products from permafrost carbon conversion, in agreement with the time profiles seen in Figure 2a. In comparison with the 375 376 Kuujjuarapik DOC, the Umiujaq DOC favors the photochemical production of VOCs 377 via decarboxylation channels. The efficient carboxyl loss follows the photo-increased 378 pH of the Umiujaq water and may be associated with more significant release of CO₂. 379 The relevant mechanisms may be much more complex than expected⁶⁰, depending on 380 the actual composition of the liquid phase.

381 Uncertainties may exist when reproducing the natural conditions. Firstly, the production and release of volatile species are obviously changing with temperature^{61,62}. 382 383 Secondly, the carbons in thermokarst ponds are refreshed because (i) the permafrost is always thawing by warming^{14,63}, rainfall⁶⁴, and wildfire⁹; (ii) wind above the shallow 384 385 pond will mix its liquid phase, including the vertical transport from deeper layers to interface⁶⁵; and (iii) the headwater streams feed the thermokarst ponds by bringing in 386 soil materials, especially during rain episodes^{57,65}. The refresh of thermokarst pond 387 water will further change its physical properties (e.g., surface tension, liquid viscosity) 388 389 that are associated with the performance of air/water interface. Third and perhaps most important, the "chemical history" (e.g., number of diurnal cycles) of permafrost carbons
will also alter the liquid properties^{20,22,26}, followed by the variation of sunlight intensity
at various depths and at the aqueous interface⁵⁷. Those effects are certainly to be taken
into account for the following studies.

394 **IMPLICATIONS**

It is worthwhile to mention that, the measurements performed by this study are too 395 396 limited in space and time to provide a robust assessment of the VOC emission over the numerous thermokarst landscapes $(3.6 \times 10^6 \text{ km}^2, \sim 20\% \text{ of the northern permafrost})$ 397 region¹⁴), but given the vast area of this continuously formed natural air/water interface 398 399 and the considerable emission fluxes of VOCs under sunlight, there is an urgent need 400 to take throughout research of this photochemistry over the thermokarst ponds/lakes on 401 Earth. Besides, temporal difference caused by the variations of carbon compositions, temperature, and irradiance are warranted in future works, as an attempt to reveal more 402

403 characteristics of the photochemistry over thermokarst landscapes.

Nevertheless, the photochemical production of atmospheric VOCs may strongly impact the local air composition by generating significant airborne O₃ in the presence of NO_x. Simply based on equations (3) and (4), the SF₀₃ is calculated to be 3.83 nmol $m^{-2} s^{-1}$ for the Kuujjuarapik sample and 7.34 nmol m⁻² s⁻¹ for the Umiujaq sample, under the experimental irradiance. Such results are much higher than the SF₀₃ over an authentic air/sea surface (0.51 nmol m⁻² s⁻¹) investigated under the same experimental condition⁴², suggesting that the photochemistry we discovered might present

411	considerable impacts on the local ozone budget. Corrected by equation (5), the daily
412	average NF ₀₃ is 0.17 nmol m ⁻² s ⁻¹ for Kuujjuarapik and 0.32 nmol m ⁻² s ⁻¹ for Umiujaq,
413	during the entire warm season. The speciated contribution to O ₃ formation is shown in
414	Figure S17. Among the Kuujjuarapik VOCs, C2H4O (acetaldehyde) is the largest O3
415	contributor and accounts for 35.7% of the NF03, followed by C3H4O (acrolein, 11.4%),
416	C ₃ H ₄ O ₂ (acrylic acid, 5.9%), C ₄ H ₆ O ₂ (methyl acrylate, 5.5%), and C ₄ H ₆ O (methyl vinyl
417	ketone, 5.5%). In contrast, the main O3 contributors among the Umiujaq VOCs are
418	C2H4O (acetaldehyde, 40.5%), C3H4O (acrolein, 9.4%), C4H8O (butanal, 4.4%), C4H8
419	(butene, 4.2%), and C ₄ H ₆ O (methyl vinyl ketone, 4.0%). The ranking of NF ₀₃ differs
420	from that of VOC concentrations, implying that the atmospheric impact of this
421	photochemistry may not be directly determined by the magnitude of VOC emission. In
422	addition to altering locally the O3 budget, the emissions of VOCs could also lead to an
423	increase in secondary organic aerosol formation, similar to what has been observed for
424	processes involving chromophoric materials at the air/sea interface ⁶⁶ . This however
425	needs to be verified by further measurements.

Even though the VOC flux data were not reported for typical thermokarst landscapes, a shallow post-glacial subarctic lake was recently explored for its in-field emission of six non-methane VOCs⁶⁷, and the comparison results are listed in Table S4. At the end of subarctic warm season (i.e., September), the lake presented net deposition fluxes of methanol, acetaldehyde, acetone, and dimethyl sulfide, in contrast to the net emission fluxes of isoprene and monoterpenes, eventually causing a net deposition flux of total

432	VOCs (-0.287 nmol C m ⁻² s ⁻¹). The negative flux measured for the lake demonstrates
433	that processes other than photochemistry were involved. In this study, both thermokarst
434	pond waters owned more significant fluxes of methanol, acetaldehyde, and acetone than
435	other three VOCs, corresponding to the photoinduced VOC emission fluxes of 0.080
436	and 0.172 nmol C m ⁻² s ⁻¹ for Kuujjuarapik and Umiujaq, respectively. While in nature
437	it is difficult to isolate the contribution of photochemical processes to VOC emissions,
438	the comparable absolute VOC fluxes obtained in the field and laboratory imply that the
439	photochemistry reported here may have considerable implications for the budget and
440	composition of atmospheric VOCs over subarctic water bodies.
441	Another obvious extension of this work is the comparison of the carbon fluxes of
442	VOCs, which are currently not in-field measured over thermokarst ponds, with those of
443	CH4 and CO2, which have been frequently measured over water bodies in the Arctic.
444	The CO ₂ and CH ₄ fluxes have been measured above ponds close to those we
445	sampled ^{19,68} . The average fluxes of CO ₂ and CH ₄ were respectively 143.84 and 5.80
446	nmol C m ⁻² s ⁻¹ over a lithalsa-affected thermokarst lake, and 676.28 and 40.60 nmol C
447	m ⁻² s ⁻¹ over a palsa-affected thermokarst lake. In comparison, the VOC fluxes we
448	measured are lower by 1-3 orders of magnitude. However, measurement conditions are

- 449 not comparable. Our experiments were performed over samples with a maximal depth 450 of ~ 0.55 cm, in contrast to the water bodies with depths of typically 2.0 m over which
- 451 the gaseous CO₂ and CH₄ were produced. Furthermore, one study on a thermokarst lake
- 452 near Kuujjuarapik²⁶ concluded that "only half of the light-associated DOC losses were

453	converted into CO ₂ ", so that the fraction of VOCs produced by pond water photolysis
454	may actually be significant. Meaningful comparisons would require operating under
455	similar conditions for CO2, CH4 and VOC measurements. This is an analytical
456	challenge, left for future research.



Figure 1. Overview of the studied thermokarst ponds.

(a) Geographical location of the sampling sites in Northern Quebec, Canada. The color scale distinguishes the type of permafrost. The study sites are located at the sporadic permafrost zone. Surrounding of the thermokarst ponds located at the (b) Kuujjuarapik and (c) Umiujaq sites.



Figure 2. Characterization of the photoproduction of VOCs.

(a) Temporal profile of the concentration of VOCs produced from thermokarst pond water. Irradiation is highlighted by the yellow shading. The surface water samples were collected at the Kuujjuarapik (denoted as K) and Umiujaq (denoted as U) sites located in Northern Quebec, Canada. (b) Venn diagram for the numbers of VOCs photoemitted from the different samples. (c) Concentration or number of the photoproduced organic vapors as a function of carbon amount or VOC subgroup, varying with the type of thermokarst pond water sample. (d) Concentration proportions of the photoinduced VOCs. The ten most abundant species are listed for each sample. In total, 38 and 55 photoinduced VOCs, accounting for the measured volume concentrations of 13.83 and 20.14 ppb, were observed for the K and U samples, respectively.



Figure 3. Evaluation of the photoproduction of VOCs.

(a) Diurnal variation of hourly natural VOC fluxes (NFvocs) over the studied thermokarst ponds. The emission flux of VOC-associated carbon can be calculated by equation (1) and further corrected by equation (2). The radiation data were recorded by a meteorological station near the Umiujaq pond during the sampling day: 4th September 2019.
(b) Contribution proportions of the photoinduced VOC-associated carbon fluxes (unit: nmol m⁻² s⁻¹) over different thermokarst ponds. The ten most abundant species are listed for each site.



Figure 4. Schematic diagram for the photoinduced VOC production.

The photochemical production of VOCs can be briefly described as the evolution of soil carbon lost from thawing permafrost to dissolved carbon in thermokarst pond along with the increased temperature, followed by the photogeneration of volatile atmospheric carbon in surface water (solid line arrows). Overall, this photochemistry connects the Earth's cryospheric carbon reservoir to the atmosphere, via the hydrosphere. In detail, it comprises the release of permafrost carbon into thermokarst ponds (dash-dot line arrow), the photo-reaction of particulate organic carbon (POC) and dissolved organic carbon (DOC) in the liquid phase (short-dash line arrow), the further conversion of secondary DOC (DOCs) acting as VOC precursors (short-dash line arrow), the diffusion of volatile reacted DOC (DOCr) into the atmosphere as VOCs (dot line arrow), as well as the retention of DOCr in the liquid phase (dash line arrow). DOCs can also be completely oxidized to CO₂.

ASSOCIATED CONTENT

Supporting Information. Description of experimental procedures, methods of data analysis, characteristics (irradiance and actinic flux) of the experimental and natural irradiation; detailed information and molecular characterization of the detected VOCs, photoinduced evolutions of liquid phase, calculation of ozone formation potentials.

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Notes

The authors declare no competing interest.

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