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# 1 Spatial heterogeneity of nitrification contributes to tree-grass

# 2 coexistence in West African savannas

3 Sarah Konaré<sup>1,2</sup>, Kouamé Fulgence Koffi<sup>2,3</sup>, Simon Boudsocq<sup>4</sup>, Jacques Gignoux<sup>2</sup>,

# 4 Jean-Christophe Lata<sup>2</sup>, Xavier Raynaud<sup>2</sup> and Sébastien Barot<sup>2</sup>

- 5 <sup>1</sup>UFR Sciences biologiques, Département Géosciences, Université Peleforo Gon Coulibaly, BP 1328
- 6 Korhogo, Côte d'Ivoire
- 7 <sup>2</sup>Sorbonne Université, Université de Paris-Cité, UPEC, IRD, CNRS, INRAE, Institute of Ecology
- 8 and Environmental Sciences, iEES Paris, 75005 Paris, France
- 9 <sup>3</sup>UFR des Sciences de la Nature, Station d'Ecologie de Lamto/CRE, Pôle de Recherche
- 10 Environnement et Développement Durable, Laboratoire d'Ecologie et du Développement Durable,
- 11 Université Nangui Abrogoua, 02 BP 801 Abidjan 02, Côte d'Ivoire
- 12 <sup>4</sup>Eco&Sols, INRAE, CIRAD, IRD, Montpellier SupAgro, Univ Montpellier, Montpellier, France
- 13 Corresponding author: Sarah Konaré
- 14 E-mail : <u>ksarahfaike@upgc.edu.ci</u>
- 15 Adress : UFR Sciences biologiques, Département Géosciences, Université Peleforo Gon Coulibaly,
- 16 BP 1328 Korhogo, Côte d'Ivoire
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#### 25 Abstract

In savannas, the coexistence between trees and grasses is determined by complex mechanisms
 based on water partitioning and disturbances. But little is known about the contribution of
 other resources such as N. In West African savannas, nitrification inhibition by grasses and
 nitrification stimulation by trees create spatial heterogeneity in nitrification fluxes and N
 stocks. Besides, savanna trees can also extend a part of their roots in the surrounding open
 area to take up N.

2. To investigate the role of the spatial heterogeneity of nitrification on tree-grass coexistence, we used a two-patch model simulating N dynamics between an open patch (without trees) and a tree clump patch (trees with grasses under their canopy). The open patch was characterized by a low nitrification rate while the tree clump patch was characterized by a high nitrification rate. Both patches were connected through horizontal fluxes due to soil horizontal exploration by tree roots. We tested coexistence for different spatial tree distributions as they are known to strongly influence savanna dynamics.

39 3. Our results show that the spatial heterogeneity of nitrification induces a spatial partitioning
40 between ammonium (NH<sub>4</sub><sup>+</sup>) and nitrate (NO<sub>3</sub><sup>-</sup>) promoting tree-grass coexistence. As
41 nitrification inhibition by grasses leads to high NH<sub>4</sub><sup>+</sup> availability in the open, the possibilities
42 of coexistence are optimized when trees have different preferences in the open versus under
43 their canopy. Tree-grass coexistence is thus observed when grasses prefer NH<sub>4</sub><sup>+</sup>, while trees
44 prefer NH<sub>4</sub><sup>+</sup> in the open and NO<sub>3</sub><sup>-</sup> under their canopy.

4. Contrary to random tree distribution, tree clumping enhances tree-grass coexistence.
 46 Intraspecific aggregation strengthens the effect of spatial heterogeneity, which decreases
 47 interspecific competition and favors tree-grass coexistence. By contrast, increasing the surface
 48 explored by tree roots in the open tends to increase tree-grass competition. This enhances the
 49 competitive ability of trees for N acquisition and consequently favors invasion by trees.

50 5. *Synthesis*. This study shows that this new coexistence mechanism based on mineral N
 51 partitioning into NH<sub>4</sub><sup>+</sup> and NO<sub>3</sub><sup>-</sup> can be determinant in the functioning of West African humid
 52 savannas. This mechanism probably interacts with disturbance-based mechanisms, but such
 53 interactions should be studied through new models.

54 Keywords: NH<sub>4</sub><sup>+</sup> versus NO<sub>3</sub><sup>-</sup> preference, nitrogen partitioning, nitrification, savanna, spatial
55 heterogeneity, spatial tree distribution, tree-grass coexistence

# 56 Introduction

57 Savanna vegetation is characterized by a continuous layer of grasses intermixed with a discontinuous 58 stratum of trees. The coexistence of many plant species has long been debated as it is in apparent 59 contradiction to the principle of competitive exclusion stipulating that two species competing for the 60 same resource cannot coexist over the long term (Hardin, 1960; Barot & Gignoux, 2004). In savannas, 61 tree-grass coexistence has been attributed to two main causes, resource partitioning and disturbances 62 (Sankaran et al., 2004)  $\Box$ . In dry savannas, the scarcity of resources, and especially water, leads to 63 niche partitioning as grasses are more competitive than trees in the topsoil layers while trees can 64 explore the soil vertically and take up water at greater depth (Schenk & Jackson, 2002). This niche 65 differentiation through different rooting depths can favor coexistence between trees and grasses (Walter, 1971; Walker & Noy-Meir, 1982). In contrast to dry savannas, wet savannas are not limited 66 67 by water and could turn into forests, but the presence of disturbances such as fire or herbivory reduces 68 the density of trees (Sankaran et al., 2004; Staver et al., 2011)  $\Box$ . Fire is the main disturbance limiting 69 woody cover by affecting the survival of tree seedlings and saplings (Gignoux et al., 2009)□. In the 70 same way, herbivores such as grazers and browsers can have negative effects on grass and tree growth 71 and therefore contribute to regulate the tree-grass ratio (Sankaran et al., 2008; Van Langevelde et al., 72 2003)□.

In West African savannas, the dominant perennial grass species inhibit nitrification, nitrification being the process of transformation of ammonium  $(NH_4^+)$  into nitrate  $(NO_3^-)$  (Lata et al., 2004; Subbarao et al. 2009; Srikanthasamy et al., 2018)  $\Box$ . This biological nitrification inhibition (BNI)

76 occurs through the release of grass root exudates impeding the activity of nitrifying microorganisms 77 (Lata et al., 2004; Srikanthasamy et al., 2018). Because  $NH_4^+$  is less prone to leaching than  $NO_3^-$ , 78 keeping nitrogen (N) in the NH<sub>4</sub><sup>+</sup> form decreases N losses by NO<sub>3</sub><sup>-</sup> leaching and therefore maintains 79 a high primary productivity (Boudsocq et al., 2009) . By contrast, in the Lamto humid savanna (Côte 80 d'Ivoire), the dominant tree species have been found to stimulate nitrification (Srikanthasamy et al., 81 2018). The mechanism explaining this stimulation is not known but one hypothesis would be that as 82 for grasses, specific molecules from the tree root system directly impact microbial communities. 83 Another hypothesis is that the observed increase of soil organic matter and soil humidity below tree 84 canopy could also increase soil microorganisms activity and therefore mineralization and nitrification 85 processes (Mordelet et al., 1993; Srikanthasamy et al., 2018)

86 In nutrient-limited ecosystems such as savannas, plant-soil feedbacks on N cycling can 87 locally alter N availability (Knops et al., 2002) by creating a spatial heterogeneity in soil resources, 88 and thereby influencing primary productivity. Trees and grasses, through their respective stimulation 89 and inhibition strategies, could induce a resource partitioning, if grasses preferentially absorb NH<sub>4</sub><sup>+</sup> 90 and trees NO<sub>3</sub>-promoting tree-grass coexistence in the Lamto savanna (Boudsocq et al., 2012; Konaré 91 et al., 2019). Some studies have shown that plant preferences for different chemical N forms (organic 92 and mineral N) can facilitate their coexistence (Ashton et al., 2010; McKane et al., 2002) . Spatial 93 heterogeneity due to nitrification heterogeneity could further reduce niche overlap, likely decreasing 94 interspecific interactions and promoting coexistence of different species even on a single limiting 95 resource (Barot & Gignoux, 2004; Chesson, 2000; Huston & DeAngelis, 1994)□. This leads to a complex picture requiring a spatially structured model to understand the consequences of plant-soil 96 97 feedbacks on NH<sub>4</sub><sup>+</sup> and NO<sub>3</sub><sup>-</sup> availability and tree-grass coexistence.

Many studies highlighted the relevance of spatial patterns in ecological dynamics (Grimm
et al., 1996; Grimm & Railsback, 2012). For example, intraspecific aggregation decreases the strength
of interspecific competition (Pacala, 1997; Stoll & Prati, 2001), which can slow down competitive
exclusion (Armstrong & McGehee, 1980). Spatial tree distribution is known to play an important

role in savanna dynamics and is strongly affected by disturbances (Menaut et al., 1990) . In addition, 102 103 tree clumps are generally considered as nutrient-rich patches (Mordelet et al., 1993) and this higher 104 soil fertility under tree canopy is partly due to the horizontal soil exploration by tree roots in the open 105 as it improves nutrient transfers between the open and tree clumps (Konaré et al., 2021). While spatial 106 heterogeneity tends to foster niche partitioning (Amarasekare, 2003), horizontal fluxes can minimize 107 the impact of this heterogeneity by homogenizing NH<sub>4</sub><sup>+</sup> and NO<sub>3</sub><sup>-</sup> availability between the two patches 108 (Barot et al., 2015; Barot et al., 2014). This could thus be influential for predictions of tree-grass 109 coexistence.

110 In contrast to the work of Konaré et al. (2019) that does not consider spatial heterogeneity 111 (all plants explore the same N pools) and Konaré et al. (2021) that focused on the effects of spatial 112 heterogeneity of nitrification fluxes on N fluxes and N budget, the goal of this study is to analyze the 113 role of the spatial heterogeneity in tree-grass coexistence. To do so, we used a modified version the 114 two-patch model previously published (Konaré et al., 2021) considering a patch of open area with a 115 low nitrification rate and a tree clump patch with a high nitrification rate. These two patches are 116 interconnected by horizontal fluxes (due to horizontal exploration of soil in the open area by tree 117 roots). The novelty of this model relies on considering the relation between two model parameters 118 (the proportion of tree roots in the open and the proportion of the savanna surface covered by tree 119 clumps) that depends on spatial tree distribution. Using this general model, we tested the following hypotheses: (i) Compared to the mean-field model (Konaré et al., 2019), accounting for the existence 120 121 of distinct N pools below and outside tree clumps fosters coexistence. (ii) Spatial heterogeneity in nitrification increases the likelihood of tree-grass coexistence with grasses preferring NH<sub>4</sub><sup>+</sup> and trees 122 123 preferring NO<sub>3</sub><sup>-</sup> under tree canopy and NH<sub>4</sub><sup>+</sup> in the open. Indeed, this spatial heterogeneity leads to a 124 spatial niche partitioning that reduces exclusion and favors coexistence. (iii) Tree-grass coexistence 125 depends on spatial tree distribution and is facilitated when trees are clumped. More specifically, tree 126 clumping reduces soil exploration by tree roots in the open and thus reduces competition between 127 trees and grasses in the open. (iv) Increasing the surface explored by tree roots in the open increases

128 horizontal fluxes between the open and the tree clump patches, which increases the competition

129 between trees and grasses for N and consequently reduces the chances of coexistence.

## 130 **Two-patch model**

#### 131 Model description

132 The two-patch model explores competitive interactions between trees and grasses through two 133 patches: an open patch occupied by grasses and some tree roots (patch 1), and a tree clump patch 134 occupied by trees with grasses growing below their canopy (patch 2) (Appendix S1). This model is a 135 spatially explicit extension of a previously published mean-field model (Konaré et al., 2019). The 136 model tracks N dynamics between plant compartments: grass biomass  $(G_1)$  in the open and, grass  $(G_2)$  and tree (T) biomass in the tree clump patch, and soil compartments: soil organic matter (O), 137 138 ammonium  $(N_A)$  and nitrate  $(N_N)$  in both patches (see Fig. 1, Konaré et al., 2021). Spatial 139 heterogeneity due to nitrification inhibition (BNI) by grasses and nitrification stimulation by trees is 140 represented by a low nitrification rate in the open  $(n_1)$  and a high nitrification rate under tree clumps 141  $(n_2)$ . Both patches are interconnected through horizontal fluxes due to the ability of trees to extend 142 horizontally their roots to take up nutrients in the open. In the model, N is supposed to be the limiting 143 factor of primary production and grass and tree growth depend on the acquisition of mineral N forms, 144 that can be acquired under two forms (NH<sub>4</sub><sup>+</sup> and NO<sub>3</sub><sup>-</sup>) with a certain preference ( $\beta$ ) for NH<sub>4</sub><sup>+</sup> versus 145 NO<sub>3</sub><sup>-</sup>. This preference ranges between 0 and 1 with high values of  $\beta$  corresponding to a high preference for NH<sub>4</sub><sup>+</sup>. N is absorbed by plants through N uptake rates  $\beta u$  for NH<sub>4</sub><sup>+</sup> uptake and  $(1-\beta)u$ 146 for NO<sub>3</sub><sup>-</sup> uptake. Plants release N into organic N pool at constant mortality rates  $d_G$ ,  $d_l$  and  $d_r$ 147 148 respectively for grasses, tree leaves and tree roots. Organic N mineralized into  $NH_4^+$  at rate m and 149  $NH_4^+$  can be transformed into  $NO_3^-$  at rate *n*. Each patch receives N inputs through dry and wet 150 depositions that provide N under organic and mineral forms into O, NA and NN compartments respectively at rates  $i_{O}$ ,  $i_{NA}$  and  $i_{NN}$ . Non-symbiotic fixation contributes to the input of NH<sub>4</sub><sup>+</sup> and is 151 152 included in  $i_{NA}$  (Abbadie, 2006). N losses from savannas result from the burning of the plant 153 compartments ( $l_G$  and  $l_T$ ), and leaching for other compartments ( $l_O$ ,  $l_{NA}$  and  $l_{NN}$ ). NO<sub>3</sub><sup>-</sup> losses by denitrification are included in  $l_{NN}$ . All parameters but nitrification rates are equal between the two patches to keep the model relatively simple and focus on the effects of nitrification heterogeneity (see model parameter Table, Appendix S2). All compartments are N stocks expressed as a quantity of N by surface unit, i.e kilograms of N per hectare of savanna (kg N ha<sup>-1</sup>). The system of differential

- 158 equations is the same used in Konaré et al. (2021):
- 159 Open patch (patch 1)

160 
$$\frac{dG_1}{dt} = (1 - \gamma)(\beta_G u_G N_{A1} G_1 + (1 - \beta_G) u_G N_{N1} G_1 - (d_G + l_G) G_1)$$
(1)

161 
$$\frac{dO_1}{dt} = (1 - \gamma) \Big( i_0 + d_G G_1 + d_r \left( \frac{\gamma}{1 - \gamma} \right) \alpha \, r \, T - (m + l_0) \, O_1 \Big)$$
(2)

162 
$$\frac{dN_{A1}}{dt} = (1 - \gamma) \left( i_{NA} + m O_1 - \left( \beta_G u_G G_1 + n_1 + l_{NA} + \beta_{T1} u_r \left( \frac{\gamma}{1 - \gamma} \right) \alpha r T \right) N_{A1} \right)$$
(3)

163 
$$\frac{dN_{N1}}{dt} = (1 - \gamma) \left( i_{NN} + n_1 N_{A1} - ((1 - \beta_G) u_G G_1 + l_{NN} + (1 - \beta_{T1}) \left( \frac{\gamma}{1 - \gamma} \right) u_r \alpha r T \right) N_{N1} (4)$$

164 Tree clump patch (patch 2)

165 
$$\frac{dG_2}{dt} = \gamma \left(\beta_G \, u_G \, N_{A2} \, G_2 + (1 - \beta_G) \, u_G \, N_{N2} \, G_2 - (d_G + \, l_G) \, G_2\right) \tag{5}$$

166 
$$\frac{dT}{dt} = \gamma (1 - \alpha) \left( \beta_{T2} u_r N_{A2} r T + (1 - \beta_{T2}) u_r N_{N2} r T \right) + \alpha$$
  
167 
$$\left( \beta_{T1} u_r N_{A1} r T + (1 - \beta_{T1}) u_r N_{N1} r T \right) - \left( d_l (1 - r) + d_r \alpha r + (1 - \alpha) d_r r + l_T \right) T$$

168 (6)

. .

169 
$$\frac{dO_2}{dt} = \gamma \left( i_0 + d_G G_2 + (1 - \alpha) d_r r T + d_l (1 - r) T - (m + l_0) O_2 \right)$$
(7)

170 
$$\frac{dN_{A2}}{dt} = \gamma \left( i_{NA} + m \, O_2 - \left( \beta_G \, u_G \, G_2 + (1 - \alpha) \, \beta_{T2} \, u_r \, r \, T + n_2 + \, l_{NA} \right) \, N_{A2} \right) \tag{8}$$

171 
$$\frac{dN_{N2}}{dt} = \gamma \left( i_{NN} + n_2 N_{A2} - \left( (1 - \beta_G) u_G G_2 + l_{NN} + (1 - \beta_{T2}) (1 - \alpha) u_r r T \right) N_{N2} \right)$$
(9)

#### 172 Spatial setting of the model

173 The description of the spatial setting in the model relies on the proportion of tree roots in the open ( $\alpha$ ) 174 and the proportion of the savanna surface covered by tree clumps ( $\gamma$ ) (leaving (1-  $\gamma$ ) to the open 175 savanna). Compared to Konaré et al. (2021) in which  $\alpha$  and  $\gamma$  had fixed values, in this version of the 176 model,  $\alpha$  is a function of  $\gamma$  and  $\gamma$  is a function of tree biomass (*T*). We assume that  $\alpha$  depends on the 177 proportion of tree cover and the spatial patterns of trees: for a given canopy and root system radius, 178  $\alpha$  should decrease with  $\gamma$ . However, this relation should also depend on tree distribution with lower 179 values of  $\alpha$  when trees are clumped and high values of  $\alpha$  when they are randomly or regularly 180 distributed. Although tree distribution can vary in different savanna ecosystems, it is often clustered 181 (Barot et al., 1999; Gignoux et al., 2006). Similarly, we assume a positive relationship between  $\gamma$  and 182 tree biomass *T*: the more trees, the more surface they occupy in savanna.

183 The shape of the relation between  $\alpha$  and  $\gamma$  was studied by simulating different distributions 184 of trees (random, clustered and regular). We assumed that trees were represented as two superimposed 185 discs describing the canopy and the root system as the majority of tree roots has been found within a 186 radius 10 m away from tree clump center (Menaut, personal communication). In our simulations, 187 canopy radius was set to 2 m and root system radius to 6 m. The relation between  $\alpha$  and  $\gamma$  was obtained 188 by calculating the proportion of root system that did not fall under the canopy of other trees for each 189 tree distribution and for different tree densities exploring a wide range of tree cover. We calculated 190 the relation between  $\alpha$  and  $\gamma$  for random, clumped and regular tree distributions. To simulate random 191 patterns, a Poisson process with a tree density varying from 1 to 100 trees ha<sup>-1</sup> by step of 10 and from 192 100 to 2000 trees ha<sup>-1</sup> by step of 200 was used. Clumped patterns were obtained by using a Matérn 193 cluster process (Matérn, 1960) with a tree density varying from 1 to 100 trees ha<sup>-1</sup> by step of 10 and 194 from 100 to 1000 trees ha<sup>-1</sup> by step of 100 with a mean clump radius of 2 m and a mean number of 195 trees per clump of 10. Finally, regular patterns were modeled using a Matérn hard-core process 196 (Matérn, 1960; 1986) with an inhibition distance of 2 m and the same tree densities as the random 197 patterns. To determine the relation between  $\alpha$  and  $\gamma$ , and between  $\gamma$  and T, we calculated the quantity 198 of tree roots outside tree canopy at the individual scale and the mean of all individual values to obtain 199  $\alpha$  for a given tree density. We then calculated  $\gamma$  for each tree density by determining the total space 200 occupied by trees. Linear regression models were then used to determine the relation between  $\alpha$  and 201 y. Besides, we used data from Menaut & César (1979) to establish the relation between y and tree 202 biomass (T) through a linear regression model. Equations from linear models and parameters were 203 included in the two-patch model to link  $\alpha$ ,  $\gamma$  and T.

# 204 Model analysis and parameterization

The analysis of the model relied on numerical simulations as it could not be analytically solved. All simulations were coded in R (R development Core Team, 2022) using deSolve package for the resolution of differential equations (Soetart et al., 2010). All simulations were run for 3000 years, which was sufficient to reach steady states for all compartments.

We used the same parameter sets as in Konaré et al. (2021; see Appendix S2) but run a completely different simulation experiment. These parameters are based on data from the Lamto savanna in Côte d'Ivoire (06°13'N, 05°02'W) (Abbadie et al., 2006) (Appendix S2).

212 Tree-grass coexistence was determined using the mutual invasibility criterion (Chesson & Ellner, 213 1989). This criterion considers a pair of invader and resident species (for example grasses invading a 214 forest, starting with a negligible biomass of 0.01 kg N/ha compared to the tree biomass, with a high 215 biomass of 10 kg N/ha, and then the reverse situation with trees invading a grassland). When the two 216 species are able to invade each other, they are assumed to mutually coexist over the long term 217 (Chesson & Ellner, 1989). Nitrification stimulation under tree clumps suggests a preference of trees 218 for NO<sub>3</sub><sup>-</sup> at least for roots under the tree canopy. Because NH<sub>4</sub><sup>+</sup> should be the dominant N form in the open due to nitrification inhibition by grasses, a preference of trees for NH<sub>4</sub><sup>+</sup> in the open is more 219 220 expected. We thus decided to distinguish the preference of trees for  $NH_4^+$  versus  $NO_3^-$  in the open 221 patch  $(\beta_{TI})$  and under tree clumps  $(\beta_{T2})$  and tree preference under tree clumps was fixed to a constant value ( $\beta_{T2} = 0.25$ ). Simulations of N stocks and fluxes were performed for different combinations of 222 223 grass and tree preference for NH<sub>4</sub><sup>+</sup> versus NO<sub>3</sub><sup>-</sup> in the open ( $\beta_G$  and  $\beta_{TI}$ ) varying between 0 and 1 with an increment of 0.005. We first simulated mutual invasion between trees and grasses for different 224 225 combinations of grass and tree preference for  $NH_4^+$  versus  $NO_3^-$  (tree preference was the same in the 226 open and under tree canopy:  $\beta_T$ ) versus different combinations of grass and tree preferences for NH<sub>4</sub><sup>+</sup> versus NO<sub>3</sub><sup>-</sup> in the open ( $\beta_{TI}$ ) to compare the effect of tree preference on tree-grass coexistence. 227 228 Nitrification rates in the open and under tree clumps were calculated by dividing nitrification fluxes 229 under grasses and under trees by their respective ammonium stocks (Srikanthasamy et al., 2018). We

230 increased the nitrification rate in the open to compare a spatial homogeneity to a spatial heterogeneous 231 nitrification rate to test the importance of heterogeneity in nitrification flux on tree-grass coexistence. 232 Moreover, because the proportion of tree roots in the open does not depend only on tree density or 233 biomass, but also on the spatial distribution of individual trees, we tested the effects of tree 234 distributions in space (random, regular and clustered) on tree-grass coexistence. We tested the 235 sensitivity of results to the root radius by increasing this radius from 6 to 12 m to assess the 236 relationship between  $\alpha$  and  $\gamma$  and then analyze the effect of these new equations on tree-grass coexistence. 237

#### 238 **Results**

# Links between soil exploration by tree roots and tree cover and between tree biomass and tree cover

241 For all spatial distributions tested, a negative linear relation was observed between  $\alpha$  and  $\gamma$  (p-value 242 < 0.0001 and R<sup>2</sup> = 0.999 for random and regular; p-value < 0.0001 and R<sup>2</sup> = 0.996 for cluster tree pattern) (Fig. 1a). As expected, the relation yields lower  $\alpha$  values in the case of clustered trees 243 244 compared to random and regular distribution. The regression lines whereas follow:  $\alpha = -0.897\gamma +$ 0.889,  $\alpha = -0.872\gamma + 0.890$  and  $\alpha = -0.698\gamma + 0.694$  respectively for random, regular and clustered 245 tree patterns. We observe a significant difference between cluster and random or regular patterns but 246 247 the difference between random and regular patterns was not significant. Fig. 1b shows that  $\gamma$ significantly increases with tree biomass (p-value < 0.005,  $R^2 = 0.962$ ), which leads to  $T = 530.56\gamma$  -248 249 2.81. Despite the potential nonlinearity between  $\gamma$  and T, we chose to use a linear relationship in our 250 model for simplicity. We used these results to parameterize simulations testing for tree-grass 251 coexistence.

# Tree-grass coexistence is affected by tree preference for NH<sub>4</sub><sup>+</sup> in the open versus under tree canopy

Model simulation for clustered tree distribution leads to three cases of coexistence: (1) a case where trees and grasses are spatially separated (no grasses under trees, only in the open hereafter called tree256 grass mosaic), (2) a case where grasses can only grow under tree clumps (no open area) hereafter 257 called savanna woodland, and (3) a case where grasses grow both under tree canopy and in the open 258 (hereafter called savanna). When tree preference is the same in the open and under tree canopy (Fig. 259 2a), grasses successfully invade and exclude trees for a large range of grass preference for NH<sub>4</sub><sup>+</sup> versus NO<sub>3</sub><sup>-</sup> and when trees prefer NO<sub>3</sub><sup>-</sup>. Tree-grass mosaic occurs when grasses have a higher 260 261 preference for NH<sub>4</sub><sup>+</sup> than trees ( $0 < \beta_T < 0.8$ ). By contrast, when tree preference for NH<sub>4</sub><sup>+</sup> is higher than 262 grasses, trees completely invade and exclude grasses. We also observe a small portion of the 263 parameter space where tree clumps establish when grasses prefer NO<sub>3</sub><sup>-</sup> and trees strongly prefer NH<sub>4</sub><sup>+</sup>  $(\beta_T > 0.98)$ . Savanna tree-grass coexistence is possible when grasses have a high preference for NH<sub>4</sub><sup>+</sup> 264 265 and trees prefer NO<sub>3</sub><sup>-</sup> and when grasses have a higher preference for NH<sub>4</sub><sup>+</sup> than trees ( $0 < \beta_T < 0.3$ ).

266 If trees have different preferences in the open ( $\beta_{TI}$  ranged between 0 and 1) versus under tree canopy ( $\beta_{T2}$  constant) (Fig. 2b), the sizes of the zone of savanna woodland and savanna strongly 267 268 increases while the zone of tree-grass mosaic decreases. Savanna is favored for a large range of tree 269 preference in the open ( $\beta_{Tl}$ ) and a grass preference for NH<sub>4</sub><sup>+</sup>. Tree clumps establishment becomes 270 possible when grasses and trees in the open prefer NH<sub>4</sub><sup>+</sup>. Taken together, a same tree preference in the 271 open and under tree canopy reduces possibilities of complete coexistence between trees and grasses 272 but favors tree-grass mosaic, while different tree preference in the open and under tree canopy increases possibilities of coexistence between trees and, grasses in the open and under tree canopy. 273 We thus consider different tree preference in the open versus under tree canopy for all other 274 275 simulations.

#### 276 Tree-grass coexistence depends on spatial tree distribution

Because results of regular and random tree distribution are virtually the same (Appendix S5), we only presented clustered and random distribution in the main text. Tree distribution strongly influences the conditions of coexistence between trees and grasses (Fig. 3). For random tree distribution, grasses invade and exclude trees when grasses have a higher preference for NH<sub>4</sub><sup>+</sup> than trees in the open ( $\beta_{T1}$ < 0.4). However, when tree roots in the open have a higher preference for NH<sub>4</sub><sup>+</sup> than grasses ( $\beta_G$  < 282 0.7), trees successfully invade and exclude grasses. The model also predicts a zone in which savanna 283 establishes when  $\beta_G > 0.7$  and  $0.4 < \beta_{TI} < 0.73$ , a zone where tree clumps invade when grasses and 284 trees in the open prefer NH<sub>4</sub><sup>+</sup> and a small zone of tree-grass mosaic.

Compared to random tree distribution, clustered pattern strongly increases the size of the zones of savanna and tree-grass mosaic. Tree-grass mosaic is enhanced when the grass preference for NH<sub>4</sub><sup>+</sup> is higher than the tree preference for NH<sub>4</sub><sup>+</sup> in the open ( $\beta_{TI}$ ). Savanna zone extends to large range of  $\beta_{TI}$  values and a preference of grasses for NH<sub>4</sub><sup>+</sup>. Switching from random to clustered tree distribution also reduces the size of the zone of invasion by grasses or by trees or by savanna woodland (Appendix S3).

#### 291 Comparison between spatial heterogeneity and spatial homogeneity cases

292 Spatial heterogeneity of nitrification due to its control by plants alters conditions of coexistence 293 between trees and grasses. We compared the case of spatial heterogeneity (Fig. 3) to a case of spatial 294 homogeneity where grasses do not inhibit nitrification (Fig. 4). In the case of a random tree 295 distribution, increasing nitrification rate in the open promotes invasion by tree clumps ( $\beta_G > 0.69$  and 296  $\beta_{TI} > 0.35$ ) or by trees only ( $\beta_G < 0.69$  and for all combinations of  $\beta_{TI}$ ). Compared to cases where 297 nitrification is spatially heterogenous (Fig. 3), we observed a zone of tree-grass mosaic for  $0.4 < \beta_G$ < 0.62 and  $\beta_{TI}$  < 0.13 and a zone of savanna when grasses prefer NH<sub>4</sub><sup>+</sup> ( $\beta_G$  > 0.62) and trees prefer 298 299 NO<sub>3</sub><sup>-</sup> in the open ( $\beta_{Tl} < 0.28$ ) in the spatial homogeneity case. For a clustered tree distribution, the 300 overall pattern is virtually the same, but a second zone of tree-grass mosaic appears when grasses and trees in the open strongly prefer respectively NO<sub>3</sub><sup>-</sup> ( $\beta_G < 0.2$ ) and NH<sub>4</sub><sup>+</sup> ( $\beta_{TI} > 0.95$ ). Increasing 301 nitrification rate reduces total coexistence between trees and grasses and facilitates zones with only 302 303 trees or trees with grasses under their canopy. This shows that different nitrification rates in the open 304 and in the tree clump patch improve chances of tree-grass coexistence.

#### 305 Soil exploration by tree roots affects tree-grass coexistence

Fig. 5 displays the outcome of mutual invasion between trees and grasses when the tree root system
radius is increased from 6 m to 12 m. As described above, savanna (grasses grow both under tree

308 canopy and in the open) in random distribution cases occurs when  $\beta_G > 0.7$  and  $0.4 < \beta_{Tl} < 0.73$  and 309 the size of this zone increases and extends to more combinations of  $\beta_{TI}$  when we switch from a random 310 to a clustered tree distribution (root radius = 6 m, Figs 5a and b). We observe the same tendencies 311 when the root radius is increased to 12 m (Appendix S4). Overall, clustered distributions tend to 312 increase the size of the savanna zone. However, increasing the surface explored by tree roots reduces 313 the sizes of the zone of tree-grass coexistence in cases of both random and clustered distributions 314 (Figs 5c and 5d). This is more perceptible for clustered patterns for which the size of the savanna and 315 tree-grass mosaic zones largely decreases while the zone of invasion by grasses increases (Fig. 5d).

# 316 **Discussion**

317 In our model, tree-grass coexistence occurs when plants have contrasted preferences for NH<sub>4</sub><sup>+</sup> versus 318 NO<sub>3</sub>. This is in agreement with theories showing that the coexistence of different species is possible 319 when they differ in their use of resources (Armstrong & McGehee, 1980). This is also in agreement 320 with published works on competition for  $NH_4^+$  and  $NO_3^-$  between plants (Boudsocq et al., 2012; 321 Konaré et al., 2019). These results are also supported by empirical studies on N partitioning indicating 322 that plant having different preferences for different chemical N forms can coexist (Ashton et al., 2010; 323 McKane et al., 2002) . Mineral N partitioning appears as a stabilizing mechanism (Chesson, 2000; 324 Barot & Gignoux, 2004) fostering coexistence between savanna trees and grasses. Here, we further 325 analyze the influences of these processes on tree-grass coexistence testing for the effects of spatial 326 heterogeneity in nitrification fluxes (hypothesis 1), mineral N partitioning into NH<sub>4</sub><sup>+</sup> and NO<sub>3</sub><sup>-</sup> 327 (hypothesis 2), spatial tree patterns (hypothesis 3) and horizontal soil exploration by tree roots (hypothesis 4). 328

#### 329 Comparison between the one-patch and the two-patch model

Compared to the mean-field model (Konaré et al., 2019), the two-patch model allows to simulate distinct pools of mineral N available below tree clumps and in the open. Although the sizes of the zone of savanna in the two-patch model (cluster distribution, Fig. 3) and in the mean-field model (Fig. 2A, Konaré et al. (2019)) are not significantly different, the two-patch model predicts different 334 possible cases of coexistence that corresponds to different landscape structures. Indeed, while the 335 mean-field model only allows one possible case of tree-grass coexistence, the two-patch model can 336 lead to three types of coexistence: savanna woodland (trees with grasses under their canopy only), 337 tree-grass mosaic (trees with grasses only in the open) or savanna (trees with grasses everywhere). 338 These results are in accordance with theories showing that coexistence in a spatially heterogeneous 339 environment is facilitated (Chesson, 2000; Amasekare, 2003) even with a single resource. In this case, 340 spatial heterogeneity separates the mineral N resource so that this resource is no longer fully shared 341 between the competitors leading to more diverse scenarios of coexistence in the two-patch than in the 342 one-patch model.

#### 343 Heterogeneity in nitrification influences tree-grass coexistence in the Lamto savanna

344 Studies on resource-based mechanisms of tree-grass coexistence in savannas have focused on water competition (Walter, 1971; Walker & Nov-Meir, 1982). Little is known, however, about other 345 346 resources that are essential for plant growth, such as N (Donzelli et al., 2013) . In our model, when 347 trees have the same preference in the open and under tree canopy, savannas occur when grasses have 348 a high preference for  $NH_4^+$  and trees a high preference for  $NO_3^-$  (Fig. 2). This confirms that tree-grass 349 coexistence is possible when grasses prefer  $NH_4^+$  and trees prefer  $NO_3^-$  at least under tree canopy. 350 Wang and Macko (2011) studied the preference of grasses depending on climatic conditions and 351 found that grasses tend to prefer NH<sub>4</sub><sup>+</sup> in humid areas. Moreover, ongoing studies on *Hyparrhenia* 352 diplandra in the Lamto savanna (Barot, personal communication) as well as a study on Andropogon 353 gavanus (Rossiter-Rachor et al., 2009) suggest that these two species that inhibit nitrification, have a preference for NH<sub>4</sub><sup>+</sup>. Unlike the mean-field model where trees and grasses coexist when grasses 354 355 prefer NO<sub>3</sub><sup>-</sup> and trees prefer NH<sub>4</sub><sup>+</sup> (Konaré et al., 2019), in the two-patch model, tree-grass coexistence 356 becomes possible when grasses prefer  $NH_4^+$  and trees prefer  $NO_3^-$ . This is more in accordance with 357 the Lamto savanna case as grasses are known to inhibit nitrification while trees stimulate it (Lata et 358 al., 2004; Srikanthasamy et al., 2018). The N limitation in the Lamto savanna (Abbadie et al., 2006)

has induced the evolution of different strategies: a conservative strategy for grasses through the BNI
capacity and an acquisitive strategy for trees through nitrification stimulation (Barot et al., 2015) .

361 Konaré et al. (2021) showed that  $NH_4^+$  uptake by tree roots outside tree canopy (35.58%) 362 of total N uptake by trees) increased tree biomass. Our results confirm this scenario because different 363 tree preference for  $NH_4^+$  versus  $NO_3^-$  in the open and under tree canopy (trees prefer  $NH_4^+$  in the open 364 and NO<sub>3</sub><sup>-</sup> under their canopy) increases the zone of savanna, i.e. tree-grass coexistence (Fig. 2). 365 Besides, simulations testing different values of tree preference for NH<sub>4</sub><sup>+</sup> versus NO<sub>3</sub><sup>-</sup> under their 366 canopy showed that increasing tree preference for NH<sub>4</sub><sup>+</sup> under tree canopy reduces the establishment 367 of savanna zones but promotes tree invasion (Appendix S6). These results suggest that plasticity of 368 tree root systems in their preferences for NH<sub>4</sub><sup>+</sup> versus NO<sub>3</sub><sup>-</sup> should occur depending on the spatial distribution of  $NH_4^+$  and  $NO_3^-$  (Britto & Kronzucker, 2013)  $\Box$ . In both cases (same versus different 369 370 tree preferences), our results show that local interactions between species and their environment can 371 induce a spatial heterogeneity leading to niche partitioning and thus promote their coexistence (Huston & DeAngelis, 1994; Amasekare, 2003). These different impacts of trees and grasses on 372 373 nitrification create small-scale heterogeneities by increasing the availability of their preferred N form, 374 which induces niche complementarity for N acquisition. In comparison, increasing nitrification rate 375 in the open reduces coexistence and favors the establishment of zones with only trees and savanna woodland (Fig. 4). Indeed, increasing this rate reduces heterogeneity in nitrification which increases 376 377 niche overlap (Chesson, 2000; Amasekare, 2003) and increases the competitive ability of trees for N. 378 Taken together, these feedbacks based on N recycling tend to stabilize the open and tree clump 379 patches and to favor tree-grass coexistence at a larger scale. Besides, the preference for NH<sub>4</sub><sup>+</sup> versus 380 NO<sub>3</sub>- qualitatively influences savanna dynamics, confirming that this preference can have important 381 consequences at the ecosystem scale on vegetation dynamics (Boudsocq et al., 2012) and 382 quantitatively on plant biomass (Appendix S7). While plasticity in plant preference for NH<sub>4</sub><sup>+</sup> versus 383 NO<sub>3</sub><sup>-</sup> has already been described, our results suggest that we should acquire experimental results to

384 measure this preference on savanna trees and its spatial variability and to analyze the underlying

385 mechanisms along Britto and Kronzucker's arguments (Britto & Kronzucker, 2013).

#### **386** Tree spatial distribution alters the conditions of coexistence

387 Many modeling studies have demonstrated that species coexistence could be facilitated by 388 intraspecific clustered distributions (Inouye, 1999; Hartley & Shorrocks, 2002). Our results confirm 389 that coexistence is easier when trees are clumped than when they are randomly distributed. In a 390 random tree distribution, grass and tree roots growing outside tree canopy easily overlap in the open. 391 This increases competitive interactions between trees and grasses for N acquisition in the open 392 through lateral N exchanges following the horizontal soil exploration by tree roots. The strong 393 competition between trees and grasses in the open prevents the establishment of tree-grass mosaic 394 and savanna zones whereas it facilitates invasions by grasses or by trees depending on their 395 preferences for  $NH_4^+$  versus  $NO_3^-$ . By contrast, the clumping of trees limits the proportion of tree 396 roots in the open as tree individuals are surrounded by more conspecifics (tree roots tend to be more 397 under the canopy of neighbouring trees). This decreases transfers of N between the open and the tree 398 clump patch through horizontal fluxes and highlights the impact of spatialization: intraspecific 399 aggregation tends to foster intraspecific competition over interspecific competition. Some studies 400 found that intraspecific aggregation is influential for species coexistence as it facilitates the 401 maintenance of weaker competitors (Stoll & Prati, 2001; Monzeglio & Stoll, 2005). Taken together, 402 tree spatial patterns influence the outcomes of competition by impacting the strength of intra relative 403 to interspecific competition (Stoll & Prati, 2001). Besides, in some savannas (Couteron & Kokou, 404 1997; Skarpe, 1991) □ and particularly in the Lamto savanna □(Barot et al., 1999; Gignoux et al., 405 2006), trees often show an aggregated spatial distribution. They form clumps leaving an important 406 proportion of the surface covered by grasses. Tree clumps are maintained by fires: reduced grass 407 biomass under tree shading decreases fire intensity and hence increases tree sapling survival 408 (Hochberg et al., 1994; Gignoux et al., 2006). This mechanism somehow increases the likelihood of 409 tree-grass coexistence by impeding fires (promoted by grass biomass) to kill all tree saplings. Our

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410 results suggest that tree clumps may favor tree-grass coexistence and savanna maintenance through

411 another mechanism: a decrease in the competition for mineral N, but empirical studies are needed to

412 support this effect of tree spatial distribution on savanna dynamics.

#### 413 The impact of horizontal soil exploration on coexistence

414 Horizontal soil exploration contributes to nutrient enrichment under tree canopy (Konaré et al., 2021) 415 but little is known about their possible effects on coexistence. Our results show that regardless of tree 416 distribution, savanna and tree-grass mosaic zones are reduced by the increase of the surface explored 417 by tree roots. Indeed, horizontal soil exploration by tree roots leads to spatial transfers of N between 418 the open and the tree clump patches. Increasing the proportion of tree roots in the open increases the 419 competitive ability of trees to take up N and thus the competition between trees and grasses for N in 420 the open. Although, cluster distribution allows a spatial partitioning by increasing intraspecific 421 competition, this proliferation of tree roots in the open increases the strength of interspecific 422 competition relative to intraspecific competition (Stoll & Prati, 2001) and consequently prevents grass 423 establishment. These lateral fluxes between these two patches lowers the spatial heterogeneity by 424 homogenizing the N availability between these two patches (Barot et al., 2015; Barot et al., 2014), 425 which tends to reduce complete coexistence and favors the zones of exclusion by trees and by grasses (Appendix S3). Moreover, when the size of the tree root system relative to the canopy increases, the 426 427 relation between the proportion of tree roots in the open and tree cover is virtually the same for each 428 distribution (Appendix S4). As a result, increasing the surface explored by tree roots decreases the 429 differences between random and clustered tree distribution.

# 430 Conclusion

Savanna ecosystems can be viewed as small scale meta-ecosystems (Loreau et al., 2003) where the spatial heterogeneity in nitrification between the two patches and the intensity of lateral N fluxes strongly influences the outcomes of competition between trees and grasses. Our findings confirm that spatial heterogeneity in nitrification promotes tree-grass coexistence when trees and grasses have different preference for  $NH_4^+$  versus  $NO_3^-$  with grasses preferring  $NH_4^+$  and trees  $NO_3^-$  under their 436 canopy. While increasing the soil exploration by tree roots in the open tends to increase the 437 competitive ability of trees to acquire N, intraspecific aggregation through cluster tree distribution 438 intensifies the effect of spatial heterogeneity, which lowers the competition between trees and grasses 439 favoring their coexistence. Mineral N partitioning into  $NH_4^+$  and  $NO_3^-$  can play an important role by 440 promoting tree-grass coexistence. Although the ability to inhibit nitrification seems to be common in 441 African perennial grasses and some crops (Lata et al., 2004; Subbarao et al., 2009; Rossiter-Rachor 442 et al., 2009, O'Sullivan et al., 2016; Lata et al. 2022), little is known about environmental conditions 443 that have selected for this behavior (but see Lata et al. (2022)). Even if this inhibition has not been 444 assessed in other savanna types, mineral N partitioning could be involved in the tree-grass coexistence 445 of all West African humid savannas that have virtually the same dominant grass and tree species as 446 Lamto savanna. The high performance of African grasses and their effects on N cycling in northern 447 Australian and South American savannas (D'Antonio & Vitousek, 1992; Rossiter-Rachor et al., 2009) 448 also seem to be linked to their BNI ability. Though the frequency of this ability in other African 449 grasses is not fully known (Lata et al., 2022), many African grasses seem to be able to inhibit 450 nitrification so that the new coexistence mechanism we emphasize should be studied at a large 451 geographical scale and particularly in Eastern and Southern African savannas. Besides, N fixing trees 452 and large herbivores, that are absent in the Lamto savanna, are known to be influential in the 453 functioning of other savannas such as East African savannas (Sankaran et al., 2008). These 454 mechanisms, that provide important N inputs through symbiotic fixation, animal dungs and urine and 455 impact N cycling differently, should be included in new models to assess the robustness of our results.

Clearly, our new mechanism of coexistence based on the partition of the mineral N resource, probably interacts with formerly identified mechanisms based on the impact of disturbances on tree demography to explain tree-grass coexistence (Higgins et al., 2000). Although fire is determinant for the maintenance of savanna structure by reducing woody cover (Gignoux et al., 2006), mineral N partitioning likely acts in interaction with fire as the intensity of fire depends on the biomass of flammable grasses, this grass biomass depending in turn on their N acquisition, essential

- 462 nutriment for their growth. For instance, nitrification inhibition by grasses has been shown to increase
- 463 grass biomass, which should increase fire intensity and its capacity to impede tree encroachment.
- 464 This would ultimately favor tree-grass coexistence. However, new models should be built to take into
- 465 account mechanisms based on both resource competition and demography.

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# 472 **Conflict of interest statement**

473 The authors declare no conflict of interest.

# 474 Author contributions

SK, KFK, S Bo, JG, JCL, XR and S Ba conceived the ideas and designed methodology. SK analyzed
the model and wrote the manuscript with the contribution of all co-authors. All authors gave final
approval for publication.

# 478 **Data availability statement**

479 All data regarding the model is included in the paper.

# 480 **References**

- 481 Abbadie, L. (2006). Nitrogen inputs to and outputs from the soil-plant system. In L. Abbadie, J.
- 482 Gignoux, X. Le Roux, & M. Lepage (Eds.), *Lamto: Structure, functioning, and dynamics of a*483 *savanna ecosystem* (pp. 255–275). Springer.
- 484 Abbadie, L., Gignoux J., Le Roux X., & Lepage, M. (2006). Lamto: Structure, functioning,
- 485 and dynamics of a savanna ecosystem. Springer.
- 486 Amarasekare, P. (2003). Competitive coexistence in spatially structured environments: A synthesis.
- 487 *Ecology Letters*, *6*, 1109–1122. https://doi.org/10.1046/j.1461-0248.2003.00530.x

- 488 Armstrong, R. A., & McGehee, R. (1980). Competitive Exclusion. *American Naturalist*, *115*, 151–
  489 170. https://doi.org/10.1086/283553
- Ashton, I. W., Miller, A. E., Bowman, W. D., & Suding, K. N. (2010). Niche complementarity due
  to plasticity in resource use: Plant partitioning of chemical N forms. *Ecology*, *91*, 3252–3260.
  https://doi.org/10.1890/09-1849.1
- Barot, S., Bornhofen, S., Boudsocq, S., Raynaud, X., & Loeuille, N. (2015). Evolution of nutrient
  acquisition: when space matters. *Functional Ecology*, 30, 283-294.
  https://doi.org/10.1111/1365-2435.12494
- Barot, S., Bornhofen, S., Loeuille, N., Perveen, N., Shahzad, T., & Fontaine, S. (2014). Nutrient
  enrichment and local competition influence the evolution of plant mineralization strategy: a
  modelling approach. *Journal of Ecology*, *102*, 357–366. https://doi.org/10.1111/13652745.12200
- Barot, S., & Gignoux, J. (2004). Mechanisms promoting plant coexistence: Can all the proposed
  processes be reconciled? *Oikos*, *106*, 185–192. https://doi.org/10.1111/j.00301299.2004.13038.x
- Barot, S., Gignoux, J., & Menaut, J.-C. (1999). Demography of a savanna palm tree: predictions from
  comprehensive spatial patterns. *Ecology*, *80*, 1987–2005. https://doi.org/10.1890/00129658(1999)080[1987:doaspt]2.0.co;2
- Boudsocq, S., Lata, J. C., Mathieu, J., Abbadie, L., & Barot, S. (2009). Modelling approach to analyze
  the effects of nitrification inhibition on primary production. *Functional Ecology*, *23*, 220–230.
  https://doi.org/10.1111/j.1365-2435.2008.01476.x
- 509 Boudsocq, S., Niboyet, A., Lata, J. C., Raynaud, X., Loeuille, N., Mathieu, J., ... Barot, S. (2012).
- 510 Plant preference for ammonium versus nitrate: a neglected determinant of ecosystem functioning?
- 511 *The American Naturalist*, *180*, 60–69. https://doi.org/10.1086/665997
- 512 Britto, D. T., & Kronzucker. H. J. (2013). Ecological significance and complexity of N-source
- 513 preference in plants. *Annals of Botany*, *112*, 957–963. https://doi.org/10.1093/aob/mct157

- 514 Chesson, P., & Ellner, S. (1989). Invasibility and stochastic boundedness in monotonic competition
- 515 models. Journal of Mathematical Biology, 27, 117–138. https://doi.org/10.1007/bf00276099
- 516 Chesson, P. (2000). General theory of competitive coexistence in spatially-varying environments.
- 517 *Theoretical Population Biology*, *58*, 211–237. https://doi.org/10.1006/tpbi.2000.1486
- 518Couteron, P., & Kokou, K. (1997). Woody vegetation spatial patterns in a semi-arid savanna of519BurkinaFaso,WestAfrica.PlantEcology,132,211–227.
- 520 https://doi.org/10.1023/a:1009723906370
- 521 D'Antonio, C. M., & Vitousek, P. M. (1992). Biological invasions by exotic grasses, the grass/fire
  522 cycle, and global change. *Annual review of ecology and systematics*, *23*, 63–87.
- 523 https://doi.org/10.1146/annurev.es.23.110192.000431
- Donzelli, D., De Michele, C., & Scholes, R. J. (2013). Competition between trees and grasses for
  both soil water and mineral nitrogen in dry savannas. *Journal of Theoretical Biology*, *332*, 181–
  190. https://doi.org/10.1016/j.jtbi.2013.04.003
- Gignoux, J., Barot, S., Menaut, J. C., & Vuattoux, R. (2006). Structure, Long-Term Dynamics, and
  Demography of the Tree Community. In L. Abbadie, J. Gignoux, X. Le Roux, & M. Lepage
- 529 (Eds.), Lamto: Structure, functioning, and dynamics of a savanna ecosystem (pp. 335–378).
  530 Springer.
- Gignoux, J., Lahoreau, G., Julliard, R., & Barot, S. (2009). Establishment and early persistence of
  tree seedlings in an annually burned savanna. *Journal of Ecology*, *97*, 484–495.
  https://doi.org/10.1111/j.1365-2745.2009.01493.x
- Grimm, V., Frank, K., Jeltsch, F., Brandl, R., Uchmanski, J., & Wissel, C. (1996). Pattern-oriented
  modelling in population ecology. *Science of Total Environnement*, 183, 151–166.
  https://doi.org/10.1016/0048-9697(95)04966-5
- 537 Grimm, V., & Railsback, S. F. (2012). Pattern-oriented modelling: A "multi-scope" for predictive
- 538 systems ecology. *Philosophical Transactions of the Royal Society B: Biological Sciences*, 367,
- 539 298–310. https://doi.org/10.1098/rstb.2011.0180

- 540 Hardin, G. (1960). The Competitive Exclusion Principle. *Science*, *131*, 1292–1297.
  541 https://doi.org/10.1126/science.131.3409.1292
- Hartley, S., & Shorrocks, B. (2002). A general framework for the aggregation model of coexistence. *Journal of Animal Ecology*, *71*, 651–662. https://doi.org/10.1046/j.1365-2656.2002.00628.x
- Higgins, S. I., Bond, W. J., & Trollope, W. S. W. (2000). Fire, resprouting and variability: a recipe
  for grass-tree coexistence in savanna. *Journal of Ecology*, *88*, 213–229.
- 546 https://doi.org/10.1046/j.1365-2745.2000.00435.x
- Hochberg, M.E., Menaut, J. C., & Gignoux, J. (1994). The influences of tree biology and fire in the
  spatial structure of the West African savannah. *Journal of Ecology*, *82*, 217–226.
  https://doi.org/10.2307/2261290
- Huston, M., & DeAngelis, D. (1994). Competition and coexistence: the effects of resource transport
  and supply rates. *The American Naturalist*, *144*, 47–79. https://doi.org/10.1086/285720
- Inouye, B. D. (1999). Integrating nested spatial scales: implications for the coexistence of competitors
  on a patchy resource. *Journal of Animal Ecology*, 68, 150-162. https://doi.org/10.1046/j.13652656.1999.00272.x
- Knops, J. M. H., Bradley, K. L., & Wedin, D. A. (2002). Mechanisms of plant species impacts on
  ecosystem nitrogen cycling. *Ecology Letters*, 5, 454–466. https://doi.org/10.1046/j.14610248.2002.00332.x
- Konaré, S., Boudsocq, S., Gignoux, J., Lata, J. C., Raynaud, X., & Barot, S. (2021). Spatial 558 559 heterogeneity in nitrification and soil exploration by trees favour source-sink dynamics in a 560 humid savanna: А modelling approach. Functional Ecology, 35, 976-988. https://doi.org/10.1111/1365-2435.13762 561
- 562 Konaré, S., Boudsocq, S., Gignoux, J., Lata, J. C., Raynaud, X., & Barot, S. (2019). Effects of mineral
- 563 nitrogen partitioning on tree-grass coexistence in West African savannas. *Ecosystems*, 22, 1679–
- 564 1690. https://doi.org/10.1007/s10021-019-00365-x

- Lata, J. C., Degrange, V., Raynaud, X., Maron, P. A., Lensi, R., & Abbadie, L. (2004). Grass
  populations control nitrification in savanna soils. *Functional Ecology*, 18, 605–611.
  https://doi.org/10.1111/j.0269-8463.2004.00880.x
- Lata, J. C., Le Roux, X., Koffi, K. F., Yé, L., Srikanthasamy, T., Konaré, S., & Barot, S. (2022). The
- 569 causes of the selection of Biological Nitrification Inhibition (BNI) in relation to ecosystem
- 570 functioning, and a research agenda to explore them. *Biology and Fertility of Soils*, *58*, 207–224.
- 571 https://doi.org/10.1007/s00374-022-01630-3
- Loreau, M., Mouquet, N., & Holt, R. D. (2003). Meta-ecosystems: a theoretical framework for a
  spatial ecosystem ecology. *Ecology Letters*, 6, 673–679. https://doi.org/10.1046/j.14610248.2003.00483.x
- 575 Matérn, B. (1960). Spatial variation stochastic models and their application to some problems in forest 576 surveys and other sampling investigations. *Medd. Statens Skogsforskn-ingsinst*, *49*, 1–144.
- 577 Matérn, B. (1986). Spatial Variation. Lecture Notes in Statistics 36. Springer
- 578 McKane, R. B., Johnson, L. C., Shaver, G. R., Nadelhoffer, K. J., Rastetter, E. B., Fry, B., ... Murray,
- 579 G. (2002). Resource-based niches provide a basis for plant species diversity and dominance in

580 arctic tundra. *Nature*, 415, 68–71. https://doi.org/10.1038/415068a

- Menaut, J. C., Gignoux, J., Prado, C., & Clobert, J. (1990). Tree community dynamics in a humid
  savanna of the Côte d'Ivoire: modelling the effects of fire and competition with grass and
  neighbours. *Journal of Biogeography*, *17*, 471–481. https://doi.org/10.2307/2845379
- 584 Menaut, J. C., & César, J. (1979). Structure and primary productivity of Lamto savannas, Ivory Coast.
- 585 *Ecology*, *60*, 1197–1210. https://doi.org/10.2307/1936967
- Monzeglio, U., & Stoll, P. (2005). Spatial patterns and species performances in experimental plant
   communities. *Oecologia*, *145*, 619–628. https://doi.org/10.1007/s00442-005-0168-3
- 588 Mordelet, P., Abbadie, L., & Menaut, J. C. (1993). Effects of tree clumps on soil characteristics in a
- 589 humid savanna of West Africa (Lamto, Côte d'Ivoire). Plant and Soil, 153, 103-111.
- 590 https://doi.org/10.1007/BF00010549

- 591 O'Sullivan, C.A., Fillery, I. R. P., Roper, M. M., & Richards, R. A. (2016). Identification of several
- wheat landraces with biological nitrification inhibition capacity. *Plant and Soil*, 404, 61–74.
  https://doi.org/10.1007/s11104-016-2822-4
- Pacala, S. W. (1997). Dynamics of plant competition. In M. J. Crawley (Ed.), *Plant Ecology*, (pp.
  532–555). Oxford.
- 596 Rossiter-Rachor, N. A., Setterfield, S. A., Douglas, M. M., Hutley, L. B., Cook, G. D., & Schmidt,
- 597 S. (2009). Invasive *Andropogon gayanus* (gamba grass) is an ecosystem transformer of nitrogen
- relations in Australian savanna. *Ecological Applications*, 19, 1546–1560.
  https://doi.org/10.1890/08-0265.1
- R Development Core Team. (2022). R: A language and environment for statistical computing.
  Version 4.2.2.
- Sankaran, M., Ratnam, J., & Hanan, N. (2008). Woody cover in African savannas: the role of
  resources, fire and herbivory. *Global Ecology and Biogeography*, *17*, 236–245.
  https://doi.org/10.1111/j.1466-8238.2007.00360.x
- Sankaran, M., Ratnam, J., & Hanan, N. P. (2004). Tree-grass coexistence in savannas revisited insights from an examination of assumptions and mechanisms invoked in existing models.
   *Ecology Letters*, 7, 480–490. https://doi.org/10.1111/j.1461-0248.2004.00596.x
- Schenk, H. J., & Jackson, R. B. (2002). Rooting depths, lateral root spreads and belowground
  aboveground allometries of plants in water limited ecosystems. *Journal of Ecology*, *90*, 480–
- 610 494. https://doi.org/10.1046/j.1365-2745.2002.00682.x
- Skarpe, C. (1991). Spatial patterns and dynamics of woody vegetation in an arid savanna. *Journal of Vegetation Science*, *2*, 565–572. https://doi.org/10.2307/3236039
- 613 Soetaert, K., Petzoldt, T., & Setzer, R. W. (2010). Solving differential equations in R: Package
  614 deSolve. *Journal of statistical software*, *33*, 1–25. https://doi.org/10.18637/jss.v033.i09
- 615 Srikanthasamy, T., Leloup, J., N'Dri, A. B., Barot, S., Gervaix, J., Koné, A. W., ..., & Lata, J. C.
- 616 (2018). Contrasting effects of grasses and trees on microbial N-cycling in an African humid

617	savanna. Soil Biology and Biochemistry, 117, 153–163.						
618	https://doi.org/10.1016/j.soilbio.2017.11.016						
619	Staver, A. C., Archibald, S., & Levin, S. (2011). Tree cover in sub-Saharan Africa: Rainfall and fire						
620	constrain forest and savanna as alternative stable states. Ecology, 92, 1063-1072.						
621	https://doi.org/10.1890/10-1684.1						
622	Stoll, P., & Prati, D. (2001). Intraspecific aggregation alters competitive interactions in experimental						
623	plat communities. Ecology, 82, 319–327. https://doi.org/10.1890/0012-						
624	9658(2001)082[0319:iaacii]2.0.co;2						
625	Subbarao, G. V., Nakahara, K., Hurtado, M. P., Ono, H., Moreta, D. E., Salcedo, A. F.,, & Ito, O.						
626	(2009). Evidence for biological nitrification inhibition in Brachiaria pastures. Proceedings of						
627	the National Academy of Sciences, 106, 17302–17307.						
628	https://doi.org/10.1073/pnas.0903694106						
629	Van Langevelde, F., Van De Vijver, C. A. D. M., Kumar, L., Van De Koppel, J., De Ridder, N., Van						
630	Andel, J.,, & Rietkerk, M. (2003). Effects of fire and herbivory on the stability of savanna						
631	ecosystems. <i>Ecology</i> , 84, 337–350. https://doi.org/10.1890/0012-						
632	9658(2003)084[0337:eofaho]2.0.co;2						
633	Walker B. H., & Noy-Meir, I. (1982). Aspects of stability and resilience of savanna ecosystems. In						
634	B. J. Huntley, & B. H. Walker (Eds.), Ecology of tropical savannas, (pp. 556-590). Springer.						
635	Walter, H. (1971). Ecology of tropical and subtropical vegetation. Oliver & Boyd.						
636	Wang, L., & Macko, S. A. (2011). Constrained preferences in nitrogen uptake across plant species						
637	and environments. Plant, Cell & Environment, 34, 525-534. https://doi.org/10.1111/j.1365-						
638	3040.2010.02260.x						
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**Figure 1**: (a) Proportion of tree roots in the open ( $\alpha$ ) as a function of tree cover ( $\gamma$ ) according to cluster, random and regular tree distribution, (b) Tree biomass as a function of tree cover ( $\gamma$ ) for a root radius equals to 6 m.

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650 Figure 2: Mutual invasibility plots between trees and grasses according to grass ( $\beta_G$ ) and tree preference for  $NH_4^+$ . (a): trees have identical preferences for  $NH_4^+$  in the open and the tree patches 651 652  $(\beta_T)$ . (b): trees have different preferences for NH<sub>4</sub><sup>+</sup> in the open  $(\beta_T)$  and the tree patch  $(\beta_{T2})$ . In these simulations,  $\beta_{T2}$  was set constant to 0.25. Simulations correspond to clustered tree distributions. 653 654 Invasion zones: G1+G2: grasses invade and exclude trees, T: Trees invade and exclude grasses, G1+T: 655 tree-grass mosaic (coexistence between trees and grasses in the open), G2+T: savanna woodland 656 (coexistence between trees and grasses under their canopy), G1+G2+T: savanna (coexistence 657 between trees and grasses under their canopy and in the open).

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Figure 3: Mutual invasibility plots between trees and grasses according to random (a) and clustered
(b) tree distributions at the savanna scale. Invasion zones: G1+G2: grasses invade and exclude trees,
T: Trees invade and exclude grasses, G1+T: tree-grass mosaic (coexistence between trees and grasses
in the open), G2+T: savanna woodland (coexistence between trees and grasses under their canopy),
G1+G2+T: savanna (coexistence between trees and grasses under their canopy).

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**Figure 4**: Mutual invasibility plots between trees and grasses according to random (a) and clustered (b) tree distributions at the savanna scale. These figures correspond to cases where grasses do not inhibit nitrification (spatial homogeneity:  $n_1 = n_2 = 4.16$  yr<sup>-1</sup>). Invasion zones: G1+G2: grasses invade and exclude trees, T: Trees invade and exclude grasses, G1+T: tree-grass mosaic (coexistence

669	between trees and grasses in the open), G2+T: savanna woodland (coexistence between trees and
670	grasses under their canopy), G1+G2+T: savanna (coexistence between trees and grasses under their
671	canopy and in the open)

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26	invasion by grasses while black lines the fluxes less influenced by the invasion by grasses.				

**Figure S4**: Proportion of tree roots in the open ( $\alpha$ ) as a function of tree cover ( $\gamma$ ) according to cluster,

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Figure S5-a: Mutual invasibility plots between trees and grasses according to grass ( $\beta_G$ ) and tree ( $\beta_{TI}$ ) preference for NH<sub>4</sub><sup>+</sup> in the open patch at the savanna scale. Simulations corresponds to regular tree distribution with cases of spatial homogeneity ( $n_1 = n_2 = 4.16 \text{ yr}^{-1}$ ) (left) and cases of spatial heterogeneity ( $n_1 \neq n_2$ ) (right). Invasion zones: G1+G2: grasses invade and exclude trees, T: Trees invade and exclude grasses, G1+T: tree-grass mosaic (coexistence between trees and grasses in the open), G2+T: savanna woodland (coexistence between trees and grasses under their canopy), G1+G2+T: savanna (coexistence between trees and grasses under their canopy and in the open).

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Figure S5-b: Mutual invasibility plots between trees and grasses according to grass ( $\beta_G$ ) and tree ( $\beta_{Tl}$ ) preference for NH<sub>4</sub><sup>+</sup> in the open patch at the savanna scale. The graph corresponds to simulations of regular tree distributions with a root system radius of 6 m (left) and 12 m (right) when determining the relation between  $\alpha$  and  $\gamma$ . Invasion zones: G1+G2: grasses invade and exclude trees, T: Trees invade and exclude grasses, G1+T: tree-grass mosaic (coexistence between trees and grasses in the open), G2+T: savanna woodland (coexistence between trees and grasses under their canopy), G1+G2+T: savanna (coexistence between trees and grasses under their canopy and in the open).

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Figure S6: Mutual invasibility plots between trees and grasses according to random (left) and clustered (right) tree distributions at the savanna scale. The preference of trees under tree canopy increases from the top to the bottom and corresponds respectively to  $\beta_{T2}=0.25$ ;  $\beta_{T2}=0.35$  and  $\beta_{T2}=0.5$ . Invasion zones: G1+G2: grasses invade and exclude trees, T: Trees invade and exclude grasses, G1+T: tree-grass mosaic (coexistence between trees and grasses in the open), G2+T: savanna woodland (coexistence between trees and grasses under their canopy), G1+G2+T: savanna (coexistence between trees and grasses under their canopy).

54	<b>Figure S7-a</b> : Grass (patterns a and b) and tree (patterns c and d) biomass as a function of grass ( $\beta_G$ )
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56	scale.
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59	$(\beta_{TI})$ preference for NH <sub>4</sub> <sup>+</sup> in the open patch according to clustered tree distributions at the savanna
60	scale.
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# 80 Appendix S1: Schematic representation of the two-patch model

Tree clump



86 Figure S1

# 92 Appendix S2: Model parameters

# 93 Table 1: Model parameters

Parameters	Definition	Unit	Values			
Grass parameters						
$d_G$	Turnover rate of grass	yr-1	0.6			
$l_G$	Rate of N losses from grass compartment	yr <sup>-1</sup>	0.4			
$u_G$	N uptake rate	ha kg <sup>-1</sup> N yr <sup>-1</sup>	0.14186			
$\beta_G$	Preference for NH <sub>4</sub> <sup>+</sup>	No unit	-			
Tree parameters						
$d_r$	Turnover rate of tree roots	yr <sup>-1</sup>	0.08			
$d_l$	Turnover rate of tree leaves	yr <sup>-1</sup>	0.073			
$l_T$	Rate of N losses from tree compartment	yr <sup>-1</sup>	0.11			
<i>u</i> <sub>r</sub>	N uptake rate by tree roots	ha kg <sup>-1</sup> N yr <sup>-1</sup>	0.08			
$\beta_{TI}$	Preference for $NH_4^+$ in the open patch	No unit	-			
$\beta_{T2}$	Preference for $NH_4^+$ in the tree clump patch	No unit	0.25			
α	Fraction of roots in the open	No unit	-			
γ	Tree clumps proportion	No unit	-			
r	Root shoot ratio	No unit	0.5			
Soil paramete	ers					
i <sub>O</sub>	N organic input to the savanna	kg N ha <sup>-1</sup> yr <sup>-1</sup>	16.5			
m	N mineralization rate	yr-1	0.025			
lo	N loss from the N organic compartment in surface soil layer	yr <sup>-1</sup>	0.0027			
i <sub>NA</sub>	$\rm NH_4^+$ inputs to the savanna	kg N ha <sup>-1</sup> yr <sup>-1</sup>	23			
<i>n</i> <sub>1</sub>	Nitrification rate in the open patch	yr-1	0.09			
<i>n</i> <sub>2</sub>	Nitrification rate in the tree clump patch	yr-1	4.16			
l <sub>NA</sub>	NH <sub>4</sub> <sup>+</sup> loss rate	yr-1	0.0133			
i <sub>NN</sub>	NO <sub>3</sub> - inputs to the savanna	kg N ha <sup>-1</sup> yr <sup>-1</sup>	4.1			
l <sub>NN</sub>	NO <sub>3</sub> <sup>-</sup> loss rate	yr-1	2.7			

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# 98 Appendix S3: Graphical representation of scenarios of invasion by trees versus invasion by

# 99 grasses

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- 102 Figure S3-a
- 103



106 Figure S3-b

#### 107 Appendix S4: Relation between soil exploration by tree roots and tree cover

We performed simulations for different tree distributions to determine the relation between the proportion of tree roots in the open ( $\alpha$ ) and tree cover ( $\gamma$ ) to analyze the effect of spatial patterns on tree-grass coexistence. This relation has been tested for random, cluster and regular tree distributions.

111 We considered trees with 2 m crown radius and 12 m root radius for each distribution in 1ha plot.

For all spatial distributions tested, a negative linear relation was observed between  $\alpha$  and  $\gamma$  (p-value < 0.0001 and R<sup>2</sup>=0.998 for random; p-value < 0.0001 and R<sup>2</sup>=0.999 for cluster; p-value < 0.0001 and R<sup>2</sup>=0.999 for regular tree pattern).  $\alpha$  significantly decreases while  $\gamma$  increases regardless tree distributions. In contrast to the results of the main text where cluster distribution was different to random and regular distributions (Figure 1a), we observe here that the relation between  $\alpha$  and  $\gamma$  is virtually the same for all patterns. This leads to these linear equations:  $\alpha = -0.984\gamma + 0.971$ ,  $\alpha = -0.977\gamma$  +0.973 and  $\alpha = -0.932\gamma + 0.924$  respectively for random, regular and clustered tree distributions.

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# Appendix S6: Effect of tree preference under their canopy depending on spatial tree distributions

We did simulations for different combinations of grass and tree preference for  $NH_4^+$  in the open. 158 159 Here, we tested different tree preference for NH<sub>4</sub><sup>+</sup> under tree canopy by gradually increasing tree 160 preference for NH<sub>4</sub><sup>+</sup> under their canopy ( $\beta_{T2} = 0.35$ ;  $\beta_{T2} = 0.5$ ) knowing that the value by default used 161 in the main text is  $\beta_{T2} = 0.25$ . Regarding random pattern, when tree preference under tree canopy for 162 NH<sub>4</sub><sup>+</sup> increases ( $\beta_{T2} = 0.35$ ), trees tend to exclude grasses, which reduces the zone of savanna. This 163 also increases conditions of coexistence between trees and grasses in the open. When tree preference 164 under tree canopy for NH<sub>4</sub><sup>+</sup> increases further ( $\beta_{T2} = 0.5$ ), the zone of savanna becomes very small 165 while the zone of tree-grass mosaic increases. The overall pattern is similar for cluster pattern but the 166 region of coexistence between trees and grasses in the open is much larger than in the random tree pattern case. 167 Increasing tree preference for NH<sub>4</sub><sup>+</sup> under tree canopy favors the establishment of trees and the 168

exclusion of grasses in random and cluster patterns. This also leads to an increase of the zone of treegrass mosaic (trees and grasses in the open) but strongly reduces the zone of savanna (grasses grow both under tree canopy and in the open) in both tree patterns.

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# 191 Appendix S7: Plant biomass depending on tree distributions



- 194 Figure S7-a







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205 Figure S7-b