

Fisheries outcomes of marine protected area networks: Levels of protection, connectivity, and time matter

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- ¹ Fisheries outcomes of marine protected area networks: levels of
- ² protection, connectivity and time matter
- 3 Short running title: Networks of protected areas benefit fisheries
- 4
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30 Abstract

31 Establishing large networks of fully protected marine protected areas (MPAs) is challenging because of displacement costs for fisheries. The use of partially protected areas is often proposed 32 33 as an alternative. However, how conservation and fisheries outcomes of MPA networks are mediated through time by the level of protection remain uncertain. Here we use a metapopulation 34 model of a commercially exploited demersal coastal fish to assess conservation and fisheries 35 outcomes of alternative management policies. We compare the temporal performances of non-36 spatial management, large MPAs, or networks of MPAs in an overfished case study. In addition, 37 we assess how the magnitude of both outcomes is mediated by larval connectivity and level of 38 protection. We distinguish the relative contribution to fisheries outcomes of unprotected areas in 39 between MPAs, and unprotected areas further away, receiving less displaced fishing effort and 40 potential biomass export. We show that spatial management outperforms non-spatial 41 management, that conservation and fisheries outcomes increase with increasing levels of 42 protection, that fisheries outcomes in areas in between MPAs are higher when connected through 43 larval dispersal, and that increases in catch are preceded by short-term decrease. Our results call 44 for an increase in protection levels to meet both ecological and fisheries management goals. 45

47 Main Text

48

49 Introduction

Increasing human pressure on the ocean is causing unprecedented impacts on marine ecosystems 50 (O'Hara et al. 2021). Overfishing is the main driver of change (IPBES, 2019) and threatens 51 sustainable fisheries. Marine Protected Areas (MPAs) are an area-based management tool that is 52 expected to deliver both conservation and fisheries benefits (Reimer et al. 2020). MPAs that are 53 well-managed, well-enforced and with strict enough protection levels accrue fish size, abundance 54 and biomass within their borders (Edgar et al. 2014; Zupan et al. 2018). These conservation 55 benefits can increase with the time of protection (Claudet et al. 2008) and be exported outside 56 protected areas via recruitment subsidy and spillover, and hence support fisheries (Pelc et al. 2010; 57 Harrison et al. 2012; Di Lorenzo et al. 2016, 2020). 58

59 Networks of MPAs are touted to balance conservation and fisheries benefits of MPAs by limiting the size of MPAs and/or offering a spatial configuration that can reduce fishers' displacement costs 60 to unprotected areas of the network in between MPAs (Roberts 2001; Grorud-Colvert et al. 2014). 61 While fishable areas within MPA networks are expected to benefit the most from larval dispersal 62 and spillover (two components of connectivity) from the networked MPAs (Barceló et al. 2021), 63 64 they also receive most of the displaced fishing effort (Halpern et al. 2004). Fisheries benefits outside individual MPAs (Pelc et al. 2010; Harrison et al. 2012; Di Lorenzo et al. 2016) or outside 65 MPA networks (Hastings & Botsford 2003; Gaines et al. 2010; Harrison et al. 2020) have been 66 extensively studied. However less attention, if any, has been given in distinguishing the potential 67 synergistic effects of networked MPAs inside a network and the overall effect outside the network. 68 This implies comparing the relative ecological and fisheries outcomes of networks on unprotected 69 areas between MPAs with those on unprotected areas further away from the area subject to 70 71 protection.

72 Another component influencing conservation and fisheries outcomes of MPA networks is the level of protection (a classification based on the potential impacts on species size and number, and on 73 habitats, of allowed activities with an MPA; Horta e Costa et al. 2016). Full and high protection 74 levels confer the largest conservation benefits (Zupan et al. 2018; Grorud-Colvert et al. 2021). 75 76 Since the level of protection regulates the amount of fishing pressure that can remain within partially protected areas, it also drives the amount of displaced fishing pressure outside those protected 77 areas. The highest levels of protection are mostly implemented in residual areas (Devillers et al. 78 2015). In areas under higher rates of fishing pressure, the majority of MPAs are of lower protection 79 levels (sensu Horta e Costa et al. 2016), largely allowing fishing activities within their borders 80

(Dureuil et al. 2018; Claudet et al. 2020). Globally, the majority of MPAs do not offer high protection
 (Pelc et al., in press).

Here, using a metapopulation model of a demersal coastal species experiencing overharvesting (Belharet et al. 2020), we compare the potential conservation (standing biomass) and fisheries (catch) outcomes through time of a set of non-spatial and spatial management scenarios: i) setting catch limits; ii) implementing large MPAs; iii) implementing networks of MPAs of smaller size. For each scenario, we assess how the conservation and fisheries outcomes are mediated by the level of protection (or of catch limit) and connectivity (in this work represented by larval dispersal), in different locations.

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91 Methods

92 1. <u>Metapopulation model</u>

We use an age-structured, discrete-time and spatially explicit metapopulation model of the white 93 seabream, Diplodus sargus (Linnée 1758), developed by Belharet et al. (2020) to assess the effects 94 of alternative configurations of management scenarios on conservation and fisheries outcomes. 95 The metapopulation model describes the key biological traits and processes influencing the 96 demographic dynamics of this demersal coastal species (i.e., reproduction, larval dispersal, 97 recruitment, body growth, sexual maturation, natural and fishing mortality). It explicitly considers 98 connectivity among different sub-populations with larval dispersal. Due to the limited vagility of 99 white seabream adults (Di Franco et al. 2018), adult displacement between cells is not represented 100 in the model (see Supporting Information for details on model description). 101

The model is first calibrated (see Supporting Information) and run for 100 years to reach equilibrium before starting the simulations described below. Our study area covers the coastal area located between latitudes 41.8-42.6 and the longitude 3.10-3.77. The spatial resolution is about 2 x 2 km (i.e., one grid cell of the model, for a total of 86 cells).

We build n=3 networks of six MPAs (each MPA is represented by one grid cell). The three networks are implemented at different locations to cover the whole study area and to better account for the spatial variability due mainly to larval connectivity. The MPAs composing each network are separated in space by unprotected cells (see Supporting Information for details on modelednetworks).

111 2. <u>Management scenarios</u>

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We create an overfishing context by increasing the fishing mortality rate that left 10% of the total 112 unexploited biomass remaining (Worm et al. 2009, see Supplementary Information). We assess 113 and compare potential conservation and fisheries outcomes by running several management 114 scenarios in the overfished context (presented below) and in the non-overfished context (fishing 115 mortality rate set as in Belharet et al., 2020, results presented in Supporting Information). Each 116 management scenario is systematically compared with its associated control scenario, i.e., a 117 scenario with the same parameters but where neither spatial nor non-spatial management is 118 implemented. Management scenarios are the following (Figure 1, Table 1): 119

- non-spatial fishery management, in which catch limits are modelled through an
 evenly distributed reduction in fishing mortality (across all 86 cells);
 - implementation of a large MPA (6 cells in a spatially contiguous arrangement);
 - implementation of a network of MPAs (6 MPAs of one cell, non-adjacent to each other).

To assess how each management scenario can be mediated by levels of protection, we model the following configurations:

- full protection: 100% reduction of fishing mortality in protected areas (large MPAs
 or networked MPAs). The equivalent fishing mortality is evenly redistributed in the
 adjacent unprotected cells;
- three levels of partial protection: respectively 75% (strong protection), 50%
 (intermediate protection) and 25% (low protection) reduction of fishing mortality in
 protected areas. The equivalent fishing mortality is redistributed in the adjacent
 unprotected cells;
- non-spatial scenarios: the three levels of reduced fishing mortality (strong,
 intermediate and low) are not concentrated in 6 protected cells but evenly
 distributed across all 86 cells.

Each management scenario is tested with the previous levels of protection (4 simulations per scenario). Simulations are run over a period of 40 years preceded by a period without protection of 4 years (before and after impact data). Control simulations are run over a period of 44 years (Figure
1, Table 1).

As we are also interested in understanding if the ability of MPA networks to deliver both conservation and fisheries benefits is dependent on the network's connectivity, we test an additional group of simulations where the connectivity matrix is modified to exclude larval exchanges among network's cells in both managed and control scenarios (Figure 1, Table 1, "unconnected network"). We assess the impact of each management scenario in different locations by defining focal areas, described in Figure 1.

147 3. <u>Statistical analysis</u>

We use a meta-analytical approach (Hedges et al. 1999) to assess the effectiveness of each management measure. First, we calculate effect sizes to compare each test simulation T with the corresponding control simulation C. The effect size associated with scenario i is calculated as the log-response ratio R of biomass (or catch) in each area j and in each year k with respect to the corresponding control scenario:

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$$R_{i,j,k} = \frac{1}{n} \sum_{j=1}^{k} \ln\left(\frac{X_{T_{i,j,k,n}}}{X_{C_{i,j,k,n}}}\right)$$

where *n* is the number of spatial replicates for scenario *i*; $X_{T_{i,j,k,n}}$ and $X_{C_{i,j,k,n}}$ are the total biomass or catch in all cells of area *j* and year *k*, in test *T* and control *C* simulations, respectively.

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For each effect size $R_{i,i,k}$, 95% confidence intervals were calculated as:

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$$CI_{R_{i,j,k}} = R_{T_{i,j,k}} \pm u_{\alpha/2} * \sqrt{\frac{S_{T_{i,j,k}}^2}{n}};$$

where *u* is the two-tailed critical value of the standard normal distribution at the significance level α , and $S_{T_{i,j,k}}^2$ is the variance associated with the effect size in area *j* and in year *k*.

All the analyses are carried out using the statistical software R (R Core Team, 2020).

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163 **Results**

Implementation of non-spatial fishery management scenarios (scenario T0, Figure 1), with reductions in fishing mortality evenly distributed throughout the region rather than concentrated in protected cells, leads to negligible differences in total biomass or total catch compared to the control scenario (Figures 2 and 3).

Implementation of large fully protected MPAs (scenario T1, Figure 1) results in overall average 168 increases of about 90% in biomass ($R = 0.64 \pm 0.11$, Table S4) and 60% in catch ($R = 0.47 \pm 0.10$, 169 Table S5) across the whole region, after 40 years of protection. This is due to a 877% increase in 170 biomass inside the fully protected area ($R = 2.28 \pm 0.16$, Table S4), and a 65% increase in biomass 171 $(R = 0.52 \pm 0.12)$, Table S4) and about the same increase in catch $(R = 0.51 \pm 0.12)$, Table S5) 172 outside the fully protected area (Figure 2 and 3). Overall gains in biomass and catch decrease as 173 levels of protection decrease (Figure 2), with only full protection and the two most restrictive levels 174 175 of partial protection providing long-term increases in biomass and catch across the region (Figure 2 and 3). Inside the partially protected areas, catches first decline, with the largest catch losses 176 associated with the strongest levels of protection (Figure 3). Then, catches recover between 5 and 177 15 years of protection, for low and high levels of protection, respectively, reaching higher values 178 than those without protection (Figure 3). 179

Implementation of connected networks of fully protected MPAs leads to similar increases, in 180 biomass (both inside the protected areas and outside the networks), and catch (outside the 181 networks) as in the large fully protected MPAs scenarios (Figure 2 and 3). In the unprotected areas 182 in between networked MPAs, there is first a slight decrease in biomass, compared to the absence 183 of protection, but biomass then starts to be larger, resulting in a 30% increase after 40 year of 184 protection ($R = 0.25 \pm 0.04$, Table S4). Catch also increases over the same period, compared to 185 the case without the implementation of a network of fully protected areas, to almost a 50% increase 186 $(R = 0.40 \pm 0.06)$, Table S5) after 40 years of protection. These gains in biomass and catch in 187 fishable areas in between the fully protected networked MPAs are higher than those observed 188 outside the large fully protected areas. Biomass and fisheries benefits in all three areas decrease 189

with decreasing levels of protection. Benefits in catch are only observed for full and the two most
 restrictive (strong and intermediate) levels of partial protection (Figures 2 and 3).

In the absence of connectivity (scenario T3, Figure 1), no biomass or catch benefits are observed
 in unprotected areas between the MPAs of the network (Figures 2 and 3). Within MPAs, biomass
 gains are similar to those of connected networks.

In a non-overfished situation, biomass increases inside connected MPAs follow the same dynamics as in the overfished situation but with a smaller magnitude of increase ($R = 1.17 \pm 0.01$ after 40 years of full protection, Table S4).

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199 Discussion

Here, we assessed for the first time the relative contribution of different levels of protection in MPA 200 201 networks to conservation and fisheries outcomes for a commercially exploited demersal coastal fish. Although our results are species-specific, the general patterns that emerge are also potentially 202 valid for other coastal species characterized by low mobility of adult life stages. We showed that 203 networks of partially protected areas can effectively support both fisheries and conservation, with 204 benefits increasing with the level of protection. We also confirm the role of connectivity for the 205 fisheries effectiveness of a network of MPAs, emphasizing the importance of distinguishing 206 unprotected areas in between MPAs from those further away. 207

Our most compelling result is that protection level matters not only for conservation outcomes but also for fisheries outcomes. For conservation, stronger protection levels generate higher biomass, as observed in recent empirical studies (Zupan et al. 2018; Turnbull et al. 2021). For fisheries outcomes, time matters. Gains in catch are not only linked to biomass gains, but also to the spatial dynamics of fisheries. For all levels of protection, the initial decrease in catch is short and quickly offset by the increase in biomass. Thus, after a few years, even with lower fishing pressure in the MPAs, catches are at least equivalent to what they would have been without protection.

Inside MPAs, patterns of response to protection are very similar for the three spatial management 215 scenarios. After a rapid increase in biomass -about 10 years for fully protected MPAs and 15 years 216 for partially protected MPAs – gains tend to stabilize after 20 years of protection, as also evidenced 217 by empirical studies on other demersal species (MacNeil et al. 2015; Magris 2021; Ferreira et al. 218 2022). In the non-overfished situation, the increase in biomass for full and partial protection (from 219 20% to 240%) is comparable to that reported in the literature on similar reef-associated demersal 220 species in the Mediterranean (Giakoumi et al. 2017) or globally (Zupan et al. 2018). In the 221 overfished situation, biomass increases for full (+800%) and partial protection (from 30 to 300%) 222

are larger, because the rate of change in fish biomass is higher when reducing fishing mortality
 inside MPAs by a large proportion. These results are consistent with those of Ziegler et al. (2022),
 who showed that MPAs response ratio increase
 in heavily fished situations, simply because of high deterioration of stocks outside MPAs.

Outside MPAs, full and high protection levels lead to a notable increase in biomass (+25% to +75%), slightly higher with an MPA network than with a large MPA. In our study, MPAs can become saturated by adults due to their limited displacement. Single large MPAs generate large numbers of larvae that are mainly retained within the protected area, but do not all become adults due to density dependence (Melià et al. 2020). In a network, MPAs are interspersed with fished areas that can benefit from contributions from several surrounding MPAs, and where larvae can survive to become adult fish that can then be caught (Hastings & Botsford 2003).

Between MPAs of a connected network, positive conservation and fisheries outcomes appear after 234 10 years of protection for full and strong protection levels. First, MPAs generate larval subsidies 235 that can offset the loss of biomass caused by fishing effort displacement and, subsequently, sustain 236 237 fisheries (Cowen 2006), representing the ecological and fisheries benefits of spillover, respectively 238 (Di Lorenzo et al. 2016). Fisheries benefits from MPA networks have been demonstrated before for full protection (Le Port et al. 2017). Barceló et al. (2021) estimated that benefits would appear 239 8-18 years after ecological benefits inside MPAs. Here, we show for the first time that those benefits 240 can occur, over a similar period, also with a strong protection level. In the unconnected network, 241 however, the overall gains in catch and biomass are lower than in a connected network. Larval 242 exchanges within the network can compensate for biomass offset between MPAs and benefit 243 fisheries in those areas. While previous empirical and modelling studies have shown how spillover 244 245 of adults and export of larvae from single MPAs can contribute to fisheries (Gell & Roberts 2003; Di Lorenzo et al. 2016, 2020; Le Port et al. 2017), fisheries benefits within fishable areas of networks 246 have never been studied specifically. When such areas have been included in models (Hastings & 247 Botsford 2003) or when studies have evaluated the magnitude of networks' fisheries benefits 248 (Fovargue et al. 2018), the impact of fishing displacement has been overlooked and thus the export 249 benefits might also have been overestimated. While Pelc et al. (2010) showed that larval export 250 from networked MPAs can be large enough to offset mortality due to displaced fishing effort, they 251 252 did not specifically focus on the dynamics of fish biomass in unprotected areas between networked 253 MPAs. Here, we show that alternating unprotected fished areas with MPAs could be a key solution to optimize the export of benefits from MPAs for the most restrictive protection levels. 254

We have shown that, in the case of overfishing, spatial management outperforms non-spatial management when comparing cases of similar overall reduction in fishing mortality, as also

evidenced elsewhere (Rassweiler et al. 2012; Carvalho et al. 2019). In the case of spatial 257 management, the reduction in fishing mortality is concentrated inside MPAs and thus locally higher 258 than in the non-spatial management scenario with the same overall reduction of fishing pressure 259 260 but distributed throughout the modelled area. Inside MPAs, even when partially protected, fish increase in size and produce more larvae as propagule production increases disproportionately to 261 the size of spawners (Marshall et al. 2019). These larval subsidies can thus support fisheries 262 outcomes by being exported from the protected areas towards the unprotected areas (Rassweiler 263 et al. 2012; Harrison et al. 2020), and contributing to the persistence of meta populations (Almany 264 et al. 2007; Saenz-Agudelo et al. 2011). In contrast, in non-spatial management scenarios, 265 reducing fishing mortality in each cell might not be sufficient to provide population-wide benefits in 266 an overfished situation. Comparing the overall effectiveness of different spatial management 267 scenarios, it appears that large MPAs and networks of MPAs can deliver similar conservation and 268 fisheries benefits when networks are connected through larval dispersal. MPAs that are not 269 270 connected by larval dispersal result in suboptimal, underperforming MPA networks (Rassweiler et al. 2012). 271

Our inferences are based on a number of assumptions. First, as detailed in Supporting Information, 272 273 we developed our model using the characteristics of a typical temperate demersal coastal species. However, our results should remain valid to a broad range of overfished species (Costello et al. 274 275 2012; Carvalho et al. 2019). Coastal areas are often places where fishing pressure is high and where space is a limiting factor, so where connected networks of fully protected areas such as 276 those developed here could prove most useful. Further developments of our model could aim to 277 278 better capture the behavior of pelagic species with large movements or to account for the population implications of time-at-risk when species cross MPA boundaries (Villegas-Ríos et al. 279 2021). Second, we did not consider density-dependent spillover. Thus, our model might 280 underestimate fisheries benefits and overestimate conservation benefits. Third, we used an 281 average larval connectivity matrix to represent larval export. Several studies have shown that larval 282 behavior produces spatial and temporal variability in connectivity patterns (Cowen & Sponaugle 283 2009; Bode et al. 2019). Nevertheless, the connectivity portfolio effect (Harrison et al. 2020) 284 suggests that MPA networks' emergent properties may provide overall stability in larval supply. 285 Finally, to reduce complexity, we only use one single size for each of the six MPAs that are part of 286 a network, and one single size (six times larger) for the large MPA. Including size as a continuous 287 288 variable could help better inform MPA network planning (Fovargue et al. 2018). Future

developments of our model could include, among other aspects, multi-species interactions, influence of habitat heterogeneity, or fishing effort increase within MPAs boundaries (Magris 2021).

291 In a world of increasing tension between conservation and resource use, there is a need to identify and improve sustainable management scenarios with multiple social and ecological outcomes. 292 Currently, the two main global strategies for implementing marine conservation consist of 293 establishing few very large (often remote) fully or partially protected areas (Lubchenco & Grorud-294 Colvert 2015) or many smaller partially protected areas with often insufficient protection levels to 295 deliver conservation benefits (Claudet et al. 2020). Our results show that a non-mutually exclusive 296 third path is possible in areas where fisheries displacement costs are high. Networks of connected 297 fully protected areas can reduce displacement costs while still delivering positive conservation and 298 fisheries outcomes. Increases in catch are preceded by a short-term decrease that calls for the 299 identification of mechanisms to compensate for those short-term losses. Our findings provide novel 300 301 evidence that can support decision making in designing network of MPAs that reconcile conservation and fisheries goals. 302

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311 **References**

- 312
- Almany, G.R., Berumen, M.L., Thorrold, S.R., Planes, S. & Jones, G.P. (2007). Local replenishment of coral reef fish populations in a marine reserve. *Science (80-.).*, 316, 742–744.
- Barceló, C., White, J.W., Botsford, L.W. & Hastings, A. (2021). Projecting the timescale of initial
 increase in fishery yield after implementation of marine protected areas. *ICES J. Mar. Sci.*,
 78, 1860–1871.
- Belharet, M., Di Franco, A., Calò, A., Mari, L., Claudet, J., Casagrandi, R., Gatto, M., Lloret, J.,
 Sève, C., Guidetti, P. & Melià, P. (2020). Extending full protection inside existing marine

- protected areas, or reducing fishing effort outside, can reconcile conservation and fisheries
 goals. J. Appl. Ecol., 1365-2664.13688.
- Bode, M., Leis, J.M., Mason, L.B., Williamson, D.H., Harrison, H.B., Choukroun, S. & Jones, G.P.
 (2019). Successful validation of a larval dispersal model using genetic parentage data. *PLOS Biol.*, 17, e3000380.
- 325 Carvalho, P.G., Jupiter, S.D., Januchowski-Hartley, F.A., Goetze, J., Claudet, J., Weeks, R.,
- Humphries, A. & White, C. (2019). Optimized fishing through periodically harvested closures. *J. Appl. Ecol.*, 1365-2664.13417.
- Claudet, J., Loiseau, C., Sostres, M. & Zupan, M. (2020). Underprotected Marine Protected Areas
 in a Global Biodiversity Hotspot. *One Earth*, 2, 380–384.
- Costello, C., Ovando, D., Hilborn, R., Gaines, S.D., Deschenes, O. & Lester, S.E. (2012). Status and
 Solutions for the World's Unassessed Fisheries. *Science (80-.).*, 338, 517–520.
- Cowen, R.K. (2006). Scaling of Connectivity in Marine Populations. *Science (80-.).*, 311, 522–527.
- Cowen, R.K. & Sponaugle, S. (2009). Larval Dispersal and Marine Population Connectivity. *Ann. Rev. Mar. Sci.*, 1, 443–466.
- 335 Devillers, R., Pressey, R.L., Grech, A., Kittinger, J.N., Edgar, G.J., Ward, T. & Watson, R. (2015).
- Reinventing residual reserves in the sea: Are we favouring ease of establishment over need for protection? *Aquat. Conserv. Mar. Freshw. Ecosyst.*, 25, 480–504.
- Díaz, S., Settele, J., Brondízio, E., Ngo, H., Guèze, M., Agard, J., Arneth, A., Balvanera, P.,
- Brauman, K., Butchart, S., Chan, K., Garibaldi, L., Ichii, K., Liu, J., Subrmanian, S., Midgley,
- G., Miloslavich, P., Molnár, Z., Obura, D., Pfaff, A., Polasky, C. (2019). Summary for
- 341 policymakers of the global assessment report on biodiversity and ecosystem services of the
- 342 Intergovernmental Science-Policy Platform on Biodiversity and Ecosystem Services.
- Edgar, G.J., Stuart-Smith, R.D., Willis, T.J., Kininmonth, S., Baker, S.C., Banks, S., Barrett, N.S.,
 Becerro, M.A., Bernard, A.T.F., Berkhout, J., Buxton, C.D., Campbell, S.J., Cooper, A.T.,

345	Davey, M., Edgar, S.C., Försterra, G., Galván, D.E., Irigoyen, A.J., Kushner, D.J., Moura, R.,
346	Parnell, P.E., Shears, N.T., Soler, G., Strain, E.M.A. & Thomson, R.J. (2014). Global
347	conservation outcomes depend on marine protected areas with five key features. Nature,
348	506, 216–220.
349	Ferreira, H.M., Magris, R.A., Floeter, S.R. & Ferreira, C.E.L. (2022). Drivers of ecological
350	effectiveness of marine protected areas: A meta-analytic approach from the Southwestern
351	Atlantic Ocean (Brazil). J. Environ. Manage., 301, 113889.
352	Fovargue, R., Bode, M. & Armsworth, P.R. (2018). Size and spacing rules can balance
353	conservation and fishery management objectives for marine protected areas. J. Appl. Ecol.,
354	55, 1050–1059.
355	Di Franco, A., Plass-Johnson, J.G., Di Lorenzo, M., Meola, B., Claudet, J., Gaines, S.D., García-
356	Charton, J.A., Giakoumi, S., Grorud-Colvert, K., Hackradt, C.W., Micheli, F. & Guidetti, P.
357	(2018). Linking home ranges to protected area size: The case study of the Mediterranean
358	Sea. Biol. Conserv., 221, 175–181.
359	Gaines, S.D., White, C., Carr, M.H. & Palumbi, S.R. (2010). Designing marine reserve networks for
360	both conservation and fisheries management. Proc. Natl. Acad. Sci., 107, 18286–18293.
361	Gell, F.R. & Roberts, C.M. (2003). Benefits beyond boundaries: the fishery effects of marine
362	reserves. Trends Ecol. Evol., 18, 448–455.
363	Giakoumi, S., Scianna, C., Plass-Johnson, J., Micheli, F., Grorud-Colvert, K., Thiriet, P., Claudet, J.,
364	Di Carlo, G., Di Franco, A., Gaines, S.D., García-Charton, J.A., Lubchenco, J., Reimer, J., Sala,
365	E. & Guidetti, P. (2017). Ecological effects of full and partial protection in the crowded
366	Mediterranean Sea: a regional meta-analysis. Sci. Rep., 7, 8940.
367	Grorud-Colvert, K., Claudet, J., Tissot, B.N., Caselle, J.E., Carr, M.H., Day, J.C., Friedlander, A.M.,
368	Lester, S.E., de Loma, T.L., Malone, D. & Walsh, W.J. (2014). Marine protected area
369	networks: Assessing whether the whole is greater than the sum of its parts. <i>PLoS One</i> , 9, 1–
370	7.

371	Grorud-Colvert, K., Sullivan-Stack, J., Roberts, C., Constant, V., Horta E Costa, B., Pike, E.P.,
372	Kingston, N., Laffoley, D., Sala, E., Claudet, J., Friedlander, A.M., Gill, D.A., Lester, S.E., Day,
373	J.C., Gonçalves, E.J., Ahmadia, G.N., Rand, M., Villagomez, A., Ban, N.C., Gurney, G.G.,
374	Spalding, A.K., Bennett, N.J., Briggs, J., Morgan, L.E., Moffitt, R., Deguignet, M., Pikitch, E.K.,
375	Darling, E.S., Jessen, S., Hameed, S.O., Di Carlo, G., Guidetti, P., Harris, J.M., Torre, J.,
376	Kizilkaya, Z., Agardy, T., Cury, P., Shah, N.J., Sack, K., Cao, L., Fernandez, M. & Lubchenco, J.
377	(2021). The MPA Guide: A framework to achieve global goals for the ocean. Science (80).,
378	373.
379	Hall, A.E., Cameron, D.S. & Kingsford, M.J. (2021). Partially protected areas as a management
380	tool on inshore reefs. <i>Rev. Fish Biol. Fish.</i> , 31, 631–651.
381	Halpern, B.S., Gaines, S.D. & Warner, R.R. (2004). Cofounding effects of the export of production
382	and the displacement of fishing effort from marine reserves. <i>Ecol. Appl.</i> , 14, 1248–1256.
383	Harrison, H.B., Bode, M., Williamson, D.H., Berumen, M.L. & Jones, G.P. (2020). A connectivity
384	portfolio effect stabilizes marine reserve performance. Proc. Natl. Acad. Sci., 201920580.
385	Harrison, H.B., Williamson, D.H., Evans, R.D., Almany, G.R., Thorrold, S.R., Russ, G.R., Feldheim,
386	K.A., Van Herwerden, L., Planes, S., Srinivasan, M., Berumen, M.L. & Jones, G.P. (2012).
387	Larval export from marine reserves and the recruitment benefit for fish and fisheries. Curr.
388	<i>Biol.</i> , 22, 1023–1028.
389	Hastings, A. & Botsford, L.W. (2003). Comparing designs of marine reserves for fisheries and for
390	biodiversity. <i>Ecol. Appl.</i> , 13, 65–70.
391	Hedges, L. V, Gurevitch, J. & Curtis, P.S. (1999). The meta-analysis of response ratio in
392	experimental ecology. <i>Ecology</i> , 80, 1150–1156.
393	Horta e Costa, B., Claudet, J., Franco, G., Erzini, K., Caro, A. & Gonçalves, E.J. (2016). A
394	regulation-based classification system for Marine Protected Areas (MPAs). Mar. Policy, 72,
395	192–198.
396	Di Lorenzo, M., Claudet, J. & Guidetti, P. (2016). Spillover from marine protected areas to

397	adjacent fisheries has an ecological and a fishery component. J. Nat. Conserv., 32, 62–66.
398	Di Lorenzo, M., Guidetti, P., Di Franco, A., Calò, A. & Claudet, J. (2020). Assessing spillover from
399	marine protected areas and its drivers: A meta-analytical approach. Fish Fish., 21,
400	faf.12469.
401	Lubchenco, J. & Grorud-Colvert, K. (2015). Making waves: The science and politics of ocean
402	protection. <i>Science (80).,</i> 350, 382–383.
403	MacNeil, M.A., Graham, N.A.J., Cinner, J.E., Wilson, S.K., Williams, I.D., Maina, J., Newman, S.,
404	Friedlander, A.M., Jupiter, S., Polunin, N.V.C. & McClanahan, T.R. (2015). Recovery
405	potential of the world's coral reef fishes. <i>Nature</i> , 520, 341–344.
406	Magris, R.A. (2021). Effectiveness of Large-Scale Marine Protected Areas in the Atlantic Ocean
407	for Reducing Fishing Activities. Front. Mar. Sci., 8, 1149.
408	Marshall, D.J., Gaines, S., Warner, R., Barneche, D.R. & Bode, M. (2019). Underestimating the
409	benefits of marine protected areas for the replenishment of fished populations. Front.
410	Ecol. Environ., 17, 407–413.
411	Melià, P., Casagrandi, R., Franco, A. Di, Guidetti, P., Gatto, M., Di Franco, A., Guidetti, P. & Gatto,
412	M. (2020). Protection reveals density-dependent dynamics in fish populations: A case study
413	in the central Mediterranean. PLoS One, 15, e0228604.
414	O'Hara, C.C., Frazier, M. & Halpern, B.S. (2021). At-risk marine biodiversity faces extensive,
415	expanding, and intensifying human impacts. Science (80)., 372, 84–87.
416	Pelc, R.A., Warner, R.R., Gaines, S.D. & Paris, C.B. (2010). Detecting larval export from marine
417	reserves. Proc. Natl. Acad. Sci., 107, 18266–18271.
418	Pike, E.P., MacCarthy, J.M.C., Hameed, S.O., Harasta, N., Grorud-Colverd, K., Sullivan-Stack, J.,
419	Claudet, J., Horta e Costa, B., Gonçalves, E.J., Villagomez, A., Morgan, L. (in press). Ocean
420	protection quality is lagging behind quantity: Applying a scientific framework to assess real
421	MPA progress against the 30 by 30 target. Cons. Lett.

422	Le Port, A., Montgomery, J.C., Smith, A.N.H., Croucher, A.E., McLeod, I.M. & Lavery, S.D. (2017).
423	Temperate marine protected area provides recruitment subsidies to local fisheries. Proc. R.
424	Soc. B Biol. Sci., 284, 20171300.
425	Rassweiler, A., Costello, C. & Siegel, D.A. (2012). Marine protected areas and the value of
426	spatially optimized fishery management. Proc. Natl. Acad. Sci., 109, 11884–11889.
427	Reimer, J.M., Devillers, R. & Claudet, J. (2020). Benefits and gaps in area-based management
428	tools for the ocean Sustainable Development Goal. Nat. Sustain., 20.
429	Roberts, C.M. (2001). Effects of Marine Reserves on Adjacent Fisheries. Science (80)., 294,
430	1920–1923.
431	Saenz-Agudelo, P., Jones, G.P., Thorrold, S.R. & Planes, S. (2011). Connectivity dominates larval
432	replenishment in a coastal reef fish metapopulation. Proc. R. Soc. B Biol. Sci., 278, 2954–
433	2961.
434	Turnbull, J.W., Johnston, E.L. & Clark, G.F. (2021). Evaluating the social and ecological
435	effectiveness of partially protected marine areas. Conserv. Biol., 0, 1–12.
436	Villegas-Ríos, D., Claudet, J., Freitas, C., Moland, E., Thorbjørnsen, S.H., Alonso-Fernández, A. &
437	Olsen, E.M. (2021). Time at risk: Individual spatial behaviour drives effectiveness of marine
438	protected areas and fitness. Biol. Conserv., 263, 109333.
439	Worm, B., Hilborn, R., Baum, J.K., Branch, T.A., Collie, J.S., Costello, C., Fogarty, M.J., Fulton, E.A.,
440	Hutchings, J.A., Jennings, S., Jensen, O.P., Lotze, H.K., Mace, P.M., McClanahan, T.R., Minto,
441	C., Palumbi, S.R., Parma, A.M., Ricard, D., Rosenberg, A.A., Watson, R. & Zeller, D. (2009).
442	Rebuilding Global Fisheries. Science (80)., 325, 578–585.
443	Ziegler, S.L., Brooks, R.O., Hamilton, S.L., Ruttenberg, B.I., Chiu, J.A., Fields, R.T., Waltz, G.T.,
444	Shen, C., Wendt, D.E. & Starr, R.M. (2022). External fishing effort regulates positive effects
445	of no-take marine protected areas. Biol. Conserv., 269, 109546.
446	Zupan, M., Fragkopoulou, E., Claudet, J., Erzini, K., Horta e Costa, B. & Gonçalves, E.J. (2018).

- 447 Marine partially protected areas: drivers of ecological effectiveness. Front. Ecol. Environ.,
- 448 16, 381–387.
- 449
- 450

Figures and Tables



Non-spatial and spatial management options

Figure 1. Schematic configurations of the non-spatial (catch limit) and spatial (large MPA and network of MPAs) management scenarios, and their associated control (without protection). Redistribution of fishing mortality once an MPA is established induces higher fishing mortality in cells adjacent to the MPA. Networks of MPAs have two focal areas: the marine protected areas, which can be fully or partially protected, and the unprotected areas in between the MPAs. In the spatial management options, MPAs of 4 protection levels are assessed: Full protection (100% of fishing mortality is removed from the protected cell and redistributed in adjacent cells); three levels of partial protection (75%, 50%, 25% of fishing mortality, respectively, redistributed in adjacent cells). In the non-spatial management scenarios, four levels of catch limit are assessed by applying the same reduction of overall fishing mortality as in the spatial management options across all cells.

Gains and losses in fish biomass of non-spatial and spatial management options, in comparison to business as usual



Figure 2. Simulated dynamics of effect sizes (log-ratio of the biomass in test simulation compared to their respective control simulation) in the whole region (first column), in each focal area (inside large MPA or inside networked MPAs, between networked MPAs, and outside large MPA or MPA network), for each set of simulations (non-spatial management, large MPA, connected MPA network, unconnected MPA network).

Gains and losses in fish catch of non-spatial and spatial management options, in comparison to business as usual



Figure 3. Simulated dynamics of effect sizes (log-ratio of the catch in test simulation compared to their respective control simulation) in the whole region (first column), in each focal area (inside large MPA or inside networked MPAs, between networked MPAs, and outside large MPA or MPA network), for each set of simulations (non-spatial management, large MPA, connected MPA network, unconnected MPA network).

Table 1. Description of the simulations used to evaluate the effectiveness of non-spatial
 management scenarios (T0) and spatial management scenarios (T1, T2, T3) at delivering
 conservation and fisheries outcomes, compared to unmanaged control scenarios (C1, C2).

Simulations	Details about the simulations
Control, connected (C1)	No MPA
Control, unconnected (C2)	No MPA Larval connectivity disactivated within the network
Catch limit (T0)	No MPA Reduction of the same proportion of fishing mortality for each cell in the model
Connected network (T1)	Activation of the MPA network the 5th year Larval connectivity activated
Large MPA (T2)	Activation of the large MPA the 5th year The large MPA has a surface equivalent to the total area covered by all the MPAs considered in scenario T1
Unconnected network (T3)	Activation of the MPA network the 5th year Larval connectivity disactivated within the network