

Disentangling plant- and environment-mediated drivers of active rhizosphere bacterial community dynamics during short-term drought

Sreejata Bandopadhyay, Xingxing Li, Alan Bowsher, Robert Last, Ashley Shade

▶ To cite this version:

Sreejata Bandopadhyay, Xingxing Li, Alan Bowsher, Robert Last, Ashley Shade. Disentangling plantand environment-mediated drivers of active rhizosphere bacterial community dynamics during shortterm drought. 2024. hal-04277894

> HAL Id: hal-04277894 https://hal.science/hal-04277894

> > Preprint submitted on 22 Jul 2024

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers. L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.



Disentangling plant- and environment-mediated drivers of active rhizosphere bacterial community dynamics during short-term drought Sreejata Bandopadhyay^{1,2,3}, Xingxing Li^{3,4}, Alan W. Bowsher^{1,2}, Robert L. Last^{3,4,5}, Ashley Shade^{6*} ¹Department of Microbiology and Molecular Genetics, Michigan State University, MI, USA ²Plant Resilience Institute, Michigan State University, MI, USA ³DOE Great Lakes Bioenergy Research Center, Michigan State University, MI, USA ⁴Department of Biochemistry and Molecular Biology, Michigan State University, MI, USA ⁵Department of Plant Biology, Michigan State University, MI, USA ⁶ Université de Lyon, France; CNRS, INRAe, Ecole Nationale Véterinaire de Lyon and Université Lyon 1, UMR 5557 Écologie Microbienne, 16 rue Dubois 69622 Villeurbanne Cedex, FRANCE *Corresponding author

Abstract Background. Mitigating the effects of climate stress on crops is important for global food security. The microbiome associated with plant roots, henceforth, the rhizobiome, can harbor beneficial microbes that alleviate stress impacts. However, the factors influencing the recruitment of the rhizobiome during stress are unclear. We conducted an experiment to understand bacterial rhizobiome responses to short-term drought for two crop species: switchgrass and common bean. We used 16S rRNA and 16S rRNA gene sequencing to investigate the impact of drought severity on the recruitment of active bacterial rhizobiome members. We included planted and unplanted conditions to distinguish the environment- versus plantmediated drivers of the active rhizobiome. Results. Though each crop had a distinct rhizobiome, there were differences in the active microbiome structure between drought and watered and between planted and unplanted treatments. Despite their different community structures, the drought rhizobiome dynamics were similar across the two crops. However, the presence of a plant more strongly explained the rhizobiome variation in bean (17%) than in switchgrass (3%), with a small effect of plant mediation during drought only observed for the bean rhizobiome. The switchgrass rhizobiome was stable despite differences in the rhizosphere metabolite profiles between planted and unplanted treatments. Specifically, steroidal saponins and diterpennoids were enriched in drought, planted switchgrass soils. Conclusions. We conclude that rhizobiome benefits to resist short-term drought are crop-specific, with the possibility of decoupling of plant exudation and rhizobiome responses, as we observed in switchgrass. We propose bacterial taxa uniquely associated with common bean plants during the short-term drought, which could be further evaluated to determine any plant benefit during drought. **Keywords** Rhizosphere microbiome, rhizosphere metabolome, drought stress, indicator species, common bean, switchgrass, 16sRNA:16SrRNA gene ratios

Background

75

76

77

78

79

80

81

82

83

84

85

86 87

88

89

90

91

92

93

94

95

96

97

98

99 100

101

102

103

104

105

106

107

108

The idea that beneficial microbes can be managed to improve crop performance is building momentum with new research directions [1-4]. It is well known that plant-associated microorganisms can promote plant performance when conditions are unfavorable [5]. For example, microbes can help plants acquire limiting nutrients such as nitrogen and phosphorus [7]. In turn, plants can shape their microbiome by secreting metabolites including sugars in their root exudates to support microbiome recruitment and maintenance [6]. Furthermore, different plant genotypes can have distinct metabolite compositions (e.g. [8]), with potential consequences for microbiome recruitment (e.g., [9]). In the bioenergy feedstock switchgrass (*Panicum virgatum*), there were different metabolite profiles for upland and lowland ecotypes [8], with in-vitro studies showing that differential metabolite accumulation across ecotypes can contribute to different microbiome compositions [9]. The increased frequency and intensity of droughts have become major challenges for crop production globally [10, 11]. During drought, soil loses its moisture content, which can be exacerbated by erratic rainfall and temperature fluctuations [12]. There also are changes in other soil and plant properties, such as root exudation [13], with concurrent changes in the composition and function of the local soil and rootassociated microbiome. Microorganisms can also play direct or indirect roles in plant drought tolerance. For example, some plant growth-promoting bacteria (PGPB) exude exopolysaccharides which can retain the soil moisture content at the beginning of drought [14]. They can also produce antioxidant enzymes during a drought that can support plant response to reactive oxygen species that results from drought related stress [15]. However, beneficial and detrimental strains that affect plant growth under drought conditions have been identified (for example, in sorghum [16]). Drought can increase the relative abundance of phyla such as Firmicutes, Chloroflexi, and Actinobacteria, while phyla such as Bacteroidetes and Planctomycetes are relatively depleted [13]. Drought responses of the rhizosphere and endosphere microbiomes have been extensively studied in crops such as rice [17, 18] and sorghum [19]. These studies have shown that drought can cause taxonomic enrichment of bacteria that may help plants recover during stress while also affecting the temporal assembly of the root microbiome. A recent study used random forest models to show that different drought regimes altered the microbiome succession as compared to the watered controls [17]. In this study, the degree of difference in the microbiome was directly proportional to the duration of drought, and the delay in microbiome succession persisted even after watering resumed, with enduring changes in the root microbiome observed as many as 62-84 days after the drought ended [17].

110

111

112

113

114

115

116

117

118

119

120121

122

123

124

125

126127

128

129

130

131

132

133

134

135

136

137

138

139

140141

142

Because of the dry soil conditions and typically low microbial biomass associated with them, it remains technically challenging to assess the active microbial members of the rhizosphere during drought, and few studies have done so [19]. As the dormant bacterial pool in soils and rhizospheres is substantial [20-22], separating the responses of active from dormant members may be insightful for targeting the most responsive populations for plant benefit. Additionally, different plants have different inherent tolerances to dry conditions. Comparing the rhizosphere microbiomes across differently drought-sensitive crops may also provide insights into the members that especially can benefit sensitive plants. Furthermore, while changes in plant root exudation during drought can shape the rhizosphere microbiome response [23-25], it is unclear how much of that response is mediated by the host or attributable to the environmental conditions of the drought that impact the microbes directly. Drought conditions include low moisture availability, changes in physical soil structure, and connectivity via soil pores, all of which can directly impact microbes. Understanding whether responses are host- or environment-mediated could inform the separate targets for microbiome modification via host and soil management. Plant changes in rhizosphere exudates during drought stress could serve as signals to reactivate and recruit from the dormant pool, though this has not been thoroughly considered. Finally, host-mediated microbiome engineering has been employed in crops, such as wheat, to select beneficial microbial communities that promote plant tolerance to drought stress [26]. In rhizosphere soils or similar environments containing ample amounts of energy-rich substrates, it is thought that much of the microbial community is active rather than dormant [27, 28]. Evidence from ¹³C-PLFA (phospholipid-derived fatty acids) shows that Gram-positive bacteria assimilated ¹³C more actively in the rhizosphere than in bulk soil [29]. Studies have also shown that Gram-negative bacteria in the rhizosphere actively assimilated root-derived carbon more successfully than Gram-positive bacteria [30-33]. For example, activity staining of bacterial cells from rhizosphere soil revealed that as much as 55% of them were active, with further evidence suggesting that rhizosphere soils have approximately 20% more active cells than bulk soils [28, 34]. However, the active rhizosphere microbiome pool can change during stress events, with members shifting to dormancy in response to the changing environment or climate, for instance, low moisture availability. The contribution of the resuscitated or active community becomes more prominent during and in the immediate aftermath of stress, when environmental conditions fluctuate, and prevalent members shift in competitiveness. Thus, it is critical to understand active community dynamics over short-term stress exposure. Several knowledge gaps exist about microbiome activation and recruitment during drought. First, the factors affecting the recruitment of the active rhizobiome during drought are unclear; specifically, the role

144

145

146

147

148

149150

151

152

153

154

155

156

157

158

159

160

161

162

163164165

166167

168169

170

171

172

173

174

175

176

factors tested (Figure 1).

of the environment and plant as drivers in that assembly. Second, there is limited knowledge as to the effect of the duration of drought on rhizobiome assembly and how that depends on or interacts with drought conditions. Finally, how these microbiome dynamics differ across different plant families with different drought tolerances and to what extent the microbiome response is attributable to the host or environment needs to be understood. To address these knowledge gaps, we conducted a greenhouse experiment using the annual legume common bean (Phaseolus vulgaris, var. Red Hawk) and the perennial grass switchgrass (Panicum virgatum, var. Cave-in-Rock). These crops substantially differ in physiology and root architecture, as bean has a taproot system, and switchgrass has a dense rhizome that extends several meters belowground. Furthermore, switchgrass genotypes are relatively more consistently drought-resistant, whereas bean genotypes have more variation in their drought tolerance [35-37]. The objectives of this experiment were to understand the active bacterial rhizobiome immediate assembly and short-term recruitment over a gradient of drought severity. For both plants, we compared planted to unplanted conditions using agricultural field soils previously planted with one of the two crops (and thus expected to have a legacy from their rhizobiomes) and partitioned the influences of plant-mediated versus environment-mediated rhizobiome responses to drought. We addressed two hypotheses. First, we hypothesized that active rhizobiomes respond to drought and change progressively with drought severity. Second, we hypothesized that the host plant mediates the responses of the active rhizobiome to drought via compositional changes in root metabolites. Methods Experimental design We performed a greenhouse experiment with common bean (Phaseolus vulgaris var. Red Hawk) and switchgrass (Panicum virgatum var. Cave-in-Rock). The experimental design consisted of two crop levels (bean, switchgrass), two planting levels (planted, unplanted), two moisture levels (drought, watered), and five destructive plant and rhizosphere sampling points correlating with increased drought severity. The drought gradient included a pre-drought baseline that was watered, and then destructive sampling over the next six days, with the day 2 samples experiencing the lowest drought severity and day 6 samples experiencing the highest. Five replicates were collected at each time point for each combination of the

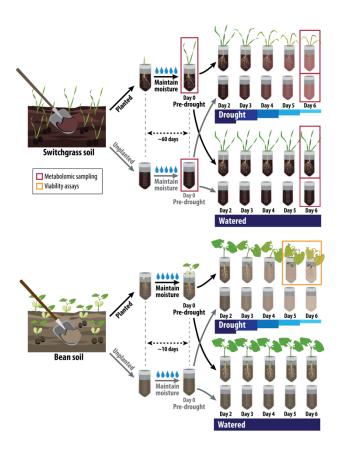


Figure 1. The experimental design used in the greenhouse study. Day 2 to day 6 correspond to increased drought severity.

Soil sample collection

Bean soil was collected from Montcalm Research Center on September 6, 2018, at Stanton, MI (43.350885, -85.177044), and was most recently planted with common bean (*Phaseolus vulgaris* var. Red Hawk) that year. Switchgrass samples were collected from the Great Lakes Bioenergy Research Center switchgrass plots at Lux Arbor Reserve (42.475224, -85.444979) in Denton, MI, on August 27, 2018. This site has been under continuous switchgrass (*Panicum virgatum* var. Cave-in-rock) cultivation since 2011. Switchgrass soil was collected and homogenized from six sites randomly situated throughout the plots. All sampling materials were sterilized with ethanol before use and between samples. A shovel was used to collect bean and switchgrass soils to a depth of 10 cm. Large rocks and root fragments were removed from the soil. Within three hours (for bean soils) and five hours (for switchgrass soils), soils were transported to the lab and stored at 4°C until sieving. Soil nutrient analysis was completed for the field soils at the Michigan State University Soil and Plant Nutrient Laboratory (**Table S1**).

196

197

198 199

200201

202

203

204

205

206

207

208

209210

211

212

213

214

215

216

217

218219

220221

222

223

224

225

226

227

228

Soil samples were sieved (4mm mesh) over three to five days immediately after the sample collection day and stored at 4°C until needed for the experiment. The drought experiment commenced after eight months and six months of soil storage at 4°C, for bean and switchgrass, respectively. Seed germination Approximately 250 switchgrass seeds (Sharp Bros. Seed of Mo., Inc) were placed in 50 ml conical tubes. 10 ml of 5% bleach was added to the tubes. The tubes were placed on a shaker at 200 rpm for 20 min. The bleach/water mix was decanted, and 10 ml of deionized water was added. The tubes were agitated again at 200 rpm for 5 minutes, then decanted. The deionized water rinse was repeated two more times. Approximately 60 switchgrass seeds were placed onto a wet filter paper (90 mm diameter) per sterile Petri plate. Seeds were moistened with deionized water, plates were wrapped in parafilm, covered with foil to block light, and incubated at 32°C incubator for three days. After three days, 1 ml of deionized water was supplemented to each plate, which was re-wrapped and returned to 32°C for an additional day. Bean seeds were obtained from the Dry Bean Breeding and Genetics program at Michigan State University. Approximately 590 seeds were placed in a 1L Erlenmeyer flask, and 1L of sterilization solution was added (0.1% Tween-20, 10% bleach, and 90% DI water) to the seeds. The seeds were soaked for 15 min at room temperature and inverted every few minutes. The seeds were washed five times with deionized water, and 12 seeds were placed onto Petri dishes lined with moistened filter paper. Plates were wrapped in parafilm and incubated at room temperature in the dark. Over the next three days, 2 ml of Milli-q water was added to each plate. After four days, approximately 60% of the switchgrass and bean seeds germinated and had an emerged radicle, and these were advanced for use in the experiment. Growth tube preparation and planting Growth tubes were made by modifying 50 ml conical tubes to better control soil water content during the experiment. An ethanol-cleaned drill bit was used to drill a hole in the bottom of 50-ml conical tubes. A hole was also drilled in every cell of the tube racks used to hold the 50 ml tubes for switchgrass and bean plants. Small squares of Kimwipe® tissue papers (5cm * 5cm) were cut, each rolled, and one rolled sheet was used to plug the drain hole in each 50 ml conical tube (Figure S1). Autoclaved perlite was combined with sieved field soil (50:50 v/v) to fill the prepared tubes to the top. Soils were watered with 10 ml of deionized water, which reduced the total soil volume to approximately 35 ml. Tubes for the planted treatment received one germinated seed. Unplanted tubes were prepared the same but without a seed.

230

231

232

233

234

235236

237

238

239

240

241

242

243

244

245

246

247

248

249

250251

252253

254

255

256

257

258

259

260

261

262

Seeds were planted so that ~1 cm of the seedling was exposed above the soil and at a depth to ensure that the emerging root would not push the seedling out of the soil. After planting, additional soil: perlite mix was added to the tube to replace the volume lost after settling and supplemented with 3 ml of deionized water. Each tube was wrapped in foil to exclude light in the root zone. The greenhouse's daily temperature range was 21-32°C, with 14 hours of supplemental daytime lighting (400-watt high-pressure sodium lamp). Seedlings were provided sufficient water to achieve healthy growth before initiating the drought. Switchgrass tubes were watered with 10 ml of deionized water every other day or every two days for two months until the plants were large enough for the experiment, which was designated as when the roots had filled the tube. Though this was close to the flowering stage, no flowers were observed. Bean tubes were watered daily with 10 ml of deionized water for the first five days and then twice daily with 10 ml for four additional days. After ~9 days of watered conditions, approximately 55 bean plants were sorted by height as a proxy for total biomass, and similarly sized plants were selected to include in the experiment. At this vegetative stage, the first trifoliate leaf emerged and started to unfold in all bean plants, and roots were completely bound within the tube. Before the drought, the planted tubes of both switchgrass and bean had tube-bound root systems such that all their soil could be directly influenced by the plant and classified as the rhizosphere. Bean roots were tube-bound within a week, but switchgrass took about two months. The timing for the start of drought treatment was chosen based on the time needed for the roots to fill the tube so that the entire soil was plant-influenced rhizosphere. Thus, the growth stage at the time of drought commencement was different between the two crops. Drought treatment and destructive sampling The same experimental design and sampling strategy were applied to bean and switchgrass (Figure 1). Twenty-five seedlings and 25 unplanted tubes were assigned to either the sufficiently watered ("watered") or to the reduced water treatment ("drought"), totaling 100 samples. In addition, five planted and five unplanted tubes were collected before the drought ("pre-drought," day 0). These samples were used as a baseline control to compare to the subsequent drought samples. The drought was initiated on the same day the pre-drought samples were collected (see Tables S2 and S3 for the watering regimes for switchgrass and bean, respectively). The drought severity was increased over six days by adding progressively less water. Five replicate soil samples (individual tubes used for destructive sampling) were collected for each experimental condition from day 2 to day 6 for five post-drought sampling points.

To ensure that the drought bean plants remained viable (and that our assessment was of the rhizobiome response to live plants rather than to dead), an additional 36 plants were in parallel subjected to drought to determine their viability on the final days of the experiment. On day 5, when bean plants were browning and lacked turgor (day 5), half (18) of the extra plants were provided 15 ml of deionized water in the morning and the afternoon. All 18 plants were resuscitated, and we observed their leaves returned to full turgor by the next morning. Thus, the drought plants from day 5 were likely alive when sampled despite appearing well-desiccated. The same was repeated on day 6 with the remaining drought plants. Thirteen of the 18 plants resuscitated (leaves back to full turgor) within two days, affirming that most drought plants on day 6 were viable. As a relatively more drought-tolerant plant, switchgrass did not appear severely desiccated on days 5 and 6, and the resuscitation check was unnecessary to confirm its viability.

Soil, root, and shoot sample collection and gravimetric soil moisture

Materials and surfaces were sterilized and/or ethanol-cleaned before use and between samples. All tubes were destructively sampled. During each sampling, each tube's contents were dumped onto sterilized foil. The soil was brushed from the root system and homogenized with a spatula. Roots were manually removed from the soil, and 0.5 g of soil was collected into a 1.5 ml microcentrifuge tube for DNA/RNA co-extractions. An additional ~6 g soil was retained as a backup. Perlite (>2mm) was avoided. Soils were flash-frozen in liquid nitrogen and stored at -80°C until nucleic acid extractions. An additional 4 g of soil was collected to determine gravimetric soil moisture content. After recording the mass of fresh soil collected, soils were dried in an oven at 50°C for at least three days and then re-weighed. Percent gravimetric soil moisture was calculated as the percent water mass lost during drying using **Equation 1** [38].

Gravimetric moisture content (%) =
$$\frac{(Mass \ of \ soil_{wet} - Mass \ of \ soil_{dry})}{Mass \ of \ soil_{dry}} * 100$$
 Equation 1

After soil collection, the shoot system was cut at the base of the stem and put in pre-labeled envelopes.

The shoots were dried at 50°C for at least five days. Masses taken after drying were used to determine the biomass of the shoot. Treatment differences in shoot biomass and gravimetric soil moisture were analyzed with a three-way ANOVA type III test upon satisfaction of normality.

An additional 30 switchgrass plants (15 watered and 15 drought) were grown for metabolite analysis on day 6. Soils were homogenized, shoots were sampled as previously described, and root systems were

cleaned in deionized water. Then, samples were flash-frozen in liquid nitrogen in 15 ml tubes. Soils of the unplanted treatments (watered and drought) from day 6, stored at -80°C, were used to compare with the planted treatments. We also analyzed metabolites from pre-drought soils (four replicates of planted and five unplanted treatments) stored at -80°C. In addition, ten unplanted soils from day 6 (five drought and five watered) and nine pre-drought samples (four planted, five unplanted) were analyzed for metabolites. In summary, 15 switchgrass plants from the day 6 watered and drought treatments each yielded 30 samples for soil, shoot, and root metabolites.

Soil DNA/RNA co-extractions

296

297

298

299

300

301

302

303304

305 306

307

308

309

310

311

312

313

314

315

316

317

318

319

320

321

322

323

324

325

326

327

328 329

DNA/RNA co-extractions were performed using the protocol specified in [39] with minor modifications. Briefly, 0.5 g of flash-frozen soil was added to 0.7 mm PowerBead® garnet bead tubes (Qiagen, Germantown, MD, purchased early 2021 before being discontinued). 0.5 ml of CTAB-phosphate buffer (120 mM, pH 8) and 0.5 ml of phenol:chloroform: isoamyl alcohol solution (25:24:1, nuclease-free, Invitrogen®) were added to the tube and placed on a bead beater to beat for 30 s. Tubes were then centrifuged at 12000g for 10 min at 4°C. The top aqueous layer was extracted and placed into a new 1.5 ml microcentrifuge tube. Next, 0.5 ml of chloroform: isoamyl alcohol (24:1, nuclease-free, Sigma®) was added to the tube and inverted several times to form an emulsion to remove residual phenol. Tubes were centrifuged at 16000g for 5 min at 4°C, the top aqueous layer extracted and placed into a new 1.5 ml microcentrifuge tube. Nucleic acids were precipitated by adding two volumes of 30% polyethylene glycol solution (PEG6000, 1.6M NaCl) and mixing a few times. Tubes were incubated on ice for two hours. Tubes were then centrifuged at 16000g for 20 min at 4°C. The supernatant was removed, and 1 ml of 70% ice-cold ethanol was added. Tubes were centrifuged at 16000g for 15 min at 4°C. Ethanol wash was pipetted out, being careful not to remove the pellet, and placed back in the centrifuge for a final spin for 10 s to collect residual ethanol. The remaining ethanol was pipetted out of tubes, and the pellet was allowed to air dry. The nucleic acid pellet was resuspended in 30 µl of nuclease-free water. Negative controls for extractions included tubes to which no soil sample was added, and only included reagents and the garnet beads to check contamination in the extraction reagents. Two negative controls were processed alongside the experimental samples for each extraction day. All DNA and RNA samples were quantified using Qubit® dsDNA BR assay kit and RNA HS assay kit on a qubit 2.0 fluorometer (Invitrogen, Carlsbad, CA, USA). All nucleic acid raw coextracts (containing DNA and RNA) were visualized using agarose gel electrophoresis and validated with a band for DNA and RNA [39].

DNase treatment and cDNA synthesis

332

333

334

335

336

337

338

339

340

341

342343

344345

346

347

348

349

350

351

352

353

354

355

356

357

358

359

360361

362

The DNA/RNA coextract was used to prepare purified RNA using the Invitrogen TURBO® DNA-free Kit with minor modifications. 1 µl of 10X TURBO® DNase Buffer and 3 µl of TURBO® DNase enzyme were added to a 6 µl aliquot of the DNA/RNA coextract. The mixture was then incubated for 30 mins at 37°C, after which 2 µl of a DNAse inactivation reagent was added to each tube. The resulting solution was mixed by flicking the tubes by hand and then incubated at room temperature for 5 mins. The samples were centrifuged at 2000g for 5 min at room temperature. The purified, DNAse-treated RNA samples were transferred to a clean, sterile tube and immediately processed to make complementary DNA (cDNA). cDNA was prepared using the Invitrogen® SuperScript III First Strand Synthesis System using random hexamers per the manufacturer's instructions. cDNA samples were stored at -20°C until further processing. Negative controls were prepared for the cDNA synthesis step to check for reagent contamination. PCR and RT-PCR DNA (PCR) and cDNA samples (RT-PCR) were amplified using a standard protocol specified by the Research Technology Support Facility at Michigan State University. A 15 µl reaction volume was prepared for each sample to amplify the V4 hypervariable region of the 16S rRNA gene with 2X GoTaq Green Mastermix, primers 515F (5'- GTGCCAGCMGCCGCGGTAA- 3') and 806R (5'-GGACTACHVGGGTWTCTAAT-3') [40] and 1 µl template. The final concentration of the GoTaq Green Mastermix was 1X, and the final concentration of each primer was 0.1 µM. The PCR cycle was run using the following cycling parameters: initial denaturation at 94°C for 3 min; 30 cycles of denaturation at 94°C for 45 s, annealing at 50°C for 60 s, and extension at 72°C for 90 s, followed by a final extension at 72°C for 10 min. Samples were kept at 4°C and immediately visualized on a 1% agarose gel using 100bp ladder. All RNA samples purified from DNA/RNA coextracts and used to prepare cDNA were also used to run PCR and check on 1% agarose gel to ensure there were no contaminating DNA bands in RNA samples. All PCR reactions included a no-template negative control and an E. coli DNA template as a positive control. All negative controls used for extractions were also included as samples in the PCRs. Results indicated no contamination from extraction or PCR reagents and consistent amplification performance with the cycling parameters. Illumina sequencing of the 16S rRNA gene and 16S rRNA

364

365

366

367

368

369

370

371372

373

374

375

376

377

378

379

380

381

382 383

384 385

386

387

388

389

390

391

392

393

394

395

396

Amplicon sequencing was performed at the Genomics Research Technology Support Facility at Michigan State University. The V4 hypervariable region of the 16S rRNA gene was amplified using dualindexed Illumina compatible primers 515F and 806R as described in [41]. PCR products were batch normalized using an Invitrogen SequalPrep DNA Normalization Plate, and the normalized products recovered from each of the six plates submitted were pooled. The pools were cleaned up and concentrated using a QIAquick PCR Purification column followed by AMPureXP magnetic beads; it was quality controlled and quantified using a combination of Qubit dsDNA HS, Agilent 4200 TapeStation HS DNA1000, and Invitrogen Collibri Library Quantification qPCR assays. Each plate submitted for sequencing had approximately 85 samples (6 plates); each plate was used for one MiSeq run. The pools from each plate were each loaded onto an Illumina MiSeq v2 standard flow cell, and sequencing was performed in a 2x250bp paired-end format using a MiSeq v2 500 cycle reagent cartridge. Custom Sequencing and index primers were added to appropriate wells of the reagent cartridge. Base calling was done by Illumina Real Time Analysis (RTA) v1.18.54, and the output of RTA was demultiplexed and converted to FastQ format with Illumina Bcl2fastq v2.20.0. In addition to experimental samples, all negative controls from each extraction day were sequenced, and these samples were evenly distributed throughout all the MiSeq runs. Two positive controls (mock communities) were sequenced with each MiSeq run. One positive control was an in-house Mock community prepared in the Shade Lab [42], and the other positive control was provided by the RTSF at MSU for library preparation. Metabolite extraction from plant tissue and soil The switchgrass shoot, root, and rhizosphere soil samples collected from 15 individual plants in the drought experiment were pooled into seven replicates (with six replicates, each including two samples and one replicate including three samples (Figure S2A, B). The plant tissues were lyophilized and ground into powders using an automated tissue homogenizer (SPEX SamplePrep, Metuchen, NJ). The plant metabolites were extracted from the powders using 80% methanol containing 1 µM telmisartan internal standard and normalized by tissue weight to achieve equal concentration as described in [8]. The soil metabolites were extracted following the same protocol with differences in the soil-to-solvent ratio (1:2, v/v) and incubation method (mixing on a laboratory tube rocker). Extracts were centrifuged at 4000g for 20 min at room temperature to remove solids. The supernatant was completely dried using a SpeedVac vacuum concentrator (ThermoFisher, Waltham, MA) and reconstituted using 1/10 the original solvent volume.

398

399

400

401

402

403

404

405

406

407

408

409

410

411

412

413

414

415

416

417

418

419

420

421 422

423

424

425

426

427

428

429

Liquid chromatography—mass spectrometry (LC-MS) based untargeted metabolomics The chromatographic separation and MS analysis for the switchgrass metabolites were performed using a reversed-phase, UPLC BEH C18, 2.1 mm x 150 mm, 1.7 µm column (Waters, Milford, MA) and an Electrospray Ionization - Quadrupole Time-of-Flight MS (ESI-QToF-MS, Waters). Mass spectra, under positive ionization, were acquired in data-independent acquisition (DIA, MS^E) and data-dependent acquisition (DDA, MS/MS) modes. The details were described in [8]. The untargeted metabolomics (DIA) data processing, including retention time (RT) alignment, lock mass correction, peak detection, adduct grouping and deconvolution, and metabolite annotation, were done using the Progenesis QI software package (v3.0, Waters) following the protocol in [8]. The identified analytical signals were defined by the RT and mass-to-charge ratio (m/z) information and referred to as the *features*. Measurement of each feature across the sample panel was filtered by interquartile range, log-transformed, and scaled for multivariate analyses using R Studio v3.1.1 and R package MetabolAnalyze (scaling function set to type "pareto"). The log-transformed and pareto scaled (normalized) abundance of features for soil were used for principal component analysis (PCA). A PLS-DA model was generated using the MetaboAnalyst 4.0 online tool platform [43] to assess each metabolite feature's variable of importance (VIP) coefficient. The top 50 features with the highest VIP coefficient were used to visualize in a heatmap. For this, the features were log-transformed and scaled by row (features) from 0 to 1. This was completed using R software and then visualized using the heatmap function in the R package ComplexHeatmap. Distance and clustering methods were set to Euclidean and Ward.D to generate hierarchical clustering for the heatmap. The DDA (MS/MS) spectra were only used for metabolite mining using CANOPUS (class assignment and ontology prediction using mass spectrometry) [44] machine learning function built in the SIRIUS 4 (https://bio.informatik.uni-jena.de/sirius/), a computational tool for systematic compound class annotation. The identification level was denoted for each annotated metabolite based on the criteria for metabolite identification as per [45]. 16S rRNA and 16S rRNA gene sequence analysis Sequence data were analyzed using QIIME2 [46]. All paired-end sequences with quality scores were compressed and denoised using the DADA2 plugin [47]. The denoising step dereplicated sequences, filtered chimeras, and merged paired-end reads. The truncation parameters to use with the DADA2 plugin were determined using FIGARO [48]. FIGARO analyzes error rates in a directory of FASTQ files to determine the optimal trimming parameters for sequencing pipelines that utilize DADA2. The truncation length was set to 123 F and 162 R for all the data, with minimum overlap set to 30 base pairs, which resulted in 93% merging success. All truncation was performed from the 3' end for consistent final read

431

432

433

434

435

436

437

438

439

440441

442

443

444

445

446

447

448

449

450

451

452

453

454

455

456

457 458

459 460

461

462

lengths. The DNA and cDNA datasets were separately denoised. The resulting DNA and cDNA count tables were merged into a single QIIME2 artifact using the feature-table merge command. Similarly, the DNA and cDNA representative sequences were merged into a single QIIME2 artifact using the feature-table merge-segs command. The representative sequences from the combined count tables were clustered at 99% identity de-novo, and the clustered representative sequences were classified using SILVA v138 [49] to generate the taxonomy file. Ninety-nine percent sequence identity was used to define OTUs to conservatively account for any potential amplification errors that may have occurred during the cDNA synthesis from the RNA. The resulting OTU (operational taxonomic unit) table and taxonomy files were exported to R for ecological analysis. Designating the active community members All downstream analyses were performed in R version 4.0. The R package decontam [50] was used to determine the number and identity of contaminants in the dataset (Figures S3A and S3B) and remove them using the prevalence method. Contaminating taxa, mitochondria, and chloroplast sequences were filtered from the datasets. Based on rarefaction curves, a subsampling depth of 15,000 reads per sample was selected (Figure S3C). After subsampling, 16S rRNA to rRNA gene ratios (hereafter, 16S rRNA:rRNA gene) were computed from the DNA and cDNA datasets as described in [20]. While we compared a few methods therein for this dataset (please see Supplemental Information for details), we ultimately chose the method that applied a 16S rRNA:rRNA gene ratio threshold >=1. The chosen method was statistically robust in overarching patterns of beta-diversity given the exclusion or inclusion of phantom taxa (taxa with detected cDNA but not DNA counts). For phantom taxa that were detected in greater than 5% of samples, the DNA counts = 0 were changed to DNA==1 (as in "method 2" in [20]) (see Supplemental Materials). All other phantom taxa were excluded. The DNA OTU table was filtered to include only sequence counts of active taxa in the samples determined to meet our ratio threshold. Consequently, while every DNA and cDNA sample of sequence counts was initially rarefied to 15,000 reads, each sample's active community varied in their total reads (2000-6000). Relativized abundances were used for ecological statistics. Microbiome data analysis for active community The OTU table of the DNA counts of active taxa, taxonomy table, and metadata files was merged using the phyloseq package [51] in R. We used Bray-Curtis dissimilarity to determine beta diversity but also tested weighted UniFrac distance, and results were comparable. Permutational analysis of variance

464

465

466

467

468

469

470

471

472

473

474 475

476

477

478

479

480

481

482 483

484

485

486 487

488

489 490

491

492

493

494

495

496

(PERMANOVA) was conducted using the adonis function in vegan package [52] to assess differences in community structure by treatment and interactions: drought treatment (watered or drought), drought severity/sampling (days 2-6), and plant treatment (planted or unplanted). For post hoc tests, pairwise comparisons between drought levels were computed using the pairwise adonis 2 function in vegan package. PERMANOVA tests using the adonis function in R vegan package were also done on metabolite feature abundances used for multivariate analysis to understand the effects of planting, drought, and sampling day factors. To understand drought dynamics, we analyzed the Bray Curtis similarity of the microbiome to the pre-drought samples over the covariate of time for planted and unplanted treatments in bean and switchgrass. This was visualized using a smoothed conditional means (geom smooth function) with a linear model. This same approach was also used to assess general, relative fluctuations in rhizobiome size across samples, as proxied by DNA concentration. Richness was the observed number of OTUs, using the estimate richness function from 100 re-samples of the community. The normality of alpha diversity metrics was assessed using a Shapiro-Wilk test with a cut-off of W>0.9 for normality assumptions. Since data were normally distributed, a parametric three-way ANOVA test was used to assess the main effects and interaction effects between factors on richness. Contrasts were set to "sum" before running the ANOVA model to ensure that type III ANOVA tests were valid. We used non-parametric Kruskal-Wallis tests (Kruskal.test in R) to calculate differences in classlevel relative abundances between treatments. We used the indicspecies package in R to determine indicator species associated with experimental conditions. We used the abundance-based counterpart of Pearson's phi coefficient of association within the multipatt function [53]. To correct the phi coefficient for unequal group sizes the "func" parameter within multipatt was set to "r.g." P values were adjusted for false discovery rates. Indicator species were calculated at the OTU level for each treatment combination and the OTU and family level for the cropspecific indicators. We created a heatmap visualization of the 50 most abundant and active OTUs in the bean dataset for indicator species. We used a maximum standardization approach using the decostand function in the vegan package in R. We distinguished taxa that changed relative abundance over the drought gradient from taxa that changed in their detection. For samples without detection of an OTU's activity or DNA (e.g., DNA = 0, cDNA = 0), that OTU's abundance was coded as NA. For samples that had no detection of an OTU's activity but had detection of its DNA (e.g., DNA >0, cDNA = 0), that OTU's abundance was coded as 0. For samples that detected an OTU's activity and DNA (e.g., DNA >0, cDNA >0), that OTU's

498

499 500 501

502 503

504 505

506

507

508

509510511

512

513

514

515

516

517

518

519

520

521

522

523524

525

526

527

528

529

abundance was the value of its DNA sequence count. We also included a special consideration of "phantom" OTUs (e.g., DNA =0, cDNA>0), which were assigned a sequence count of 1. Results Overview We conducted a short-term drought experiment over 6 days and assessed the active rhizobiome dynamics of bean and switchgrass in the greenhouse. Rhizosphere soils were collected from fields recently planted with each crop, and treatments included planted/unplanted and watered/drought. We collected predrought and post-drought samples over time and with increased drought severity (Figure 1). Efficacy of the short-term drought treatment The watered samples in both bean and switchgrass had relatively high and stable gravimetric soil moisture content over the different time points for both planted and unplanted conditions (Figure 2A). In bean, the drought samples had a decrease in moisture content over time for both planted and unplanted soils, as expected. In switchgrass, drought samples had comparable moisture content over time for both planted and unplanted soils. The final sample (day 6) for switchgrass had the lowest mean soil moisture in planted soils and the highest mean moisture in unplanted soils. In both bean and switchgrass, the unplanted drought treatments retained higher soil moisture than the planted ones. There were interaction effects of drought, sampling day, and plant presence on bean soil moisture (Three-way ANOVA type III test, F=2.84, P=0.03, Table S4), suggesting the effects of each of these factors depended on the levels of the other factors. In switchgrass, there were main effects of drought on soil moisture (three-way ANOVA Type III test, F=28.43, P<0.001) and of plant presence on soil moisture (F=6.66, P=0.01). In bean plants, the shoot biomass was always higher in the watered samples as compared to drought (twoway ANOVA Type III test, F=21.18, P<0.001), and the watered plants increased in biomass over time (Figure 2B). There were no differences in switchgrass biomass between watered and drought samples over time (two-way ANOVA, F=0.30, P=0.59). Switchgrass shoot biomass fluctuated mildly over time but had no obvious trend (two-way ANOVA Type III test, F= 3.43, P=0.02, Table S4).

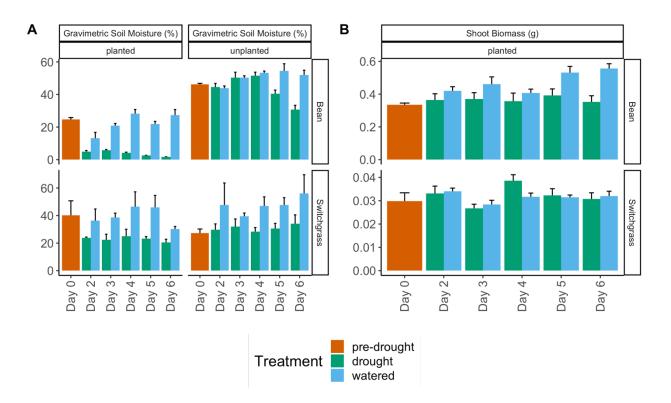


Figure 2: A) Gravimetric soil moisture of bean (sample size (n)=110) and switchgrass (sample size (n)=110) planted and unplanted soils at different time points for drought and watered samples. B) Shoot biomass of bean (sample size (n)=55) and switchgrass (sample size (n)=55) at different sampling days for drought and watered samples.

Metabolome responses of drought for switchgrass

Previous work has shown that switchgrass alters its metabolism during drought [54]. Given the absence of switchgrass shoot biomass response to the drought, we performed untargeted LC-MS-based metabolomics of switchgrass soils to assess any drought-induced changes in plant metabolites that could impact its microbiome. In addition, we also profiled the globalchanges that occurred in the metabolomes of switchgrass shoots and roots caused by the drought treatment. The metabolite profiling revealed 3,532 distinct metabolite features whose maximum abundance among the biological samples was \geq 500 counts (Supplemental Data 1-3).

A comparative analysis revealed 1051 root-, 538 shoot- and 231 soil-specific features (**Figure S4A**). The root and shoot shared 1551 features, the root and soil shared 883 features, and the leaf and soil shared 750 features, while 736 features were shared by all three groups (**Figure S4A**). The planted rhizosphere soils differed from the unplanted ones and the planted soil collected on day 0 (PERMANOVA, planted versus

550 551

552

553

554

555

556

557

558

559

560

561

562

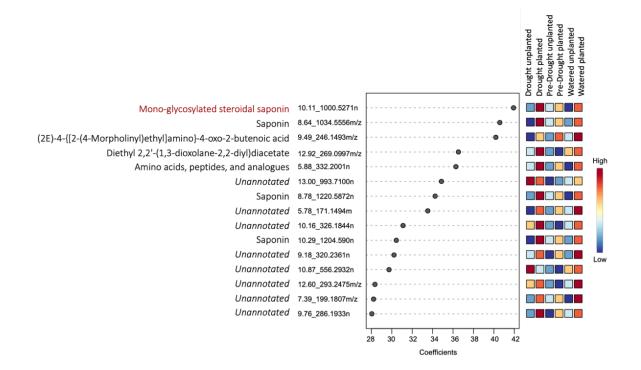
563

564

565

566

unplanted, F=11.26, R²=0.27, P=0.001, **Figure S4B**). There also were differences between day 0 and day 6 soils, indicating an overall effect of time (PERMANOVA F=3.55, R²=0.1, P=0.02). While the scatter plot suggested divergence between the drought and watered metabolomes for planted soils (Figure S4C), these overarching trends were not statistically supported (PERMANOVA F=2.2, R²=0.15, P=0.07). However, several metabolite features strongly differentiated the treatments. Saponins most differentiated the switchgrass metabolite soil profiles across the planted and drought samples (the metabolite annotation confidence levels can be found in Supplemental Data 4). Among the discriminating features, 10.11 1000.5271n was the most explanatory (highest variable of importance, VIP, Figure 3). This metabolite feature was previously identified as a mono-glycosylated steroidal saponin [8] (Figure S5). Some other most important features were also annotated as the specialized metabolites, including the root-accumulating diterpenoids with elevated concentrations in drought-treated planted soils (Figure S4D, Supplemental Data 4). This finding is consistent with the studies investigating longer drought exposure [8, 54]. The higher abundances of these specialized metabolites in the drought samples from the planted soils suggest that switchgrass could release them into the rhizosphere soil when stressed by drought. Additional metabolite results are provided in the Supplemental Materials.



568

569

570

571

572573

574

575

576

577

578579

580

581

582 583

584

585

586

587

588

589 590

591

592

593

594 595

596

597598

599

600

Figure 3. VIP coefficient score plot from a PLS-DA model showing the top 15 variables (features) of importance to differentiate the different switchgrass soil types. Feature 10.11 1000.5271n, a previously identified switchgrass saponin, had the highest coefficient and is thus the most important detected variable during the drought treatment. At the end of each feature name, 'm/z' stands for mass-to-charge ratio and 'n' stands for neutral mass. No detection of relative changes in rhizobiome size over time To proxy relative changes in microbiome size, we considered the recovered DNA concentration from the rhizosphere soil before dilution and normalization for sequencing. Linear models suggested that there were no supported trends in DNA concentrations over time for any of the treatments (Figure S6, all P > 0.05) Rhizobiome sequencing summary: total and active bacteria communities For the total community (DNA) data, total non-chimeric merged reads for both bean and switchgrass were 20,526,147. 19,938,535 reads remained after filtering out mitochondria, chloroplast, and contaminant OTUs, out of which 10,204,741 and 9,733,794 reads were attributed to switchgrass and bean, respectively. For the active community (cDNA) data, total non-chimeric merged reads for both bean and switchgrass were 22,018,663, of which 21,796,243 reads remained after filtering out mitochondria, chloroplast, and contaminant OTUs, out of which 11,799,699 and 9,996,544 reads were attributed to switchgrass and bean, respectively. The total number of taxa (DNA) and active taxa (based on DNA/RNA ratios) were 21,407 and 8,732, for bean samples and 17,331 and 7,539, respectively, for switchgrass samples. The microbiome data presented, henceforth, all pertain to the DNA counts of the active community members (16SrRNA:rRNA gene >= 1). The percent active taxa ranged from 40-60% of the total observed taxa in a sample. This establishes that an estimated half of the taxa detected in the DNA dataset were not observed or did not meet our activity criterion in the RNA dataset. Bean and switchgrass had distinct active rhizobiomes The literature shows that different plants harbor microbiomes [55-57]. Here, we also found different total (PERMANOVA Psuedo-F=163.02, R²=0.44, P=0.001) and active rhizobiome structures for bean and switchgrass (Figure S7, Figure 4A, active community PERMANOVA Pseudo-F= 91.21, P= 0.001, R²= 0.30), with 3883 active OTUs shared between the crops. However, the overall beta dispersion was comparable across bean and switchgrass (PERMDISP F=2.95, P=0.09).

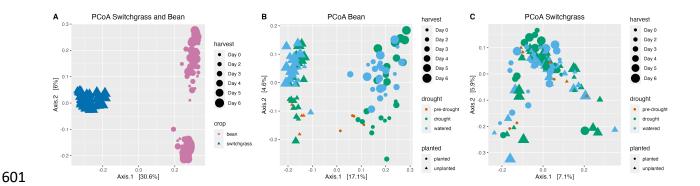


Figure 4: A) Principal Coordinate Analysis (PCoA) comparing bean and switchgrass active community composition (sample size (n)=212). PCoA of B) bean (sample size (n)=105) and C) switchgrass (sample size (n)=107) comparing active communities of planted and unplanted soil samples, as well as comparisons across drought treatments: pre-drought, drought, and watered. Samples collected across all sampling days are shown.

Addressing hypothesis: 1 Active rhizobiomes respond to drought and change progressively with drought severity.

Effects of drought, sampling day (colinear with drought severity for the drought treatment), and drought: sampling day interactions were significant and of similar explanatory value for both bean and switchgrass (**Figure 4B-C**, PERMANOVA all P < 0.05, see **Table S5**). We further investigated the effect of the drought treatments (levels: drought, pre-drought, and watered) using post hoc tests. In switchgrass, there were differences between pre-drought and watered treatments (PERMANOVA Pseudo-F = 1.36, R^2 = 0.02, P=0.03) and between the watered and drought condition (Pseudo-F= 1.55, R^2 =0.02, P=0.004). In bean, there were differences between the pre-drought and drought condition (Pseudo-F= 1.79,

For the drought treatment in bean samples, planted rhizobiomes had lower richness (number of active OTUs) than unplanted ones (Three-way ANOVA, F=60.12, P<0.001, **Table S6).** For switchgrass, there were no richness differences between the levels of different factors tested in the experiment, except for a weak interaction observed between planted and drought treatments (Three-way ANOVA F=4.30, P<0.05, **Table S6**).

We hypothesized that rhizobiomes in the drought samples would change faster and to a greater extent than the watered ones. We also hypothesized that the presence of the plant might suppress or stabilize

630

631

632

633

634

635636637

638

639

640

641

642

643

644 645

646

647

648

649

650 651

652

653

654

655

656

657

658

659

660

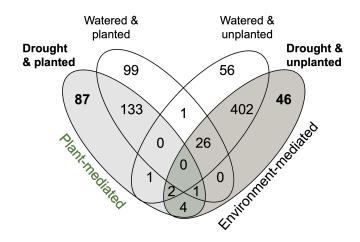
661

662

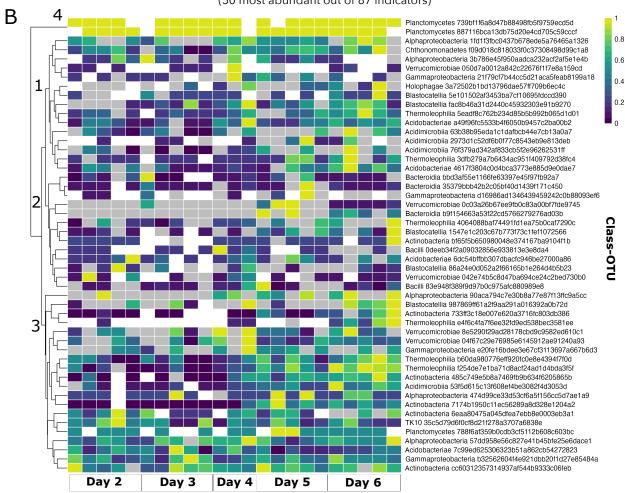
these dynamics due to host-microbiome feedback, resulting in reduced and slower change in planted than unplanted soils. To test this, we assessed linear models of the active rhizobiome structure (beta diversity) over time and as compared to the pre-drought condition (Figure S8). In these models, the slopes represent the rates of change, and the intercepts represent their magnitudes. However, the linear models generally were not supported for changes in beta diversity, suggesting only incremental beta diversity changes given the short-term drought. The exception was detecting significant linear regression for changes in beta diversity in the switchgrass watered and planted rhizobiome ($R^2 = 0.23$, P = 0.02). While there were no detected differences in beta-dispersion (community variance) among switchgrass experimental factors, for bean rhizobiomes, there were differences detected among the drought treatments, with post-hoc tests revealing differences in dispersion between the watered and drought rhizobiomes (mean dispersion(watered) = 0.41, mean dispersion (drought) = 0.44, PERMDISP Pseudo-F= 13.56, d.f.= 1, P= 0.001) and between the pre-drought and drought rhizobiomes (mean dispersion (predrought) = 0.40, mean dispersion (drought) = 0.44, PERMDISP Psuedo-F= 8,15, d.f.= 1, P= 0.01), but not between the pre-drought and watered (PERMDISP P= 0.47). There were statistically supported differences in community structure with drought and time/drought severity for both plants. Still, it is notable that there were no apparent differences in the temporal dynamics (rate and magnitude of change) across these treatments. In addition, average distance to the median in community structure (dispersion) was modestly higher in droughted bean rhizobiomes than in pre-drought and watered ones. Thus, hypothesis 1 was supported by the data. Addressing hypothesis 2: The plant mediates the responses of the active rhizobiome members to drought. The most striking difference between the rhizobiomes of the two crops was whether a plant was present. Though the presence of a plant was a significant explanatory factor for both crops' rhizobiomes, it had the highest explanatory value of all the factors tested for bean (17% explanatory value, Figure 4B) and relatively low explanatory for switchgrass (3% explanatory value, Figure 4C, Table S5). Furthermore, the interaction between plant presence and drought was only significant in bean rhizobiomes (PERMANOVA Pseudo-F=1.84, P=0.03) and not in switchgrass (PERMANOVA Pseudo-F=1.13, P=0.03) and not in switchgrass (PERMANOVA Pseudo-F=1.13). 0.18, **Table S5**). We conclude that whether the plant mediates rhizobiome responses to short-term drought depends on the plant. Thus, hypothesis 2 was partially supported and conditional on the plant species investigated. There were plant-mediated responses to drought for bean rhizobiomes but not for switchgrass.

Motivated by the plant-mediated effects of drought for bean, we next investigated the activated taxa that were uniquely associated with the planted drought condition in bean rhizobiomes. We identified 87 such taxa using indicator analysis (**Figure 5A**). We reasoned that the responses of these activated taxa are likely to be mediated, directly or indirectly, by the bean plant during drought.





Activity dynamics of indicator taxa in bean drought planted condition (50 most abundant out of 87 indicators)



668 669

671

672

673

674

675

676

677

678

679 680

681

682

683

684

685

686

687

688

689 690

691

692

693 694

695

696

697

698

699

700

701

702

703

Figure 5: A. Venn diagram showing the indicator species associated with each drought/planted condition in common bean (sample size (n)=203). B. heatmap showing the relative abundance dynamics of the top 50 most abundant and active OTUs from the 87 indicators associated with the bean planted drought condition. Inactive taxa are coded as "0" (zero), corresponding to the light gray cells. More abundant OTUs are more vellow, and less abundant are bluer. Taxa not detected in either DNA or RNA dataset is denoted as NA and colored as white in the heatmap. We standardized counts within an OTU relative to the maximum observed abundance value detected for that OTU across samples. Phantom taxa detected in >5% of samples were included. The legends on the right indicate classification at the Class level and the OTUID (sample size (n)=23). Focusing on the most abundant 50 indicators of the planted drought condition, we examined their activity dynamics over drought severity (Figure 5B). Several OTUs increased over time and with drought severity (enriched in Days 5-6, within clades 1 and 3), including members belonging to the classes Actinobacteria, Verrucomicrobiae, Acidobacteriae, Acidimicrobiia, Thermoleophilia, and Alphaproteobacteria. Several OTUs were detected consistently over time despite increasing drought severity (clade 4), including some belonging to Alphaproteobacteria, Planctomycetes, Gammaproteobacteria, and Actinobacteria (Figure 5B). These planted drought indicator taxa that were either enriched over time or stable are targets for follow-up research to understand their exact fitness advantages and potential benefits to the host in the drought environment. Notably, while the indicator species analysis revealed several OTUs strongly associated with each drought and planting level for bean rhizobiomes, no indicators were discovered for switchgrass, further supporting the previous finding of no interaction of planting and drought for switchgrass. In the drought samples planted with bean, we noted that many stable or enriched OTUs belonged within few classes. Thus, we next compared the overall differences and changes in the relative abundances of these classes between the treatments for bean rhizobiomes (Figure 6). One motivation for this analysis was to understand what classes the plant retains during drought by comparing the planted and unplanted drought conditions. Only taxa associated with Blastocatellia were absent in the unplanted drought versus the planted drought rhizobiomes. However, Gammaproteobacteria (Kruskal-Wallis chi-squared = 23.96, df = 1, P<0.001), Actinobacteria (Kruskal-Wallis chi-squared = 5.89, df = 1, P=0.02) and Alphaproteobacteria (Kruskal-Wallis chi-squared = 26.75, df = 1, P<0.001) were significantly higher in relative abundance in the planted compared to the unplanted treatments in bean under drought condition. Among these, Gammaproteobacteria (Kruskal-Wallis chi-squared = 31.59, df = 1, P<0.001) and

Alphaproteobacteria (Kruskal-Wallis chi-squared = 16.65, df = 1, P<0.001) were significantly increased in planted compared to unplanted treatments in watered bean treatments samples. At the same time, Actinobacteria was not statistically different between planted and unplanted treatments.

Another motivation was to identify potentially "lost" taxa during a drought that are normally associated with the bean. We explored drought-sensitive taxa by comparing the planted drought to the planted watered samples, reasoning that if any of these taxa are beneficial for the plant, they could be targets to prioritize for recovery after a drought. However, there were no notable absences or decreases of the major active classes in the planted soils across watered and drought conditions. Rather, the most notable comparative difference among treatment groups was between the planted and unplanted conditions, as several classes were only associated with the unplanted soils, including vadinHa49, Subgroup 5, Pla4_lineage, Oligoflexia, FCPU426, and bacteriap25. These classes are potentially selected against or non-competitive in the presence of the plant.

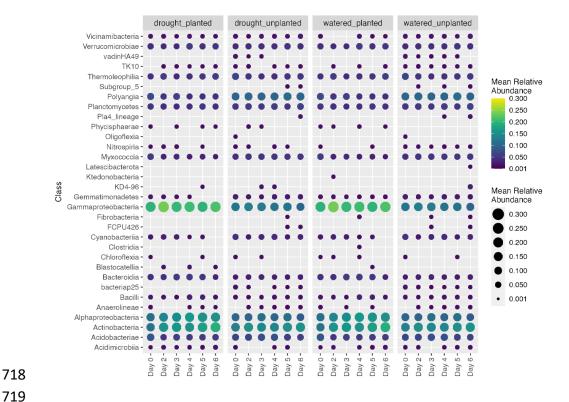


Figure 6. Temporal changes in relative abundances of active taxa that were the most abundant bacterial classes in bean (n=56) pre-drought and drought conditions.

725726

727

728

729

730

731

732

733

734

735736

737

738

739

740

741

742

743

744

745

746

747

748

749

750

751

752

753754

755

756

757

Discussion Soil microbes can be important in plant drought tolerance [58]. Improving our understanding of rootassociated soil microbial responses during drought stress can help improve a plant's resilience through targeted microbiome manipulation. This study used microbiome and metabolomics data to understand microbial community shifts during short-term drought in two plant species - common bean and switchgrass. We focused on the subset of the bacterial community that was likely to be active (RNAbased assessment) during drought by comparing unplanted and planted rhizosphere soils to watered soils. These comparisons were critical to tease apart plant-mediated responses to drought, which otherwise can be challenging to explore as the stress of drought independently will alter soil microbial communities and direct many taxa to dormancy. We deliberately included rhizosphere soils from fields with a recent legacy of growing each crop and thus had an imprint of their typically-associated rhizosphere microbiota. We predicted that in doing so, the microbiome assembly during short-term drought would be biologically relevant to the field and crop. The major differences in the active community structures between bean and switchgrass soils were maintained in drought and watered conditions and over time, with no indication of convergence across the two communities given the drought. We detected no specific taxa that were generally selected in drought across the different crops, neither at the OTU or family taxonomic level. While other studies have suggested that there could be similarities in the drought response of the microbiome across phylogenetically distant and related host species [59], our study suggests that developing universal, "drought-supportive" targets for microbiome manipulation could instead require solutions tailored to different plant families. Despite differences in rhizobiome compositions and different host sensitivities to drought, we observed that both bean and switchgrass rhizobiomes changed comparably given water reduction and over time with increasing drought severity. The major difference between the crops was that the presence of a plant was more important for the bean rhizobiome response to drought. At the same time, the plant presence was largely inconsequential for the switchgrass rhizobiome. Generally, switchgrass rhizobiome differences between the planted and unplanted treatments were weak relative to the bean. For the bean rhizobiomes, several OTUs within classes Actinobacteria and Alphaproteobacteria were enriched in the planted samples compared to unplanted samples during drought. Actinobacteria have generally been seen to be enriched during drought in plant roots [60, 61] and are well-known for their versatile capabilities ranging from stress tolerance to bioremediation. We also found that

Alphaproteobacteria and Actinobacteria increased relative abundance over drought severity for beanplanted treatments during drought.

Physiological and morphological differences between bean and switchgrass could explain the differences observed in their assembly of the rhizobiome during drought stress. Switchgrass is known to be generally drought-tolerant [35]. Its extensive and deep root architecture could indicate better resource and water allocation that benefits switchgrass during drought, making it less reliant on beneficial support from the plant's microbiome. On the other hand, bean plants could benefit from directed manipulation of the microbiome during short-term drought, given its wide range of drought susceptibility across genotypes [37]. A more expansive analysis of the microbiomes of perennial versus annual life histories and among plants with different physiologies could provide insights into expectations of different microbiome drought responses. It could be that the rhizosphere microbiomes of perennial plants are more generally more resistant to stress. Notably, in our study, the metabolome results show that the switchgrass plants were responsive to the drought and had clear shifts in some expected features of their metabolite profile as compared to the watered and unplanted conditions. Thus, the metabolite and microbiome changes were not necessarily coupled at the temporal scale of the experiment (days), nor were they collectively indicative of the plant-mediated shift in metabolism that subsequently drove a shift in its rhizosphere microbiome. The observation that the switchgrass microbiome was relatively stable drought could be due in part to specialized metabolites that switchgrass produces to combat stress.

Saponins are a group of specialized plant metabolites that can exhibit biological activities such as antibacterial, antifungal, and cytotoxic properties [9, 62]. Saponins are secreted from plant roots into the rhizosphere and can alter soil bacterial communities [63]. Saponins also have a role in detoxifying free radicals [64] thus helping combat the oxidative stress imposed by drought. In this study, saponins were increased in concentration in the planted switchgrass rhizospheres under drought. In other studies, saponin concentrations also increased in leaves of different plant species under water-deficit conditions [65, 66], suggesting similar impacts of drought in other plan. Actinobacteria are known to degrade saponins in the rhizosphere. They are also known as assimilators of saponins, which can explain the co-occurrence of saponins with Actinobacterial OTUs in our study [63]. In addition, the increased accumulations of diterpenoids were also observed in the planted rhizosphere soils stressed by drought. This is consistent with a recent metabolomics study that showed that diterpenoids are also enriched in switchgrass roots after a longer period of drought stress (five weeks) [54].

792

793

794

795

796 797

798

799

800

801

802

803

804

805

806

807

808

809

810811

812

813

814

815

816

817

818

819 820

821

822

823

824

We identified several bean OTUs that selectively increased only with increasing drought severity. This included OTUs belonging to classes Alphaproteobacteria, Blastocatellia, Actinobacteria, Thermoleophilia, Gammaproteobacteria, Verrucomicrobiae, Acidobacteriae, and Acidimicrobiia. These short-term stress responders are critical for further exploration as they may be potential bioinoculant targets. Our study suggests that considering the active members of the rhizobiome provides complementary insights into the ecology of drought and can suggest different responses of taxa. This study is different in its focus on detecting active taxa by rRNA:rRNA gene ratios. Other studies typically consider in aggregate the total rhizobiome community using DNA-based sequencing, including active, inactive, and deceased members. From these DNA-based studies, the literature reports that soils with limited moisture and under drought conditions have reduced the richness of Proteobacteria and Verrucomicrobia [67]. However, in our study, members of these groups were active and abundant in the rhizobiome of drought, unplanted bean rhizosphere soil. Out of the 46 bean taxa uniquely detected in the unplanted drought, several OTUs belonging to these phyla were detected as active, persistent, and in some cases, relatively high abundance. This included, for example, two Verrucomicrobiae OTUs (OTU#543e7a460e393e133062fbfae2f5675e and OTU#37184df09489f56490af1670195a8be5) as well as a Gammaproteobacteria OTU (OTU#da195c41827a582bf1b4bd7cb9007452) and Alphaproteobacteria OTU (OTU#fa5570e95914a27603f466f8d4789a26). Furthermore, though Actinobacteria are widely detected in drought soil [68-70], out of the 46 unique taxa for the unplanted drought samples, we only found one OTU belonging to Actinobacteria (OTU# 78b14ac22ebd6a508a7f122a42beb87a). Interestingly, we did not find this Actinobacteria OTU in the planted drought samples in bean rhizobiomes, though several other OTUs were detected from Actinobacteria. It could be that Actinobacteria are persistent but perhaps dormant or not highly active under some drought conditions and that specific OTUs have distinct responses during drought treatments that could change under the influence of a plant. Thus, broad generalizations at class or family levels remain challenging and specific cases must be evaluated at deeper taxonomic resolution. The RNA: DNA ratio method also has limitations that must be carefully considered in interpreting these results. For example, we set a ratio threshold of 1 for the detection of active taxa, which removed several taxa from consideration and, as such, retained only those that met our criteria. Thus, we could be missing the detection of additional active OTUs, especially among rare taxa. Other studies have applied RNA: DNA gene ratios ranging from 0.5-2 [28], with five considered overly conservative [20]. However, after

826

827

828829

830

831

832

833

834

835

836837

838

839

840

841

842

843

844

845

846

847 848

849

850

851852

853854855

856

857

858

the included taxa were designated, it was important to use the DNA sequence counts to designate taxon relative abundances (and not the RNA counts) to reduce bias due to the expected variation in transcription from taxon to taxon [71]. Another consideration of this study is the intense and short duration of the drought exposure, accumulating over six days. Our goal was to understand immediate drought response and selection of members in the rhizosphere, so we deliberately chose a shorter timeframe and pulse event to avoid the inclusion of longer-term assembly dynamics. As the rhizobiome assembly continues with drought and rewetting, its activity dynamics are expected to change. Thus, this work presents the immediate consequences of short-term rhizobiome drought, and additional work is needed to understand longer-term consequences. Conclusions This study assessed the active rhizobiome responses to short-term drought for two different crops: switchgrass and bean. Both plants' active rhizobiomes changed given the drought and over time with drought intensity, but the response was more muted for switchgrass than for bean. Using unplanted controls, we determined that there were plant-mediated effects of drought on the active rhizobiome only for bean plants and not for switchgrass, suggesting that different plants have different reliance on or engagements with or their rhizobiomes in drought. There was unique recruitment of taxa to drought bean plants, and we could distinguish among the plant- and environment-responsive taxa. This work identifies rhizobiome taxa that may be recruited by or lost by bean plants during drought, which can be targeted to improve understanding and tested for plant benefits. **Declarations** Ethics approval and consent to participate Not applicable **Consent for publication** Not applicable Availability of data and materials Sequence data for this study has been deposited to NCBI Sequence Read Archive under project number PRJNA862978. Scripts for analyzing the microbial community for this study can be found on GitHub at https://github.com/ShadeLab/PAPER DroughtRhizobiome Bandopadhyay 2023.

861

862863

864

865

866

867

868

869

870871

872

873

874

875876

877

878

Competing interests The authors declare that they have no competing interests. **Funding** This work was supported by the National Science Foundation Award Number MCB 1817377 to AS, by the Great Lakes Bioenergy Research Center, U.S. Department of Energy, Office of Science, Office of Biological and Environmental Research under Award Number DE-SC0018409, and by the National Science Foundation Long-term Ecological Research Program (DEB 1832042) at the Kellogg Biological Station. AS acknowledges support from the USDA National Institute of Food and Agriculture and Michigan State University AgBioResearch. **Authors contributions** AS and AWB designed the experiment. SB, XL, AWB conducted experiments. SB, XL, and AS generated figures. SB, XL, RLL and AS analyzed the data. SB prepared the first draft. All authors read and approved the final manuscript. Acknowledgments The authors acknowledge the Great Lakes Bioenergy Research Center Communications team for assistance in generating graphic design.

REFERENCES

883

884

- Naseri B, Younesi H: Beneficial microbes in biocontrol of root rots in bean crops: A metaanalysis (1990–2020). *Physiological and Molecular Plant Pathology* 2021, 116:101712.
- O'Callaghan M, Ballard RA, Wright D: Soil microbial inoculants for sustainable agriculture: Limitations and opportunities. *Soil Use and Management* 2022, 38(3):1340-1369.
- Hone H, Mann R, Yang G, Kaur J, Tannenbaum I, Li T, Spangenberg G, Sawbridge T: Profiling, isolation and characterisation of beneficial microbes from the seed microbiomes of drought tolerant wheat. *Scientific Reports* 2021, 11(1):11916.
- 4. Aslam MM, Okal EJ, Idris AL, Qian Z, Xu W, Karanja JK, Wani SH, Yuan W: Rhizosphere microbiomes can regulate plant drought tolerance. *Pedosphere* 2022, 32(1):61-74.
- Springer; 2013. Arora NK: Plant microbe symbiosis: fundamentals and advances: Springer; 2013.
- Xing-Feng H, M. CJ, F. RK, Ruifu Z, Qirong S, M. VJ: Rhizosphere interactions: root exudates, microbes, and microbial communities. *Botany* 2014, 92(4):267-275.
- Singh SK, Wu X, Shao C, Zhang H: Microbial enhancement of plant nutrient acquisition. *Stress Biology* 2022, 2(1):3.
- 899 8. Li X, Sarma SJ, Sumner LW, Jones AD, Last RL: Switchgrass Metabolomics Reveals Striking
 900 Genotypic and Developmental Differences in Specialized Metabolic Phenotypes. *Journal of*901 *Agricultural and Food Chemistry* 2022, 70(26):8010-8023.
- 902 9. Li X, Chou M-Y, Bonito GM, Last RL: Identification of anti-fungal bioactive terpenoids from the
 903 bioenergy crop switchgrass (Panicum virgatum). bioRxiv
 904 2023:2023.2002.2024.529965.
- 905 10. Meliho M, Khattabi A, Jobbins G, Sghir F: Impact of meteorological drought on agriculture in the Tensift watershed of Morocco. *Journal of Water and Climate Change* 2019, 11(4):1323-1338.
- Ziska L, Crimmins A, Auclair A, DeGrasse S, Garofalo J, Khan A, Loladze I, de León A,
 Showler A, Thurston J: Ch. 7: Food safety, nutrition, and distribution. *The impacts of climate change on human health in the United States: a scientific assessment* 2016:189-216.
- 910 12. Tietjen B, Schlaepfer DR, Bradford JB, Lauenroth WK, Hall SA, Duniway MC, Hochstrasser T,
 911 Jia G, Munson SM, Pyke DA *et al*: Climate change-induced vegetation shifts lead to more
 912 ecological droughts despite projected rainfall increases in many global temperate drylands.
 913 Global Change Biology 2017, 23(7):2743-2754.
- 914 13. Naylor D, Coleman-Derr D: Drought Stress and Root-Associated Bacterial Communities.
 915 Frontiers in Plant Science 2018, 8.
- Naseem H, Ahsan M, Shahid MA, Khan N: Exopolysaccharides producing rhizobacteria and their role in plant growth and drought tolerance. *J Basic Microbiol* 2018, 58(12):1009-1022.
- 918 15. Camaille M, Fabre N, Clément C, Ait Barka E: Advances in Wheat Physiology in Response to
 919 Drought and the Role of Plant Growth Promoting Rhizobacteria to Trigger Drought Tolerance.
 920 Microorganisms 2021, 9(4):687.
- 921 16. Qi M, Berry JC, Veley KW, O'Connor L, Finkel OM, Salas-González I, Kuhs M, Jupe J, Holcomb E, Glavina del Rio T *et al*: Identification of beneficial and detrimental bacteria impacting sorghum responses to drought using multi-scale and multi-system microbiome comparisons. *The ISME Journal* 2022, 16(8):1957-1969.
- 925 17. Santos-Medellín C, Liechty Z, Edwards J, Nguyen B, Huang B, Weimer BC, Sundaresan V:
 926 Prolonged drought imparts lasting compositional changes to the rice root microbiome. *Nature* 927 *Plants* 2021, 7(8):1065-1077.
- Santos-Medellín C, Edwards J, Liechty Z, Nguyen B, Sundaresan V, Ausubel FM: Drought
 Stress Results in a Compartment-Specific Restructuring of the Rice Root-Associated
 Microbiomes. *mBio* 2017, 8(4):e00764-00717.
- Yu L, Naylor D, Dong Z, Simmons T, Pierroz G, Hixson KK, Kim Y-M, Zink EM, Engbrecht
 KM, Wang Y et al: Drought delays development of the sorghum root microbiome and enriches

- 933 for monoderm bacteria. *Proceedings of the National Academy of Sciences* 2018, 115(18):E4284-934 E4293.
- 935 20. Bowsher AW, Kearns PJ, Shade A: 16S rRNA/rRNA Gene Ratios and Cell Activity Staining
 936 Reveal Consistent Patterns of Microbial Activity in Plant-Associated Soil. mSystems 2019, 4(2).
- Jones SE, Lennon JT: Dormancy contributes to the maintenance of microbial diversity.
 Proceedings of the National Academy of Sciences 2010, 107(13):5881-5886.
- Cole JJ: Aquatic microbiology for ecosystem scientists: new and recycled paradigms in ecological microbiology. *Ecosystems* 1999, 2(3):215-225.
- Trivedi P, Batista BD, Bazany KE, Singh BK: Plant-microbiome interactions under a changing world: responses, consequences and perspectives. *New Phytologist* 2022, 234(6):1951-1959.
- 943 24. Bennett JA, Klironomos J: Mechanisms of plant–soil feedback: interactions among biotic and abiotic drivers. *New Phytologist* 2019, 222(1):91-96.
- Pugnaire FI, Morillo JA, Peñuelas J, Reich PB, Bardgett RD, Gaxiola A, Wardle DA, van der
 Putten WH: Climate change effects on plant-soil feedbacks and consequences for biodiversity and
 functioning of terrestrial ecosystems. *Science Advances* 2019, 5(11):eaaz1834.
- Jochum MD, McWilliams KL, Pierson EA, Jo Y-K: Host-mediated microbiome engineering (HMME) of drought tolerance in the wheat rhizosphere. *PLoS ONE* 2019, 14.
- 950 27. Hund K, Schenk B: The microbial respiration quotient as indicator for bioremediation processes.
 951 *Chemosphere* 1994, 28(3):477-490.
- 952 28. Blagodatskaya E, Kuzyakov Y: Active microorganisms in soil: critical review of estimation criteria and approaches. *Soil Biology and Biochemistry* 2013, 67:192-211.
- Denef K, Roobroeck D, Wadu MCM, Lootens P, Boeckx P: Microbial community composition
 and rhizodeposit-carbon assimilation in differently managed temperate grassland soils. *Soil Biology and Biochemistry* 2009, 41(1):144-153.
- 957 30. Butler JL, Williams MA, Bottomley PJ, Myrold DD: Microbial community dynamics associated with rhizosphere carbon flow. *Applied and Environmental Microbiology* 2003, 69(11):6793-6800.
- Treonis AM, Ostle NJ, Stott AW, Primrose R, Grayston SJ, Ineson P: Identification of groups of metabolically-active rhizosphere microorganisms by stable isotope probing of PLFAs. *Soil Biology and Biochemistry* 2004, 36(3):533-537.
- 32. Lu Y, Abraham WR, Conrad R: Spatial variation of active microbiota in the rice rhizosphere revealed by in situ stable isotope probing of phospholipid fatty acids. *Environmental Microbiology* 2007, 9(2):474-481.
- Tian J, Dippold M, Pausch J, Blagodatskaya E, Fan M, Li X, Kuzyakov Y: Microbial response to rhizodeposition depending on water regimes in paddy soils. *Soil Biology and Biochemistry* 2013, 65:195-203.
- 968 34. Norton JM, Firestone MK: Metabolic status of bacteria and fungi in the rhizosphere of ponderosa pine seedlings. *Applied and Environmental Microbiology* 1991, 57(4):1161-1167.
- Barney JN, Mann JJ, Kyser GB, Blumwald E, Van Deynze A, DiTomaso JM: Tolerance of switchgrass to extreme soil moisture stress: Ecological implications. *Plant Science* 2009, 177(6):724-732.
- Ye Z, Sangireddy SR, Yu C-L, Hui D, Howe K, Fish T, Thannhauser TW, Zhou S: Comparative
 Proteomics of Root Apex and Root Elongation Zones Provides Insights into Molecular
 Mechanisms for Drought Stress and Recovery Adjustment in Switchgrass. *Proteomes* 2020,
 8(1):3.
- 977 37. López CM, Pineda M, Alamillo JM: Differential Regulation of Drought Responses in Two
 978 Phaseolus vulgaris Genotypes. *Plants* 2020, 9(12):1815.
- 38. JoVE: Determination of Moisture Content in Soil. In: *Environmental Microbiology*. Cambridge,
 980 MA: Science Education Database; 2022.
- 981 39. Griffiths RI, Whiteley AS, O'Donnell AG, Bailey MJ: Rapid Method for Coextraction of DNA
 982 and RNA from Natural Environments for Analysis of Ribosomal DNA- and rRNA-Based

- 983 Microbial Community Composition. *Applied and Environmental Microbiology* 2000, 66(12):5488-5491.
- Gaporaso JG, Lauber CL, Walters WA, Berg-Lyons D, Lozupone CA, Turnbaugh PJ, Fierer N,
 Knight R: Global patterns of 16S rRNA diversity at a depth of millions of sequences per sample.
 Proceedings of the national academy of sciences 2011, 108(supplement 1):4516-4522.
- 41. Kozich JJ, Westcott SL, Baxter NT, Highlander SK, Schloss PD: Development of a Dual-Index
 989 Sequencing Strategy and Curation Pipeline for Analyzing Amplicon Sequence Data on the MiSeq
 990 Illumina Sequencing Platform. Applied and Environmental Microbiology 2013, 79(17):5112 991 5120.
- Golovas J, Bintarti AF, Mechan Llontop ME, Grady KL, Shade A: Do-it-Yourself Mock
 Community Standard for Multi-Step Assessment of Microbiome Protocols. *Current Protocols* 2022, 2(9):e533.
- 995 43. Chong J, Wishart DS, Xia J: Using MetaboAnalyst 4.0 for comprehensive and integrative metabolomics data analysis. *Current protocols in bioinformatics* 2019, 68(1):e86.
- 997 44. Dührkop K, Nothias L-F, Fleischauer M, Reher R, Ludwig M, Hoffmann MA, Petras D, Gerwick
 998 WH, Rousu J, Dorrestein PC: Systematic classification of unknown metabolites using high 999 resolution fragmentation mass spectra. *Nature Biotechnology* 2021, 39(4):462-471.
- Sumner LW, Amberg A, Barrett D, Beale MH, Beger R, Daykin CA, Fan TWM, Fiehn O,
 Goodacre R, Griffin JL *et al*: Proposed minimum reporting standards for chemical analysis.
 Metabolomics 2007, 3(3):211-221.
- Bolyen E, Rideout JR, Dillon MR, Bokulich NA, Abnet CC, Al-Ghalith GA, Alexander H, Alm
 EJ, Arumugam M, Asnicar F: Reproducible, interactive, scalable and extensible microbiome data
 science using QIIME 2. *Nature biotechnology* 2019, 37(8):852-857.
- 1006 47. Callahan BJ, McMurdie PJ, Rosen MJ, Han AW, Johnson AJA, Holmes SP: DADA2: High-1007 resolution sample inference from Illumina amplicon data. *Nature methods* 2016, 13(7):581-583.
- Weinstein MM, Prem A, Jin M, Tang S, Bhasin JM: FIGARO: An efficient and objective tool for optimizing microbiome rRNA gene trimming parameters. *bioRxiv* 2019:610394.
- 1010 49. Quast C, Pruesse E, Yilmaz P, Gerken J, Schweer T, Yarza P, Peplies J, Glöckner FO: The
 1011 SILVA ribosomal RNA gene database project: improved data processing and web-based tools.
 1012 Nucleic acids research 2012, 41(D1):D590-D596.
- Davis NM, Proctor DM, Holmes SP, Relman DA, Callahan BJ: Simple statistical identification and removal of contaminant sequences in marker-gene and metagenomics data. *Microbiome* 2018, 6(1):226.
- 1016 51. McMurdie PJ, Holmes S: phyloseq: An R Package for Reproducible Interactive Analysis and Graphics of Microbiome Census Data. *PLOS ONE* 2013, 8(4):e61217.
- 1018 52. Dixon P: VEGAN, a package of R functions for community ecology. *Journal of Vegetation Science* 2003, 14(6):927-930.
- 1020 53. Chytrý M, Tichý L, Holt J, Botta-Dukát Z: Determination of diagnostic species with statistical fidelity measures. *Journal of Vegetation science* 2002, 13(1):79-90.
- Tiedge K, Li X, Merrill AT, Davisson D, Chen Y, Yu P, Tantillo DJ, Last RL, Zerbe P:
 Comparative transcriptomics and metabolomics reveal specialized metabolite drought stress responses in switchgrass (Panicum virgatum). New Phytologist 2022, 236(4):1393-1408.
- 1025 55. Li Y, Kong Y, Teng D, Zhang X, He X, Zhang Y, Lv G: Rhizobacterial communities of five co-1026 occurring desert halophytes. *PeerJ* 2018, 6:e5508.
- 1027 56. Matthews A, Pierce S, Hipperson H, Raymond B: Rhizobacterial community assembly patterns vary between crop species. *Frontiers in microbiology* 2019, 10:581.
- 1029 57. Pérez-Jaramillo JE, Mendes R, Raaijmakers JM: Impact of plant domestication on rhizosphere microbiome assembly and functions. *Plant molecular biology* 2016, 90(6):635-644.
- Fitzpatrick CR, Copeland J, Wang PW, Guttman DS, Kotanen PM, Johnson MTJ: Assembly and ecological function of the root microbiome across angiosperm plant species. *Proceedings of the National Academy of Sciences* 2018, 115(6):E1157-E1165.

- Naylor D, DeGraaf S, Purdom E, Coleman-Derr D: Drought and host selection influence bacterial community dynamics in the grass root microbiome. *The ISME Journal* 2017, 11(12):2691-2704.
- 1036 60. Bogati K, Walczak M: The impact of drought stress on soil microbial community, enzyme activities and plants. *Agronomy* 2022, 12(1):189.
- Breitkreuz C, Herzig L, Buscot F, Reitz T, Tarkka M: Interactions between soil properties, agricultural management and cultivar type drive structural and functional adaptations of the wheat rhizosphere microbiome to drought. *Environmental Microbiology* 2021, 23(10):5866-5882.
- Sparg S, Light M, Van Staden J: Biological activities and distribution of plant saponins. *Journal* of ethnopharmacology 2004, 94(2-3):219-243.
- 1043 63. Nakayasu M, Yamazaki S, Aoki Y, Yazaki K, Sugiyama A: Triterpenoid and Steroidal Saponins Differentially Influence Soil Bacterial Genera. *Plants* 2021, 10(10):2189.
- 1045 64. Chen Y, Miao Y, Huang L, Li J, Sun H, Zhao Y, Yang J, Zhou W: Antioxidant activities of saponins extracted from Radix Trichosanthis: an in vivo and in vitro evaluation. *BMC* 1047 Complementary and Alternative Medicine 2014, 14(1):1-8.
- 1048 65. Odjegba VJ, Alokolaro AA: Simulated drought and salinity modulates the production of phytochemicals in Acalypha wilkesiana. *Journal of Plant Studies* 2013, 2(2):105.
- El-Sayed A, Razin A, Swaefy H, Mohamed S, Abou-Aitah K: Effect of water stress on yield and bioactive chemical constituents of Tribulus species. *J Appl Sci Res* 2008, 4(12):2134-2144.
- 1052 67. Mohammadipanah F, Zamanzadeh M: Bacterial Mechanisms Promoting the Tolerance to
 1053 Drought Stress in Plants. In: Secondary Metabolites of Plant Growth Promoting
 1054 Rhizomicroorganisms: Discovery and Applications. Edited by Singh HB, Keswani C, Reddy MS,
 1055 Sansinenea E, García-Estrada C. Singapore: Springer Singapore; 2019: 185-224.
- Jurburg SD, Natal-da-Luz T, Raimundo J, Morais PV, Sousa JP, van Elsas JD, Salles JF:
 Bacterial communities in soil become sensitive to drought under intensive grazing. Science of The
 Total Environment 2018, 618:1638-1646.
- 1059 69. Acosta-Martinez V, Cotton J, Gardner T, Moore-Kucera J, Zak J, Wester D, Cox S: Predominant bacterial and fungal assemblages in agricultural soils during a record drought/heat wave and linkages to enzyme activities of biogeochemical cycling. *Applied Soil Ecology* 2014, 84:69-82.
- 1062 70. Berard A, Sassi MB, Kaisermann A, Renault P: Soil microbial community responses to heat wave components: drought and high temperature. *Climate Research* 2015, 66(3):243-264.
- 1064 71. Blazewicz SJ, Barnard RL, Daly RA, Firestone MK: Evaluating rRNA as an indicator of microbial activity in environmental communities: limitations and uses. *The ISME Journal* 2013, 7(11):2061-2068.

1071 1072

1073

10741075

1076

1077

1078

1079 1080

1081

1082

1083

1084

10851086

1087

1088

1089

10901091

1092

1093

1094

1095

1096

1097

1098

1099

1100 1101

1102

1103 1104 1105

Figure Legends Figure 1. The experimental design used in the greenhouse study. Day 2 to day 6 correspond to increased drought severity for the 6-day drought treatment. Figure 2. A) Gravimetric soil moisture of bean (sample size (n)=110) and switchgrass (sample size (n)=110) planted and unplanted soils at different sampling days for drought and watered conditions. B) Shoot biomass of bean (sample size (n)=55) and switchgrass (sample size (n)=55) at different sampling days for drought and watered conditions. **Figure 3.** VIP coefficient score plot from a PLS-DA model showing the top 15 variables (features) of importance to differentiate the different switchgrass soil types. Feature 10.11 1000.5271n, a previously identified switchgrass saponin, had the highest coefficient and is thus the most important detected variable during the drought treatment. At the end of each feature name, 'm/z' stands for mass-to-charge ratio and 'n' stands for neutral mass. Figure 4. A) Principal Coordinate Analysis (PCoA) comparing bean and switchgrass active community composition (sample size (n)=212). PCoA of B) bean (sample size (n)=105) and C) switchgrass (sample size (n)=107) comparing active communities of planted and unplanted soil samples, as well as comparisons across drought treatments: pre-drought, drought, and watered. Figure 5. A) Venn diagram showing the indicator species associated with each drought/planted condition in common bean (sample size (n)=203). B) Heatmap showing the relative abundance dynamics of the top 50 most abundant and active OTUs from the 87 indicators associated with the bean planted drought condition. Inactive taxa are coded as "0" (zero), corresponding to the light gray cells. More abundant OTUs are more yellow, and less abundant are bluer. Taxa not detected in either DNA or RNA dataset is denoted as NA and colored as white in the heatmap. We standardized counts within an OTU relative to the maximum observed abundance value detected for that OTU across samples. Phantom taxa detected in >5% of samples were included. The legends on the right indicate classification at the Class level and the OTUID (n=23). Figure 6. Dot plots demonstrate the temporal changes in relative abundances of active taxa that were the most abundant bacterial classes in bean (n=56) pre-drought and drought conditions.

 Table S1. Chemical and physical analysis of field-collected soils for bean and switchgrass.

Table S2. Water addition schedule by treatment for switchgrass.

Table S3. Water addition schedule by treatment for bean plants.

Table S4. Differences in gravimetric soil moisture and shoot biomass by experimental factors as assessed by a three-way ANOVA with interactions (Type III). The F-statistics are reported with significance indicated at p<0.05*, p<0.01**, and p<0.001***.

Table S5. Permuted analysis of variance analysis to determine differences in microbiome structure given experimental treatments (a.k.a. factors). Sums of squares, degrees of freedom, R-squared values, and Pseudo F values are reported. The significance of each test is indicated at p<0.05*, p<0.01**, and p<0.001***.

Table S6. Differences in the number of observed active OTUs (richness) across experimental factors, as assessed using a three-way ANOVA with interactions (Type III tests). F values are reported, and significance is indicated at p<0.05*, p<0.01**, and p<0.001***.

1129 1130

1131

11321133

1134

1135

11361137

1138

1139

1140

1141

11421143

1144

1145

1146

1147

1148

1149

1150

11511152

1153

1154

11551156

1157

11581159

1160

1161

Supplemental Figures Figure S1. Photo of the experimental setup in the greenhouse for A) switchgrass (inset-tube showing switchgrass roots grown to fill to the inside edge of the tube) and B) bean. Figure S2. Sample collection for the switchgrass untargeted metabolomics. A. Sample panel with the numbers of replicates indicated. B. The drought and watered switchgrass tissue and planted soil replicates were combined from the samples collected from individual plants. Figure S3. Sequencing quality control checks. (A) Library sizes of true samples (blue) and negative control (red) samples for both DNA and cDNA-sequenced samples. (B) OTUs were designated as either true or false contaminants based on their prevalence in true samples and negative controls using package decontam in R. (C) Rarefaction curves for all DNA and cDNA samples (pre-rarefaction). Curves start to plateau out at 13-15k reads (n=438). Figure S4. The untargeted metabolomics revealed metabolite differences between the soil types. A. Venn diagram documenting comparative metabolite profiles and the number of specific and shared features identified in switchgrass root, shoot, and soil. B. PCA scores plot of the metabolomes of different soil types showing the principal components PC1 and PC3. C. 3-D scatter plot of PCA scores plot of the metabolomes of different soil types showing PC1, 2, and 3. E. Heatmap showing relative abundances of the top 50 PLS-DA most important features (vertical axis) across the biological samples (horizontal axis). For the hierarchical clustering of features, Euclidean and Ward.D were used as the distance measure and clustering method, respectively (sample size root (n)=30, shoot (n)=30, soil(n)=49). **Figure S5.** MS/MS spectra for the feature 10.11 1000.5271n, obtained by the positive LC-MS analysis, DDA mode. Top trace: survey scan; bottom trace: MS/MS. The molecular ions and sapogenin core fragment ions were annotated. M, molecular ion. Figure S6. Recovered DNA concentrations of samples by crop, treatment, and time point. DNA was measured using Oubit 2.0 with the dsDNA BR assay kit. Figure S7. Stacked bar plots depicting the Class-level, active bacterial community composition by treatment for bean and switchgrass. All samples collected are shown (n=212).

1163

1164

1165

1166

1167 1168 1169

1170

1171

11721173

1174

11751176

1177

11781179

1180

1181

1182

11831184

1185

1186

1187

Figure S8. For each experimental treatment, changes in Bray Curtis similarity to the pre-drought community over time and/or with increasing severity. Sampling day is indicated by the symbol size and treatment by colors (blue is planted, green is unplanted) and line type (solid is drought, and dashed is watered). The linear regressions were insignificant for any treatment except for the watered, planted switchgrass (blue, dashed line on the right panel). **Supplemental Data Data 1.** The metabolite features identified from Cave-in-Rock shoot tissues (with abundance ≥ 500) by the positive mode LC-MS analysis. **Data 2.** The metabolite features identified from Cave-in-Rock root tissues (with abundance ≥ 500) by the positive mode LC-MS analysis. **Data 3.** The metabolite features identified from all soil samples (with abundance ≥ 500) by the positive mode LC-MS analysis. **Data 4.** The top 50 PLS-DA important metabolite features (obtained by positive mode LC-MS analysis). Annotations were performed by searching the online mass spec databases through the Progenesis QI or using the CANOPUS machine function built in the SIRIUS 4 (https://bio.informatik.uni-jena.de/sirius/) (Materials and Methods). The ones annotated by CANOPUS are in boldface. ^a Compounds were previously reported in [54]; ^b Compounds were previously reported in [8]. ^c Identification level (A; B; C)-(A) standard or NMR; (B(i)) confident match based on MS/MS and (B(ii)) confident match using in-silico MS/MS approaches and (B(iii)) partial match based on MS/MS and (C(i)) confident match based on MSn and (C(ii)) confident match using in-silico MSn approaches and (C(iii)) partial match based on MSn; (D) MS only. These are based on the criteria for metabolite identification in [45].