

Enzymatic degradation of tetracycline by Trametes versicolor laccase in a fluidized bed reactor

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| 1 | Enzymatic degradation of tetracycline by <i>Trametes Versicolor</i> laccase in a |
|----|-----------------------------------------------------------------------------------------------|
| 2 | fluidized bed reactor |
| 3 | |
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| 9 | |
| 10 | Abstract |

Laccase from Trametes Versicolor was successfully immobilized on gelatin beads by a 11 crosslinking reaction with glutaraldehyde. Immobilized laccases showed better stability 12 13 towards pH and temperature than free laccases. Moreover, the immobilized laccases retained a good relative activity of 85% after 20 days of storage at 4°C. The degradation of 14 tetracycline (TC) was studied with immobilized enzymes in both batch and fluidized bed 15 reactors (FBR). The average degradation rate (1.59 mg h⁻¹ U_{enzymes}⁻¹) estimated over 24 hours 16 in the FBR was almost 5 times higher than in the stirred tank reactor. Maximum degradation 17 rate achieved was $72 \pm 1\%$ with a circulation flow rate of 80 mL min⁻¹ and addition of air at a 18 19 flowrate of 15 mL min⁻¹. Study of the stability of the active beads under reaction conditions, 20 shows that 45% of the TC was degraded after 5 cycles of 24 hours each. The toxicity of the TC solution before and after treatment was also investigated with microtox assays. 21

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 tetracycline degradation

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27 **1. Introduction**

28 Over the last few decades, worldwide population growth and ageing have led to increased 29 consumption of pharmaceutically active compounds (PhACs). However, many of these 30 compounds are resistant to conventional treatments used in Wastewater Treatment Plants 31 (WWTPs), and their residues have been detected in surface water and sometimes even in 32 drinking water (Couto et al., 2019; Petrie et al., 2015; Mirsaei et al., 2018). The occurrence and persistence of PhACs in water circuits represents a major global concern due to their 33 34 negative impact on the environment and health (aquatic toxicity, endocrine disruption, 35 development of antibiotic resistance in some bacteria, etc.) (Grenni et al, 2018; Vieira et al., 2021). Consequently, a number of recent research projects have focused on the development 36 of tertiary treatments to remove the resistant compounds. They include adsorption with 37 38 activated carbon (Rajapaksha et al., 2019), membrane processes (Khanzada et al., 2020), 39 advanced oxidation processes (Kanakaraju et al., 2018) and hybrid processes (Grandclement 40 et al., 2017).

Among these processes, bioremediation involving enzymes appears to be a promising option
because enzymatic reactions are carried out under mild conditions (pH, temperature).

43 Particular attention has been focused on oxydo-reductases and especially laccases, which only

44 need the dissolved oxygen in water to degrade a broad range of micro pollutants (Varga et al.,

45 2019; Bilal et al., 2019; Morsi et al., 2020). However, due to the high cost of enzymes and the

difficulty of recovering them once they are dissolved in the reaction medium, immobilization 46 47 of the enzymes onto a support would appear to be a cost-effective solution for improving the biocatalyst stability and allowing their re-used (Basso & Serban, 2019). Furthermore, the use 48 of immobilized catalysts enables continuous operation, which is important for large-scale 49 operations. In a recent review on the potential of free and immobilized oxidative enzymes for 50 wastewater treatment, Zdarta et al (2022) highlighted the key role of the immobilization 51 52 method and the type of solid support on the activity and stability of the biocatalyst. Indeed, a significant loss of activity during the immobilization step can considerably reduce the 53 economic interest of the process. 54

55 Among the large family of oxidative enzymes, laccases have been successfully used for antibiotic degradation after being immobilized on various inorganic or organic supports 56 including membranes (de Cazes et al., 2014; Taheran et al., 2017), monoliths (Ahmad et al., 57 58 2020) and solid particles (Shi et al., 2014; Wen et al. 2019; Zdarta et al., 2020). Enzymes immobilized on solid particles are currently used in stirred tanks. However, if a continuous 59 process is envisaged, a packed-bed (PBR) or a fluidized-bed reactor (FBR) would be more 60 convenient. In FBRs, the continuous ascendant fluid flow maintains the particles of 61 biocatalyst in suspension, avoiding risks of column clogging and preferential flow paths. 62 63 Mass transfer in such reactors is also facilitated, but high shear stress can denature the biocatalyst (Dal Magro et al., 2021). Despite this drawback, fluidized-bed reactors are widely 64 used for wastewater treatment with whole cells (Swain et al., 2018) and recently, the use of 65 66 immobilized laccases in FBRs has been reported for the biotransformation of bisphenol A (Piao et al., 2019) and estrogens (Sun et al., 2021). 67

A previous study (Harguindeguy et al., 2021) reported a promising sustainable process for
preparing a gelatin-based support for enzyme immobilization. The gelatin beads, which were
obtained by coacervation of gelatin droplets in a polyethylene glycol solution, were cross-

71 linked with glutaraldehyde allowing, on one hand, the improvement of their mechanical and 72 thermal properties and, on the other hand, the covalent bonding of laccases on the surface of the beads. These biocatalytic beads present low density, requiring low fluid velocity to 73 74 fluidize them. Energy consumption and shear stress are limited, avoiding any risk of laccase denaturation. Such active beads therefore appear to be good candidates for use in FBRs. 75 76 The present work aims to study the feasibility of micro-pollutant depletion by the laccase of 77 Trametes versicolor immobilized onto gelatin beads in an FBR. As explained in the introduction, immobilization generally enhances enzymes stability whereas FBR are very 78 interesting because they operate under continuous conditions and improve the mass transfer of 79 80 substrates towards active sites. From different types of micro-pollutants (pharmaceuticals, personal care products, pesticides), we selected an antibiotic (tetracycline (TC)) as a model 81 substrate. TC is one of the most widely used veterinary antibiotics in Europe (European 82 83 Medicine Agency, 2021) and is consequently frequently found in aquatic systems, from municipal wastewaters to seawaters (Abejon et al., 2015). Moreover, the accumulation of TC 84 85 in wastewater treatment plants is responsible for the proliferation of antibiotic resistant bacteria (ARB), which represents a real health risk for humans and animals (Nguyen et al., 86 2021). Furthermore, previous studies have demonstrated that TC can be degraded by the 87 88 laccase of *Trametes versicolor* without requiring a mediator (de Cazes et al., 2014). Firstly, the stability of the immobilized laccases with respect to pH, temperature and during 89 storage was studied in a continuously stirred-tank reactor. Then, the active-beads were used in 90 an FBR and experiments were carried out to evaluate the effect of operating parameters (fluid 91 92 velocity, oxygen supply by bubbling air) on TC removal. Finally, the acute toxicity of the TC solution before and after enzymatic treatment was evaluated by Microtox® assays. 93

94

95 **2. Materials and Methods**

96 2.1 Chemicals and enzyme

| 97 | Gelatin porcine skin, (MW \approx 50 – 100 kDa, bloom 300, type A, CAS 9000-70-8), |
|-----|----------------------------------------------------------------------------------------------------------|
| 98 | polyethylene glycol (PEG, molecular weight 10000, CAS CAS 25322-68-3), glutaraldehyde |
| 99 | aqueous solution grade II (25% in H ₂ O, CAS : 1111-30-8), laccase (from Trametes versicolor, |
| 100 | CAS 80498-15-3), and 2'-Azino-bis(3-ethylbenzothiazoline-6-sulfonic acid)diammonium salt |
| 101 | (CAS 30931-67-0), also known as ABTS, were purchased from Sigma-Aldrich. Stock |
| 102 | solution of TC (Sigma Aldrich, CAS 60-54-8) was prepared in ultra-pure water (pH 6). Buffer |
| 103 | solutions were made from citric acid (CAS 77-92-9), and di-sodium hydrogen phosphate |
| 104 | dehydrate (CAS 10028-24-7). All these chemicals were of ACS reagent grade. |
| 105 | 2.2 Active bead preparation and characterization |
| 106 | 2.2.1 Bead preparation: |
| 107 | Active beads were prepared according to a previously reported three-step process |

(Harguindeguy et al, 2021). Firstly, an aqueous gelatin solution (22% wt, 60°C) was added 108 dropwise with a syringe pump (SP200, flow rate 4 mL min⁻¹) inside an aqueous PEG₁₀₀₀₀ 109 solution (40% w/v, 50°C). The PEG solution was then cooled down in an ice bath at 5°C and 110 the formed beads were recovered and washed with distilled water to remove excess PEG. The 111 112 gelatin beads were then cross-linked and activated by immersion for 1 hour in a 3% v/v glutaraldehyde solution prepared in a citrate phosphate buffer solution (50 mM, pH 7) at 25°C 113 under magnetic stirring (250 rpm). After the activation step, the beads were thoroughly rinsed 114 115 with the same buffer solution to remove the glutaraldehyde that did not react and were immersed for 2 hours in a solution of *Trametes versicolor* laccase (prepared at 3 U mL⁻¹in the 116 117 same citrate phosphate buffer solution) at 25°C under magnetic stirring (250 rpm). At the end of the grafting step, the beads were washed 3 times with the same citrate phosphate buffer 118 solution and stored at 4°C in the buffer until being used. For blank experiments, deactivated 119

| 120 | beads were obtained by immersion of active beads in boiling water (100°C) for 30 minutes to |
|-----|----------------------------------------------------------------------------------------------------|
| 121 | deactivate the enzymes. The characteristics of the active beads are reported in Table 1. |
| 122 | (Table 1) |
| 123 | |
| 124 | <u>2.2.2 Activity assay</u> |
| 125 | Enzyme activity measurements are based on the oxidation of 2,2'-Azino-bis (3- |
| 126 | ethylbenzothiazoline-6-sulfonic acid)diammonium salt (ABTS). 900 μ L of ABTS (1 mM, |
| 127 | citrate-phosphate buffer (50 mM, pH 4)) was mixed with 100 μ L of laccase solution (10 mg |
| 128 | L ⁻¹ in the citrate-phosphate buffer) in a spectrophotometric cell and the reaction was |
| 129 | monitored with a Shimadzu UV-2401 PC spectrophotometer at 420 nm along with the |
| 130 | software UV Probe version 2.21. One unit of enzyme activity (U) was defined as the amount |
| 131 | of enzymes that converted 1 μ mol of ABTS to its cation radical per minute at pH 4 and 25°C. |
| 132 | For immobilized enzymes, 0.7 g of active beads were immersed in 25 mL of ABTS solution |
| 133 | (1 mM in citrate-phophate buffer (50 mM, pH 4, 25°C)), under magnetic stirring (250 rpm), |
| 134 | and the absorbance of the solution was monitored every minute for 15 minutes. Both |
| 135 | conditions of activity measurement described above are referenced as standard conditions. |
| 136 | |
| 137 | 2.2.3 pH and temperature stability |
| 138 | pH and thermal stability were determined by incubating free and immobilized enzymes in |
| 139 | three citrate phosphate buffer solutions (50 mM, pH 3, 5 or 9) during 24 hours at 25°C or in |

citrate phosphate buffer solution (50 mM, pH 7) at 60°C during 6 hours. At different times,
samples were withdrawn and enzymatic activity was measured with ABTS under standard

142 conditions (1 mM, pH 4, 25°C). Results are expressed as relative activity (i.e. ratio between

the activity measured and the initial activity). The experiments were carried out in triplicateand the average value is displayed.

145 <u>2.2.4 Kinetics constants</u>

| 146 | The apparent kinetic parameters of the Michaelis–Menten equation (K_m and V_{max}) of free and |
|-----|------------------------------------------------------------------------------------------------------|
| 147 | immobilized laccase towards TC were determined from initial reaction rate measurements. |
| 148 | TC solutions of $2 - 5 - 10 - 15 - 25$ ppm were prepared in ultra-pure water (pH 6). The |
| 149 | reactions were carried out in batch at 25°C, with an enzyme concentration equivalent to 30 U |
| 150 | L ⁻¹ for both free and immobilized enzymes. |
| 151 | 2.2.5 Storage stability |
| 152 | The storage stability of the free and immobilized enzymes was evaluated, by incubating both |
| 153 | laccase forms in a citrate phosphate buffer (50 mM, pH 7) at 4 °C. The residual activity was |
| 154 | obtained by measuring the remaining activity every 5 days for 20 days under standard |
| 155 | conditions. |
| 156 | |
| 157 | 2.3 Tetracycline degradation |
| 158 | |
| 159 | All the TC degradation assays were carried out with a TC solution prepared at 20 ppm in ultra |
| 160 | pure water (pH 6). Firstly, TC degradation with both free and immobilized enzymes was |
| 161 | studied in a stirred tank reactor (10 mL) at 25°C during 24 hours with an enzyme activity |
| 162 | equivalent to 100 U L^{-1} of laccase. Then, degradation assays with active beads were carried |
| 163 | out in fluidized-bed reactor operating in recirculation (see Figure 1). A jacketed glass column |
| 164 | (height 23 cm; internal diameter 2 cm) was filled with wet active beads (~ 4 g corresponding |

to a final enzyme activity of 60 U L^{-1}). The height and porosity of the fixed bed were equal to

166 5 cm and 0.36 respectively. The minimum fluidizing velocity (U_{mf}) was calculated according 167 to eq (1) since the Re value (eq. 2) at U_{mf} was below 20 (Gidaspow, 1994).

168

169
$$U_{mf} = \frac{d_p^2}{150} \cdot \frac{\varepsilon_{mf}^3}{(1 - \varepsilon_{mf})} \cdot \frac{(\rho_s - \rho_f)}{\mu_f} \cdot g \qquad (Eq.1)$$

170
$$R_e = \frac{\rho_f \cdot U_{mf} \cdot d_p}{\mu_f} \qquad (Eq.2)$$

171

172 Where dp is the bead diameter, ε_{mf} is the porosity of the bed at U_{mf} which is assumed to be 173 similar to that of the fixed bed, ρ_s and ρ_f are the density of bead and fluid respectively, μ_f is 174 the fluid viscosity and g is the normal gravity.

175 In the present work, U_{mf} calculated with the mean value of bead diameter (2.0 mm), was 0.21

176 cm min⁻¹ (a fluid flow of 40 mL min⁻¹) which corresponded to a Re value of 4.3.

177 TC solution (100 mL) was pumped continuously upwards through the column using a gear

pump (micro pump serial number 951990005, Cole Parmer) at different flow rates: 30, 80 and

179 130 mL min⁻¹ corresponding to approximately 0.8 U_{mf} , 2 U_{mf} and 3.2 U_{mf} respectively. The

180 column and the feed tank temperatures were controlled at 25° C by a water bath with a

181 circulation system. In some experiments, a continuous flow of air (15 mL min⁻¹) was bubbled

in the reservoir of feed solution maintaining the reaction medium under oxygen saturation,

183 which was checked by an oxygen sensor (Hamilton oxygen probe).

184

(Fig 1)

Blank experiments were carried out without any enzyme or with deactivated beads to ensure that TC self-degradation or adsorption were negligible. To determine the operational stability of the immobilized enzymes, 5 cycles of TC degradation were carried out in the FBR with the same active beads by replacing the TC solution by a fresh one every 24 hours. Between each cycle, the beads were thoroughly rinsed with 50 mL of citrate phosphate buffer (50 mM, pH 7) as well as the reactor. The initial activity was considered as the reference value of 100%.

192 2.4 Analysis

193 <u>2.4.1 TC analysis</u>

TC concentration was measured by high-performance liquid chromatography coupled to
triple-quadrupole mass spectrometry (HPLC–MS). Samples were injected through a MacheryNagel C18 column (50 mm x 2 mm) with a Waters e2695 Separations Module, and the 410
m/z fragment was detected with a Micromass Quattro micro API device. Elution was done
with an isocratic mobile phase (HPLC grade water, 99% formic acid, and acetonitrile
(74.9:0.1:25)) at a flow rate of 2 mL min⁻¹. The amount of TC was quantified with calibration

standards prepared by dilution of the TC stock solution in ultra-pure water.

201 <u>2.4.2 Toxicity assessments</u>

202 Acute toxicity of the TC solution before and after enzymatic treatment was evaluated by the Microtox® bacteria toxicity test based on the general principles described by Bulich (1979) 203 204 and Johnson (2005). The toxicity of samples was determined by measuring the effect on the luminescence of marine bacteria Vibrio fischeri NRRL B-11177 using a Microtox® Model 205 500 Analyzer with MicrotoxOmni software (Modern Water Inc.; United Kingdom). 206 207 The toxicity measurements were carried out for TC solution (1 ppm) before and after 24 hours of laccase treatment (with both free and immobilized laccase). The pH of all the samples was 208 209 adjusted between 6.5 and 7.5 and the toxicity was evaluated after 5 and 15 minutes of contact with the bacteria. Sample toxicity was expressed in percentage of inhibition effect I (%) 210 compared to a control (salt solution) and calculated by equation 3, where E_0 is the 211 212 bioluminescence intensity of the control, and E is the bioluminescence intensity of the sample. 214

214

213
$$I\% = \frac{E_0 - E}{E_0} * 100$$
 (Eq.3)

Each sample was analyzed in triplicate and the average value is reported in the results section.

217 **3. Results**

In the figures of this section, each point presented is a mean value of at least three
experimental results, the bars of error were calculated by the standard deviation of the
measurements.

221 3.1 Stability of immobilized enzymes

222 The stability of free and immobilized enzymes was studied for 3 different pHs (pH 3, 5 and 9) and the results are presented in Figure 2-A. According to this figure, grafting seems to have 223 only a slight effect on laccase stability at pH 5 and 9. Both enzyme forms showed a similar 224 225 decrease of activity in the range of 20-30% after 24 hours of incubation at pH 5 or 9. On the contrary, when the enzymes were incubated at pH 3, the positive effect of grafting is clearly 226 227 observed. After 24 hours of incubation, the residual activity of the free enzyme was equal to $30 \pm 3\%$ while the immobilized laccase still presented a remaining activity of $52 \pm 3\%$. The 228 229 protecting effect of grafting towards deactivation induced by acidic conditions has already been reported in the literature (Wang et al., 2008, Asgher et al., 2017, Guardado et al. 2021). 230 This could be related to the stiffening of the enzyme structure, making it less sensitive to 231 232 denaturing conditions, especially when multi-covalent bonds are involved, but the hydrophilic properties of the gelatin support could also contribute to maintaining a favorable environment 233 234 around the enzymes (Rodriguez et al., 2021).

Enzymatic stability can also be affected by temperature increase; the denaturation effect depends on the enzyme strain and can be mitigated with grafting. Figure 2-B displays the activity evolution of free and immobilized enzymes incubated at 60°C during six hours. A large decrease in enzyme stability is noticed, especially for free enzymes. In this case, relative activity remains at just $40 \pm 2\%$ after 1 hour of incubation at 60°C; this value drops to $12 \pm$ 1% after 6 hours incubation. On the contrary, the immobilization step seems to improve

thermostability of enzymes, as relative activity remains at 38 ± 1% after 6 hours at 60°C. As
previously reported for acidic pH conditions, grafting appears to prevent structural
transformation of enzymes when confronted with relatively high temperatures (Rodriguez et
al., 2021).

245

(Fig 2)

The storage stability of free and immobilized enzymes was studied in citrate phosphate buffer 246 247 (50mM, pH7) at 4°C during 20 days. During that time (results not shown), free laccase activity decreased slightly but continuously to $90 \pm 5\%$ of the initial value at the end of the 248 experiment. In the case of immobilized enzymes, an initial decrease in activity was observed 249 within the first days of storage, probably due to the leakage of enzymes that were not 250 covalently bound to the gelatin beads. Non-covalent interactions (Van der Waals or ionic 251 252 interactions) between laccase and the gelatin chains can also occur during the grafting step. Due to the weakness of these bonds, biocatalysts that have not been covalently grafted can be 253 254 released into the storage solution. However, the final concentration of enzymes in the storage 255 solution was probably too low and did not allow enzyme activity to be measured. The decrease of bead activity could also result in enzyme distortion promoted by further 256 257 interactions between laccases and the support as has been already reported (Rodriguez et al., 258 2021). Nevertheless, after 10 days and until the end of the experiment, the relative activity 259 remained constant and equal to $85 \pm 4\%$ of the initial value. Although no stability increase was observed compared to free enzymes, the covalent grafting method used in this work leads 260 to stable laccase derivatives. 261

262 3.2 TC degradation

Previous studies have demonstrated that TC is degraded by *Trametes versicolor* laccase (De
Cazes). TC was therefore chosen as a PhACs model compound to evaluate the performance of

enzymatic beads in a stirred tank reactor (STR) and an FBR. Firstly, apparent kinetics 265 266 constants of laccase towards TC were determined with free and immobilized enzymes; experiments were carried out with the same enzyme activity (30 U L^{-1}). The V_{max} values 267 (0.001 and 0.002 µmol.L⁻¹.min⁻¹ for free and immobilized enzymes respectively) towards TC 268 were very low compared to those estimated at the same enzyme concentration with ABTS 269 $(0.029 \text{ and } 0.033 \text{ } \mu\text{mol } \text{L}^{-1} \text{ min}^{-1} \text{ for free and immobilized enzymes respectively})$ 270 (Harguindeguy et al, 2021). Since ABTS is known as a specific substrate for laccases, it is not 271 surprising that this enzyme shows a higher reactivity for this substrate. In addition, as the 272 V_{max} values of both enzyme forms were quite similar, it is not possible to conclude whether or 273 274 not immobilization improves the reactivity of the enzymes towards TC. Nevertheless, it appears that immobilized laccase has a slightly higher affinity towards TC than free laccase; 275 the K_m value is lower for the immobilized form (62.8 μ mol L⁻¹) than for the free form (88.2 276 umol L⁻¹). However, it should be noted that this positive effect on laccase affinity was lesser 277 than that observed by De Cazes et al. These authors reported that the K_m of laccase covalently 278 279 grafted to a gelatin layer deposited on a ceramic membrane was 5 times lower than that of free laccase (De Cazes et al., 2014). 280

The first experiments of TC gradations of were carried out with a 20 ppm TC solution in a 281 282 stirred tank reactor with both free and immobilized laccases at the same concentration (100 $U.L^{-1}$); the results obtained are presented in Figure 3. After a similar decrease of TC 283 concentration for both enzyme forms within the first two hours of reaction, the degradation 284 rate slowed down more drastically with free enzymes to reach a final TC removal of $50 \pm 2\%$ 285 286 after 24 hours. When active beads were used, the residual TC concentration was $30 \pm 1\%$. As 287 experiments carried out with inactivated beads resulted on a TC degradation less than 4 %, it can be concluded that TC removal is mainly the result of laccase activity. 288

The immobilization clearly improves the stability of enzymes in operational conditions and 289 290 seems to preserve the biocatalyst from shear stress denaturation. De Cazes et al. have observed similar results with the same laccase of Trametes versicolor covalently grafted to a 291 292 ceramic membrane previously coated with a gelatin layer. They reported that TC degradation increased from $30 \pm 7\%$ up to $56 \pm 8\%$ when the laccases were immobilized (De Cazes et al., 293 294 2014). This improvement of enzyme stability can be related to a favorable environment 295 created by gelatin. More generally, grafting to a biopolymer helps enhance the stability of proteins without altering the flexibility necessary for their catalytic activity (Ma et al., 2018; 296 Asgher et al., 2017). 297

298 The TC degradation tests in the FBR were carried out at 25°C with a 20 ppm TC solution and recirculating flow of 80 mL.min⁻¹ (i.e. 2 U_{mf}). It should be noted that the reaction volume and 299 enzyme concentration were different from those used for experiments done in the stirred-tank 300 reactor (100 mL instead of 10 mL and 60 U L⁻¹ instead of 100 U L⁻¹ respectively). The results 301 obtained are reported in Table 2. As already reported for stirred tank reactor, no adsorption or 302 303 TC auto degradation occurred in the FBR. Blank experiments showed that $98 \pm 2\%$ of the TC remained in the reactor after 24 hours whereas this percentage decreased to $56 \pm 4\%$ when 304 active beads were used. This later value is lower than that observed in the stirred-tank reactor 305 306 $(35 \pm 1\%)$. However, when reaction volume and enzyme concentration are taken into account, the average degradation rate (estimated over 24 hours) is almost 5 times higher in the FBR 307 $(1.59 \text{ mg h}^{-1} \text{ U}_{\text{enzymes}}^{-1})$ than in the stirred tank reactor (0.35 mg h $^{-1} \text{ U}_{\text{enzymes}}^{-1})$. The larger 308 amount of TC in the FBR probably reduced the substrate limiting issues and thus explained 309 the higher degradation rate observed. In addition, even if we are still in laminar conditions 310 (flow velocity = 0.42 cm s^{-1} , Re= 8.5), the hydrodynamic conditions and beads movement 311 observed could also play a role by reducing the boundary layers and thus facilitating the 312 contacts and exchanges between the substrate and the enzymes. 313

(Fig 3)

315

(Table 2)

_ . .

316 3.3 Effect of the recirculation flow rate and air bubbling

To improve the efficiency of TC degradation, the effect of the circulation flow rate and air 317 318 bubbling were also investigated. Experiments were carried out at 3 different recirculation rates (30 mL min⁻¹, 80 mL min⁻¹, and 130 mL min⁻¹, corresponding respectively to the 319 following linear velocities: 0.16, 0.42 and 0.69 cm.s⁻¹ and Re values equal to 3.2, 8.5 and 320 13.8), without or with air bubbling (15 mL min⁻¹). An oxygen sensor was used to verify that 321 the air bubbling maintained the oxygen saturation of the reaction medium in the feed tank. 322 323 The results are reported in figure 4. When no air was added (open symbols), the percentage of degradation increased with the 324 velocity of the fluid. At a velocity of 0.16 cm s⁻¹, which is lower than the minimum fluidizing 325 velocity (0.21 cm s⁻¹, corresponding to 40 mL min⁻¹), the degradation rate of TC is very low 326 and only 7% of TC is degraded in 24 hours. On the contrary, 50% and 70% of degradation 327 were achieved in the reactor when the circulation flow were fixed to 2 and $3.2 U_{mf}$ 328 respectively. Increasing the flow rate allows particles to be separated from each other. At flow 329 rates of 80 and 130 mL min⁻¹, 95% and 100% respectively of the beads were in fluidized 330 331 conditions. The circulation of the fluid around the particles was thus facilitated leading to 332 better exchange between the catalyst and the substrate. On the contrary, in fixed-bed reactors, since all particles are in contact with each other, the access to all active sites is probably 333 334 limited; it is also possible that the fluid does not flow homogeneously through the bed creating inhomogeneous reaction zones inside the reactor. Furthermore, even though the 335 increased flow rate reduces the residence time in the reactor, it enables renewal of the 336 337 substrate close to the active sites of the enzymes that favors degradation. Similar results were

reported by Piao et al. (2019) who observed an increase in bisphenol degradation from 50% to 338 80% in 6 hours by increasing the flow rate from 1.2 mL min⁻¹ to 7.2 mL min⁻¹ in an FBR 339 using porous silica as a laccase immobilization support. 340

Figure 5 shows that at low recirculation rates (30 and 80 mL min⁻¹) the injection of air (closed 341 symbols) considerably improves the degradation rate of TC. Indeed, when the velocity of the 342 solution is lower than minimum fluidizing velocity (0.8 U_{mf}), adding air to the system 343 344 increased the degradation of TC from 7% to 40% in 24 hours. This could be due to the presence of small air bubbles observed around the active beads. These small air bubbles 345 transported from the feed tank by the flow caused expansion of the bed and increased 346 347 agitation around the beads, leading to better exchanges between the catalyst and the substrate. Moreover, the addition of air also allowed a high concentration of oxygen to be maintained 348 near the active sites. The same effect was observed in the FBR for a flow rate of 80 mL.min⁻¹ 349 350 (i.e 2 U_{mf}) where the TC degradation was enhanced from 45% to 70%. Nevertheless, the opposite behavior was obtained at the highest recirculation flow applied (squares : 130 351 352 mL.min⁻¹, 3.2 U_{mf}). In this case, the air bubbling resulted in a decrease of the percentage of TC degradation at 24 hours from 70% (without air bubbling) to 50% (with air bubbling). This 353 354 decrease probably results from the presence of numerous and large air bubbles dragged by the 355 higher flow into the FBR. Indeed, the presence of these air bubbles may mean that some active sites were not continuously in contact with the solution, leading to a decrease of the 356 reaction rate. It is also possible that the high recirculation flow combined with the presence of 357 358 gas bubbles, induced unfavorable shear stresses on the enzymes, leading to their denaturation and thus to a decrease in the efficiency of the degradation process. 359

360

(Fig 4)



3.4 Operational stability of the biocatalyst

The operational stability of the immobilized laccase inside the FBR was evaluated in cyclic 362 experiments carried out during 5 degradation cycles of 24 hours. The experiments were 363 carried out at a laccase concentration of 100 U L⁻¹ and at recirculation rate of 100 mL min⁻¹ 364 without air addition. At the end of each cycle, the reaction medium was removed and the 365 active beads were rinsed with a citrate-phosphate buffer (50 mM, pH7). The reactor was then 366 fed with a fresh TC solution. The evolution of TC concentration during these 5 cycles is 367 reported in the figure 5. It should be noted that no loss of activity was observed during the 368 three first cycles since the evolution of TC concentration was guite similar and a TC 369 degradation of 57% was achieved after 24 hours for these three cycles. Afterwards, a slight 370 decrease in the degradation efficiency was observed during the 4th and 5th cycles and the TC 371 degradation was only 47% at the end of the cycles. Nevertheless, these experiments 372 demonstrate the excellent stability of the immobilized laccases which were still active and 373 374 able to degrade more than 45% of the TC of a 20 ppm TC solution after 120 h of use.

375

(Fig 5)

376 3.5 Toxicity assays

Degradation does not always mean that the toxicity of the effluent is reduced as has already 377 been reported in the literature (Beker et al., 2016). Therefore Microtox assays were performed 378 on TC samples before and after enzymatic treatment with both free and immobilized laccase. 379 As the concentrations used in this project are quite far from those found in real effluents, the 380 degradation tests were carried out in batch with a 1 ppm TC solution and 60 U L⁻¹ of free or 381 immobilized enzymes. The results obtained are presented in figure 7. The initial TC solution 382 showed inhibition effects of 29% and 42% after 5 and 15 minutes of contact respectively. 383 After 24 hours of enzyme treatment, these values almost doubled when free enzymes were 384 385 used (57 and 73% respectively) whereas they decreased to 19 and 25 %, respectively, when

the enzymes were immobilized on the gelatin beads. These results are surprising and seem to 386 387 indicate that the degradation products obtained with free enzymes could be different and more toxic towards Vibrio fischeri than those generated by immobilized enzymes. A difference 388 between the oxidation mechanisms of free and immobilized laccases is possible, resulting in 389 various degradation products with different toxicity levels. However, the difference in toxicity 390 may also be the result of the presence of additives in the commercial enzyme solution. Indeed, 391 Microtox assays performed on a solution of free laccase prepared at 60 U mL⁻¹ showed 392 inhibition effects of 39% and 50% after 5 and 15 minutes of contact respectively. On the 393 contrary, no inhibition was observed when Microtox assays were performed on water placed 394 395 in contact with active beads for 24 hours. The rinsing of the active beads carried out after the grafting step probably allowed the removal of the toxic species present in the commercial 396 397 enzyme solution.

Microtox results show that the treatment of TC solution with immobilized laccases enable a
deacrease in the toxicity of TC towards a specific bacterium (*Vibrio fischeri*). Nevertheless,
other bioassays with other microorganisms must be performed to conclude on the safety of
TC degradation products.

402

(Fig 6)

403 **4.** Conclusion

Laccase of *Trametes versicolor* was covalently immobilized onto gelatin beads to be used in an FBR for degradation of TC. The results obtained showed that TC depletion depends on the operating conditions. High feed flow rates and air bubbling in reaction mixture can prevent mass transfer limitation and/or reaction limitation due to lack of oxygen. A good operational stability was observed for 5 successive degradation cycles of 24 hours. Thus, this grafted biocatalyst combined with an FBR seems to be a promising alternative to assure a high degradation rate as well as possible recyclability of the catalyst. Furthermore, the

biodegradability of gelatin beads is a real advantage for the development of a sustainable
depolluting process. However, the active beads preparation needs further improvement in
order to reduce bead sizes and increase the immobilization yield (only 30% in the present
study). Modeling and simulation studies will be also necessary to scale-up the process for
waste-water treatment and to conclude on its economic viability.

CRediT authorship contribution statement

Marine Harguindeguy: investigation, formal analysis, writing of the original draft; Céline
Pochat-Bohatier: conceptualization, supervision, formal analysis, writing - review & editing;
José Sanchez-Marcano : conceptualization, formal analysis, writing - review & editing; MariePierre Belleville : conceptualization, supervision, formal analysis, writing - review & editing

Conflicts of interest

424 There are no conflicts to declare.

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| 569 | |

| Table 1: A | ctive bead | characte | eristics. | |
|------------|------------|----------|-----------|--|
| | _ | | ~ | |

| Diameter | Density | Specific activity |
|-----------|-----------------------|----------------------|
| (mm) | (kg m ⁻³) | (U g ⁻¹) |
| 2 ± 0.2 | 1112 ± 13 | 0.1±0.01 |

Table 2: Degradation of a TC solution (20 ppm) after 24 hours of reaction in two types of
reactors: stirred tank reactor 10 mL in volume and 100 U L⁻¹ of enzymatic activity and FBR
100 mL and 60 U L⁻¹ respectively (temperature 25°C ; flow rate in FBR: 80 mL min⁻¹).

| Reactor type | | TC residual after 24 hours (%) | Mean degradation rate (mg h ⁻¹ .Uenzymes ⁻¹) |
|--------------|-----------------|--------------------------------------|---------------------------------------------------------------------|
| Stirred tank | Inactived beads | 96 ± 3 | - |
| | Active beads | 35 ± 1 | 0.35 |
| Fluidized | Inactived beads | 98 ± 2 | - |
| bed | Active beads | 56 ± 4 | 1.59 |









Figure 2: Stability of free and immobilized enzymes (A): pH stability of free enzymes (red
dotted line, open symbols) and immobilized enzymes (black continuous line, closed symbols)
during 24 hours at pH 3 (diamonds), pH 5 (circles), and pH 9 (triangles), incubation at 25 °C
in a citrate-phosphate buffer (50 mM); (B) temperature stability of free enzymes (dotted line)
and immobilized enzymes (continuous line) incubated at 60 °C in a citrate phosphate solution
buffer (50 mM, pH 7).





Figure 3: TC (20 ppm) degradation with free (red diamonds – dotted line) or immobilized
(black squares – continuous line) enzymes at 100 U.L⁻¹, at 25°C for 24 hours. Blanks are
represented in open red diamonds (free enzymes) or open black squares (immobilized
enzymes).



Figure 4: Effect of circulation flow rate and oxygen bubbling on TC degradation. (black
triangle curve : 30 mL min⁻¹; red circle curve : 80 mL min⁻¹; blue square curve : 130 mL min⁻¹,
open symbols : without air bubbling, closed symbols : with air bubbling).



Figure 5: Operational stability of immobilized laccase inside the FBR (100mL of 20 ppm TC solution prepared in ultra-pure water pH 6, enzymes concentration of 100 U L-1, recirculation flow rate of 100 mL min-1 at 25°C). After each 24 hours period (dotted red line) the reactor was emptied, washed and refilled with a TC fresh solution and a new experiment started. A total of 5 cycles were carried out (total of 120 h).





Figure 6: Inhibition effects of TC solution before and after laccase treatment (with free and
immobilized enzymes) after 5 minutes (stripes) or 15 minutes (dots) of contact with bacteria
Vibrio fischeri.