

# Potential marine benthic colonisers of offshore wind farms in the English channel: A functional trait-based approach

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1 2	Potential marine benthic colonisers of offshore wind farms in the English Channel: a functional trait-based approach			
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13	Highlights			
14 15 16 17 18 19 20 21 22	<ul> <li>Communities of hard substrates in English Channel differed from those of offshore wind farms and oil and gas platforms.</li> <li>157 taxa have been identified as potential colonisers of Dunkirk's offshore wind farms</li> <li>Functional profiles of potential colonisers were created based on the categories of functional traits represented most in each community.</li> <li>Communities of young offshore wind farms (&lt; 3 y) and hard substrates had similar functional volumes, while that of oil and gas platforms was the smallest.</li> </ul>			

#### 24 Abstract

Offshore wind farms (OWFs) have gained attention as a promising alternative to conventional 25 energy sources. However, their installation and operation may have multiple ecological 26 impacts on the marine environment, including the "reef effect". The reef effect is the 27 colonisation of wind turbines and other artificial substrates by benthic organisms, which has a 28 major impact on marine biodiversity as it changes community assemblages and ecosystem 29 functioning. We conducted a two-step study to predict the reef effect of a future OWF 30 (Dunkirk, northern France). First, we explored similarities between colonisers of existing 31 32 OWFs and those of other hard substrates (oil and gas platforms (O&GP) and hard substrates 33 in the English Channel (HSEC)). We then analysed functional traits to determine a trait profile of potential colonisers of Dunkirk's OWF. Statistical analyses revealed that OWF and 34 O&GP communities were more similar to each other than to that of HSEC. Comparing the 35 three communities revealed that they shared 157 taxa, which could be potential colonisers of 36 Dunkirk's future OWF. The functional profile revealed that OWF colonisers were species 37 38 ranging from 10–100 mm in size, with gonochoric reproduction, pelagic and planktotrophic larvae, a life span of less than 2 years or 5-20 years, were sessile, and were carnivores or 39 suspension feeders. Functional trait analysis revealed that during their intermediate stage of 40 development, OWF benthic communities have a functional richness and diversity (0.68 and 41 0.53, respectively) similar to those of HSEC communities (0.54 and 0.50, respectively). 42 However, based on using O&GP as a long-term view of the colonisation of OWFs, functional 43 richness and diversity could decrease during the climax stage (0.07 and 0.42, respectively). 44

Keywords: offshore wind farms, colonisation; benthic species; biological traits; functional
diversity; English Channel.

#### 47 **1. Introduction**

It is now evident that greenhouse gas emissions from human activities are responsible for global warming. According to the Sixth IPCC report, total net anthropogenic greenhouse gas emissions continued to increase from 2010–2019, as did cumulative net emissions since 1850. Mean annual emissions from 2010–2019 were higher than those in previous decades (IPCC, 2022). To reduce greenhouse gas emissions and meet the demand for energy, renewable energy became a relevant alternative and continues to develop.

In the context of the European Green Deal, which aims to make Europe a climate-54 neutral continent by 2050, the European Union has set several targets to transition energy use, 55 such as having 32.5% of renewable energy in total energy consumption and reducing 56 57 greenhouse gas emissions 40% by 2030. In Europe and around the world, such as China and the United States of America, the development of offshore wind turbines is a response to the 58 need for renewable energy in the energy mix (de Castro et al., 2019). The influence of this 59 type of infrastructure on the marine environment is complex from technical, legal, societal, 60 and environmental points of view for all stakeholders. This renewable source of marine 61 energy must be integrated into ecosystems that are already subject to multiple human 62 disturbances (e.g. climate change, overexploitation of resources, pollution). Installing and 63 operating OWFs (Offshore Wind Farms) can have many ecological impacts on the marine 64 environment, including wildlife populations, habitats, water quality, and nutrient cycles 65 (Degraer et al., 2020; Galparsoro et al., 2022). For example, noise pollution and sea bottom 66 destruction can occur immediately during installation of OWFs, whereas other ecological 67 impacts can occur over longer periods, such as the colonisation of benthic species on new 68 structures, known as the "reef effect" (Wilhelmsson and Malm, 2008). 69

The reef effect is the colonisation of wind turbines and other artificial substrates (e.g.
OWFs, shipwrecks, harbours, oil and gas platforms) by benthic species, which has a large

influence on marine biodiversity (Svane and Petersen, 2001; Petersen and Malm, 2006; 72 Kerckhof et al., 2010; Langhamer, 2012; Dannheim et al., 2020; Degraer et al., 2020). During 73 the operational phase, wind turbine foundations and their scour protection layers create new 74 75 habitats that are quickly colonised, mainly by larvae of benthic species such as gastropods, bivalves, crustaceans, and polychaetes (Wilhelmsson and Malm, 2008; Zupan et al., 2023). 76 Installing an OWF modifies local marine habitats, which implies changes in community 77 assemblage due to the installation of artificial hard substrates on natural soft sediment 78 79 (Kerckhof et al., 2010, 2019; Dannheim et al., 2020). The OWF increases the rapid and abundant colonisation of benthic marine species via larval dispersal, as most of these species 80 have a bentho-pelagic life cycle. At first, the scientific community thought that the new 81 community followed the same temporal dynamics as a colonised natural habitat (Kerckhof et 82 al., 2010, 2019). First, the pioneer stage of early colonisers lasts for 2 years after the turbines 83 84 are installed. The early colonisers are opportunistic, and the species composition depends on the type of foundation (Kerckhof et al., 2019). Next, an intermediate stage, 3-5 years after 85 installation, has high species richness (Kerckhof et al., 2019; Degraer et al., 2020). Finally, 86 the climax stage, 6 years after installation, has biomass dominated by mussels and anemones 87 (Kerckhof et al., 2019; Degraer et al., 2020). But a recent study challenges this theory, 88 89 discussing the unstable nature of the community even after 11 years (Zupan et al., 2023). The difficulty of understanding and anticipating succession at the community level could be 90 overcome by both longer-term and large-scale research, where biological interactions are 91 92 included into the analysis.

Colonisers can influence the diversity of the community (Walker and Chapin, 1987; Connell and Slatyer, 1997). Some species can help increase its species richness, such as the filter-feeding blue mussel (*Mytilus edulis*), which is considered positive engineer that facilitates establishment of other species (Connell and Slatyer 1977; Kerckhof et al., 2019).

Conversely, other species are considered negative engineers, such as the plumose anemone 97 (Metridium senile), which inhibits establishment of other species by predating their free-98 swimming larvae (Connell and Slatyer 1977; Kerckhof et al., 2019). This biofouling provides 99 100 an additional food resource for higher trophic levels within the local marine food web. For example, the foundation of a wind turbine can contain up to 2 t of mussels (Shields and 101 Payne, 2014), which in turn attracts predators. On some OWF, such as a Belgian OWF in the 102 North Sea, the climax community is dominated by the plumose anemone and the blue mussel 103 104 (i.e. 'slimeification' or 'mytilisation', respectively) (Krone et al., 2013; Kerckhof et al., 2019).

Installing an OWF can influence species richness of the local community assemblage 105 106 as well as ecosystem functioning. In the past few decades, the development of trait databases (e.g. FishBase, Polytraits, Arctic Traits) has supported the development of the functional trait-107 based approach in ecology (Martini et al., 2021). Instead of understanding and classifying 108 109 species based only on their taxonomy, it does so based on their functional traits, including morphological (e.g. maximum body size), life-history (e.g. mode of reproduction, life span), 110 physiological (e.g. maximum growth rate, feeding regime) and behavioural (e.g. mobility), 111 112 which determine their role and impact in an ecosystem. Analysing these traits can increase understanding of mechanisms that underlie community assembly, species interactions, and 113 114 ecosystem functioning (Naeem and Wright, 2003; Petchey and Gaston, 2006; Nock et al., 2016; Degen et al., 2018). 115

The functional trait-based approach can address a wide range of ecological issues, including climate change and human impacts. It can be used to compare how functioning differs between assemblages and can highlight initial responses to changes in the functional structure that traditional diversity analysis cannot detect (Mouillot et al., 2013). Installing human-made structures, including OWFs and the structures they are built on (e.g. foundations, cables), may encourage colonisation of species not present in the local soft-

bottom marine habitat that have novel functional traits. This could increase functional 122 123 diversity and change ecosystem functioning by increasing the functional niche space. The functional trait-based approach can help identify which species are colonising these new 124 habitats and how their functional traits may influence the local ecosystem. We considered the 125 future OWF of Dunkirk (northern France) as a case study to predict consequences of 126 installing OWFs in the eastern EC. The Dunkirk OWF (planned for 2028) is particularly 127 128 interesting because it is located in the transition zone between the eastern French EC, where 129 no OWF currently exists (but seven are planned), and the southern North Sea. Our goal was to identify potential colonisers and their functional trait profiles. To this end, we compared 130 benthic communities in the EC on natural (e.g. rocks, gravel, pebbles) and artificial hard 131 substrates (e.g. shipwrecks, harbours) (HSEC) to those on OWFs in the North Sea and Irish 132 Sea (OWF), and on oil and gas platforms in the North Sea (O&GP). This was done using 133 134 traditional community-structure analysis and the functional trait-based approach. The objectives of this study were as follows: 135

- Describe the community structure of the three hard-substrates habitats (i.e. HSEC,
   OWF, and O&GP) and their similarities.
- 1382. Develop a traits profile of the species that colonise each of these three habitats,139 with emphasis on the dominant traits of species that colonise OWFs.
- Compare functional characteristics (e.g. functional diversity, species richness) of
  the three community assemblages to predict the community of the artificial reef
  created by installing the Dunkirk OWF.
- 143
- 144 2. Material and methods
- 145 2.1. Data collection

146 The first step in the study was to compile the literature on colonisers of the three habitats of147 interest (OWF, O&GP, HSEC), and to be able to compare their communities with each other.

Data on species community composition were collected from the EC, Irish Sea and the North 148 Sea since the Dunkirk OWF will be located at the transition between these two biogeographic 149 provinces. The literature review identified 7 OWFs (i.e. C Power, Belwind, Egmond aan Zee, 150 151 Princess Amalia, Horns Rev, Hywind, and North Hoyle) for which lists of colonising species were available, in 13 references and 1 shared data set (Supplementary Material Table A). For 152 O&GP, 28 platforms had lists of colonisers available in 5 references (Supplementary Material 153 Table A). Finally, for HSEC we identified three types of hard substrates - natural hard 154 155 substrates (e.g. rocks, gravel, pebbles), artificial reefs, and shipwrecks - using 8 references (Supplementary Material Table A). A literature review was performed using Google Scholar 156 and ISI Web of Science, where 28 references have been identified with species lists of 157 colonisers. In these references, species were collected and/or identified via scuba diving, scrape 158 sampling, ROV imagery and Rallier du Baty dredge. We selected only macrofauna (> 1 mm) 159 160 for our study. We used the WORMS database to verify that all benthic species cited in the literature were identified by the correct species name. A list of species presence (1) or absence 161 (0) was then created and transformed into a data frame for each habitat. 162



Figure 1. Locations of the three hard substrate habitats (oil and gas platforms, hard substrates
in the English Channel, and offshore wind farms) with a species list found in the literature
(see Supplementary Material Table A for information on the habitats).

163

## 168 2.2. Community structure analysis

We built a list of species for each of the three hard-substrates habitats (OWF, O&GP and 169 HSEC). As the characteristics of these three habitats are potentially different (age, shore 170 distance, depth...), we carried out an analysis of the community structure and we made the 171 172 hypothesis that OWF and O&GP would have closer characteristics compared to HSEC. To this end, we combined the three lists into a single presence/absence data set (Supplementary 173 Material Table B) and calculated a matrix of the Bray-Curtis dissimilarity index using the 174 175 vegdist function of the vegan package of R software (R Core Team, 2022). We then performed a PERMANOVA (999 permutations, Bray-Curtis dissimilarity index (McArdle 176 and Anderson, 2001)) on the dissimilarity matrix using the Adonis function of the vegan 177

package. To illustrate similarities among colonisers of the three habitats, we performed hierarchical ascending classification on the dissimilarity matrix using the Ward method, using the hclust function of the *stats* package of R. Finally, we used a Venn diagram to identify which species the three habitats shared in common.

182

## 2.3. Functional trait-based approach

After comparing the communities structure between the three hard-substrates habitats, we 183 compared the functional traits of the main colonisers of these three habitats in order to 184 explore whereas it follows the same trend. Based on Degen and Faulwetter (2019), seven 185 biological traits were selected: one morphological trait (i.e. maximum body size), four life-186 history traits (i.e. mode of reproduction, type of larval development, mode of larval 187 dispersal, and life span), and two behavioural traits (i.e. mobility and feeding regime of 188 juveniles/adults) (Table 1). Some functional traits give information on the state of the 189 community (successional stage and stability) such as maximum body size and longevity. 190 Other functional traits such as the mode of reproduction, the type of larva and its diet give 191 insights into their potential dispersal strategy and then, colonisation abilities of artificial 192 substrates at a larger scale. Each trait was divided into 3-5 categories, resulting in 27 193 categories (Table 1). The biological traits were obtained from the literature (Supplementary 194 Material Table C) and from databases (e.g. Polytraits, Marlin, Biotic, Encyclopedia of 195 life). We used a fuzzy coding approach to allocate each species to the categories of a given 196 biological trait (Chevenet et al., 1994), attributing 0 or 3 when a species had no affinity or 197 complete affinity, respectively, with a category. A species could have affinities with 198 multiple categories for a given trait, but the sum of all categories always equalled 3. We then 199 200 created a data set that contained the main species in each habitat with their fuzzy coded 201 biological traits. Due to the large number of species, an occurrence index (Ir) was calculated for each habitat: 202

$$Ir = \frac{\text{Number of references in which the species was cited}}{\text{Total number of references}} \times 100$$
(1)

203

A species was selected for functional trait-based analysis when it was cited in at least 30% of the references for the targeted habitat. Overall, 68, 56, and 28 species were included in the biological trait matrix for HSEC, OWF, and O&GP, respectively (Supplementary Material Table D). All trait matrices were at the species level and double-checked for accuracy.

To test for independence between trait categories and habitats (HSEC, OWF, 209 O&GP), the fuzzy code values for each category were summed and then transformed into a 210 percentage. This is repeated for each of the three habitats (Bolam et al., 2017; Zhulay et al., 211 2021). Percentage values were used to test the independence using a Chi<sup>2</sup> test on R. 212 213 Significant differences were assessed using Pearson's Chi<sup>2</sup> post-hoc test in R. In order to obtain a functional profile, we calculated a cumulative score of the functional categories 214 with a threshold greater than or equal to 50%, thus identifying the most represented 215 216 functional categories for each of the three habitats.

217

### 2.4. Functional diversity and structure

To take the functional trait approach a step further, trait diversity and community structure 218 219 were studied using statistical tools. Two functional diversity indices were calculated to 220 better understand the functional characteristics of each habitat's community. First, the mean 221 functional richness (FRic) was estimated as the percentage of the functional volume of each habitat's community derived from a multidimensional functional space (Mouchet et al., 222 223 2010), which revealed how much of the functional niche space was filled by the species present (Degen et al., 2018). The second index was Rao's quadratic entropy (RaoQ), which 224 measures trait dissimilarity and ranges from 0 (i.e. low functional diversity: species in the 225 community have similar functional trait profiles) to 1 (i.e. high functional diversity: species 226 in the community have unique functional trait profiles) (Botta-Dukát, 2005). The two 227

functional indices were calculated using the dbFD function of the *FD* package (Laliberté and
Legendre, 2010) of R. The functional spaces were constructed using principal coordinate
analysis (PCoA), which encapsulated unique trait combinations of species by substrate
(Laliberté and Legendre, 2010), performed on species-by-trait and habitats-by-species data
sets (Supplementary Materials E and F).

To visualise which trait influenced the functioning of each hard substrate, we 233 performed multiple factor analysis (MFA) (Escofier and Pagès, 1994) of species using trait 234 235 data sets in which species were grouped by habitat (Supplementary Material G). MFA is a modified PCoA that can consider categorical data and thus gives equal weight to traits that 236 237 have different numbers of categories (McLean et al., 2021). Associations between traits and substrates were assessed based on the contributions of trait values to the two first axes. MFA 238 was performed using the MFA function of the FactoMineR package of R. Principal 239 240 component analysis was then performed using data from the MFA to obtain a correlation circle of the functional traits using the fviz\_contrib function of the *factoextra* package of R. 241 242 Finally, the functional space of each habitat was plotted using the coordinates of the species 243 in the overall functional space provided by the MFA.

The overlap of functional polygons between the three hard-substrates habitats was 244 calculated using a Monte Carlo method, usually performed to study niche regions (Swanson 245 et al., 2015). The Bayesian probabilistic method estimates size and pairwise polygon overlap 246 without considering sample size (e.g. number of main species per habitats in our study). 247 Overlaps were estimated from the species coordinates of the two first axes of the PCoA for 248 each of the main species. Average overlap between functional polygons with 95% Bayesian 249 intervals based on 10,000 iterations in bivariate dimension. The analyses were performed 250 using the *nicheROVER* package (Lysy et al., 2021) of R. 251

252

The functional diversity indices and MFA visualisation were repeated on trait

- 253 matrices by removing species with direct development (trait L3, Table 1) due to their
- absence from the main species list of O&GP colonisers (but not from the entire species list)
- but not in HSEC and OWF species' lists.

Table 1. The biological traits and their categories used in the functional trait-based approachto identify the main species colonising hard substrates

<b>Biological trait</b>	Category	Definition		
Maximum body	<b>S</b> 1	Small (< 10 mm)		
size	<b>S</b> 2	Small-medium (10-50 mm)		
	<b>S</b> 3	Medium (50-100 mm)		
	<b>S</b> 4	Medium-large (100-300 mm)		
	S5	Large (> 300 mm)		
Reproduction	<b>R</b> 1	Asexual (e.g. budding, fission)		
	R2	Gonochoric		
	R3	Hermaphrodite		
Larval	L1	Pelagic		
development	L2	Benthic		
	L3	Direct (no larval stage)		
Larval diet	LD1	Planktotrophic		
	LD2	Lecithotrophic		
	LD3	Direct (no larval stage)		
Life span	A1	Short (< 2 years)		
	A2	Short-medium (2-5 years)		
	A3	Medium-long (5-20 years)		
	A4	Long (> 20 years)		
Mobility	MO1	Sessile		
	MO2	Burrower		
	MO3	Crawler		
	MO4	Swimmer		
Feeding habit FH1		Deposit feeder		
	FH2	Suspension feeder		
	FH3	Herbivore		
	FH4	Opportunist/Scavenger		
	FH5	Carnivore		

## 259 **3. Results**

- 260 3.1. Community structure analysis of hard substrates
- 261 The literature review of colonisers identified 384, 265, and 559 species for OWF, O&GP, and
- 262 HSEC, respectively (Supplementary Material Table B). A total of 14 phyla were represented,
- of which 4 dominated (i.e.  $\geq$  58% of species in each habitat): Arthropoda, Annelida,
- 264 Mollusca, and Cnidaria (Fig. 2).



Figure 2. Percentage of each species colonising the three habitats, by phylum. (HSEC: hard substrates in the English Channel, O&GP: oil and gas platforms, and OWF: offshore wind farms)

The PERMANOVA revealed that habitat had a significant (P < 0.05) effect on species communities (Supplementary Material H). The hierarchical classification showed that the community of HSEC (Height = 0.741) differed from those of O&GP and OWF (Height = 0.576), which were more similar to each other.

The Venn diagram indicated 84 species in common among all three habitats, with OWF and HSEC having the most (n = 73) and HSEC and O&GP having the fewest (n = 29) (Fig. 3).



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Figure 3. Venn diagram of the taxa shared among the substrates for all species
inventoried (HSEC: hard substrates in the English Channel, O&GP: oil and gas platforms,
and OWF: offshore wind farms)

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## 281 3.2. Biological trait profiles of hard substrates

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When categories of each of the seven biological traits were compared among the habitats, at least one of the categories differed among habitats, except for the mobility trait (P > 0.05) (Fig. 5, Supplementary Material I). For the reproduction trait, category hermaphrodite was represented less in O&GP than in the other two habitats. Benthic, direct larval development, and the feeding mode associated to direct development mode were not represented in O&GP.
The percentage of planktotrophic larvae was higher in OWF and O&GP than in HSEC. For
the life span trait, category < 2 years dominated in OWF but not in the other two habitats. Life</li>
span category 2–5 years was absent in O&GP. For the diet of the colonisers, category
suspension feeder was represented less in O&GP than in the other two habitats, and category
carnivore was represented less in HSEC than in the other two habitats.



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Figure 5. Percentage of each category of the seven biological traits selected (A) maximum
body size, B) reproduction, C) larval development, D) larval diet, E) life span, F) mobility,
and G) feeding habit) for the three habitats (HSEC: hard substrates in the English Channel,
O&GP: oil and gas platforms, OWF: offshore wind farms)

The percentage contributions of the most represented categories of the seven biological traits varied among habitats (Supplementary Material Table J). For HSEC and O&GP, the categories represented most for maximum body size were 10–50 mm and 100–300 mm.

303	For OWF, the categories represented most were 10-50 mm and 50-100 mm. For
304	reproduction, category gonochoric dominated in HSEC, O&GP, and OWF For HSEC,
305	O&GP, and OWF, larval development was mainly pelagic and larval diet was mainly
306	planktotrophic, however, lecithotrophic larva was also highly represented in HSEC. For life
307	span, the most represented categories were 2-5 years and 5-20 years for HSEC, 5-20
308	years and > 20 years for O&GP, and < 2 years and 5–20 years for OWF. For mobility,
309	sessile organisms dominated in HSEC, O&GP, and OWF. For feeding habit, for HSEC and
310	OWF, colonisers were either suspension feeders or carnivores, whereas for O&GP,
311	carnivores dominated
312	
313	3.3 Functional diversity and structure
314	FRic and RaoQ were highest for HSEC (0.68 and 0.53, respectively) and lowest for O&GP
315	(0.07 and 0.42, respectively).

Table 2. Functional diversity indices (functional richness (FRic) and Rao's quadratic
entropy (RaoQ)) based on the biological traits of the main species (n = number of species).
Bold indicates the highest indicator values.

Habitat	n	FRic	RaoQ
Hard substrates in the English Channel	68	0.68	0.53
Offshore wind farms	56	0.54	0.50
Oil and gas platforms	28	0.07	0.42

320

In the MFA, the first two axes of the correlation circle explained 27.7% (14.8% and 12.9%, respectively) of the variability in the data set (Fig. 6). Eight categories had a high contribution: organisms with pelagic larval development, direct development, a planktotrophic larval diet, and sessile and crawling adult organisms with a life span less than 2 years or 5–20 years.



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Figure 6. Correlation circle resulting from the multifactorial analysis, showing the
 percentage of contribution (contrib) of all 27 categories of the 7 functional traits to the two
 first axes (see Table 1 for definitions of the category codes)

331 Species were grouped by their biological traits within the functional space created by the MFA. Polygons were drawn for each habitat by connecting the outermost species in the 332 scatterplot (Fig. 7). According to the FRic, the functional volumes of HSEC and OWF were 333 relatively similar in size, whereas that of O&GP was smaller. The larger functional volumes of 334 335 HSEC and OWF were explained by species with direct larval development, such as, for 336 HSEC, the molluscs Buccinum undatum, Nucella lapillus, Calliostoma zizyphinum, and 337 *Musculus discors*; the crustaceans *Melita palmata* and *Phtisica marina*; and the echinoderm Amphipholis squamata; and for OWF, the crustaceans Idotea pelagica, Monocorophium 338 339 sextonae, Stenothe marina, S. valida, P. marina, Jassa herdmani, J. marmorata, and



342

**Figure 7.** Functional polygons of the three habitats, top left is the overlap of hard substrates in the English Channel (red), oil and gas platforms (green), and offshore wind farms (blue), species are represented by points. Dimensions 1 and 2 explained 14.8% and 12.9% of the variance, respectively.

347 Overlap analyses revealed that HSEC functional polygon is covered by 35.1% and

348 91.6% of the O&GP and OWF functional polygons respectively. O&GP functional

polygon is covered by 96.9% and 94.5% of HSEC and OWF functional polygons

- respectively. Finally, OWF functional polygon is covered by 85.3% of HSEC and 27.1%
- of O&GP functional polygons respectively (Supplementary materials L).

## 352 **4. Discussion**

353 Installation of OWFs is known to influence marine ecosystems, such as by for instance

causing the reef effect (Wilhelmsson and Malm, 2008) or stepping-stone effect (De Mesel et

al., 2015), or by rendering habitats more complex (i.e. transition from soft substrates to hard 355 356 substrates) (Degraer et al., 2020). As an example to anticipating the impact of colonisation by benthic organisms on artificial man-made structures, we applied our method to the English 357 358 Channel marine ecosystem, in relation to the French energy transition objectives, that plans to install seven OWFs within the EC in the next decade. This requires predicting the influence of 359 360 OWFs on marine ecosystems, particularly on benthic communities in the EC. The present study focused specifically on the reef effect of benthic species and showed that community 361 structure was more similar between the two human-made structures (OWFs and O&GP), 362 whereas functional diversity was more similar between the communities of OWF and HSEC. 363 Beyond the French case, our results could be of interest for the whole English Channel. 364

365

366 4.1 Comparison of communities in the three habitats

367 The list of colonisers highlighted similarities and differences in community structure among the three habitats: the species composition of HSEC differed from those of OWF and O&GP, 368 which were more similar to each other. However, the community structure analysis was based 369 370 only on the presence/absence of species in each habitat, and thus provided a qualitative overview of the composition of assemblages. Natural and artificial substrates have been 371 compared (Coolen et al., 2020; Pinto et al., 2021) to identify parameters that can explain 372 differences in their communities. Several hypotheses were explored, such as the substrate's 373 type (e.g. rock vs. steel), complexity (e.g. surface area, rocks, crevices (Atilla et al., 2005; 374 Coolen et al., 2015)). Zupan et al. (2023) showed that the age and type of substrate had 375 376 significant effects on species composition. The type of substrate (concrete, steel and complexity) is a well-known factor in the marine environment according to multiple 377 colonisation experiments with different types of substrates that influenced the assemblage of 378 colonisers (Taniguchi and Tokeshi, 2004; Burt et al., 2009; Gaudron et al., 2010; Osman, 379

2015). The type of substrate could explain the differences in species composition of HSEC vs.
those of OWF and O&GP as HSEC samples are often located on seabed, while OWF and
O&GP samples are usually taken at mid-water depth. A recent study showed the impact of
depth on marine species richness and diversity in benthic colonisers (Coolen et al., 2022).

Another hypothesis is that biofilm that develops on hard substrates can influence the 384 substrates that larvae choose during colonisation (Hadfield, 2010). It can also influence 385 386 metamorphosis during the transition from larval to post-larval stages (Shitema, 2012). Thus, it would be interesting to study the biofilm composition of hard substrates to identify whether it 387 can explain community differences. Tan et al. (2015) compared biofilm composition on 388 389 natural (i.e. rocky shores) and artificial (i.e. seawall) substrates, and observed no difference in biofilm composition, but did observe a difference in the relative abundance of biofilm 390 391 organisms. Other environmental factors could explain the differences in species composition 392 of HSEC vs. those of OWF and O&GP. For example, depth is known to influence species richness on hard substrates, with maximum richness from 15 or 25 m to 75 m in depth (Clarke 393 and Lidgard., 2000; Van der Stap et al., 2016; Coolen et al., 2020). Distance to shore can also 394 impact species diversity (Van der Stap et al., 2016; Reeves et al., 2018). However, the spatial 395 aspect should be considered, as OWF and O&GP were not present in the EC during our study. 396 397 Although species assemblages of OWF and O&GP (mainly in the North Sea) differed from that of HSEC, 84 taxa were common to all three habitats. These species are likely to colonise 398 the future Dunkirk OWF and, more broadly, all OWFs installed in the EC in the near future. 399 Interestingly, 73 taxa were common to HSEC and OWF, which increases the number of 400 401 potential colonisers of the Dunkirk OWF to 157. From a broader perspective, this means that a fraction of benthic organisms already present in the area where the offshore parks are 402 located will potentially colonise them. The use of the functional profile of colonisers could 403 then allow identifying this fraction of colonisers thanks to their functional traits. But this 404

method is likely to overestimate the colonisers potential, because the number of taxa observed 405 in the literature was always less than 157 (e.g. 78 species by Kerckhof et al. (2012), 63 by 406 Kerckhof et al. (2010), 121 by Karlsson et al. (2021), 85 by Vanagt et al. (2013), 70 by 407 Leonhard and Pedersen (2004), and 55 by Bouma and Lengkeek (2013)). In this study, the 408 case of the future OWF of Dunkirk is particularly interesting because it is located in the 409 transition zone between the eastern French EC, where no OWF currently exists (but seven are 410 planned), and the southern North Sea where numerous OWF are already implemented. The 411 methodology and the results obtained for the French case study could be of interest to any 412 other future OWF. 413

414

#### 415 4.2 Biological traits of species on hard substrates

Over the past decade, the functional trait-based approach has been widely adopted and seen as 416 a complementary to species-based approaches (Degen and Faulwetter, 2019). Biological trait 417 analysis of colonisers of the three habitats highlighted that gonochoric reproduction, pelagic 418 419 larval dispersal, and sessile juveniles/adults were the categories represented most in each of the three habitats. However, dominant traits differed among the habitats, including maximum 420 body size, larval diet, life span, and adult feeding. Colonisers were smallest for OWF (i.e. 421 maximum body size of 50-100 mm), mostly intermediate for HSEC (i.e. 10-50 or 100-300 422 mm), and largest for O&GP (i.e. 100-300 or 10-50 mm). A similar trend was observed for 423 life span, which was shortest for OWF colonisers (i.e. < 2 or 5–20 years), intermediate for 424 HSEC (i.e. 2-5 or 5-20 years), and longest for O&GP colonisers (i.e. 5-20 or > 20 years). 425 According to the literature, the maximum body size and life span of colonisers provide 426 427 information on the stage of the community (Bolam and Eggleton, 2014; Bremner, 2008; Cain et al., 2014; Costello et al., 2015; Emmerson, 2012; Micheli and Halpern, 2005; Norkko et al., 428 429 2013; van der Linden et al., 2016). As OWF colonisers were smaller and had a shorter life

span, the OWF community could be considered fast growing, highly productive, and less 430 sensitive to disturbance, which indicates a younger community. Due to their larger body size 431 and longer life span, O&GP colonisers could be considered slow growing, less productive, 432 433 and more sensitive to disturbance. Biological traits of the HSEC community could be intermediate because the substrates are of different types (i.e. intertidal rocky shore, pebbles, 434 shipwrecks, concrete) and ages (i.e. 4 to ca. 80 years). The first theory about benthic 435 436 community succession on offshore marine structures was that it follows the same rules as that in other ecosystems, with pioneer, intermediate, and climax stages (Kerckhof et al., 2010). 437 Degraer et al. (2020) identified these three stages of succession on wind turbines on the 438 Belgian coast in the North Sea, with the pioneer stage from 0–2 years, the intermediate stage 439 from 3–5 years, and the climax stage from 6 years after installation. But a recent study (Zupan 440 et al., 2023) showed that after 11 years communities are still changing and two other theories 441 442 are put forward, first several stable or climax communities theory (Sutherland, 1974) and the repeated successional pathway of the cycle climax theory (Oshurkov, 1992). Therefore, the 443 444 age of the community should be considered when using a functional trait profile. Pioneer species are generally small due to harsh abiotic conditions and relatively low nutrient 445 availability (MacArthur and Wilson, 1967; Thornton, 2007; Dykman et al., 2021), whereas 446 species at the climax stage are relatively larger, with slower growth, and influence the 447 establishment of other species (Hirata, 1986; Tilman, 1988, Teixido et al., 2004). Therefore, 448 according to their functional profiles, OWF assemblages in the present study were in the 449 transition between pioneer and intermediate stages of succession, whereas O&GP 450 assemblages were at the climax stage. These observations for a mean age of the OWFs (3 451 years) and O&GP (21 years) during sampling agree with the succession observed by Degraer 452 453 et al. (2020). Our results suggest that O&GP can be used as long-term predictions of

454 communities of OWF colonisers at the climax stage, information during the younger stages455 are needed to fully support this statement.

Our results also revealed relevant information on feeding habits. In the study of 456 Degraer et al. (2020), offshore hard substrates were generally dominated by suspension 457 feeders, but our study showed that carnivores dominated in OWF and O&GP. This could be 458 due to having used a qualitative approach instead of a quantitative approach based on biomass 459 460 or abundance, which were not available for this study. The main functional traits of species in a targeted habitat could be used to predict colonisation of human structures. More concretely, 461 the functional profile of OWF colonisers defined in the present study could be used to identify 462 463 which potential colonisers already living in the EC could colonise the seven future OWFs based on their functional traits. 464

465

466 4.3 Functional diversity and structure

The functional diversity indices showed greater functional similarity between the assemblages 467 of OWF and HSEC, which differed from those of O&GP. HSEC had the highest functional 468 469 diversity indices, but those of OWF were higher than those of O&GP, which had the smallest functional volume (perhaps mainly due to the absence of species with direct larval 470 471 development within the main species list) and thus seemed to be an unstable habitat (Tilman, 2001). For HSEC and OWF, arthropods and molluscs were the main species with direct larval 472 development. Although not on the list of main species used for functional trait analysis, 473 arthropods with direct larval development were present in O&GP (e.g. in Amphipoda, 474 Decapoda, and Isopoda Orders), but they were not identified to the species level. 475

Taxonomic resolution can influence the results of functional analysis of a community.
According to Poff et al. (2006), family-level resolution can result in loss of information,
particularly for families with genera that have multiple trait characteristics. Thus, genus-level

resolution could detect biological or ecological responses more accurately. While the species 479 480 level is the highest resolution and can detect the smallest biological or ecological responses, it is also sensitive to redundancy due to congeneric species that are functionally equivalent and 481 thus considered ecologically equivalent (Wiggins and Mackay, 1978). This most accurate 482 level was able to detect subtler biological and ecological responses than using family or 483 genus, which are less accurate (Poff et al., 2006). However, identifying taxa to the species 484 485 level can also be negative because it may result in excluding many individuals that cannot be identified, as in our study, which may have biased comparison of the ecological functioning 486 of the three habitats. Repeating the biological trait analyses in this study after removing 487 488 species with a direct development trait showed the same trends as the original biological trait analyses (i.e. smaller functional indices and volumes for O&GP than for the two other 489 490 habitats) (Supplementary Material K).

491 Overall, HSEC seems to be the most stable and resilient of the three habitats studied. At the mean age of OWF (i.e. 3 years) our study indicates a high functional diversity 492 493 supported by the functional indexes values and high overlap values of the functional polygons. OWF had a functional diversity similar to those of the communities already present 494 in HSEC. As Coolen et al. (2022) and Zupan et al. (2023) demonstrated that the greatest 495 specific richness and diversity were reached around 6 years, the maximum functional 496 similarity between OWF and HSEC could be reached at this period. Conversely, using O&GP 497 as a long-term reference indicates that community functional diversity may decrease on 498 OWFs during years. This agrees with Kerckhof et al. (2019), who observed that mussels and 499 500 anemones dominated long-term communities colonising wind turbines. Because Zupan et al. (2023) recently challenged the idea that OWFs will tend to a stable climax stage as their 501 502 longest monitoring program, the current state of knowledge requires further studies on longterm colonisation of artificial substrates. 503

## 505

## 506 **5.** Conclusion

Installing an OWF creates a new habitat that benthic organisms can colonise. Focusing on 507 functional traits helped to identify species responses to environmental disturbances and their 508 effects on ecosystems (Carmona et al., 2016), which is relevant for studying the influence of 509 510 OWFs on ecosystems. Our case study indicated how this new artificial substrate could 511 influence benthic communities and their functioning. As functional traits can be used across species and geographic ranges (Bernhardt-Römermann et al., 2011), we compared different 512 communities that form many habitats (i.e. OWF, O&GP, and HSEC) in different areas (i.e. 513 514 EC, North Sea, and Irish Sea). Based on functional trait profiles, we identified 157 taxa that 515 could colonise future OWFs built in the EC. Qualitative analysis of OWFs in the North Sea showed that their functional diversity is relatively similar to that of hard substrates already 516 517 present in the EC during the intermediate successional stage of the community. It would be interesting to replicate this approach using abundance and biomass data and to apply it to all 518 519 of the species sampled to quantify the functional diversity of OWFs.

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526

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## 534 **References**

53Atilla, N., Fleeger, J.W., Finelli, C.M., 2005. Effects of habitat complexity and hydrodynamics on the

- abundance and diversity of small invertebrates colonizing artificial substrates. Journal of Marine
   Research 63, 1151–1172. <u>https://doi.org/10.1357/002224005775247580</u>
- 53Bernhardt-Römermann, M., Gray, A., Vanbergen, A.J., Bergès, L., Bohner, A., Brooker, R.W., De Bruyn,
- L., De Cinti, B., Dirnböck, T., Grandin, U., Hester, A.J., Kanka, R., Klotz, S., Loucougaray, G.,
- 540 Lundin, L., Matteucci, G., Mészáros, I., Oláh, V., Preda, E., Prévosto, B., Pykälä, J., Schmidt, W.,
- 541 Taylor, M.E., Vadineanu, A., Waldmann, T., Stadler, J., 2011. Functional traits and local environment
- 542 predict vegetation responses to disturbance: a pan-European multi-site experiment. Journal of Ecology
- 543 99, 777–787. <u>https://doi.org/10.1111/j.1365-2745.2011.01794.x</u>

54Bolam, S.G., Eggleton, J.D., 2014. Macrofaunal production and biological traits: Spatial relationships along

- the UK continental shelf. Journal of Sea Research 88, 47–58.
- 546 <u>https://doi.org/10.1016/j.seares.2014.01.001</u>
- 54Bolam, S.G., Garcia, C., Eggleton, J., Kenny, A.J., Buhl-Mortensen, L., Gonzalez-Mirelis, G., van Kooten,
- 548 T., Dinesen, G., Hansen, J., Hiddink, J.G., Sciberras, M., Smith, C., Papadopoulou, N., Gumus, A.,
- 549 Van Hoey, G., Eigaard, O.R., Bastardie, F., Rijnsdorp, A.D., 2017. Differences in biological traits
- composition of benthic assemblages between unimpacted habitats. Marine Environmental Research
   126, 1–13. https://doi.org/10.1016/j.marenvres.2017.01.004
- 55**B**otta- Dukát, Z., 2005. Rao's quadratic entropy as a measure of functional diversity based on multiple traits. Journal of Vegetation Science 16, 533–540. https://doi.org/10.1111/j.1654-1103.2005.tb02393.x
- 55**B**ouma, S., Lengkeek, W., 2013. Benthic communities on hard substrates within the first Dutch offshore wind farm (OWEZ). Nederlandse Faunistische Mededelingen 41, 59–67.
- 55Bremner, J., 2008. Species' traits and ecological functioning in marine conservation and management.
- 557 Journal of Experimental Marine Biology and Ecology 366, 37–47.
- 558 <u>https://doi.org/10.1016/j.jembe.2008.07.007</u>
- 55Brierley, A.S., Kingsford, M.J., 2009. Impacts of Climate Change on Marine Organisms and Ecosystems.
  Current Biology 19, R602–R614. <u>https://doi.org/10.1016/j.cub.2009.05.046</u>
- 56Burt, J., Bartholomew, A., Bauman, A., Saif, A., Sale, P.F., 2009. Coral recruitment and early benthic
- 562 community development on several materials used in the construction of artificial reefs and
- 563 breakwaters. Journal of Experimental Marine Biology and Ecology 373, 72–78.
- 564 <u>https://doi.org/10.1016/j.jembe.2009.03.009</u>
- 566 Sain, William D. Bowman, Sally D. Hacker, n.d. Ecology, Third Edition.

566 armona, C.P., de Bello, F., Mason, N.W.H., Lepš, J., 2016. Traits Without Borders: Integrating Functional

- 567 Diversity Across Scales. Trends in Ecology & Evolution 31, 382–394.
- 568 <u>https://doi.org/10.1016/j.tree.2016.02.003</u>

56@hevenet, Fran., Dolédec, S., Chessel, D., 2006. A fuzzy coding approach for the analysis of long-term
ecological data. Freshwater Biology 31, 295–309. <u>https://doi.org/10.1111/j.1365-2427.1994.tb01742.x</u>

57Clarke, A., Lidgard, S., 2000. Spatial patterns of diversity in the sea: bryozoan species richness in the North
Atlantic. Journal of Animal Ecology 69, 799–814. <u>https://doi.org/10.1046/j.1365-2656.2000.00440.x</u>

576 onnell, J.H., Slatyer, R.O., 1977. Mechanisms of Succession in Natural Communities and Their Role inCommunity Stability and Organization. The American Naturalist 111, 1119–1144.

576 oolen, J.W.P., Bos, O.G., Glorius, S., Lengkeek, W., Cuperus, J., van der Weide, B., Agüera, A., 2015.

- 576 Reefs, sand and reef-like sand: A comparison of the benthic biodiversity of habitats in the Dutch
- 577 Borkum Reef Grounds. Journal of Sea Research 103, 84–92.
- 578 <u>https://doi.org/10.1016/j.seares.2015.06.010</u>

57@oolen, J.W.P., van der Weide, B., Cuperus, J., Blomberg, M., Van Moorsel, G.W.N.M., Faasse, M.A.,

- 580 Bos, O.G., Degraer, S., Lindeboom, H.J., 2020. Benthic biodiversity on old platforms, young wind
- farms, and rocky reefs. ICES Journal of Marine Science 77, 1250–1265.
- 582 <u>https://doi.org/10.1093/icesjms/fsy092</u>

586 ostello, M.J., Claus, S., Dekeyzer, S., Vandepitte, L., Tuama, É.Ó., Lear, D., Tyler-Walters, H., 2015.
Biological and ecological traits of marine species. PeerJ 3, e1201. <u>https://doi.org/10.7717/peerj.1201</u>

58Dannheim, J., Bergström, L., Birchenough, S.N.R., Brzana, R., Boon, A.R., Coolen, J.W.P., Dauvin, J.-C.,

- 586 De Mesel, I., Derweduwen, J., Gill, A.B., Hutchison, Z.L., Jackson, A.C., Janas, U., Martin, G.,
- 587 Raoux, A., Reubens, J., Rostin, L., Vanaverbeke, J., Wilding, T.A., Wilhelmsson, D., Degraer, S.,
- 588 2020. Benthic effects of offshore renewables: identification of knowledge gaps and urgently needed

research. ICES Journal of Marine Science 77, 1092–1108. <u>https://doi.org/10.1093/icesjms/fsz018</u>

59De Mesel, I., Kerckhof, F., Norro, A., Rumes, B., Degraer, S., 2015. Succession and seasonal dynamics of the epifauna community on offshore wind farm foundations and their role as stepping stones for non-

592 indigenous species. Hydrobiologia 756, 37–50. <u>https://doi.org/10.1007/s10750-014-2157-1</u>

59Degen, R., Aune, M., Bluhm, B.A., Cassidy, C., Kędra, M., Kraan, C., Vandepitte, L., Włodarska-

- 594 Kowalczuk, M., Zhulay, I., Albano, P.G., Bremner, J., Grebmeier, J.M., Link, H., Morata, N.,
- 595 Nordström, M.C., Shojaei, M.G., Sutton, L., Zuschin, M., 2018. Trait-based approaches in rapidly
- changing ecosystems: A roadmap to the future polar oceans. Ecological Indicators 91, 722–736.
- 597 <u>https://doi.org/10.1016/j.ecolind.2018.04.050</u>

59Begen, R., Faulwetter, S., 2019. The Arctic Traits Database – a repository of Arctic benthic invertebrate
traits. Earth System Science Data 11, 301–322. <u>https://doi.org/10.5194/essd-11-301-2019</u>

60Degraer, S., Carey, D., Coolen, J., Hutchison, Z., Kerckhof, F., Rumes, B., Vanaverbeke, J., 2020. Offshore
Wind Farm Artificial Reefs Affect Ecosystem Structure and Functioning: A Synthesis. Oceanography
33, 48–57. <u>https://doi.org/10.5670/oceanog.2020.405</u>

60Dykman, L.N., Beaulieu, S.E., Mills, S.W., Solow, A.R., Mullineaux, L.S., 2021. Functional traits provide
new insight into recovery and succession at deep- sea hydrothermal vents. Ecology 102(8), e03418.
https://doi.org/10.1002/ecy.3418

60Emmerson, M.C., 2012. The importance of body size, abundance, and food-web structure for ecosystem607 functioning, in: Solan, M., Aspden, R.J., Paterson, D.M. (Eds.), Marine Biodiversity and Ecosystem

Functioning: Frameworks, Methodologies, and Integration. Oxford University Press, pp. 85–100.
 https://doi.org/10.1093/acprof:oso/9780199642250.003.0007

61Escofier, B., Pagès, J., 1994. Multiple factor analysis (AFMULT package). Computational Statistics & Data
611 Analysis 18, 121–140. <u>https://doi.org/10.1016/0167-9473(94)90135-X</u>

61Galparsoro, I., Menchaca, I., Garmendia, J.M., Borja, Á., Maldonado, A.D., Iglesias, G., Bald, J., 2022.

613 Reviewing the ecological impacts of offshore wind farms. npj Ocean Sustain 1, 1–8.

614 https://doi.org/10.1038/s44183-022-00003-5

61Gaudron, S.M., Pradillon, F., Pailleret, M., Duperron, S., Le Bris, N., Gaill, F., 2010. Colonization of

organic substrates deployed in deep-sea reducing habitats by symbiotic species and associated fauna.

617 Marine Environmental Research 70, 1–12. <u>https://doi.org/10.1016/j.marenvres.2010.02.002</u>

61Bladfield, M.G., 2010. Biofilms and Marine Invertebrate Larvae: What Bacteria Produce That Larvae Use to 619 Choose Settlement Sites. Annual Review of Marine Science 3, 453–470.

620 <u>https://doi.org/10.1146/annurev-marine-120709-142753</u>

62Hirata, T., 1986. Succession of sessile organisms on experimental plates immersed in Nabeta Bay, Izu

622 Peninsula, Japan. I. Algal succession. Marine Ecology Progress Series 34, 51–61.

623 <u>https://doi.org/10.3354/meps034051</u>

624PCC, 2022. Climate Change 2022: Impacts, Adaptation and Vulnerability.

62Karlsson, R., Tivefälth, M., Duranović, I., Martinsson, S., Kjølhamar, A., Murvoll, K.M., 2022. Artificial
hard-substrate colonisation in the offshore Hywind Scotland Pilot Park. Wind Energy Science 7, 801–
814. <a href="https://doi.org/10.5194/wes-7-801-2022">https://doi.org/10.5194/wes-7-801-2022</a>

62K erckhof, F., Rumes, B., Norro, A., Houziaux, J. S., & Degraer, S. (2012). A comparison of the first stages

of biofouling in two offshore wind farms in the Belgian part of the North Sea. Offshore wind farms in

- 630 the Belgian Part of the North Sea: Heading for an understanding of environmental impacts. Royal
- 631 Belgian Institute of Natural Sciences, Management Unit of the North Sea Mathematical Models,
- 632 Marine ecosystem management unit, Brussels, 17-39.

63K erckhof, F., Rumes, B., Degraer, S., 2019. About "Mytilisation" and "Slimeification": a decade of

succession of the fouling assemblages on wind turbines off the Belgian coast. Memoirs on the Marine

- Environment Environmental impacts of offshore wind farms in the Belgian part of the North Sea:
- 636 making a decade of monitoring, research and innovation. Memoirs on the Marine Environment., 73–
- 637 84.

63Kerckhof, F., Rumes, B., Jacques, T., Degraer, S., Norro, A., 2010. Early development of the subtidal

- 639 marine biofouling on a concrete offshore windmill foundation on the Thornton Bank (southern North
- 640 Sea): first monitoring results. Underwater Technology 29, 137–149. <u>https://doi.org/10.3723/ut.29.137</u>

64Krone, R., Gutow, L., Joschko, T.J., Schröder, A., 2013. Epifauna dynamics at an offshore foundation –

Implications of future wind power farming in the North Sea. Marine Environmental Research 85, 1–
<u>https://doi.org/10.1016/j.marenvres.2012.12.004</u>

64<sup>1</sup> aliberté, E., Legendre, P., 2010. A distance-based framework for measuring functional diversity from
multiple traits. Ecology 91, 299–305. <u>https://doi.org/10.1890/08-2244.1</u>

64Eanghamer, O., 2012. Artificial Reef Effect in relation to Offshore Renewable Energy Conversion: State of
the Art. The Scientific World Journal 2012, 1–8. <u>https://doi.org/10.1100/2012/386713</u>

64Beonhard, S.B., Pedersen, J., 2004. Hard bottom substrate monitoring Horns Rev offshore wind farm.649 Annual status report 2004 79.

65MacArthur and Wilson, 1967. The Theory of Island Biogeography, Princeton Landmarks in Biology. ed.651 Princeton University Press.

65Martini, S., Larras, F., Boyé, A., Faure, E., Aberle, N., Archambault, P., Bacouillard, L., Beisner, B.E.,

Bittner, L., Castella, E., Danger, M., Gauthier, O., Karp- Boss, L., Lombard, F., Maps, F., Stemmann,

L., Thiébaut, E., Usseglio- Polatera, P., Vogt, M., Laviale, M., Ayata, S., 2021. Functional

trait- based approaches as a common framework for aquatic ecologists. Limnology and Oceanography

656 66, 965–994. <u>https://doi.org/10.1002/lno.11655</u>

65McArdle, B.H., Anderson, M.J., 2001. Fitting multivariate models to community data: a comment on

distance-based redundancy analysis. Ecology 82, 290–297. <u>https://doi.org/10.1890/0012-</u>
 <u>9658(2001)082[0290:FMMTCD]2.0.CO;2</u>

66McLean, M., Stuart-Smith, R.D., Villéger, S., Auber, A., Edgar, G.J., MacNeil, M.A., Loiseau, N.,

Leprieur, F., Mouillot, D., 2021. Trait similarity in reef fish faunas across the world's oceans.

662 Proceedings of the National Academy of Sciences 118, e2012318118.

663 <u>https://doi.org/10.1073/pnas.2012318118</u>

66Micheli, F., Halpern, B.S., 2005. Low functional redundancy in coastal marine assemblages. Ecology
665 Letters 8, 391–400. <u>https://doi.org/10.1111/j.1461-0248.2005.00731.x</u>

66 Mouchet, M.A., Villéger, S., Mason, N.W.H., Mouillot, D., 2010. Functional diversity measures: an

overview of their redundancy and their ability to discriminate community assembly rules. Functional

668 Ecology 24, 867–876. <u>https://doi.org/10.1111/j.1365-2435.2010.01695.x</u>

66Mouillot, D., Graham, N.A.J., Villéger, S., Mason, N.W.H., Bellwood, D.R., 2013. A functional approach
reveals community responses to disturbances. Trends in Ecology & Evolution 28, 167–177.
https://doi.org/10.1016/j.tree.2012.10.004

67Naeem, S., Wright, J.P., 2003. Disentangling biodiversity effects on ecosystem functioning: deriving

solutions to a seemingly insurmountable problem. Ecology Letters 6, 567–579.

674 https://doi.org/10.1046/j.1461-0248.2003.00471.x

67Nock, C.A., Vogt, R.J., Beisner, B.E., 2016. Functional Traits, in: John Wiley & Sons, Ltd (Ed.), ELS.
676 Wiley, pp. 1–8. <u>https://doi.org/10.1002/9780470015902.a0026282</u>

67Norkko, A., Villnäs, A., Norkko, J., Valanko, S., Pilditch, C., 2013. Size matters: implications of the loss of
large individuals for ecosystem function. Scientific Reports 3, 2646. <u>https://doi.org/10.1038/srep02646</u>

67Øsman, R.W., 1977. The Establishment and Development of a Marine Epifaunal Community. Ecological
680 Monographs 47, 37–63. <u>https://doi.org/10.2307/1942223</u>

68Pacesila, M., Burcea, S.G., Colesca, S.E., 2016. Analysis of renewable energies in European Union.
682 Renewable and Sustainable Energy Reviews 56, 156–170. <u>https://doi.org/10.1016/j.rser.2015.10.152</u>

68Betchey, O.L., Gaston, K.J., 2006. Functional diversity: back to basics and looking forward. Ecol Letters 9,
684 741–758. <u>https://doi.org/10.1111/j.1461-0248.2006.00924.x</u>

68Betersen, J.K., Malm, T., 2006. Offshore Windmill Farms: Threats to or Possibilities for the Marine

686 Environment. AMBIO: A Journal of the Human Environment 35, 75–80.

687 https://doi.org/10.1579/0044-7447(2006)35[75:OWFTTO]2.0.CO;2

68Binto, T.K., Barros, F.P.A., Nunes, J.A.C.C., Miranda, R.J., Pereira, B.M.S., Pereira, P.H.C., Sampaio,

689 C.L.S., 2021. Shipwrecks fouling community: Similarities with natural rocky reefs at different

distances and wave exposure regimes. Ocean & Coastal Management 213, 105895.

691 <u>https://doi.org/10.1016/j.ocecoaman.2021.105895</u>

- 69Poff, N.L., Olden, J.D., Vieira, N.K.M., Finn, D.S., Simmons, M.P., Kondratieff, B.C., 2006. Functional
- trait niches of North American lotic insects: traits-based ecological applications in light of
- 694 phylogenetic relationships. Journal of the North American Benthological Society 25, 730–755.
  695 https://doi.org/10.1899/0887-3593(2006)025[0730:FTNONA]2.0.CO;2
- 69Reeves, D., Chesney, E., Munnelly, R., Baltz, D., 2018. Barnacle settlement and growth at oil and gas
- platforms in the northern Gulf of Mexico. Marine Ecology Progress Series 590, 131–143.
   <u>https://doi.org/10.3354/meps12468</u>

699 hields, M.A., Payne, A.I.L. (Eds.), 2014. Marine Renewable Energy Technology and Environmental
700 Interactions, Humanity and the Sea. Springer Netherlands, Dordrecht. <u>https://doi.org/10.1007/978-94-</u>
701 017-8002-5

708 himeta, J., Cutajar, J., Watson, M., Vlamis, T., 2012. Influences of biofilm-associated ciliates on the
settlement of marine invertebrate larvae. Marine Ecology Progress Series 449, 1–12.

704 <u>https://doi.org/10.3354/meps09638</u>

70Svane, I., Petersen, J.K., 2001. On the Problems of Epibioses, Fouling and Artificial Reefs, a Review.
706 Marine Ecology 22, 169–188. <u>https://doi.org/10.1046/j.1439-0485.2001.01729.x</u>

70**T**an, E.L.-Y., Mayer-Pinto, M., Johnston, E.L., Dafforn, K.A., 2015. Differences in Intertidal Microbial
 Assemblages on Urban Structures and Natural Rocky Reef. Frontiers in Microbiology 6: 1276.

70**T**aniguchi, H., Tokeshi, M., 2004. Effects of habitat complexity on benthic assemblages in a variable
environment. Freshwater Biology 49, 1164–1178. <u>https://doi.org/10.1111/j.1365-2427.2004.01257.x</u>

71Teixidó, N., Garrabou, J., Gutt, J., Arntz, W.E., 2004. Recovery in Antarctic benthos after iceberg
disturbance: trends in benthic composition, abundance and growth forms. Marine Ecology Progress
Series 278, 1–16. <u>https://doi.org/10.3354/meps278001</u>

71**T**hornton, I., 2007. Island Colonization: The Origin and Development of Island Communities. Cambridge715 University Press.

716 ilman, D., 2001. Functional Diversity, in: Encyclopedia of Biodiversity. Elsevier, pp. 109–120.
 https://doi.org/10.1016/B0-12-226865-2/00132-2

71\%ilman, D., 1988. Plant Strategies and the Dynamics and Structure of Plant Communities. (MPB-26),
Volume 26. Princeton University Press. <u>https://doi.org/10.2307/j.ctvx5w9ws</u>

720an der Linden, P., Marchini, A., Dolbeth, M., Patrício, J., Veríssimo, H., Marques, J.C., 2016. The
721 performance of trait-based indices in an estuarine environment. Ecological Indicators 61, 378–389.

722 <u>https://doi.org/10.1016/j.ecolind.2015.09.039</u>

723an der Stap, T., Coolen, J.W.P., Lindeboom, H.J., 2016. Marine Fouling Assemblages on Offshore Gas
Platforms in the Southern North Sea: Effects of Depth and Distance from Shore on Biodiversity. PLoS

725 ONE 11, e0146324. https://doi.org/10.1371/journal.pone.0146324

726/anagt T., Van de Moortel L. and Faasse M, 2013. Development of hard substrate fauna in the Princess
Amalia Wind Farm Monitoring 3.5 years after construction (No. 2013009). eCOAST.

728 Valker, L.R., Chapin, F.S., 1987. Interactions among Processes Controlling Successional Change. Oikos
 50, 131–135. <u>https://doi.org/10.2307/3565409</u>

73Wiggins, G.B., Mackay, R.J., 1978. Some Relationships between Systematics and Trophic Ecology in

731 Nearctic Aquatic Insects, with Special Reference to Trichoptera. Ecology 59, 1211–1220.

732 <u>https://doi.org/10.2307/1938234</u>

73 Wilhelmsson, D., Malm, T., 2008. Fouling assemblages on offshore wind power plants and adjacent

- substrata. Estuarine, Coastal and Shelf Science 79, 459–466.
- 735 <u>https://doi.org/10.1016/j.ecss.2008.04.020</u>

73Zhulay, I., Bluhm, B.A., Renaud, P.E., Degen, R., Iken, K., 2021. Functional Pattern of Benthic Epifauna in
the Chukchi Borderland, Arctic Deep Sea. Frontiers Marine Science 8, 609956.

738 https://doi.org/10.3389/fmars.2021.609956

739