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Impact of the *Phaeocystis globosa* spring bloom on the intertidal benthic compartment in the eastern English Channel: A synthesis

Nicolas Spilmont^{a, *}, Lionel Denis^a, Luis Felipe Artigas^b, Frédéric Caloin^c, Lucie Courcot^b, Anne Créach^d, Nicolas Desroy^e, François Gevaert^a, Pascal Hacquebart^a, Cédric Hubas^{b, g}, Marie-Andrée Janquin^a, Yves Lemoine^d, Christophe Luczak^{a, c, h}, Aline Migné^{f, g}, Mathieu Rauch^a and Dominique Davoult^{f, g}

^h Université d'Artois, IUFM, Centre de Gravelines, 40 rue Victor Hugo – BP 129, 59820 Gravelines Cedex, France

*: Corresponding author : Nicolas Spilmont, email address : nicolas.spilmont@univ-lille1.fr

Abstract:

From 1999 to 2005, studies carried out in the frame of regional and national French programs aimed to determine whether the Phaeocystis globosa bloom affected the intertidal benthic communities of the French coast of the eastern English Channel in terms of composition and/or functioning. Study sites were chosen to cover most of the typical shore types encountered on this coast (a rocky shore, an exposed sandy beach and a small estuary). Both the presence of active Phaeocystis cells and their degradation product (foam) did have a significant impact on the studied shores. The primary production and growth rates of the kelp Saccharina latissima decreased during the bloom because of a shortage of light and nutrient for the macroalgae. On sandy sediments, the benthic metabolism (community respiration and community primary production), as well as the nitrification rate, were enhanced during foam deposits, in relation with the presence of bacteria and active pelagic cells within the decaying colonies. In estuarine sediments, the most impressive impact was the formation of a crust at the sediment surface due to drying foam. This led to anoxic conditions in the surface sediment and resulted in a high mortality among the benthic community. Some organisms also tended to migrate upward and were then directly accessible to the higher trophic level represented by birds. Phaeocystis then created a shortcut in the estuarine trophic network. Most of these modifications lasted shortly and all the systems considered came back to their regular properties and activities a few weeks after the end of the bloom, except for the most impacted estuarine area.

Keywords: Intertidal; *Phaeocystis globosa*; Benthic fluxes; *Saccharina latissima*; Bacteria; Macrozoobenthos

^a Université des Sciences et Technologies de Lille, Laboratoire d'Océanologie et de Géosciences, CNRS UMR 8187 LOG, Station Marine de Wimereux, 28 av. Foch – BP 80, 62930 Wimereux, France

^b Université du Littoral Côte d'Opale, Laboratoire d'Océanologie et de Géosciences, CNRS UMR 8187 LOG, Maison de la Recherche en Environnement Naturel, 32 av. Foch, 62930 Wimereux, France

^c Groupe Ornithologique et Naturaliste du Nord-Pas-de-Calais (GON), MNE 23 rue Gosselet, 59000 Lille, France ^d Université des Sciences et Technologies de Lille, UMR CNRS 8016, Bât SN2, 59655 Villeneuve d'Ascq Cedex, France

^e IFREMER, LER FBN, station de Saint-Malo, 38 rue du Port Blanc, 35800 Dinard, France

^f UPMC Univ Paris 6, UMR 7144 AD2M, Station Biologique de Roscoff, Place Georges Teissier – BP 74, Roscoff Cedex, France

⁹ CNRS, UMR 7144 AD2M, Station Biologique de Roscoff, Place Georges Teissier – BP 74, Roscoff Cedex, France

52 **1. Introduction**

53 Because of their interface position, coastal ecosystems are highly productive 54 environments (e.g. Gazeau et al., 2004). They are characterised by intense physical and 55 chemical gradients, and human activities have been proved to impact on their functioning, 56 especially through eutrophication (Cloern, 2001; Gray et al., 2002). Within coastal 57 environments, the intertidal zones constitute favourable areas for exchanges with the adjacent 58 terrestrial systems and coastal water masses. Thus, their functioning is mainly driven by these 59 exchanges of matter and energy and they constitute highly variable systems, both in space and 60 time. In particular, exchanges between the sediment and the water column, especially during 61 phytoplanktonic blooms, often enhanced by eutrophication, have been reported to drastically 62 affect the biogeochemical processes in surficial sediments (e.g. Boon et al., 1998; Boon et al., 63 1999; Grenz et al., 2000).

64 Phaeocystis globosa is one of the three Phaeocystis species that have been reported to 65 form blooms (for a review, see Schoeman et al., 2005). P. globosa is a temperate species found on the coast of the eastern English Channel, southern North Sea and the south coast of 66 67 China. Though not being toxic, P. globosa is considered to be responsible for harmful algal 68 blooms (HAB; Veldhuis and Wassmann, 2005). Its life cycle is characterized by different 69 forms of solitary cells alterning with colonial stages (Whipple et al., 2005 and references 70 therein). These colonies are constituted of cells packed within a matrix and are surrounded by 71 a "skin" (Hamm et al., 1999) which protects the cells from minute organisms (Hamm, 2000). 72 During the colonial phase, one of the most spectacular manifestations of the bloom, because 73 visible to the naked eye, can be observed on the seashore, consisting in large accumulations of 74 mucilaginous aggregates (foam; Fig.1). This foam formation occurs under specific windy conditions (Lancelot, 1995) and the amount of foam deposited on the beach is directly 75 76 dependent on the wind speed and direction that control foam formation and deposit on the

beach, respectively. The intensity of the bloom in the eastern English Channel and in Belgian
coastal waters appeared to be related to the North Atlantic Oscillation (NAO) and nutrient
loads by rivers (Seuront and Souissi, 2002; Breton et al., 2006; Gypens et al., 2007), and the
differences in blooms amplitude between the two areas have been assigned to differences in
the turbulence regime (Schapira et al., 2006).

82 Though the effect of the *P. globosa* bloom within the water column has been widely 83 studied regarding for instance seawater properties (Seuront et al., 2006) or interactions with 84 planktonic organisms (e.g. Ruardij et al., 2005), few studies concerned its effect on benthic 85 organisms and/or processes. Accumulation of colonies on the sediment and their 86 incorporation into the food web have been described (Cadée, 1996), as well as the occurrence 87 of meiofauna feeding on the foam (Armonies, 1989) and the efficiency of filtration and 88 absorption of the different life stages by benthic organisms (Smaal and Twisk, 1997). 89 Although the intertidal zone is potentially the most impacted area because of foam 90 accumulations, no study had been devoted to this specific system before the late nineties. 91 However, foam deposits might constitute huge organic matter inputs into the sediment and 92 might then interfere with both the living biota and the sediment biogeochemical processes. Furthermore, the high turbidity levels encountered during the spring bloom (Seuront et al., 93 94 2006 and references therein) might interfere with the photosynthetic activity of the benthic 95 autotrophic organisms.

From 1999 to 2005, studies aimed to determine whether the *Phaeocystis* bloom affected the intertidal benthic communities of the French coast of the eastern English Channel in terms of composition and/or functioning. The eastern English Channel is an epicontinental sea that constitutes a transition zone between the Atlantic Ocean and the North Sea. It is characterized by a semi diurnal megatidal regime, the tidal magnitude reaching 9 m in spring tides in estuaries. In the area, the shoreline exhibits a complex morphology composed of sandy beaches, cliffs, estuaries, wetlands and polders (Battiau-Queney et al., 2001; Dauvin, in

press). Thus, the study sites were selected along the coast to cover most of the shore types that
might be impacted during the *Phaeocystis* bloom (i.e. a rocky shore, an exposed sandy beach
and a small estuary).

While the consequences of the bloom in the subtidal area are described in Denis and
Desroy (this volume), the present paper synthesises the results that were obtained in the
different intertidal environments regarding benthic metabolism and fluxes and the
macrozoobenthic communities composition.

110

111 **2. Materials and Methods**

112 Three study sites representing the major sediment types of the eastern English Channel 113 were investigated along the French coast (Fig. 2): a rocky shore in Audresselles, an exposed 114 sandy beach in Wimereux and the Canche Bay as an estuary.

115

116 *2.1. Macroalgal photosynthetic activity and growth rate in the rocky shore*

117 In Audresselles, the photosynthesis and growth rate of the kelp Saccharina latissima 118 (formerly Laminaria saccharina) were measured in 2001. The photosynthetic activity was 119 measured *in situ*, during the late phase and after the *Phaeocystis* bloom (in early May and 120 early June for the year considered, respectively). A submersible pulse amplitude modulated 121 (Diving PAM, Walz) fluorometer was used to estimate the relative electron transport rate 122 (rETR) following Gevaert et al. (2003). The photosynthetically active radiation (PAR) 123 reaching the algae was measured simultaneously to each rETR measurement, and the surface 124 PAR was continuously recorded during the experiments (Li-Cor SA-190 quantum sensor). 125 Additionally, 75 individuals (of an initial length between 22.5 cm and 74.1 cm) were 126 marked with coloured thread attached to their stipe in early March. From March to June 2001, during spring tides when algae were accessible by foot (every 15 or 30 days), the total length 127

128 (from the base on the holdfast to the top of the frond) of each individual was measured. After

129	each measurement, a hole (1 cm diameter) was pierced in the frond at a distance of 10 cm of
130	the junction between the frond and the stipe. The growth rate was expressed as follow:
131	
132	$GR = 100 \left[(Lh_{t+1}-Lh_t)/L_t \right] / TL $ (1)
133	
134	with GR the growth rate (%.d ⁻¹); Lh_t and Lh_{t+1} the length between the hole and the junction
135	between the frond and the stipe (cm) measured at time t and t+1, respectively; L_t the total
136	length measured at time t (cm); TL the time lag between t and t+1 (d). Due to storm events
137	and human activities (fishing), the number of individuals at the end of the experiment was 34.
138	
139	2.2. Benthic metabolism, oxygen and nutrient fluxes, scanning electron microscopy of
140	foam in the exposed sandy beach
141	The study site located in Wimereux was a typical exposed sandy beach characterised
142	by medium size sand. A survey of benthic metabolism under emersed conditions (community
143	primary production and community respiration measured as CO ₂ fluxes following Migné et
144	al., 2002) was performed from March 2000 to September 2005 (Spilmont et al., 2005; Hubas
145	et al., 2007). During the period, seven additional measurements were carried out on
146	Phaeocystis foam deposits (one in 2001, two in 2002, 2003 and 2004). The bacterial biomass
147	within the first cm of sediment was concurrently estimated together with bacterial production
148	and growth efficiency during the 2004-2005 survey.
149	Benthic fluxes at the water-sediment interface (O_2, NH_4^+) were estimated in 2004 and
150	2005 on sediment cores incubated in the laboratory, by measuring changes in concentrations
151	in the overlying water and by measuring solute concentrations in the interstitial water (Rauch
152	et al., in press).
153	For the scanning electron microscopy observation, samples were fixed with
154	lugol/glutaraldehyde solutions (0.2% v/v final concentration) and filtered through

155 polycarbonate membrane (Millipore, 2µm). They were then rinsed 3 times in MilliQ water 156 (Millipore), dehydrated in a graded ethanol series (30, 50, 70, 80, 90 and 100%, Fisher 157 Scientific, P.A.) and critical point dried in liquid CO₂ using critical point drying 158 apparatus(Bal-Tec CPD 030). Filters were mounted on a aluminium stubs, sputter-coated with 159 Au/Pd (Polaron SC7620) and observed at 15kV with a SEM LEO 438 VP.

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2.3. Macrozoobenthos, oxygen and nutrient fluxes and waders frequentation in the 162 estuary

The Canche Bay is a small estuary (ca. 6.3 km^2 of intertidal area) characterised by a 163 relatively low river discharge (100 m³.s⁻¹ in winter and 60 m³.s⁻¹ in summer). Four intertidal 164 165 sites were studied for macrozoobenthos (abundance, species richness and diversity), sediment organic matter content, sediment pigments contents (chlorophyll *a* and phaeopigments) and 166 167 oxygen microprofiles from April to December 2001 (Desroy and Denis, 2004). One sandy 168 and one muddy station were studied both in the northern and the southern part of the Bay 169 (medium to fine sands and mud in the north; muddy fine sands and mud in the south). Data on 170 macrofauna were also collected using a multilevel core sampler (Luczak and Menu, 171 submitted) that provided complementary information on the vertical distribution of the 172 macrofauna within the sediment. This device allowed the sampled core to be sliced in sub-173 samples corresponding to five layers in the sediment (0-1 cm; 1-2.5 cm; 2.5-5 cm; 5-10 cm 174 and 10-20 cm). Sediment oxygen demand and total fluxes of dissolved inorganic nitrogen 175 fluxes at the sediment/water interface were also estimated in the northern part of the Bay in 176 March and May 2003 (Rauch and Denis, in press).

177 The number of waders staging in the estuary was based on the results of counts 178 performed every ten days (at least one count per ten days period) in 2001. For the 1998-2000 179 period, data come primarily from the Groupe Ornithologique et Naturaliste du Nord-Pas-de-180 Calais (G.O.N.) database. Counts were made by a combination of experienced volunteers and

professionals. High water counts were used, when birds concentrate on roosts. When several
counts were performed in a ten days period, the highest count for each species was used
(Yates and Goss-Custard, 1991).

184

185 **3. Results and discussion**

186 *3.1. Shortage of light and nutrients for* Saccharina latissima

187 The results obtained during the *in situ* measurements of the photosynthetic activity and 188 growth rate of Saccharina latissima are exhibited on Figs 3 and 4. During the late phase of the 189 bloom (early May), almost no light reached the bottom as soon as the study site was submerged, with an average light extinction of 99% (average PAR of $11 \pm 7 \mu mol.m^{-2}.s^{-1}$ 190 191 during immersion, n = 5; Fig. 3A) whereas light kept reaching the algae all along the 192 immersion period during the experiment conducted after the bloom (early June: 79% average light extinction, $118 \pm 88 \,\mu\text{mol.m}^{-2}\text{.s}^{-1}$ during immersion, *n*=6; Fig. 3B). Thus, the 193 194 photosynthetic activity of S. latissima during immersion, estimated through the measurement of rETR, was much lower when measurements were performed during the bloom $(4 \pm 2,$ 195 196 n = 5; Fig. 3A) than for individuals in regular conditions (28 ± 17, n = 6; Fig. 3B). As a 197 consequence, the average growth rate of S. latissima decreased during the bloom (down to $0.61 \pm 0.02\%$.d⁻¹; Fig. 4), though March-April is usually a period of intense growth for the 198 species. Towards the end of the bloom, the growth rate became higher $(1.37 \pm 0.05\% d^{-1})$ and 199 200 then the algae came back to their regular growth cycle (limited growth rate throughout 201 summer). The time course of the carbon content (Fig 5A in Gevaert et al., 2001) exhibited a 202 typical trend for the species (directly linked to the growth cycle), being the lowest in March 203 $(23.9 \pm 0.5\% \text{ DW})$ and the highest in September $(31.4 \pm 0.6\% \text{ DW})$. However, the time 204 course of the nitrogen content (Fig. 5B in Gevaert et al., 2001) was characterised by a 205 decrease from March $(3.4 \pm 0.1\% \text{ DW})$ to June $(2.2 \pm 0.1\% \text{ DW})$, that matched with the 206 Phaeocystis bloom period for the year of the study (1999 in Gevaert et al., 2001). The

207 nitrogen being taken up from the medium by *S. latissima* as NO_3^- , it can be assumed that the 208 seawater was depleted in nitrate during this period and led the algae to use their internal N 209 stock. The growth of *Phaeocystis globosa* is known to be mainly sustained by the massive use 210 of NO_3^- (Schapira et al., 2006), with concentrations in the seawater almost decreasing down to 211 0 µmol.l⁻¹ in March-April (Gentilhomme and Lizon, 1998); this suggests that *P. globosa* and 212 *S. latissima* compete for the same nitrogen source.

During the bloom it appeared that *S. latissima* was limited both in light and nitrate, as a consequence of (i) the turbidity induced by the presence of *Phaeocystis* cells and and (ii) a competition for NO_3^- . After the bloom, the algae recovered its regular photosynthetic activity and growth rate. This testifies to the plasticity of this species to adapt to environmental variations, as also previously underlined at the day scale (Gevaert et al., 2003).

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3.2. Enhancement of the benthic metabolism due to foam deposits in the exposed sandy
beach

222 Figure 5 shows the results obtained during emersion in terms of gross community 223 production (GCP) and community respiration (CR), both on bare sediment and on sediment 224 covered by foam deposits. CR was always higher on foam deposits than on bare sediment and the highest CR ever measured on this study site was on foam deposits on May 17 in 2001 225 (11.95 mgC.m⁻².h⁻¹). By compiling data obtained by Spilmont et al. (2005) and Hubas et al. 226 227 (2007), there was a significant positive correlation between sediment bacterial biomass and 228 benthic community respiration during emersion (r = 0.547, n = 41, p < 0.01; Fig. 6). As 229 mentioned by these authors the community respiration in this sandy beach was dominated by 230 bacterial respiration all along the year; the organic matter input made up of *Phaeocystis* foam 231 would constitute a substratum for autochtonous bacterial populations and would also bring

allochtonous bacteria within the matrix, bacterial abundance and production also being linkedto the bloom within the water column (Becquevort et al., 1998; Lamy et al., 2006).

Gross community production was always detected on foam deposits, except on May 234 235 17 in 2001. Furthermore, GCP was often higher on deposits than on bare sediment (on May 6 and 21 in 2002 and on May 5 and June 9 in 2004; Fig. 5) and was as high as 23.13 mgC.m⁻².h⁻ 236 ¹ on foam on June 06 in 2004. This suggested that active phytoplanktonic cells were deposited 237 238 within the foam. This hypothesis was corroborated by high assimilation numbers typical of 239 phytoplanktonic cells calculated by Spilmont et al. (2005) and Hubas et al. (2007). The 240 scanning electron microscopy photograph of a foam sample taken during the deposit in spring 241 2004 (Fig. 7), brought a new evidence of phytoplanktonic cells trapped within the foam (a 242 *Phaeocystis* cell and a diatom are visible). Bacteria responsible for the high respiration rates 243 were also visible on this picture.

244 As for air-sediment CO₂ fluxes, the sediment oxygen demand (SOD) and ammonium 245 release increased at the water-sediment interface during the spring bloom of *Phaeocystis* 246 globosa, whereas these processes remained limited during the rest of the year (Fig. 8). The 247 increase in SOD obviously appeared to be a direct response to the organic matter input and 248 was correlated to bacterial abundances. The concomitant release of nitrates described by 249 Rauch et al. (in press) was not only a diffusive flux from the sediment to the water column but 250 was also linked to high nitrification rates coupled with lower denitrification rates in surficial 251 sediments.

As mentioned by Spilmont et al. (2005) and Rauch et al. (in press), both the phase of the bloom and the meteorological conditions control the response of the sedimentary processes. During the early stage of the bloom (exponential growth of both solitary cells and healthy colonies), low community primary production and respiration were measured, but calm condition might allow phytoplanktonic cells to settle and lead to an increase in primary production at the air-sediment interface (Hubas et al., 2007). Nutrients being rapidly

consumed in the water column during this phase, the diffusive nutrient fluxes increased.
During the late stage of the bloom (decaying colonies), windy conditions are favourable for
foam accumulations to be brought to the beach and to settle at low tide. Thus, both relatively
high community production and high respiration rates are expected. Nutrient cycling is also
affected during this stage, the nitrification and denitrification processes then being uncoupled
(Rauch and Denis, in press).

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3.3. Benthic fluxes and macrozoobenthic community modifications in the Canche Bay

266 The results obtained on benthic fluxes in the Canche Bay gave the same conclusions as 267 those previously described for Wimereux, i.e. an increase in the sediment oxygen demand and 268 a high dissolved inorganic nitrogen release during the bloom, these modifications being 269 reversible a short time after the bloom (Rauch and Denis, in press). In a previous analysis of 270 the macrozoobenthic community at four sampling stations in this study site, Desroy and Denis 271 (2004) had the opportunity to compare both low-impacted and high-impacted areas 272 (respectively the south side and the north side of the Bay) during the 2001 bloom. The authors 273 reported values of usual community descriptors (density, specific richness and diversity) 274 exhibiting a rapid and significant reduction during the late phase of the *Phaeocystis* bloom in 275 the impacted area (see their Figs. 7 and 8). These modifications, in terms of quantity and 276 quality of macrozoobenthos, were linked to the development of anoxic conditions, sometimes being as extreme as the formation of an encrusting foam avoiding O₂ to penetrate further than 277 278 1 mm (see their Fig. 5), when foam deposits were combined with high temperatures. These 279 anoxic conditions led either to the direct death of some organisms, or to an upward migration 280 of some individuals. The results obtained with the multilevel corer showed that the 281 macrofauna tended to migrate upward during the first fortnight of May (Fig. 9) together with 282 a decreasing number of species in all the layers (Fig. 10), especially in the northern sandy sediment where the encrusting foam was encountered. In these conditions, preys are directly 283

284 accessible for birds, without the need to dig them out. Interestingly enough, waders have been 285 observed feeding on the macrofauna lying on the sediment during anoxic conditions (authors 286 pers.obs). It appears that the upward migration of the macrofauna perfectly matched with the 287 second annual peak of abundance of waders in the Canche Bay during refuelling on their 288 migration route (Fig. 11). Following the diagram proposed by Zwarts and Wanink (1993), in 289 regular conditions the harvestable fraction of benthic prey for waders depends on the fraction 290 being profitable within preys that are available. In this context, small preys are usually 291 ignored for an optimisation of the ratio between the energy needed to handle them and the 292 energy they bring. In our study, since the remaining living macrofauna was lying on the 293 sediment or very close to the surface, all preys (i.e. small and large ones) were profitable since 294 waders did not have to spend much energy to feed on them. In this context, Phaeocystis did 295 temporally and locally modify the energy flow through the estuarine ecosystem. As 296 underlined by Desroy and Denis (2004), the macrozoobenthic communities recovered slowly 297 from the disturbance, especially the north mud site which still exhibited very low species 298 richness and density six months after the bloom.

299

300 Conclusion

301 *Phaeocystis* is considered to play a key role in global biogeochemical cycles, climatic 302 regulation and fisheries yield (Schoeman et al., 2005). However, this statement was mainly (if 303 not only) derived from studies on its impact on the physical, chemical and biological 304 properties of the water column. In this synthesis, we strengthened this theory through results 305 obtained on the intertidal benthic compartment. As exhibited in Table 1, both *Phaeocystis* 306 active cells and their degradation product (foam) did have a significant impact on the studied 307 shores. *Phaeocystis* actually modified the nitrogen cycle within the sediment and thus the N 308 fluxes from the sediment, and also appeared to compete with macroalgae for nitrates. 309 Furthermore, throughout the *Phaeocystis* bloom, the rates of macroalgae primary production

310 on rocky shores decreased, whereas the rates of gross community production and community 311 primary production were enhanced on the sandy shore, then leading to changes in the carbon 312 fluxes from the benthos to the air and the water column. The macrozoobenthos quality 313 (diversity) and quantity (abundances) decreased during foam accumulation, some of these 314 changes leading to energetic shortcuts within the estuarine trophic network (i.e. from 315 macrozoobenthos to birds). However, most of these modifications lasted shortly. Indeed, all 316 the systems and organisms came back to their regular activity or presented their initial 317 properties a few weeks after the end of the bloom. The only notable exception was one of the 318 benthic communities from muddy sediments studied in the Canche Bay which did not recover 319 more than 6 months after the end of the bloom. Nevertheless, this represented a very singular 320 case since this site was particularly impacted that year (2001). This last remark points out the 321 fact that the intensity of the *Phaeocystis* bloom remains locally unpredictable in time and 322 highly variable in space (see Fig. 1). For example, although the abundance of the species has 323 been stated to be related to the NAO (Seuront and Souissi, 2002; Breton et al., 2006), the 324 abundance and accumulation of foam is very dependant on local hydrodynamic and 325 topographic conditions. Thus, the surface covered by foam is not predictable and remains 326 difficult to estimate at a large spatial scale due to its short residence time (the foam is wiped 327 out during each rising tide); the calculation of an accurate impact of this deposit on carbon 328 and nitrogen fluxes remains very hazardous.

329

330 Aknowledgements

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473	Sea depends on the variation in energy density, body weight, biomass, burying depth and
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477 Figure captions

478 Figure 1. Accumulations of mucilaginous aggregates (foam) of *Phaeocystis globosa* on the
479 beach of Wimereux during the 2002 spring bloom.

480

481 Figure 2. Location of the study sites along the French coast of the Eastern English Channel.482

483 Figure 3. Photosynthetic activity of *Saccharina latissima* in spring 2001. The photosynthetic

484 activity was measured as the relative electron transfer rate (rETR, dots) and

485 photosynthetically active radiations (PAR, μ mol.m⁻².s⁻¹, dashed line) reaching the algae

486 (bottom PAR) were measured simultaneously. The surface PAR was measured throughout the

487 experiments (dark line). The grey bars symbolise the periods of immersion.

488

489 Figure 4. Growth rate(C) of *Saccharina latissima* measured from March to June 2001,

490 expressed in $\%.d^{-1}$. The grey bar roughly represents the bloom intensity (the darker, the higher

491 the chlorophyll concentration in the water column).

492

493 Figure 5. Gross community production (GCP, positive values in $mgC.m^{-2}.h^{-1}$) and community

494 respiration (CR, negative values in mgC.m⁻².h⁻¹) measured during emersion in the sandy

495 beach of Wimereux on bare sediment (open bars) and on foam deposits (dashed bars). 0

496 indicates that no flux was detected and \times indicates that fluxes were not measured. Redrawn

497 from Spilmont et al. (2005) and Hubas et al. (2007).

498

499 Figure 6. Sediment bacterial biomass (mgC.m⁻²) plotted against benthic community

500 respiration during emersion (CR, $mgC.m^{-2}$) and linear regression forced through the origin

501 (CR = $0.48 \times$ Bacterial biomass, r = 0.547, n = 41, p < 0.01). Data compiled from Spilmont et

502 al. (2005) and Hubas et al. (2007).

504

505 a SEM LEO 438 VP) showing a *Phaeocystis globosa* solitary flagellate cell (P), a diatom (D) 506 and some bacteria (B). 507 Figure 8. Average Sediment Oxygen Demands and ammonium releases (μ mol.m⁻².h⁻¹; ± SD; 508 509 n=4) at the sediment-water interface during the year 2004. Black diamonds represent values 510 recorded in the absence of *Phaeocystis globosa*, whereas open squares are measurements 511 performed during the spring 2004 *Phaeocystis globosa* bloom (late March-early May). 512 513 Figure 9. Time course of the average burying depth (cm) of the macrofauna at the four 514 sampling stations surveyed in the Canche Bay in 2001. The grey bars symbolise the period of 515 maximal foam accumulation on the sediment. The north site was highly impacted, the south 516 site was low impacted. 517 518 Figure 10. Time course of the number of species found in each layer sampled using the 519 multilevel corer (layer 1 (L1): 0-1 cm, layer 2 (L2): 1-2.5 cm, layer 3 (L3): 2.5-5 cm, layer 4 520 (L4): 5-10 cm, layer 5 (L5): 10-20 cm) at the four sampling stations surveyed in the Canche 521 Bay for the year 2001. The grey bars symbolise the period of maximal foam accumulation on 522 the sediment. The north site was highly impacted, the south site was low impacted. 523 524 Figure 11. Total number of waders counted within the Canche Bay for ten day periods (3 per 525 each month of the year) for the period 1998-2001 (grey bars, mean \pm s.d.) and for the year 526 2001 (dashed bars).

Figure 7. Scanning electron microscopy photograph of a foam sample (observed at 15kV with

- 527
- 528

529 **Table captions**

- 530 Table 1. Synthesis of the effects of the *Phaeocystis* bloom on the physical and/or biological
- 531 environment with their consequences on the benthic communities from the different shore
- 532 types studied.

533





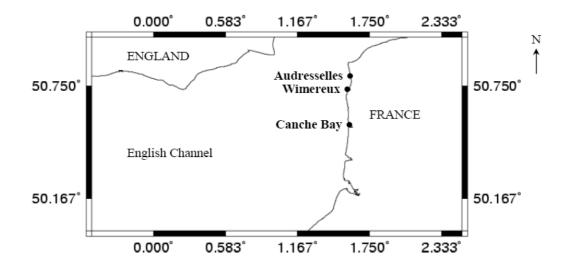


Figure 2



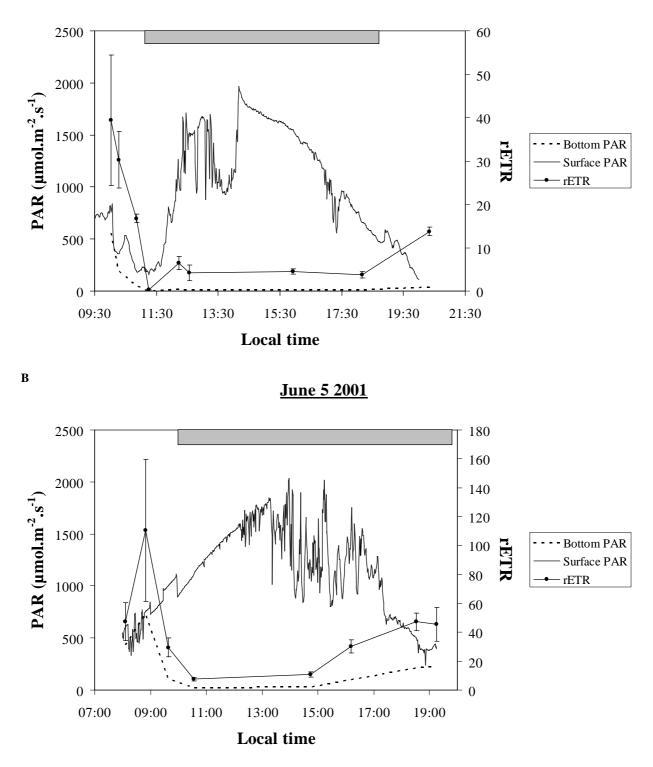


Figure 3

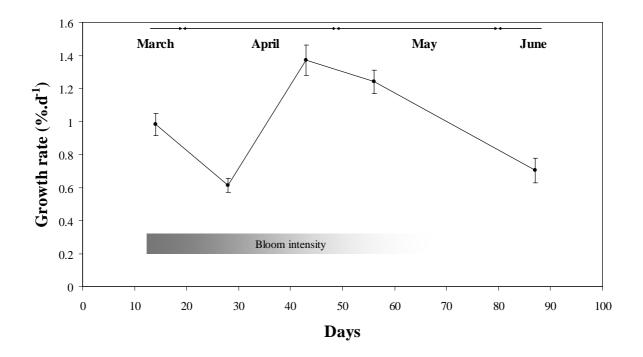


Figure 4

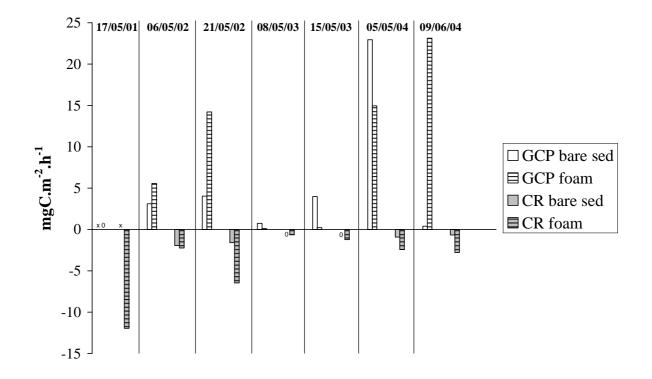
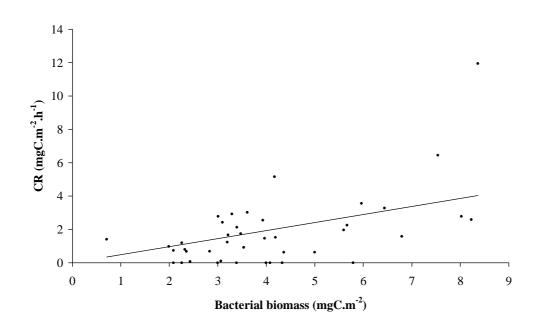


Figure 5.





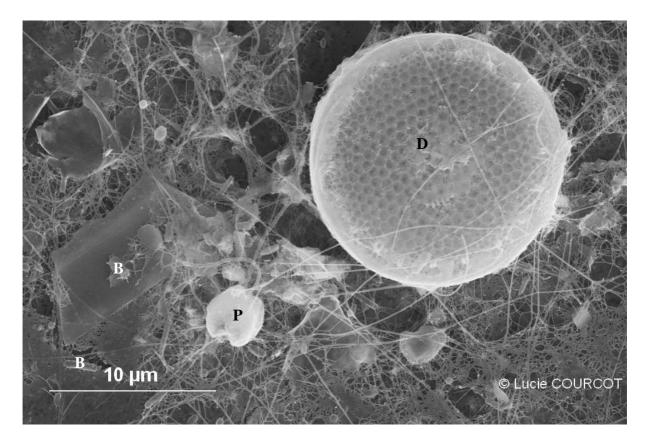


Figure 7.

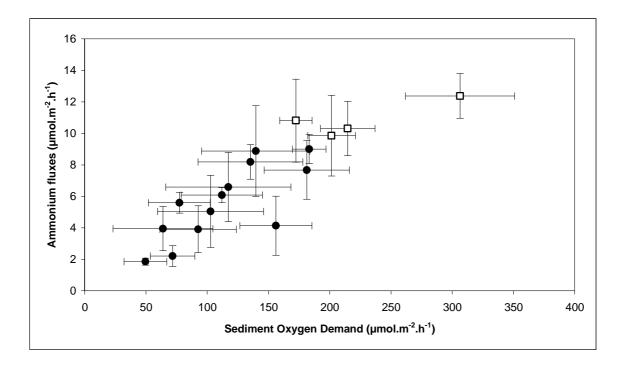
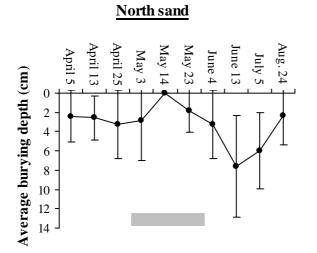
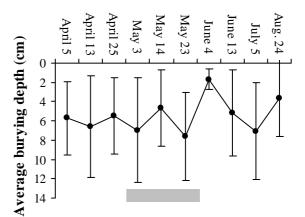


Figure 8.



South sand



South mud

May 14 \pm

May 3

April 25 +

April 13 +

April 5 +

0

Average burying depth (cm)

May 23

June 4

June 13

Aug. 24

July 5 H

North mud

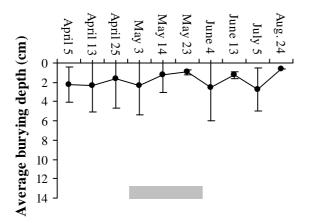


Figure 9.

North sand

South sand

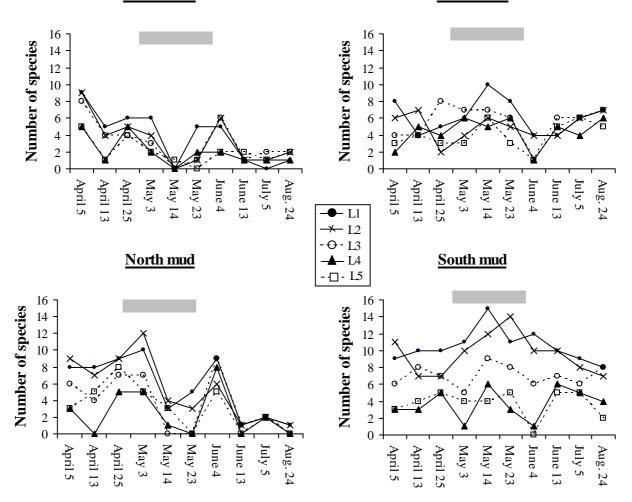


Figure 10.

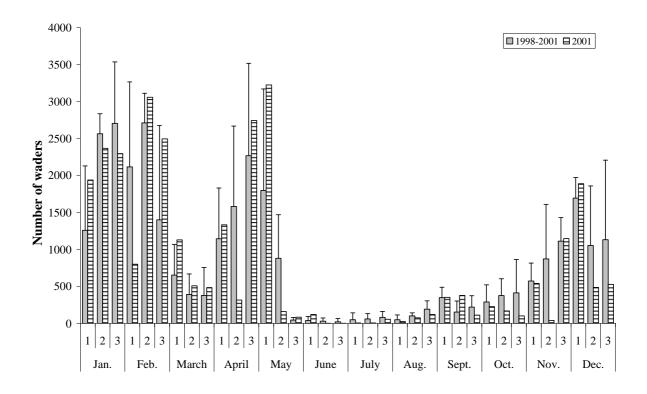


Figure 11.

Table 1

Bloom stage	Early stage		Late stage	
Shore type	Rocky shore	Sandy shore	Sandy shore	Estuarine sediments
<i>Phaeocystis</i> effect on the physical and/or biological environment	Decrease in the light penetration and NH4 ⁺ availability in the water column	Active cells deposits in calm conditions	Settlement of foam including active phytoplanktonic cells and bacteria	Anoxic conditions due to a limited O ₂ penetration into the sediment
Consequences on the benthic communities composition/functionning	Low primary production and low growth rate for underlying macroalgae	Increase in the sediment primary production	Increasing community primary production and respiration. NH4 ⁺ release	Rapid dramatic decrease in the species richness and density Macrofauna directly accessible to shorebirds. Energy flow modification
Recovery after the bloom	Rapid	Rapid	Rapid	Slow