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FEATURE ARTICLE

Population structure and environmental niches of *Rimicaris* shrimps from the Mid-Atlantic Ridge

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ABSTRACT: Among the endemic and specialized fauna from hydrothermal vents, *Rimicaris* shrimps constitute one of the most important and emblematic components of these ecosystems. On the Mid-Atlantic Ridge, 2 species belonging to this genus co-occur: *R. exoculata* and *R. chacei* that differ in their morphology, trophic regime and abundance. *R. exoculata* forms large and dense aggregations on active vent chimney walls in close proximity to vent fluid emissions, whereas *R. chacei* is much less conspicuous, living mostly in scattered groups or solitary further away from the fluids. However, the recent revision of *Rimicaris* juvenile stages from the Mid-Atlantic Ridge shows that *R. chacei* abundance would be higher than expected at these early life stages. Here, we describe and compare the population structure of *R. exoculata* and *R. chacei* at the Snake Pit and Trans-Atlantic Geotraverse (TAG) vent fields. We show distinct population demographics between the 2 co-occurring shrimp species with a large post-settlement collapse in *R. chacei* populations suggesting high juvenile mortality for this species. We describe important spatial segregation patterns between the 2 species and their different life stages. Additionally, our results highlight distinct niches for the earliest juvenile stages of both *R. exoculata* and *R. chacei*, compared with all other life stages. Finally, we discuss the potential factors, including predation and competitive interactions, that could explain the differences we observed in the population structure of these 2 species.

KEY WORDS: Hydrothermal vents · Life cycle · Population structure · Environmental niches · Juvenile mortality · *Rimicaris* shrimps



Rimicaris exoculata shrimp adults (light grey color) and juveniles (red color) at Snake Pit vent field (Beehive edifice) on the Mid-Atlantic Ridge.

Photo: BICOSE2 (2018)-Nautile@ifremer

1. INTRODUCTION

Fueled by the chemosynthetic energy arising from the mixing between vent fluids and surrounding seawater, hydrothermal vent ecosystems host dense and lush communities of endemic fauna which are organized in distinct species assemblages along steep thermal and chemical gradients. The dynamics of each assemblage depend both on local processes affecting birth and survival rates through local biotic and abiotic interactions, and on broader-scale processes influencing connections between spatially distant assemblages through individual exchanges (e.g. larval dispersal) (Mullineaux et al. 2018). Within this meta-community framework, species survival over genera-

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tions thus depends on the dispersal of larvae, their ability to locate suitable habitats, successfully settle on the bottom and finally enter existing benthic populations or develop into new ones. This implies adaptations at each stage of the life cycle of vent species.

As in other marine benthic ecosystems (Menge 1991), population demographics of species from hydrothermal vents depend on the number of competent larvae reaching the vent site, the probability of success of their settlement and any mortality events occurring between their successive life stages (Kelly & Metaxas 2008). Therefore, any biotic or abiotic factor affecting one of these 3 steps will likely impact the capacity of individuals to reach sexual maturity and the size of the adult populations. These environmental factors act differentially depending on the life stages considered, as each stage can occupy its own ecological niche and differs in its vulnerability to the different mortality factors (Werner & Gilliam 1984).

Whereas high juvenile mortality, mostly related to size-selective predation, is often characteristic of the post-settlement phase in shallow-water species (Gosselin & Qian 1997), a similar generalization has not yet been established in hydrothermal vent ecosystems. This can be attributed to the difficulty of conducting long-term population surveys and temporal observations along appropriate timescales in such remote places. Settlement and recruitment patterns of these species have thus mostly been inferred from cohort monitoring. According to these studies, populations of some vent species such as *Bathymodiolus azoricus* mussels or some gastropods are largely dominated by small-sized individuals (Comtet & Desbruyères 1998, Kelly & Metaxas 2008), implying either an important juvenile mortality or large variations between the different recruitment events. Conversely, a limited number of juveniles has been reported for several other taxa, including alvinellid polychaetes and many pelto-spirid or lepetodrilid gastropods (Zal et al. 1995, Faure et al. 2007, Matabos & Thiebaut 2010, Marticorena et al. 2020), suggesting frequent arrival of larvae at the vent field followed by a fast growth rate and an accumulation within the adult populations. Still, spatial or temporal limitation in sampling could also bias our understanding of post-settlement processes in these populations, as different life stages sometimes inhabit different parts of the field. Important spatial zonation between juveniles and adults has been reported for instance in *Kiwa tyleri* crabs, in which adults aggregate near the fluid exits

while juveniles remain restricted to the periphery of the field (Marsh et al. 2015). Similarly, segregation by size was observed along a thermal gradient in *B. azoricus* mussels, with juveniles and small mussels preferentially found in areas with lower temperatures (Husson et al. 2017).

Rimicaris shrimps are important components of hydrothermal vent ecosystems, as they constitute some of the most visually dominant megafauna in several regions such as in the Central Indian Ridge (Watanabe & Beedessee 2015), the Mid-Cayman Rise (Plouviez et al. 2015) or the Mid-Atlantic Ridge (MAR) (Zbinden & Cambon-Bonavita 2020). In the northern part of the MAR, 2 species of this genus co-occur: *R. exoculata*, which forms large aggregations close to vent fluid emissions, with densities of several thousands of individuals per m², and *R. chacei*, presumed to be less abundant than its congener (several tens to hundreds of individuals per m²), as individuals are less conspicuous within the vent fields (Segonzac et al. 1993).

It is assumed that population demography of congeneric species should be comparable especially when they evolve in similar habitat conditions (Münzbergová 2013). However, observations of contrasting population sizes and distribution patterns in many biologically similar congeners suggest that distinct population dynamics can exist even in cases of syntopy when species jointly occupy the same habitat at the same time (Münzbergová 2013, Jones & Ricciardi 2014, Bouchemousse et al. 2017). A stable co-occurrence at a small spatial scale of closely related species usually presupposes some differentiation in resource use as well as mechanisms that prevent interbreeding (Beer-mann & Franke 2012). When congeners display similar biological traits and ecological niches, competitive processes are expected to be intensified (Franke et al. 2007, Rius et al. 2009), although they do not necessarily constitute a major driver of their population dynamics (Bouchemousse et al. 2017). The 2 *Rimicaris* species differ by their trophic regime, with *R. exoculata* being highly dependent on the dense and diversified chemosynthetic episympiotic communities hosted in their enlarged cephalothorax (Gebruk et al. 2000, Ponsard et al. 2013, Zbinden & Cambon-Bonavita 2020) and *R. chacei* relying on symbiotrophy, bacterivory and scavenging (Gebruk et al. 2000, Apremont et al. 2018). This mixotrophic regime of *R. chacei* has been hypothesized to result from the competition for space with *R. exoculata* that would maintain *R. chacei* at a distance from vent fluids essential for fueling their symbionts (Apremont et al. 2018).

Zonation patterns between *R. exoculata* life stages have been reported in different studies (Segonzac et al. 1993, Shank et al. 1998, Gebruk et al. 2010) which highlighted the existence of patches of red juveniles adjacent to the dense aggregations of adults in several MAR vent fields. More recently, spatial segregation between males, found at the periphery away from the fluid mixing zones, and females, mainly occurring within the dense aggregations on active chimney walls, have also been observed (Hernández-Ávila et al. preprint: <https://doi.org/10.1101/2021.06.27.450066>). In addition, dense aggregations of small juveniles were also collected close to peripheral and diffuse fluid emissions arising from cracks at the Trans-Atlantic Geotraverse (TAG) vent field (Methou et al. 2020, Hernández-Ávila et al. preprint). Surprisingly, these individuals were revealed to be juvenile stages of *R. chacei*, indicating a misidentification in the previous species affiliation of these life stages (Methou et al. 2020, Hernández-Ávila et al. preprint). Such a large abundance of *R. chacei* early life stages was unexpected considering the low abundance of the adults, and implied for *R. chacei* either a low post-settlement survival or a larger adult population than previously stated, or both. Therefore, a full characterization of the population structures of *R. exoculata* and *R. chacei* is required to properly assess the demography of each species and whether the differences in the relative abundance of life stages could be related to their different trophic strategies or to other external factors such as predation, or physiological stress arising from their environment.

In this study, we used the case of *R. exoculata* and *R. chacei* to explore mechanisms allowing the co-existence of populations of 2 congeneric species within vent habitats, where trophic resources are abundant but very restricted spatially. We hypothesized that different demographic patterns, as well as niche partitioning may both be involved in the co-existence of these vent shrimps. To test these hypotheses, we conducted a detailed analysis of population structures of each species comparing their global size-frequency distributions within 2 vent fields along the northern MAR, namely TAG and Snake Pit. We also examined the distribution of sexes and life stages within visually distinct types of shrimp assemblages for which we also obtained temperature as well as chemical data. This allows us to characterize the realized environmental niches of each species through juvenile development to the adult stage, and discuss how biotic and abiotic factors may influence their demography, ultimately providing insights into mechanisms potentially involved in their co-existence.

2. MATERIALS AND METHODS

2.1. Field sampling

Rimicaris shrimps were collected on board the RV 'Pourquoi pas?' during the BICOSE 2 cruise (<http://dx.doi.org/10.17600/18000004>) between 26 January and 10 March 2018 at the TAG (26° 08.2' N, 44° 49.5' W, 3620 m depth) and the Snake Pit (23° 22.1' N, 44° 57.1' W, 3470 m depth) hydrothermal vent fields using the suction sampler of the HOV 'Nautile'. Samples were collected all over the active venting areas at the TAG vent field and on the 3 active edifices of Snake Pit: the Beehive, the Moose and the Nail (see Fig. S1 in the Supplement at www.int-res.com/articles/suppl/m684p001_supp/). In total, we collected 3919 *R. exoculata* shrimps (1533 from TAG and 2386 from Snake Pit) and 1201 *R. chacei* shrimps (886 from TAG and 315 from Snake Pit).

Spatially discrete samples were targeted to explore the fine-scale variations in the distribution of populations of each species, from 10s of centimeters for high-density areas to several meters in sparsely populated areas, with 13 different spatial samples at TAG and 17 at Snake Pit (Table S1). These samples were classified according to the type of vent fluid emissions and their proximity to them. During the dives, each sample was kept separate using the different chambers of the suction sampler carousel and the closeable bioboxes of the HOV 'Nautile'. All shrimps were stored in 80 % EtOH. For 20 of the sampling points, temperature measurements were carried out prior to sampling shrimps, moving the HOV temperature probe across several points (0.1°C precision for a temperature below 50°C) within the targeted sampling area, for 3 to 6 min 45 s with a sampling frequency of 1 Hz.

Additionally, a more precise environmental characterization of the vent fluid chemistry was obtained for 14 of these sampling points with pH, iron (Fe^{2+} and total iron, Fe_{tot}) and H_2S concentration measurements. Water samples were collected with the PEPITO sampler implemented on the HOV 'Nautile' (Sarradin et al. 2008) and equipped with blood bags (Terumo®). The suction inlet was coupled to the HOV temperature probe. Prior to use, all equipment used for sampling was rigorously washed 3 times with diluted hydrochloric acid (pH 2; Suprapur, Merck) and then thoroughly rinsed with ultrapure water (Milli-Q element system). Immediately after the recovery of the HOV, samples were processed in the chemical lab onboard (clean lab, P 100 000; ISO8), and pH was measured on a subsample using a Metrohm pH

meter. Measurements were carried out after calibration of the pH meter with NBS buffers (pH 4, 7 and 10) at 25°C. Total iron (Fe_{tot}) concentration was determined onboard on water samples taken by PEPITO using the Ferrozine spectrophotometric method (Stookey 1970). Free inorganic sulfides ($\Sigma\text{S} = \text{HS}^- + \text{S}^{2-} + \text{H}_2\text{S}$; Le Bris et al. 2000) were also determined onboard on recovered water samples using the spectrophotometric method of Cline (1969).

In addition to the PEPITO sampling, total dissolvable Fe^{2+} was measured *in situ* using a chemical miniaturized analyzer (CHEMINI; Vuillemin et al. 2009). The *in situ* measurement was based on flow injection analysis with a colorimetric detection (methylene blue method). A calibration of the analyzer was performed *in situ* at the beginning and at the end of each HOV dive using Fe(II) stock solutions. Hydrothermal samples were pumped without any filtration, and the signal acquisition (~3 min) was initiated at the same time as the water sampling since the same inlet was used for both PEPITO and CHEMINI. Finally, the *in situ* concentration of dissolved oxygen was monitored during each sampling using an optode (Aanderaa).

2.2. Life stage identifications and measurements

For each individual, carapace length (CL) was determined to the nearest 0.1 mm by Vernier calipers from the rear of the eye socket to the rear of the carapace in the mid-dorsal line. Each shrimp was identified and sorted according to its life stage and sex. Sex was identified in adults by the occurrence of the appendix masculina on the second pleopod in males, and by the shape of the endopod of the first pleopod (Komai & Segonzac 2008). Since these sexual characters appear only at the adult stage, sex of juvenile/subadult specimens could not be determined.

Juvenile stages A and B of *R. exoculata* and juvenile stage A of *R. chacei* were identified according to Methou et al. (2020). Because pleopod morphologies are identical between subadults and small females, the distinction between them was made according to the onset of sexual differentiation (OSD) size defined previously for *R. exoculata* and *R. chacei*, respectively, for a CL of 10 and 5.98 mm, respectively (Methou et al. 2020, Hernández-Ávila et al. preprint: <https://doi.org/10.1101/2021.06.27.450066>). All individuals without male characteristics were sorted as females if their size exceeded OSD, whereas individuals with a smaller size than OSD were sorted as subadults. Subadults and juvenile stages were some-

times considered together as immature individuals in the analyses.

The dataset compiling size, sex and life stage identifications of each shrimp individual is available in the SEANO database (<https://doi.org/10.17882/84192>).

2.3. Statistical analysis of population structure

Rimicaris shrimps were grouped first by species and then by life stage for each vent field. Visual examination of our dataset and Shapiro-Wilk normality tests revealed that the size distributions of the shrimps collected during the BICOSE 2 expedition did not follow a Gaussian distribution. Therefore, non-parametric tests were used for intergroup comparisons, using a Mann-Whitney test to compare differences in size between vent fields (Snake Pit vs. TAG), species (*R. exoculata* vs. *R. chacei*) or sex (females vs. males). Size frequency distributions between groups were compared with Kolmogorov-Smirnov 2-sample goodness of fit tests. Significant variations in the proportions of the 2 *Rimicaris* species or in proportions of immature to adult individuals for each species were tested using a χ^2 test with a Yates correction for 1 df. Frequencies of males and females in samples were tested for significant variation from a 1:1 sex ratio using a χ^2 test with a Yates correction for 1 df (detailed p-values for these tests are given in Table S2). The different spatial samples were grouped by *k*-means hierarchical clustering based on a Euclidean distance matrix and displayed with a heatmap. The optimal number of groups was determined both visually and statistically using the gap statistic method (Tibshirani et al. 2001). All tests were performed using R version 3.6.1 (R Core Team 2020).

2.4. Cohort analysis of the population structure

Histograms of the size frequency distributions were analyzed as mixtures of normal distributions for samples containing at least 40 individuals. Size-frequency histograms were plotted using a 1 mm size class interval. This interval was chosen to meet the following criteria described by Jollivet et al. (2000): (1) most size classes have at least 5 individuals each, (2) adjacent empty size classes have been minimized, and (3) the interval is clearly larger than the estimated error rate in measurements. We determined the overlapping component distribution that gives the best fit to the histogram with the 'mixdist' R package (Macdonald & Du 2012). Our modal decomposi-

tions were considered as valid according to the test from the 'mixdist' package, which is based on the χ^2 approximation to the likelihood ratio statistic. This allowed the identification of gamma components and their parameters, i.e. mean, sigma and estimated proportion, each corresponding respectively to the mean CL size, the standard deviation of the CL size and the proportion of a defined cohort.

2.5. Thermal and environmental niches of each species, sex and life stage

Four temperature descriptors were calculated for each of the 20 population samples where temperature was measured: the mean (Avg.T), the minimum (Min.T), the maximum (Max.T) and the standard deviation (Std.T) (Table S3). Four additional environmental measures (pH, H₂S, Fe²⁺ and Fe_{tot} concentrations) were used for 14 of these samples (Table S3) for a second, more refined, niche analysis. Values below the detection limit of the measurement devices were considered as null for the niche analysis. Pearson's correlations were used to select non-redundant variables.

The niches of males, females, subadults and juveniles of each shrimp species were examined similarly to the study of thermal niches of species encountered in *Bathymodiolus azoricus* mussel beds at the Lucky Strike vent field (Husson et al. 2017). The niche of a set of specimens (i.e. males, females, subadults or juvenile stages) was studied using the outlier mean index (OMI) created by Dolédec et al. (2000), and computed using the 'niche' function in the R package 'ade4' (Dray & Dufour 2007). On a PCA of environmental parameters, sampling points corresponding to the different population samples collected are weighted by the number of specimens. The center of gravity of each of these weighted points is the average position of a group of individuals, defined here by species and life stage or sex, in the scatterplot defined by the PCA, i.e. the average environmental conditions in which this group thrives. Different indexes are given by the OMI analysis, including the OMI index, also called 'marginality'. A permutation test is used to check the significance of this index, indicating if species marginality is significantly higher than expected by chance. If this test is not significant, this means that the species distribution is independent of the environmental conditions. The OMI analysis also gives a measure of the niche breadth: the tolerance index (Tol), which corresponds to the variance around the centroid. Associated with tolerance (Tol), the residual tolerance (RTol) index, defined as the part of the variance that is not

explained by the environmental variables used in the PCA, indicates whether the chosen variables are suitable for the niche analysis. These 3 indexes (OMI, Tol and RTol) constitute the total inertia of the niche, and each can be expressed as a percentage of this inertia. At the same time, OMI analysis also identifies the variables that best differentiate the environmental niches of the studied group of individuals.

3. RESULTS

3.1. Population structures of *Rimicaris* shrimps at TAG and Snake Pit in February–March 2018

At each vent field, overall size distributions of the 2 *Rimicaris* species were different from each other with significantly distinct size-frequency distributions both at TAG (Kolmogorov-Smirnov 2-sample test, $D = 0.949$, $p < 0.001$) and at Snake Pit vent fields ($D = 0.545$, $p < 0.001$). Unlike *R. exoculata*, *R. chacei* size distribution was indeed skewed towards the smaller sizes, corresponding to the juvenile and subadult life stages (Fig. 1). This resulted in a significantly higher proportion of immature individuals for *R. chacei* compared with *R. exoculata*, both at TAG (*R. chacei*: 91.9%, *R. exoculata*: 23.8%; $\chi^2 = 1036.4$, $p < 0.001$; Table 1) and at Snake Pit (*R. chacei*: 52.1%, *R. exoculata*: 32.6%; $\chi^2 = 45.74$, $p < 0.001$; Table 1). Immature individuals represented a significantly higher proportion of the whole *R. chacei* population at TAG compared with Snake Pit ($\chi^2 = 240.3$, $p < 0.001$), whereas immature individuals were significantly more represented in Snake Pit populations of *R. exoculata* compared with TAG ($\chi^2 = 34.2$, $p < 0.001$). Nonetheless, a similar proportion of *R. chacei* stage A juveniles were observed in the populations of both vent fields (TAG: 41.2%, Snake Pit: 38.1%; $\chi^2 = 0.79$, $p = 0.373$). However, *R. exoculata* stage A juveniles were still in lower proportions at the TAG vent field (TAG: 3.7%, Snake Pit: 13.9%; $\chi^2 = 108.03$, $p < 0.001$).

Overall, larger sizes were observed for *R. exoculata* compared with *R. chacei* at any life stage, either for adults, subadults or juvenile stages (Table S4; Mann-Whitney test, $p < 0.001$, all cases). The largest recorded adults of the 2 species were nonetheless of similar size at TAG (20.1 and 20.5 mm CL respectively for *R. exoculata* and *R. chacei*) but were bigger for *R. exoculata* at Snake Pit (24.4 and 21.1 mm CL respectively for *R. exoculata* and *R. chacei*). Males of *R. exoculata* exhibited significantly larger average sizes than females in both vent fields (TAG: $W = 24\,460$, $p < 0.001$; Snake Pit: $W = 80\,537$, $p < 0.001$).

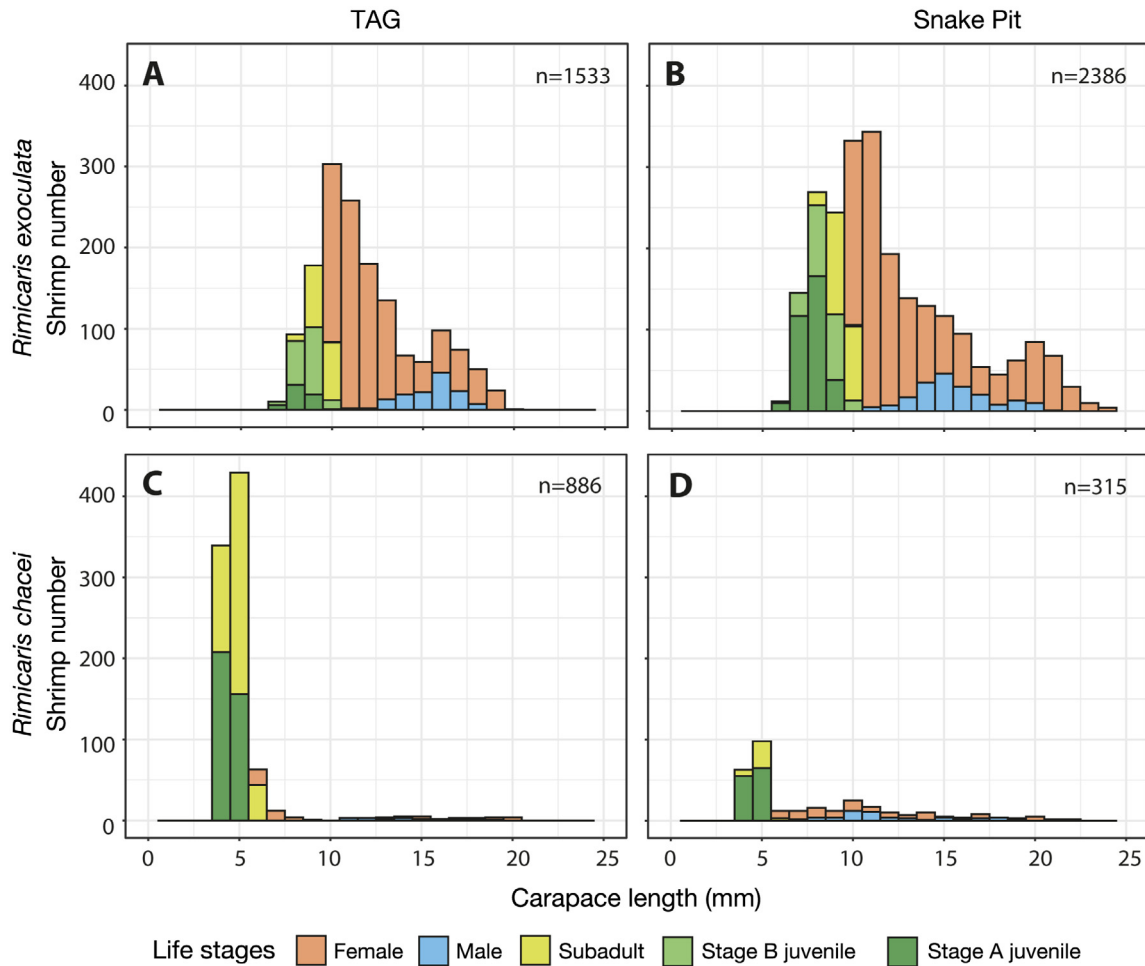


Fig. 1. Size-frequency distribution of *Rimicaris exoculata* and *R. chacei* from 2 vent fields (Trans-Atlantic Geotraverse [TAG] and Snake Pit), denoting life stages and sexes: (A) *R. exoculata* from TAG, (B) *R. exoculata* from Snake Pit, (C) *R. chacei* from TAG, (D) *R. chacei* from Snake Pit

However, the largest recorded *R. exoculata* adults were females both at TAG and at Snake Pit (20.1 and 24.4 mm CL, respectively). In *R. chacei*, the average size of males and females was not significantly different at Snake Pit ($W = 2180$, $p = 0.065$), and males were slightly larger than females on average at TAG ($W = 249.5$, $p = 0.001$). As with *R. exoculata*, the

largest recorded *R. chacei* adults were females at both vent fields (20.5 and 21.8 mm CL, respectively).

At each vent field, MIX modal decomposition detected 4 cohorts in the populations of *R. exoculata* and 3 cohorts in the populations of *R. chacei* (Fig. S2). In *R. exoculata*, the second and third cohorts had the highest proportions of individuals, roughly two-

Table 1. Sample and population data of *Rimicaris exoculata* and *R. chacei* from the Trans-Atlantic Geotraverse (TAG) and Snake Pit vent fields, Mid-Atlantic Ridge. Significance levels: ** $p < 0.01$, *** $p < 0.001$; -: not applicable

Species	Vent field	Total specimens	Adults			Immatures				Sex ratio TM:TF	χ^2 (1 df)	Significance
			Total	Female (TF)	Male (TM)	Total	Subadult	Juvenile B	Juvenile A			
<i>Rimicaris exoculata</i>	TAG	1533	1168	1033	135	365	155	153	57	0.13:1	690.41	***
	Snake Pit	2386	1609	1415	194	777	232	212	333	0.14:1	926.56	***
<i>Rimicaris chacei</i>	TAG	884	72	53	19	812	448	-	364	0.36:1	16.06	***
	Snake Pit	315	151	95	56	164	44	-	120	0.59:1	10.07	**

thirds of the populations, both at TAG and at Snake Pit (Table 2). These 2 cohorts mostly included adults with some immature stages below the OSD in the second cohort. The first cohort only comprised immature individuals and represented at most 1/5 of the total population. The fourth cohort comprised the larger adults, and represented at most 1/5 of the total population. In *R. chacei*, however, the first cohort that included mostly immature individuals comprised half of the population at Snake Pit and up to 92.5% of the population at TAG. The second cohort of *R. chacei* comprised small adults mostly, with proportions varying between 3 and nearly 40% of the whole population. The third cohort comprised large adults, representing only 4–12% of the total population (Table 2).

Populations of the 2 *Rimicaris* species also differed by their sex ratio. As reported for 2014 (Hernández-Ávila et al. preprint: <https://doi.org/10.1101/2021.06.27.450066>), sex ratios of *R. exoculata* collected in 2018 from TAG and Snake Pit were strongly biased towards females at both vent fields (TAG: $\chi^2 = 690.41$, $p < 0.001$; Snake Pit: $\chi^2 = 926.56$, $p < 0.001$; Table 1). Similarly, sex ratios of the *R. chacei* collected at these vent fields deviated significantly from 1:1 (TAG: $\chi^2 = 16.06$, $p < 0.05$; Snake Pit: $\chi^2 = 10.07$, $p < 0.001$; Table 1). However, this deviation from a 1:1 sex ratio was less marked in *R. chacei* populations compared with *R. exoculata*, with a much larger proportion of males among adult individuals at TAG (*R. chacei*: 26.4%, *R. exoculata*: 11.6%; $\chi^2 = 12.38$, $p < 0.001$) and at Snake Pit (*R. chacei*: 37.1%, *R. exoculata*: 12.1%; $\chi^2 = 68.92$, $p < 0.001$). For each species, sex ratios did not vary significantly between populations of the 2 vent fields (*R. chacei*: $\chi^2 = 0.12$, $p = 0.732$; *R. exoculata*: $\chi^2 = 2.04$, $p = 0.153$).

3.2. Spatial variation in the population structure of *Rimicaris* shrimps from TAG and Snake Pit

Our 30 samples were assigned to 6 visually distinct types of assemblages (Fig. 2) according to the domi-

nant *Rimicaris* life stage/species/sex and the proximity of the sampled assemblage to the hydrothermal activity. More specifically, these assemblages were classified as close to active emissions for those adjacent to black smokers, close to diffuse emissions for those in a close proximity to transparent shimmering flows, or in inactive areas when no vent fluid could be detected visually in their surroundings. Six clusters were also determined statistically to be the optimal number of clusters for our population dataset using the gap statistic method (Fig. 3).

The most visually dominating assemblage type was the dense aggregations observed close to the active vent fluid emissions (Fig. 2A; Fig. S3). These assemblages comprised almost exclusively *R. exoculata*, with *R. chacei* sometimes also present but in very low proportions (*R. chacei*: 0.9–10.4% of all *Rimicaris* shrimps; $\chi^2 = 93.5$ –714.1, $p < 0.001$). Most of the dense aggregations were dominated by adults of *R. exoculata*, with a larger proportion of adults in all samples (67.4–94.2% of all *R. exoculata* of a given sample, $n = 13$; $\chi^2 = 47.1$ –501.1, $p < 0.001$), except in one from Snake Pit (The Nail edifice) that exhibited similar proportions of adults and immature individuals ($\chi^2 = 0.46$, $p = 0.49$). The majority of immatures from this population sample were at an advanced juvenile stage (i.e. subadults or stage B juveniles), with early juveniles only representing 2.6% of all the *R. exoculata* individuals. Large variations were observed in the proportions of immature stages between the different samples from dense aggregations from 5.8 to 32.6% of all the *R. exoculata* individuals. For nearly all of the dense aggregation samples (13 out of 14 samples), sex ratio of *R. exoculata* strongly deviated from 1:1 ($\chi^2 = 31.2$ –311, $p < 0.001$), with a large bias toward females. One exception was found in a dense aggregation sample from TAG exhibiting an equilibrated sex ratio for *R. exoculata* ($\chi^2 = 0.69$, $p > 0.05$).

At the periphery of the dense aggregations, in parts less impacted by the turbulent mixing between hydrothermal fluid and seawater, or where fluid exits

Table 2. Cohort analysis with proportions (%), mean and SD of the modal components estimated from the size-frequency distributions of *Rimicaris exoculata* and *R. chacei* collected from TAG and Snake Pit vent fields. –: not applicable

Species	Vent field	Cohort 1			Cohort 2			Cohort 3			Cohort 4		
		%	Mean	SD	%	Mean	SD	%	Mean	SD	%	Mean	SD
<i>Rimicaris exoculata</i>	TAG	14.5	8.6	0.6	25.2	10.3	0.6	39.6	11.9	1.3	20.7	16.5	1.3
	Snake Pit	21.7	7.9	0.7	33.5	10.5	0.9	33.6	13.7	2.4	11.2	20.4	1.4
<i>Rimicaris chacei</i>	TAG	92.5	4.7	0.4	3.3	6.8	0.3	4.2	15.3	3.2	–	–	–
	Snake Pit	48.9	4.6	0.3	39.4	9.4	2.5	11.7	17.4	2.4	–	–	–

were dampened, a second type of assemblage was observed at Snake Pit (The Beehive), also largely dominated by *R. exoculata* (Fig. 2B; Fig. S4A). Unlike dense aggregations, these 2 population samples

exhibited large proportions of immatures ($\chi^2 = 202.2$, $p < 0.001$; 231.2 , $p < 0.001$) representing 90.4 and 91.3% of all *R. exoculata* individuals collected there. These assemblages were considered nurseries of *R.*

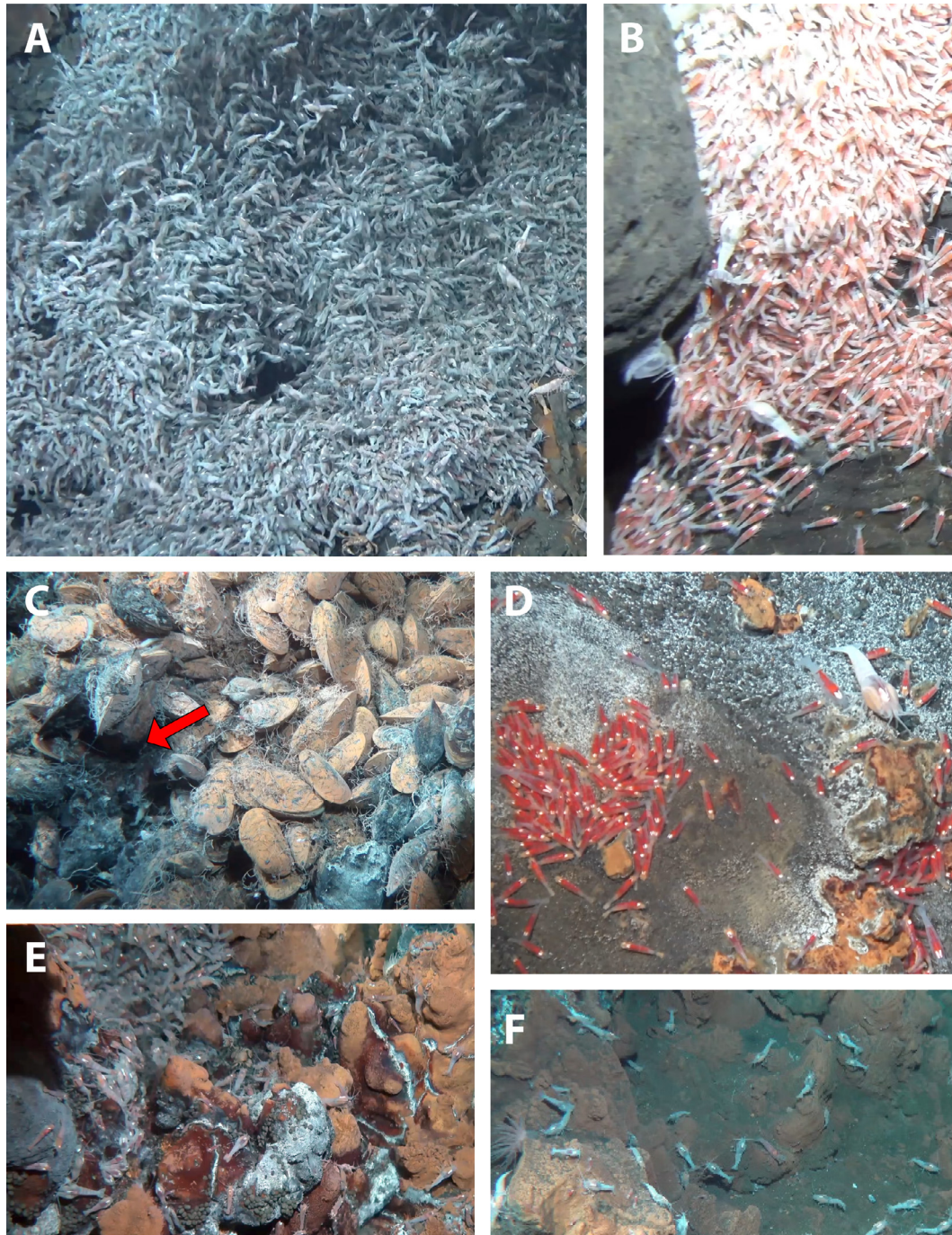


Fig. 2. Overview of the different *Rimicaris* assemblage types from TAG and Snake Pit vent fields. (A) Dense aggregation of *R. exoculata* adults observed near vent fluid exits at TAG. (B) Nursery of *R. exoculata* adjacent to the dense adult aggregation on the flank of chimneys at Snake Pit. (C) Hidden aggregation of *R. chacei* (red arrow) observed behind a mussel bed at Snake Pit. (D) *R. chacei* nursery associated with low temperature diffusions at TAG. (E) Low-density alvinocaridid assemblage at the periphery of a dense aggregation of *R. exoculata* at TAG. (F) Scattered *Rimicaris* individuals, in areas away from any visible vent fluid emissions at TAG

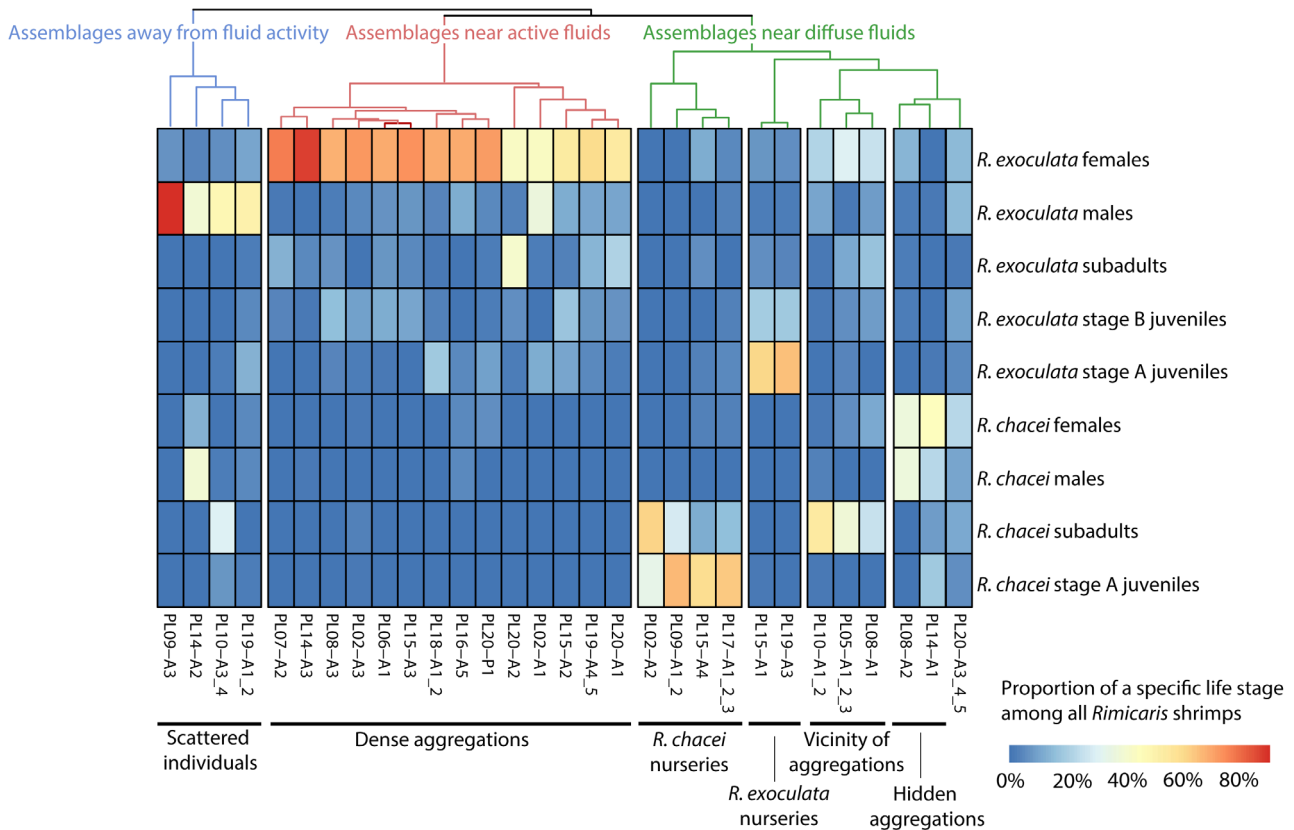


Fig. 3. Heat map displaying hierarchical clustering of the different samples collected in this study, according to their composition (in terms of species, sexes and life stages). Each column of the heat map corresponds to one sample with its ID provided below

exoculata. Most of these immature individuals were at an early juvenile stage, with stage A juveniles representing 70.4 and 74.5% of all the immature individuals. In addition, the sex ratio of the few *R. exoculata* adults from these assemblages did not significantly deviate from 1:1 ($\chi^2 = 3.2$, $p = 0.07$) or was only slightly biased towards females ($\chi^2 = 5.4$, $p = 0.02$). As for dense aggregations, *R. chacei* shrimps were nearly absent from *R. exoculata* nurseries, with only 2 *R. chacei* juveniles collected in total.

Two other types of shrimp assemblages observed close to diffuse vent fluid emissions were dominated by *R. chacei* ($\chi^2 = 16.5$ – 888.1 , $p < 0.001$) representing 77.4–99.5% of all the *Rimicaris* shrimps collected. These were generally not directly adjacent to the dense aggregations, but clearly separated even if the distance could be as short as a few tens of centimetres. One of these 2 types of assemblages was collected in a hole between rocks at TAG or between *Bathymodiolus puteoserpentis* mussels at Snake Pit (The Moose edifice). Although individuals were rather densely packed, they were hard to see due to the relatively small size of the assemblages and their ‘hidden’ localization; these assemblages were there-

fore called ‘hidden aggregations’ of *R. chacei* (Fig. 2C; Fig. S4B). These hidden aggregations clearly exhibited larger proportions of *R. chacei* adults than immatures ($\chi^2 = 35.6$ – 60.1 , $p < 0.001$), with adults representing 71.7% of all *R. chacei* individuals collected there. The sex ratio of *R. chacei* from these assemblages was strictly equilibrated for the sample collected at TAG ($\chi^2 = 0$, $p = 0.99$) but slightly deviated toward females for the one collected at Snake Pit ($\chi^2 = 7.45$, $p = 0.006$).

The second type of assemblage dominated by *R. chacei* was observed close to diffuse emissions exiting from cracks at the bottom of vent chimneys or in flat areas of the vent fields (Fig. 2D; Fig. S4C). Contrary to the hidden aggregations, significantly larger proportions of immature individuals were observed in these nurseries of *R. chacei* ($\chi^2 = 36.8$ – 868.5 , $p < 0.001$), representing at least 96.7% of all individuals of the species. In general, most of these immature individuals were at an early juvenile stage, with stage A juveniles representing 69.9–82.6% of all immature individuals. One *R. chacei* nursery from TAG exhibited a larger proportion of *R. chacei* subadults than stage A juveniles.

A last assemblage type found in areas of moderate fluid influence was observed at TAG in the vicinity of dense aggregations (Fig. 2E; Fig. S4D). These assemblages were characterized by similar proportions of the 2 *Rimicaris* species ($\chi^2 = 2.9$, $p > 0.05$) or only slightly larger proportions of one of the 2 species ($\chi^2 = 7.5\text{--}7.9$, $p < 0.05$). Larger proportions of *R. exoculata* adults ($\chi^2 = 24\text{--}30.1$, $p < 0.001$) and larger proportions of *R. chacei* immatures ($\chi^2 = 56.4\text{--}208.9$, $p < 0.001$) were usually observed in these assemblages. However, in one of these assemblages, similar proportions of adults and immature individuals were found for both species ($\chi^2 = 0.4\text{--}2.5$, $p > 0.05$). In all of them, most of the immature shrimps were at an advanced juvenile stage for both species, stage A juveniles representing at most 7.3 and 2.4% of all immature individuals respectively for *R. exoculata* and for *R. chacei*. In general, sex ratios of *R. exoculata* from these assemblages did not significantly deviate from 1:1 ($\chi^2 = 0.1\text{--}3$, $p > 0.05$; $n = 2$), except in 1 assemblage with a large bias toward females ($\chi^2 = 117.4$, $p < 0.001$).

Rimicaris shrimps were also collected at the periphery of the vent chimneys, on inactive sulfides away from the vent fluid activity (Fig. 2F; Fig. S4E). These assemblages of scattered individuals exhibited either similar proportions of the 2 *Rimicaris* species ($\chi^2 = 0.1\text{--}2$, $p > 0.05$) or larger proportions of *R. exoculata* individuals ($\chi^2 = 26.4$, $p < 0.001$). As in dense aggregations, larger proportions of *R. exoculata* adults over immature individuals were observed in all of these assemblages ($\chi^2 = 16.2\text{--}44.6$, $p < 0.001$). The sex ratio of *R. exoculata* from these assemblages deviated from 1:1 ($\chi^2 = 6.4\text{--}16.3$, $p < 0.001$), with a large bias toward males (81.8–93.3% of all *R. exoculata* adults). For *R. chacei*, both larger proportion of adults ($\chi^2 = 20.2$, $p < 0.001$) and larger proportions of immature individuals ($\chi^2 = 28.9$, $p < 0.001$) were reported in these 2 different population samples. Unlike *R. exoculata*, the sex ratio of *R. chacei* from these assemblages did not significantly deviate from 1:1 ($\chi^2 = 3$, $p = 0.11$).

One sample collected from a shrimp assemblage in a flat area with moderate fluid emission away from an active venting chimney at Snake Pit (The Nail edifice), exhibited no clear structure corresponding to one of the 6 assemblages defined above. Rather, it clustered with the hidden aggregations of *R. chacei* adults (Fig. 3). However, unlike samples from hidden aggregations, this population showed similar proportions of both *Rimicaris* species ($\chi^2 = 0.2$, $p > 0.05$). Moreover, proportions of adults and immature individuals were similar ($\chi^2 = 0.1\text{--}1.4$, $p > 0.05$), and sex ratio did not significantly deviate from 1:1 ($\chi^2 =$

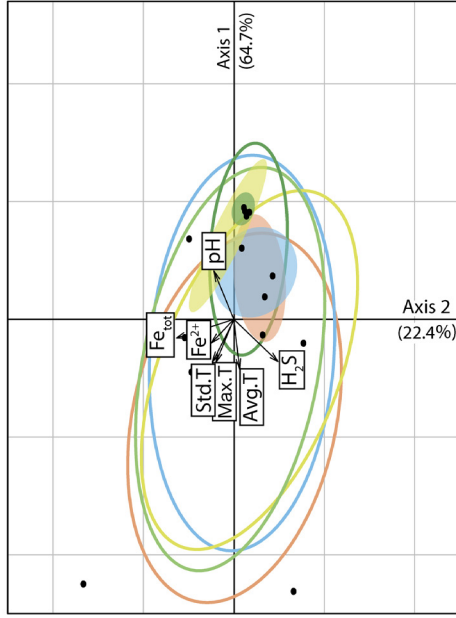
0.1–0.11, $p > 0.05$) both for *R. exoculata* and *R. chacei*. This assemblage also differed from the others by being dominated by *Mirocaris fortunata* (58.9% of all the shrimps sampled), another co-occurring alvinocaridid at these sites that was sporadically observed in very low proportions in the other samples.

3.3. Thermal and environmental niches of *Rimicaris* shrimps and their life stages from TAG and Snake Pit

Average temperatures around *Rimicaris* assemblages (i.e. mean of temporal observations at each location) ranged between 3.4 and 22.5°C, with maximum records reaching up to 38.1°C and minimum records down to 3.1°C. Standard deviations that can be considered as the thermal variability recorded in the assemblages varied from 0.3 to 6.9°C (with thermal recordings lasting 3 to 6 min 45 s). Regarding the vent fluid chemistry, average H₂S concentrations around *Rimicaris* assemblages were between 0 and 8.5 µM, with pH variations between 6.86 and 7.9. It is worth mentioning that total sulfide concentrations were measured onboard on recovered samples and may be underestimated compared with the actual *in situ* conditions. *In situ* iron concentrations in *Rimicaris* assemblages were highly variable, with Fe²⁺ concentrations between 0 and 239 µM and Fe_{tot} concentrations between 0 and 316 µM. As minimal temperatures were strongly correlated with average temperature ($r^2 = 0.97$, $p < 0.05$), only average temperature (Avg.T), temperature standard deviation (Std.T) and maximum temperature (Max.T) were used in the 2 PCAs. No clear significant correlation was observed between any of the thermal descriptors and the concentrations of the different reduced chemical elements (Pearson correlations, $p > 0.01$ in all cases).

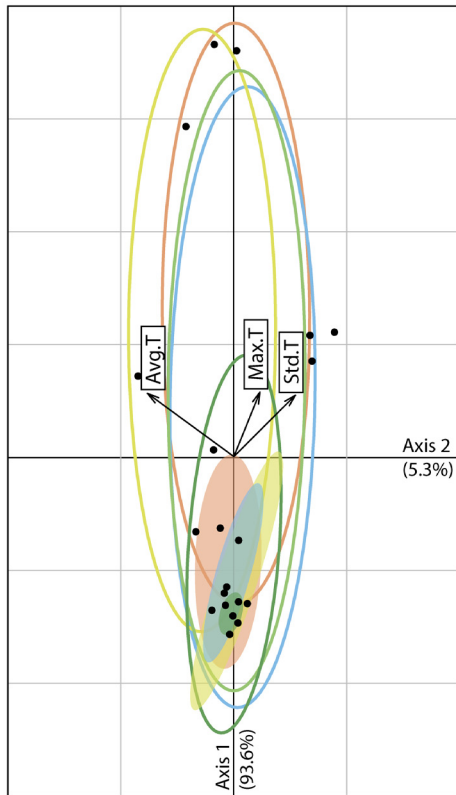
For the thermal niches (Fig. 4A,B; 20 sampling points), most of the variance was explained by the 2 first axes of the PCA representing respectively 93.6 and 5.3% of the inertia. All 3 temperature metrics were very well represented on the PCA ($\Sigma\cos^2 = 0.99$, 0.99 and 0.98 for Avg.T, Std.T and Max.T, respectively) and were strongly correlated with the first axis (cor = 0.96, 0.95 and 0.99 for Avg.T, Std.T and Max.T, respectively). Avg.T, Std.T and Max.T almost contributed equally to the construction of the first axis (contribution = 32.7, 32.4 and 34.9%, respectively). Similarly, Avg.T and Std.T also contributed equally to the construction of the second axis (contribution = 46.4 and 53.4%, respectively).

Environmental niches



C

Thermal niches



A

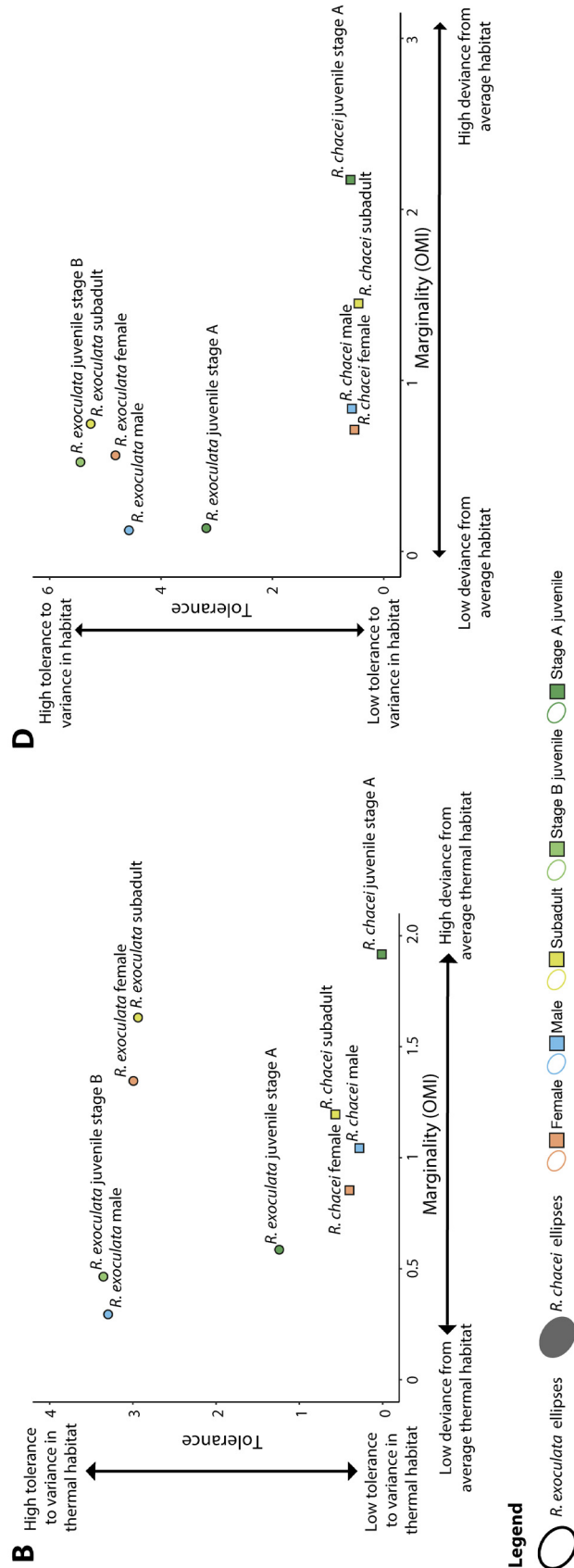


Fig. 4. Niches of *Rimicaris* shrimps according to their species, sex and life stages. Two-dimensional plot of the outlier mean index (OMI; marginality) analysis. Each ellipse represents the (A) thermal niche and (C) environmental niche of a *Rimicaris* group (defined by species, life stage and sex in adults). Each black point represents 1 assemblage. Also shown are total tolerance of their (B) thermal and (D) environmental niche

In the case of the environmental niches (Fig. 4C,D; 14 sampling points), a large part of the variance was explained by the 2 first axes of the PCA representing respectively 64.7 and 22.4% of the inertia. All environmental parameters were relatively well represented on the first 2 axes of the PCA, with $\Sigma\cos^2 > 0.85$ for all parameters except for Std.T and pH ($\Sigma\cos^2 = 0.81$ and 0.70 , respectively). Thermal descriptors as well as pH correlated more with the first axis (cor = -0.97 , -0.88 , -0.95 and 0.83 for Avg.T, Std.T, Max.T and pH, respectively) whereas H₂S measurements were more correlated with the second axis (cor = -0.83). Iron concentrations were equally correlated with the first (cor = -0.75 and -0.67 for Fe²⁺ and Fe_{tot} respectively) and the second axes of the PCA (cor = 0.62 and 0.64 for Fe²⁺ and Fe_{tot} respectively). The thermal descriptors contributed the most to the construction of the first axis (contribution = 21, 17 and 20.1% for Avg.T, Std.T and Max.T, respectively), followed by pH and Fe²⁺ and Fe_{tot} (contribution = 15.4, 12.5 and 10%, respectively). On the other hand, H₂S contributed in large part to the construction of the second axis followed by iron concentrations (contribution = 24.3 and 25.8% respectively for Fe²⁺ and Fe_{tot}).

As revealed by the OMI analysis, stage A juveniles of *R. exoculata* occupied a distinct thermal niche compared with those of their older counterparts (Fig. 4A), differing both in terms of niche breadth (lower tolerance [Tol] in stage A juveniles) and overall position (Fig. 4B). Although *R. exoculata* adults, subadults and stage B juveniles exhibited similar niche breadths, the mean niche position significantly deviated from the average sampled thermal conditions only for females and subadults (permutation tests: $p < 0.05$, Table 3), suggesting that the distributions of these two were driven by thermal conditions. Overall, the same trends were observed in the OMI analyses with both thermal and chemical descriptors, although significant marginality was only found for *R. exoculata* females ($p < 0.05$; Table 4).

Thermal niches of all *R. chacei* life stages were smaller than those of each *R. exoculata* life stage (Fig. 4A). Stage A juveniles of *R. chacei* displayed a smaller niche breadth as other *R. chacei* life stages, as well as a mean niche position that significantly deviated from the average sampled thermal conditions (Fig. 4B; permutation test, $p < 0.05$). Conversely, niche positions of *R. chacei* adults and subadults did not significantly depart from the average sampled thermal conditions (Fig. 4B). Environmental niches of *R. chacei* exhibited the same trends as those observed for thermal niches (Fig. 4C,D, Table 4).

Table 3. Thermal range and parameters of the thermal niche outlier mean index (OMI; marginality) analysis for the different *Rimicaris* groups. Tol: tolerance index (variance around the centroid); Rtol: residual tolerance (part of the variance not explained by the variables used in the PCA); omi, tol and rtol: % of the total inertia of their respective indexes. Perm: permutation. * $p < 0.05$, ** $p < 0.005$, NS: not significant ($p > 0.05$)

Rimicaris group	n	Average temperatures (°C)			Thermal variability (°C)			Thermal range (°C)			OMI analysis			Perm. test	p		
		mean	min	max	mean	min	max	min	max	inertia	OMI	Tol	Rtol			omi	tol
<i>R. exoculata</i> females	1508	12.8	3.4	22.5	3.6	0.3	6.9	3.1	38.1	4.85	1.63	2.94	0.28	33.7	60.6	5.7	**
<i>R. exoculata</i> male	272	10.1	3.4	22.5	2.9	0.3	6.9	3.1	38.1	3.90	0.29	3.30	0.31	7.5	84.5	8.0	NS
<i>R. exoculata</i> subadults	265	13.5	3.4	22.5	3.2	0.3	6.9	3.2	38.1	4.83	1.35	2.99	0.49	27.8	61.9	10.2	*
<i>R. exoculata</i> stage B juveniles	298	10.8	3.4	22.5	3.0	0.5	6.9	3.2	38.1	4.05	0.46	3.35	0.23	11.5	82.8	5.7	NS
<i>R. exoculata</i> stage A juveniles	330	6.0	3.4	22.5	1.3	0.3	6.9	3.1	38.1	1.93	0.59	1.24	0.10	30.4	64.4	5.2	NS
<i>R. chacei</i> females	61	5.6	3.6	14.5	1.0	0.5	4.8	3.3	20.7	1.30	0.85	0.40	0.04	65.8	30.8	3.4	NS
<i>R. chacei</i> male	36	5.1	3.4	9.7	0.9	0.3	3.9	3.1	20.7	1.36	1.04	0.28	0.03	76.8	20.7	2.5	NS
<i>R. chacei</i> subadults	283	4.7	3.4	14.5	0.9	0.3	3.9	3.1	20.7	1.85	1.19	0.57	0.08	64.4	31.0	4.6	NS
<i>R. chacei</i> stage A juveniles	391	3.9	3.4	8.2	0.5	0.3	3.9	3.1	20.7	1.93	1.91	0.02	0.005	98.9	0.8	0.3	*

Table 4. Chemical range and parameters of the environmental niche outlier mean in dex (OMI; marginality) analysis for the different *Rimicaris* groups. Abbreviations as in Table 3

<i>Rimicaris</i> group	n	Average temperatures (°C)		Thermal variability (°C)		Thermal range (°C)		H2S (µM)		Fe2+ (µM)		Fetot (µM)		OMI analysis					Perm. test P	
		mean	min	max	mean	min	max	min	max	mean	mean	mean	mean	inertia	OMI	Tol	Rtol	omi		tol
<i>R. exoculata</i> females	1249	12.1	3.4	20.3	3.4	0.3	5.9	3.1	38.1	3.7	4.1	87.9	9.01	0.56	4.82	3.63	6.2	53.5	40.3	**
<i>R. exoculata</i> male	188	9.2	3.4	20.3	2.2	0.3	5.9	3.1	38.1	2.7	25.6	71.6	8.17	0.12	4.58	3.47	1.5	56.0	42.5	NS
<i>R. exoculata</i> subadults	143	10.3	3.4	20.3	2.8	0.3	5.9	3.2	38.1	3.8	36.4	80.5	10.55	0.75	5.26	4.54	7.1	49.9	43.0	NS
<i>R. exoculata</i> stage B juveniles	263	10.1	3.4	20.3	2.9	0.5	5.9	3.2	38.1	2.7	35.2	79.2	9.75	0.52	5.45	3.78	5.4	55.9	38.8	NS
<i>R. exoculata</i> stage A juveniles	300	5.4	3.4	20.3	1	0.3	5.9	3.1	38.1	0.8	4.3	16.6	5.71	0.14	3.19	2.39	2.4	55.8	41.9	NS
<i>R. chacei</i> females	36	6.1	3.6	9.7	1.2	0.5	3.9	3.3	20.7	2.1	8.8	22.4	2.21	0.83	0.56	0.82	37.6	25.4	37.0	NS
<i>R. chacei</i> male	26	5.4	3.4	9.7	1.1	0.3	3.9	3.1	20.7	2.3	14.5	34.2	2.42	0.71	0.52	1.18	29.6	21.5	49.0	NS
<i>R. chacei</i> subadults	187	4.6	3.4	8.2	1.2	0.3	3.9	3.1	20.7	0	10.3	36.5	3.59	1.45	0.45	1.69	40.3	12.6	47.1	NS
<i>R. chacei</i> stage A juveniles	322	3.8	3.4	8.2	0.5	0.3	3.9	3.1	20.7	0.02	0.3	12.2	3.79	2.18	0.6	1.01	57.5	15.9	26.6	NS

For the thermal niches, residual tolerance (RTol) ranged from 0.3 to 10.2% of total inertia; thus, the variables used to describe these niches explained 89.8–99.7% of the niche distribution of the different *Rimicaris* life stages (Table 3). On the other hand, RTol for the environmental niches ranged from 26.6 to 49% of total inertia; thus, the variables used to describe these niches only explained 51–74.4% of the niche distribution of the different *Rimicaris* life stages (Table 4).

4. DISCUSSION

4.1. Small-scale distribution of *Rimicaris* life stages along an environmental gradient

Our temperature measurements indicate that thermal conditions around the different *R. exoculata* assemblages of TAG and Snake Pit ranged between 3.4 and 22.5°C, with maximal peaks of temperatures up to 38.1°C (Table 3). These results slightly extend the known thermal range occupied by these shrimps compared with previous *in situ* temperature measurements in *R. exoculata* aggregations, which reported average thermal conditions between 4.6 and 16°C, and up to 25°C at maximum (Desbruyères et al. 2001, Geret et al. 2002, Zbinden et al. 2004, Schmidt et al. 2008). Although *R. exoculata* experiences 100% of mortality when exposed for 1 h to 33°C in closed pressurized chambers, these shrimps tolerate shorter exposures of 10 min at 39°C with only 35% mortality (Ravaux et al. 2019). With a defined critical thermal maximum (CTmax) of 38°C for this species, it thus appears that *R. exoculata* occupies most of its fundamental thermal niche *in situ*. We also noticed important movements of shrimps in aggregations where average temperature exceed 20°C (samples PL06A1, PL15A3 and PL20A1) whereas the others tended to remain steady, confirming the *in vivo* behavioral observations of Ravaux et al. (2019).

In contrast, we observed that *R. chacei* inhabits a narrower thermal niche, between 3.4 and 14.5°C, up to 20.7°C at maximum, regardless of the life stage considered (Table 3). In the absence of experimental work on *R. chacei* thermotolerance, it is unclear if their more limited distribution within the hydrothermal vent gradient stems from a physiological limitation or from biological interactions such as competition with its congeners. This question can be answered in part by a comparison with *Mirocaris fortunata*, another alvinocaridid shrimp from the MAR, which has a thermal tolerance similar to that of *R.*

exoculata (Shillito et al. 2006) but occupies a more restricted thermal habitat (Desbruyères et al. 2001, Methou 2019). The narrower distribution of *R. chacei* could also simply be related to a lower thermal preference, as seen in many vent species that preferentially select habitats with cooler temperatures than their upper thermal limit (Bates et al. 2010). This all suggests that temperature per se is not a sufficient limiting factor to explain differences in alvinocaridid shrimp distributions.

Similarly, our dataset also points out that *R. exoculata* inhabits a wider range of reduced compound concentrations (traced through *in situ* total dissolvable iron) than *R. chacei* for a given life stage, although this trend was much less marked between the earliest juvenile stages of the 2 species. Our *in situ* chemistry measurements around biological communities indicated significant deviation from the conservative dilution model as has already been reported in previous studies (Le Bris et al. 2006, 2019, Nakamura & Takai 2014). This deviation from a conservative dilution of the endmember originates from the highly reactive nature of the sulfides and iron species and the consumption or production of these reduced compounds by microbial activity.

At both fields, an important zonation between adults and juvenile stages was also found for *R. exoculata* and for *R. chacei*. This spatial segregation along the thermal gradient is reflected in the distinct thermal niches observed between the earliest juveniles and the other life stages for the 2 *Rimicaris* species. Similar differences in distribution of bathymodiolin mussel life stages have been hypothesized to be related to intraspecific competition for space between adults and juveniles (Comtet & Desbruyères 1998, Husson et al. 2017). This was further supported by evidence of recruitment inhibition by high-density assemblages of adult mussels from the East Pacific Rise (Lenihan et al. 2008). Still, an influence of intraspecific competitive interactions among *Rimicaris* shrimps is tempered by their high motility as well as the non-negligible presence of early juveniles within several high-density assemblages dominated by adult stages. A progressive acquisition of a high thermal tolerance of these recruiting stages, as seen in post-larval stages of the intertidal shrimp *Palaemon elegans* (Ravaux et al. 2016), could also limit their upper thermal distribution. Additionally, distributions of *R. exoculata* and *R. chacei* juveniles further differ by their habitat topography, i.e. on steep chimney walls for *R. exoculata* nurseries and within flat areas or smooth slopes for *R. chacei* nurseries. As topography strongly influences the overall current hydrodynam-

ics, with low-topography areas generally associated with low mixing rates (Girard et al. 2020), these structural differences between habitats might reflect the relative dependency of each species on the vent fluid mixing gradient essential to fuel their ongoing developing symbiosis (Methou et al. 2020). Although our measurements were limited, this variability of the fluid chemistry between nursery habitats of the 2 species could be observed in our dataset with lower H₂S and Fe concentrations in *R. chacei* nurseries compared with *R. exoculata* ones.

We also confirmed the segregation pattern between males and females of *R. exoculata* observed at TAG in January–February 2014 (Hernández-Ávila et al. preprint: <https://doi.org/10.1101/2021.06.27.450066>) at each of the studied vent fields. Such a pattern was not observed in *R. chacei*, which also exhibited a more equilibrated sex ratio overall, more similar to what has been observed for its sister species *R. hybisae* (Nye et al. 2013).

Despite apparent relative stability evidenced by the maintenance of an assemblage type in the same area over the years, the spatial zonation of the different *Rimicaris* life stages also seems dynamic with a succession of assemblages in other areas on short temporal scales. During one of our dives at TAG, we witnessed the disappearance of an entire dense aggregation assemblage of *R. exoculata* adults replaced by an aggregate of *R. chacei* hidden between rocks in just about 10 d (Fig. S5). Unfortunately, we are currently unable to confirm if this succession of assemblages is related to a modification of the environmental conditions, as we could not obtain initial temperature measurements prior to the replacement of the dense aggregate, due to technical difficulties during the dive.

In addition to the assemblages dominated by a particular life stage of one of the 2 *Rimicaris* species, different types of mixed assemblages were also observed in areas a little further away from the fluids. Interestingly, these assemblages with a mixed proportion of the 2 species were also the assemblages with the lowest densities of *Rimicaris* individuals, visually countable in contrast to the other more dense assemblage types (Fig. 2). These more peripheral assemblages with less marked structuration could correspond to assemblages with much lower densities due to suboptimal habitat conditions for *Rimicaris* spp. Lower supply of reduced compounds to feed their cephalothoracic symbionts, or an absence of a symbiont pool to renew their symbiotic communities along their molt cycle, would limit the establishment of these species in these areas. A finer environmental characterization of these particular areas would be

required to better understand the subtle ecological differences that could exist between these habitats and the denser shrimp assemblage types. In any case, these density-dependent trends suggest that competitive processes might occur between *R. chacei* and *R. exoculata* life stages.

4.2. Population dynamics of *Rimicaris* shrimps from the MAR are driven by biotic interactions

Similar to most species from hydrothermal vents, *R. exoculata* and *R. chacei* exhibited multimodal size-frequency distributions, suggesting a discontinuous recruitment pattern in both species, thus confirming previous observations made for *R. exoculata* in 2014 (Hernández-Ávila et al. preprint). Still, population structures of the 2 congeneric *Rimicaris* species from the MAR differed widely with a large post-recruitment collapse of *R. chacei* populations at both vent fields. This collapse was even more pronounced for the TAG population, where immature individuals represented more than 90 % of all the individuals collected. This implies either a high mortality rate of *R. chacei* early post-settlement stages in comparison with *R. exoculata* and/or a large variability in the magnitude of the different larval settlement events. Although temporal tracking of these populations could not be performed, important aggregations of *R. chacei* juveniles in nurseries were reported between January and February 2014 (Hernández-Ávila et al. preprint) as well as between March and April 2017 (Methou et al. 2020). Therefore, it is less likely that the observed variations between *R. chacei* life stages abundance would be related to variations in larval supply and/or settlement success between years.

Therefore, we hypothesize that the post recruitment collapse observed in *R. chacei* populations results more probably from a higher juvenile mortality that could be explained by different, not necessarily mutually exclusive, factors (Fig. 5). Among them, predation could have a major impact on *R. chacei* demography, as it has been recognized as one of the most important causes of juvenile mortalities in shallow-water ecosystems (Gosselin & Qian 1997). The more peripheral and exposed distribution of *R. chacei* juveniles within the vent fields could expose them to a higher predation pressure by populations of *Maractis rimicarivora* anemones found in these areas and particularly abundant at TAG (Copley et al. 2007). Indeed, we witnessed a juvenile from the *R. chacei* nursery area being caught by these anemones during a dive at TAG in 2014 (Video S1 at www.int-res.com/

[articles/suppl/m684p001_supp/](http://www.int-res.com/articles/suppl/m684p001_supp/)). Predation by these carnivorous anemones on *R. exoculata* adults has also been observed (Van Dover et al. 1997) and was proposed to affect male individuals from the peripheral areas (Hernández-Ávila et al. preprint) which could explain, at least partially, the large sex ratio bias toward females of this species. Additionally, a larger vulnerability to predation by macrourid fishes could also preferentially affect *R. chacei* juveniles. Despite a limited observation time in these areas, we recorded 2 predation events by these fishes within or close to *R. chacei* nursery assemblages during our samplings at TAG, both in 2014 and in 2018 (Video S1). This higher predatory pressure at TAG would also be in agreement with the greater collapse observed in the populations of *R. chacei* at this site compared with those from Snake Pit (Fig. 1). In contrast, the distribution of *R. exoculata* early juveniles, sheltered on the flanks of large chimney structures or within the dense aggregations of *R. exoculata* adults, probably protects them from this high predatory pressure. These shelters for *R. exoculata* juveniles could also arise from the different chemical regime of those aggregations, acting as a possible deterrent from 'non-vent' vagrant predators such as macrourid fishes.

In addition to their distinct spatial distribution, the juvenile stages of both *Rimicaris* shrimps differed in their average settlement size, i.e. twice as large for *R. exoculata* compared with *R. chacei* (Methou et al. 2020). Numerous studies on marine fishes and shrimps have reported body size as a key factor in the survival of early juveniles, with a lower overall mortality for larger individuals (Meekan et al. 2006, Gagliano & McCormick 2007, Mace & Rozas 2018). Thus, survival of *Rimicaris* juveniles could fit into the 'bigger is better' hypothesis, whether larger body size would provide them a lower vulnerability against predation, a better physiological state and/or a higher competitive ability. As an alternative to high juvenile mortality, important emigration of *R. chacei* juveniles, i.e. secondary migration, to a different location following their initial settlement at the vent fields could also lead to similar patterns in their population structure. Long-term monitoring with video imagery of the areas of *Rimicaris* nurseries or individual tracking of early-stage shrimps could help to refine our hypotheses and determine more precisely the relative influence of these different factors on their population dynamics.

Interspecific competition for space between the 2 species, already suggested by their spatial segregation patterns, could also have a major influence on their demography. It is commonly accepted that

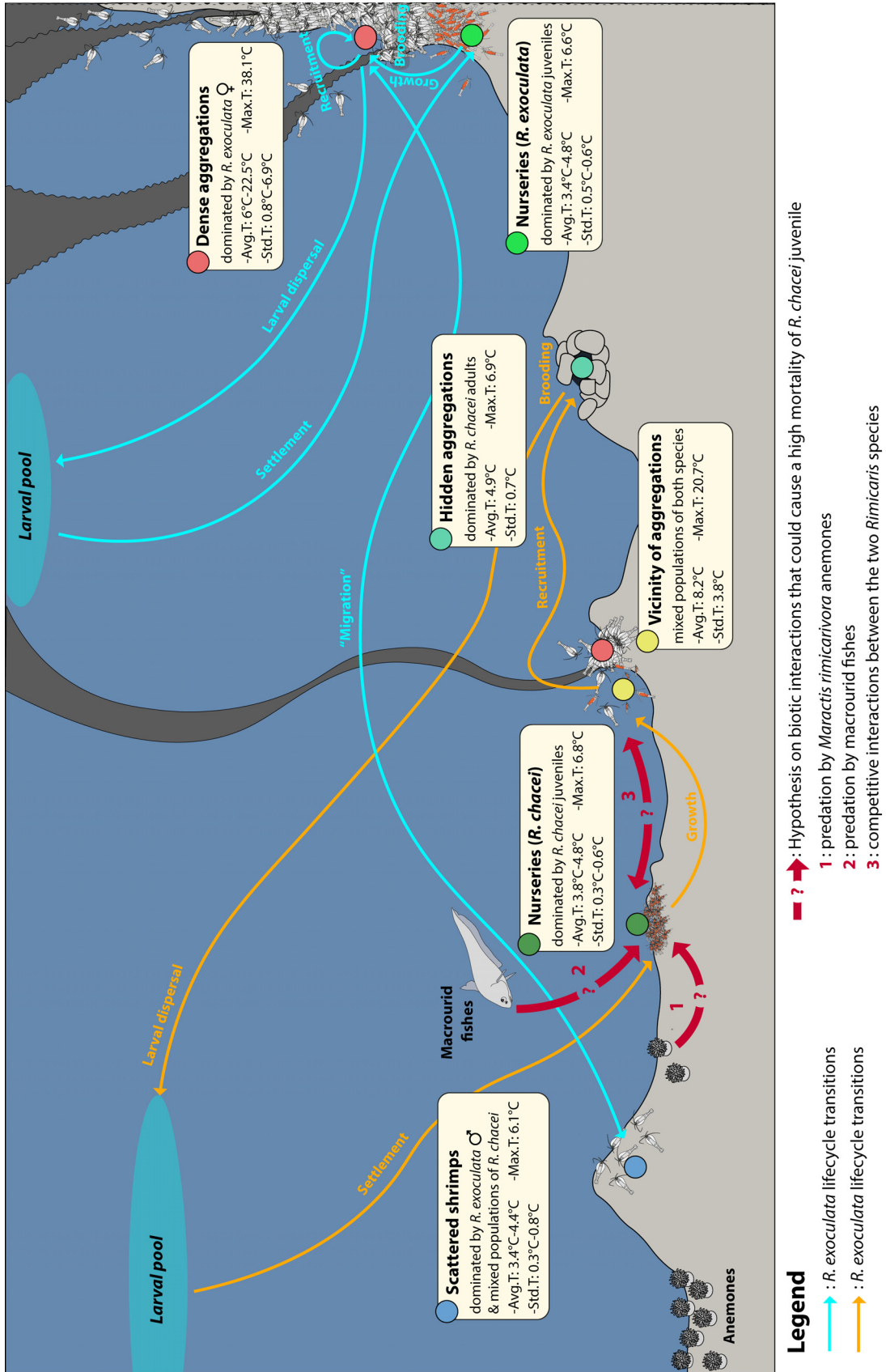


Fig. 5. Schematic view of the distribution of different *Rimicaris* assemblages at hydrothermal vents from the Mid-Atlantic Ridge with current hypotheses on the factors affecting their population dynamics. Avg. T: mean temperature; Max. T: maximum temperature; Std. T: standard deviation of temperature

abundant species within an assemblage outcompete less common ones. However, this general statement has been contradicted by various manipulative studies which did not report any effect of being naturally rare or common on the competitive ability of a species (Angel et al. 2006, Matias et al. 2012). Still, marine lobsters can experience important demographic bottlenecks before recruitment, due to the more specific requirements of their early benthic juvenile stages for sheltered areas, which are limited in space (Wahle & Steneck 1991). In a similar fashion, the low abundance of *R. chacei* adults might result from a lower competitive ability of their early juvenile stages to access areas of active fluid emissions, as suggested by their narrower thermal niche compared with similar stages of *R. exoculata* (Fig. 4).

In hydrothermal vents, competitive processes for space are confounded with competition for food in symbiotrophs whose nutrition depends entirely on their access to the vent fluid reduced elements. Both *R. exoculata* and *R. chacei* are fueled by large communities of filamentous ectosymbionts (Ponsard et al. 2013, Apremont et al. 2018), which progressively make up during their ontogeny an increasing part of their diet after settlement (Methou et al. 2020). A more limited access to areas of active fluids for *R. chacei* juveniles could potentially lead to a limitation to primarily acquire their episybionts and develop a sufficient nutritive symbiotic community within their cephalothorax, which could thus cause mass mortality in these populations. In this scenario, the mixotrophic behavior of *R. chacei* would be the result of a past competition with *R. exoculata* leading to a trophic niche differentiation in *R. chacei* adults over evolutionary timescales, allowing the stable coexistence of the 2 species in the northern MAR vent fields. Additionally, the more limited local diversity at TAG (Desbruyères et al. 2001), likely related to the absence of *Bathymodiolus* mussel beds known to house large macrofaunal diversity (Desbruyères et al. 2001, Rybakova & Galkin 2015), enhances even more the limitation in alternative food resources for *R. chacei*, thus possibly explaining the larger post-recruitment collapse of their population at this site. These shrimps are not the first case of population demographics driven by competitive interactions for food in hydrothermal vent ecosystems. Populations of *Paralvinella pandorae* from Juan de Fuca Ridge exhibited similar post-settlement collapses when found in sympatry with their congener *P. palmiformis*, whose isotopic niches largely overlap (Levesque et al. 2003).

This link between nutrition modes and population demographics of these shrimps is also further sup-

ported when their population structures are compared with those of *R. hybisae* from the Mid-Cayman Spreading Center. Although COI divergence was initially stated between the latter and *R. chacei* (Nye et al. 2012), probably because of inaccurate species identifications in public genetic databases, these 2 species are indeed so close phylogenetically that they cannot be distinguished on the sole basis of the genetic markers used to date (Vereshchaka et al. 2015). In this way, *R. hybisae* populations from the Mid-Cayman Rise can almost be seen as a case of *R. chacei* allopatry with no known competitors like *R. exoculata*. However, *R. hybisae* size-frequency structures largely differ from those of *R. chacei* and appear to be more similar to those of *R. exoculata* (Nye et al. 2013).

It should be noted that, unlike *R. exoculata*, *R. chacei* adults were more difficult to catch with the suction sampler during our attempts to collect them, as they sometimes escaped the tip of the sampler when we approached them. This escape behavior was not observed for the juvenile specimens of the same species inhabiting the nurseries. It is therefore possible that our dataset might slightly underestimate the exact proportion of *R. chacei* adults in the field. Nonetheless, it is unlikely that this potential underestimation of *R. chacei* adult numbers would overturn the extremely marked population structure observed for this species at both vent fields. Similarly, we believe that the disequilibria in sampling effort between the different assemblage types, with an overrepresentation of the dense aggregations against the other assemblage types, is representative of their higher occurrence within the hydrothermal vent fields. A detailed cartography giving the distribution of these assemblages, with shrimp density estimations, when possible, over the entire vent fields would help to better assess our dataset representativeness and if our sampling strategy could have partially affected our conclusions on the population dynamics of these 2 species. Temporal monitoring of these animals would also be necessary to clarify if their population dynamics could not be affected, at least in part, by long-term cyclicities.

The mechanisms by which *R. chacei* populations maintain a high supply of settlers despite a limited adult number comparatively to *R. exoculata* are still unclear. Larger *R. chacei* populations from yet unknown vent fields in the MAR representing source populations, with TAG and Snake Pit as sink populations, at the metapopulation level or differences in reproductive output between *R. exoculata* and *R. chacei* are possible. Additional work on the repro-

ductive biology of *R. chacei* will be needed to better understand differences in the life history traits of these species. Further experimental studies on the thermal physiology and tolerance to the vent fluid toxic compounds of both *R. chacei* adults and *Rimicaris* juvenile stages would also be required to understand if the distribution patterns observed between these different life stages are related to physiological limitations or other biotic factors.

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