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RESEARCH

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Unveiling the hidden economic toll of biological invasions in the European Union

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Abstract

Background Biological invasions threaten the functioning of ecosystems, biodiversity, and human well-being by degrading ecosystem services and eliciting massive economic costs. The European Union has historically been a hub for cultural development and global trade, and thus, has extensive opportunities for the introduction and spread of alien species. While reported costs of biological invasions to some member states have been recently assessed, ongoing knowledge gaps in taxonomic and spatio-temporal data suggest that these costs were considerably underestimated.

Results We used the latest available cost data in *InvaCost* (v4.1)—the most comprehensive database on the costs of biological invasions—to assess the magnitude of this underestimation within the European Union via projections of current and future invasion costs. We used macroeconomic scaling and temporal modelling approaches to project available cost information over gaps in taxa, space, and time, thereby producing a more complete estimate for the European Union economy. We identified that only 259 out of 13,331 (~ 1%) known invasive alien species have reported costs in the European Union. Using a conservative subset of highly reliable, observed, country-level cost entries from 49 species (totalling US\$4.7 billion; 2017 value), combined with the establishment data of alien species within European Union member states, we projected unreported cost data for all member states.

Conclusions Our corrected estimate of observed costs was potentially 501% higher (US\$28.0 billion) than currently recorded. Using future projections of current estimates, we also identified a substantial increase in costs and costly species (US\$148.2 billion) by 2040. We urge that cost reporting be improved to clarify the economic impacts of greatest concern, concomitant with coordinated international action to prevent and mitigate the impacts of invasive alien species in the European Union and globally.

Keywords Projection, *InvaCost*, Monetary impacts, Invasion costs, Temporal trends, Missing data

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Background

Invasive alien species—those introduced into regions where they are not native and in which they cause negative impacts to nature and/or society—are among the main drivers of global biodiversity decline [1, 2], and considered a growing threat with multiple feedbacks to ecosystem services and human well-being [3–6]. The ecological and socio-cultural impacts of invasive alien species are substantial and expected to continue increasing due to climate [7], geopolitical [8], and economic changes [9].

Most assessments of the impacts of invasive alien species have relied on scoring systems, such as those proposed by Panetta [10] and subsequently updated by Kumschick et al. [11] and Vilà et al. [12]. However, scoring systems are often context-dependent, with some scoring criteria lacking objectivity or requiring advanced assessor expertise [13], rendering them unsuitable for broader generalisation or prediction. There is therefore a need to improve the reliability of impact assessments by considering monetary costs arising from the loss and damage of resources as well as management actions. Attempts to quantify the economic costs of invasive alien species at broad spatial scales have occurred in recent decades [14], although early attempts had considerable shortcomings [15] or were limited to economic inventories of specific sectors [16, 17].

Most recently, the *InvaCost* project and the structured and publicly available database it has produced [18–20], have provided an unprecedented opportunity to investigate taxonomic and geographic trends in the monetary costs of invasive alien species worldwide (e.g., [21, 22]). Studies with a national (e.g., [23]), regional (e.g., [24]), continental (e.g., [25]), taxonomic (e.g., [26]), or pathway (e.g., [27, 28]) focus have, however, all identified difficulties with making comprehensive monetary cost estimates owing to spatial and taxonomic gaps. These data gaps include: (i) costs for many known, established invasive alien species that have not been assessed; (ii) cost information that is often inaccessible or otherwise not available publicly, or (iii) many cost estimates that do not fulfil quality criteria (i.e., peer-reviewed and/or transparent calculations).

In Europe, the monetary costs incurred by invasive alien species from 1960 to 2020 were recently estimated to be US\$140.2 billion [22], although persistent data gaps suggest that this is an underestimate [29]. This inferred cost might appear low relative to the annual gross domestic product of some European countries (e.g., 3.3% when compared to Germany's 2021 gross domestic product [GDP] of US\$4.3 trillion; data.worldbank.org). However, it is a more substantial quantity when considering

countries with smaller economic turnover such as Hungary (US\$182.3 billion) or Slovakia (US\$115.0 billion).

The combined economic turnover of European Union member states makes it the third largest economy worldwide (data.worldbank.org), and Europe as a whole has been a historic centre of trade, human migration, and tourism. As such, both continental Europe and the European Union (with its current 27 member states) have been particularly vulnerable to biological invasions, despite recent transboundary legislation to tackle invasive alien species [30–33]. The European Union Invasive Alien Species Regulation 1143/2014 [34] directs efforts to tackle the threats presented by invasive alien species to the European Union. However, many such species, including the 88 identified species of Union concern (of which 12 are thought not to be present yet in the Union's territory), lack any evidence of economic costs [19], despite the European Union being an economic and monetary union with coordinated economic and fiscal policies (e.g., a common monetary policy and currency). The problem is compounded because there are already several thousand established alien species in the European Union [35], with large discrepancies between the total number of established invasive alien species and those with assessed costs (for instance, France has a conservative minimum of 2621 alien species, but only 98 species [<4%] with reported costs; [23]).

An absence of information on economic impacts risks downplaying the threats posed by invasive alien species (including costs and other impacts), because it underestimates the economic threat biological invasions represent and hinders national policies, governance, research, and educational efforts along with broad-scale multinational initiatives to combat the problem [36–38]. Improved quantification of these species' impacts is also essential for prioritisation, mitigation, and eradication that underlie the Convention on Biological Diversity 2020 [39], the 2022 Kunming-Montreal Global Biodiversity Framework, and the European Union Invasive Alien Species Regulation 1143/2014 [34]. However, even when implemented, such actions are most often locally focused, and are frequently of insufficient length or magnitude to address the scale of the problems because of budget limitations. This lack of allocated funding compromises outcomes while simultaneously increasing both long-term management and damage costs [40, 41]. Thus, the feasibility of management remains impaired by the political choice not to ensure adequate resources [42].

Because cost information for many invasive alien species is missing, national and transnational estimates would benefit from a more complete accounting. Given that invasion history is one of the best predictors of eventual impacts [43], using existing cost estimates of a

measured species would seem the most appropriate and straightforward approach to interpolate missing costs in other invaded countries. However, context is also important, and we would not necessarily expect impacts to be identical across countries [5]. For instance, macroeconomic differences could influence the cost of labour (and management) or the per-unit currency value of a sector, and hence, the magnitude of damages. Alternatively, wealthier or larger countries could be better prepared to respond to invasions, resulting in inverse relationships between costs and macroeconomic indicators [44]. Given the challenges triggered by a lack of detailed cost information, we projected the available data to fill both spatial and taxonomic gaps to assess the extent to which we have so far underestimated the costs of invasive alien species to the economies of European Union member states. We also forecast how baseline invasion costs and factors correlated with costs (i.e., the number of costs reported in the literature and number of species with reported costs) will develop up to 2040 using several temporal modelling approaches. As a result, we develop a more comprehensive accounting of the costs of alien species to the European Union using invasion history and fitted macroeconomic scaling factors.

Our analyses are therefore important methodological and applied advances towards improving spatial and temporal estimates of costs. We expected that (i) the projected amount of costs of invasive alien species will be considerably higher than those reported in the *InvaCost* (v4.1) database due to (ii) mismatches between known invaders and those with recorded costs [23, 45–47], and (iii) expected increasing future trajectories of costs of invasions [19, 48, 49]. Our study makes a critical step towards improving cost predictions of biological invasions for the European Union as well as elsewhere and represents a substantial development beyond existing representations of their monetary impacts.

Methods

Cost data

To estimate the cost of biological invasions on the combined economy of the European Union, we used the latest available version of the *InvaCost* database (v4.1; [19]). *InvaCost* currently includes 13,553 entries of reported economic costs from invasive alien species retrieved from peer-reviewed, official, or grey-literature sources in both English and 21 other languages [45], and over 60 descriptor variables (i.e., impacted sector, type of cost, etc.). *InvaCost* (v4.1) has standardised individual cost records to a common currency and year to account for variation and inflation: 2017 US\$ (see [19] for detailed information on conversion; exchange rate for 2017: US\$1 = €0.8852; World Bank, 2022). We

applied the conversion (except for the temporal projection; see below) for the period following 1960, because we could not obtain monetary exchange rates from official institutions (e.g., World Bank) prior to that year. We then converted costs from 2017 US\$ to 2022 EU€ values using an inflation correction factor = 1.10545941 for the period between 2017 and 2022 and an average currency exchange rate for 2022 of US\$1 = €0.9515 (World Bank, 2022).

To ensure that we used cost data from only the 27 member states, we selected costs for every individual country currently in the European Union (excluding candidate countries), using the 'Official_country' column from the *InvaCost* database: Austria, Belgium, Bulgaria, Croatia, Cyprus, Czech Republic, Denmark, Estonia, Finland, France (excluding French overseas territories), Germany, Greece, Hungary, Ireland, Italy, Latvia, Lithuania, Luxembourg, Malta, Netherlands, Poland, Portugal, Romania, Slovakia, Slovenia, Spain, and Sweden. This filtering step identified 5,442 cost entries from the original global dataset (13,553 entries).

To obtain comparable costs of invasive alien species, we considered all costs lasting for a period of < 1 year as 'annual costs', and re-calculated (and thus 'expanded') costs covering multiple years on an annual basis by spreading the total cost across the time covered, using the *expandYearlyCosts* function in the *invacost* package version 0.3–4 [20] in R version 4.1.3 [50]. Deriving the total cumulative cost of invasions over time requires taking into account the probable duration of each cost occurrence. The duration is the number of years between the probable starting ('Probable_starting_year') and ending ('Probable_ending_year') years of the costs reported by each publication included in the *InvaCost* database [19].

Expansion resulted in 7,214 annualised, European Union-relevant cost entries. To exclude cost estimates with doubtful reliability, we first filtered the *InvaCost* database to obtain a 'highly reliable, observed' dataset of only 'observed' costs using the 'Method_reliability', the 'Method_reliability_refined' and 'Implementation' columns reflecting (i) the perceived reliability of the type of publication and/or cost-estimation approach ('high', when originating from peer-reviewed articles, official reports, or from grey material but with documented, reproducible and traceable methods, and 'low' otherwise), and (ii) whether the cost was realised or empirically incurred ('observed') or whether it was based on predictions or costs expected over time or space ('potential'). The resulting dataset of 'highly reliable' and 'observed' costs contained 5,592 entries. Further analyses focused on two descriptors: (i) *type of cost*: 'damage' refers to damage or losses incurred by invasion, and 'management' includes

control-related expenditures (i.e., costs due to monitoring, prevention, management, eradication); (ii) *impacted sector*: the activity, societal, or market sector incurring the cost (see Additional file 1).

Taxonomic gaps

To identify taxonomic gaps in the reporting of potentially costly invasive alien species, we (i) quantified the ratio between recorded invasive alien species in *InvaCost* and established alien species nationally within European Union member states, and (ii) assessed the number of species of Union concern (last update, Commission Implementing Regulation (EU) 2022/1203 of 12 July 2022 Amending Implementing Regulation (EU) 2016/1141 to Update the List of Invasive Alien Species of Union Concern. Available online: <https://eur-lex.europa.eu/legal-content/EN/TXT/HTML/?uri=CELEX:32022R1203&from=EN>) on invasive alien species recorded in *InvaCost*. For this, we assembled the most comprehensive dataset of established alien species in the European Union by combining the *SInAS* database of alien species occurrences [51] with Casties et al. [52], as well by adding species that were not yet included in either of the two lists. We used the *SInAS_AlienSpeciesDB_2.4.1* [51] file as the base file for our dataset. We removed species without assignment of invaded country/region, or those not reported in the European Union from the dataset. Then, we also removed species with assignment only as CASUAL and ABSENT in the columns 'degreeOfEstablishment' (N) and 'occurrenceStatus' (L), respectively, due to their unclear establishment status in those regions [53]. Finally, we checked species identity and the spelling of scientific names against the Global Biodiversity Information Facility [54]. If we did not find a species in the Global Biodiversity Information Facility, we did general internet searches in June and July 2022 to confirm species authenticity. We corrected misspelt species names and removed any duplicate species from the dataset.

We applied a paired *t*-test to compare the total number of recorded, established alien species and those of Union concern listed within the dataset of established alien species in the European Union (github.com/LeungEcoLab/EU_costs), with the respective total number of invasive alien species recorded in *InvaCost* at the level of individual European Union member states. We also removed species not authenticated and thus, included only species-specific entries from the *InvaCost* database, thereby excluding any mixed or unspecified data.

Spatial and sectoral projections

We used invasion history to interpolate costs to countries where a species was confirmed present using the dataset of established alien species in the European Union, but

where the cost was not estimated, and also tested for differences due to macroeconomic factors. We restricted our interpolations only to species and sectors (e.g., agriculture, health, management and control, etc.) where a given species' impact had been estimated in at least one European Union member state, and we only used highly reliable observed, country-level costs reported in the *InvaCost* database. We focused our analyses on country-level costs because they constitute most (92%) costs and are directly comparable. For full transparency, we also reported interpolated costs at the site level (see Additional file 1), although these are more difficult to compare because spatial scales of analyses varied widely, and no area information was available. Thus, our interpretation of interpolated costs at the site-level is speculative. After filtering for only highly reliable, observed costs, we obtained a total of 162 entries at the country level encompassing monetary costs of 49 species from 22 European Union member states. The site-level costs represented 2907 entries in 12 countries of the European Union for 179 species.

We also tested for differences resulting from macroeconomic factors and to account for potential socio-economic differences, we analysed the following: GDP, population size, country area, and importance of the impacted sector (in %). We expected costs to increase with GDP and population size because these are typical macroeconomic predictors with broad data availability, and they were relevant socio-economic predictors for the costs of invasive alien species in previous studies [22, 24, 55]. Further, the size (value) of a given sector logically relates to the potential magnitude of effect [56]. This last predictor scaled health and agriculture pest costs by the current health expenditure and the value added by agriculture, forestry, and fishery (both in % of GDP) (data.worldbank.org), respectively.

We scaled costs by the relative magnitudes of socio-economic factors, but costs need not change proportionally with macroeconomic factors, and a model should be flexible to account for this. For example, if the value of agriculture in country A was lower than in country B, the impact of an agricultural pest in A should also be fractionally, but not necessarily proportionally, lower than in B. We modelled relative differences in economic factors as a ratio (e.g., GDP_A/GDP_B), and accounted for non-proportional effects using a fitted coefficient (γ). In notation:

$$\hat{X}_{s,c_i,h,j} = \frac{1}{n} \sum_{k=1}^n X_{s,c_k,h,j} \prod_{m=1}^M \left(\frac{V_{m,c_i}}{V_{m,c_k}} \right)^{\gamma_m},$$

where X is the *InvaCost*-listed cost, \hat{X} is the interpolated cost based on averaged, scaled observed costs, s is species, c is country, h is type of cost (damage, management,

or both), and j is impacted sector. V denotes each of the M macroeconomic variables analysed. c_i refers to each country i where species s has been observed, where we are trying to interpolate, whereas c_k refers to the country where species k has been both observed and measured in *InvaCost*. To avoid unreasonable extrapolations beyond the data, we truncated scalar ratios by the maximum (or minimum) observed in the fitted model. The model allows for effects ranging from proportional scaling of costs with macroeconomics ($\gamma = 1$) to no effect ($\gamma = 0$), and to either positive ($\gamma > 0$) or negative ($\gamma < 0$) effects.

We fitted values of γ through maximum likelihood (using the *Optim* function in R), using the deviation between predicted costs (Eq. 1) and costs observed in *InvaCost* for species–sector combinations where there were entries across multiple countries. We considered alternative models using all possible combinations of our four macroeconomic variables and calculated the Akaike's information criterion (AIC) for each model (Additional file 1). Finally, we used the AIC-weighted average across all our models [57] to estimate the cost of each species–sector–country combination. We applied these values to species present in countries (Briski et al., unpublished data) but that were missing economic estimates. We estimated uncertainty in the parameter values (γ) as well as the resulting economic estimates through bootstrapping.

Temporal projection

To project the temporal dynamics of the monetary impacts at the European Union-level, we used the highly reliable and observed data subset. We then identified 'extreme' cost entries that can distort statistical analyses and violate model assumptions. These were any cost value exceeding the third quartile + 1.5 times the interquartile range for each year [21]. Using the *summarizeCosts* function implemented in the *invacost* package, we estimated the average cost for each decade in the time range 1970–2020 to visit the raw cost trends over time because the first highly reliable, observed cost entry occurred in 1970. Accordingly, we did this analysis with the highly reliable, observed dataset with ($n = 5592$ entries) and then without ($n = 4872$ entries) 'extreme' cost entries.

To identify periods in which the variance of costs reported within *InvaCost* (v4.1) changed over time, we applied a first-order autoregressive AR (1) process for the residuals [58]. This enabled the computation of derivatives of fitted splines using the method of finite differences to estimate the rate of change (slope) in the fitted smoother [59]. This produces diagnostic plots of the costs over time, where we could identify periods of non-random change(s) and superimpose them on the respective

temporal trend [60]. Following the non-random changes identified throughout the entire period, we identified a non-random change in costs after 1980 at which point a change in the trend is apparent and the few costs repeated over time prior to this point limited our ability to estimate the variance in costs. We therefore restricted this analysis to the period 1980–2017, excluding (i) the values reported from 2018 to 2020 due to lag times in reporting of costs after they are incurred [61, 62], and (ii) all years prior to 1980. We then modelled the annual total of this subset over time (using the 'Impact_year' of each cost entry as the year of occurrence) using both linear and quadratic robust regressions on the highly reliable, observed data via the *lmrob* function of the R *robust-base* package [63], with maximum iterations set to $n = 1,000$. Ultimately, we projected the trend in total costs until 2040, by using the *predict* function in the *forecast* R package [64]. Because we were interested in determining the projected trend of management expenditure in the European Union, we repeated the procedure considering only management costs—which constituted 90% of the entries with, and 93% of the entries without extreme values—by filtering the data using the 'Type_of_cost_merged' column. We therefore estimated both 'total' and 'management' costs separately and applied this analysis to our highly reliable observed data set and another subset without the identified extreme values (resulting in a final dataset of 3,685 cost entries). We did all analyses in R version 4.1.3 [50].

Results

Trends in documented costs

The total cost of biological invasions across all European Union member states from 1960 to 2020 reported in *InvaCost* v4.1 amounted to US\$129.9 billion (2017 value), corresponding to €138.6 billion (2022 value). Of these, US\$7.3 billion (€7.7 billion; 5.9%) was highly reliable and observed, with US\$752.9 million (€791.9 million; 10.28%) dedicated to management costs. Both total and highly reliable, observed costs varied substantially across European Union member states (Fig. 1; Additional file 1).

Including extreme values, the average annual cost from 1970 to 2020 was US\$143.9 million (€151.4 million) although this varied substantially by decade, falling from an initial US\$19.17 million (€20.17 million) annually in the 1970s to US\$0.5 million in the 1980s, before increasing to US\$53.3 million (€56.1 million) in the 1990s and ultimately reaching US\$461.1 million (€485.0 million) annually in the 2010s (Fig. 2a). Excluding extreme values (Fig. 2a), the average annual cost from 1970 to 2020 was US\$8.3 million (€8.7 million) and followed a similar decadal trend overall, increasing to US\$11.6 million (€12.2 million) annually in the 2000s before a decrease in

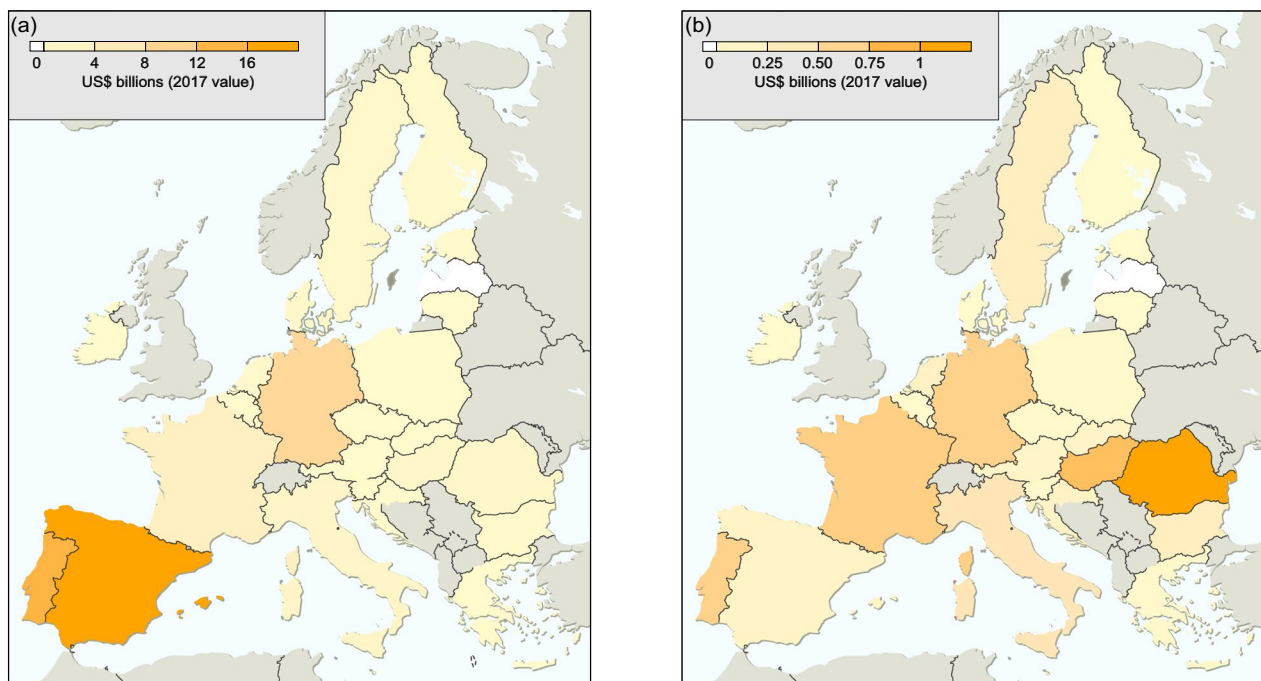


Fig. 1 **a** Total invasion costs across European Union member states. **b** Highly reliable, observed cost subset across European Union member states. Note the different scales

the 2010s. Decadal trends followed the projected change in reported costs over time, indicating the strongest increase between the 1980s and 1990s (Fig. 2b).

Taxonomic gaps

We identified 1459 (± 1011) established alien species on average recorded per European Union member state, ranging between 217 alien species established in Luxembourg and 3691 in France (Fig. 3; Additional file 1). In contrast, there was only an average of 15 (± 32) costly invasive alien species per European Union member state within *InvaCost* v4.1, ranging between one species in Luxembourg to 161 in Spain. This disparity implies that only ~1% of established alien species reported from European Union member states have recorded costs. Of the 88 species of Union concern, 66 are established in the European Union (Additional file 1), with 48 having reported economic cost in *InvaCost* (Additional file 1). On average, 21 (± 12) species of Union concern occur per European Union member state, but only 4 (± 8) have reported costs within that country. The number of established alien species by European Union member states recorded in our data and those listed as of Union concern were both statistically different from the member states' respective number of invasive alien species in *InvaCost* (Briski et al.: $t_{26} = 7.3786$, $p < 0.001$; European Union Regulation 1143/2014: $t_{45} = 6.0722$, $p < 0.001$; Fig. 3).

Spatial and sectoral taxonomic projections

Highly reliable, observed, species-specific costs totalled at US\$6.5 billion (€6.8 billion), from which country-level costs (from 49 species; $n = 162$ entries) totalled US\$4.7 billion (€4.9 billion; representing 89.8% of all highly reliable, observed costs entries for European Union member states). On this conservative basis of highly reliable, country-level costs, the different projections (Additional file 1) estimated the missing cost data to be ~869 cost entries across all European Union member states, contributing an additional US\$23.4 billion (€24.6 billion) with uncertainties around the estimated missing costs after model averaging of a minimum 8.39 billion and maximum 54.9 billion, summing to a total of US\$28.0 billion (€30.9 billion; +501%). However, when excluding two high-leverage points from the final dataset (reported annualised costs of >\$500 million as reported damage from Romania and Hungary), the averaged models (Additional file 1) estimated the missing costs for European Union member states at US\$7.2 billion (€7.6 billion) with uncertainties around the estimated missing costs after model averaging of a minimum 6.43 billion and maximum 20.26 billion. Adding the missing costs resulted in a total of US\$11.8 billion (€12.4 billion; +153%) in this scenario. For a breakdown of cost imputations excluding high-leverage points, see Additional file 1.

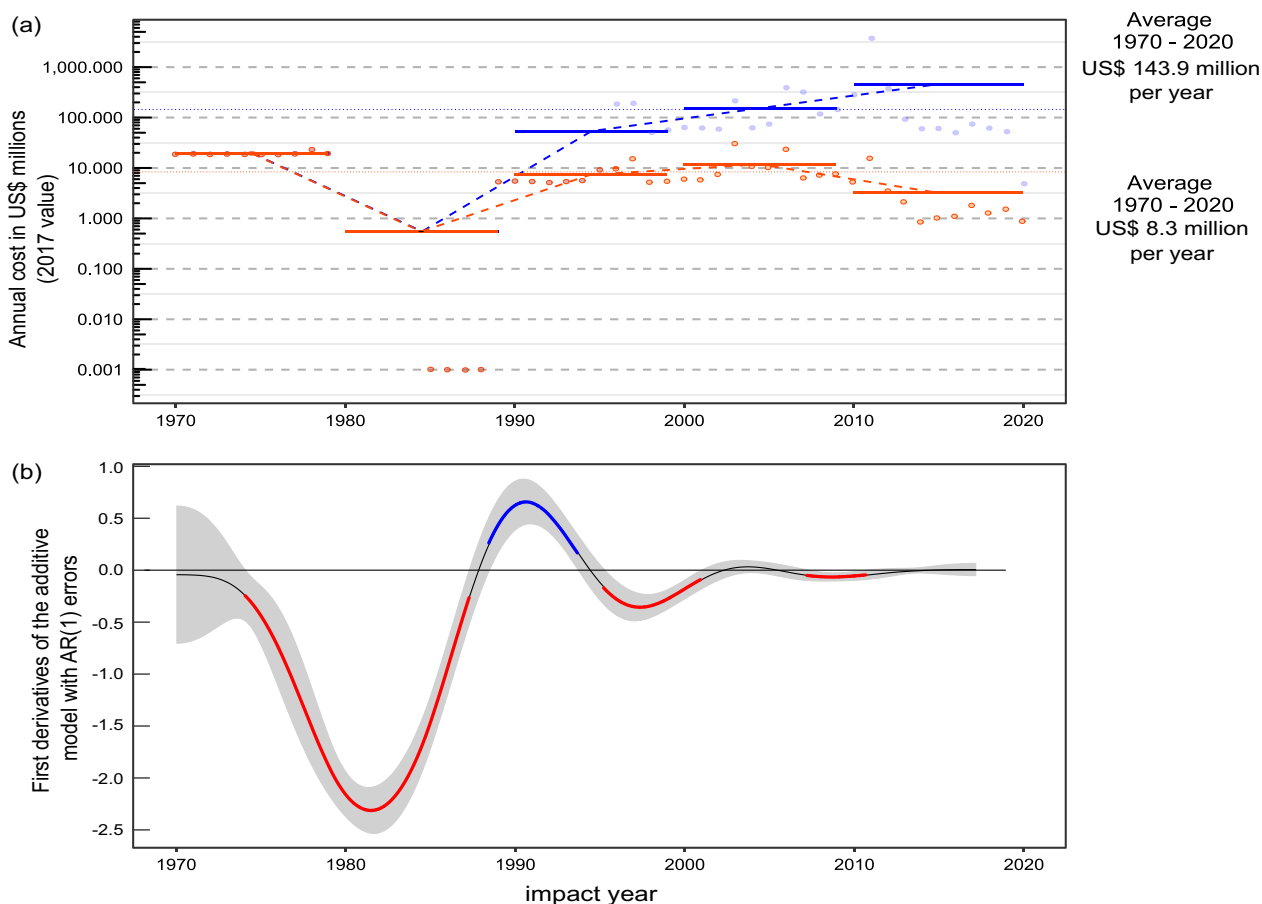


Fig. 2 **a** Temporal trend of highly reliable, observed average annual costs per decade (solid bars) with (blue) and without (orange) extreme values recorded in *InvaCost* for European Union member states (annual costs are presented on the log₁₀ scale). Points are annual totals, and the dash-dot line is the overall annual average. **b** Diagnostic plots showing changes in the total annual cost (highly reliable, observed subset) over time (considered against impact year that refers to the year that the cost incurred). Red: periods of non-random declines. Blue: periods of non-random increases

The extent of projected costs across European Union member states varied substantially. The final dataset revealed that the highest increase in projected costs occurred in Lithuania (US\$201.19 million; 1,093,525%), followed by Malta (US\$193.25 million; 129,923%), Czech Republic (US\$ 3,818.83 million; 60,093%), and Denmark (US\$524.67 million; 54,670%). In terms of raw costs, the highest increase was in Czech Republic (US\$3.8 billion; €4.0 billion), followed by Germany (US\$2.4 billion; €2.5 billion), France (US\$2.3 billion; €2.4 billion), the Netherlands and Poland both with US\$2.0 billion (€2.1 billion). Added costs for all other countries were < US\$2 billion (€2.1 billion). We provide a detailed breakdown of projected country-level costs in Fig. 4 and Additional file 1. On average, the projection of costs increased national invasion costs by 4750% ($\pm 15,216\%$).

We also found that the projection of costs in European Union member states increased by an average of 2340%

for impacted sectors (Additional file 1). Following the projection, costs to public and social welfare increased the most (US\$4,243.49 million; 7950%), followed by costs to forestry (US\$8,157.39; 4588%), authorities and stakeholders (US\$9,333,86 million; 2611%), fisheries (US\$88.28 million; 2112%), health (US\$273.23 million; 30%), and the agricultural sector (US\$549.73 million; 18%) (see Additional file 1 for a detailed description of the sectors considered in the *InvaCost* database). Finally, scenarios considering the missing costs of the two main types of costs resulted in projections increasing management costs by US\$8,786.76 million (3933%), and by US\$14,400.38 million (325%) for damage costs (Additional file 1).

Temporal projection

Using the 1980–2017 subset of highly reliable (observed costs), we identified substantial increases with both linear

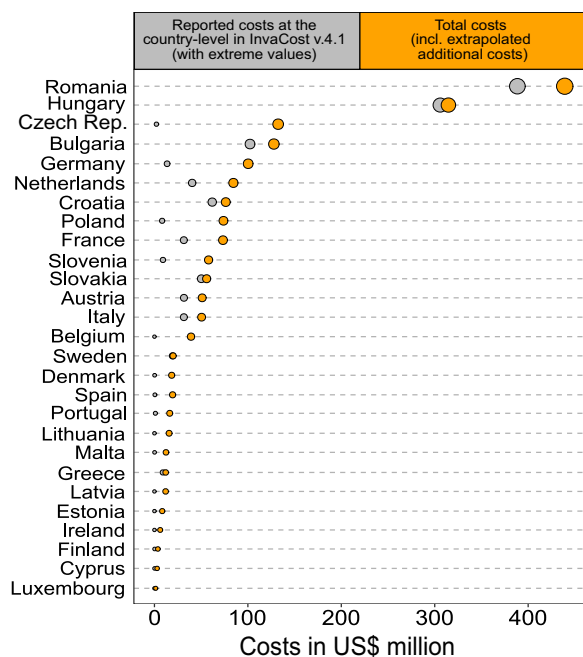


Fig. 4 Comparison of highly reliable observed country-level costs and projected costs at the level of individual European Union member states. The size of the circles is scaled to the respective cost

the magnitude of invasion costs was at least 1.5 to 6 times higher, even while using a conservative subset of the available data. We expect these costs to rise further towards 2040 in accordance with the temporal increase in numbers of alien species [48]. In addition, we found that the presence of high-impact (i.e., highly costly) species increased decadal averages by a factor of 10, underlining the importance of understanding the impacts and effective management actions for these particularly costly invasive alien species.

Spatial, sectoral, and taxonomic projections

Economic costs are available for just 1% of alien species recorded as established across the European Union. For the 88 invasive alien species currently identified as being of Union concern, a much higher proportion of 54.5% had reported costs (although many still lacked costs evaluated at the level of individual member states). This absence of data is concerning given that ~14,000 alien taxa are listed in the *European Alien Species Database* [65] for Europe, and that invader distribution assessments can be incomplete or decades out of date [66]. This disparity between species with cost records and invader richness assessments indicates that the full economic burden on the European Union will be much higher than currently reported. Indeed, it is precisely this low proportion of alien species with reported

costs at national levels that is one of the major limitations to estimating true total costs and raising awareness of the severe impacts of invasions across the European Union.

This underestimation is further underlined for multiple, highly impactful taxa that either have extensive potential distributions under current and future scenarios of environmental change [67, 68], or are present and have the potential to spread from the 32 neighbouring countries into the European Union. Further, we did not include in our analysis the European Union candidate countries (Albania, Moldova, Republic of North Macedonia, Montenegro, Serbia, Turkey, Ukraine), nor other European countries beyond the European Union, such as Switzerland, the United Kingdom, and Norway. When accounting for some of these gaps, cost projections revealed a potentially > six-fold (>500%) rise using only a conservative subset including only country-level entries from 49 species, thereby excluding 172 species ($n=3080$ entries) with costs inferred only at the level of specific sites. We found that growth due to the scaling of missing costs was particularly burdensome to the public and social-welfare sector, followed by sectors encompassed by the circular economy initiatives and the European Green Deal (i.e., fisheries and forestry, both economic sectors of high importance to the European Union; [69, 70]).

Although the projected cost increased by only 164% when we excluded two high-leverage points, we emphasise that this model was a poorer fit than the model including all values. While this does not necessarily mean that the higher projection is the most realistic, it does underline the importance of assessing the robustness of outliers. Here, these two high-leverage costs were both highly reliable and observed and thus not unreliable per se, as exemplified by massive costs arising from national eradication programmes of single species [71] or a few 'hyper-costly' invasive alien species [55].

Our cost modelling has taken into account all management costs, regardless of whether they were effective or inefficient, and hence we did not differentiate between, e.g., costly eradication measures and cheaper mitigation measures (i.e., based on efficacy or effectiveness). This information is often unavailable in underlying studies, but would constitute a major step forward in understanding management decisions for future work [72]. However, while costs from eradication efforts can be substantial, these can be short-term expenses for long-term savings [73, 74], and inevitably increase with per-capita effort over time as the populations or individuals most difficult to eradicate are targeted last, leading to an unequal temporal distribution of management costs. There is the possibility that such a substantial increase in projected management costs could undermine enthusiasm for future management efforts. However, it is also likely that when such control

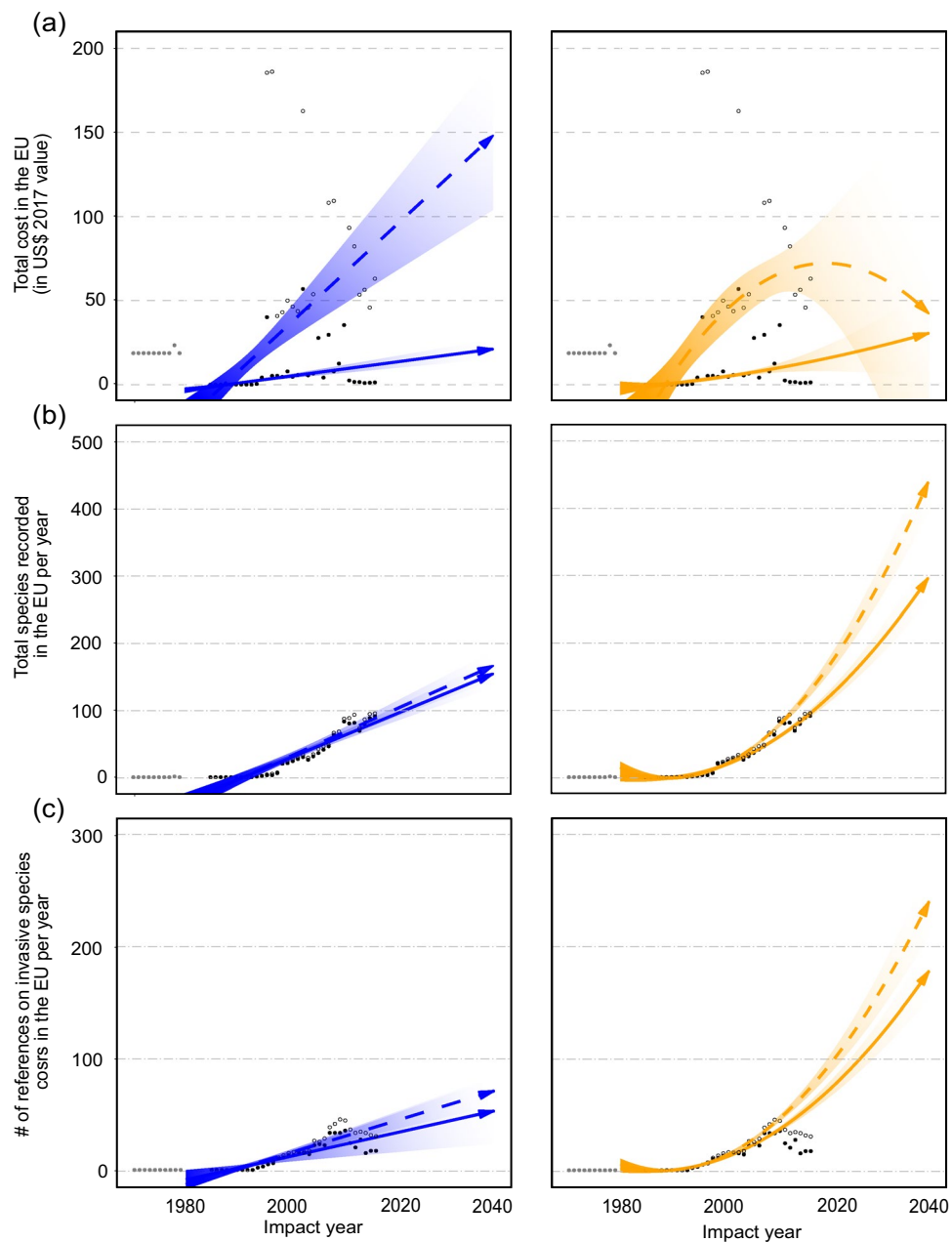


Fig. 5 Temporal projection of total annual invasion costs **(a)**, recorded species per year with costs **(b)**, and number of reported references in *InvaCost* v4.1 per year **(c)** using linear (left, in blue) and quadratic (right, in orange) robust regressions with their respective confidence intervals. Solid dots and lines represent trends without the inclusion of extreme values. Open circles and dashed lines represent the respective trends when including all data (i.e., including extreme values). Grey dots present data for 1960–1979 that we did not consider in the projection due to an absence of repeated measures of individual cost entries

and eradication (a component of management costs) are successful, future costs are reduced to routine monitoring while avoiding the costs of inaction [61, 75].

Although not all alien species do or will cause monetisable damage or elicit management expenditure (e.g., non-monetised negative environmental impacts or incompatibility with current economic sectors; [76, 77]), the identified disparity with invader richness indicates that the full economic burden on the European Union will be much higher than the estimate we provide here. This reinforces the need for a coordinated effort across the European Union to adopt better reporting of invasion costs, and for central and accessible collation, as found within *InvaCost*, a living database that has grown considerably in only a few years (v0, 2017: 260 entries; v1.0: 2,419; v3.0: 9,823; v4.1, 2022: 13,553). Simplifying the process of adding information to such databases would facilitate use and uptake. This should include increasing outreach to industry groups, health insurance sectors, and others where a lack of communication among those who are affected by damage or loss, those who assess costs, and those who synthesise costs results in currently recorded but inaccessible data. Such efforts would provide a more complete and accurate picture of the total costs and the full economic impacts of biological invasions. While technical updates to database accessibility are complex and ongoing, our scaled projection approach provides a more complete estimate of costs and a major advance in the interpolation of invasion costs from limited data. Unlike previous attempts to investigate the temporal dynamics of the damage costs of invasive alien species [61, 75], the scaling model we developed permits spatial cost projections from one region to another given regional-specific species records or context-specific differences in conditions (e.g., economic factors). Such extrapolated costs then provide a basis for modelling incurred costs across a range of spatio-temporal scales.

Temporal projection

Early cost reporting for invasive alien species was particularly sporadic (e.g., reporting high cost from the rabbit *Oryctolagus cuniculus* in Germany in the 1970s; [78]), with countries joining the European Economic Community in the 1980s and the first international research programmes (ERASMUS) being founded, before the foundation of the European Union in 1993. Hence, our models revealed further increases in costs, new costly invasive alien species, but also cost reports across the European Union over the next 20 years until 2040. Variation in decadal costs remained low, likely due to the conservative approach to include costs used to inform these temporal projections and exclude costs characterised as 'potential', meaning that they had not materialised at

the time of reporting. However, it is possible that some of those costs could have been incurred following the year of reporting and that they might not have yet been recorded in *InvaCost*. The projected increase in the number of invasive alien species reflects a robust predicted rise identified in invasion rates and an increasing awareness of their impacts [79].

The well-characterised effect of reporting lags in the costs of invasive alien species [22, 20, 80] most likely explains the declines in total annual costs reported in recent years. An increasing number of alien species [7, 79] enabled by changing global climates and transport networks [81] will demand increasing management interventions and thereby inflate their future monetary burden, even though not all of them will become invasive [82]. This is particularly emphasised by reported management costs contributing 90–93% of the available data, while contributing only ~11% of the highly reliable, observed costs and ~28% when we excluded extreme values, and showing the steepest projected incline (Additional file 1). This increasing burden could also be exacerbated by continued economic growth, given the positive relationship between costs and economic output [22], changes in the economic output of sectors over time where resident invasive alien species impose the most damage, as well as reduced efficacy of management (e.g., via chemical resistance).

Uncertainties

While projecting unknown costs of biological invasions in the European Union, our research has revealed large disparities between the total number of invasive alien species that have established themselves and those that have had their associated costs assessed (1%). Out of all the European Union member states, Romania and Hungary have notably higher reported costs resulting from invasive species at the country level. These high costs are driven by estimated damages caused by the common ragweed *Ambrosia artemisiifolia* and can potentially be explained by the large proportion of agricultural land invaded (36% and 61% in Romania and Hungary, respectively; [83]). This indicates how single species can have a disproportionate effect on cost reporting depending on the prevalence of impacted sectors within a location or region.

Because we estimated potential costs caused by invasive species for European Union member states where monetary impacts were unknown, it is important to emphasise the inherent uncertainty involved. Consequently, projected costs, whether over space or time, should be considered with caution. Nevertheless, we also note that these estimations were based only on a conservative subset of the richest country-level data, while

considering that for invasive alien species from which models were fit, the costs were still fragmented and underestimated, thereby highlighting the importance of the costs involved. An additional source of uncertainty lies in lack of knowledge about the efficiency of management costs. Information on whether management expenditure has turned out to be efficient and to what degree, in relation to their objectives, can be essential for building a more comprehensive understanding of the true costs associated with biological invasions.

Our analyses do not consider any positive effects of invasive alien species, which are generally not captured in the *InvaCost* database. Examples could include, for instance, an increase in biological production in semi-arid ecosystems, the restoration of industrial or contaminated areas where native species cannot survive, with benefits from the acceleration of succession and soil formation or reductions in nutrient leakage and water basin eutrophication [84, 85], or fast-growing plants that increase carbon sequestration while outcompeting native species [86]. Alternatively, some invasive alien species such as *Buddleja davidii* might provide certain local economic benefits [87, 88], although these cannot discount or undermine the presence of costs because they often affect different actors and ecosystem functions [89], and nor have they been documented or known to be anywhere close to the magnitude of costs.

Outlook

The number of invasive alien species and resultant economic costs will likely increase in the future, even if projections remain uncertain [7, 90], due to reporting lags [20], the emergence of new invasive alien species as costly [9], the high variance in reported costs between 1980 and 2020, and the low sample of cost estimates before 1980 [91]. Even so, the reported total of US\$7.3 billion in highly reliable observed costs is a substantial sum, especially considering that fewer than 1% of alien species within the European Union have documented costs. Our results strongly indicate the need for collective mitigation and prevention actions within the European Union to prevent biological invasions and their associated costs. These include, in addition to existing legislation: (i) improved biosecurity protocols; (ii) well-coordinated, large-scale management; (iii) targeted research to improve pre- and post-invasion mitigation; and (iv) horizon scanning to include a consideration of costs [92]. Well-communicated, timely actions and coordination substantially mitigate the negative impacts of invasive alien species [75, 76, 93, 94]. Based on the projected increases in future costs, we urge the European Union member states to expand their efforts in Union-wide coordination to combat the threat posed by invasive alien species. We further urge

investments in a more granular understanding of management-expenditure efficiency that will allow for a more comprehensive idea of the costs of management actions that contribute to lowering invasion costs. This can in turn be expected to help capture with more accuracy the true toll of biological invasions, and identify and fill any knowledge gaps.

Cost assessments play an important role in improving management efficiency and effectiveness by promoting early, focused, and evidence-based management interventions towards alien species, ultimately leading to long-term cost savings, relieving the burden on taxpayers, and enabling governments to allocate their financial resources more effectively. This is further substantiated because prior work has emphasised the ability of early investments (e.g., in biosecurity) to lower impacts of biological invasions in the long term [75, 93]. However, economic downplaying of risks or impacts, often triggered by insufficient budget allocation for conservation, further reduces the likelihood of cooperation among and within countries to implement risk assessments and management planning. Additional economic burdens imposed by invasive alien species could exacerbate an already difficult period of geopolitical tension [95] regarding the cost of natural resources and biodiversity. The identified growing socio-economic impacts of invasive alien species [9, 80] and the current persistent misallocation of strategic investment [96], mean that invasive alien species are degrading many aspects of human life, and placing an additional burden on human and social well-being [2]. Rising resource losses and damages due to invasive alien species could fuel further political debates among European Union member states and affect conservation and research budgets. This relative instability could weaken the European Union in the future following recent crises (e.g., COVID-19 pandemic, Brexit, Russian invasion of Ukraine) [97–99] and could drive policy changes that have negative consequences for trade and biosecurity. For example, the case of changing food-export pathways and cost increases in selling prices received by domestic producers for their output occurred after the invasion of Ukraine [100], a phenomenon that could in turn reduce attention on environmental challenges such as biological invasions. Considering these ongoing economic stressors within the European Union [98], our results warn against the pattern of reduced investment in environmental management [101, 102]. Without improved actions, the rates of biological invasions and associated costs will continue to rise, degrading natural capital, and hampering industries, sustainable development, and sustainability targets [103].

Supplementary Information

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Additional file 1: Description of the sectors considered in the InvaCost database. **Additional file 2:** Site-level costs recorded in InvaCost for the European Union member states (a), by type of cost (b) and impacted sector (c) with and without high-leverage points and the respective projected additional cost and percent increase. **Additional file 3:** Detailed breakdown of the considered models included in the model averaging. **Table 1:** Country-level analyses. Values of the fitted parameters for each predictor combination and model performance estimated using Akaike's information criterion (AIC). We calculated Akaike's weights for model averaging following Burnham & Anderson [57]. **Table 2:** Country-level analyses without high-leverage points. Values of the fitted parameters for each predictor combination and model performance estimated using Akaike's information criterion (AIC). Model outputs for each combination of predictors. We calculated Akaike's weights for model averaging following Burnham & Anderson [57]. **Table 3:** Site-level analyses (A) with all data and (B) without high-leverage points. Values of the fitted parameters for each predictor combination and model performance estimated using Akaike's information criterion (AIC). Model outputs for each combination of predictors. We calculated Akaike's weights for model averaging following Burnham & Anderson [57]. **Additional file 4:** Invasion costs (total costs and highly reliable, observed costs) for European Union member states recorded in *InvaCost* v4.1, in 2017 US\$ billion. **Additional file 5:** Comparison of the number of established alien species in the European Union and alien species in *InvaCost* v4.1. **Additional file 6:** Established alien species of Union concern recorded among the established alien species in the European Union with the first European Union member state they were recorded in and the year of first record [48]. **Additional file 7:** Breakdown of interpolated costs excluding high-leverage points. **Additional file 8:** Country-level costs recorded in *InvaCost* with and without high-leverage points and the respective projected additional cost and percent increase. **Additional file 9:** Country-level costs recorded in *InvaCost* by type of cost (a) and impacted sector (b) with and without extreme values and the respective projected additional cost and percent increase. **Additional file 10:** Temporal projection of total annual management costs (a), reported species per year with management costs (b), and number of reported references on management expenditure in *InvaCost* v4.1 per year (c) using linear (left, in blue) and quadratic (right, in orange) robust regressions and their respective confidence intervals. Solid dots and lines represent trends without the inclusion of extreme values. Open circles and dashed lines represent the respective trends when including all data (i.e., including extreme values). Grey dots present data for the period 1960–1979 that are not considered in the projection due to an absence of repeated measures of individual cost entries.

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Author contributions

MH, PJH, and BL analyzed the data. PJH, CJAB, RNC, MH, BL, and TWB contributed to the writing of the first version of the manuscript. EB, and FC contributed valuable data. DAA, EA, PB, EB, FC, PEH, AK, MK, CL, RLM, FJC, DR, IS, AST, and AJT contributed to the writing of the manuscript and the interpretation of results. All authors read and approved the final manuscript.

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Availability of data and materials

All data generated or analysed during this study are openly available: github.com/LeungEcoLab/EU_costs.

Declarations

Ethics approval and consent to participate

Not applicable.

Consent for publication

Not applicable.

Competing interests

The authors declare that they have no competing interests.

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References

- Bellard C, Cassey P, Blackburn TM (2016) Alien species as a driver of recent extinctions. *Biol Lett* 12:20150623
- Pyšek P, Hulme PE, Simberloff D, Bacher S, Blackburn TM, Carlton JT, Dawson W, Essl F, Foxcroft LC, Genovesi P, Jeschke JM, Kühn I, Liebhold AM, Mandrak NE, Meyerson LA, Pauchard A, Pergl J, Roy HE, Seebens H, van Kleunen M, Vilà M, Wingfield MJ, Richardson DM (2020) Scientists' warning on invasive alien species. *Biol Rev* 95:1511–1534
- Shackleton RT, Shackleton CM, Kull CA (2019) The role of invasive alien species in shaping local livelihoods and human well-being: a review. *J Environ Manage* 229:145–157
- Shackleton RT, Richardson DM, Shackleton CM, Bennett B, Crowley SL, Dethnen-Schmutz K, Estévez RA, Fischer A, Kueffer C, Kull CA, Marchante

- E, Novoa A, Potgieter LJ, Vaas J, Vaz AS, Larson BM (2019) Explaining people's perceptions of invasive alien species: a conceptual framework. *J Environ Manage* 229:10–26
5. Sinclair JS, Brown JA, Lockwood JL (2020) Reciprocal human-natural system feedback loops within the invasion process. *NeoBiota* 62:489–508
 6. Vilà M, Hulme PE (2017) Impact of biological invasions on ecosystem services. Springer, Cham
 7. Seebens H, Bacher S, Blackburn TM, Capinha C, Dawson W, Dullinger S, Genovesi P, Hulme PE, van Kleunen M, Kühn I, Jeschke JM, Lenzner B, Liebhold AM, Pattison Z, Pergl J, Pyšek P, Winter M, Essl F (2021a) Projecting the continental accumulation of alien species through to 2050. *Glob Change Biol* 27:970–982
 8. Ribeiro J, Bingre P, Strubbe D, Santana J, Capinha C, Araújo MB, Reino L (2022) Exploring the effects of geopolitical shifts on global wildlife trade. *Bioscience* 72:560–572
 9. Essl F, Lenzner B, Bacher S, Bailey S, Capinha C, Daehler C, Dullinger S, Genovesi P, Hui C, Hulme PE, Jeschke JM (2020) Drivers of future alien species impacts: an expert-based assessment. *Glob Change Biol* 26:4880–4893
 10. Panetta FD (1993) A system of assessing proposed plant introductions for weed potential. *Plant Prot Q* 8:10–14
 11. Kumschick S, Vimercati G, Villiers FA, Mokhatla MM, Davies SJ, Thorp CJ, Rebelo AD, Measey GJ (2017) Impact assessment with different scoring tools: how well do alien amphibian assessments match? *NeoBiota* 33:53–66
 12. Vilà M, Gallardo B, Preda C, García-Berthou E, Essl F, Kenis M, Roy HE, González-Moreno P (2019) A review of impact assessment protocols of non-native plants. *Biol Invasions* 21:709–723
 13. González-Moreno L, Lazzaro Vilà Preda Adriaens Bacher Brundu Coppessl García-Berthou Katsanevakis Moen Lucy Nentwig Roy Srébalie nêTalgø Vanderhoeven Andjelković Arbačauskas Auger-Rozenberg Bariche Boets Boeiro Borges Canning-Clode Cardigos Chartosia Cottier-Cook Crocetta D'hondt Foggia Follak Gallardo Gammelman Giakoumi Giuiliani Fried Jelaska Jeschke Jover Juárez-Escario Kalogirou Kočić Kytinou, Laverty Lozano Maceda-Veiga Marchante Marchante Martinou Meyer Michin-Montero-Castaño Morales, Morales-Rodriguez Muhthassim Nagy Ogris Onen, Pergl Puntilla Rabitsch Ramburn Rego Reichenbach Romeralo Saul Schrader Sheehan Simonović Skolka Soares Sundheim Tarkan Tomov Tricarico Tsiamis, Uludağ van Valkenburg Verreycken Vetraino Vilar Wiig Wittzell Zanetta Kenis PLMCTSGGHFFESTLFEWHEGVSAKM-AM-JMPMPAJFNEJFBB-SBØSCGLSJMMASAECAEVAEHAFSDAMCCNZANHJRWTTCFCW-CGRP-MAOLASREKAJHAMLØJAM (2019) Consistency of impact assessment protocols for non-native species. *NeoBiota* 44:1–25
 14. Pimentel D, Zuniga R, Morrison D (2005) Update on the environmental and economic costs associated with alien-invasive species in the United States. *Ecol Econ* 52:273–288
 15. Hensley MP (2012) A critique on the current standards for evaluating costs for invasive species in economic literature. Dissertation, Southern Illinois University Carbondale
 16. Lovell SJ, Stone SF, Fernandez L (2006) The economic impacts of aquatic invasive species: a review of the literature. *Agric Resour Econ Rev* 35:195–208
 17. Paine DR, Sheppard AW, Cook DC, De Barro PJ, Worner SP, Thomas MB (2016) Global threat to agriculture from invasive species. *Proc Natl Acad Sci USA* 113:7575–7579
 18. Ahmed DA, Haubrock PJ, Cuthbert RN et al (2023) Recent advances in availability and synthesis of the economic costs of biological invasions. *Authorea*. <https://doi.org/10.22541/au.167907497.77590504/v1>
 19. Diagne C, Leroy B, Gozlan RE, Vaissière AC, Assailly C, Nuninger L, Roiz D, Jourdain F, Jarić I, Courchamp F (2020) InvaCost, a public database of the economic costs of biological invasions worldwide. *Sci Data* 7:277
 20. Leroy B, Kramer AM, Vaissière AC, Kourantidou M, Courchamp F, Diagne C (2022) Analysing economic costs of invasive alien species with the invacost R package. *Methods Ecol Evol* 13:1930–1937
 21. Cuthbert RN, Pattison Z, Taylor NG, Verbrugge L, Diagne C, Ahmed DA, Leroy B, Angulo E, Briski E, Capinha C, Catford JA, Dalu T, Essl F, Gozlan RE, Haubrock PJ, Kourantidou M, Kramer AM, Renault D, Wasserman RJ, Courchamp F (2021) Global economic costs of aquatic invasive alien species. *Sci Total Environ* 775:145238
 22. Haubrock PJ, Turbelin AJ, Cuthbert RN, Novoa A, Taylor NG, Angulo E, Ballesteros-Mejia L, Bodey TW, Capinha C, Diagne C, Essl F, Golivets M, Kirichenko N, Kourantidou M, Leroy B, Renault D, Verbrugge L, Courchamp F (2021) Economic costs of invasive alien species across Europe. *NeoBiota* 67:153–190
 23. Renault D, Manfrini E, Leroy B, Diagne C, Ballesteros-Mejia L, Angulo E, Courchamp F (2021) Biological invasions in France: alarming costs and even more alarming knowledge gaps. *NeoBiota* 67:191–224
 24. Kourantidou M, Cuthbert R, Haubrock PJ, Novoa A, Taylor N, Leroy B, Capinha C, Renault D, Angulo E, Diagne C, Courchamp F (2021) Economic costs of invasive alien species in the Mediterranean basin. *NeoBiota* 67:427–458
 25. Crystal-Ornelas R, Hudgins EJ, Cuthbert RN, Haubrock PJ, Fantle-Lepczyk J, Angulo E, Kramer AM, Ballesteros-Mejia L, Leroy B, Leung B, López-López E, Diagne C, Courchamp F (2021) Economic costs of biological invasions within north America. *NeoBiota* 67:485–510
 26. Kouba A, Oficialdegui FJ, Cuthbert RN, Kourantidou M, South J, Tricarico E, Gozlan RE, Courchamp F, Haubrock PJ (2022) Identifying economic costs and knowledge gaps of invasive aquatic crustaceans. *Sci Total Environ* 813:152325
 27. Balzani P, Cuthbert RN, Briski E, Galil B, Castellanos-Galindo GA, Kouba A, Kourantidou M, Leung B, Soto I, Haubrock PJ (2022) Knowledge needs in economic costs of invasive species facilitated by canalisation. *NeoBiota* 78:207–223
 28. Turbelin AJ, Diagne C, Hudgins EJ, Moodley D, Kourantidou M, Novoa A, Haubrock PJ, Bernery C, Gozlan RE, Francis RA, Courchamp F (2022) Introduction pathways of economically costly invasive alien species. *Biol Invasions* 24:2061–2079
 29. Zenni RD, Essl F, García-Berthou E, McDermott SM (2021) The economic costs of biological invasions around the world. *NeoBiota* 67:1–9
 30. Capinha C, Essl F, Porto M, Seebens H (2023) The worldwide networks of spread of recorded alien species. *Proc Natl Acad Sci USA* 120:e2201911120
 31. Carboneras C, Genovesi P, Vilà M, Blackburn TM, Carrete M, Clavero M, D'hondt B, Orueta JF, Gallardo B, Gerales P, González-Moreno P, Gregory RD, Nentwig W, Paquet J-Y, Pyšek P, Rabitsch W, Ramirez I, Scalera R, Tella JL, Walton P, Wynne R, (2018) A prioritised list of invasive alien species to assist the effective implementation of European union legislation. *J Appl Ecol* 55:539–547
 32. Genovesi P, Carboneras C, Vilà M, Walton P (2015) EU adopts innovative legislation on invasive species: a step towards a global response to biological invasions? *Biol Invasions* 17:1307–1311
 33. Turbelin AJ, Malamud BD, Francis RA (2017) Mapping the global state of invasive alien species: patterns of invasion and policy responses. *Glob Ecol Biogeogr* 26:78–92
 34. EU, (2014) Regulation No 1143/2014 of the European parliament and of the council on the prevention and management of the introduction and spread of invasive alien species. *Off J Eur Union* L315:35–55
 35. Dawson W, Moser D, Van Kleunen M, Kreft H, Pergl J, Pyšek P, Weigelt P, Winter M, Lenzner B, Blackburn TM, Dyer EE, Cassey P, Scrivens SL, Economo EP, Guénard B, Capinha C, Seebens H, García-Díaz P, Nentwig W, García-Berthou E, Casal C, Mandrak NE, Fuller P, Meyer C, Essl F (2017) Global hotspots and correlates of alien species richness across taxonomic groups. *Nat Ecol Evol* 1:0186
 36. Caffrey JM, Baars JR, Barbour JH, Boets P, Boon P, Davenport K, Dick JTA, Early J, Edsman L, Gallagher C, Gross J, Heinimaa P, Horrill C, Hudin S, Hulme PE, Hynes S, MacIsaac HJ, McLoone P, Millane M, Moen TL, Moore N, Newman J, O'Conchuir R, O'Farrell M, O'Flynn C, Oidtmann B, Renals T, Ricciardi A, Roy H, Shaw R, van Valkenburg JLCH, Weyl O, Williams F, Lucy FE (2014) Tackling invasive alien species in Europe: the top 20 issues. *Manag Biol Invasions* 5:1–20
 37. Keller RP, Geist J, Jeschke JM, Kühn I (2011) Invasive species in Europe: ecology, status, and policy. *Environ Sci Eur* 23:1–17
 38. Sharp RL, Larson LR, Green GT (2011) Factors influencing public preferences for invasive alien species management. *Biol Conserv* 144:2097–2104
 39. Essl F, Latombe G, Lenzner B, Pagad S, Seebens H, Smith K, Wilson JRJ, Genovesi P (2020) The convention on biological diversity (CBD)'s post-2020 target on invasive alien species—what should it include and how should it be monitored? *NeoBiota* 62:99–121
 40. Pergl R, Hooft R, Suchánek M, Knaisl V, Slifka J (2019) "Data Stewardship Wizard": a tool bringing together researchers, data stewards, and data experts around data management planning. *Data Sci J* 18:59

41. Sutcliffe C, Quinn CH, Shannon C, Glover A, Dunn AM (2018) Exploring the attitudes to and uptake of biosecurity practices for invasive non-native species: views amongst stakeholder organisations working in UK natural environments. *Biol Invasions* 20:399–411
42. Heink U, Van Herzele A, Bela G, Kalóczkai Á, Jax K (2018) Different arguments, same conclusions: how is action against invasive alien species justified in the context of European policy? *Biodivers Conserv* 27:1659–1677
43. Kulhanek SA, Ricciardi A, Leung B (2011) Is invasion history a useful tool for predicting the impacts of the world's worst aquatic invasive species? *Ecol Appl* 21:189–202
44. Early R, Bradley BA, Dukes JS, Lawler JJ, Olden JD, Blumenthal DM, Gonzalez P, Grosholz ED, Ibañez I, Miller LP, Sorte CJB, Tatem AJ (2016) Global threats from invasive alien species in the twenty-first century and national response capacities. *Nat Commun* 7:12485
45. Angulo E, Diagne C, Ballesteros-Mejia L, Adamjy T, Ahmed DA, Akulov E, Banerjee AK, Capinha C, Dia CAKM, Dobigny G, Duboscq-Carra VG, Golivets M, Haubrock PJ, Heringer G, Kirichenko N, Kourantidou M, Liu C, Nuñez MA, Renault D, Roiz D, Taheri A, Verbrugge LNH, Watari Y, Xiong W, Courchamp F (2021) Non-english languages enrich scientific knowledge: the example of economic costs of biological invasions. *Sci Total Environ* 775:144441
46. Angulo E, Ballesteros-Mejia L, Novoa A, Duboscq-Carra V, Diagne C, Courchamp F (2021) Economic costs of invasive alien species in Spain. *NeoBiota* 67:267–297
47. Cuthbert RN, Bartlett AC, Turbelin AJ, Haubrock PJ, Diagne C, Pattison Z, Courchamp F, Catford JA (2021) Economic costs of biological invasions in the United Kingdom. *NeoBiota* 67:299–328
48. Seebens H, Blackburn TM, Dyer EE, Genovesi P, Hulme PE, Jeschke JM, Pagad S, Pyšek P, Winter M, Arianoutsou M, Bacher S, Blasius B, Brundu G, Capinha C, Celesti-Grapow L, Dawson W, Dullinger S, Fuentes N, Jäger H, Kartesz J, Kenis M, Kreft H, Kühn I, Lenzner B, Liebhold A, Mosena A, Moser D, Nishino M, Pearman D, Pergl J, Rabitsch W, Rojas-Sandoval J, Roques A, Rorke S, Rossinelli S, Roy HE, Scalera R, Schindler S, Štajerová K, Tokarska-Guzik B, van Kleunen M, Walker K, Weigelt P, Yamanaka T, Essl F (2017) No saturation in the accumulation of alien species worldwide. *Nat Commun* 8:14435
49. Seebens H, Clarke DA, Groom Q, Wilson JR, García-Berthou E, Kühn I, McGeoch M (2020) A workflow for standardising and integrating alien species distribution data. *SUNScholar Research Repository*. <http://hdl.handle.net/10019.1/112353>
50. R Core Team (2022) R: A Language And Environment for Statistical Computing. R Foundation for Statistical Computing, Vienna <http://www.r-project.org/index.html>.
51. Seebens H (2021b) SInAS database of alien species occurrences (2.4.1), Zenodo. <https://doi.org/10.5281/zenodo.5562892>; Accessed 16 May 2022
52. Casties I, Seebens H, Briski E (2016) Importance of geographic origin for invasion success: a case study of the north and Baltic seas versus the great lakes-St. Lawrence River region *Ecol Evol* 6:8318–8329
53. Groom Q, Desmet P, Reyserhove L, Adriaens T, Oldoni D, Vanderhoeven S, Baskauf SJ, Chapman A, McGeoch M, Walls R, Wiczorek J, Wilson JR, Zermoglio PFF (2019) Improving Darwin core for research and management of alien species. *Biodivers Inf Sci Stand* 3:e38084
54. GBIF (2022) Global Biodiversity Information Facility. <https://www.gbif.org>. Accessed 1 Jul 2022
55. Heringer G, Angulo E, Ballesteros-Mejia L, Capinha C, Courchamp F, Diagne C, Duboscq-Carra VG, Nuñez MA, Zenni RD (2021) The economic costs of biological invasions in central and south America: a first regional assessment. *NeoBiota* 67:401–426
56. Turbelin AJ, Hudgins EJ, Catford JA, Cuthbert RN, Diagne C, Kourantidou M, Roiz D, Courchamp F (2023) Biological invasions as burdens to primary economic sectors. <https://doi.org/10.21203/rs.3.rs-2444595/v1>
57. Burnham KP, Anderson DR (2002) Model selection and multimodel inference: a practical information-theoretic approach. Springer-Verlag, Berlin
58. Nathan et al. 1999. On the application of generalised additive models to the detection of trends in hydrologic time series data. In: *Handbook and Proceedings of the Water 99 Joint Congress, 25th Hydrology & Water Resources Symposium, 2nd International Conference on Water Resources & Environment Research, Institution of Engineers, Barton, Australia*. Available at search.informit.org/doi/https://doi.org/10.3316/INFO
59. Le Hen G, Balzani P, Haase P, Kouba A, Liu C, Nagelkerke LA, Theissen N, Renault D, Soto I, Haubrock PJ (2023) Alien species and climate change drive shifts in a riverine fish community and trait compositions over 35 years. *Sci Total Environ* 867:161486
60. Simpson GL (2018) Modelling palaeoecological time series using generalised additive models. *Front Ecol Evol* 6:1–21
61. Ahmed DA, Hudgins EJ, Cuthbert RN, Haubrock PJ, Renault D, Bonnaud E, Diagne C, Courchamp F (2022) Modelling the damage costs of invasive alien species. *Biol Invasions* 24:1949–1972
62. Haubrock PJ, Cuthbert RN, Ricciardi A, Diagne C, Courchamp F (2022) Economic costs of invasive bivalves in freshwater ecosystems. *Divers Distrib* 28:1010–1021
63. Maechler M, Rousseeuw P, Croux C, Todorov V, Ruckstuhl A, Salibian-Barrera M, Verbeke T, Koller M, Conceicao EL, Anna di Palma M (2022) robbustbase: Basic Robust Statistics. R package version 0.95–0
64. Hyndman R, Athanasopoulos G, Bergmeir C, Caceres G, Chhay L, O'Hara-Wild M, Petropoulos F, Razbash S, Wang E, Yasmeeen F (2023) forecast: forecasting functions for time series and linear models. R package version 8:20
65. EASIN (2022) European Alien Species Information Network. [EASIN \(2022\) European Alien Species Information Network. https://easin.jrc.ec.europa.eu/easin](https://easin.jrc.ec.europa.eu/easin). Accessed 01 Dec 2022
66. Coughlan NE, Lyne L, Cuthbert RN, Cunningham EM, Lucy FE, Davis E, Caffrey JM, Dick JT (2020) In the black: information harmonisation and educational potential amongst international databases for invasive alien species designated as of union concern. *Glob Ecol Conserv* 24:e01332
67. Gillard M, Thiébaud G, Deleu C, Leroy B (2017) Present and future distribution of three aquatic plants taxa across the world: decrease in native and increase in invasive ranges. *Biol Invasions* 19:2159–2170
68. Macêdo RL, Franco ACS, Kozłowski-Suzuki B, Mammola S, Dalu T, Rocha O (2022) The global social-economic dimension of biological invasions by plankton: grossly underestimated costs but a rising concern for water quality benefits? *Water Res* 222:118918
69. Carus M, Dammer L (2018) The circular bioeconomy—concepts, opportunities, and limitations. *Ind Biotechnol* 14:83–91
70. Siddi M (2020) The European Green Deal: Assessing its current state and future implementation. The Finnish Institute of International Affairs. FIIA Working Paper 114. <https://www.fiaa.fi/en/publication/the-european-green-deal>.
71. Haubrock PJ, Bernery C, Cuthbert RN, Liu C, Kourantidou M, Leroy B, Turbelin AJ, Kramer AM, Verbrugge LNH, Diagne C, Courchamp F, Gozlan RE (2022) Knowledge gaps in economic costs of invasive alien fish worldwide. *Sci Total Environ* 803:149875
72. Cuthbert RN, Diagne C, Hudgins EJ, Turbelin A, Ahmed DA, Albert C, Bodey TW, Briski E, Essl F, Haubrock PJ, Gozlan RE, Kirichenko N, Kourantidou M, Kramer AM, Courchamp F (2022) Biological invasion costs reveal insufficient proactive management worldwide. *Sci Total Environ* 819:153404
73. Carter ZT, Lumley T, Bodey TW, Russell JC (2021) The clock is ticking: temporally prioritizing eradications on islands. *Glob Chang Biol* 27:1443–1456
74. Russell JC, Innes JG, Brown PH, Byrom AE (2015) Predator-free New Zealand: conservation country. *Bioscience* 65:520–525
75. Ahmed DA, Hudgins EJ, Cuthbert RN, Kourantidou M, Diagne C, Haubrock PJ, Leung B, Liu C, Leroy B, Petrovskii S, Beidas A, Courchamp F (2022) Managing biological invasions: the cost of inaction. *Biol Invasions* 24:1927–1946
76. Cuthbert RN, Diagne C, Haubrock PJ, Turbelin AJ, Courchamp F (2022) Are the “100 of the world's worst” invasive species also the costliest? *Biol Invasions* 24:1895–1904
77. Vaissière AC, Courtois P, Courchamp F, Kourantidou M, Diagne C, Essl F, Kirichenko N, Welsh M, Salles JM (2022) The nature of economic costs of biological invasions. *Biol Invasions* 24:2081–2101
78. Heikkilä, J. (2006) Economics of invasive alien species: pre-emptive versus reactive control. Licentiate Thesis, MTT Discussion Papers
79. Seebens H, Blackburn TM, Dyer EE, Genovesi P, Hulme PE, Jeschke JM, Pagad S, Pyšek P, van Kleunen M, Winter M, Ansong M, Arianoutsou M, Bacher S, Blasius B, Brockhoff EG, Brundu G, Capinha C, Causton

- CE, Celesti-Grapow L, Dawson W, Dullinger S, Economo EP, Fuentes N, Guénard B, Jäger H, Kartesz J, Kenis M, Kühn I, Lenzner B, Liebhold AM, Mosen A, Moser D, Nentwig W, Nishino M, Pearnan D, Pergi J, Rabitsch W, Rojas-Sandoval J, Roques A, Rorke S, Rossinelli S, Roy HE, Scalera R, Schindler S, Štajerová K, Tokarska-Guzik B, Walker K, Ward DF, Yamanaka T, Essl F (2018) Global rise in emerging alien species results from increased accessibility of new source pools. *Proc Natl Acad Sci USA* 115:E2264–E2273
80. Diagne C, Leroy B, Vaissière AC, Gozlan RE, Roiz D, Jarić I, Salles J-M, Bradshaw CJA, Courchamp F (2021) High and rising economic costs of biological invasions worldwide. *Nature* 592:571–576
 81. Tobin I, Vautard R, Balog I, Bréon FM, Jerez S, Ruti PM, Thais F, Vrac M, Yiou P (2015) Assessing climate change impacts on European wind energy from ENSEMBLES high-resolution climate projections. *Clim Change* 128:99–112
 82. Russell JC, Blackburn TM (2017) Invasive alien species: denialism, disagreement, definitions, and dialogue. *Trends Ecol Evol* 32:312–314
 83. Bullock JM, Chapman D, Schafer S, Roy D, Girardello M, Haynes T, Beal S, Wheeler B, Dickie I, Phang Z, Tinch R (2012) Assessing and controlling the spread and the effects of common ragweed in Europe. Final Report to the European Commission, DG Environment. Nat Environ Res Council, UK
 84. Huxman TE, Cable JM, Ignace DD, Eilts JA, English NB, Weltzin J, Williams DG (2004) Response of net ecosystem gas exchange to a simulated precipitation pulse in a semi-arid grassland: the role of native versus non-native grasses and soil texture. *Oecologia* 141:295–305
 85. Mingorance MD, Franco I, Rossini-Oliva S (2017) Application of different soil conditioners to restorate mine tailings with native (*Cistus ladanifer* L.) and non-native species (*Medicago sativa* L.). *J Geochem Explor* 174:35–45
 86. Zhang Z, Wang Y, Zhu Y, He K, Li T, Mishra U, Peng Y, Wang F, Yu L, Zhao X, Zhu L (2022) Carbon sequestration in soil and biomass under native and non-native mangrove ecosystems. *Plant Soil* 16:1–6
 87. Britton JR, Orsi ML (2012) Non-native fish in aquaculture and sport fishing in Brazil: economic benefits versus risks to fish diversity in the upper river Paraná basin. *Rev Fish Biol Fisheries* 22:555–565
 88. Sheppard AW, Shaw RH, Sforza R (2006) Top 20 environmental weeds for classical biological control in Europe: a review of opportunities, regulations and other barriers to adoption. *Weed Res* 46(2):93–117
 89. Kourantidou M, Haubrock PJ, Cuthbert RN, Bodey TW, Lenzner B, Gozlan RE, Courchamp F (2022) Invasive alien species as simultaneous benefits and burdens: trends, stakeholder perceptions and management. *Biol Invasions* 24:1905–1926
 90. Armstrong JS (1984) Forecasting by extrapolation: conclusions from 25 years of research. *Interfaces* 14:52–66
 91. Cassey P, García-Díaz P, Lockwood JL, Blackburn TM, Jeschke J, Heger T (2018) Invasion biology: searching for predictions and prevention, and avoiding lost causes. In: Jeschke JM, Heger T (eds) *Invasion biology: hypotheses and evidence*. CAB International, Wallingford
 92. Roy HE, Bacher S, Essl F, Adriaens T, Aldridge DC, Bishop JDD, Blackburn TM, Branquart E, Brodie J, Carboneras C, Cook EJ, Copp GH, Dean HJ, Eilenberg J, Gallardo B, García M, García-Berthou E, Genovesi P, Hulme PE et al (2019) Developing a list of invasive alien species likely to threaten biodiversity and ecosystems in the European union. *Glob Change Biol* 25:1032–1048
 93. Leung B, Lodge DM, Finnoff D, Shogren JF, Lewis MA, Lamberti G (2002) An ounce of prevention or a pound of cure: bioeconomic risk analysis of invasive species. *Proc R Soc B: Biol Sci* 269:2407–2413
 94. Patoka J, Magalhães ALB, Kouba A, Faulkes Z, Jerikho R, Vitule JRS (2018) Invasive aquatic pets: failed policies increase risks of harmful invasions. *Biodivers Conserv* 27:3037–3046
 95. Analytica O (2022) Higher inflation and lower growth test corporate bonds. *Emerald Expert Briefings*. <https://doi.org/10.1108/OXAN-DB271600>
 96. Hermoso V, Clavero M, Villero D, Brotons L (2017) European union's conservation efforts need more strategic investment to meet continental commitments. *Conserv Lett* 10:231–237
 97. Preston C (1995) Obstacles to EU enlargement: the classical Community method and the prospects for a wider Europe. *J Common Mark Stud* 33:451–463
 98. Quaglia L, Verdun A (2022) Explaining the response of the ECB to the COVID-19 related economic crisis: inter-crisis and intra-crisis learning. *J Eur Public Policy*. <https://doi.org/10.1080/13501763.2022.2141300>
 99. Sedelmeier U (2011) Europeanisation in new member and candidate states. *Living Rev Eur Gov* 6:1
 100. Mottaleb KA, Kruseman G, Snapp S (2022) Potential impacts of Ukraine-Russia armed conflict on global wheat food security: a quantitative exploration. *Glob Food Sec* 35:100659
 101. Dasgupta P (2021) *The economics of biodiversity: the Dasgupta review*. Hm Treasury, London
 102. Fletcher R (2021) *The economics of biodiversity: the Dasgupta review*. *J Political Econ* 28:1–5
 103. Hermoso V, Carvalho SB, Giakoumi S, Goldsborough D, Katsanevakis S, Leontiou S, Markantonatou V, Rumes B, Vogiatzakis IN, Yates KL (2022) The European union biodiversity strategy for 2030: opportunities and challenges on the path towards biodiversity recovery. *Environ Sci Policy* 127:263–271

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