

# Preferences for climate change policies: the role of co-benefits

Jens Abildtrup, Jette Bredahl Jacobsen, Suzanne Elizabeth Vedel, Udo Mantau, Robert Mavsar, Davide Pettenella, Irina Prokofieva, Florian Schubert, Anne Stenger, Elsa Varela, et al.

# ▶ To cite this version:

Jens Abildtrup, Jette Bredahl Jacobsen, Suzanne Elizabeth Vedel, Udo Mantau, Robert Mavsar, et al.. Preferences for climate change policies: the role of co-benefits. Journal of Environmental Economics and Policy, 2024, 13 (1), pp.110-128. 10.1080/21606544.2023.2223182 . hal-04132398

# HAL Id: hal-04132398 https://hal.science/hal-04132398

Submitted on 6 Jul 2023

**HAL** is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers. L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

## 1 Manuscript submitted Journal of Environmental Economics and Policy

2 14/05/2023, published 18/06/2023: https://doi.org/10.1080/21606544.2023.2223182

3

## 4 Title: Preferences for climate change policies: the role of co-benefits

- 5 Abildtrup, Jens<sup>1,†</sup>
- 6 Jacobsen, Jette Bredahl<sup>2,3</sup>
- 7 Vedel, Suzanne Elizabeth<sup>4</sup>
- 8 Mantau, Udo<sup>5</sup>
- 9 Mavsar, Robert<sup>6</sup>
- 10 Pettenella, Davide<sup>7</sup>
- 11 Prokofieva, Irina<sup>8</sup>
- 12 Schubert, Florian<sup>9</sup>
- 13 Stenger, Anne<sup>10</sup>
- 14 Varela, Elsa<sup>8</sup>
- 15 Vidale, E.<sup>7</sup>
- 16 Thorsen, Bo Jellesmark<sup>2,3</sup>
- 17
- 18

36

41

### 19 Affiliations

- 20 <sup>1</sup> Université de Lorraine, Université de Strasbourg, AgroParisTech, CNRS, INRAE, BETA, 54000 Nancy, France
- <sup>21</sup>
   <sup>2</sup> Department of Food and Resource Economics, University of Copenhagen, Rolighedsvej 23, 1958 Frb C, Denmark.
- <sup>23</sup>
   <sup>3</sup> Centre for Macroecology, Evolution and Climate, University of Copenhagen, Rolighedsvej 23, 1958 Frb C, Denmark.
- <sup>4</sup> Department of Analytics, The Technical and Environmental administration, City of Copenhagen, Njalsgade 13,
   DK-2300, Denmark.
- <sup>5</sup> INFRO-Informationssysteme f
  ür Rohstoffe, An der Lachte 13, 29223 Celle, Germany.
   31
- <sup>6</sup> European Forest Institute, Yliopistokatu 6, 80100 Joensuu, Finland
- <sup>7</sup> Department of Land, Environment, Agriculture and Forestry, University of Padova, Vale Università 16, 35020
   Legnaro PD, Italy
- <sup>8</sup> Forest Science and Technology Centre of Catalonia, Ctra. St. Llorenç de Morunys km 2, 25280 Solsona (Lleida),
   Spain.
- 40 <sup>9</sup> INTEND Geoinformatik GmbH, Johanna-Waescher-Straße 5, 34212 Kassel
- <sup>10</sup> Université de Strasbourg, Université de Lorraine, AgroParisTech, CNRS, INRAE, BETA, 67000, Strasbourg
   France
- 44 <sup>†</sup> Corresponding author

## 46 Acknowledgement

- 47 This research was funded by the European Union under EU FP7 (Grant no.: 243950 NEWFOREX).
- 48 Disclosure statement: The authors report there are no competing interests to declar

#### 49 Preferences for climate change policies: the role of co-benefit

#### 50 Abstract:

51 Policies mitigating climate change provide a global public good but are also likely to imply local co-benefits where implemented. 52 This may affect citizens' preferences for what policy to implement as well as where to implement it. This aspect remains 53 understudied despite its relevance for international climate negotiations, national policies, and the development of voluntary 54 carbon credit markets. The results of a discrete choice experiment show that citizens in five countries (Denmark, France, 55 Germany, Italy and Spain) have quite similar mean willingness to pay for carbon emission reductions and agree on the ranking of 56 policies targeting different sectors. Specifically, policies targeting renewable energy use, are preferred over policies targeting 57 industrial energy efficiency or carbon sequestration and biomass production in forests. Applying follow-up questions shows that 58 concerns over co-benefits, notably air pollution, is linked to preferences for implementation in the home country. In the absence 59 of co-benefits, citizens are indifferent or prefer policies implemented in other countries. 60 Key policy insights Citizens in five European countries share preferences for climate change mitigation 61 62 policies, though significant intra-national heterogeneity in preferences exist Policies targeting increased use of renewables are preferred over policies targeting 63 64 improved energy efficiency in the industry. 65 Citizens express preferences for policies implemented in their own country. This is associated with their perception of co-benefits. In particular, consideration of reduced air 66 pollution as a side effect of investing in renewable energy and in energy efficiency in the 67 industry are important determinants of preferences for national implementation of 68 policies. 69 Preferences for national co-benefits may both enhance policy acceptance and reduce 70 71 willingness to support policies implemented in other countries. The latter aspect may reduce cost-effectiveness across countries but ease effort-sharing negotiations. 72

Keywords: Carbon emissions, co-benefits, willingness to pay, choice experiment, cross-country study, policy
 acceptability

75 76	<b>1. Introduction</b> Reaching international agreements on climate change mitigation remains high on the agenda of
77	governments and intergovernmental institutions (Vasconcelos, Santos, and Pacheco 2013). As
78	greenhouse gases distribute globally in the atmosphere, reducing emissions is a global public
79	good as no exclusion or rivalry is possible with regard to the climate benefits from carbon
80	emission reductions. Therefore, the international policy debate has focused on instruments that
81	could promote cost-effective policies and reduce emissions at the lowest marginal costs across
82	countries (Hepburn and Stern 2008; Kriegler et al. 2014, Stiglitz 2019).
83	While a global agreement on carbon emission reductions in Paris 2015 (COP21) was reached,
84	international agreements on concrete actions have lagged behind (Wagner and Zeckhauser 2012;
85	Vasconcelos et al. 2013), and explanations for this include the sheer complexity of multiparty
86	negotiations (Wagner and Zeckhauser 2012; Smead et al. 2014). The pursuit of cost-
87	effectiveness in some cases clashes with views on justice (or distributional and historical
88	fairness) (Grubb 1995; Gardiner 2004; Adger et al. 2009; Wei et al. 2012; Anderson and
89	Bernauer 2016; Ščasný et al. 2017; Jagers et al. 2021). People in different countries often differ
90	in their perceptions of climate change and in their support for mitigation policies (Leiserowitz
91	2006; Reiner et al. 2006; Adger et al. 2009; Ščasný et al. 2017; Kàcha, Vintr, and Brick 2022).
92	They also differ in their trust in the ability of (foreign) governments and institutions to undertake
93	effective mitigation actions (Akter, Bennett, and Ward 2012), or their perception of burden-
94	sharing design (Carlsson et al. 2013; Ščasný et al. 2017). Finally, while mitigation is a global
95	public good, any specific mitigation policy is likely to have side effects in the form of co-
96	benefits or collateral damages that are often predominantly local, for example changes in local
97	air pollution, visual disamenities, traffic risk, etc. (Pearce 2000; IPPC 2001 chapter 8; Bollen,
98	Jamet, and Corfee-Morlot 2009; West et al. 2013; Sovacool et al. 2020). The not-in-my-

99 backyard (NIMBY) effect is also a well-known example of local disamenities, well-studied

100 within renewable energy (Batel, 2020). However, much less emphasis has been given to co-

101 benefits. Yet, both mayaffect public policy preferences (Pittel and Ruebbelke 2008; Svenningsen

and Thorsen 2020), and thereby policy choices at a larger scale.

103 If we look at current policies, public perceptions play a large role. The European Union's Council 104 recently decided Fit-for-55 plan of reducing emissions by 55% in 2030 relative to 2005, and details are currently being negotiated and adopted<sup>1</sup>. It is interesting in this perspective as the plan contains 105 both a common goal and national aggregate targets (yet to be distributed) but also sector-specific 106 targets. It is a part of the Green Deal, which contains multiple objectives, not only emission 107 reductions. For example a just transition, climate adaptation, and ambitions of increased welfare. 108 If we look at instruments, we see both a cross-national instrument: a quota trading scheme for 109 certain sectors, but also areas where national regulation is by large decided independently from the 110 common EU policies. In addition, we see fast-growing voluntary carbon credit markets (Kreibich 111 112 and Hermwille 2021). This complexity of goals and instruments can be interpreted as a result of political reality, and therefore, it is of relevance to investigate the general public's perception, as 113 114 they are ultimately the ones choosing the politicians.

115

How much the public is willing to pay for carbon emission reductions has been investigated in the economic literature on climate change (Alló and Loureiro 2014; Drews and van den Bergh 2015) and an overview of studies can be found in Alberini et al. (2018). Studies have predominantly focused on carbon emission reductions from a single policy measure targeting

specific sectors, e.g. addressing transport choices (Brouwer, Brander, Van Beukering 2008;

 $<sup>^1\,</sup>EU's\ fit-for-55\ plan:\ https://www.consilium.europa.eu/en/policies/green-deal/fit-for-55-the-eu-plan-for-a-green-transition/$ 

121	Achtnicht 2012), energy choices (Ek 2005; Longo, Markandya, Petrucci 2008; Solomon and
122	Johnson 2009; Scarpa and Willis 2010; Akter and Bennett 2011; Adaman et al. 2011; Hanemann,
123	Labandeira, and Loureiro2011; O'Keeffe 2014; energy efficiency (Longo, Hoyos, and
124	Markandya 2012), household reduction measures (Faure et al. 2022), or more loosely defined
125	policies (Carlsson et al. 2012; Ščasný et al. 2017). Thus, policies targeting alternative sectors are
126	rarely evaluated against each other. This is unfortunate as different sectors may face different
127	opportunity costs of emissions, and different policies are likely to come with different co-
128	benefits, costs and potentially negative externalities. The willingness to pay (WTP) for co-
129	benefits and other ancillary effects of mitigation measures and carbon offset programs has been
130	investigated (MacKerron et al. 2009; Kotchen, Boyle, and Leiserowitz 2013; Rodriguez-Entrena,
131	Espinosa-Goded, and Barreiro-Hurle 2014; Torres et al. 2015; Baranzini, Borzykowski, and
132	Carattini 2018), but not across different policies and countries. Furthermore, few studies have
133	elicited comparable WTP for carbon reductions across different countries (Carlsson et al. 2012;
134	Ščasný et al. 2017; Faure et al. 2022). Some studies have reported that citizens think mitigation
135	policies should be implemented in their own country rather than in other countries (Carlsson et
136	al. 2012; Buntaine and Prather 2018; Svenningsen and Thorsen 2020), while other studies
137	(Diederich and Goeschl 2017; Baranzini, Borzykowski, and Carattini 2018) have found no
138	significant preferences for mitigation locally (European) over mitigation in developing countries.
139	However, to the best of our knowledge, the perceived role of co-benefits for the preferences for
140	an action to be undertaken in peoples' home country relative to alternative countries has not been
141	analyzed in detail before. Anderson and Bernauer (2016) suggest for future research, based on a
142	study analyzing factors determining preferences for carbon offsetting in the US, that 'it would
143	be interesting to examine how various types of co-benefits influence preferences for domestic

versus international offsetting. Examples of such co-benefits are green technology innovation, *local air quality improvement, and biodiversity protection*'. Torres et al. (2015) find a decreasing
WTP for forest carbon services with the distance to the site of implementation but only
considering mitigation projects within Mexico.

148

149 This paper applies a discrete choice preferences elicitation technique (Train 2003) to estimate the WTP for various carbon emission reduction policies. As in Ščasný et al. (2017) and Faure et al. 150 (2022), the estimated WTP is used as an indicator of the policy acceptance. Therefore, the focus 151 is not to estimate the size of the marginal WTP for carbon emission reductions per se, but to test 152 the following three hypotheses: The first null hypothesis is that the WTP for carbon reduction 153 policies does not vary between policies. Three policies (types of measures) are considered: 154 investment in energy-saving production technologies in the industry, investment in renewable 155 energy and investment in forest management and wood use to increase CO<sub>2</sub>-sequestration. These 156 157 three groups do not cover all potential mitigation options, but represent important options (Kriegler et al. 2014) that differ substantially in how mitigation takes place, i.e. energy 158 159 efficiency, energy production and carbon sequestration, respectively. We have no a priori 160 expectations about the relative preferences for the three policies. The second null hypothesis is that the WTP for implementing measures in the home country is identical to implementing it in 161 162 other countries (Carlsson et al. 2012; Buntaine and Prather 2018). In other words, respondents 163 will not give up cost-effectiveness to ensure local implementation even if they are reminded that the impact is independent of the location of the mitigation measures. The empirical literature 164 does not show a clear tendency regarding the impact of location, e.g. Diederich and Goeschl 165 166 (2017) did not find a positive preference for local reductions in emissions. The third null

hypothesis is that the WTP for local implementation of a measure does not depend significantly 167 on the respondents' perception of co-benefits associated with the measure. However, according 168 169 to other findings in the literature the a priori expectation is, that local implementation matters for some measures. If the co-benefits are perceived to be positive and considered important, the 170 WTP is higher and if considered negative and important the WTP is lower (Torres et al. 2015, 171 172 Longo, Hoyos, and Markandya 2012). Considering side effects of forest measures on recreation possibilities and biodiversity (the use value of local nature protection) protection are expected to 173 strengthen the preferences for local implementation (Elbakidze and McCarl 2007; Glenk and 174 Colombo 2011; Torres et al. 2015; Bakhtiari et al. 2018). Similarly, we expect a positive impact 175 on preferences for local implementation of renewable energy and industry efficiency measures if 176 the respondents consider reduction of air pollution a positive side effect (Nemet, Holloway, and 177 Meier 2010; West et al. 2014). On the other hand, we would expect that the WTP is lower for 178 these reduction measures if the respondents consider negative local employment effects resulting 179 180 from these measures (Krupnick, Burtraw, and Markandya 2000; O'Keeffe 2014; Bernauer and Gampher 2015). If any of these assumptions holds, we must reject the null hypothesis. However, 181 again the existing empirical evidence is not clear. For example, Baranzini, Borzykowski, and 182 183 Carattini (2018) find, in an experiment with students in Geneva, no significant positive effect on preferences for local implementation with an information treatment emphasizing local co-184 benefit<sup>2</sup>. 185

186

These hypotheses are tested among representative samples of the citizens of five European
 countries (Denmark, France, Germany, Italy and Spain). The following section presents the data

<sup>&</sup>lt;sup>2</sup> The fact that Baranzini, Borzykowski, and Carattini (2018) did not find an effect of an information treatment emphasizing local co-benefits, is not the same as rejecting that respondents consider co-benefits.

collection and econometric analysis. Section 3 presents the results followed by a discussion,concluding remarks and perspectives.

191

#### 192 Material and Methods

#### 193 Survey design and data collection

A first version of the questionnaire was developed in English in co-operation between national teams of researchers. This allowed a common and careful deliberation of each component of the questionnaire and ensured a rigid translation process. Following this, the questionnaire was translated into the five languages of the five countries participating in the survey and was tested on the local populations. Based on results and feedback from these pilot tests, the English language common questionnaire was revised and translated again.

200 The final questionnaire (see supplementary material) had the following structure: First, a part

with general questions on knowledge and attitudes towards climate change was presented,

followed by the introduction of a choice experiment (see e.g. Mariel et al. 2021) which was used

to elicit the preferences for carbon reduction policies among citizens of the five European

204 countries. After the choice experiment, followed a section, which gathered information on the

205 respondents' attitude towards and consideration of co-benefits.

Respondents were given a careful introduction to the different policy aspects included in the choice experiment. They were informed that the EU member states have already committed themselves to carbon emission reductions and that they were now to consider the choice among policies relevant to achieve this goal. They were also informed that while their responses were entirely anonymous, the results of the survey would eventually be communicated to policy makers and could influence future policy developments. This to ensure consequentiality (Mariel

212	et al. 2021). Respondents were then asked to consider that, should further policies be
213	implemented along the lines suggested from the results of the experiment, the financing of these
214	policies would come from the member states and their citizens. More precisely, the costs of the
215	policies would be financed by an additional tax levied on all European households' electricity
216	bill <sup>3</sup> , and in each alternative they were asked to consider the stated specific costs to their
217	household. Respondents were also reminded that the proposed policies do not solve the problem
218	of CO <sub>2</sub> in the atmosphere entirely, but are only some policy initiatives among others, which will
219	contribute to the reduction of the total $CO_2$ emissions by the EU countries. Thereby we stressed
220	the marginal aspect. They were also informed about the global public good nature of carbon
221	emission reductions, including that CO <sub>2</sub> emissions distribute in the atmosphere, and therefore it
222	does not matter for the climate where emissions are reduced. To provide an intuition about the
223	scale of the CO <sub>2</sub> emission reductions, this information was presented in terms of annual
224	emissions from the corresponding number of average European households.
225	Next, the different aspects (attributes) of the policies were explained, including the levels these
226	could take. In the choice sets the attribute levels were described with text and icons (see Figure 1
227	for an example). The policy alternatives would target effects in one of three sectors: Either
228	increased energy efficiency in the industry or investing more in renewable energy technology, or
229	changes in forest management to increase carbon sequestration and use of wood. The policy
230	alternatives were briefly explained in general terms without reference to potential co-benefits
231	(Figure 2). As our purpose was to assess the preferences for different main sectors where
232	reductions could take place, the descriptions did not detail the specific potential technologies, but

<sup>&</sup>lt;sup>3</sup> The electricity bill as payment vehicle is also used in Aldy, Kotchen, and Leiserowitz (2012) to reveal WTP for climate mitigation policies. As no common EU level taxation scheme exists, this was considered the most coherent payment vehicle across countries - even though energy also comes from other sources than electricity. The payment vehicle was also accepted as realistic among participants in the focus groups.

were kept in generic terms. For example, the measure "forest management and wood use" may 233 reflect reduced harvest to increase carbon stock, changes in tree species to increase sequestration, 234 or a change in wood use, either to increase sequestration of carbon in wood products or to 235 increase substitution of fossil fuel-based products. The policy would be implemented in only one 236 of the five countries, and which one was shown in a separate line. The policy differed also in the 237 amount of carbon emission reductions it produced and finally also in the cost to the household. 238

	Policy 1	Policy 2	Policy 3	No policy
Sector targeted	Forest management and use of wood	Industry	Renewable energy	None of them
Country where measure is implemented	Denmark	Spain	France	-
Annual reduction of CO <sub>2</sub> in million of tonnes (Mt)	4 Mt CO <sub>2</sub> (1 million households' emission)	<b>10</b> Mt CO <sub>2</sub> ( <b>2.5</b> million households' emission)	8 Mt CO <sub>2</sub> (2 million households' emission)	0
Extra <b>annual</b> cost of energy for <b>your</b> household for the next 5 years	€ 8	€2	€ 60	€ 0
TICK here →	$\downarrow$	$\downarrow$	$\downarrow$	$\downarrow$

Figure 1. A presentation of a choice set. Each respondent was asked to indicate their preferred policy among three 239

240 different policies and a 'No policy' alternative. Note the cost to the household being explicitly stated and tied to

241 each alternative. Attribute levels vary between alternatives and choice sets. Each respondent answered eight such

242 choice sets.

# Industry

This includes investments in energy-saving production technologies to improve the effectiveness of raw material consumption, and to improve transport modes and logistic systems to decrease fuel consumptions.

# **Renewable energy**

This includes investments in renewable energy technology, for example, establishing windmill parks, solar energy projects and the enhanced use of biomass-based energy.

## Forest management and the use of wood

Trees and other plants bind  $CO_2$  when they grow. Therefore, we can reduce  $CO_2$  in the atmosphere by e.g. increasing forest stock or by changes in wood use.

- Figure 2. The information given to the respondents about the reduction strategies in the questionnaire
- 246 The design of the choice experiment involved four different attributes each with three to 10
- 247 different levels, all explained to the respondents. These are outlined in Table 1. In each choice
- task the respondent had the option to pick either of three alternative policies or no additional
- 249 policy, a status quo option. A d-efficient design was generated using the software Ngene for a
- 250 multinomial logit model with zero priors for main effects.

251

Table 1 Attributes and attribute levels used to describe the policy alternatives, see the questionnaire extract abovefor details of the presentation to respondents.

	Attribute	Policy attribute levels	Status quo (no policy)
	Sector targeted	-Forest management and use of wood -Industry	No measure
	Country where measure is implemented	-Kenewable energy	
	Country where measure is implemented	-Germany	-
		-France	
		-Italy	
		-Spain	
	Annual CO <sub>2</sub> reduction, million tonnes (Mt)	4, 6, 8, 10, 12	0
	Extra annual cost of energy for your		€0 (status quo)
	household for the next 5 years		
254 255 256	<sup>a</sup> In Denmark these units were converted into I 200, 500, 1000	DKK ( $\in 1=7.5$ DKK) and rounded to the levels 3	3, 7, 15, 30, 60, 100,
257 258	The survey data were collected through	h an online survey platform using samp	les of the
259	population in Denmark, France, Germa	any, Italy, and Spain, where representat	iveness was
260	targeted with regard to gender and age	. A sample size of around 400 was targe	eted and achieved
261	in all countries, cf. Table 2 below. The	survey was carried out using the Surve	eyXact software
262	and distributed by local polling agencies	es in the different countries during Apri	l 2013. The sample
263	is representative with regards to gende	r. There are some national deviations of	n other socio-
264	demographics, e.g. in Denmark, Germa	any, and Spain the samples are slightly	underrepresented
265	with regard to young people (< 25 year	rs old) and relatively overrepresented w	ith older people
266	(>65 years old). In France and Italy san	mples are relatively underrepresented w	vith respondents in
267	the oldest age class. While Table 2 rev	eal the representativeness with respect t	to socio-
268	demographic characteristics we are not	t able to evaluate to which degree, the r	espondents are
269	representative with respect to knowled	ge and preferences for climate mitigation	on policies.
270	However, we have no reason to believe	e that the probability of membership of	the panels of the
271	polling agencies is dependent on their	attitude to climate mitigation policies. I	Furthermore, when
272	respondents were invited to answer the	e questionnaire they were not told that the	he questionnaire

was about climate policy and therefore it is less likely that respondents self-selected into the

274 questionnaire.

Table 2 Age and gender distribution in the samples compared to the distributions in each country (2013) for the age

276	range 18-	70 targeted	in survey	(%)
-----	-----------	-------------	-----------	-----

	Age class	Denmark	Germany	France	Italy	Spain	Average
Population	18-24	13.2	11.5	12.8	10.4	10.2	11.6
Sample	18-24	5.0	2.4	11.9	10.0	8.4	7.5
Population	25-34	17.0	17.8	18.6	17.0	20.1	18.1
Sample	25-34	8.8	18.9	27.0	24.1	23.5	20.4
Population	35-44	19.9	19.0	20.2	22.6	24.2	21.2
Sample	35-44	13.8	18.4	22.7	29.1	21.0	21.0
Population	45-54	21.0	24.2	20.6	22.3	21.2	21.9
Sample	45-54	23.0	22.9	19.7	22.6	11.3	19.9
Population	55-64	18.2	18.9	19.2	18.2	16.2	18.1
Sample	55-64	32.3	22.9	14.6	13.0	19.7	20.5
Population	65-70	10.7	8.5	8.6	9.4	8.1	9.1
Sample	65-70	17.3	14.6	4.0	1.3	16.2	10.7
Population	Female	50.4	50.8	51.5	51.6	50.7	51.0
Sample	Female	49.8	49.6	50.9	50.6	49.8	50.1

Source Eurostat appsso.eurostat.ec.europa.eu/nui/show.do?dataset=demo\_pjan&lang=fr Table : Population on 1
 January by age and sex Code : demo\_pjan

- 279
- 280

281

A total of 2208 completed questionnaires were gathered (DK: 402, D: 429, F: 419, I: 470, E: 282 283 488). 78 observations were excluded, because respondents had reported an age outside the targeted interval of 18-70 years. Respondents who had chosen the answer "no policy" in all 284 285 choice sets where asked a follow-up question. If they replied "I do not want to pay more in taxes and I think the government should finance it from currently collected taxes" or "I think the large 286 polluters (industries) should pay as they are polluting the most" they were excluded as their 287 288 answers indicate that they did not make a deliberate trade-off between the choice alternatives. This was the case for only 30 respondents distributed over all countries. 289

#### 291 Measures of co-benefit concern

The survey design was prepared for an analysis of the importance of co-benefits, which was 292 formulated to the respondents as 'side effects' in their local language as 'co-benefits' was 293 considered technical jargon. Specifically, respondents were asked follow-up questions after the 294 choice experiment about to what degree they considered a set of different side effects, when 295 296 making the policy choices they just made. Respondents rated the consideration they assigned to each type of co-benefit on a four-point Likert scale with the intervals "Not at all", "Hardly", "A 297 little" to "A lot". The side effects considered was an outcome of statements from respondents 298 during the pre-testing phase of the questionnaire. Thus, it reflects effects that respondents expect 299 - not necessarily the actual effects. As our aim was to investigate co-benefits, we emphasized 300 positive such. There are obviously other externalities, depending on which specific technology is 301 considered within each sector. Our aim with the co-benefit concerns was to investigate their 302 correlation to the prior stated WTP. Thus, to avoid too lengthy a survey, we opted for a single 303 304 statement for each co-benefit (rather than e.g. formulate opposing statements with both a positive and a negative effect). 305

306

#### 307 *Econometric models and analyses*

The choice experiment relies on Lancaster's consumption theory (Lancaster 1966), stating that the utility of a good, in the present case an alternative policy, is derived from the attributes of that alternative good. Further, this is combined with random utility theory (McFadden 1974, Train 2003), where it is assumed that each individual choose the alternative among a set of alternatives that maximize expected utility.

Specifying the utility function in preference-space, the respondents are indexed by *n*, their chosen policy by *i*, the household cost of the policy alternative by  $p_i$  and the attributes of the alternative are all included in the vector  $x_i$ . Thus, the utility, *U*, of choosing a particular alternative may be described as:

317

$$\begin{array}{l} 318 \\ 319 \end{array} \quad U_{ni} = asc + \alpha p_i + \boldsymbol{\beta}' \boldsymbol{x}_{ni} + \varepsilon_{ni} \tag{1}$$

320 Here,  $\alpha$ , and  $\beta$  are parameters to be estimated and  $\mathcal{E}_{ni}$  is the unobserved component of utility.

321 asc = 1 if the "*Policy 1*", "*Policy 2*" or "*Policy 3*" are chosen; otherwise *zero*.  $\varepsilon_{ni}$  is assumed 322 to be a stochastic, *iid* extreme value distributed error term.

To enable an easy comparison across country specific models in the analyses, equation (1) is reformulated to obtain a representation of the respondents' preferences in so-called willingness to pay (WTP) space (Train and Weeks 2005):

- 326
- 327

328 
$$U_{ni} = \alpha (\overline{asc} + p_i + c' x_{ni}) + \varepsilon_{ni}$$
(2)

Here  $c = \frac{\beta}{\alpha}$  and  $\overline{asc} = \frac{asc_i}{\alpha}$  where c'is a vector of marginal WTP estimates, each related to an attribute of the policy alternative in question. Notice that this implies that the WTP is estimated directly thereby avoiding the problem of unstable marginal WTP estimates based on ratios of parameters estimated in preference-space models (Thiene and Scarpa 2009).

To account for heterogeneity in preferences and hence WTP for the different aspects of policies across the respondents in each country, for each country specific data set a mixed logit specification was applied. This is a flexible and computationally practical econometric model, which in principle may be used to approximate any discrete choice model derived from random utility maximization (Train 2003; Train and Weeks 2005). It allows for heterogeneity across respondents by specifying random variation in the coefficients *c* of equation (2). The distributions of the random coefficients are defined as  $f(c|\omega)$  where  $\omega$  denotes a matrix of parameters of these distributions. This allows writing the unconditional choice probability of a sequence of the *T* choices  $y_n = \{i_{1n}, i_{2n}, ..., i_{Tn}\}$  made by respondent *n* as:

342 
$$P(y_n) = \int \prod_t^T \frac{e^{\alpha(\overline{asc} + p_i + c'x_{ni})}}{\sum_i e^{\alpha(\overline{asc} + p_j + c'x_{nj})}} f(\boldsymbol{c}|\boldsymbol{\omega}) d(\boldsymbol{c})$$
(3)

343 Continuous distributions for the WTP of the policy attributes are assumed normal.

344 The test of the third hypothesis that co-benefit concerns are of no concern for local

implementation of reduction measures is undertaken by including the replies to the question on

co-benefit concern in the econometric model (Longo, Hoyos, and Markandya 2012). More

347 specifically interaction terms,  $z_{kn}$ , between the various co-benefit concerns and the dummies

representing the different policy sectors (renewable energy, industry, and forest management) arecreated and included in (2):

350

351 
$$U_{ni} = \alpha(\overline{asc} + p_i + c'x_{ni} + \gamma_1 z_{1n}h_{ni} + \dots + \gamma_6 z_{6n}h_{ni}) + \varepsilon_{ni}$$

352

where  $z_{kn}$  takes on a value of one if the respondent *n* has replied "*a lot*" or "*a little*" for combination, *k*, of side effect concern and policy measure as defined in table 3; otherwise zero.  $h_{ni}$  takes the value of one if the policy measure is implemented in the country where the respondent is living; otherwise 0.

(4)

Table 3 Definition of the interaction terms  $z_{kn}$  by side effect concern question and policy measure.

	Forest management and use of wood	Industry	Renewable energy
Side effect concern "A little" or "A lot"			
Forest management's impact on biodiversity	$z_{In}=1$		
Forest management's impact on recreational opportunities That reducing CO <sub>2</sub> -emissions from industry will give cleaner air	<i>z</i> <sub>2<i>n</i></sub> =1		
locally		$z_{3n}=1$	
That reducing CO <sub>2</sub> -emissions by increased use of renewable energy			
will give cleaner air locally			$z_{4n}=1$
That implementing reduction measures may have a negative impact			
on local employment		$z_{5n} = 1$	$z_{6n} = 1$

#### 360 **Results**

#### 361 *The main effects across countries*

The main results for the choice model are shown in Table 4. Note that the country where the 362 policies are implemented and the different policies enter as 0-1 dummy variables in each of the 363 364 estimations. Therefore, one country variable and one policy variable are omitted for the estimation to be feasible. In each country's regression, the home country is omitted and the forest 365 management policy is omitted. These are confounded with the constant of the regression; the 366 ASC. Thus, the estimated parameters show if people on average prefer implementation in other 367 countries more or less than in their own country, and if they on average prefer a policy targeting 368 energy efficiency in the industry or investments in renewable energy more or less than policies 369 targeting forest management and use of wood. For each attribute the rows show the estimated 370 mean WTP in the countries populations and the standard deviation (SD) of the WTP distribution. 371

**Table 4:** Mean annual willingness to pay (WTP) for policy attributes, measured in EURO (2012 PPP EU27;

EUROSTAT ESA95 aggregated) per household per year for the next five years. For the sector and country of

implementation attributes, we have used "forest management and wood use" and home country as reference
 variable, respectively. S.D. is the estimated standard deviation of the WTP distribution. Standard errors of estimated

parameters (WTP means and standard deviations) are reported in parentheses. Bold figures are significantly different

378 from zero at the 5% level and italics at 10% level.

	Respondents in:									
Policy		Denmark	F	rance	Germ	nany	Italy		Spain	
$CO_2$ (Mt/year)	WTP	<b>6.78</b> (1.11)	7.22	(0.80)	6.93	(0.88)	8.29	(1.26)	11.6	(1.49)
CO <sub>2</sub> (Witrycar)	S.D.	<b>15.67</b> (1.47)	9.5.2	(0.87)	11.73	(1.00)	15.8	(1.66)	20.3	(1.83)
Target industry	WTP	-7.08 (5.59)	-20.36	(4.65)	-11.00	(4.97)	2.94	(7.88)	-14.42	(8.45)
	S.D.	<b>54.02</b> (7.30)	38.33	(6.29)	47.41	(6.67)	87.22	(11.03	94.98	(10.95)
Target renewable energy	WTP	<b>28.54</b> (4.91)	21.80	(3.75)	33.62	(4.50)	66.20	(8.04)	48.19	(7.42)
	S.D.	<b>55.01</b> (6.20)	39.00	(5.30)	49.32	(5.31)	82.88	(9.97)	85.91	(9.60)
Implementation in Denmark	WTP		-42.87	(6.41)	-65.14	(8.30)	-52.82	(12.17	-89.82	(12.61)
	S.D.		52.21	(7.63)	48.33	(8.18)	108.88	(15.22	113.19	(14.01)
Implementation in France	WTP	<b>-29.08</b> (4.89)			-44.59	(5.40)	-22.85	(8.11)	-52.66	(7.86)
	S.D.	1.32 (12.99			1.31	(9.35)	4.24	(18.20	1.01	(13.67)
Implementation in Germany	WTP	<b>-26.87</b> (6.59)	-15.04	(5.26)			-5.36	(8.83)	-55.61	(9.61)
	S.D.	0.46 (11.15	31.53	(10.82)			13.41	(30.53	3.73	(23.07)
Implementation in Italy	WTP	<b>-44.84</b> (6.27)	-42.84	(5.52)	-71.86	(7.72)			-94.04	(10.96)
	S.D.	12.16 (21.13	1.98	(13.06)	8.23	(40.68)			1.92	(17.23)
Implementation in Spain	WTP	<b>-22.56</b> (5.07)	-23.43	(4.21)	-50.54	(5.82)	-9.41	(8.34)		
	S.D.	<b>26.60</b> (10.66	0.46	(11.18)	1.04	(11.21)	0.90	(45.06		
Technical constant (ASC)	WTP	<b>86.0</b> (9.92)	106.8	(9.88)	114.90	(10.59)	204	(21.2)	182.2	(18.23)
Number of respondents		391 (1955)		422 (2110)	38	34 (1920)	462	(2310)	44	1 (2205)
Log Likelihood		-2027.6		-2062.5		-1895.8		-2375.5		-2225.3
McFadden pseudo R <sup>2</sup>		0.252		0.295		0.288		0.258		0.272

<sup>379</sup> 

It is noted that in several cases there is considerable heterogeneity in the preferences among 380 respondents in the different countries. This is indicated by the significant standard deviations of 381 the preference distributions; cf. equation (3). While the respondents in different countries have a 382 very similar mean WTP for carbon reductions there is considerable heterogeneity in the 383 population with respect to this attribute in all the countries. The opposite is the case for e.g. the 384 "Implementation in Italy" attribute, where the estimated standard deviations of the parameters 385 are not significant in any of the four countries. The overall model fit as indicated by the 386 McFadden pseudo  $R^2$  between 0.252 and 0.295 across the models. 387

A first finding is the large similarity across countries in the mean marginal value per Mt for reduced CO<sub>2</sub> emissions. It is almost identical for households from Denmark, Germany and France but higher for Italy, and in particular higher for Spain. However, the difference in WTP is only statistically significant between Spain and Denmark, Germany and France, applying a two-

393 sided t-test and a 5% level of significance.

The first hypothesis, that the WTP is the same for all three groups of policies is rejected. 394 Households in all countries on average significantly prefer policies targeting investment in 395 renewable energy use to policies targeting either of the other sectors. Policies targeting energy 396 efficiency in the industrial sector are not preferred over policies targeting the forest sector; for 397 France and Germany, they are even preferred less. It should be noted here that these results do 398 not compare specific forest management or wood use measures with specific measures in the 399 renewable energy or industry sectors, but only how the population overall consider measures in 400 401 these sectors.

Households in most countries on average agree that policy implementation in their home country
is preferable to implementation elsewhere, and thereby we reject the second hypothesis that WTP
for reduction abroad is the same as for reductions in the home country. Only Italy is marginally
indifferent to implementation in Germany, Spain or at home. Implementation in Italy or
Denmark is the least preferred by other countries.

Another finding of interest is evident from the regression constant. The constant can be
 interpreted as the willingness to support any mitigating actions above the status quo (irrespective
 of place, sector and impact), plus a correction for the confounded baseline variables for a policy

410 targeting forest management in the respondent's home country. The citizens of the

Mediterranean countries, Spain and Italy, have a significantly higher tendency to support any 411 alternatives to the status quo, compared to the three other countries and the citizens in France 412 have a significant higher tendency to support measures than in Denmark (5% level). 413 414 Accounting for co-benefit concerns 415 416 Co-benefits were, generally, considered important (Table 5). Especially, the impacts on air quality were considered important. On average over the five countries, 80% of the respondents 417 did consider the positive impact of increased use of renewable energy on the local air quality "a 418 lot" or "a little", while only 20% did "not at all" or "hardly" consider these co-benefits in their 419 choice of policy. Similarly, 77% of the respondents did consider the impact of reducing industry 420 emission on local air quality "a lot" or "a little". The impacts of policy measures on employment 421 were, on average considered the least important. 422

424 **Table 5.** Share of respondents replying "a lot" or "a little" on the question "To what extent did you take the

following side effects into account when choosing between the policies above?" The alternative answers were

426 "hardly" and "not at all".

	Denmark	Germany	France	Italy	Spain	Average
Forest management's impact on biodiversity	0.54	0.58	0.76	0.75	0.78	0.68
Forest management's impact on recreational opportunities That reducing CO2-emissions from industry will give		0.63	0.65	0.87	0.69	0.67
cleaner air locally	0.62	0.70	0.75	0.89	0.89	0.77
That reducing CO2-emissions by increased use of renewable energy will give cleaner air locally	0.68	0.70	0.78	0.91	0.91	0.80
That implementing reduction measures may have a negative						
impact on local employment	0.44	0.32	0.44	0.59	0.56	0.47
Average	0.56	0.59	0.68	0.80	0.77	0.68

427

428

429 A model for the preferences elicited in the choice sets, was estimated taking into account this

430 source of preference heterogeneity using interaction terms between the home country variable,

and the variable for associated policy-specific co-benefit concern cf. equation (4) and Table 3.

432 The results are reported in Table 6.

Table 6 Households' annual willingness to pay for policy attributes in Euros (PPP EU27) per household the next five years taking into account concerns about co-benefits. 

The base reference category is a policy targeting "forest management and wood use" for respondents not considering recreation and biodiversity co-benefits. Bold figures are significantly different from zero at the 5% level and italics on a 10% level. S.D. is the estimated standard deviation of the estimated WTP. Standard errors are in 

parentheses.

Policy attribute			Denmark		France	(	Germany		Italy	5	Spain	
CO <sub>2</sub> (Mt/year)	Row	WTP	6.85	(1.10)	6.29	(0.72)	6.97	(0.89)	7.87	(1.26)	10.83	(1.42)
		S.D.	15.4	(1.46)	8.63	(0.79)	11.6	(1.02)	15.8	(1.66)	19.0	(1.69)
The base reference category "Forest management and wood use" implemented in	(0)											
home country <u>and not concerned about biodiversity</u> and recreation co-benefits.												
"Forest management and wood use" with biodiversity concern and implementation	(1)	WTP		(2.0.5)		(7.00)		(= 00)		(1		(0.50)
in home country $(z_{In}=1, h=1)$ vs reference (0)			2.52	(7.95)	13.05	(5.90)	11.53	(7.09)	-12.16	(15.50)	11.15	(9.73)
<b>"Forest management and wood use</b> " with recreation concern <u>and</u> implementation in home country $(a + 1, b, 1)$ are reference $(0)$	(2)	WTP	20.27	(0,00)	5.00	(5.00)	10 (0	(7.01)	42.07	(10.00)	4.02	(0.22)
in nome country $(z_{2n}=1, h=1)$ vs reference (0)	(2)	WTD	30.27	(8.08)	5.98	(5.99)	18.00	(7.01)	42.97	(16.09)	-4.92	(9.23)
pollution and employment effect vs reference (0)	(3)	WIP	-1.47	(5.98)	-14.15	(4.67)	-8.80	(5.49)	5.59	(8.07)	-24.34	(9.15)
		S.D.	54.28	(7.27)	35.38	(5.66)	44.98	(6.81)	88.70	(10.96)	91.67	(10.27)
"Industry " with air pollution concern and implementation in home country $(z_{3n}=1,$	(4)	WTP										
h=1) vs reference (3)			22.73	(8.54)	9.96	(4.22)	24.77	(5.09)	16.52	(9.95)	26.09	(10.42)
"Industry" with employment concern and implementation in home country $(z_{5n}=1,$	(5)	WTP										
h=1) vs reference (3)			-3.42	(9.95)	-1.90	(4.64)	2.60	(5.67)	17.32	(8.17)	-12.77	(11.48)
" <b>Renewable energy</b> " not implemented in home country and/or without concerns for air pollution and employment effect vs reference (0)	(6)	WTP	31.46	(5.54)	30.17	(3.95)	37.50	(5.21)	60.67	(8.18)	32.84	(8.25)
		S.D.	53.62	(6.15)	36.71	(4.81)	46.27	(5.35)	86.08	(10.25)	82.83	(9.04)
"Renewable energy" air pollution concern and implementation in home country	(7)	WTP										
$(z_{4n}=1, h=1)$ vs reference (6)			21.47	(5.72)	1.16	(4.50)	23.25	(6.41)	41.22	(13.47)	23.53	(9.55)
"Renewable energy" Employment concern and implementation in home country	(8)	WTP										
$(z_{6n}=1, h=1)$ vs reference (6)			4.93	(5.98)	0.49	(5.02)	-0.32	(7.07)	2.71	(14.09)	-5.48	(9.99)
<b>Implementation in Denmark</b> relative to implementation in home country for	(9)	WTP			-12.32	(11.56)	13.41	(15.23)	60.13	(34.83)	-39.11	(25.94)
respondents not considering local co-benefits		S.D.			49.71	(6.77)	48.34	(8.16)	110.4	(15.58)	108.11	(12.95)
Implementation in France relative to implementation in home country for	(10	WTP	47.03	(13.71)			33.48	(14.43)	86.26	(34.46)	-2.35	(24.65)
respondents not considering local co-benefits		S.D.	2.28	(14.68)			1.20	(8.97)	2.05	(17.29)	0.74	(13.09)
Implementation in Germany relative to implementation in home country for	(11)	WTP	48.24	(14.35)	16.68	(11.11)			103.00	(34.99)	-3.42	(25.11)
respondents not considering local co-benefits		S.D.	0.46	(11.15)	28.30	(10.30)			6.16	(27.04)	4.51	(23.05)
Implementation in Italy relative to implementation in home country for respondents	(12)	WTP	30.51	(13.79)	-9.25	(11.26)	5.57	(14.98)			-42.85	(25.36)
not considering local co-benefits		S.D.	9.29	(22.24)	2.94	(14.41)	17.78	(15.78)			2.29	(16.75)
Implementation in Spain relative to implementation in home country for	(13)	WTP	54.70	(14.04)	7.59	(10.78)	28.40	(14.67)	100.11	(34.87)		
respondents not considering local co-benefits		S.D.	25.46	(10.67)	1.31	(12.68)	1.25	(10.80)	4.04	(38.07)		(
Technical constant (ASC)		WTP	6.10	(15.25)	67.21	(13.22)	35.02	(15.89)	99.47	(37.13)	133.6	(28.28)
Number of respondents (choices)			391	(1955)	422	(2110)	384	(1920)	462	(2310)	441	(2205)
Loglikelihood				-1998.1		-2039.2		-1873.3		-2361.1		-2216.8
McFadden pseudo $\mathbb{R}^2$				0.263		0.303		0.296		0.263		0.275

438	The parameters for the main effects of sector and country variables are estimated as random and
439	normally distributed in this model. The interaction variables between the 'co-benefit concern'
440	measures and the relevant sector targeted and the presence of 'own country' in the alternatives
441	chosen are not, the reason being that these variables themselves are explaining heterogeneity in
442	choices made. As was the case in the model of Table 4, there is also considerable preference
443	heterogeneity remaining in this more developed model. In particular, there is heterogeneity in the
444	preferences related to the different policy sectors and the actual mitigation measure. In this
445	model, preferences are generally more homogenous when it comes to the country of
446	implementation, and mainly finding considerable heterogeneity for implementation in Denmark.
447	The model extension increased the model fit somewhat, though not dramatically.
448	
449	The third hypothesis, that the perception of co-benefit has no impact on WTP can be rejected.
450	Accounting for concerns for local co-benefits eliminates or even reverses the preferences for
451	country of implementation. Note that row (9) to (13) in Table 6 show that the marginal WTP of
452	implementing a policy abroad is now relative to implementing at home for citizens not
453	considering co-benefits <sup>4</sup> . Thus, preferences for implementation at home are driven by a
454	substantial group of citizens concerned with co-benefits. When the reference is citizens not
455	concerned with co-benefits Danes and Italians tend to favor implementation abroad, while
456	German, French and Spanish residents are largely indifferent. It is also noted that the estimated
457	average WTP in the population for the reduced carbon emissions are largely unaffected in size
458	from this model extension. Similarly, after having controlled for local implementation for
459	citizens considering co-benefits, the variables for the sectors still show that investments in

<sup>&</sup>lt;sup>4</sup> The change in reference category between Table 4 and Table 6 explain the differences in the WTP of the ASC.

renewable energy is systematically preferred over the alternatives, whereas citizens in three of the countries are indifferent between policies targeting the two other sectors. France and Spain have a small but significant preference for policies targeting forest management and use of wood over policies targeting energy efficiency in industry (row (3)). The small sensitivity of these variables to the model extension shows that the covariance between co-benefit concern variables and the remaining model is centered on the country of implementation variables.

The coefficients on the interaction terms show that preference for reduced air pollution is the 466 main reason to prefer implementation at home of the 'Renewable energy' policy and the 467 'Industry efficiency' policy. Furthermore, positive or negative impacts on employment rarely 468 affects preferences, with Italians' concern with industry jobs being the only exception. For 469 Italians, concern for employment effects influence unexpectedly positively the preferences for 470 energy efficiency measures implemented locally. Finally, people in Denmark, Germany and 471 Spain concerned with local recreation effects of a policy targeting forest management will pay a 472 473 premium for implementing forest policies at home, for example to benefit from increased forest areas or similar. Concern for biodiversity effects of a forest sector targeted policy does not spill 474 over to WTP except to a small extent in France. 475

476

#### 477 Discussion

The present study analyzed the acceptance of different emission reduction policies in Europe, applying a discrete choice experiment. While this approach allows us to estimate the determinants of the WTP for alternative polices, it was not the objective to provide an estimate of the social value of CO<sub>2</sub> emission reductions.

The study design allows evaluating how citizens in different countries view climate mitigation 482 policies targeting different sectors and the tradeoffs they see herein, and this is largely new to the 483 484 literature. The finding that the citizens across countries on average share a very similar idea about how the different policies rank relative to each other is of considerable policy interest as it 485 sheds new light on the popular undercurrent reflected in the national and international policy 486 agendas on renewable energies. In this way, the study contributes significantly to a literature 487 where earlier studies have predominantly focused on carbon emission reductions from a single 488 policy measure targeting specific sectors, e.g. addressing transport choices (Brouwer, Brander, 489 and Van Beukering 2008; Achtnicht 2012), energy choices (Akter and Bennett 2011; Hanemann, 490 Labandeira, and Loureiro 2011), or more loosely defined policies (Carlsson et al. 2012). An 491 exception is Longo, Hoyos, and Markandya (2012) who also find that WTP for a policy 492 promoting renewable energy is higher than for promoting energy efficiency<sup>5</sup>. The drawback of 493 considering a larger number of alternative policy targets is that it is not possible to make 494 495 respondents relate to more specific details of how each policy might be designed for the different sectors. Therefore, respondents may vary considerably in how they imagine and believe each 496 policy to be designed. This variation will be embedded and captured in the preference 497 498 heterogeneity found across all countries in the models used in this paper. One of the sectors in our study was defined rather loosely as "Forest management and wood use". One could argue 499 500 that this could cover a wide range of different measures and strategies which may be mutually 501 exclusive; reducing harvesting, increasing the carbon stock in the forest, harvesting more to 502 stock carbon in wood products, or substituting fossil-intensive products. These different measures may have different co-benefits and the preference for this measure may therefore 503

<sup>&</sup>lt;sup>5</sup> The policies considered had different scope, i.e. the WTP per ton of reduction were highest for a policy promoting energy efficiency.

depend on how the respondents perceived the measure. To investigate if lack of detailed 504 information on the definition of the reduction measures may have influenced choices in the 505 506 choice experiment we analyzed the answers to an open-ended question which in the questionnaire was given to respondents who said they never choose the sector "Forest 507 management and wood use" in the five choice tasks. We asked them to explain why they have 508 509 not been choosing this measure. Out of 319 answers, only 65 indicated that they have not chosen this measure due to lack of knowledge or not believing in the measure. Thus, this ambiguity does 510 not seem to have caused concerns for most respondents. Nevertheless, to shed light on 511 preferences for specific forest management and wood use options, we suggest that future studies 512 should explore different forest measures, the importance of information provision and 513 respondents' perceptions. 514

515

The same limitation of using a generic sector description applies to reduction measures targeting 516 517 renewable energy sector, where different measures (e.g. windmills, solar panels, and biomass use) may have different co-benefit and the preferences will therefore depend on which measure 518 considered by the respondent. For example, in our questionnaire it was assumed that the effect of 519 520 measures in the renewable energy sector would have a positive effect on air pollution, However, this may not be the case if the renewable energy measure was burning biomass for energy. In this 521 522 sense, our phrasing of the co-benefit follow-up questions could be criticized for being 523 suggestive<sup>6</sup>. For all the co-benefits, we only included single statements of co-benefits to which 524 people could state agree or disagree on a five point Lickert scale, not reversed formulations. While this is indeed a limitation if the study was focused on analyzing these statements alone, it 525

<sup>&</sup>lt;sup>6</sup> A point raised by a reviewer

is of minor importance here, where our focus is on evaluating the correlation between the answer to this question and the WTP elicited in the prior choice experiment, which is unaffected by the formulations. Possible other side effects that respondent find relevant will be captured in the sector specific variable.

One possible consideration is that the payment vehicle is linked to electricity consumption. It may seem less clear why reductions in industry or by carbon sequestration in forests, should affect the electricity bill. However, in reality they are linked to some extent as the grid costs are shared and paid for, not only depending on consumption. Likewise, carbon compensation mechanisms cause carbon sequestration to be included as well. Therefore, and because we did not find evidence for such concerns in focus groups, we do not see it likely that this drives the results.

The majority of earlier studies of policies with positive local externalities have reported people to 537 prefer policies implemented in their home country rather than abroad (Carlsson et al. 2012; 538 539 Buntaine and Prather 2018) and to be concerned about local co-benefits (Longo, Hoyos, and Markandya 2012; Torres et al. 2015). The present results provide a much stronger basis for these 540 findings as the topics are combined in one cross-country experiment. On average, people will 541 542 prefer a policy implemented at home, and this preference has a significant size in the sense that people may be willing to trade their preferred choice of sector target for a policy that targets their 543 544 second or even third choice of sector, but is implemented in their home country. This pattern 545 appears particularly strong for Germany and Spain, but the preference for implementation at 546 home is another general preference, which is shared across all five countries. Interestingly, there is much less heterogeneity on this part of the preference set. It is only for implementation in 547 548 Denmark (which together with Italy are the on average least preferred two countries by the four

other countries to have a policy implemented in) that there is significant heterogeneity in

550 preferences. For all the other countries most heterogeneity estimates are not significant. Thus,

these preferences are more stable and generally shared not only across, but also within,

552 populations.

Importantly, results show that preferences for implementation at home may reflect concerns over 553 554 local co-benefits. When accounting for the shared commonalities in choices made by people who have expressed strong concerns for the potential co-benefits, the only parameters that change 555 significantly are those related to the preferred country of implementation. Thus, when controlled 556 for co-benefit concerns, respondents who did not consider co-benefits in their choices are much 557 less likely to prefer implementation at home. In fact, we find that not only do a lot of the 558 parameters on target countries become insignificant, but many in fact change sign and become 559 significantly positive. This is true for all countries, but in particular for citizens in Denmark and 560 Italy. This result has important implications for the choice of regulations for emission reductions. 561 562 While international and inter-sectoral transferal quotas are normally preferred as they are considered cost-efficient, sector and national dependent ancillary effects may create externalities 563 on the carbon reduction market. Therefore, it may be relevant to consider a tariff system when 564 565 trading carbon credits between different sectors or countries (Lutter and Shogren 2002). Understanding the population's preferences for carbon mitigation measures and their location 566 should also guide the development of business models for voluntary carbon credit markets 567 568 (Anderson and Bernauer 2016).

The results show that citizens of the Mediterranean countries have a higher WTP, than citizens in northern countries, which may indicate that WTP may also vary with local expected impacts. In Europe, the Mediterranean countries are among those expected to feel climate change impact the

most, with higher temperatures and in particular potentially severe problems with draught 572 periods (Giorgi and Lionello 2008a). Evidence from studies of resource owners and citizens 573 574 suggest that personal experiences with climate anomalies and extremes is a key driver in both beliefs about and actions in relation to climate change (Owen et al. 2012), and studies have also 575 found indications of a similar North-South gradient related to personal experience (Blennow and 576 577 Persson 2009; Blennow et al. 2012). The citizens samples in the present sample are likely to vary considerably in the way they look upon all these different aspects and in their experiences and 578 579 this will lead to preference heterogeneity. In the econometric model, this heterogeneity is captured explicitly and it is significant. Thus, while the mean estimates suggest that most are 580 willing to contribute towards mitigation, the standard deviations also show that some are willing 581 to contribute a lot more than the mean, whereas as others are not willing to contribute. 582 Nevertheless, the similarity of preferences and willingness to support policies suggests a strong 583 popular basis for the joint European Union policies on climate change mitigation. 584

585

In this study, we have had emphasis on co-benefits and country of implementation, but other 586 factors may matter for preferences for where to implement mitigation policies. Negative (local) 587 588 externalities (e.g. NIMBY aspects) may be an issue, but probably more at a local scale than a national one. Given the results of preference for home country location for policy 589 590 implementation, other issues important for the respondents have outweighed any concerns for 591 local negative externalities in our results. Another possible reason could be a question of trust in 592 institutions. If people have higher trust in their national governments than in other countries governments, it may drive the results. Finally, there may ethical considerations of burden sharing 593 594 (see e.g. van den Berg et al, 2020) which may also affect the results. We leave it for future

studies to investigate the interplay between these factors. While choice experiments are a useful method for assessing the determinants of policy acceptability (Ščasný et al 2017; Faure et al. 2022), carrying out complementary studies, applying, for example, qualitative interviews, could improve the understanding of the underlying perceptions and knowledge for the choices made in the choice experiment.

600

### 601 **Concluding remarks and perspectives**

This study has undertaken cross-country analysis of citizens' preferences for climate mitigation 602 policies that vary in their degree of mitigation in the form of carbon emission reductions, in what 603 sector they target, and in which of the countries included, the policies would be implemented. 604 605 Such a study requires that it is credible that policy implementations in one country can potentially affect the income of citizens in other countries. The European Union represents a 606 credible institutional framework for such a study. The EU has a joint budget to which all 607 608 countries and hence tax paying citizens contribute, and it has a joint climate mitigation policy ambition and has already implemented significant, mandatory and judicially binding largescale 609 policy instruments across and within countries, most notably the ETS. Thus, targeting citizens in 610 the five European Union member states Denmark, France, Germany, Italy and Spain, the 611 research questions set forth could be addressed. 612 Overall there is considerable support and willingness to pay (WTP) for significant climate 613

mitigation policies, though there is also considerable variation in the support and degree of
willingness to pay – and more so within each country than between them.

A substantial part of the WTP relates directly to the public good of carbon emission reductions,

617 which in all countries and models remains a stable and dominant component of citizens average

WTP. The mean WTP per unit of carbon emission reductions are surprisingly similar acrosscountries, though it tends to be higher in the Mediterranean countries.

620 Three other elements, however, are also of importance to people and their support for climate mitigation policies: namely which sector is targeted by the policy and in which country of the 621 five, the policy will be implemented. Again, the citizens in all countries share a similar pattern of 622 623 preferences. First of all, in all countries, citizens on average prefer policies targeting investment in renewable energy over policies targeting either energy efficiency in the industry sector or 624 forest management and the use of wood. Secondly, in all countries, citizens prefer policies 625 implemented at home *ceteris paribus*, and this is a preference where homogeneity is strong 626 within countries as well as between. Thirdly, it is shown that the citizens' concern for potential 627 co-benefits, notably reductions in air pollution from measures targeting renewables energy or the 628 industry sector, is the driver of preferences for implementation in the home country. Correcting 629 for this, and in the absence of co-benefits or co-benefit concerns, people are indifferent to where 630 631 a policy is implemented or may even prefer it elsewhere.

These results are important and raise three immediate points of relevance for international and 632 national policy debates. First, they suggest that it is likely easier to pursue and obtain public 633 634 support for cost-effectiveness goals across borders, when the policy instrument selected has no co-benefits of concern to people. Secondly, national policies on the other hand, might benefit 635 636 from taking into account the role of co-benefits in their design, making policies more robust 637 (Bernauer and Gampher 2015). Articulation and quantification of co-benefit aspects of climate 638 mitigation actions implemented domestically may increase popular support for policies. Together and thirdly, however, these policy implications could also spell trouble for international climate 639 640 policy negotiations as the most cost-effective policies could potentially also come with

- 641 significant co-benefits. In such cases, national governments might tend to focus these policies on
- a national implementation, which may violate an overall cost-effective implementation.

- 644645 Acknowledgements:
- 646 [excluded in review version]
- 647

#### 648 **6. References**

- Achtnicht M., 2012. German car buyers' willingness to pay to reduce CO2 emissions. Climatic
   Change, 113:679-697.
- Adaman, F., Karalı, N, Kumbaroğlu, G., Or, İ., Özkaynak, B., and Zenginobuz, Ü. 2011. What
- determines urban households' willingness to pay for CO2 emission reductions in Turkey:
   A contingent valuation survey. Energy Policy, 39(2):689–698.
- Adger W.N., Dessai S., Goulden M., Hulme M., Lorenzoni I., Nelson D.R., Naess L.O., Wolf J.,
   and Wreford A. 2009. Are there social limits to adaptation to climate change? Climatic
- 656 Change, 93:335-354.
- Akter S., Bennett J., and Ward M.B. 2012. Climate change scepticism and public support for
   mitigation: Evidence from an Australian choice experiment. Global Environmental
   Change-Human and Policy Dimensions, 22:736-745.
- Akter S. and Bennett J., 2011. Household perceptions of climate change and preferences for
   mitigation action: the case of the Carbon Pollution Reduction Scheme in Australia.
   Climatic Change, 109: 417-436.
- 002 Oninade Change, 109: 117 150.
- Alberini, A., Bigano, A., Ščasný, M., and Zvěřinová, I. 2018. Preferences for Energy Efficiency

vs. Renewables: What Is the Willingness to Pay to Reduce CO2 Emissions? Ecological
Economics, 144: 171–185.

- Aldy, J. E., Kotchen, M. J., and Leiserowitz, A. A. 2012. Willingness to Pay and Political
- 667 Support for a U.S. National Clean Energy Standard. Nature Climate Change 2(5)
- Alló, M. and Loureiro, M.L. 2014. The Role of Social Norms on Preferences towards Climate
- 669 Change Policies: A Meta-Analysis. Energy Policy 73: 563–574.

670	Anderson, B. and Bernauer, T. 2016. How much carbon offsetting and where ? Implications of
671	efficiency, effectiveness, and ethicality considerations for public opinion. Energy Policy,
672	94:387–395.
673	Anthoff D. and Tol R.S.J. 2010. On international equity weights and national decision making on
674	climate change. Journal of Environmental Economics and Management, 60:14-20.
675	Anthoff D., Hepburn C., and Tol R.S.J. 2009a. Equity weighting and the marginal damage costs
676	of climate change RID A-2105-2009 RID D-5245-2011. Ecological Economics, 68:836-
677	849.
678	Anthoff D., Tol R.S.J., and Yohe G.W. 2009b. Risk aversion, time preference, and the social cost
679	of carbon. Environmental Research Letters, 4:024002.
680	Bakhtiari, F., Jacobsen, J.B., Thorsen, B.J., Lundhede, T.H., Strange, N., and Boman, M. 2018.
681	Disentangling distance and country effects on the value of conservation across national
682	borders. Ecological Economics 147, 11-20.
683	Baranzini, A., Borzykowski, N., and Carattini, S. 2018. Carbon offsets out of the woods?
684	Acceptability of domestic vs. international reforestation programmes in the lab. Journal
685	of Forest Economics, 32:1–12.
686	Batel, S., 2020. Research on the social acceptance of renewable energy technologies: Past,
687	present and future. Energy Research and Social Science, 68: 101544.
688	Bernauer, T. and Gampfer, R. 2015. How robust is public support for unilateral climate policy?
689	Environmental Science and Policy, 54, 316–330
690	Blennow K., Persson J., Tome M. and Hanewinkel M., 2012. Climate Change: Believing and
691	Seeing Implies Adapting. Plos One, 7:e50182.

692	Blennow K. and Persson J., 2009. Climate change: Motivation for taking measure to adapt.
693	Global Environmental Change-Human and Policy Dimensions, 19:100-104.
694	Bollen, J. C., Jamet, S., and Corfee-morlot, J. 2009. Co-Benefits of Climate Change Mitigation
695	Policies: Literature review and new results. OECD, Economics Department Working
696	Paper Series No. 693.
697	Brouwer R., Brander L., and Van Beukering P. 2008. "A convenient truth": air travel passengers'
698	willingness to pay to offset their CO <sub>2</sub> emissions. Climatic Change, 90:299-313.
699	Buntaine, M. T. and Prather, L. 2018. Preferences for Domestic Action Over International
700	Transfers in Global Climate Policy. Journal of Experimental Political Science, 5(2), 73–
701	87
702	Carlsson F. and Martinsson P., 2001. Do hypothetical and actual marginal willingness to pay
703	differ in choice experiments? Application to the valuation of the environment. Journal of
704	Environmental Economics and Management, 41:179-192.
705	Carlsson F., Kataria M., Krupnick A., Lampi E., Lofgren A., Qin P., and Sterner T., 2013. A fair
706	share: Burden-sharing preferences in the United States and China. Resource and Energy
707	Economics, 35:1-17.
708	Carlsson F., Kataria M., Krupnick A., Lampi E., Lofgren A., Qin P., Chung S., and Sterner T.
709	2012. Paying for Mitigation: A Multiple Country Study. Land Economics, 88:326-340.
710	Carson R.T. and Groves T. 2007. Incentive and informational properties of preference questions.
711	Environmental & Resource Economics, 37:181-210.
712	Diederich, J. and Goeschl, T. (2017). Does Mitigation Begin At Home? Discussion Paper Series
713	634, Department of Economics, Heidelberg University, Heidelberg pp26.

714	Drews, S. and van den Bergh, J. C.J.M. 2015. What Explains Public Support for Climate
715	Policies? A Review of Empirical and Experimental Studies, Climate Policy 3062: 1–22.
716	Elbakidze, L., and McCarl, B. A. (2007). Sequestration Offsets versus Direct Emission
717	Reductions : Consideration of Environmental co-effects. Ecological Economics, 60, 564-
718	571.
719	Ek, K. 2005. Public and private attitudes towards "green" electricity: The case of Swedish wind
720	power. Energy Policy, 33(13):1677–1689.
721	Faure, C., Guetlein, M. C., Schleich, J., Tu, G., Whitmarsh, L., and Whittle, C. 2022. Household
722	acceptability of energy efficiency policies in the European Union: Policy characteristics
723	trade-offs and the role of trust in government and environmental identity. Ecological
724	Economics, 192
725	Gardiner S. 2004. Ethics and global climate change. Ethics, 114:555-600.
726	Giorgi F. and Lionello P. 2008a. Climate change projections for the Mediterranean region.
727	Global and Planetary Change, 63:90-104.
728	Giorgi F. and Lionello P. 2008b. Climate change projections for the Mediterranean region.
729	Global and Planetary Change, 63:90-104.
730	Glenk, K., Colombo, S. 2011. Designing policies to mitigate the agricultural contribution to
731	climate change : an assessment of soil based carbon sequestration and its ancillary effects.
732	Climatic Change, 43–66.
733	Grubb M., 1995. Seeking Fair Weather - Ethics and the International Debate on Climate-Change.
734	International Affairs, 71:463-496.
735	Hanemann M., Labandeira X., and Loureiro M.L. 2011. Climate change, energy and social
736	preferences on policies: exploratory evidence for Spain. Climate Research, 48:343-348.

737	Hepburn C. and Stern N., 2008. A new global deal on climate change. Oxford Review of
738	Economic Policy, 24:259-279.
739	IPCC 2001. Mitigation of Climate Change. Contribution of Working Group III to the Third
740	Assessment Report of the Intergovernmental Panel on Climate Change. Cambridge
741	University Press, Cambridge.
742	Jagers, S. C., Lachapelle, E., Martinsson, J., and Matti, S. 2021. Bridging the ideological gap?
743	How fairness perceptions mediate the effect of revenue recycling on public support for
744	carbon taxes in the United States, Canada and Germany. Review of Policy Research,
745	38(5): 529–554.
746	Kácha, O., Vintr, J., and Brick, C. 2022. Four Europes: Climate change beliefs and attitudes
747	predict behavior and policy preferences using a latent class analysis on 23 countries.
748	Journal of Environmental Psychology, 81(june).
749	Kotchen M.J., Boyle K.J., and Leiserowitz A.A. 2013. Willingness-to-pay and policy-instrument
750	choice for climate-change policy in the United States. Energy Policy, 55:617-625.
751	Kriegler, E., Weyant, J.P., Blanford, G. J., Krey, V., Clarke, L., Edmonds, J., Fawcett, A.,
752	Luderer, G., Riahi, K., Richels, R., Rose, S.K., Tavoni, M., and Vuuren, D. P. Van. 2014
753	The role of technology for achieving climate policy objectives : overview of the EMF 27
754	study on global technology and climate policy strategies. Climatic Change, 123: 353-
755	367.
756	Kreibich, N. and Hermwille, L. 2021. Caught in between: credibility and feasibility of the
757	voluntary carbon market post-2020. Climate Policy, 21(7): 939-957.

758	Krupnick, A., Burtraw, D., and Markandya, A. (2000). The Ancillary Benefits and Costs of
759	Climate Change Mitigation. In OECD (Ed.), Assessing the Ancillary Benefits and Costs
760	of Greenhouse Gas Mitigation (pp. 517–560). OECD Publishing.
761	Lancaster K.J., 1966. New Approach to Consumer Theory. Journal of Political Economy,
762	74:132-157.
763	Leiserowitz A., 2006. Climate change risk perception and policy preferences: The role of affect,
764	imagery, and values. Climatic Change, 77:45-72.
765	List J. and Gallet C. 2001. What experimental protocol influence disparities between actual and
766	hypothetical stated values? Environmental & Resource Economics, 20:241-254.
767	Longo A., Hoyos D., and Markandya A. 2012. Willingness to Pay for Ancillary Benefits of
768	Climate Change Mitigation. Environmental & Resource Economics, 51:119-140.
769	Longo, A., Markandya, A., and Petrucci, M. 2008. The internalization of externalities in the
770	production of electricity: Willingness to pay for the attributes of a policy for renewable
771	energy. Ecological Economics, 67(1):140–152.
772	Lutter, R., and Shogren, J. F. 2002. Tradable Permit Tariffs: How Local Air Pollution Affects
773	Carbon Emissions Permit Trading. Land Economics, 78(2):159–170.
774	Mariel, P., Hoyos, D., Meyerhoff, J., Czajkowski, M., Dekker, T., Glenk, K., Jacobsen, J.B.,
775	Liebe, U., Olsen, S.B., Sagebiel, J., and Thiene, M. 2021. Environmental Valuation with
776	Discrete Choice Experiments, Guidance on Design, Implementation and Data Analysis.
777	Springer Briefs in Economics, Springer International Publishing, 129 pp.
778	MacKerron G.J., Egerton C., Gaskell C., Parpia A., and Mourato S. 2009. Willingness to pay for
779	carbon offset certification and co-benefits among (high-)flying young adults in the UK.
780	Energy Policy, 37:1372-1381.

781	McFadden D., 1974. Conditional logit analysis of qualitative choice behaviour. In: P. Zarembka
782	(Editor), Frontiers in Econometrics. Academic Press, New York, pp. 105-142.
783	Murphy J., Allen P., Stevens T., and Weatherhead D., 2005. A meta-analysis of hypothetical bias
784	in stated preference valuation. Environmental & Resource Economics, 30:313-325.
785	O'Keeffe, L. 2014. A Choice Experiment Survey Analysis of Public Preferences for Renewable
786	Energy in the United States. Journal of Environmental and Resource Economics at Colby,
787	1(1).
788	Nemet, G. F., Holloway, T., and Meier, P. (2010). Implications of incorporating air-quality co-
789	benefits into climate change policymaking. Environmental Research Letters, 5(1):
790	Owen, A., Conover, L. E., Videras, J., and Wu. S. 2012. Heat Waves, Droughts, and Preferences
791	for Environmental Policy. Journal of policy analysis and management 31(3): 556–77.
792	Pearce, D. 2000. Policy frameworks for the ancillary benefits of climate change. In OECD (Ed.),
793	Assessing the Ancillary Benefits and Costs of Greenhouse Gas Mitigation (pp. 517–560).
794	OECD Publishing.
795	Pittel K. and Ruebbelke D.T.G., 2008. Climate policy and ancillary benefits: A survey and
796	integration into the modelling of international negotiations on climate change. Ecological
797	Economics, 68:210-220.
798	Reiner D., Curry T., De Figueiredo M., Herzog H., Ansolabehere S., Itaoka K., Johnsson F., and
799	Odenberger M. 2006. American exceptionalism? Similarities and differences in national
800	attitudes toward energy policy and global warming. Environmental science & technology,
801	40:2093-2098.
802	Rodríguez-Entrena, M., Espinosa-Goded, M., and Barreiro-Hurlé, J. 2014. The Role of Ancillary
803	Benefits on the Value of Agricultural Soils Carbon Sequestration Programmes: Evidence

804	from a Latent Class Approach to Andalusian Olive Groves. Ecological Economics 99:
805	63–73.

- Scarpa, R. and Willis, K. G. 2010. Willingness-to-pay for renewable energy: Primary and
   discretionary choice of British households' for micro-generation technologies. Energy
- Ščasný M., Zvěřinová, I., Czajkowski, M., Kyselá, E., and Zagórska, K. 2017. Public

Economics, 32(1), 129–136

- acceptability of climate change mitigation policies: a discrete choice experiment, Climate
  Policy 17: S11-S130.
- Smead R., Sandler R.L., Forber P., and Basl J. 2014. A bargaining game analysis of international
  climate negotiations. Nature Climate Change, 4:442-445.
- Solomon, B. D. and Johnson, N. H. 2009. Valuing climate protection through willingness to pay
  for biomass ethanol. Ecological Economics, 68(7), 2137–2144
- 816 Sovacool, B. K., Martiskainen, M., Hook, A., Baker, L. 2020. Beyond cost and carbon : The
- 817 multidimensional co-benefits of low carbon transitions in Europe. Ecological Economics,
  818 169(November 2019)
- Stiglitz, J.E. 2019. Addressing climate change through price and non-price interventions.
  European Economic Review, 119: 594-612.
- Svenningsen. L.S. and Thorsen, B. J. 2020. Preferences for distributional impacts of climate
   policy. Environmental and Resource Economics, 75: 1-24
- 823 Thiene M. and Scarpa R. 2009. Deriving and Testing Efficient Estimates of WTP Distributions
- in Destination Choice Models. Environmental & Resource Economics, 44:379-395.

825	Torres, A. B., MacMillan, D. C., Skutsch, M., and Lovett, J. C. 2015. 'Yes-in-My-Backyard':
826	Spatial Differences in the Valuation of Forest Services and Local Co-Benefits for Carbon
827	Markets in México. Ecological Economics 109: 130-41.
828	Train K. and Weeks M. 2005. Discrete choice models in preference space and willingness-to-pay
829	space. In: R. Scarpa and A. Alberini (Editors), Applications of simulation methods in
830	environmental and resource economics, Springer Publisher, Dordrecht, The Netherlands,
831	pp. 1-16.
832	Train K.E. 2003. Discrete Choice Methods with Simulation. Cambridge University Press,
833	Cambridge, Cambridge, 334 pp.
834	van den Berg, N.J., van Soest, H.L., Hof, A.F., den Elzen, M.G.J., van Vuuren, D.P., Chen, W.,
835	Drouet, L., Emmerling, J., Fujimori, S., Höhne, N., Kõberle, A.C., McCollum, D.,
836	Schaeffer, R., Shekhar, S., Vishwanathan, S.S., Vrontisi, Z., Blok, K., 2020. Implications
837	of various effort-sharing approaches for national carbon budgets and emission pathways.
838	Climatic Change 162: 1805–1822.
839	Vasconcelos V.V., Santos F.C., and Pacheco J.M. 2013. A bottom-up institutional approach to
840	cooperative governance of risky commons. Nature Climate Change, 3:797-801.
841	Wagner G. and Zeckhauser R.J. 2012. Climate policy: hard problem, soft thinking. Climatic
842	Change, 110:507-521.
843	Wei T., Yang S., Moore J.C., Shi P., Cui X., Duan Q., Xu B., Dai Y., Yuan W., Wei X., Yang Z.,
844	Wen T., Teng F., Gao Y., Chou J., Yan X., Wei Z., Guo Y., Jiang Y., Gao X., Wang K.,
845	Zheng X., Reng F., Lv S., Yu Y., Liu B., Luo Y., Li W., Ji D., Feng J., Wu Q., Cheng H.,
846	He J., Fu C., Ye D., Xu G., and Dong W. 2012. Developed and developing world

847	responsibilities for historical climate change and CO2 mitigation. Proceedings of the
848	National Academy of Sciences of the United States of America, 109:12911-12915.
849	West, J. J., Smith, S. J., Silva, R. A., Naik, V., Zhang, Y., Adelman, Z., Fry, M.M., Anenberg, S.,
850	Horowitz, L.W., and Lamarque, JF. 2014. Co-benefits of Global Greenhouse Gas
851	Mitigation for Future Air Quality and Human Health. Nature Climate Change, 3(10):
852	885–889.
853	