

The Structure of Plant–Pollinator Networks is Affected by Crop Type in a Highly Intensive Agricultural Landscape

Claire Gay, Sabrina Gaba, Vincent Bretagnolle

▶ To cite this version:

Claire Gay, Sabrina Gaba, Vincent Bretagnolle. The Structure of Plant–Pollinator Networks is Affected by Crop Type in a Highly Intensive Agricultural Landscape. SSRN Electronic Journal, 2023, 10.2139/ssrn.4378048 . hal-04071662

HAL Id: hal-04071662

https://hal.science/hal-04071662

Submitted on 19 Apr 2023

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers. L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

1 The structure of plant-pollinator networks is affected by crop

type in a highly intensive agricultural landscape

1

3

2

- 4
- 5 Claire Gay^{a,b,*}, Sabrina Gaba^{a,b,c}, Vincent Bretagnolle^{a,b,c}
- 6 a UMR 7372 Centre d'Etudes Biologiques de Chizé, CNRS & La Rochelle Université 79360
- 7 Villiers-En-Bois (France)
- 8 b USC 1339 Centre d'Etudes Biologiques de Chizé, INRAE, CNRS & La Rochelle Université 79360
- 9 Villiers-En-Bois (France)
- 10 ° LTSER « Zone Atelier Plaine & Val de Sèvre » 79360 Villiers-En-Bois (France)
- 11 *Author to whom correspondence should be addressed: claire.gay@cebc.cnrs.fr

12

- Abstract
- 14 In agricultural landscapes, bipartite networks formed by pollinators and the flowers they
- forage on, are characterised by the presence of species managed by humans, whether
- honeybees (*Apis mellifera*) or crop plants. These managed species can affect the structural
- properties of these networks because of spatial and temporal variation in the availability of
- resources, and competition for these resources; for example, crop plants such as oilseed rape
- and sunflower produce a large number of flowers during a short blooming period. Here, we
- 20 examined the structure of plant–pollinator networks in an intensive agricultural landscape, the
- 21 LTSER Zone Atelier Plaine & Val de Sèvre (West of France). We compiled a six-year
- 22 monitoring dataset of plant–pollinator interactions, sampling by sweep-nets along transects in
- 23 the main six crop types (811 fields in total). To describe networks, we computed six metrics:
- 24 connectance, nodes number, modularity, specialisation and two nestedness measures. Strong
- 25 differences between crops were observed in nestedness and specialisation: in oilseed rape,

- 26 networks were much more nested and much less specialised than in sunflower. In addition,
- 27 the link 'honeybee-crop flower' differed between the two mass-flowering crops. Honeybees
- and sunflowers appeared as specialist species in sunflower crops, interacting strongly and
- 29 almost exclusively with each other, whereas they are usually considered highly generalist
- 30 species. Indeed, sunflower pollination may be almost entirely driven by honeybees,
- 31 conversely to oilseed rape crops, where the presence of wild bees and other insects tended to
- 32 produce a more diversified network.

34

Key words

35 Bipartite networks, Mass-flowering crops, Weeds, Honeybees, Spatio-temporal heterogeneity

36

37

Highlights

- Honeybees and crop flowers drive the network structure in agricultural environment
- Main mass-flowering crops got very distinct network properties
- Specialisation of managed species is highly variable among mass-flowering crops

1. Introduction

41

Insect biodiversity is declining worldwide, and numerous local insect extinctions have already 42 been documented (Habel et al., 2019). This decline has high consequences for the ecosystem 43 services that rely upon insects, not least pollination (Frankie et al., 2009). In temperate 44 regions, 88% of plants depend on zoogamous pollination for their reproduction (Ollerton et 45 al., 2011) and 70% of crops depend to some extent on insect pollination (Klein et al., 2007), 46 which improves yields of insect-pollinated crops and ensure agricultural production, as in 47 oilseed rape (OSR; Brassica sp.) and sunflower (SF; Helianthus sp.) (Perrot et al., 2019, 48 2018). In arable farming systems, insect pollinators forage partly – and sometimes mainly – 49 on mass-flowering crops, but availability of these crops does not cover the entire season of the 50 pollinators' activity (Odoux et al., 2014). Between OSR and SF blooms (the so-called 'dearth 51 period'), the resources provided by weeds are essential for maintaining pollinators, including 52 the honeybee Apis mellifera (Carvalheiro et al., 2011; Requier et al., 2015). Weeds and other 53 54 wild plants are therefore critical for pollinators (Bretagnolle and Gaba, 2015), and contribute to their diversity and abundance at both field and landscape scales (Carvalheiro et al., 2011). 55 Among insect pollinators, bees and hoverflies contain generalist species that can interact with 56 many wild flowers (Dunn et al., 2020) or crop plants, and both improve pollination service 57 (Garibaldi et al., 2011). Hoverflies are more resilient to agricultural modifications than bees in 58 agricultural landscapes (Jauker et al., 2009); conversely, butterflies which also pollinate many 59 wild plants, have been dramatically affected by agricultural intensification (Johst et al., 2006). 60 Overall, weeds and pollinators are closely intertwined in agricultural farmland, and form 61 ecological interaction networks, defined by a set of nodes (species) interconnected by one or 62 more links (interactions). There is evidence that such networks have changed in structure over 63 the last century, due to pollinator and plant community changes, with many interactions 64 65 involving specialist species being lost (Mathiasson and Rehan, 2020), mostly as a result of

land use intensification (Ferreira et al., 2013). Indeed, the persistence of pollinators – together 66 with the services they provide – depends on stability, hence the structure of plant–pollinator 67 ecological networks. 68 However, plant–pollinator ecological networks in agricultural landscapes present unique 69 properties: first, they harbour managed and over-abundant species; second, they are highly 70 dynamic in space and time due to crop rotations. Such characteristics are expected to impact 71 the nature and structure of plant-pollinator networks. First, farmland ecological networks are 72 structured around managed honeybees in addition to wild pollinators, and managed crops in 73 addition to weeds (Geslin et al., 2017). The effect of mass-flowering crops on one hand, and 74 75 the presence of the honeybee on the other hand, can be related to 'invasive species' in arable systems (Geslin et al., 2017; Stanley and Stout, 2014), i.e. super-generalist species that 76 interact with multiple groups of species (Giannini et al., 2015), leading to asymmetry and 77 nestedness of the network (Ferreira et al., 2013). Perhaps counterintuitively, this asymmetry 78 and nestedness contributes to the stability and the resistance of networks to land use change 79 (Memmott et al., 2004). Links that emerge from these super-abundant species may, however, 80 come at the expense of others, leading to species role shifts (Albrecht et al., 2014), or even 81 loss of species in both plants or pollinators, which should affect the structure and functioning 82 83 of networks (Valido et al., 2019). For instance, OSR flowers number c. 500,000/ha (Hoyle and Cresswell, 2007), are brightly coloured and attractive, and produce both pollen and 84 nectar. OSR can thus create a distortion in mutualistic networks and has already been shown 85 to be related to the decline of specialist long-tongued bumblebees in favour of short-tongued 86 species better adapted to its flower traits (Diekötter et al., 2010). Honeybees in return may 87 decrease wild bee richness, leading to networks with fewer links (Lázaro et al., 2021). 88 However, a decrease or even a removal of supergeneralist (key) species may destabilise the 89 network (Kovács-Hostyánszki et al., 2019). Secondly, agricultural landscapes display very 90

high spatio-temporal unpredictability (due to crop rotations), and seasonal heterogeneity 91 (Rollin et al., 2016): honeybees are present from April to September (Odoux et al., 2014), 92 while OSR and SF, the main super-generalist plants, have very short flowering peaks 93 (Bretagnolle and Gaba, 2015), and wild bees show generally have a short presence within the 94 season. Crop rotation and spatial heterogeneity is therefore expected to affect the temporal 95 and spatial dynamics of plant–pollinator networks (Chakraborty et al., 2021). 96 There is currently a lack of descriptive information on plant–pollinator networks in farmland 97 habitats, especially at large spatial scale or over longer time-scales (Olesen et al., 2011). 98 Indeed, most studies investigated the crop compartment in relation to the semi-natural 99 100 elements or its plants (e.g. Magrach et al., 2018; Stanley and Stout, 2014), while few studies aimed at investigating differences among crops (but see Kovács-Hostyánszki et al., 2019) and 101 even fewer the effect of crop diversity. This is unfortunate, since crop effect was found to be 102 the most statistically important effect in a vast study of predation-prey networks in 103 agricultural habitats (Ma et al., 2019). Indeed, descriptive information is critical to 104 understanding community stability and robustness, which are linked to the number of species 105 and the number of links between them (Carpentier et al., 2021), and therefore network 106 resilience, which relies on redundancy in the number of pollinator species per plant species 107 108 (Memmott et al., 2004) and network nested structure (Bascompte et al., 2003). Network stability is improved by pollinator diversity (Bendel et al., 2019), nestedness and low 109 modularity, though possibly at the expense of persistence (Thébault and Fontaine, 2010). 110 Along environmental gradients from less- to more- impacted by human footprint, e.g. from 111 natural sites to monoculture farming, an increase in connectance and nestedness but a 112 decrease in the number of nodes as well as modularity were shown (Morrison et al., 2020). 113 Increased connectance is expected to decrease secondary extinctions (Blüthgen et al., 2007) 114

and improves robustness at the community level as well as stability (Thébault and Fontaine, 2010).

In this study, we used a descriptive approach to qualify plant–pollinator networks in a highly intensive farming system, at a large spatial scale (study site of 450km²) with data collected on six different crop types (811 fields in total). Because our sampling effort was high, we first paid attention to two generally overlooked patterns, the effect of sample size and the degree distribution. Sampling effort has been shown to affect many network metrics, through sampling intensity and sampling bias, but also underlying species abundances (Blüthgen et al., 2007). Degree distribution is highly skewed, resulting in part from many 'forbidden links' (Jordano, 2016). The frequency distribution of links is claimed to be best described by power-law distributions (Blüthgen, 2010), which have been shown to fit most natural networks. We also describe the overall visual structure of networks by seasons and crop types. Then, we investigated major features of the network, from simple network descriptors (number of nodes, connectance) to more complex ones (nestedness, modularity, specialisation), quantifying their level of variation among crop types and seasons, while accounting for

sample size. We expected plant–pollinator network structure to vary seasonally because of the

presence of mass-flowering crops, and the spillover or dilution effects that result from such

presence. Indeed, we expected strong variations in network structure among crop types, in

particular between mass-flowering and other crops, but also between arable crops and

grasslands. We further predicted a clear dominance of the managed honeybee, especially

during the flower blooms of OSR and SF, which dominate landscapes when flowering. In

particular, we predicted that networks would be more specialised during OSR and SF blooms

than in other periods and other crops. We also predicted that networks would be more nested

during the blooming periods of mass-flowering crops than in the dearth periods, and thus in

these crops – as a result of super-generalist and abundant species presence – more able to include the same partners as the specialists in their spectrum of interactions.

141

142

143

144

145

146

147

148

149

150

151

152

153

154

155

156

157

158

159

160

140

139

2. Materials and methods

2.1. Study site

Our study site is the Long Term Socio-Ecological Research (LTSER) site 'Zone Atelier Plaine & Val de Sèvre', located in the Nouvelle-Aquitaine region of southwest France, south of the city of Niort (Bretagnolle et al., 2018). It covers 450 km², harbouring c. 13,000 fields. We focused on six years of survey, 2015–2020, which is one of the largest time intervals of plant-pollinator network studies (but see Chacoff et al., 2018; Petanidou et al., 2008; Ponisio et al., 2017). Each year, 100 to 220 different fields were surveyed (some fields were studied for more than one year) with a standardised protocol. Fields were selected in a stratified random process (see Bretagnolle et al. (2018) for details). From 2015–2020, the number of fields sampled per respective year were 174, 217, 195, 160, 117 and 115 (see ESM1a and 1b in Appendix A). Crop surveyed included 7% corn (Zea mays) and sorghum (Sorghum bicolor), 8% alfalfa (Medicago sativa), 12% meadows (temporary or permanent), 18% oilseed rape (Brassica napus), 22% sunflower (Helianthus annuus), and 33% cereal crops. Cereals included barley (Hordeum vulgare), durum wheat (Triticum turgidum), einkorn (Triticum monococcum), oat (Avena sativa), rye (Secale cereale), triticale (×Triticosecale) and winter wheat (Triticum aestivum). Sampling was performed from April to August, i.e. all crops were surveyed at least once during their respective flowering periods: OSR from April–May; sunflower from July-August; alfalfa, meadows and cereals from May-June to the end of August; and corn and sorghum from July-August.

162

163

161

2.2. Sampling of plant-pollinator interactions

Insects surveyed included bees (both honeybees and wild bees), butterflies and hoverflies (Syrphidae), i.e. the three main pollinator groups in the LTSER. This study site hosts a large number of pollinator species (Rollin et al., 2016). Each year, two (2015) or three (2016–2020) 50 m long transects were performed per field: one in the field centre, one in the field margin (section between the field edge and the first seeding line) and the last in between, at c. 25 m from the margin (except in 2015). Field margins generally harbour a more diverse fauna compared to field cores. Transects were parallel to tram lines. For each transect, the surveyed area was a theoretical volume, 2.5 m each side, 2.5 m in front of the observer and c. 1 m above the observer's head. Surveys were conducted under calm weather conditions, i.e. sunny days with air temperatures in full sun mostly above 15°C (95% T >16°C; 12° to 45°C, in sunlight), and between 7:30 a.m. and 6:30 p.m. In 27.9% of the fields (226 of 811), two sampling sessions were conducted (early- and late-season). Each plant-pollinator interaction (i.e. an insect landing on a flower) was tallied, with both plant and pollinator being identified (pollinators being caught with a sweep-net for later identification by a professional entomologist if necessary). Transects were walked along very slowly (c. 12 s/m), for a total duration of 10 min each. The observer stopped the timer each time an insect had to be identified or captured. Butterflies were always identified by sight; however, given the time spent in flight in butterflies, all individuals crossing the sampled theoretical volume were tallied, even if not on a flower. Almost all pollinators and plants (94% and 96% respectively) were identified to species level.

184

185

186

187

188

164

165

166

167

168

169

170

171

172

173

174

175

176

177

178

179

180

181

182

183

2.3. Statistical analyses

2.3.1. Sampling completeness and frequency distribution of degrees

We estimated sampling completeness of plants and pollinators using the Chao estimator of asymptotic species richness (Chacoff et al., 2012), which is based on rare species (those that

occur only once or twice). We calculated the Chao 2 estimator using the *vegan* R package (v2.5.7; Oksanen et al., 2020) and evaluated the percentage of completion (ratio between the observed value and estimated value) of our sampling of plants and pollinators.

We analysed the frequency distribution of degrees in total as well as in subnetworks (per crop and per period), testing for best-fit with *a priori* distributions, i.e. Exponential, Poisson, Weibull, Lognormal, Negative binomial, Power law and Fisher's logseries. We used the *fitdistrplus* package (Delignette-Muller and Dutang, 2015), except for the Fisher's logseries (which is not implemented in this package), which we computed. Power law was tested with the *poweRlaw* package (Gillespie, 2015). Best fit was checked using the Akaike Information Criterion (AIC; Burnham et al., 2011). We fixed a threshold of 2 units' difference in AIC to select the best model among the candidate models.

2.3.2. Number of nodes and number of interactions: crop and temporal effects

From floral visitation events, we constructed bipartite networks with plant and bee nodes, using a visitation event as a plant–pollinator interaction. To analyse how the network structure varied among crops and seasons, we built interaction subnetworks. Six crop categories were retained: OSR, cereals, alfalfa, meadows, SF and corn–sorghum. We also split the data into three equal-length seasons, each 45 days long: April 1st–May 20th (OSR flowering peak, characterised by high resource availability for insects), May 21st–July 9th, characterised by a general lack of flowering crop resources available (i.e. the dearth period) – when honeybee colony sizes are, perhaps surprisingly, at their highest (Odoux et al., 2014) – and finally, July 10th–August 22nd, corresponding to the SF flowering peak. We also divided our dataset into six main taxonomic groups: honeybees, bumblebees, other social bees (including some Halictidae, e.g. Halictus rubicundus, Lasioglossum pauxillum), solitary bees, hoverflies and butterflies. To describe seasonal patterns in the number of interactions, we standardised our

data, since the number of transects per field and the number of fields sampled per crop and per year varied. We thus converted the number of interactions per transect per field into a number per hectare (ha), since each transect surveyed an area of 250 m². Next, we grouped samples per fortnight by averaging values (with standard-error, se) of interactions per ha per fortnight for each crop. Obtained values were compared with ANOVA and post-hoc Tukey tests. Data were log-transformed.

220

221

222

223

224

225

226

227

228

229

230

231

232

233

234

235

236

237

238

214

215

216

217

218

219

2.3.3. Network structure descriptors

A large number of indices is available to describe network structure (see ESM2, which presents their respective properties in Appendix A). We selected the following ones: number of nodes (i.e. the number of species in interaction), connectance (the ratio between the number of links observed and the total number of possible links), modularity – which gives insights into the presence of subnetworks among the network (called modules, i.e. weakly interlinked subsets of species that are constituted of strongly interacting species) and which was calculated using the greedy Louvain algorithm method (Blondel et al., 2008) – and nestedness, described by NODF (Nestedness metrics based on Overlap and Decreasing Fill; Almeida-Neto et al., 2008). The latter is a property describing to what extent the partners of specialists are a subset of the partners of generalists. We also used WNODF (Weighted Nestedness metrics based on Overlap and Decreasing Fill; Almeida-Neto & Ulrich, 2011), the quantitative alternative to NODF that accounts for the number of similar interactions. Finally, we focused on the specialisation H₂' (standardised two-dimensional Shannon entropy) – the degree of specialisation between the two levels of the entire network (Blüthgen et al., 2006) – and the d'species-level specialisation measure (standardised Kullback-Leibler distance, Blüthgen et al., 2006), corrected by de Manincor et al. (2020), representing the degree of interaction specialisation at the species level. Network descriptors were obtained using the R

packages bipartite (Dormann et al., 2009), igraph (Csardi and Nepusz, 2006), NetIndices (Kones et al., 2009) and vegan (Oksanen et al., 2020). Some of these metrics (i.e. connectance, nestedness) are sensitive to sample or network size (Dormann et al., 2009). We thus checked how indices varied with increasing numbers of samples (according to periods and crops) by calculating all metrics (excluding d') with increasing numbers of sampled fields. Five to 195 fields (every 10 field steps) were randomly sampled (with replacement) from the overall sample. At each step (i.e. field number), 100 groups of fields were constituted, the six network metrics were computed and the mean and the 95% confidence interval (CI) were calculated. The maximum bootstrap threshold to compare the three seasons was 175 field groups, when OSR period samples stopped, but we also went up to 195 fields to compare only dearth period and sunflower period. When comparing the six different crop types, we had to stop at a threshold of 35 fields per group (100 groups) in the bootstrap because we had very few fields of corn-sorghum and alfalfa. To increase the sampling effort, we had to lose crops, and were left with only four crops with 100 groups of 85 fields. However, when focusing on the two mass-flowering crops, OSR and SF, we were able to analyse 100 groups of 165 fields. Therefore, the number of fields bootstrapped varied according to the crops to be compared.

256

257

258

259

260

261

239

240

241

242

243

244

245

246

247

248

249

250

251

252

253

254

255

2.3.4. Field selection

To calculate Chao 2 estimates and the number of interactions per hectare, we included the fields in which no insects had been observed (138 fields, 17.0%). However, to represent interaction networks and calculate metrics on aggregated fields, we only retained the fields on which at least one interaction was observed in the dataset.

All analyses were run with *R software v.4.1.0* (R Core Team, 2021).

3. Results

264

265

266

267

268

269

270

271

272

273

274

275

276

277

278

279

280

281

282

283

284

3	1	Sampling	completeness	and frac	manev distri	hution of	dograss
J.	1.	Samping	compictioness	and nec	luciicy aistri	սաստո տւ	ucgices

A total of 37,133 plant–pollinator interactions were tallied over 811 fields (127 of which were studied in at least two years between 2015 and 2020), occurring between 200 species of pollinators foraging on 165 species of plants (including individuals that were only identified to genus level). We found 96 species of solitary bees, 35 species of social bees (including the honeybee A. mellifera and 14 species of bumblebees), 38 species of butterflies and 31 species of hoverflies. A total of 758 interactions were obtained between one plant or one pollinator occurring only once. No interaction was found between one plant and one pollinator occurring both only once. In regard to pollinators, over the full network, 80 species (40.0%) interacted with only one plant (crop or weed), 29 species (14.5%) interacted with two, while 91 species (45.5%) were more generalist (>2 plants, up to 92 plants). Similar values were obtained for plants, with 38 weed species interacting with only one insect species (23%), and 16 with two insect species (9.7%). There were 138 fields without any observed interaction (17.0%). We caught 64.5% of pollinators, and recorded 78.6% of the plant community, based on Chao 2 estimators. Overall, the frequency distribution of degrees (i.e. interacting plants/pollinators) fitted better to the Fisher's log-series (see ESM3 in Appendix A) than to Lognormal ($\Delta_{AIC} = 73$) or Power laws ($\Delta_{AIC} = 77$). When splitting by crop or per season, similar results were found. Differences in the goodness-of-fit of Fisher's log-series and Lognormal were, however, much lower in alfalfa ($\Delta_{AIC} = 12$) and in corn and sorghum crops ($\Delta_{AIC} = 3$) than in other crops.

285

286

287

288

3.2. Number of nodes and number of interactions: crop and temporal effects

The total number of nodes varied highly among crops (165, 114, 186, 178, 161 and 75 in meadows, alfalfa, cereals, SF, OSR and corn–sorghum, respectively). Accumulation curves

per crop indicated that these differences among crops (Fig. 1) could be detected even from 289 290 low sample sizes (c. n = 50). Thus, independently of the number of sampled fields, meadows showed the highest values of node numbers, followed by alfalfa, cereals, OSR and SF (Fig. 291 1). Year did not affect the pattern of the accumulation curves (see ESM4 in Appendix A), 292 although 2017 was a species-rich year regardless of the sampling effort. 293 The number of plant–pollinator interactions, here standardised per hectare of sampling effort 294 (which can be considered as an instantaneous visitation rate per hectare), was on average 295 296 338.43±23.19 interactions/ha (mean±se). This value, however, showed strong seasonal variations, with most interactions occurring in early July when considering all crops, or mid-297 298 June when excluding SF (Fig. 2). There were also strong variations among crops, with interactions in SF being overwhelmingly dominant and driving seasonal variation. In early 299 July, we found significant differences in the average number of interactions among crops 300 $(F_{5.62} = 20.16, p\text{-value} < 0.0001)$, with significantly higher values for SF than in alfalfa (T_{HSD} 301 p-adj < 0.001), meadows (T_{HSD} p-adj < 0.001), cereals (T_{HSD} p-adj < 0.001) and OSR (T_{HSD} p-302 adj = 0.0015). 303 Network structure differed strongly among crops in relation to variation of the total number of 304 nodes, either in the repartition of interactions between the main groups of plants or between 305 306 the main groups of pollinators (Fig. 3). In flowering crops, the crop flower was generally the most visited flower in the crop considered. For instance, 95.27% of pollinators visited SF 307 plant in SF fields, and 86.07% OSR plants in OSR fields. However, this was true only for 308 309 mass-flowering crops: alfalfa represented approximately a third of all visited flowers in alfalfa fields (36.12%) and corn was poorly visited in corn–sorghum fields (7.55%). In those crops, 310 pollinators preferentially foraged on wild flowers (89.62%). Moreover, honeybees' 311 dominance was not of the same magnitude in all crops: they were overrepresented in SF 312 (91.95%) and OSR (76.60%), while hoverflies rather than honeybees were the most-313

314	interacting group in other crops (32.55% in corn and sorghum crops, 56.42% in cereal crops,
315	36.02% in alfalfa crops and 53.38% in meadows). Interestingly, when found in other crops,
316	OSR and SF flowers showed the same interaction pattern, i.e. interacting mainly with
317	honeybees. Moreover, hoverflies foraged more on wild flowers than did any other insect
318	group, potentially explaining their high presence in weed-rich crops such as meadows or
319	alfalfa. For instance, whereas the total percentage of weed visits by all pollinator groups was
320	63.78% in alfalfa crops, it reached 77.94% when considering only hoverfly visits.
321	The overall pattern of the network is depicted on Fig. 4, which represents standardised
322	interactions between pollinators and plants according to their degree (i.e. number of
323	interaction partners). Species with the highest degrees were not necessarily the most frequent,
324	although species' frequency of occurrence and degree were strongly related (i.e. decreasing
325	pattern of species degrees were concomitant with decreasing species frequencies) (Fig. 4). In
326	terms of the number of interactions, honeybees and bumblebees were most prevalent (83.72%
327	of the total interactions number, $n = 31,089$), followed by hoverflies (8.21%, $n = 3,050$),
328	solitary bees (2.69%, $n = 998$) and butterflies (1.36%, $n = 506$). Honeybees were by far the
329	dominant insect species (80.80% of all interactions, $n = 30,007$), the next two being
330	Sphaerophoria scripta (3.80%, $n = 1,412$) and Episyrphus balteatus (2.19%, $n = 812$), two
331	hoverflies. However, although S. scripta was found 20 fewer times than honeybees in total, its
332	number of plant partners (92) was higher than that of honeybees (75). Butterflies were far less
333	often observed, with only 81 interactions for the most common butterfly, <i>Pieris rapae</i> .
334	Concerning plants, SF represented 72.36% (n = $26,871$) of the whole sample, seven times
335	more than OSR (10.69%, $n = 3,971$), and 60 times more than the most-interacting weed,
336	Torilis arvensis (1.20%, $n = 446$). Therefore, the complete network was dominated by
337	honeybees and mass-flowering crops: honeybees interacting with OSR or SF represented
338	77.44% of all interactions.

341

342

343

344

345

346

347

348

349

350

351

352

353

354

355

356

357

358

359

360

361

362

363

3.3. Network structure

3.3.1. Seasonal effect on network structure

The average number of nodes during the dearth period (mid-May to early July) was similar to the average number of nodes for all periods, regardless of the sampling effort (Fig. 5a), reaching 214 (CI_{95%}: 195–230) species when considering a sampling effort of 175 fields. The average number of species was nearly equal during SF and OSR flowering blooms (154 (142– 165) and 148 (146–151), respectively, in 175 fields); that is to say, 1.4-times lower than during the dearth period, but with a weaker variability among fields during OSR flowering. Mean connectance did not vary a lot with period, being 3% (2.6–3.4) during bloom periods and 2.5% (2.2-2.8) during the dearth period in 175 fields. Mean modularity was slightly higher for the OSR blooming network (0.42 (0.41–0.43)) than for other periods (0.39 (0.36– 0.41) in SF blooming and 0.37 (0.35–0.38) in dearth period)). Specialisation of the network remained constant during dearth and SF periods when considering network sizes above 100 fields (see ESM2 for details in Appendix A), but decreased much more quickly in the OSR blooming period, with no overlapping 95% in CIs. This highlights that OSR blooming period was represented by more generalist species than the other periods. Specialisation was higher during the SF bloom and dearth periods, with average values between 0.49 (0.43–0.55) and 0.50 (0.46–0.52), respectively, compared to the OSR flowering bloom (0.30 (0.29–0.33)). The OSR blooming period was characterised by a higher nestedness (37.2 (35.6-38.6) in 175 fields) than the SF blooming period (24.0 (21.3–26.9)) and dearth period (27.0 (24.4–29.9)). Similar patterns were observed when using weighted NODF: 14.8 (12.7–16.8) during the dearth period, 13.3 (11.8–15.0) during the SF blooming period, and 21.9 (20.6–23.1) during the OSR blooming period. Thus, it seems that the flowering period of the OSR was the most prone to host specialist species that had partners similar to generalist species.

366

367

368

369

370

371

372

373

374

375

376

377

378

379

380

381

382

383

384

385

386

387

388

3.3.2. Crop effect on network structure

All crops' subnetworks were similarly poorly connected (c. 3% at a threshold of 85 fields) (Fig. 5b). Although OSR blooming period was the most modular period (see above), mean modularity was almost identical whatever the crop considered, at about 0.5. At an equal sampling survey effort of 85 fields, modularity was 0.43 (CI_{95%}: 0.39–0.47) for meadows, 0.45 (0.40-0.49) for SF and 0.48 (0.45-0.52) for OSR. With a higher sampling rate (n = 165) fields), but at the cost of losing crop types in the comparisons, OSR showed significant higher modularity values than SF. The two mass-flowering crops' networks also had very different specialisation (H2'), showing the two most extreme values among our six crops. OSR was far less specialised (0.31 (0.29–0.33) at 165 fields) than all other crops, while SF (0.58 (0.55– 0.61)) showed the most specialised network; the other crops showed intermediate values, e.g. 0.45 (0.43–0.47) for meadows and 0.40 (0.34–0.46) for cereals at a survey effort of 85 fields. Low specialisation in OSR could be partially explained by the low specialisation of the most abundant insect, the honeybee (d' = 0.27) and the low specialisation of the crop flower, OSR (d' = 0.21). Indeed, in OSR, the average d' of all species was 0.26 for insects (n = 97, median)= 0.20) and 0.23 for plants (n = 68, median = 0.44), meaning that the network was lowly specialised at both trophic levels and highly driven by the honeybee and OSR flower dominance. In SF, specialisation was 0.67 for SF (more than three times higher than in OSR) and 0.42 for the honeybee (1.5-times higher than in OSR). SF was thus highly connected to honeybees despite honeybees being slightly less specialised than SF, indicating a larger partners' spectrum. The average d' of all insect species in SF was 0.41 (n = 109, median = (0.50) and (0.67) for plants (n = 65, median = 0.51), again reflecting the preponderant roles of honeybees and SF. Nestedness (NODF) increased with sampling effort, being higher in OSR than in any other crops (mean value of 29.8 (22.5–36.8) when sampling effort was low, i.e. 35 fields), especially in meadows (16.1 (11.0–21.3)), cereals (21.3 (15.1–28.4)) and SF (21.6 (16.8–29.2)), rather than in alfalfa (25.5 (21.0–30.8)). When increasing sampling effort (85-field threshold), nestedness reached 33.0 (27.7–38.2) in OSR networks, c. 25% more than in cereals (23.9 (19.9–28.5)) and in SF (23.4 (19.5–28.1)). Weighted NODF average values showed similar patterns: SF reached a WNODF mean value of 14.5 (13.1–15.7) when considering a threshold of 165 fields, whereas OSR networks peaked on average at 21.6 (20.3–22.5).

396

397

398

399

400

401

402

403

404

405

406

407

408

409

410

411

412

413

389

390

391

392

393

394

395

4. Discussion

The detection of several hundreds of species in the plant–pollinator network indicates that the LTSER ZAPVS is characterised by a high diversity of species despite its intensive farming system. While the honeybee, SF and OSR are by far the species that establish most interactions, they are not necessarily those with the greatest number of interacting partners. Although OSR flower is the species with the highest degree (number of interaction partners), confirming its structuring role in the network (Diekötter et al., 2010; Stanley and Stout, 2014), it is closely followed by wild plant species, generally hosted in less-intensively managed but more diverse crops such as alfalfa or meadows. Moreover, the hoverfly Sphaerophoria scripta showed far more partners than honeybee, including OSR and SF flowers, suggesting that its role in the network has to be explored in more detail. *Bombus* species are also frequent visitors of SF, as Syrphidae of OSR (Jauker and Wolters, 2008). Although these are rarely studied as key species in networks, they might be strongly involved in OSR and SF pollination (Breeze et al., 2011), and could act as pollination insurance if honeybees were to decline dramatically (Sánchez-Bayo and Wyckhuys, 2019). Completeness of survey effort was, however, moderate: 65% for insects but 80% for plants according to the Chao 2 estimator, despite our dataset covering six months (from April to

414	September) and six years. The phenology of some plants and insects being very short may
415	explain why not all species were detected, in addition to the classic and inherently imperfect
416	detection method. In addition, the low connectance within crops (around 2-3%) suggests that
417	few interactions are carried out among all possible ones. These forbidden links (Jordano,
418	2016) could result from spatial or temporal constraints (Valdovinos, 2019). Indeed, one of our
419	possible explanations is that each crop network corresponds to a set of different fields that
420	belongs to different landscape entities and different sampling weeks, which generates a high
421	community variability in terms of phenology and location. But a low connectance, i.e. high
422	proportion of forbidden links, is in certain cases one of the indicators of destabilised networks
423	(Landi et al., 2018).
424	The hypothesis of the dominance of honeybee interactions with crop flowers during OSR
425	flowering and SF flowering, and not in other periods, was largely verified. Overdominance of
426	managed species may destabilise networks (Valido et al., 2019). The stability of a network
427	can be evaluated through the values of some network metrics (Thébault and Fontaine, 2010).
428	A stable bipartite mutualistic network is often a network with a high connectance and many
429	species, as well as a nested – but not very modular – network (Thébault and Fontaine, 2010).
430	In our case, connectance was always very low (<5%), whether we considered the total
431	network or the subnetworks per season or per crop. When the connectance is so low,
432	modularity and nestedness are usually positively correlated (Fortuna et al., 2010), hence a
433	situation with a nested but not modular network is unlikely, therefore possibly suggesting
434	instability in our networks. In particular, the corn-sorghum network, which is highly modular
435	and comprises few species, may show particularly low stability, though we had relatively
436	small sample sizes. Meadows were a special case, endowed with a large number of species
437	but not very nested and not very connected compared to other crops. Thus, although meadow
438	networks appear to be resilient networks thanks to the number of species they host (Okuyama

and Holland, 2008), they display a certain vulnerability because the sharing of partners is weak, like in the studied SF crops, but contrary to the OSR crops. Indeed, the most striking differences in networks were found between OSR and SF, both being mass-flowering crops. As expected, these two crops drove the overall behaviour of networks during their blooming season (approximately the same metrics of behaviour between the OSR crop and OSR period, and between the SF crop and SF period), although the meadows, weeds in cereals, and alfalfa were flowering at the same time. However, the OSR and SF crops' networks were somewhat opposed in some of the key metrics, hence possibly showing different network stability (Duchenne et al., 2022). The OSR network did not harbour many species, but was highly nested compared to the other crop networks, which therefore suggests it was a resilient network. Its nestedness may result from the fact that the OSR network was mostly composed of species with a high degree of generalisation (i.e. the average species specialisation was quite low in OSR crops), leading to a relatively nested structure (Fontaine et al., 2009). Thus, as the OSR network was the most nested network in our study, it was expected that the general core of species it included were generalists rather than specialists, which was indeed the case. The interaction between OSR and honeybees has a structuring effect for the network, corroborating studies that highlight OSR flower and its partners as a hub that positively influences the visitation and pollination of neighbouring wild plants, especially those with similar morphology (Thompson et al., 2021). OSR is also known to host diverse insect communities (Stanley et al., 2013), while honeybees remain one of the most efficient pollinators and, as demonstrated here, the most important partner of OSR. Conversely, SF appears less stable than OSR because it is much less nested for an equivalent diversity of species, and is composed of more specialised species, including the honeybee-SF duo, which both qualify as specialist species, interacting strongly and almost entirely with each other – a pattern that was not necessarily found in other studies (e.g. Valido et al., 2019).

439

440

441

442

443

444

445

446

447

448

449

450

451

452

453

454

455

456

457

458

459

460

461

462

464	Such a quasi-exclusive relationship between honeybees and SF as we found here may reflect
465	either competition within pollinators or plant communities in SF crops, or a strong niche
466	differentiation between wild and managed species. Conversely, OSR may host more
467	generalist species that interact with specialist species and therefore enhance the persistence of
468	the network (Liao et al., 2022).
469	
470	5. Conclusion
471	Our results suggest that the honeybee and crop flower pair may differ between OSR and SF,
472	being an almost exclusive relationship in SF that prevents other partners by competition,
473	while allowing other modules in OSR hence increasing nestedness and consequently
474	improving network stability. This divergence may possibly result from phenological
475	difference, or as a consequence of flower shape – a divergence that can be further explored
476	with species trait approaches.
477	
478	Author Contributions
479	Conceptualisation, V.B. and S.G.; Methodology, C.G., V.B. and S.G.; Formal Analysis, C.G.;
480	Writing-Original Draft Preparation, C.G., S.G. and V.B; Writing-Review and Editing, C.G.,
481	S.G. and V.B. All authors have read and agreed to the published version of the manuscript.
482	
483	Declaration of Competing Interest
484	The authors declare that they have no known competing financial interests or personal
485	relationships that could have appeared to influence the work reported in this paper.
486	

Acknowledgements

We thank in particular Jean-Luc Gautier, Alexis Saintilan and Marilyn Roncoroni for the
monitoring, sampling and identification of the pollinator' data. We also thank the farmers of
the LTSER 'Zone Atelier Plaine & Val de Sèvre' involved in the research program. CG
received funding for her PhD grant from the project SHOWCASE (SHOWCASing synergies
between agriculture, biodiversity and ecosystem services to help farmers capitalising on
native biodiversity) within the European Union's Horizon 2020 Research and Innovation
Programme under grant agreement N° 862480. This study was supported by the 'Programme
Pollinisateurs' (2015–2020) funded by the Ministère de l'Ecologie et de l'Environnement.

Appendix A. Supporting information

Supplementary data associated with this article can be found in the online version at [...]

References

501	Albrecht, M., Padrón, B., Bartomeus, I., Traveset, A., 2014. Consequences of plant invasions on
502	compartmentalization and species' roles in plant-pollinator networks. Proc. R. Soc. B. 281,
503	20140773. https://doi.org/10.1098/rspb.2014.0773
504	Almeida-Neto, M., Guimarães, P., Guimarães, P.R., Loyola, R.D., Ulrich, W., 2008. A consistent
505	metric for nestedness analysis in ecological systems: reconciling concept and measurement.
506	Oikos 117, 1227–1239. https://doi.org/10.1111/j.0030-1299.2008.16644.x
507	Almeida-Neto, M., Ulrich, W., 2011. A straightforward computational approach for measuring
508	nestedness using quantitative matrices. Environmental Modelling & Software 26, 173–178.
509	https://doi.org/10.1016/j.envsoft.2010.08.003
510	Bascompte, J., Jordano, P., Melián, C.J., Olesen, J.M., 2003. The Nested Assembly of Plant-Animal
511	Mutualistic Networks. Proceedings of the National Academy of Sciences of the United State
512	of America 100, 9383–9387.

513	Bendel, C.R., Kral-O'Brien, K.C., Hovick, T.J., Limb, R.F., Harmon, J.P., 2019. Plant–pollinator
514	networks in grassland working landscapes reveal seasonal shifts in network structure and
515	composition. Ecosphere 10. https://doi.org/10.1002/ecs2.2569
516	Blondel, V.D., Guillaume, JL., Lambiotte, R., Lefebvre, E., 2008. Fast unfolding of communities in
517	large networks. Journal of Statistical Mechanics: Theory and Experiment 2008, P10008.
518	https://doi.org/10.1088/1742-5468/2008/10/P10008
519	Blüthgen, N., 2010. Why network analysis is often disconnected from community ecology: A critique
520	and an ecologist's guide. Basic and Applied Ecology 11, 185–195.
521	https://doi.org/10.1016/j.baae.2010.01.001
522	Blüthgen, Nico, Menzel, F., Blüthgen, Nils, 2006. Measuring specialization in species interaction
523	networks. BMC Ecology 13.
524	Blüthgen, Nico, Menzel, F., Hovestadt, T., Fiala, B., Blüthgen, Nils, 2007. Specialization, Constraints,
525	and Conflicting Interests in Mutualistic Networks. Current Biology 17, 341–346.
526	https://doi.org/10.1016/j.cub.2006.12.039
527	Breeze, T.D., Bailey, A.P., Balcombe, K.G., Potts, S.G., 2011. Pollination services in the UK: How
528	important are honeybees? Agriculture, Ecosystems & Environment 142, 137-143.
529	https://doi.org/10.1016/j.agee.2011.03.020
530	Bretagnolle, V., Berthet, E., Gross, N., Gauffre, B., Plumejeaud, C., Houte, S., Badenhausser, I.,
531	Monceau, K., Allier, F., Monestiez, P., Gaba, S., 2018. Towards sustainable and
532	multifunctional agriculture in farmland landscapes: Lessons from the integrative approach of a
533	French LTSER platform. Science of The Total Environment 627, 822-834.
534	https://doi.org/10.1016/j.scitotenv.2018.01.142
535	Bretagnolle, V., Gaba, S., 2015. Weeds for bees? A review. Agron. Sustain. Dev. 35, 891–909.
536	https://doi.org/10.1007/s13593-015-0302-5
537	Burnham, K.P., Anderson, D.R., Huyvaert, K.P., 2011. AIC model selection and multimodel inference
538	in behavioral ecology: some background, observations, and comparisons. Behav Ecol
539	Sociobiol 65, 23–35. https://doi.org/10.1007/s00265-010-1029-6

540	Carpentier, C., Barabás, G., Spaak, J.W., De Laender, F., 2021. Reinterpreting the relationship
541	between number of species and number of links connects community structure and stability.
542	Nature Ecology & Evolution 5, 1102–1109. https://doi.org/10.1038/s41559-021-01468-2
543	Carvalheiro, L.G., Veldtman, R., Shenkute, A.G., Tesfay, G.B., Pirk, C.W.W., Donaldson, J.S.,
544	Nicolson, S.W., 2011. Natural and within-farmland biodiversity enhances crop productivity:
545	Weeds maximize nature benefits to crops. Ecology Letters 14, 251–259.
546	https://doi.org/10.1111/j.1461-0248.2010.01579.x
547	Chacoff, N.P., Resasco, J., Vázquez, D.P., 2018. Interaction frequency, network position, and the
548	temporal persistence of interactions in a plant-pollinator network. Ecology 99, 21-28.
549	https://doi.org/10.1002/ecy.2063
550	Chacoff, N.P., Vázquez, D.P., Lomáscolo, S.B., Stevani, E.L., Dorado, J., Padrón, B., 2012.
551	Evaluating sampling completeness in a desert plant-pollinator network: Sampling a plant-
552	pollinator network. Journal of Animal Ecology 81, 190-200. https://doi.org/10.1111/j.1365-
553	2656.2011.01883.x
554	Chakraborty, P., Chatterjee, S., Smith, B.M., Basu, P., 2021. Seasonal dynamics of plant pollinator
555	networks in agricultural landscapes: how important is connector species identity in the
556	network? Oecologia 196, 825-837. https://doi.org/10.1007/s00442-021-04975-y
557	Csardi, G., Nepusz, T., 2006. The Igraph Software Package for Complex Network Research.
558	InterJournal. Complex Systems 1695.
559	de Manincor, N., Hautekèete, N., Mazoyer, C., Moreau, P., Piquot, Y., Schatz, B., Schmitt, E.,
560	Zélazny, M., Massol, F., 2020. How biased is our perception of plant-pollinator networks? A
561	comparison of visit- and pollen-based representations of the same networks. Acta Oecologica
562	105, 103551. https://doi.org/10.1016/j.actao.2020.103551
563	Delignette-Muller, M.L., Dutang, C., 2015. fitdistrplus: An R Package for Fitting Distributions.
564	Journal of Statistical Software 64. https://doi.org/10.18637/jss.v064.i04
565	Diekötter, T., Kadoya, T., Peter, F., Wolters, V., Jauker, F., 2010. Oilseed rape crops distort plant-
566	pollinator interactions. Journal of Applied Ecology 47, 209–214.
567	https://doi.org/10.1111/j.1365-2664.2009.01759.x

568	Dormann, C.F., Frund, J., Bluthgen, N., Gruber, B., 2009. Indices, Graphs and Null Models:
569	Analyzing Bipartite Ecological Networks. TOECOLJ 2, 7–24.
570	https://doi.org/10.2174/1874213000902010007
571	Duchenne, F., Wüest, R.O., Graham, C.H., 2022. Seasonal structure of interactions enhances
572	multidimensional stability of mutualistic networks. Proceedings of the Royal Society B:
573	Biological Sciences 289. https://doi.org/10.1098/rspb.2022.0064
574	Dunn, L., Lequerica, M., Reid, C.R., Latty, T., 2020. Dual ecosystem services of syrphid flies
575	(Diptera: Syrphidae): pollinators and biological control agents. Pest Manag Sci 76, 1973-
576	1979. https://doi.org/10.1002/ps.5807
577	Ferreira, P.A., Boscolo, D., Viana, B.F., 2013. What do we know about the effects of landscape
578	changes on plant-pollinator interaction networks? Ecological Indicators 31, 35-40.
579	https://doi.org/10.1016/j.ecolind.2012.07.025
580	Fontaine, C., Thébault, E., Dajoz, I., 2009. Are insect pollinators more generalist than insect
581	herbivores? Proceedings of the Royal Society B: Biological Sciences 276, 3027–3033.
582	https://doi.org/10.1098/rspb.2009.0635
583	Fortuna, M.A., Stouffer, D.B., Olesen, J.M., Jordano, P., Mouillot, D., Krasnov, B.R., Poulin, R.,
584	Bascompte, J., 2010. Nestedness versus modularity in ecological networks: two sides of the
585	same coin? Journal of Animal Ecology. https://doi.org/10.1111/j.1365-2656.2010.01688.x
586	Frankie, G.W., Thorp, R.W., Hernandez, J., Rizzardi, M., Ertter, B., Pawelek, J.C., Witt, S.L.,
587	Schindler, M., Coville, R., Wojcik, V.A., 2009. Native bees are a rich natural resource in
588	urban California gardens. California Agriculture 63, 113-120.
589	https://doi.org/10.3733/ca.v063n03p113
590	Garibaldi, L.A., Steffan-Dewenter, I., Kremen, C., Morales, J.M., Bommarco, R., Cunningham, S.A.,
591	Carvalheiro, L.G., Chacoff, N.P., Dudenhöffer, J.H., Greenleaf, S.S., Holzschuh, A., Isaacs,
592	R., Krewenka, K., Mandelik, Y., Mayfield, M.M., Morandin, L.A., Potts, S.G., Ricketts, T.H.
593	Szentgyörgyi, H., Viana, B.F., Westphal, C., Winfree, R., Klein, A.M., 2011. Stability of
594	pollination services decreases with isolation from natural areas despite honey bee visits:

595	Habitat isolation and pollination stability. Ecology Letters 14, 1062–1072.
596	https://doi.org/10.1111/j.1461-0248.2011.01669.x
597	Geslin, B., Gauzens, B., Baude, M., Dajoz, I., Fontaine, C., Henry, M., Ropars, L., Rollin, O.,
598	Thébault, E., Vereecken, N.J., 2017. Massively Introduced Managed Species and Their
599	Consequences for Plant-Pollinator Interactions. Advances in Ecological Research 57, 147-
600	199. https://doi.org/10.1016/bs.aecr.2016.10.007
601	Giannini, T.C., Garibaldi, L.A., Acosta, A.L., Silva, J.S., Maia, K.P., Saraiva, A.M., Guimarães, P.R.,
602	Kleinert, A.M.P., 2015. Native and Non-Native Supergeneralist Bee Species Have Different
603	Effects on Plant-Bee Networks. PLoS ONE 10, e0137198.
604	https://doi.org/10.1371/journal.pone.0137198
605	Gillespie, C.S., 2015. Fitting Heavy Tailed Distributions: The poweRlaw Package. J. Stat. Soft. 64.
606	https://doi.org/10.18637/jss.v064.i02
607	Habel, J.C., Ulrich, W., Biburger, N., Seibold, S., Schmitt, T., 2019. Agricultural intensification drives
608	butterfly decline. Insect Conserv Divers icad.12343. https://doi.org/10.1111/icad.12343
609	Hoyle, M., Cresswell, J.E., 2007. The effect of wind direction on cross-pollination in wind-pollinated
610	gm crops. Ecological Applications 17, 1234–1243. https://doi.org/10.1890/06-0569
611	Jauker, F., Diekötter, T., Schwarzbach, F., Wolters, V., 2009. Pollinator dispersal in an agricultural
612	matrix: opposing responses of wild bees and hoverflies to landscape structure and distance
613	from main habitat. Landscape Ecol 24, 547–555. https://doi.org/10.1007/s10980-009-9331-2
614	Jauker, F., Wolters, V., 2008. Hover flies are efficient pollinators of oilseed rape. Oecologia 156, 819-
615	823. https://doi.org/10.1007/s00442-008-1034-x
616	Johst, K., Drechsler, M., Thomas, J., Settele, J., 2006. Influence of mowing on the persistence of two
617	endangered large blue butterfly species: Influence of mowing on butterfly persistence. Journal
618	of Applied Ecology 43, 333–342. https://doi.org/10.1111/j.1365-2664.2006.01125.x
619	Jordano, P., 2016. Sampling networks of ecological interactions. Functional Ecology 30, 1883–1893.
620	https://doi.org/10.1111/1365-2435.12763

621	Klein, AM., Vaissière, B.E., Cane, J.H., Steffan-Dewenter, I., Cunningham, S.A., Kremen, C.,
622	Tscharntke, T., 2007. Importance of pollinators in changing landscapes for world crops. Proc.
623	R. Soc. B. 274, 303–313. https://doi.org/10.1098/rspb.2006.3721
624	Kones, J.K., Soetaert, K., Van Oevelen, D., Owino, J., 2009. Are network indices robust indicators of
625	food web functioning? a Monte Carlo approach. Ecological Modelling 370–382.
626	Kovács-Hostyánszki, A., Földesi, R., Báldi, A., Endrédi, A., Jordán, F., 2019. The vulnerability of
627	plant-pollinator communities to honeybee decline: A comparative network analysis in
628	different habitat types. Ecological Indicators 97, 35–50.
629	https://doi.org/10.1016/j.ecolind.2018.09.047
630	Landi, P., Minoarivelo, H.O., Brännström, Å., Hui, C., Dieckmann, U., 2018. Complexity and stability
631	of ecological networks: a review of the theory. Population Ecology 60, 319–345.
632	https://doi.org/10.1007/s10144-018-0628-3
633	Lázaro, A., Müller, A., Ebmer, A.W., Dathe, H.H., Scheuchl, E., Schwarz, M., Risch, S., Pauly, A.,
634	Devalez, J., Tscheulin, T., Gómez-Martínez, C., Papas, E., Pickering, J., Waser, N.M.,
635	Petanidou, T., 2021. Impacts of beekeeping on wild bee diversity and pollination networks in
636	the Aegean Archipelago. Ecography 44, 1353-1365. https://doi.org/10.1111/ecog.05553
637	Liao, J., Bearup, D., Strona, G., 2022. A patch-dynamic metacommunity perspective on the
638	persistence of mutualistic and antagonistic. Ecology 103. https://doi.org/10.1002/ecy.3686
639	Ma, A., Lu, X., Gray, C., Raybould, A., Tamaddoni-Nezhad, A., Woodward, G., Bohan, D.A., 2019.
640	Ecological networks reveal resilience of agro-ecosystems to changes in farming management.
641	Nature Ecology & Evolution 3, 260–264. https://doi.org/10.1038/s41559-018-0757-2
642	Magrach, A., Holzschuh, A., Bartomeus, I., Riedinger, V., Roberts, S.P.M., Rundlof, M., Vujic, A.,
643	Wickens, J.B., Wickens, V.J., Bommarco, R., Gonzalez-Varo, J.P., Potts, S.G., Smith, H.G.,
644	Steffan-Dewenter, I., Vila, M., 2018. Plant-pollinator networks in semi-natural grasslands are
645	resistant to the loss of pollinators during blooming of mass-flowering crops. Ecography 41,
646	62–74. https://doi.org/10.1111/ecog.02847

647	Mathiasson, M.E., Rehan, S.M., 2020. Wild bee declines linked to plant-pollinator network changes
648	and plant species introductions. Insect Conservation and Diversity 13, 595-605.
649	https://doi.org/10.1111/icad.12429
650	Memmott, J., Waser, N.M., Price, M.V., 2004. Tolerance of pollination networks to species
651	extinctions. Proc. R. Soc. Lond. B 271, 2605–2611. https://doi.org/10.1098/rspb.2004.2909
652	Morrison, B.M.L., Brosi, B.J., Dirzo, R., 2020. Agricultural intensification drives changes in hybrid
653	network robustness by modifying network structure. Ecology Letters 23, 359-369.
654	https://doi.org/10.1111/ele.13440
655	Odoux, JF., Aupinel, P., Gateff, S., Requier, F., Henry, M., Bretagnolle, V., 2014. ECOBEE: a tool
656	for long-term honey bee colony monitoring at the landscape scale in West European intensive
657	agroecosystems. Journal of Apicultural Research 53, 57-66.
658	https://doi.org/10.3896/IBRA.1.53.1.05
659	Oksanen, J., Blanchet, F.G., Friendly, M., Kindt, R., Legendre, P., McGlinn, D., Minchin, P.R.,
660	O'Hara, R.B., Simpson, G.L., Solymos, P., Stevens, M.H.H., Szoecs, E., Wagner, H., 2020.
661	vegan: Community Ecology Package. R package version 2.5-7. https://CRAN.R-
662	project.org/package=vegan.
663	Okuyama, T., Holland, J.N., 2008. Network structural properties mediate the stability of mutualistic
664	communities. Ecology Letters 11, 208–216. https://doi.org/10.1111/j.1461-0248.2007.01137.x
665	Olesen, J.M., Stefanescu, C., Traveset, A., 2011. Strong, Long-Term Temporal Dynamics of an
666	Ecological Network. PLoS ONE 6, e26455. https://doi.org/10.1371/journal.pone.0026455
667	Ollerton, J., Winfree, R., Tarrant, S., 2011. How many flowering plants are pollinated by animals?
668	Oikos 120, 321–326. https://doi.org/10.1111/j.1600-0706.2010.18644.x
669	Perrot, T., Gaba, S., Roncoroni, M., Gautier, JL., Bretagnolle, V., 2018. Bees increase oilseed rape
670	yield under real field conditions. Agriculture, Ecosystems & Environment 266, 39-48.
671	https://doi.org/10.1016/j.agee.2018.07.020
672	Perrot, T., Gaba, S., Roncoroni, M., Gautier, JL., Saintilan, A., Bretagnolle, V., 2019. Experimental
673	quantification of insect pollination on sunflower yield, reconciling plant and field scale
674	estimates. Basic and Applied Ecology 34, 75–84. https://doi.org/10.1016/j.baae.2018.09.005

675	Petanidou, T., Kallimanis, A.S., Tzanopoulos, J., Sgardelis, S.P., Pantis, J.D., 2008. Long-term
676	observation of a pollination network: fluctuation in species and interactions, relative
677	invariance of network structure and implications for estimates of specialization: High
678	plasticity in plant-pollinator networks. Ecology Letters 11, 564–575.
679	https://doi.org/10.1111/j.1461-0248.2008.01170.x
680	Ponisio, L.C., Gaiarsa, M.P., Kremen, C., 2017. Opportunistic attachment assembles plant-pollinator
681	networks. Ecol Lett 20, 1261–1272. https://doi.org/10.1111/ele.12821
682	R Core Team, 2021. R: A language and environment for statistical computing. R Foundation for
683	Statistical Computing, Vienna, Austria. URL https://www.R-project.org/.
684	Requier, F., Odoux, JF., Tamic, T., Moreau, N., Henry, M., Decourtye, A., Bretagnolle, V., 2015.
685	Honey bee diet in intensive farmland habitats reveals an unexpectedly high flower richness
686	and a major role of weeds. Ecological Applications 25, 881-890. https://doi.org/10.1890/14-
687	1011.1
688	Rollin, O., Benelli, G., Benvenuti, S., Decourtye, A., Wratten, S.D., Canale, A., Desneux, N., 2016.
689	Weed-insect pollinator networks as bio-indicators of ecological sustainability in agriculture. A
690	review. Agron. Sustain. Dev. 36, 8. https://doi.org/10.1007/s13593-015-0342-x
691	Sánchez-Bayo, F., Wyckhuys, K.A.G., 2019. Worldwide decline of the entomofauna: A review of its
692	drivers. Biological Conservation 232, 8–27. https://doi.org/10.1016/j.biocon.2019.01.020
693	Stanley, D.A., Gunning, D., Stout, J.C., 2013. Pollinators and pollination of oilseed rape crops
694	(Brassica napus L.) in Ireland: ecological and economic incentives for pollinator conservation.
695	Journal of Insect Conservation 17, 1181-1189. https://doi.org/10.1007/s10841-013-9599-z
696	Stanley, D.A., Stout, J.C., 2014. Pollinator sharing between mass-flowering oilseed rape and co-
697	flowering wild plants: implications for wild plant pollination. Plant Ecol 215, 315–325.
698	https://doi.org/10.1007/s11258-014-0301-7
699	Thébault, E., Fontaine, C., 2010. Stability of Ecological Communities and the Architecture of
700	Mutualistic and Trophic Networks. Science 329, 853–856.
701	https://doi.org/10.1126/science.1188321

702	Thompson, A., Ştefan, V., Knight, T.M., 2021. Oilseed Rape Shares Abundant and Generalized
703	Pollinators with Its Co-Flowering Plant Species. Insects 12, 1096.
704	https://doi.org/10.3390/insects12121096
705	Valdovinos, F.S., 2019. Mutualistic networks: moving closer to a predictive theory. Ecology Letter
706	22, 1517–1534. https://doi.org/10.1111/ele.13279
707	Valido, A., Rodríguez-Rodríguez, M.C., Jordano, P., 2019. Honeybees disrupt the structure and
708	functionality of plant-pollinator networks. Scientific Reports 9, 4711.
709	https://doi.org/10.1038/s41598-019-41271-5