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The role of fire disturbances, human activities and climate change for long-term forest dynamics in upper-montane forests of the central Dinaric Alps

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20 Abstract

- 21 We present the first high-resolution Holocene pollen, plant-macrofossil, and charcoal records from the
- 22 upper-montane zone in the central Dinaric Alps. Drawing on these new records from well-dated
- 23 lacustrine sediments of Zminje Jezero (ca. 1500 m a.s.l.; Montenegro) and on independent chironomid-
- 24 inferred summer temperatures, we explore long-term ecosystem responses to variations in climate, fire
- disturbances and land use, as well as legacy effects of past environmental changes. 25
- A mixed spruce-fir forest established in the upper-montane zone around 9500 cal BP, and Fagus 26 27 sylvatica became co-dominant with the two conifers after 5000 cal BP. Prehistoric land-use pressure was overall remarkably low, but increased since 2000 cal BP and was highest after the Middle Ages. 28
- 29 We found a significant positive relationship between biomass burning and summer temperature,
- 30 indicating that fires were mostly climate driven. Picea abies was insensitive to summer temperature,
- 31 biomass burning and human impact, which supports the view that spruce forests may not be significantly
- impacted by fire. In contrast, Abies alba and other disturbance-sensitive trees (Tilia, Ulmus, Fraxinus 32
- 33 excelsior-type) show significant negative responses to land-use pressure and positive responses to
- 34 summer temperature. This supports the notion that these species may be well-adapted to warmer-than
- 35 present summer temperatures and that their populations declined in recent millennia due to land-use
- 36 activities. Conversely, F. sylvatica was sensitive to summer temperatures but was promoted by low
- 37 biomass burning, indicating that its expansion in the spruce-fir dominated forest was enhanced by the
- 38 onset of cooler and possibly also moister climatic conditions as well as by fire disturbances.
- 39

40 **Keywords**

Holocene, vegetation dynamics, mountain forests, fire history, vegetation-fire interactions, Montenegro 41 42

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48 Introduction

Assessing the responses of European mountain forests to changing environmental conditions is crucial to develop adaptation and management strategies to possible ecosystem shifts that may cause the loss of important ecosystem services, such as risk prevention (from avalanches, landslides or rockfall), recreation, and the maintenance and promotion of biodiversity (Klopčič et al., 2017). Such assessments are particularly important given that there is strong support for the hypothesis that climate change could markedly modify disturbance regimes, with a likely increase of some disturbances (e.g. fire) in a warming world (Seidl et al., 2017).

56 However, understanding natural ecological processes is challenging as processes often play out over long time scales (centuries, millennia), particularly when long-lived species such as trees are involved. 57 The long time horizon of palaeoecological records permits to study ecosystem responses under 58 59 substantially different environmental conditions than the present ones, including temperatures analogous to those predicted by future climate changes (approximately 1-2°C higher than present; Carter et al., 60 2018; Morales-Molino et al., 2021; Samartin et al., 2017). Moreover, current ecosystems and ecological 61 62 processes often carry legacy effects of past environmental changes and anthropogenic impacts (Cagliero 63 et al., 2022; Feurdean et al., 2009; Grindean et al., 2019; Morales-Molino et al., 2022), which have often altered species-environment relationships (Tinner et al., 2013). Thus, taking into account long-term 64 65 records may be important to gain more accurate environmental-change response assessments, especially 66 in Europe, where anthropogenic manipulation of nature started several millennia ago (Birks and Tinner, 67 2016).

68 The long-term history of mountain ecosystems varies across regions and is arguably contingent on patterns of human settlement, land use, and socioeconomic development. In this context, the central 69 70 Dinaric Alps are an interesting area as its land-use history may differ from those of other European 71 mountain regions. There is little doubt that the Balkan Peninsula acted as a land bridge from the Near 72 East, across Asia Minor and towards Central and Western Europe (Forenbaher et al., 2013). While it seems that the Neolithic spread mainly across the fertile plains along the Danube valley in the east 73 (Starčevo culture) and along the Adriatic coastal strip in the west (Impresso culture) around the 6th 74 millennium BCE (8000-7000 cal BP; Borić et al., 2019), there is evidence to suggest that the 75

mountainous hinterland of the central Dinaric Alps remained an agricultural frontier zone, eventually 76 even until the Late Neolithic (Forenbaher and Miracle, 2005). The frontier-zone hypothesis is consistent 77 78 with modelled anthropogenic deforestation (Kaplan et al., 2009), which suggests that during the past 79 3000 years relatively low forest clearance occurred when compared to other regions of Europe, and that a large fraction of usable land in the region was not heavily exploited until the Middle Ages (600-1460 80 CE; 1350-490 cal BP). By contrast, charcoal production, mining activities, and grazing likely occurred 81 82 in the interior area of the Dinaric Alps since the Bronze Age (3500-600 BCE; 5450-2550 cal BP), and 83 intensified during the Iron Age (600-200 BCE; 2550-2150 cal BP) and the Ottoman Empire (1460-1800 CE; 490-150 cal BP) (Kranic, 2009; Longman et al., 2018). However, it is not clear if these activities 84 were widespread across the mountain region, or whether human activities rather concentrated in some 85 areas. In the latter case, large forested regions in the interior range may have remained relatively intact 86 87 until the present (Nagel et al., 2017) or managed with low intensity silvicultural systems during recent centuries (Boncina, 2011). Low historical forest clearance in the montane belt may have preserved fir 88 (Abies alba Mill.), beech (Fagus sylvatica L.), and spruce (Picea abies (L.) Karst.) old-growth forests 89 90 (Cagliero et al., 2022; Motta et al., 2011; Sabatini et al., 2018). As these tree species are widespread and 91 often form mixed forests in European mountains (Hilmers et al., 2019), exploring their long-term responses under low land-use pressure may be useful to improve knowledge on the role of disturbances 92 by fire and climate. However, the impact of Neolithic and of more recent cultures on the interior 93 94 mountain region is still weakly constrained. In contrast to other European mountain regions, the area is 95 both archeologically and palaeoecologically under-documented (Borić et al., 2019; Finsinger et al., 96 2017; Vander Linden et al., 2014).

97 Thus, to provide new insights into the Holocene vegetation dynamics and into vegetation responses to 98 disturbances by fire and to changes in climate and land use we studied sediments from Zminje Jezero 99 (Fig. 1), a lake at 1535 m a.s.l., thus in the montane belt of the central Dinaric Alps. Our main aims are 100 to (1) assess the long-term Holocene vegetation dynamics of mixed montane forests, (2) characterize 101 the Holocene land-use and fire histories, (3) investigate the influence of the main drivers of fire 102 dynamics (climate and human impact), and (4) track responses of the dominant montane trees (*A. alba*, F. sylvatica, P. abies, as well as of other disturbance-sensitive trees such as Ulmus, Tilia, and Fraxinus
excelsior) to variations in climate, fire and human impact.

105

106

107 Material and methods

108

109 *Study area and study site*

110 In the Dinaric Alps, vegetation is influenced by the interaction of mountainous reliefs, pedological conditions, and proximity to the Adriatic Sea (Horvat et al., 1974; Nagel et al., 2017). Mediterranean 111 evergreen sclerophyllous vegetation dominated by *Pinus halepensis Mill.*, *Ouercus ilex L*. and *Olea* 112 113 europaea L. occurs near the coast. Mixed submediterranean deciduous woodlands with Carpinus betulus 114 L., Fraxinus ornus L., deciduous Quercus L., Ostrya carpinifolia Scop., Carpinus orientalis Mill., and Castanea sativa Mill. occur up to c. 700 m asl. The valley floodplains are commonly occupied by Alnus 115 glutinosa (L.) Gaertn., Ouercus robur L., Fraxinus sp. and Ulmus minor Mill.. In the montane belt (from 116 117 700 to 2000 m asl), forests are dominated by Pinus nigra J.F. Arnold, F. sylvatica, and mixed beech-fir-118 spruce forests. Other species occurring in the montane belt include Pinus sylvestris L., O. carpinifolia, Acer L., Ulmus L., Tilia L., and Fraxinus excelsior L.. The subalpine vegetation belt is generally above 119 c. 2000 m asl and is dominated by Pinus mugo Turra, Juniperus communis L., and Alnus viridis (Chaix) 120 121 DC.

I22 Zminje Jezero (Snake lake; 43°09'21" N 19°04'14" E; 1535 m asl) is a small 1.2-ha large and 9.5-m I23 deep glacial-origin lake (Fig. 1) whose water is drained into the Crno Jezero (1460 m asl) by a small I24 stream. The lake is located on the northwest part of a high karst plateau in the Durmitor massif (Annys I25 et al., 2014) in a valley adjacent to the one currently hosting the last surviving glacier in Montenegro, I26 Debeli namet (Hughes et al., 2011). Given its elevation, Zminje Jezero is within the limits of terminal I27 moraines that were deposited before the Lateglacial Interstadial (Hughes et al., 2011).

128 The vegetation in the surroundings of the lake is dominated by a dense spruce-fir forest (Fig. 1) with 129 lesser amounts of beech, *Acer* sp., *Sorbus aucuparia* L., *Rhamnus fallax* Boiss., and *Vaccinium myrtillus*

130 L. However, P. sylvestris, Pinus nigra subsp. nigra, Betula pendula Roth, and Pinus heldreichii H.

131 Christ are also well represented in the Durmitor massif. To the east, there is a town (Žabljak; 1450 m 132 asl) with pastures and agriculture lands. The lake shores host a rich community of wetland plants, 133 including *Molinia caerulea* (L.) Moench, *Comarum palustre* L., *Carex* sp., *Sphagnum* sp., *Parnassia* 134 *palustris* L., *Filipendula ulmaria* (L.) Maxim., and *Eriophorum vaginatum* L.. The current tree line is at 135 c. 1950 m asl and is mainly formed by spruce with beech (Bui, 1975). Above the tree line are *P. mugo* 136 and *J. communis* shrublands, alpine meadows, and rocks.

137 Climate is continental with a mean annual air temperature of 5.1°C and a mean annual precipitation of
138 c. 1450 mm at Žabljak (Annys et al., 2014). In the northern and central parts of the mountain range,
139 bedrock consists of Mesozoic sedimentary rocks and thick Middle and Upper Triassic and Upper
140 Jurassic limestones, while in the southern part Upper Cretaceous flysch are predominant (Mirković,
141 1985).

142 Zminje Jezero is located in a zone of special protection of the Durmitor National Park, where only 143 cutting of naturally collapsed trees (so-called "sanitary cuttings"), clearing of forest roads, and tourism 144 are allowed (Srdanović and Pavić, 2013; UNESCO and IUCN, 2018). Before the establishment of the 145 National Park in 1952 CE, the area was informally protected since 1907 CE.

146

147 *Field work and sediment-core correlation*

We studied the lake floor with an echo sounder connected to a Garmin GPS device (Fig. 1d) and 148 collected two overlapping sediment cores (sections 1-m long and 6 cm in diameter) from the deepest 149 150 part of the lake at a water depth of 9.3 m using a modified Livingstone piston corer and an UWITEC 151 gravity corer. All core sections were split longitudinally and photographed at ISEM (University of Montpellier) to visually describe the sediments and align the sections with the aid of Corelyzer v2.1.1 152 (CSDCO/LacCore-University of Minnesota). To assess sediment composition, we used an X-ray 153 154 fluorescence (XRF) AVAATECH core scanner at 5 mm resolution (with the two following settings: 10 kV/0.09 mA for 15 s and 30 kV/0.09 mA for 30 s), and then normalised the values using the centre-log-155 ratio method (CLR, Weltje et al., 2015). We correlated the sections based on marker layers and on 156 selected XRF records. Thereafter, one core half was cut into 1-cm slices and samples were stored in zip-157 lock bags at 4°C. 158

160 *Chronology*

161 We modelled the depth-age relationship (Fig. 2) with RBacon v2.4.1 (Blaauw and Christen, 2011) using 41 control points (Table 1 and Table S1), including 19 AMS ¹⁴C dates from terrestrial plant macrofossils 162 calibrated using the IntCal20 dataset (Reimer et al., 2020), one pollen-inferred age (onset of the 163 Holocene; Giesecke et al., 2014), and 21 control points derived from a chronology of short-lived 164 radionuclides (²¹⁰Pb and ¹³⁷Cs; Fig. S1) that was built with the constant flux constant sedimentation 165 166 (CFCS) using the serac package (Bruel and Sabatier, 2020). We excised visually identified event-deposit layers matching peaks of K, Ti, Rb, Zr, and Sr (Fig. 2), as they most probably represent sediment 167 deposition that occurred over very short time spans (Finsinger et al., 2021; Heiri et al., 2003). 168

169

170 Pollen, spores and plant-macrofossils analyses

We processed 78 1-cm³ samples for pollen analysis following standard physical and chemical 171 treatments, including treatments with HCl, KOH, sieving with a 500-um mesh and decanting, HF, 172 173 acetolysis, and mounted the fuchsin-stained residues on slides with glycerol. To calculate influxes (# cm⁻² yr⁻¹), we added *Lycopodium* tablets (Stockmarr, 1971). We identified and counted pollen, stomata 174 175 and the dung-fungi Sporormiella spore using determination keys and photographic atlases (Beug, 2004; Cagliero et al., 2022; Davis, 1987; Finsinger and Tinner, 2005; Moore et al., 1998; Reille, 1992) as well 176 177 as the pollen and stomata reference collections at ISEM (University of Montpellier). We counted at least 178 300 terrestrial pollen grains at 400x magnification and calculated percentages relative to the terrestrial 179 pollen sum, which includes pollen of trees, shrubs, upland herbs, ferns, and vascular cryptogams and excludes pollen of obligate aquatic plants. 180

The pollen diagram was divided into pollen assemblage zones using optimal partitioning by sums-ofsquares with square-root-transformed proportions of pollen types included in the terrestrial pollen sum, and exclusion of rare taxa (abundance <5%; Birks and Gordon 1985). The number of statistically significant zones was determined by comparison with the broken-stick model (Bennett, 1996) with Psimpoll v4.26 (Bennett, 2008).

For plant-macrofossils analyses, we processed 203 samples. We measured sample volumes (5.6±1.0 186 cm³) by water displacement and then sieved them with a 100-µm mesh sieve. We identified and counted 187 188 macrofossils with a stereomicroscope at 7.5-60x magnification using atlases (Birks, 2017; Katz et al., 189 1965; Schoch et al., 1988) and the reference collection at ISEM (University of Montpellier) and stored them at -18°C in Milli-Q water. Plant-macrofossil abundances are expressed as concentrations (# cm⁻³). 190 To assess land-use pressure, we used the abundances of Secale cereale pollen (a primary anthropogenic 191 192 indicator), secondary indicators (as the sum of adventives and apophytes), and cultural indicators (as the 193 sum of Cerealia-type and Plantago lanceolata-type pollen) (Behre, 1981; Tinner et al., 2003). These indexes include pollen of cultivated plants, ruderals, and plants of meadows (Deza-Araujo et al., 2020). 194 Veratrum album was added to the apophyte category, as it is avoided by cattle due to its toxicity 195 (Spiegelberger et al., 2006). We also assessed land-use pressure in the wider region using the OJC index 196 (Mercuri et al., 2013), which is based on the abundance of woody crops (Olea europaea L., Juglans 197 regia L., and Castanea sativa Mill.) that grow at lower elevations and along the coast of the Adriatic 198 199 Sea. Despite their presence in uncultivated areas, an increase in their abundances is often due to broad-200 scale cultivation (Deza-Araujo et al., 2020).

201

202 *Charcoal analyses*

We counted microscopic charcoal particles in 42 pollen slides spread along the sediment core (Finsinger 203 204 and Tinner, 2005; Tinner and Hu, 2003) to obtain a low-resolution record of regional fire activity. For macroscopic charcoal analyses, we took contiguous 1-cm³ samples and treated them without the use of 205 206 an orbital shaker with a 5% NaPO₃ solution overnight and then with a 15% H₂O₂ solution for 24h (Schlachter and Horn, 2010). Subsequently, samples were sieved through a 150-µm mesh (Higuera et 207 al., 2005) and charcoal particles were identified and counted under a stereomicroscope at 7.5-60x 208 209 magnification. The area of charcoal particles was not estimated since it is generally highly correlated to charcoal counts (Finsinger et al., 2014; Tinner et al., 1998). To account for variations in sedimentation 210 rate (Whitlock and Larsen, 2001), both microscopic and macroscopic charcoal counts were converted 211 to charcoal-accumulation rates (influx; # cm⁻² yr⁻¹), hereafter denoted mCHAR and MCHAR, 212 respectively. 213

215 *Fire episodes and fire-regime changes*

To reconstruct regional fire activity and landscape-scale biomass burning we used the long-term trends of two complementary proxies, mCHAR and MCHAR. While their potential source areas are similar (c. 40-km distance from study sites; Adolf et al., 2018), the records may differ, with MCHAR potentially reflecting fires within a smaller distance (Finsinger et al., 2017). To determine significant changes in landscape-scale biomass burning, we identified change points based on the mean and the variance in MCHAR and tested their independence from variations in the modelled sediment-accumulation rates (Finsinger et al., 2018).

To detect fire episodes within a few km distance around the study site (Higuera et al., 2010), we analysed 223 the MCHAR record with tapas v0.1.2 (Finsinger and Bonnici, 2022). This analysis involved resampling 224 the record to a constant resolution of 33 years (i.e. the 3rd quartile of the sampling-interval distribution). 225 226 Thereafter, we decomposed the record into a low-frequency background and a high-frequency peak 227 component using a loess robust to outliers with a smoothing-window width of 2000 years. The suitability 228 of the record for peak detection was evaluated through the signal-to-noise index (Kelly et al., 2011). Peaks were evaluated using the 95th percentile of the modelled noise distribution obtained with locally 229 230 defined Gaussian mixture models for each 2000-year portion of the record (Higuera et al., 2009). The peaks were further screened (Higuera et al., 2010) before calculating the fire return intervals (FRI: years 231 232 between adjacent fire episodes).

233

234 *Ecosystem responses*

We used generalized additive models (GAMs) to investigate ecosystem responses to variations in climate, fire and land use. GAMs are suited for this purpose as they can unfold nonlinear relationships between a response variable and a smoothed function of the predictor variable (Hastie and Tibshirani, 1986). Specifically, we explored the responses of fires to variations in climate and land-use pressure, and the responses of dominant forest canopy taxa (*P. abies, A. alba, F. sylvatica, F. excelsior-type, Tilia,* and *Ulmus*) to variations in climate, fires and land-use pressure (Carter et al., 2020; Colombaroli et al., 2010; Morales-Molino et al., 2021). As climate data, we used chironomid-inferred summer (July) surface-air temperature anomalies from Lago Verdarolo (1390 m asl) (Samartin et al., 2017) that is in a biogeographically comparable ecosystem in the Tuscan-Emilian Apennines where current vegetation is dominated by fir and beech. Comparisons between instrumental data and climate model runs (Samartin et al., 2017) show that at interannual to multidecadal timescales summer temperature variations in the Tuscan-Emilian Apennines are well correlated with grid cells in Montenegro (Pearson's correlation coefficients are 0.8-1 for multidecadal variability and 0.6-0.7 for interannual variability).

We assumed a quasi-Poisson distribution, log-transformed both the response and the predictor variables
to stabilize their variances, and fitted GAMs using restricted maximum likelihood (REML) smoothness
selection (Carter et al., 2020) with mgcv v1.8-38 (Wood, 2017). We chose a base period from 8000 cal
BP to the present because secondary anthropogenic indicators were very abundant during the early
Holocene, when human impact was negligible (Carter et al., 2020).

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- 255

256 **Results and interpretation**

257

258 Chronology and sediment composition

The sediments cover part of the Younger Dryas and the entire Holocene, from c. 12,340 cal BP to present (Fig. 2). The reliability of the chronology in its most recent part is broadly confirmed (Fig. S2) by the match between the expected and modelled ages of Pb peaks during the Roman period and the Middle Ages (McConnell et al., 2018; Renberg et al., 2001). Sediment deposition times vary between 6-65 years cm⁻¹ (median = 21.6 years cm⁻¹).

At the base, from 615 cm to 575 cm (12,340-11,640 cal BP), the sediments consist of light-brown sandy and silty clay (Fig. 2). From 575 cm to 479 cm (11,640-10,300 cal BP), there are 1-2-cm thick light grey silty-clay layers that alternate with darker and organic-richer layers (gradual boundaries), and a c. 28cm thick sequence of five graded deposits (529-500 cm), each one characterized by a distinct finingupward trend. Above 479 cm, sediments are mainly composed of dark-brown organic gyttja, which is interrupted by three minerogenic layers of fine grey detrital clay with sharp stratigraphic boundaries at 445-441 cm, 413-410 cm, and 283-281 cm depth. In keeping with the visual assessment of the sediments,

271 minerogenic elements (K, Ti, Rb, Zr, and Sr) are most abundant below 479 cm (10,300 cal BP) as well
272 as in conjunction with the minerogenic layers.

273

274 *Long-term vegetation dynamics*

The pollen record (Figs. 3-4) was divided into four statistically significant assemblage zones at 566.5, 455.5, and 359.5 cm (corresponding to 11,500, 9410, and 6400 cal BP) and three non-significant subzones with boundaries at 200.5 cm and 58.5 cm (2090 and 430 cal BP).

Pollen assemblages from 12,340 to 11,500 cal BP (ZMN-1; 615-566.5 cm) are dominated by Pinus and 278 herb pollen (mainly Artemisia, Poaceae and Chenopodiaceae). Both stomata and plant macrofossils 279 280 attest the local presence of Betula and Pinus, the latter presumably as low-density stands of dwarf mountain pine (P. mugo) or pine trees (P. sylvestris, P. heldreichii, or P. nigra). Cerealia-type pollen 281 282 are likely associated with wild Poaceae, as Cerealia-type pollen were also found elsewhere long before the development of agriculture (Dörfler, 2013). Such assemblages are typical of the Central European 283 Younger Dryas (12,850-11,650 cal BP; Rasmussen et al., 2014) and are indicative of an open and arid 284 285 steppe-tundra with cold-tolerant pine woodlands. Rising arboreal pollen (from c. 45 to 65%) indicate a gradual closure of the woodland during the Younger Dryas cold stage, including mixed oak woodlands 286 at lower altitude. 287

288 Picea abies stomata unequivocally indicate a rapid altitudinal shift of this tree species at the Younger 289 Dryas-Holocene transition, when a rapid climatic change including a warming of c. 2-4°C occurred 290 (Heiri et al., 2014; Tóth et al., 2015). The woodlands closed further from 11,500 to 9410 cal BP (ZMN-2; 566.5-455.5 cm), as attested by an increase of arboreal pollen (from c. 65% to 80%). F. excelsior-291 292 type, Ulmus, and Tilia pollen first appeared at 11,500 cal BP marking the expansion of meso-293 thermophilous trees at lower altitude, a characteristic feature of pollen records from the Balkan Peninsula and the adjacent European Alps (Caf et al., 2022; Vescovi et al., 2007; Willis, 1994). Betula 294 populations were denser than during the Younger Dryas, as shown by both pollen and plant macrofossils. 295 However, the abundance of light-demanding pioneer trees and shrubs (Ephedra, Juniperus and Betula) 296 decreased in conjunction with the local expansion of spruce and fir populations starting between 10,400 297

and 9600 cal BP. Finds of *Sporormiella* dung-fungi spores indicate the presence of wild herbivores
around the lake between 11,500 and 10,400 cal BP, as also observed on the Swiss Plateau (Rey et al.,
2017) and in the Maritime Alps (Finsinger et al., 2021).

301 The shift to a Picea-Abies dominated forest was mostly concluded shortly after 9410 cal BP (onset of ZMN-3; 455.5-359.5 cm), as also observed at Prokoško Jezero (1670 m asl, Bosnia Herzegovina; 302 303 Dörfler, 2013). Plant-macrofossils show that both *Pinus* and *Betula* were still intermixed in the spruce-304 fir dominated forest until around 8000 cal BP, when a low-density population of beech likely 305 established, as suggested by the continuous pollen curve. At lower altitudes, vegetation was characterized by mixed oak forests with a higher share of Corvlus between 9410 and 6000 cal BP. 306 Around 7250 cal BP, Ostrva-type pollen increased indicating the expansion of O. carpinifolia, C. 307 308 orientalis, or of both species.

This expansion was followed by an increase of Fagus and C. betulus pollen starting around 6400 cal BP 309 (onset of ZMN-4a; 359.5-200.5 cm). Although the Fagus expansion at Zminje Jezero occurred 310 approximately 2500-1000 years later than in pollen records from the Central Dinaric mountain area 311 312 (Prokoško Jezero, Crveni Potok), it broadly coincided with the expansion of Ostrya and Carpinus 313 betulus, as often seen in pollen records from the region (Prokoško Jezero, Lake Ohrid, Lake Prespa, and Crveni Potok; Dörfler, 2013; Finsinger et al., 2017; Panagiotopoulos et al., 2013; Wagner et al., 2009). 314 The admixture of Fagus in the Picea-Abies dominated forest involved a c. 2000-years long build-up 315 316 phase (6400-4600 cal BP), as attested by rising pollen abundance, after which Fagus plant macrofossils 317 and stomata regularly occur. As both C. betulus and Fagus are shade-tolerant and favoured by moist 318 habitats on upland soils (Houston Durrant et al., 2016; Sikkema et al., 2016), an increase in humidity during the growing season may have occurred (Finsinger et al., 2017). In keeping with this, plant 319 320 macrofossils of more light-demanding and better drought-resistant *Pinus* are absent in this zone, though 321 stomata occasionally indicate their scattered presence. Secale cereale pollen first appear in this zone and occur mostly between 5800 and 3800 cal BP, suggesting its presence as a weed among other cereals 322 during the Bronze Age (Behre, 1992). Some secondary anthropogenic-indicator pollen types (e.g. P. 323 lanceolata-type, Melampyrum) increased as well. However, tree pollen abundance was highest in this 324 time interval, suggesting that land-use pressure was minor. Similarly, a slightly lower tree-pollen 325

abundance between 3800 and 2100 cal BP suggests moderate presence of forest openings that are,however, not matched by rising abundance of anthropogenic-indicator pollen types.

328 By contrast, anthropogenic-indicator pollen document higher land-use pressures from Roman times onwards (c. after 2100 cal BP; ZMN-4b and ZMN-4c). Secale cereale pollen is mostly present around 329 the Roman period and regularly occurs since the Ottoman Empire (ZMN-4c; 58.5-0 cm; 430 cal BP to 330 present). Similarly, both cultural and secondary indicators are slightly higher during the Roman period 331 and substantially rise since 430 cal BP in conjunction with a substantial tree-pollen decrease. Thus, 332 333 pollen overall document two land-use phases characterized respectively by moderate land-use pressure (Roman period) and increasing and reaching highest land-use pressure since the Ottoman Empire. These 334 two land-use phases were interrupted by a phase with minor land-use pressure during the Middle Ages 335 (ZMN-4b; 200.5-58.5 cm; 2090-430 cal BP). 336

The frequent occurrence of *C. sativa* and *O. europaea* pollen since 1800 cal BP is consistent with a human-mediated spread of chestnut populations since Roman times (Conedera et al., 2004) and the cultivation of olive trees in Roman territories (Mercuri et al., 2013), including Montenegro, which was conquered by the Romans during the 2nd century BCE. Further, as in other records in the region (Dörfler, 2013; Finsinger et al., 2017), pollen of *Juglans* first appear at c. 1500 cal BP and occur more regularly after 500 cal BP.

Although *Abies* pollen abundance was substantially lower after 2100 cal BP, a mixed *Abies-Picea-Fagus*was still present locally until 430 cal BP, as confirmed by stomata and plant macrofossils. However,
lower pollen percentages and the absence of plant macrofossils of *Fagus* in conjunction with rare finds
of *Abies* macrofossils since 430 cal BP suggest a major shift towards a *Picea*-dominated forest. Forest
cover increased during the last century, perhaps due to the protection of the area since 1907 CE.

348

349 Charcoal-inferred fire history

Regional fire activity, inferred based on microscopic charcoal (mCHAR), was highest over the past 2000 years and was moderately high between 12,000 and 10,000 cal BP (Figs. 3-4). By contrast, landscapescale biomass burning (MCHAR trend) was highest between 11,400 and 6000 cal BP, and overall decreasing from 6000 cal BP to the present (Figs. 3-4 and S4). The signal-to-noise index (Fig. S5) shows that the macroscopic charcoal-accumulation rate (MCHAR) record is suitable for peak-detection analysis (Kelly et al., 2011). A total of 35 fire episodes were identified over the past 12,500 years, with a median fire-return interval (mFRI) of 260 years (FRI range: 79-790 years). Longest FRI values (>530 years) occur before 8000 and after 2000 cal BP. Charred *A. alba* and *P. abies* needles document the occurrence of local fires in the conifer-dominated mountain ecosystem, mostly between 5600 and 500 cal BP (Figs. 3-4).

360

361 *Response of fire to variations in climate and land-use pressure*

Landscape-scale biomass burning (MCHAR trend) and regional fire activity (mCHAR) show different responses to climate and land-use pressure. Whereas biomass burning significantly increases in response to summer temperatures, fire activity significantly decreases (Fig. 5a; Table 2). Similarly, whereas biomass burning significantly decreases in response to land-use pressure, regional fire activity shows a hump-shaped response including an increase for the lower range of anthropogenic indicator pollen values (Fig. 5b).

368

369 *Response of forest canopy species to variations in climate, fire and land use*

Picea's responses to variations in climate, land use and biomass burning are not statistically significant
(Fig. 5c-e). *Fagus* shows a significant negative response, and *Abies*, *F. excelsior*-type, *Ulmus*, and *Tilia*show significant positive responses to both summer temperature and biomass burning (Fig. 5c and e;
Table 2). *Abies*, *F. excelsior*-type and *Ulmus* show significant negative responses to land-use pressure
(Fig. 5d).

375

376

377 Discussion

378

381

379 *Regional vs local land-use*

380 The investigation of long-term vegetation dynamics in relation to land-use pressure and disturbances

requires a detailed understanding of land-use history. This is particularly important for southern Europe,

where forest clearances and agriculture often involved the use of fire. Moreover, land-use phases were often interspersed by periods of abandonment and secondary forest establishment as the result of the development and expansion of more permanent land-use practices, such as animal husbandry, ploughing, crop cultivation, and woodland management (Birks and Tinner, 2016). However, identifying the start and inferring the intensity of land-use pressure using palaeoecological records is challenging, particularly in regions where independent documentation by archaeological evidence is poor (Deza-Araujo et al., 2020), as is the case for the Durmitor.

389 Stomata and plant macrofossils indicate the local presence of taxa, mostly within some decametres around a site (Ammann et al., 2014; Birks, 2017). Thus, the continuous presence of plant macrofossils 390 391 and stomata of trees (*Picea*, Abies, and Fagus) indicates that tree cover persisted in the surroundings of Zminje Jezero at least throughout the past 9000 years (Fig. 3). In keeping with this, high arboreal pollen 392 (generally >80%) and AP/NAP ratio values (generally >4; Fig. 4) qualitatively indicate the presence of 393 closed forest (Deza-Araujo et al., 2020). A persistently high forest cover was also found in the internal 394 395 Dinaric Alps at Crveni potok (Finsinger et al., 2017), but contrasts with results from Prokoško Jezero 396 (Dörfler, 2013) where arboreal pollen values decreased to 50% during the past 2000 years. This 397 contrasting evidence would support the hypothesis that some of the forested regions in the interior range 398 of the Dinaric Alps have been left relatively intact until the present (Nagel et al., 2017).

Conversely, if land-use pressure was regionally heterogeneous, reconstructions based on pollen may not 399 400 necessarily covary with those based on plant macrofossils and stomata. Pollen from the wider region 401 may reach small lakes such as Zminje Jezero even if their relevant pollen source area is relatively small 402 (some kilometres at most; Sugita, 1994). Indeed, despite persistence of closed forest in the surroundings of Zminje Jezero, anthropogenic indicators unambiguously document increasing land-use pressure in 403 404 the region since the Roman period. Cultural indicators, secondary indicators, as well as OJC-index 405 values rise around 2000 cal BP, when the region currently belonging to Montenegro was conquered by the Romans (2nd century BCE). At that time, land-use pressure probably reached the Durmitor, as 406 attested by the construction of roads, including the one connecting Nikšić and Komini. The further rise 407 of anthropogenic indicators around 500 cal BP (15th century CE) is coherent with the first local 408 409 settlements in the Durmitor. For instance, the so-called stećci sites (medieval tombstones and graveyards from the 13th-17th century; Erdeljan, 2018) attest to the local presence of settlements on the Durmitor during the Middle Ages and the Ottoman Empire, and churches and monasteries from the 15th-19th century CE document the presence of local settlements during the Ottoman Empire. Thus, these results add a finer-scale scale documentation of past land-use pressure in comparison to modelled anthropogenic deforestation patterns (Kaplan et al., 2009), which suggest that relatively low forest clearance occurred during the past 3000 years and that the region was not heavily exploited until the Middle Ages.

417 Similarly, regional heterogeneity of land-use pressure may also have occurred in pre-Roman times. Anthropogenic indicators do not show distinct signs of Bronze and Iron Age land use, despite the 418 probable connection between coastal and inland settlements (Bulatović et al., 2020) and Illyrian and 419 420 Celtic Iron Age settlements (Cozzolino et al., 2020). However, Secale cereale pollen finds at Zminje Jezero suggest its presence as a weed among other cereals (Behre, 1992) since about 6000 cal BP, as 421 422 also observed in archaeobotanical samples from Neolithic sites in the mountain hinterland (e.g. at Okolište, Bosnia-Herzegovina; de Vareilles et al., 2022). While this is about 1000 years later than the 423 424 age of Neolithic settlements on the shores of the lowland Lake Ohrid (Hafner et al., 2021) and 2000 425 years later than the earliest Starčevo and Butmir culture Neolithic settlements in the inland of Bosnia-Herzegovina (Vander Linden et al., 2014), the Secale pollen finds support the view that Neolithic 426 427 novelties, albeit with a certain time lag, spread in the hinterland of Montenegro (Borić et al., 2019). 428 Although Neolithic sites have not yet been found in the Durmitor area, there is evidence of Neolithic 429 occupation sites at 20-35 km distance (Odmut rock shelter; Borić et al., 2019), suggesting that inland 430 mountain areas may be archeologically under surveyed (Vander Linden et al., 2014). Thus, the palaeoecological records agree with the hypothesis that the mountain hinterland remained an agricultural 431 432 frontier zone for much of the Neolithic (Forenbaher and Miracle, 2005) and possibly even until the 433 Roman Age.

434

435 *Regional vs local fires*

We found contrasting responses of fire to climate and land-use pressure over the past 8500 years (Fig.
5a-b), pointing to different drivers of fire dynamics at landscape and regional scales. Landscape-scale

biomass burning, as inferred based on MCHAR, responded positively and significantly to summer 438 temperature, as expected based on the strong influence (summer) temperature has on fire activity (Jain 439 440 et al., 2022; Westerling et al., 2006) and as often observed with sedimentary charcoal records (Daniau 441 et al., 2012; Power et al., 2008). As fuel was not a limiting factor in the moist and productive mountain forests, higher mean July temperatures perhaps coupled with prolonged dry seasons (Holocene Thermal 442 Maximum; Samartin et al., 2017) likely improved fuel flammability (Pausas and Paula, 2012). By 443 contrast, surprisingly regional fire activity, as inferred based on mCHAR, responded negatively to 444 445 summer temperature. A disconnect between fire-history reconstructions can arise both when comparing paleofire records from sites located near each other (Finsinger et al., 2018; Gavin et al., 2006), as well 446 as different paleofire proxies from the same site (Finsinger et al., 2017) if fire occurrence was not 447 spatially homogeneous. While charcoal records from additional sites may be necessary to better 448 constrain the spatial heterogeneity of fire occurrence in the region, it is likely that mCHAR captures fire 449 450 activity at greater distance from the site than MCHAR (Finsinger et al., 2017). This interpretation is supported by the positive response of fire activity (mCHAR) to land-use pressure (Fig. 5b), which likely 451 452 occurred in a wider region, including lower-elevation areas where submediterranean deciduous 453 woodlands occur. The slightly earlier rise of regional fire activity, which started around the Iron Age (Fig. 4), may be connected to land use in the region that remained undetected by the anthropogenic 454 455 indicators from Zminje Jezero. Biomass burning, instead, responded negatively to land-use pressure 456 (Fig. 5c). It is unlikely that this relationship is the result of feedbacks arising from human-induced 457 deforestation (e.g. lower fuel availability, fragmentation; Bowman et al., 2011) because biomass burning 458 decreased predominantly at the Neolithic/Bronze Age transition, when land-use pressure was low. The overall decreasing biomass burning suggests that humans did not extensively use fire in the Durmitor. 459 460 While this overall contrasts with paleofire evidence from other southern European mountains (e.g. the 461 Apennines; Morales-Molino et al., 2021), the lower biomass burning rates in recent millennia fit with 462 the notion that fires are currently rare in the Dinaric Alps (Nagel et al., 2017). However, charred Picea and Abies needles unambiguously document the occurrence of fires in the mixed spruce-fir-beech 463 mountain forests of the Durmitor. 464

466 Species-response curves and legacies of past environmental changes

Species-response curves are useful to assess factors that have driven long-term vegetation dynamics, 467 468 thereby unfolding legacies of past environmental changes (Carter et al., 2020; Colombaroli et al., 2010; 469 Morales-Molino et al., 2021). We found similar species responses to summer temperature and biomass burning (Fig. 5c and 5e). Given the strong and positive relationship among these two predictors (Fig. 470 5a), it is difficult to disentangle their roles separately. However, as (summer) temperature currently has 471 a strong influence on fire activity (Jain et al., 2022; Westerling et al., 2006), the response curves may 472 473 yield valuable insights into the long-term fire ecology of the analysed species (P. abies, F. sylvatica, A. alba, Ulmus, Fraxinus excelsior, and Tilia), which are widespread and often dominant in European 474 mountain forests. The right-skewed responses of Fagus to biomass burning and temperature (Fig. 5c 475 476 and 5e) suggest that it was favoured under low biomass-burning and cooler conditions. Overall, these 477 response curves support the notion that Fagus became more competitive in mixed fir-spruce forests with the onset of cooler and moister summers and fire disturbance, as observed in the ecologically similar 478 479 Northern Apennines (Morales-Molino et al., 2021). Indeed, Fagus requires moist summers for its 480 germination (Giesecke et al., 2017) and is more sensitive than Abies to drought due to its shallow root 481 system (Cheddadi et al., 2016; Tinner et al., 2013). Fagus is a shade-tolerant species and its regeneration 482 can establish and grow with low light levels under the forest canopy for years before rapidly expanding 483 and achieving dominance when light conditions improve after disturbances (Gardner and Willis, 1999). 484 Thus, fire occurrence after 5000 cal BP may have promoted the recruitment of Fagus and its admixture 485 into the mixed *Picea-Abies* forest. This matches with evidence suggesting that during the initial phase 486 of stand establishment, Fagus is favoured by mixed-severity fire and intermediate human disturbances (Bradshaw and Lindbladh, 2005; Giesecke et al., 2007; Tinner and Lotter, 2006) thanks to its capacity 487 488 to occupy newly exposed soils and canopy openings (Carter et al., 2018) more rapidly than its two main 489 competitors (Abies and Picea). Despite lacking ecological adaptations to fire and being identified as a 490 fire-sensitive species that is affected by higher biomass burning and temperatures (Carter et al., 2018; Morales-Molino et al., 2021; Tinner et al., 2000), Fagus can be advantaged by fire incidence as long as 491 it is not excessive (Maringer et al., 2020). In keeping with this, the change-point analysis shows a 492 decrease in biomass burning during the establishment of Fagus (Fig. 4), corroborating the view that 493

494 beech establishment occurs in conjunction with a shift to a lower fire activity (Bobek et al., 2019;495 Feurdean et al., 2017).

496 Further, similar to observations from the ecologically similar Northern Apennines (Morales-Molino et 497 al., 2021), we found that Abies and disturbance-sensitive deciduous taxa such as Ulmus, F. excelsior, and *Tilia* were promoted by warmer-than-present summer temperatures (up to ca. 2.5°C; Fig. 5c). 498 499 However, our findings also suggest that these taxa responded positively to increasing biomass burning 500 (Fig. 5e). This contrasts with several records showing that both high-severity infrequent fires and low-501 severity frequent fires led to their rapid decline and to local extinctions in recent millennia (Morales-Molino et al., 2021; Tinner et al., 2000, 2013). On the one hand, the striking positive response of these 502 disturbance-sensitive taxa to biomass burning could simply arise from the positive relationship between 503 landscape-scale biomass burning and summer temperatures (Fig. 5a). On the other hand, the significant 504 505 negative response of Abies, Ulmus, and F. excelsior to land-use pressure (Fig. 5d) as well as to regional 506 fire activity (Fig. S6 and Table S2) indicates that their populations declined even in the wider region 507 due to land use, probably with the use of fires. Indeed, the decline of *Abies* coincides with the rise in 508 land-use pressure (pasturing and small-scale cultivation) and of regional fire activity since 2000 cal BP. 509 Conversely, the local persistence of Abies in the surroundings of Zminje Jezero until present-day 510 suggests that land-use pressure was not widespread across the montane forest belt, leaving relatively undisturbed patches of forests in a landscape mosaic (Fig. 1). Indeed, Abies can be resistant to fire 511 512 disturbances when fires are rare (Henne et al., 2013) and its populations strongly reduce under excessive 513 human-induced fire activity (Feurdean and Willis, 2008; Finsinger et al., 2021; Tinner et al., 2013), 514 which was not the case around Zminje Jezero but may have been the case to the East of Žabljak and elsewhere in the upper-montane zone of the central Dinaric Alps. 515

Picea's response curves suggest that neither temperature nor biomass burning and land use significantly influenced its populations (Fig. 5c-e). Indeed, *Picea* remained relatively stable even under higher-thanpresent summer temperatures (Fig. 4), as also observed with palaeoecological records from the Carpathians (Carter et al., 2018). Its insensitivity to summer temperatures may simply imply that past climatic conditions did not exceed the species' tolerance range. Moreover, the lack of a significant response to biomass burning agrees with evidence showing that *Picea* can persist thanks to vigorous post-fire regeneration from wind-dispersed seeds (Bobek et al., 2018) even with median FRI as low as
260 years as at Zminje Jezero (Brown and Giesecke, 2014; Feurdean et al., 2017; Finsinger et al., 2018).
Although fires may have been fuelled by *Picea's* traits that promote fire spread (high resin content,
flammable litter, and a ladder-canopy structure), its well-established populations persisted at Zminje
Jezero despite fire occurrence in the mixed spruce-fir-beech forests.

The species-response curves may also be useful to explain the relatively recent shift from a *Picea-Abies*-527 Fagus mixed forest to one dominated by Picea that occurred in the surroundings of Zminje Jezero 528 around 430 cal BP (15th century CE; Fig. 4). Fir populations, which are more sensitive to land-use 529 pressure than *Picea* and *Fagus* (Fig. 5d), could have been reduced due to the establishment of the local 530 pastoral katun society (c. 1477 CE; Turkish Land Registry). Around 1400 CE, the Balkan Peninsula was 531 under the influence of the Republic of Venice and from 1496 CE Montenegro fell under the domain of 532 the Ottoman Empire (Rastoder, 2003). In the mid-16th century, local land-owners started seizing 533 monastery estates and hunting grounds in the lower parts of Montenegro. This resulted in the retreat of 534 considerable parts of the population in mountain areas (Rastoder, 2003), including the Durmitor, whose 535 536 forests probably started to be more intensively exploited. An alternative, or additional explanation could 537 involve the lower sensitivity of spruce to declining winter temperatures (Cheddadi et al., 2016). In this 538 scenario, the development of a *Picea*-dominated forest could be linked to the onset of cooler conditions during the Little Ice Age (1450-1850 CE). Indeed, due to its shallow root system Picea is more 539 540 vulnerable to drought than Abies (Henne et al., 2011), and spruce-dominated stands in the Dinaric Alps 541 mostly occur in habitats with significantly colder and moister microclimates (Nagel et al., 2017).

542

543

544 Conclusions

545 Vegetation of forest ecosystems is often the result of complex interactions between vegetation dynamics 546 and variations in climate, disturbances (e.g. fire), and human activities. However, understanding the 547 ecological processes underlying these interactions is challenging as processes often play out over long 548 time scales (centuries, millennia), particularly for forests where long-lived species such as trees are involved. The long-term perspective of palaeoecological records permits to study ecosystem responsesunder substantially different environmental conditions than the present ones.

In this study, we sought to investigate the Holocene vegetation and fire history in the upper-montane zone of the central Dinaric Alps, where fires are currently rare and forests mainly include beechdominated stands and mixed forests dominated by varying amounts of beech, fir, and occasionally spruce (Nagel et al., 2017). While records from different elevations may be necessary to better support our interpretation, our results support the view that at least some parts of the central Dinaric Alps remained an agricultural frontier zone during the Neolithic (Forenbaher and Miracle, 2005) and that land-use pressure was low until the Middle Ages (Kaplan et al., 2009).

Response curves indicate that fires in the upper-montane zone were mostly driven by variations in 558 climatic conditions, as expected based on the strong influence (summer) temperature has on fire activity 559 today (Jain et al., 2022; Westerling et al., 2006). Picea was insensitive to variations in summer 560 temperature, biomass burning and human impact, which supports the view that spruce forests may not 561 be significantly impacted by fire (Carter et al., 2018). In contrast, Abies and other disturbance-sensitive 562 563 deciduous trees (F. excelsior-type, Tilia, and Ulmus) show a significant positive relationship with 564 summer temperatures and a significant negative relationship with land-use pressure. This supports the 565 notion that these tree species may be well-adapted to warmer-than-present summer temperatures and 566 that their populations declined in recent millennia due to land-use activities (Morales-Molino et al., 567 2020; Tinner et al., 2000, 2013). Conversely, Fagus emerges as a species sensitive to summer 568 temperatures whose expansion in the Picea-Abies dominated forest was enhanced by the onset of cooler 569 and possibly moister climatic conditions as well as by fire disturbances.

570

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579	

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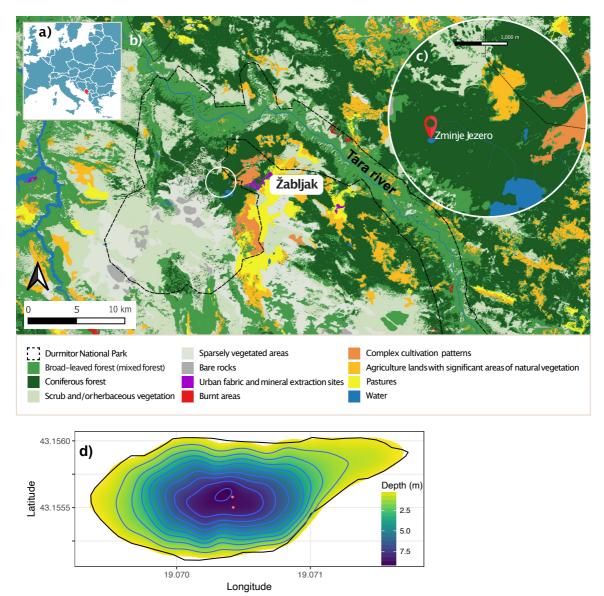
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584

- 585 Data
- 586 The palaeoecological data (pollen, spores, stomata, plant macrofossils, charcoal, and XRF records) will
- 587 be uploaded to the Neotoma database (DOI: 10.21233/XYZZ-XYZZ) through the European Pollen

588 Database.



591 Fig. 1 a-c) Location of the study site at decreasing spatial scales. b-c) Map showing the distribution of

land cover types (CORINE Land Cover 2018) in the Durmitor National Park and surroundings. d)
bathymetric map of Zminje Jezero with contour lines at 1-m intervals and coring locations (red circles).

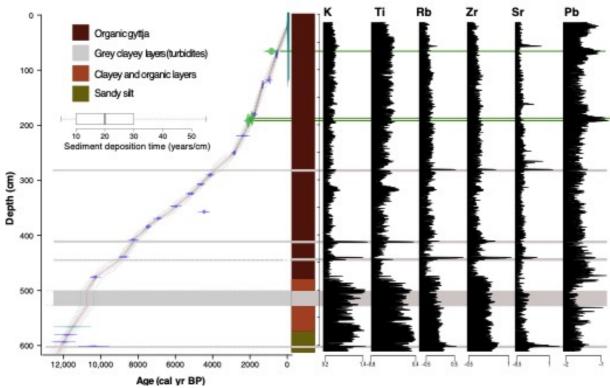
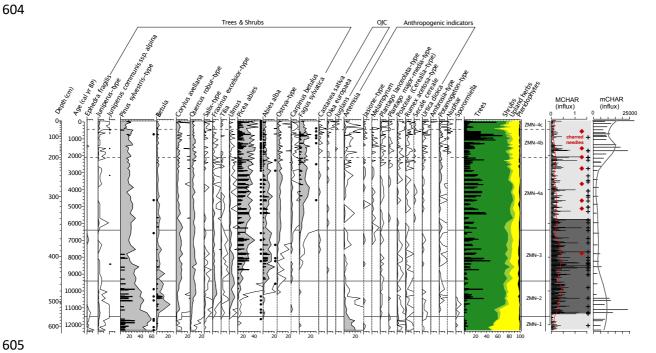


Fig. 2 Depth-age model for the Zminje Jezero sediment core. The active control points used to constrain the model (Table 1) are based on ¹⁴C, ²¹⁰Pb, and ¹³⁷Cs dates as well as on a pollen-inferred age estimate (onset of the Holocene). Passive control points based on local maxima in the XRF-Pb record (McConnell et al., 2018; Renberg et al., 2001) are shown in green (see also Fig. S2). Grey horizontal bands indicate turbidite layers that were excluded from the model. Lithological and XRF geochemical profiles of the sediment core are shown on the right. The inset shows the distribution of sediment deposition time within the sequence (median = 21 years cm⁻¹)



606 Fig. 3 Vegetation and fire history from Zminje Jezero sediments. Left: synthetic pollen diagram with selected pollen and spore percentages of the terrestrial pollen sum (grey curves; empty curves show 10x 607 exaggerations of the percentage values), presence/absence of stomata (black circles), and plant-608 macrofossil concentrations (black vertical bars, dimensionless) for selected genera and for all woody 609 taxa of the Zminje Jezero. The continuous horizontal and dashed lines indicate statistically significant 610 611 and non-significant pollen-assemblage zones, respectively. See Fig. S3 for a more detailed plantmacrofossil diagram. To the right: macroscopic-charcoal accumulation rate (MCHAR) and microscopic-612 613 charcoal accumulation rate (mCHAR) as black bars, their long-term trends (red and black lines), fire 614 episodes (black crosses), and presence of charred conifer needles (red diamonds).

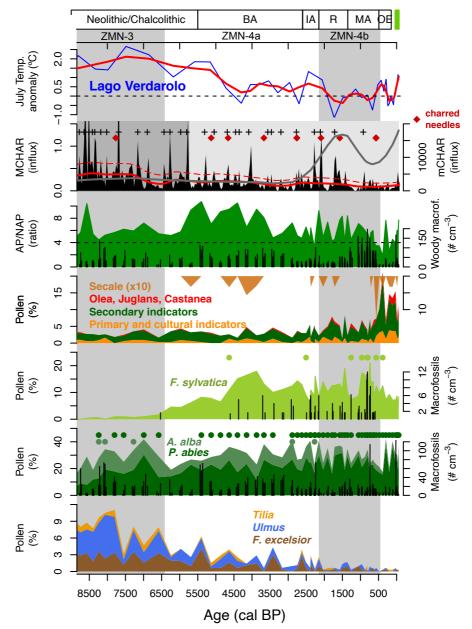


Fig. 4 Comparison of the main proxies for vegetation dynamics, fire and land-use history from Zminje 617 Jezero, and chironomid-inferred July-air temperature (T_{July}) anomalies from Lago Verdarolo (Samartin 618 et al., 2017). Microscopic charcoal (mCHAR, thick dark-grey line) documenting regional fire activity 619 and macroscopic charcoal (MCHAR, black filled area) documenting landscape-scale biomass burning 620 621 (red continuous line), local fire episodes (black crosses), periods of biomass burning as determined by 622 the change-point analysis (grey-shaded areas), and charred *Picea* and *Abies* needles (red diamonds). Temporal changes of dominant forest canopy taxa and of anthropogenic indicator pollen types (filled 623 624 polygons: pollen percentages; black vertical bars: plant macrofossils concentrations; grey circles: stomata). Grey vertical shaded areas represent pollen zone boundaries (see Fig. 3). 625

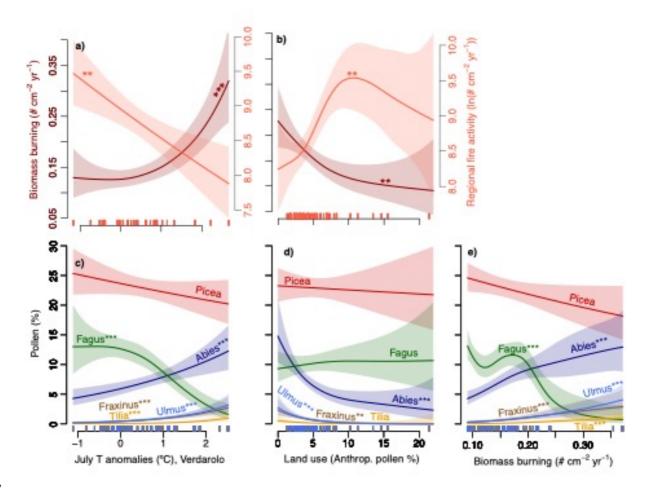




Fig. 5 Generalized Additive Models showing (a-b) the responses of landscape-scale biomass burning
 and regional fire activity to climate and land-use pressure, and (c-e) the responses of dominant forest

630 canopy taxa to climate, land-use pressure, and biomass burning.

Table 1 Control points used to establish the depth-age model for the ZMN-19 sediment core from Zminje jezero (Durmitor Massif, Montenegro). For a detailed description of the terrestrial plant macrofossils used for radiocarbon dating see Table S1.

Depth (cm)	Lab-code	Material	Dry weight	Sample Thickness	Age (¹⁴ C yr BP)
(cili)			(mg)	(cm)	(cal yr BP)*
0	²¹⁰ Pb	Bulk sediment	/	0	$-69 \pm 5*$
0.5	²¹⁰ Pb	Bulk sediment	/	0	$-65 \pm 1*$
1.5	²¹⁰ Pb	Bulk sediment	/	0	$-62 \pm 1*$
2.5	²¹⁰ Pb	Bulk sediment	/	0	$-58 \pm 1*$
3.5	²¹⁰ Pb	Bulk sediment	/	0	$-54 \pm 1*$
4.5	²¹⁰ Pb	Bulk sediment	/	0	$-51 \pm 1*$
5.5	²¹⁰ Pb	Bulk sediment	/	0	$-47 \pm 2^{*}$
6.5	²¹⁰ Pb	Bulk sediment	/	0	$-44 \pm 2^{*}$
7.5	²¹⁰ Pb	Bulk sediment	/	0	$-40 \pm 2^{*}$
8.5	²¹⁰ Pb	Bulk sediment	/	0	$-36 \pm 3*$
9.5	¹³⁷ Cs peak	Bulk sediment	/	0	$-36 \pm 1*$
9.5	²¹⁰ Pb	Bulk sediment	/	0	$-33 \pm 3*$
10.5	²¹⁰ Pb	Bulk sediment	/	0	$-29 \pm 3*$
12.5	²¹⁰ Pb	Bulk sediment	/	0	$-22 \pm 4*$
13.5	²¹⁰ Pb	Bulk sediment	/	0	$-19 \pm 4*$
14.5	NWT ¹³⁷ Cs	Bulk sediment	/	0	$-15 \pm 5*$
14.5	²¹⁰ Pb	Bulk sediment	/	0	$-13 \pm 1*$
15.5	²¹⁰ Pb	Bulk sediment	/	0	-11 ± 5*
16.5	¹³⁷ Cs	Bulk sediment	/	0	$-8 \pm 5^{*}$
16.5	²¹⁰ Pb	Bulk sediment	/	0	-5 ± 5*
17.5	²¹⁰ Pb	Bulk sediment	/	0	$-4 \pm 6^{*}$
71.5	DeA-33892	Terrestrial plant macrofossils	2.25	1	565 ± 19
118.5	DeA-23961	Terrestrial plant macrofossils	20	1	1054 ± 24
126.5	DeA-32618	Terrestrial plant macrofossils	16.8	1	1427 ± 20
180.5	DeA-31441	Terrestrial plant macrofossils	9.8	1	1844 ± 23
219.5	DeA-33893	Terrestrial plant macrofossils	3.5	1	2335 ± 55
250.5	DeA-31442	Terrestrial plant macrofossils	8	1	2755 ± 22
290.5	DeA-31091	Terrestrial plant macrofossils	8.5	1	3762 ± 22
307.5	DeA-23962	Terrestrial plant macrofossils	194	1	4108 ± 30
324.5	DeA-31093	Terrestrial plant macrofossils	17.8	1	4532 ± 25
347.5	DeA-33890	Terrestrial plant macrofossils	1.18	1	5212 ± 51
357.5	DeA-23963	Terrestrial plant macrofossils	266.5	1	3980 ± 30
369.5	DeA-31440	Terrestrial plant macrofossils	17.6	1	6060 ± 28
384.5	DeA-23964	Terrestrial plant macrofossils	170.2	1	6552 ± 37
408.5	DeA-33891	Terrestrial plant macrofossils	1.6	1	7449 ± 60
439.5	DeA-23967	Terrestrial plant macrofossils	148.5	1	7908 ± 39
476	DeA-23965	Terrestrial plant macrofossils	532.3	0.5	9139 ± 44
566.5	/	Pollen inferred (onset	/	1	$11500 \pm 250*$
		Holocene)			
581	Poz-10170	Terrestrial plant macrofossils	0.4	2	10170 ± 60
594	Poz-143126	Terrestrial plant macrofossils	3	2	10260 ± 60
601.5	Poz-143573	Terrestrial plant macrofossils	0.08	3	9190 ± 120

Table 2 Ecosystem responses to variations in climate, fire and land use at Zminje jezero (Durmitor Massif, Montenegro), as determined by Generalized Additive Models (GAMs) based on the pollen percentages and charcoal-accumulation rate records (this study) as well as chironomid-inferred July temperatures from Lago Verdarolo (Samartin et al., 2017). Effective degrees of freedom indicate the complexity of the smooth (a value = 1 indicates a straight line); significance of smooth terms: $* = P \le 0.05$, $** = P \le 0.01$, $*** = P \le 0.001$.

Independent variable	Response variable	Effective degrees of freedom	Reference degrees of freedom	F- value	Deviance explained (%)
	Biomass burning***	2.54	3.08	11.2	52.2
July es)	Regional fire**	1	1	12.18	28.3
rred . mali	P. abies	1	1	2.2	4.07
Climate nid-infer ture ano	A. alba***	1	1	14.31	22.4
Clin nid- ture	F. sylvatica***	3.07	3.88	11.47	57
Climate (chironomid-inferred July temperature anomalies)	F. excelsior-type***	1	1	21.72	28.9
(chir tem	Tilia***	1.88	2.35	10.02	31.1
-	Ulmus***	1	1	33.14	37.2
<u>د</u>	P. abies	1.0	1.0	3.88	7.1
ning o ho	A. alba***	2.32	2.78	10.65	39.1
Biomass burning (long-term trend of MCHAR)	F. sylvatica***	4.33	5.2	5.61	53
mass burr 5-term tre MCHAR)	F. excelsior-type***	1.83	2.21	0.2	32.4
Bion long-	Tilia***	1.79	2.16	12.26	37
	Ulmus***	1.65	1.99	23	46.5
	Biomass burning**	1.89	2.37	6.55	26.3
ttor	Regional fire***	3.32	4.07	7.13	38.4
sure ndica	P. abies	1	1	0.11	0.21
pres nic ir en)	A. alba***	2.26	2.83	11.14	41.7
Land-use pressure (anthropogenic indicator pollen)	F. sylvatica	1.34	1.61	0.12	1.6
hrop	F. excelsior-type**	1.08	1.16	10.37	25.2
I (ant)	Tilia	1.48	1.82	2.59	13.7
	Ulmus***	1.03	1.05	12.8	30.7

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 - Table 2 EcosMontenegro),charcoal-accunVerdarolo (Sa:1 indicates a st

Independent variable

Response variable

Effective degrees of freedom

Reference degrees of freedom

F- Deviance value explained (%)