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Identifying mature fish aggregation areas during spawning season by combining catch declarations and scientific survey data

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Abstract :

Identifying and protecting essential fish habitats like spawning grounds requires an accurate knowledge of fish spatio-temporal distribution. Commercial declarations coupled with Vessel Monitoring System provide fine scale information on the full year to map fish distribution and identify essential habitats. We developed an integrated framework to infer fish spatial distribution on a monthly time step by combining scientific and commercial data while explicitly considering the preferential sampling of fishermen towards areas of higher biomass. We developed a method to identify areas of persistent aggregation of biomass during the spawning season and interpret these as spawning areas. The model is applied to infer maps of relative biomass for three species (sole, whiting, squids) in the Bay of Biscay on a monthly time step over a 9-year period. Integrating several fleets in inference provides a good coverage of the area and improves model predictions. The preferential sampling parameters give insights into the temporal dynamics of the targeting behavior of the different fleets. Last, persistent aggregation areas reveal consistent with the available literature on spawning grounds, highlighting the potential of our approach to identify reproduction areas.

29 **1 INTRODUCTION**

30 Integrating fisheries into Marine Spatial Planning (MSP) to preserve ecosystem functions
31 and ensure a sustainable exploitation requires an accurate knowledge of fish spatio-
32 temporal distribution and more specifically of fish essential habitats such as reproduction
33 and nursery grounds (Janßen et al. 2018). However, such knowledge is still missing for
34 many species due to a lack of data with sufficient spatial, temporal or demographic
35 resolution (Delage and Le Pape 2016; Regimbart et al. 2018).

36 The available data to map fish distribution and identify essential habitats mainly rely on
37 either scientific survey data (fishery-independent data) or commercial data available
38 through on-board observer programs (fishery-dependent data) (Pennino et al. 2016). Both
39 data sources benefit of direct on-board recording of catches and are usually considered
40 as high quality data. Furthermore, both data sources were proved to be complementary
41 (Rufener et al. 2021). Scientific data benefit from a standardized sampling plan, a
42 standardized catchability and occur each year at the same period. Consequently, they
43 provide standardized data on a large spatial extent for most species and size classes
44 (Hilborn and Walters 2013; Nielsen 2015). Observer data potentially provide data over the
45 full year for all caught species, even though they do not follow a standardized protocol as
46 survey data. However, both scientific survey and onboard observer data are characterized
47 by a relatively low sampling intensity in space and time. Because of material limitations,
48 surveys occur only once or twice a year and provide a limited number of samples each
49 time (ICES 2005) and observer programs only cover a limited fraction of the entire fleet
50 (e.g. only 1% of all sea trips are covered by the French observer programs - Cornou et
51 al., 2021). The low sampling density of both data sources may lead to imprecise

52 predictions (ICES, 2005; Alglave et al., 2022) and constrains to consider only rough
53 temporal resolution (e.g. semesters, quarters or seasons – see for instance Kai et al.,
54 2017; Pinto et al., 2019; Rufener et al., 2021) to ensure a satisfying spatial coverage of
55 the data at each time step. However, the temporality of key biological events, such as the
56 reproduction peak, may be much tighter than the temporal resolution of data (Biggs et al.
57 2021). Hence, those data alone are likely not sufficient to provide accurate inferences on
58 essential fish habitats such as spawning grounds.

59 Commercial catch declarations combined with their fishing locations available from VMS
60 (Vessel Monitoring System) were proven to be an interesting alternative to obtain landing
61 per unit effort (LPUE) data with fine spatial and temporal resolution (Pedersen et al. 2009;
62 Bastardie et al. 2010; Gerritsen and Lordan 2010; Hintzen et al. 2012; Murray et al. 2013;
63 Azevedo and Silva 2020). However, considering commercial fisheries data to infer fish
64 spatial distribution remains highly challenging. Among other challenges, this implies
65 accounting for fisher sampling behavior. Fisher typically tend to preferentially sample
66 areas of higher biomass (a process referred to as preferential sampling, PS - Diggle et al.
67 2010). Hence, because data preferentially represent areas of highest biomass, ignoring
68 PS in the distribution of fishing effort when estimating spatial distribution on larger areas
69 can lead to overestimated biomass predictions (Conn et al. 2017; Pennino et al. 2019;
70 Alglave et al. 2022).

71 In a recent paper, Alglave et al. (2022) developed an integrated modeling framework to
72 infer spatial distribution of fish abundance by combining scientific survey CPUE and
73 commercial LPUE data while accounting for PS in the distribution of fishing effort. They
74 applied their framework to commercial data of a single month to match with the scientific
75 survey and did not consider any temporal dimension.

76 In this paper, we extend the modeling framework from Alglave et al. (2022) by adding a
77 temporal dimension to estimate fish spatio-temporal distribution at a monthly time step.
78 Our new model accounts for the variation over time (monthly time step) in the biomass
79 field as well as in the intensity of PS for distinct fishing fleets. To demonstrate the value of
80 the method, we selected and applied the model to 3 demersal species in the Bay of Biscay
81 (common sole, whiting and squids) characterized by contrasted configurations regarding
82 the available knowledge of their spawning grounds. We used those applications to
83 reinforce results obtained in Alglave et al. (2022) demonstrating how the integrated
84 framework allows to combine the information from several fleets in order to produce
85 accurate maps of spatio-temporal biomass. To illustrate the capacity of the framework to
86 identify areas of aggregation during the spawning season, we processed model outputs
87 to identify areas of recurrent aggregation occurring during the reproduction season and
88 compared these to the information available in the literature.

89

90 **2 MATERIAL AND METHODS**

91 In this section, we first present the different species, the datasets and how we process
92 and combine them to produce LPUE data in space and time. Second, we extend the model
93 proposed by Alglave et al. (2022) to introduce a temporal dimension on a discrete monthly
94 time step. In our applications, the models were fitted to data from 2010 to 2018 on a
95 monthly time step (108 time steps). Then, we illustrate how the PS component modifies
96 model predictions and can be interpreted, and how integrating several fleets in the
97 analysis further improves model predictions. Last, we detail the method used to
98 investigate spatio-temporal dynamics from model outputs and identify reproduction
99 grounds based on the aggregation patterns of each of three species.

100 **2.1 Case studies**

101 Sole is a data-rich case. Direct information about reproduction grounds is available
102 through egg and larvae surveys (Arbault et al. 1986). Reproduction period fall between
103 January and April, but the peak of the reproduction fall in February. Discard rate is also
104 very low, which makes the landings data a good proxy of the catch (ICES 2019a).

105 By contrast, Whiting is a data-poor case study where only indirect information of
106 reproduction period exists through spring trawl surveys (House and Forest 1993).
107 Reproduction period fall between March and May. Discard rates can be high (about 30%)
108 and thus landing data may provide a biased picture of the real catches (ICES 2019b).

109 Our third case study is Squids that represent a mixture of several species declared under
110 a common denomination in the catch (Loliginidae here referred as squids): *Loligo*
Vulgaris

111 (Lamarck, 1798), *Loligo forbesii* (Steenstrup, 1856) and *Alloteuthis sp* (Lamarck, 1798
).

112 Overall scientific survey suggests that the predominant species in the Bay of Biscay is

113 *Loligo Vulgaris* (ICES, 2020a, p.17). All 3 species are data-poor: no information exists
114 regarding their reproduction grounds but some information of the reproduction period
115 exists for *Loligo Vulgaris* (Moreno et al. 2002). For this species, the reproduction period
116 fall between January and April.

117 **2.2 Data**

118 **2.2.1 High spatial resolution catch per unit effort data for the mature component** 119 **of the populations**

120 We pre-processed the VMS and catch declaration (logbook) data to obtain high spatial
121 resolution LPUE data for the mature component of those three stocks, for three different
122 fishing fleet, and for each month of the 2010-2018 time series. In the text, we used the
123 term Landing Per Unit of Effort (LPUE) to refer to commercial observations expressed in
124 kg of (mature) biomass per hour fished. Discards are neglected in our approach, hence
125 LPUE are considered as biomass indices.

126 Our model can integrate data of different fishing fleets. For the purpose of our application,
127 we selected three different métiers that belong to the same fleet of trawlers (in this case,
128 the 'métier' term refers to a combination of gear and of a set of species that are targeted
129 by the vessels – see the Data Collection Framework and EC (2008)): OTB_DEF (bottom
130 otter trawl targeting demersal species), OTB_CEP (bottom otter trawl targeting
131 cephalopods) and OTT_DEF (multi-rig otter trawl targeting demersal species). In the
132 following those three métiers are referred as three fleets. These three fleets were selected
133 as they offer three main advantages. First, their targeting behaviors and technical
134 characteristics are similar. Second, catch per unit effort of trawlers are generally good
135 indicator of fish relative abundance while other gears (longline, gillnet) may face saturation
136 effects leading to non-linear relationship between catches and fishing time (Hovgêrd and

137 Lassen 2008). Third, the combination of the three fleets cover the full spatial domain
138 (Figure 1).

139 Because one of our primary goal is to identify spawning grounds, we filtered only the
140 mature fraction of the landings (i.e. the fraction of the individuals that can potentially
141 reproduce, not per se the fraction of the population that are spawning – this is detailed
142 further in the discussion). This was done by crossing the landings data with length class
143 and maturity data. For most of the landings, information on the commercial size categories
144 is available from the sales notes. These commercial categories are regularly sampled to
145 derive length structure of each commercial category. This allows us to estimate the
146 proportion of potentially mature fish in each commercial category by applying maturity
147 ogives and in turn estimate the proportion of mature fish for each landing declaration. See
148 SM1 for more detail. Note that this procedure was not possible for squids, as there are no
149 data on maturity and size classes for this species group.

150 Landing data were then combined with VMS data to finally obtained high spatial resolution
151 LPUE data discretized on a $0.05^\circ \times 0.05^\circ$ grid (i.e. 5.5 km x 3 km) on a monthly time step
152 (see SM2). This combination requires:

153 (1) to identify the fishing locations within the VMS data. This is realized individually
154 for each fishing vessel trajectory based on a speed threshold similarly as in
155 common data processing methods (Hintzen et al. 2012).

156 (2) to reallocate the logbooks declaration on the related VMS fishing locations. This
157 reallocation is realized individually for each fishing vessel trajectory by uniformly
158 reallocating the landings on all fishing locations. The link between both data
159 sources is realized through the combination 'vessel identifier x statistical

160 rectangle x fishing trip x day'. LPUE are then computed by simply dividing the
161 reallocated landings by the related fishing time.

162 **2.2.2 Scientific data**

163 We also integrated scientific data in the analysis. For whiting and squids we used the
164 survey data from the EVHOE survey. The Orhago survey was used for sole (ICES, 2020
165 - see SM3, Figure S3). The data were extracted from the DATRAS database on the period
166 2010 - 2018. Only the mature fraction of the survey catches were kept in the analysis to
167 make it comparable with commercial data.

168 Orhago is an annual beam trawl survey occurring in November and designed to assess
169 sole stock status in the Bay of Biscay. Each year 50 stations are sampled within 4 strata
170 all along the Bay of Biscay. Note that this survey is mainly coastal and does not sample
171 offshore areas. EVHOE is an annual bottom trawl survey occurring in late October,
172 November and early December with a stratified sampling plan. It is designed for demersal
173 fishes in the Bay of Biscay and in the Celtic Sea. In the Bay of Biscay, 80 to 90 sampling
174 hauls are recorded each year.

175 **2.3 Spatio-temporal integrated model**

176 Alglave et al. (2022) developed a hierarchical integrated statistical model to infer spatial
177 distribution of fish density through scientific survey data and commercial data. It is
178 structured in 4 layers:

- 179 - the latent field that represents biomass spatial distribution, and that is the
180 main target of the inferences;

- 181 - the observations from scientific surveys and commercial declarations that
 182 are considered as direct zero-inflated observations of the latent field at the
 183 registered fishing locations;
- 184 - the fishing sampling intensity that relates fishing locations to the latent field
 185 and model explicitly the PS of commercial fleets towards areas of higher
 186 biomass;
- 187 - unknown parameters that control the shape of the biomass latent field and
 188 the sampling process.

189 This first model was purely spatial as no temporal dimension was included in the model.

190 In this paper, we extend the framework by incorporating a temporal component to model
 191 the evolution of the latent field of biomass across the monthly time steps (Figure 2).

192 **2.3.1 Biomass field**

193 As a notable extension of Alglave et al. (2022), the biomass field (eq.1) is modeled as a
 194 spatio-temporal Gaussian Random Field (GRF) through a log link as:

$$195 \quad \log(S(x,t)) = \alpha_S(t) + \delta(x,t) \quad (1)$$

196 where $x \in D \subset R^2$ stands for the spatial locations and $t \in \llbracket 1, T \rrbracket$ for the monthly time steps.

197 $S(x,t)$ is in the same unit as the data (here kg/hr fished as data are CPUE for survey trawls
 198 or LPUE for commercial landing declarations). The term $\alpha_S(t)$ is a time varying intercept

199 modeled as a fixed effect and $\delta(x,t)$ is a GRF spatio-temporal process which represents

200 the spatio-temporal correlation structure of the biomass field. As commercial data may not

201 always cover all time steps, the temporal correlation is a critical component that allows to

202 interpolate between time-steps. Here, the spatio-temporal term has a classical stationary

203 first-order autoregressive form (eq.2) following (Cameletti et al. 2013):

204
$$\delta(x,t) = \varphi \cdot \delta(x,t-1) + \omega(x,t) \quad \text{for } t = 2, \dots, T \quad (2)$$

205 The autocorrelation coefficient φ is a scalar with $\varphi \in -1, 1$, $\omega(x,t)$ are spatial random
206 effects that represents the spatial innovation, modeled as a 0 mean GRF (with no temporal
207 correlation). Spatial random effects $\omega(x,t)$ are parameterized through a range parameter
208 ρ that corresponds to the distance at which spatial autocorrelation falls below 0.1.

209 Note that no covariate is included in the latent field to keep the model as simple as possible.
210 If any, the covariates effects are captured through the spatio-temporal term $\delta(x,t)$.
211 Similarly, the intercept $\alpha_S(t)$ was modeled through a simple fixed effect but more complex
212 specifications including some seasonal, yearly and interaction effects could be adopted
213 such as in Thorson et al. (2020).

214 **2.3.2 Sampling process for the commercial fishing points**

215 For most scientific surveys targeting a variety of species or covering wide areas, even if
216 the sampling plan is a stratified-random sampling and sampling takes into account the
217 most important commercial species in the area, the final distribution of the sampling
218 location can be considered independent from the biomass field distribution (see S3 for the
219 scientific survey used in this paper), and scientific sampling locations do not need to be
220 modeled explicitly (Diggle et al. 2010). By contrasts, the dependence between the
221 commercial fishing locations and the biomass field has to be modeled and included in the
222 likelihood function to capture preferential sampling. We extended the model proposed by
223 Alglave et al. (2022) to account for temporal variations in PS. Because the fishing behavior
224 is potentially different among fishing fleet, PS is modeled specifically for each fleet
225 denoted j (in the next section, j is also used to denote the scientific survey). Observed
226 fishing locations of any fishing fleet j are integrated in the likelihood through an

227 inhomogeneous point process (X_{comj} in the Figure 2) whose intensity $\lambda_j(x,t)$ (eq.3) controls
 228 the expected number of fishing points within a given area:

$$229 \quad \log(\lambda_j(x,t)) = \alpha_{X_j}(t) + b_j(t) \cdot \log(S(x,t)) + \eta_j(x,t) \quad (3)$$

230 where:

231 - the time varying intercept $\alpha_{X_j}(t)$ quantifies the average fishing intensity on the
 232 whole area; it is modeled as a fixed effect;

233 - the time varying $b_j(t)$ quantifies the strength of PS towards biomass $S(x,t)$; it is
 234 modeled as a fixed effect too. If $b_j(t) = 0$, then PS is null. If $b_j(t) > 0$, then PS
 235 occurs and the greater, the stronger PS. Alternatively, $b_j(t) < 0$ would indicate that
 236 fisher have a repulsive behavior towards the resource.

237 - the pure spatial GRF $\eta_j(x,t)$ captures the remaining spatial variability in the fishing
 238 point pattern not captured by the PS term (for instance, dependence of the fishing
 239 locations towards management regulations, distribution of other targeted species,
 240 habits/tradition).

241 **2.3.3 Observation process**

242 All observations for both scientific and commercial data of any fleet j are assumed all
 243 mutually independent conditionally on the latent field of biomass and the sampling
 244 locations. As data (both scientific and commercial) eventually present a high proportion of
 245 zero values, we model the observations through a Poisson-link zero-inflated model
 246 introduced by Thorson (2018) and already used in Alglave et al. (2022). The observation
 247 model explicitly considers that each fleet can have its own catchability and its own zero
 248 inflation parameter.

249 The probability to observe a catch data y_i conditionally on the location x_i (with i the
 250 observation index), the time-step t_i , the biomass field value $S(x_i, t_i)$ and the fleet j is
 251 expressed as follow:

$$252 \quad P(Y_i = y_i | x_i, t_i, S(x_i, t_i), j) = \begin{cases} p_i & \text{if } y_i = 0 \\ (1 - p_i) \cdot L\left(y_i \frac{\mu_j(x_i, t_i)}{(1 - p_i)}, \sigma_j^2\right) & \text{if } y_i > 0 \end{cases} \quad (4)$$

$$253 \quad p_i = \exp(-e^{\xi_j} \cdot \mu_j(x_i, t_i)) \quad (5)$$

254 $\mu_j(x_i, t_i) = q_j \cdot S(x_i, t_i)$ is the expected catch of fleet j at location x_i and time step t_i . It is the
 255 product of the latent field value $S(x_i, t_i)$ and of the relative catchability coefficient specific
 256 for fleet j denoted q_j . Specifically for each fleet j , ξ_j is a zero-inflation parameter controlling
 257 the proportion of zero in the data, σ_j^2 is the observation variance when the catch is positive.
 258 Equation (4) shows the two components that compose the probability to observe a catch
 259 y_i :

- 260 • the probability to obtain a zero catch ($y_i = 0$). It is modeled as a Bernoulli variable
 261 with probability $p_i = \exp(-e^{\xi_j} \cdot \mu_j(x_i, t_i))$. p_i is equivalent to the probability to obtain a
 262 0 value with a Poisson distribution of intensity $e^{\xi_j} \cdot \mu_j(x_i, t_i)$. The value of ξ_j controls
 263 the intensity of the zero inflation, when ξ_j increases, the amount of zero in the data
 264 decreases. Then the probability to obtain a positive catch is given by $1 - p_i$.
- 265 • the value of the positive catch is modeled through a lognormal distribution L with
 266 expected value $\frac{\mu_j(x_i, t_i)}{(1 - p_i)}$ and observation error σ_j^2 . The standardization by $(1 - p_i)$
 267 allows to keep the expectancy of the observation model to $\mu_j(x)$.

268 The catchabilities q_j are not identifiable per se and some additional constraints need to be
269 set to estimate the relative catchability of each fleet (Alglave et al., (2022)). To ensure
270 identifiability, one fleet catchability is set as reference level (e.g. $q_{ref} = 1$, here OTB_DEF
271 was used as the reference fleet) and the other fleets' catchabilities are estimated relatively
272 to the reference fleet through the equation:

$$273 \quad q_j = k_j \cdot q_{ref} \quad (6)$$

274 **2.3.4 Maximum likelihood estimation**

275 The estimation of the spatio-temporal model is achieved through maximum likelihood
276 estimation. We used the Stochastic Partial Differential Equation (SPDE) approach that
277 allows to benefit from the nice computational properties of Gaussian Markov Random
278 Fields while working on a continuous domain (Lindgren et al. 2011). In practice our model
279 is coded with Template Model Builder (TMB - Kristensen et al., 2016) which benefits from
280 the Laplace approximation to integrate over random effects, automatic differentiation and
281 sparse matrix computation technics for a fast estimation of the model through maximum
282 likelihood estimation. Details on estimation are provided in SM 4, 5 and 6.

283 **2.4 Evaluating the interest of integrating multiple fleets**

284 Integrating several fleets in inference allows to cover the whole area (Figure 1) and is
285 expected to improve inferences. To illustrate the value of integrating the data from multiple
286 fleets within a single integrated model, we compared the spatial predictions obtained by
287 fitting the model to all available data with those obtained by integrating only one fleet. In
288 addition, we investigated if integrating all the fleets in inference increased the correlation
289 between scientific data and model predictions.

290 We also compared the coefficient of variation of the prediction between each model on a
291 single time step (here November 2018).

292 Note that scientific data was systematically integrated into inference (either in the
293 integrated model or in the single-fleet models). However, due to the low sample size
294 compared with intensive 'VMS x logbooks' data (about 80 scientific samples each year in
295 November compared with 17000 samples per month on average), they have very low
296 contribution to inference. This was extensively discussed in Alglave et al. (2022). Here
297 they mainly provide some standardized and reference data to assess the performance of
298 the framework.

299 **2.5 Evaluating the value of modeling PS**

300 **2.5.1 Comparing the inferences with and without PS**

301 We first assessed the impact of PS on the distribution of biomass by comparing
302 estimations obtained from integrated models (i.e. models fitted to all data sources)
303 accounting for PS with those obtained when ignoring PS. We computed the log-likelihood
304 related to each data source (commercial and scientific data) to assess if there is an
305 improvement in model goodness-of-fit when accounting or not for PS. Note that fitting a
306 model without PS is straightforward as it only requires to remove the sampling process
307 component from the likelihood function.

308 **2.5.2 Interpreting the intensity of preferential sampling**

309 The estimates of PS parameters $b_j(t)$ in eq. (3) may bring valuable information on the
310 dynamics of the fishery as they inform on the strength of the relationship between
311 commercial sampling distribution and species distribution. We investigate the variability of
312 the PS parameters among the three species and the different fleets. Then, focusing on

313 the sole case study, we highlight the insights brought by the model on the temporal
314 evolution of PS and its seasonal variations.

315 **2.6 Investigating spatio-temporal distribution and identifying reproduction** 316 **grounds**

317 The spatio-temporal model provides some insight on the temporal dynamics of species
318 distribution both at inter- and intra-annual levels. We applied a method to identify recurrent
319 aggregation areas from the maps of abundance inferred at each time step.

320 **2.6.1 Aggregation index**

321 We used the Getis and Ord index $G_d(x,t)$ (Getis and Ord 1992; Ord and Getis 1995) to
322 determine persistent aggregation areas (see for instance Milisenda et al., (2021)). The
323 generalized version of the Getis and Ord index is given in Bivand and Wong (2018) and
324 Ord and Getis (1995). Basically, $G_d(x,t)$ is a normalized version of the ratio between the
325 sum of the log-biomass (denoted $s(x,t)$) within a fixed neighborhood d and the sum of s
326 (x,t) on the entire area (for a fixed time step) (Getis and Ord 1992). We computed these
327 indices on $s(x,t) = \log(S(x,t))$ so that the $s(x,t)$ are Gaussian, which makes G_d Gaussian
328 too. In the application, we used a neighborhood distance $d=7.5$ km which defines a small
329 neighborhood of 8 cells (the direct neighbors of each cell grid) and allows to identify very
330 localized aggregation areas. Positive values for the aggregation index G_d indicates that s
331 (x,t) fall within a local patch of high values while negative G_d indicates that $s(x,t)$ fall within
332 a local patch of low values. Near 0 values G_d , indicates that $s(x,t)$ does not fall in some
333 local aggregation patch. As G_d follows a standardized Gaussian distribution, the
334 comparison between the value of the index and the quantiles of a standard Gaussian
335 distribution can be used to evaluate whether or not the latent field of biomass fall within a
336 statistically significant high or low aggregation patch. We used the quantile 99% (2.58) as

337 a threshold to ensure a high level of significance for patch detection (only local patch of
338 positive values are considered) and applied the Bonferroni correction to account for the
339 multiple statistical tests that are conducted.

340 Then, we define the persistence indices $IP(x,m)$ as the proportion of times point x falls
341 significantly within an aggregation area for a specific month/season m (can be either a
342 month or several months) among several years. Areas marked with high values of IP are
343 persistent aggregation areas throughout the time series.

344 **2.6.2 Comparing the results with the available literature**

345 Persistent aggregation areas derived from our model during the reproduction period (as
346 defined from the literature) were interpreted as potential recurrent reproduction grounds.
347 We compare those inferences with the information of reproduction ground available from
348 the literature (for sole and whiting).

349 Arbault et al. (1986) investigated the reproduction of sole along the Bay of Biscay based
350 on several egg surveys occurring in 1982. Five surveys were conducted between January
351 and May. Egg density was sampled in different locations from Hendaye to Pointe du Raz
352 ($43^{\circ}30'N-48^{\circ}N$) and allowed to map the distribution of egg production on the full study
353 domain. The peak of reproduction occurred in February; thus we compare the maps
354 obtained from the February survey with the persistence index obtained from our model in
355 February.

356 For whiting, only two EVHOE trawl surveys occurred during spring (considered as the
357 reproduction period of whiting) between 1987 and 1992 in the Bay of Biscay (House and
358 Forest 1993). For each haul, the individuals were counted and aged. Individual of two
359 years and older were considered mature. We compare the distribution of mature

360 individuals obtained with these surveys and the index of persistence from our model
361 during spring (March to May).

362 No available information exists regarding the reproduction grounds of squids in this area,
363 however the study from Moreno et al. (2002) investigated the reproduction period for
364 *Loligo vulgaris* in the Eastern Atlantic and highlighted that their reproduction falls in winter
365 and spring with a peak from January to April. We compute the persistence index for this
366 period to identify the spatial aggregation patterns that emerge from the model outputs and
367 that could be considered as spawning grounds.

368 To assess whether the aggregation patterns within the reproduction period are stable over
369 the time period, we iteratively computed the persistence index over a 5-year mobile time-
370 span while pushing forward one year each time.

371 **3 RESULTS**

372 **3.1 Assessing the contribution of each data sources to inference**

373 Results highlight how combining several commercial fleets in the framework brings a
374 better picture of the spatial distribution on the whole domain. For instance, when
375 comparing model predictions with survey data (for the month of the survey), integrating
376 several fleets into the analysis improves correlation with scientific data (Figure 3). It also
377 reduces the standard deviation of the predictions on the full domain (SM 7, Figure S7).

378 When looking at the predictions within the spatial range of the fleets, single-fleet models
379 logically provide similar spatial predictions compared with the integrated model (Figure 4;
380 red dots). However, when using single fleet data, predictions realized outside the spatial
381 range of the fleet largely depart from the ones realized through the integrated models
382 (black dots, Figure 4), emphasizing the contribution of the other fleets to improve
383 inferences on areas poorly covered by single fleet. This is particularly evidenced with the
384 OTB_CEP and OTT_DEF fleets that partially cover the study area compared with
385 OTB_DEF that better cover the whole study area (Figure 1).

386 **3.2 Interpreting estimates PS intensity**

387 Estimates of the PS intensity (b parameters in eq. (3)) for the different species, the
388 different fleets and the different time steps provide information on the targeting behavior
389 that are consistent with expertise. Estimates of b are positive for each species and each
390 fleet (Figure 5, left column). For squids, PS is the strongest for OTB_CEP followed by
391 OTB_DEF and OTT_DEF. This is consistent with the expert knowledge of the targeting
392 behavior of these fleets: OTB_CEP target cephalopods and catch on average 15% of
393 squids while OTB_DEF and OTT_DEF catch respectively 5% and 1% of squid). A similar
394 pattern can be identified for whiting ($b_{OTB_{CEP}} > b_{OTB} > b_{OTT}$); this is consistent with species

395 spatial distribution as whiting (like squids) are found in coastal areas where the OTB_CEP
396 fleet is preferentially operating (Figure 1). For sole, the strength of PS is on average higher
397 for OTB_CEP and OTT_DEF than for OTB_DEF but with less contrast between the three
398 commercial fleets which is also consistent with expertise as those three fleet target this
399 high commercial value species.

400 Interestingly, some of the b parameters time series emphasize seasonal patterns (Figure
401 6, top). For instance, in the sole case study for the OTB_CEP fleet, the b parameters are
402 higher in summer and autumn emphasizing relatively stronger PS, while being lower in
403 winter and early spring (but see section 3.4 below for a more detailed interpretation of this
404 seasonality pattern).

405 **3.3 Evaluating the influence of PS on spatial inferences**

406 Because estimates of b are positive, spatial density of fishing points is positively correlated
407 with biomass density. Then logically, ignoring PS leads to a positive bias (overestimate)
408 in biomass estimates in areas not sampled by the commercial fleets compared to
409 estimates obtained while considering (Figure 5, right column, black points), but does not
410 strongly affect predictions in locations within the range of the fleets (blue points).
411 Considering PS only slightly improves the fit of the model to the data. For the Sole case
412 study, some improvement of the likelihood occurred for both the likelihood associated with
413 the commercial and the scientific data (Table 1). For whiting and squids, there are no
414 strong modifications in both scientific and commercial likelihoods. Overall, accounting or
415 not for PS does not strongly modify the overall pattern of species abundance (SM 8).

416 **3.4 Investigating spatio-temporal dynamics of fish biomass**

417 Results provide biomass density maps on a monthly time step that emphasize seasonal
418 distribution patterns and from which aggregation index were calculated. The temporal

419 correlation parameter (φ) is estimated around 0.8 for all the species emphasizing strong
420 between months temporal correlations in the biomass field values. The range parameters
421 are estimated to 55 km for sole and squids while being estimated to 67 km for whiting
422 emphasizing wider spatial autocorrelation for this species.

423 Concerning the sole case study, model predictions highlight the relatively offshore
424 distribution from November to April and a more coastal distribution from June to October
425 suggesting some offshore-coastal migrations between these 2 periods (Figure 6, bottom).
426 In particular, the migration in June/July is associated with a contraction of the sole
427 distribution around the Vendée coast, the Gironde Estuary and the Landes coast (45.5°N-
428 46°N) while the migration in November leads to an expansion of the species distribution
429 towards the offshore areas all along the Bay of Biscay. Interestingly, such seasonality
430 coincides with the seasonality of PS intensity for the OTB_CEP (Top of Figure 6 and SM
431 9). Higher PS parameter values are associated with a coastal distribution of sole while
432 lower values correspond to offshore distribution of sole.

433 Similar maps can be computed for the other species and are presented in SM10.

434 **3.5 Aggregation index and reproduction grounds**

435 For both sole and squids, areas of aggregation persistent over the years identified during
436 the spawning period exhibit strong aggregation patterns that match the available
437 knowledge of their reproduction grounds. For sole, the aggregation areas globally match
438 with the observed area of maximum egg concentration (Figure 7), although the spawning
439 grounds identified by egg maps are slightly further East of those identified by our method.
440 This slight discrepancy could be interpreted as an effect of the larval drift as the maps
441 provided by Arbault et al. (1986) are concentration of eggs and not reproduction grounds
442 per se. This is consistent with the simulation analysis of Ramzi et al. (2001) showing that

443 the eggs and larval drift in this area of the Biscay Bay is oriented to the East. Overall,
444 these aggregation areas are stable over time (Figure 8).

445 For whiting, similar patterns can be identified during the reproduction period (Figure 7);
446 they match with previous studies investigating the spatial distribution of mature whittings
447 (House and Forest 1993). In particular, the Northern (3°W - 47°N) and the Southern (2°W -
448 45.5°N) aggregation patches are almost systematically significantly considered as
449 aggregation areas (aggregation index equals 1) while the other middle one (2.5°W - 46.5°N)
450 is classified as an aggregation area that appears less frequently over the years. An
451 additional persistent aggregation area can be identified in the North of the Bay of Biscay
452 (4.5°W - 48°N) suggesting that reproduction may also occur in this area which was not
453 identified in the report of House and Forest (1993). Interestingly, the Northern aggregation
454 area (3°W - 47°N) is more intense at the end of the period (Figure 8).

455 For squids, no information related to any reproduction ground exists in the literature, only
456 the time period of the reproduction is known (the peak fall between January to April). On
457 this time period, some persistent aggregation areas can be evidenced in coastal areas
458 (Figure 7) along the Vendée coast (2.5°W - 46.5°N), the Landes coasts (1.5°W - 44°N to
459 45°N) and around Belle-Île-en-Mer (3°W - 47.25°N). Interestingly, the two Northern
460 aggregation areas are more intense at the end of the time series compared to the
461 beginning of the time series (Figure 8).

462 Maps of persistent aggregation areas are available for all month and evidence some other
463 aggregation areas outside of the reproduction period (SM11). For instance, for sole a
464 persistent patch can be identified offshore the Gironde Estuary (1.5°W – 45.5°N) from
465 August to December.

466 **4 DISCUSSION**

467 **Main findings**

468 In this paper, we develop a framework to infer fish spatio-temporal distribution on a
469 monthly time step while combining scientific survey data and commercial catch
470 declarations from several fleets. Commercial catch data constitute a valuable data source
471 that complements scientific survey or onboard sampling programs by providing much
472 higher spatio-temporal sampling density. Those complementary sources of data were
473 integrated through a spatio-temporal hierarchical model taking into account spatio-
474 temporal variation within the biomass field and PS on a monthly time step. We fitted the
475 model to VMS-logbooks data filtered and processed over the period 2010-2018 for 3
476 demersal species (sole, squids and whiting) in the Bay of Biscay.

477 We emphasize the benefit of integrating several spatially complementary fleets to infer
478 fish distribution throughout the year. We demonstrate how the within-year dynamic of the
479 PS parameters can be interpreted with regards to the joint dynamics of species distribution
480 and fishing distribution and to the overall targeting behavior of the fleets (e.g. OTB_CEP
481 for the squids case study). Even though PS parameters are not fishing intention per se
482 (Bourdaud et al. 2019), these could advantageously complement information provided by
483 landing profiles to estimate the targeting behavior of any group of vessels (either
484 métier/fleet or any group that would seem appropriate).

485 Interestingly, although interpretation of the PS parameters provide insight into the spatio-
486 temporal fleet dynamics, accounting for PS in the inferences does not significantly improve
487 model fitting even when some fleets emphasize strong PS (e.g. squids, OTB_CEP). These
488 results contrast with Alglave et al. (2022), and could result from the integration of several

489 fleets in the analysis that allow a full coverage of the area. Indeed, in Alglave et al. (2022),
490 the fleet emphasizing strong PS only covered a restricted (and coastal) part of the area.
491 As introducing PS mainly affects inferences on poorly sampled areas, predictions in the
492 offshore areas where mostly affected. Here, as the fleets are all estimated to have a
493 positive PS and cover the whole area, PS only downscale the predictions in the few areas
494 that remain unsampled.

495 Filtering the mature fraction of the population in both the scientific and the commercial
496 data make possible to infer the spatio-temporal distribution of the fraction of the biomass
497 that could be potentially mature through the year on a monthly time step. We developed
498 an index to infer aggregation areas of the potentially mature fraction of the biomass that
499 are persistent across years. When calculated on a temporal window predefined following
500 the available information on the reproduction period for each species, the aggregation
501 index enables to identify the main recurrent spatial aggregation areas within the
502 reproduction period. Results demonstrate that the recurrent aggregation areas identified
503 from our method for Sole and Whiting were highly consistent with those already identified
504 in the literature. Our results demonstrate how the aggregation index can provide new
505 insights on the spawning grounds for species like squids for which no information on the
506 spawning grounds is available on the literature. Areas of high aggregation persistent
507 across years were identified during the expected period of reproduction and could be
508 interpreted as spawning grounds. This opens perspectives for applying more
509 systematically the approach for species where no information of reproduction grounds is
510 available to fill the gaps in our knowledge with minimum cost (Delage and Le Pape 2016;
511 Regimbart et al. 2018).

512

513

514 **Combining our results with other data sources to refine the identification of**
515 **spawning grounds**

516 Persistent aggregation areas should be considered as potential spawning areas rather
517 than actual spawning areas. Indeed, although the mature fraction of the biomass was
518 filtered in the data, our maps do not directly inform whether individuals are actually
519 reproducing or not. The outputs of the model provide maps of the mature fraction of the
520 population (i.e. the individuals that can reproduce) and not the spawning fraction of the
521 population (i.e. the individuals that are actually spawning). However, by focusing on the
522 temporal window identified as reproduction period in the literature, we limit the risk of
523 misinterpreting the aggregation areas as reproduction areas.

524 Our results can also be used to help gathering additional data to identify reproduction
525 grounds. Typically, our maps could be of great help to design surveys to record eggs,
526 larvae and spawning individuals that would provide direct information of species
527 reproduction (Fox et al. 2008). Because developing such additional surveys would be
528 highly expensive, our maps could provide valuable a priori information to optimize the
529 survey design and potentially find a compromise between the cost, the spatial extent, the
530 temporal coverage of the survey and the accuracy of the expected estimates/predictions.
531 Similar ideas were already applied to the sole case study to investigate more precisely
532 the space-time variation of sole reproduction. Arbault et al. (1986) work provided a priori
533 information of reproduction grounds that allowed designing more localized surveys to
534 study inter- and intra-annual variability of one specific sole spawning area (Petitgas 1997).
535 Several statistical methods have been developed since and are suitable to optimize such
536 adaptive sampling design; see for instance the recent work of Leach et al. (2021).

537 Our results could also be combined with fisher expert knowledge (Yochum et al. 2011) to
538 complement our knowledge of fish reproduction (Delage and Le Pape 2016). For instance,
539 Bezerra et al. (2021) and Silvano et al. (2006) proved the usefulness of fishers knowledge
540 to determine the temporality of fish spawning and to identify some spawning grounds by
541 crossing the information of aggregation areas provided by several fisher. These were
542 proved complementary with scientific data as they can be available at low cost and provide
543 local knowledge of fish ecology.

544 **Limits and perspectives for the approach**

545 Our framework has several limitations that are all material for future research avenues.
546 Our model remains relatively simple with regards to all the temporal processes that
547 actually occur within a fishery. It is both a strength and a weakness: the model remains
548 relatively generic, but one might want to extend it further to account for other temporal and
549 spatio-temporal processes affecting fisheries dynamics. For instance, we opted for a non-
550 seasonal representation of the model. One could make it seasonal by decomposing the
551 intercepts $\alpha_S(t)$ and $\alpha_{X_j}(t)$ as well as the random effects $\delta(x,t)$ and $\eta(x,t)$ into yearly and
552 seasonal terms in addition to some 'season x year' interaction terms as performed in
553 Thorson et al. (2020). In their work, such specification mainly allowed to provide
554 information over the time-steps where data was lacking. In the configuration of our case
555 studies, data is available for all time steps and have a relatively good coverage of the
556 study domain. Consequently, even though it provides a nice conceptual view of
557 seasonality, complexifying our model in that direction should not deeply modify our
558 inference of the biomass field. Alternatively, our framework could integrate orthogonal
559 spatio-temporal terms in the latent field to capture the main mode of variability of the

560 biomass field (Thorson et al. 2020b). Such orthogonal terms would allow to capture the
561 main spatial patterns that structure the latent field as well as their variation in time. These
562 could prove very useful to identify the structuring processes that affect species distribution
563 and could give a valuable insight in the space-time dynamics of the species. Another
564 exciting research avenue would consist in integrating population dynamics in the latent
565 field of biomass (Cao et al. 2020). This would require to refine further the demographic
566 resolution of the VMS-logbooks data (see for instance Azevedo and Silva 2020), but once
567 done, it would give access to huge data for inferring the space-time dynamics of fish
568 populations. Finally, our model considers fisher preferentially sample areas where the
569 biomass is higher (preferential sampling), but does not consider any other drivers and
570 specifically the temporal and spatio-temporal relations that can affect fishers behavior.
571 These can be highly complex and may depend on the distribution of the resource,
572 tradition/habits, management regulations (Abbott et al., 2015; Girardin et al., 2017; Salas
573 and Gaertner, 2004; Hintzen, 2021). These drivers are rarely studied in both space and
574 time (although see Tidd et al., 2015). Our framework could allow to jointly model the
575 dynamics of the species, the distribution of the effort, the link that relates species
576 distribution and effort in space/time and all the other spatial and/or temporal drivers that
577 affect the distribution of fishing effort. For instance, we could relate the fishing intensity to
578 the biomass field from the previous time steps, or alternatively consider that the locational
579 choice depends on the catches of the previous time steps. Adding such covariates and
580 spatio-temporal dependencies in the sampling equation (eq.3) will probably not modify the
581 overall pattern of biomass distribution, but it would make possible to quantify the drivers
582 of fisher behavior and give valuable insight to the fishery dynamics.

583 Including discards would potentially improve our approach. Indeed, logbooks data are
584 landings declarations data which means they inform on the landings and not on the true
585 catch. Thus, by assuming the landings per unit effort are proportional to the biomass, we
586 make the hypothesis that the discard rate is constant in space and time and does not
587 affect model predictions. This should not be a problem for sole and squids as the discards
588 are low and TAC have not been really binding during the studied period. However, the
589 issue might be more stringent for whiting and/or other species with a high and non-
590 stationary level of discards. Integrating discards data in the analysis could help solving
591 this issue. Stock et al. (2019) and Yan et al. (2022) used observer data to model bycatch
592 in both space and time and Breivik et al. (2017) used bycatch data from onboard surveys
593 to predict the temporal evolution of bycatch realized in the full commercial data. Similarly,
594 we could integrate into the same analysis the logbooks and the observer data by assuming
595 that the catch of observer represents the sum of landings (which is also observed in the
596 logbooks data) and discards (which is unobserved in the logbooks data). This way, the
597 discards information available from observer data would be shared with the logbooks data
598 and would allow correcting for the missing portion of catch declarations data while possibly
599 accounting for possible space or time variation in the discard rate.

600 Our analysis rely on the hypothesis that the spawning season is known a priori. Extending
601 the approach to infer the spawning season based on the temporal dynamics of the
602 aggregation patterns could improve our knowledge of species spatio-temporal distribution.
603 In particular, identifying the main species phenomenological phases and their consistency
604 (or shift) in time is crucial in the context of global change (Thorson et al. 2020a). In our
605 study, we computed the aggregation index on a predefined temporal window based on
606 literature assumed to be the reproduction period (Arbault et al. 1986; Houise and Forest

607 1993; Moreno et al. 2002). Several methods exist and could be adapted to extract the
608 spatial patterns that shape model outputs, their related temporal variation and identify the
609 main phenological phases that characterize species distribution (e.g. reproduction,
610 feeding - see for instance Empirical Orthogonal Functions or Principal Oscillation Patterns
611 - Cressie and Wikle 2015; Wikle et al. 2019). While the approach we adopted in the
612 manuscript requires to know the reproduction period of the species and would be
613 inappropriate in a context of a changing reproduction time-span, those alternative
614 methods would not require any a priori. Hence, these methods would be more appropriate
615 to identify phenological modifications in species life cycle in response to climate change.
616 Applying those kind of methods to the huge amount of data available from mandatory
617 declarations, would make possible to track the effect of climate change on fish phenology
618 at a monthly/seasonal scale, while it is generally only possible at a yearly time step
619 through scientific survey data (Maureaud et al. 2020).

620 Last, confidentiality remains a major limitation to the massive use of VMS data (Hintzen
621 2021). Indeed, there are strong confidentiality constraints on these data due to the huge
622 information available on fisher fishing grounds. Few countries are now giving free access
623 to their data (e.g. Norway), but in most cases administrative procedures to get access to
624 the data remain a burden and still constitute a limitation for the use of 'VMS x logbooks'
625 data for routine operational use.

626 **Future use for Marine Spatial Planning**

627 Our model has potential application for Marine Spatial Planning (MSP). Janßen et al.
628 (2018) highlighted that one of the main requirements for implementing MSP is the
629 availability of fine scale information on species distribution and of their essential habitats.
630 Here we propose a method which can provide such information for the fraction of the

631 population available through catch declarations (i.e. mainly the adult fraction and in some
632 cases part of the juvenile fraction). This knowledge is required to design Marine Protected
633 Areas (MPA – see for instance Lambert et al. (2017) or Loisel et al. (2003)), Fishery
634 Conservation Zones (Delage et Le Pape, 2016 ; Regimbart et al., 2018), or alternatively
635 identify areas that should be kept for fishing in a context where many other human
636 activities are competing in space and time with fishing (Campbell et al. 2014; Bastardie et
637 al. 2015). This would require to integrate our results into bio-economic models in order to
638 evaluate alternative management regulations and assess their tradeoffs in regards to all
639 the sets of ecosystem services provided through activities such as fishing, aquaculture,
640 energy, shipping, recreation and conservation (Nielsen et al. 2018).

641

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653 **COMPETING INTERESTS**

654 The authors declare there are no competing interests.

655 **DATA AVAILABILITY STATEMENT**

656 Survey data are available through the DATRAS portal ([https://www.ices.dk/data/data-
657 portals/Pages/DATRAS.aspx](https://www.ices.dk/data/data-
657 portals/Pages/DATRAS.aspx)) with the package 'icesDatras' ([https://cran.r-
658 project.org/web/packages/icesDatras/index.html](https://cran.r-
658 project.org/web/packages/icesDatras/index.html)). Logbooks and VMS data are
659 confidential data and they are available on specific request to DPMA.

660 **CODES**

661 Toy example codes of the model are available on the github link:

662 https://github.com/balglave/sdm_vms_logbooks

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886

887

888 **TABLES**

889

890 Table 1. Ratio between the negative log-likelihood values (either commercial or
 891 scientific) from the integrated model accounting for preferential sampling and the
 892 integrated model ignoring preferential sampling.

Species	Negative log-likelihood ratio	
	Scientific data	Commercial data
Sole	0.97	0.92
Squids	1.00	1.01
Whiting	0.99	1.00

893

Note. The ratio between negative log-likelihoods ($-\log(lkl)$) is given as: $r = \frac{-\log(lkl_{PS})}{-\log(lkl_{noPS})}$.

894

If $r < 1$, the model accounting for PS better fits the data than the model ignoring PS (no PS).

895

896 **FIGURES CAPTION**

897 Figure 1. Spatial distribution of each fleet on the whole period (2010-2018). Unit: fishing
898 effort in fishing hour. Coordinate system: WGS84.

899

900 Figure 2. Diagram of the integrated spatio-temporal model.

901

902 Figure 3. Sole case study. Comparison between the observed scientific CPUE (y-axis)
903 and the corresponding model predictions (x-axis) on the month of the survey, based on
904 model integrating data from one commercial fleet only (either OTB_CEP, OTB_DEF,
905 OTT_DEF) or from all commercial fleets (Integrated model). x-axis: model predictions. y-
906 axis: scientific data observations (CPUE in kg/hour). Black line: linear regression
907 'log(scientific observations) ~ log(model predictions)'. r : Spearman correlation coefficient.
908 Scientific data are integrated to inference for all models. *** stands for the level of
909 significance. Each point is a grid cell in which fall a scientific data point. We compare the
910 scientific observation values to the related prediction values.

911

912 Figure 4. Sole case study. Comparison between predictions (each point is a combination
913 'grid cell x time step'.) from the integrated model (using all fishing fleets) and the model
914 integrating only one commercial fleet for the 12 months of year 2018. Left: OTB_CEP
915 fleet, middle: OTB_DEF fleet, right: OTT_DEF fleet. x-axis: integrated model predictions.
916 y-axis: single-fleet model predictions. The prediction values are log-scaled. Red points:
917 predictions within the sampling area of the related fleets (i.e. the cells sampled by the
918 fleet). Black points: predictions outside the sampling area of the related fleets. Black line:
919 $x = y$ axis. Note that the intercept of the x-y line has been scaled to account for differences
920 in the intercept values between models. Scientific data are integrated to inference for all
921 models.

922

923

924 Figure 5. Estimates of PS parameters for each commercial fleet (left) and effect of PS
925 on model outputs (right). Left: boxplot represent the variability of maximum likelihood
926 estimates of parameters b across all monthly time steps. Right: log-predictions of the
927 integrated model accounting for PS (y-axis) versus log-predictions of the integrated model
928 ignoring PS (x-axis) for the 12 months of year 2018. Blue points: predictions within the
929 sampling area of the commercial fleets (i.e. the cells sampled by commercial fleets). Black
930 point: predictions outside the sampling area of the commercial fleets. Black line: $x = y$
931 axis.

932

933 Figure 6. Sole case study. (Top) Temporal evolution of the b parameters for the three
934 commercial fleets fitted to the integrated model. Blue vertical lines: January. (Bottom)
935 Monthly biomass distribution averaged on the full period. Only quantile values are

936 represented. Model predictions come from the integrated model accounting for PS.
937 Coordinate system: WGS84.

938

939 Figure 7. Left: index of persistence (average over 2010-2018) during the reproduction
940 period of sole (February), whiting (March-May) and squids (January-April). Reproduction
941 period defined from ecological expertise. Right: literature information on reproduction
942 grounds when available. For sole, the map represents egg concentration from an egg and
943 larvae survey conducted in 1982 (Arbault et al., 1986). For whiting, the map represents
944 records of age-2+ whiting (i.e. mature individuals), from two spring trawl surveys that
945 occurred between 1987 and 1992 (House and Forest, 1993). Model predictions come
946 from the integrated model accounting for PS. Coordinate system: WGS84.

947

948 Figure 8. Persistence indices within the reproduction period computed on a 5-years
949 mobile time-span for each 3 species (5-years time span indicated on the top of each map).
950 Model predictions come from the integrated model accounting for PS. Coordinate system:
951 WGS84.

952

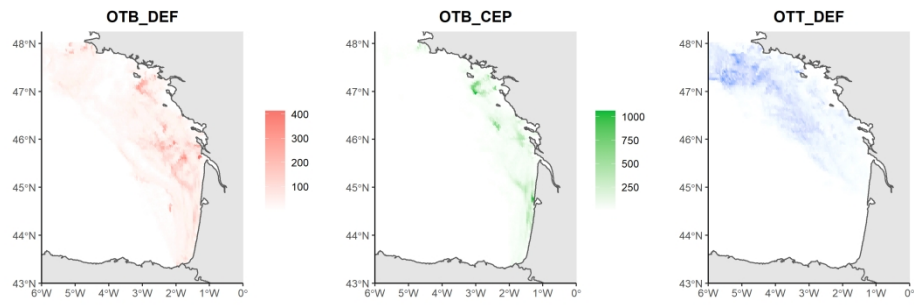


Figure 1

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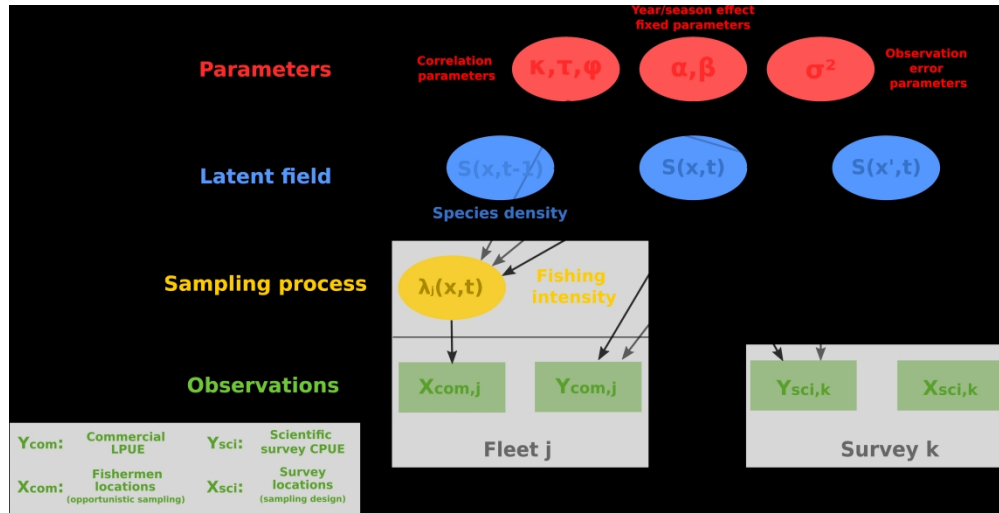


Figure 2

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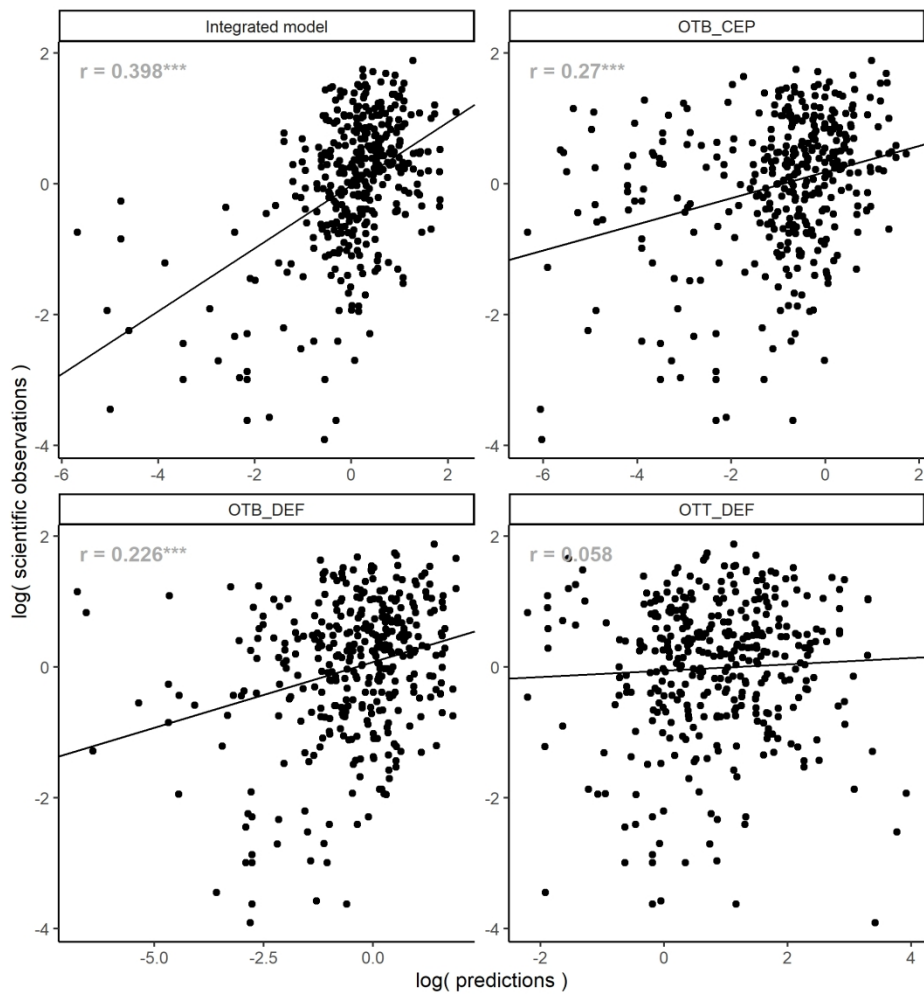


Figure 3

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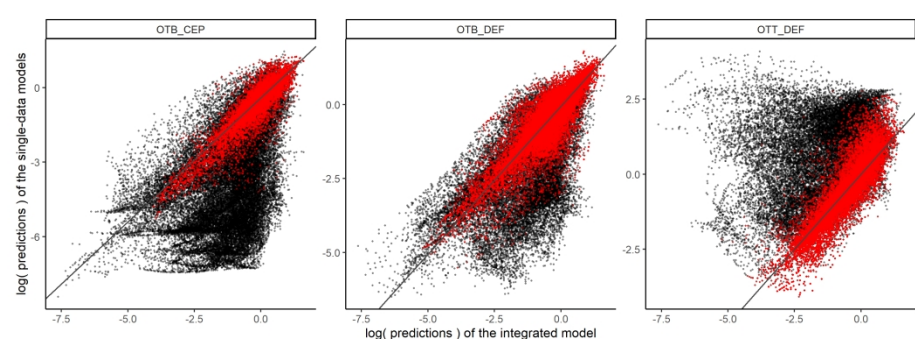


Figure 4

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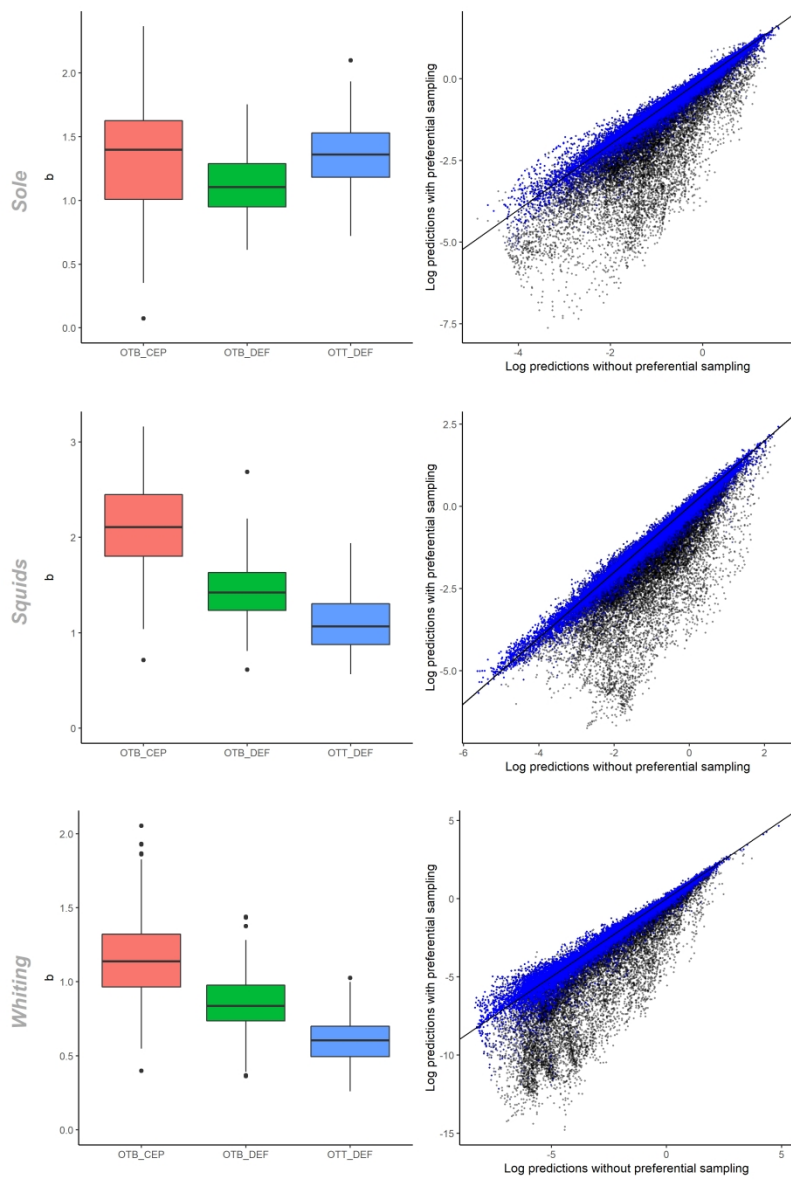


Figure 5

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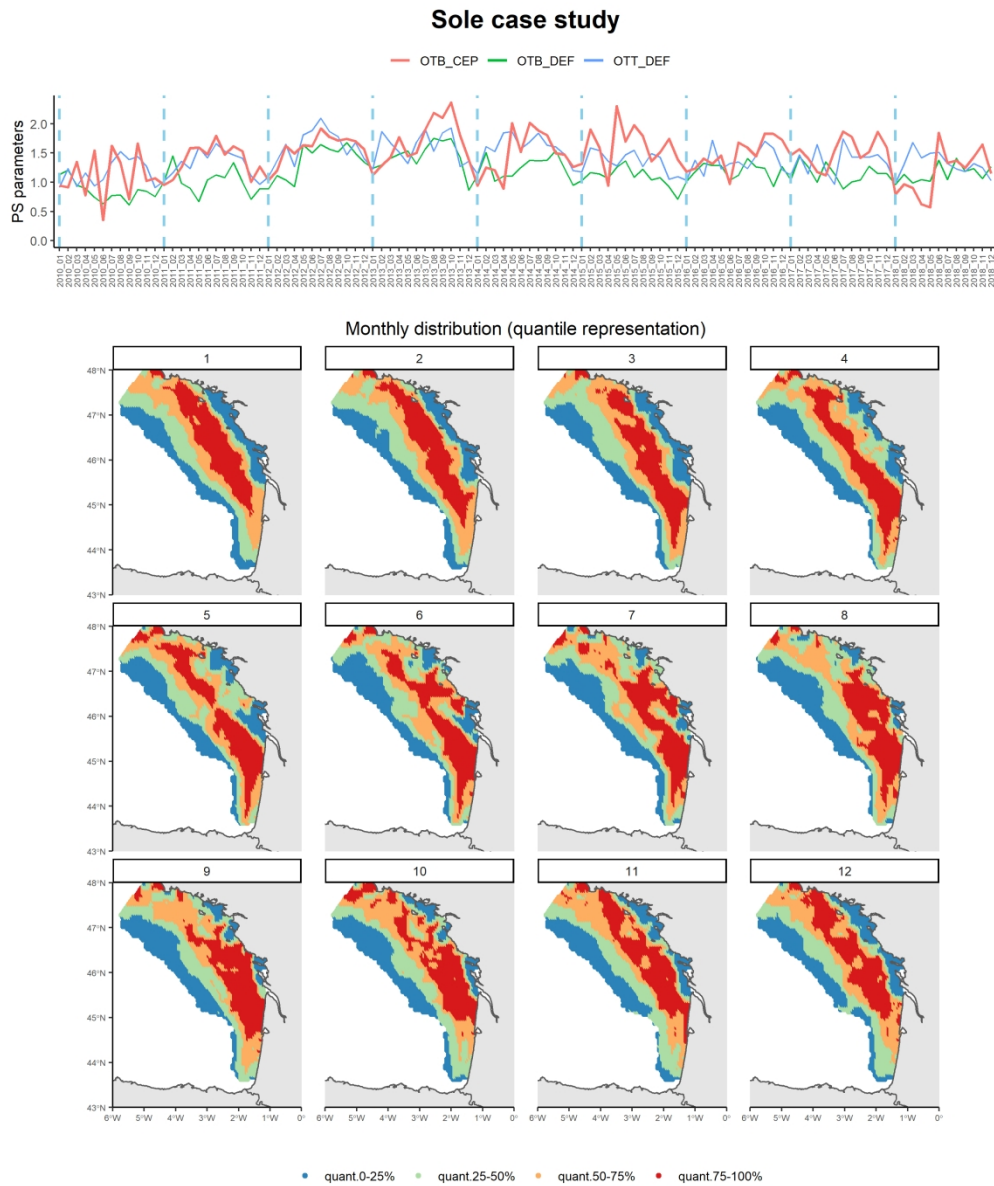


Figure 6

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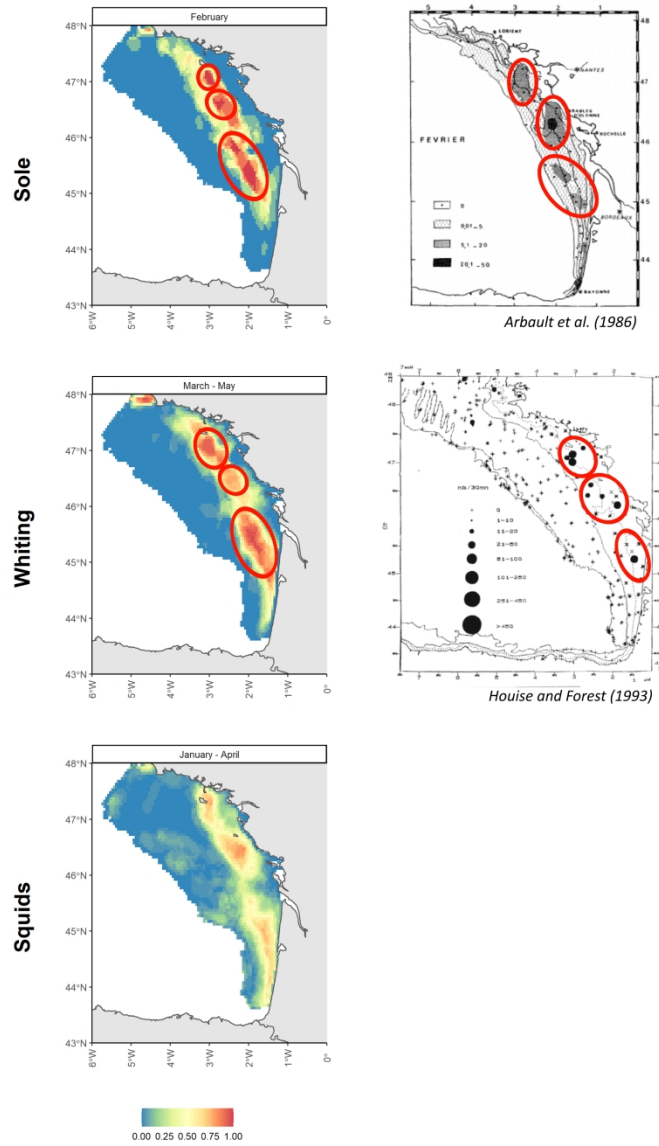


Figure 7

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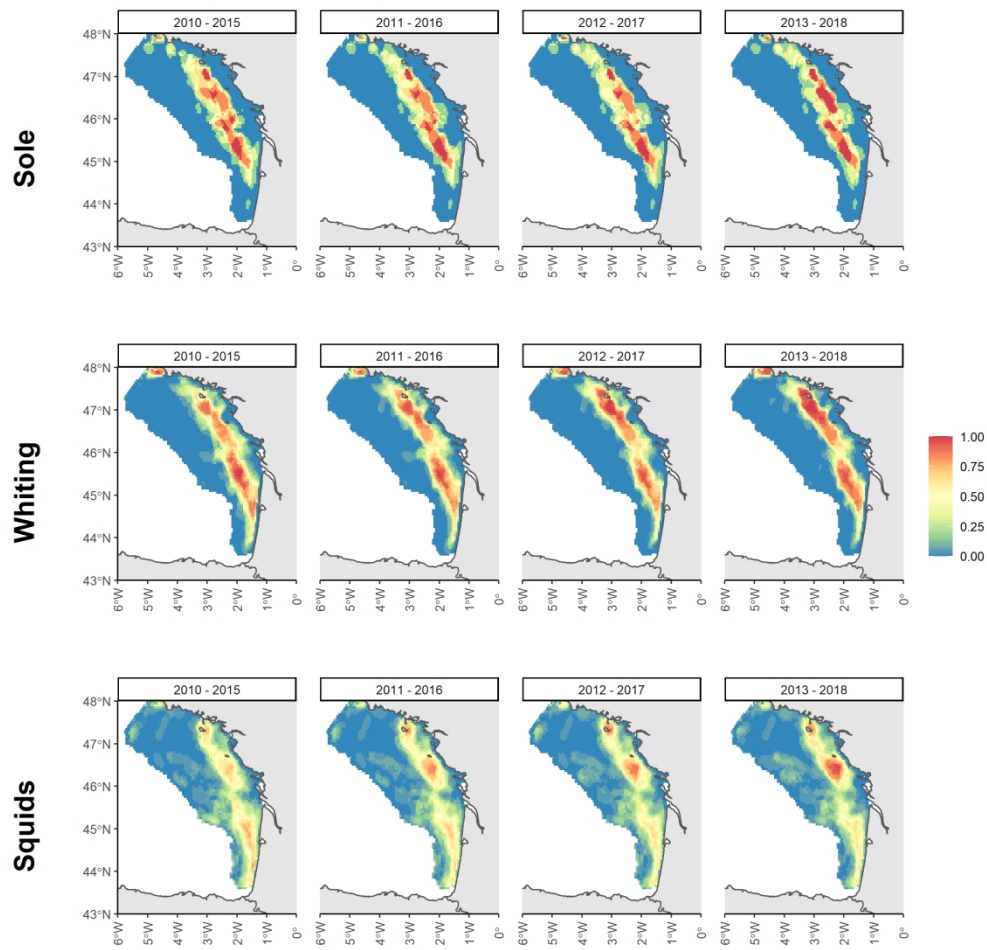


Figure 8

645x645mm (118 x 118 DPI)