



HAL
open science

Lead from Notre-Dame fire plume caught 15 km from Paris

Gaël Le Roux, Olivier Masson, Christophe Cloquet, Marie-José Tavella, Magali Beguin-Leprieur, Olivier Saunier, David Baqué, Thierry Camboulive, Julien Berger, Thierry Aigouy, et al.

► **To cite this version:**

Gaël Le Roux, Olivier Masson, Christophe Cloquet, Marie-José Tavella, Magali Beguin-Leprieur, et al.. Lead from Notre-Dame fire plume caught 15 km from Paris. *ACS Earth and Space Chemistry*, 2023, 7 (2), pp.310-314. 10.1021/acsearthspacechem.2c00321 . hal-04028513

HAL Id: hal-04028513

<https://hal.science/hal-04028513v1>

Submitted on 17 Oct 2023

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

Copyright

1 **Lead from Notre Dame fire plume caught 15 km from Paris**

2

3 Gaël Le Roux¹, Olivier Masson², Christophe Cloquet³, Marie-José Tavella¹, Magali Beguin-Leprieur⁴,

4 Olivier Saunier⁵, David Baqué¹, Thierry Camboulive¹, Julien Berger⁶, Thierry Aigouy⁶, Françoise

5 Maube⁶, Sandrine Baron⁷, Sophie Ayrault⁸, Maxime L’Héritier⁹

6 **1. Laboratoire d’Ecologie Fonctionnelle et Environnement CNRS-Université de Toulouse**

7 Campus Ensat - Avenue de l’Agrobiopole- BP 32607 Auzeville tolosane
8 31326 Castanet-Tolosan, France

9 **2. Institut de Radioprotection et de Sûreté Nucléaire- PSE-ENV/SEREN**

10 Cadarache BP3 13115 Saint-Paul-Lez-Durance Cedex, France

11 **3. CRPG, UMR 7358, Université de Lorraine-CNRS, Nancy**

12 15 rue Notre Dame des Pauvres, 54500 Vandœuvre les Nancy – France

13 **4. Institut de Radioprotection et de Sûreté Nucléaire- IRSN/PSE-ENV/SIRSE**

14 31 rue de l’écluse. BP 40035 78116 Le Vésinet Cedex, France

15 **5. Institut de Radioprotection et de Sûreté Nucléaire -PSE-SANTÉ/SESUC BMCA**

16 31 Av. de la Division Leclerc, 92260 Fontenay-aux-Roses, France

17 **6. GET-OMP CNRS-IRD-Université de Toulouse**

18 14 Av. Edouard Belin, 31400 Toulouse, France

19 **7. TRACES, CNRS UMR 5608, Université de Toulouse**

20 Université Toulouse Jean Jaurès, Maison de la Recherche

21 5, allées Antonio Machado F 31058 Toulouse cedex 9, France

22 **8. Laboratoire des Sciences du Climat et de l’Environnement (LSCE-IPSL), UMR 8212 (CEA/CNRS/UVSQ), Université**
23 **Paris-Saclay, Gif-sur-Yvette, France**

24 CEA Saclay, Bat 714, Site de l’Orme des Merisiers

25 Chemin de Saint Aubin - RD 128 F-91191 Gif sur Yvette Cedex, France

26 **9. Université Paris 8, ArScAn UMR 7041.**

27 Maison de l’Archéologie et de l’Ethnologie

28 21 allée de l’Université

29 F-92023 Nanterre Cedex, France

30

31

32 Corresponding author :

33 Gaël Le Roux

34 Tel : 33 (0) 5.34.32.37.79

35 fax : 33 (0) 5.34.32.39.01

36

37 **Abstract**

38 Thanks to the use of filters collected daily in the west of Paris, we confirm the passage of the lead-
39 laden plume following the fire on the roof and spire of Notre Dame Cathedral in April 2019. The
40 measured concentrations on the filter ($Pb= 1.4 \mu g m^{-3}$), scanning electron microscopy and Hysplit
41 simulation correspond to an estimate of a few hundred kilograms of lead that would have been
42 volatilized and then oxidized in the form of micronic and submicronic aerosols. The concentrations
43 found in the plume are, however, much lower than those found in the environment in the 1980s and
44 1990s but are 100 times higher than those found in times prior and after the fire. The isotopic
45 signature of the plume is almost identical to that of the fine Pb dust found inside Notre Dame
46 building. It is different from the isotopic signature of the Parisian atmosphere before and after the
47 fire, but it is similar to that of the atmospheric Pb legacy recorded by peat cores over the last 300
48 years in France. The presence of very fine lead-bearing particles makes them potentially
49 transportable over long distances after large urban fires. Our study shows the value of daily aerosol
50 sampling to retrospectively trace the plumes of air pollutants from industrial accidents but also from
51 historical monument fires such as Notre Dame in 2019.

52

53 **Keywords:**

54

55 Historical monument fire; Notre-Dame de Paris; lead plume; aerosols; Pb isotopes; air monitoring;
56 urban fire

57

59 Introduction

60 Lead is a chemical element that is not essential and is harmful for living organisms, but which has been
61 used by humans since the beginning of metallurgy.¹ Its global dispersion as a result of the industrial
62 era, and in particular the use of leaded gasoline, was highlighted by the work of Clair C. Patterson, who
63 showed that this chemical element is present in all compartments of the environment: from the ocean
64 abyss to the poles and at the highest mountains on the Earth.^{2,3} Lead is a by-product of silver mining
65 in galena (PbS), is common in many ores as an oxide or carbonate and easy to work. It has been used
66 in construction since Antiquity.¹ Many major European churches used lead roofs as tiles and lead
67 roofing is still commonly used. Notre Dame cathedral was no exception and was built with 240 metric
68 tons of Pb for its roof and spire.⁴ Lead has been used in the construction of the cathedral since the
69 12th-13th centuries, and was still largely used in the 19th century for the renovation of the roof and
70 reconstruction of the spire by Viollet le Duc. The Notre Dame fire of 15 April 2019 destroyed a large
71 part of the upper parts of the building, attacking the roof and the spire. The temperatures reached are
72 still debated by the scientific community but the melting temperature of Pb (327.5°C) was undoubtedly
73 reached. It is likely that in some places, the temperatures exceeded 600°C and Pb oxidized into an
74 aerosol. The Parisian air monitoring network (“AirParif”) is no longer monitoring Pb following the large
75 decrease in Pb concentrations after the ban on leaded gasoline. Only a single weekly AirParif filter
76 located under the plume in the west along the Seine River showed a significant increase in airborne Pb
77 concentrations.⁵ In addition, since the aerosols are only sampled on a weekly basis this leads to a large
78 uncertainty in the concentration of the plume during its passage. To complete this single measurement
79 point, we benefit from the OPERA network, a network of atmospheric air collectors developed by the
80 IRSN (Institut de Radioprotection et Sûreté Nucléaire) to monitor radioactivity in the environment.⁶
81 More specifically, samples from these high flow atmospheric collectors allow us to follow the plume
82 over a greater distance and coupled with data from a 24-hour aerosol sample from a location 15 km
83 from the fire, to determine the Pb isotopic composition and chemical form of the aerosolized Pb.

85 Material and Methods

86 Sampling strategy

87 We assessed the trajectory of the Notre Dame fire plume using the online Hysplit Lagrangian
88 atmospheric model⁷ (Figure S1). Based on the plume trajectory, we requested permission from the
89 French Institute of Radioprotection and Nuclear Safety (IRSN) to use weekly filters from the OPERA
90 observatory collected before and after the fire from North West of France as well as at the Pic du Midi
91 in the Pyrenees for background reference. Since IRSN has also a campus in Le Vésinet, the laboratory
92 collects air filters on a daily basis at this location, which makes it exceptional because the collector was
93 under the plume trajectory and less than 15 km from the Ile de la Cité where Notre Dame is located.
94 Few data on lead concentrations in the air exist since their sharp decline following the ban on leaded
95 petrol. However, the Limay station monitored by Airparif (weekly filter) also recorded the passage of
96 the plume.⁸

97 Geochemical and mineralogical analyses

98 The filters (polypropylene filter (3 MTM) made of several layers were split and sub-samples were taken
99 before digestion using ultrapure HNO₃ in Digtubes (SCP Science) on a hot plate at 100°C in a class
100 10000 clean room at Toulouse. Samples were then centrifuged and an aliquot was analyzed using a
101 AMETEK SPECTRO ARCOS FHX22 ICP-OES. Three different layers of the Vésinet filters were analyzed
102 and confirms that aerosols and trace metals were homogeneously distributed over the whole filter.

103 For Pb isotope analysis, after evaporation, samples were diluted by passage of 1 ml of HBr 0.8M. The
104 solution was then loaded onto 0.05ml of AG1X8 resin contained in micro-column (Bio-spin column)
105 form Bio-Rad. The matrix is eluted using 2 ml (4*0.5 ml) of HBr 0.8M. Lead was then recovered into 2
106 ml HCl 6.0 M⁹. Samples were dried down and subsequently recovered in 1 ml 0.5 ml HNO₃ 0.3 M.
107 Samples were then analyzed with MC-ICP-MS (Neptune + from Thermo). After dilution to adjust the
108 signal to the standard (NIST 981), Tl (NIST 997) was added to the samples (similarly to the standard) to
109 a ratio 1/10 compared to Pb and used to correct for mass bias (White et al. 2000). The Tl values used
110 to correct were those obtained by DS TIMS¹⁰. The total procedural blank (<500 pg) was negligible in
111 front of the Pb introduced by the samples.

112

113 Le Vésinet filters were placed on an aluminum stub and covered by a thin (~20 nm) carbon layer. They
114 were analyzed with a JEOL JSM 6360LV scanning electron microscope (SEM) set to an accelerating
115 voltage of 20 kV. The Pb-bearing particles were localized and imaged using the backscattered and

116 secondary electron detectors and the presence of Pb was determined by energy dispersive X-Ray
117 spectroscopy (using a Bruker Quantax 10 silicon drift detector).

118

119 Hysplit modelling

120

121 As an example, we performed different simulations using the offline Lagrangian Hysplit model¹¹ with
122 a potential atmospheric release of Pb from the Notre Dame fire. In a first simulation, we first ran the
123 model with an emission rate of 40kg/h of Pb for 3 hours. These conditions are relatively similar to a
124 more extensive modelling by INERIS estimating that 63 kg and 75 kg of lead were emitted into the
125 atmosphere¹² respectively from the roof and the spire. Based on our scanning electron microscope
126 results, we considered that Pb can be aerosolized in a similar manner as as ¹³⁷Cs, that is to say a part
127 of Pb was aerosolized by high temperature processes and is of micronic in size and we used thus the
128 initial deposition parameters of Hysplit model.¹¹ We thus then simulated a higher release of Pb of the
129 order of 125 kg/hour for 4 hours at 50 m altitude. Our simulations are not intended to be totally
130 realistic but to show the potential synergy between our geochemical and mineralogical data and future
131 more sophisticated atmospheric modelling.

132

133 Results and Discussion

134

135 Only the filters located near Paris (Le Vésinet and Limay) show an unambiguous increase in Pb
136 concentrations following the Notre Dame fire (Figure 1 & Supplementary excel spreadsheet). Le
137 Vésinet filter was also characterized for copper and iron, and those metals do not show any increase
138 showing the uniqueness of the plume with mainly Pb as a constituent. From electron microscope
139 imaging, lead is found in Pb oxide aerosols in spherical form with a size of 5-10µm as well as smaller
140 particles (~1µm) whose shape is more difficult to distinguish due to entanglement in the filter (Figure
141 2). Daily Le Vesinet Pb concentrations before and after Notre Dame fire are about 0.01 µg m⁻³ whereas
142 daily concentrations of Pb at le Vesinet in the filter following Notre Dame fire is 1.4 µg m⁻³. No lead
143 particles were found in the Le Vésinet filters before and after the fire. The daily filter sampled at Le
144 Vésinet shows a higher concentration than the weekly filter in Limay (0.108 µg m⁻³) due to the dilution
145 effect from non-contaminated air over the sampling period. According to the Hysplit model, the plume
146 passed over Le Vésinet and Limay in less than 5 hours. If we consider that Pb collected in the filters
147 comes exclusively from the plume, it is possible to cross multiply and estimate the Pb volume
148 concentrations during the 5 hours passage of the plume, respectively 6.72 µg m⁻³ and 3.5 µg m⁻³ for Le

149 Vésinet and Limay. Hysplit model shows also relative higher Pb concentrations at Le Vésinet compared
150 to Limay, which is located 50 km from Paris downtown (Figure S2**Erreur ! Source du renvoi**
151 **introuvable.**).

152 The lead concentration, even when adjusted for the duration of the plume, has a similar magnitude to
153 that found in the 1990s in France when leaded petrol was used.¹³ In 1991, the concentration in the
154 vicinity of the site in heavy traffic was $0.7\mu\text{g m}^{-3}$ and decreased to $0.15\mu\text{g m}^{-3}$ in 1999.¹⁴ However, the
155 concentrations are 20,000 times higher than the lead concentration measured at our reference site at
156 Pic du Midi and 100 times more than in Le Vésinet on the day before and the day after. Our work also
157 shows the value of filters taken on a daily basis that can be made available to the scientific and health
158 community following an accident. Only this filter sampling interval can truly validate the extent of
159 particulate air pollution from such incidents, based on Pb concentrations.

160 The other weekly filters to the west and north-west do not show an increase in lead although
161 potentially under the passage of the more dilute plume. In contrast, in the North-East, at Charleville-
162 Mézières, Pb concentration is also high ($0.0847\mu\text{g m}^{-3}$), as are other metals (Cu and Fe), which
163 highlights the importance of monitoring of air pollution composition over over the long term
164 complementary to PM10 and PM2.5. In general, lead concentrations in the air in Northern France
165 appear to be higher than in Normandy (Figure S5), which is logical given the Atlantic influence on the
166 Normandy sites and the mining and industrial heritage of Northern France. Pb concentrations
167 measured in Western France (Penly, Paluel) are in agreement with Pb concentrations measured in
168 Normandy in 2019 (info AirNormandie- Figure S5) and by the EMEP network in France where
169 concentrations in 2019 from the MERA-EMEP network (6 locations-<https://ebas-data.nilu.no/>) ranged
170 from $4.10^{-4}\mu\text{m}^{-3}$ to $7.10^{-3}\mu\text{g m}^{-3}$ in PM10 in France on a bi-weekly network. Little is known about fires
171 in lead-involved buildings^{15,16} and our study has shown the influence of a building fire on air quality at
172 a distance of 15 km away. We also assumed that lead-laden particles could travel further than assumed
173 as shown by atmospheric modelling and thus potentially influence atmospheric records such as ice
174 cores.¹⁷

175 A powerful tracer for identifying the origin of lead is Pb isotope composition, which can generally
176 discriminate between natural and anthropogenic sources of lead. Here, the isotopic signature
177 measured at Le Vésinet ($^{206}\text{Pb}/^{207}\text{Pb} = 1.1697\pm 1.10^{-5}$; $^{208}\text{Pb}/^{206}\text{Pb} = 2.0979\pm 1.10^{-4}$) during the passage
178 of the plume is clearly distinct from the signatures of the filters before and after the fire (Figure 3). It
179 is also distinct from the signatures at other air monitoring stations (Figure 3).

180 The isotopic signature at the site is variable depending on the lead objects or molten or emitted lead
181 collected at the site. The first study to look at the isotopic signature of lead was based on single

182 collector mass spectrometry measurements which did not allow for a very accurate measurement of
183 lead-204.¹⁸ A second, more recent study¹⁹ succeeded in determining the isotopic signature of lead
184 with a more precise selection of dusts inside the building and a MC-ICP-MS measurement protocol
185 strictly identical to ours. This selection of indoor dust seems appropriate since the isotopic signature
186 of lead in the plume is in particular similar to that measured in sample “ND140619-06 (I)”, made of
187 fine dust collected on the wood of the Notre Dame organ. Briard et al.¹⁹ showed that the isotopic
188 signature was distinct from the legacy lead stored in different compartments in Paris. However, put in
189 a more general spatial and historic context, the signature of Notre Dame fire plume is relatively
190 indistinguishable from the recent isotopic signature trend of the air in France as determined by
191 atmospheric environmental archives such as *Sphagnum* bogs (Figure 3). We used the isotope ratios of
192 the most abundant isotopes to lead-204 to try to distinguish a potential influence of the Notre Dame
193 plume on the more distant filters. The Maubeuge and Villeneuve d'Ascq filters collected during the
194 plume passage are isotopically different from the filters collected after the passage. Their signature
195 seems to tend towards that of the plume. However, the Charleville Mézières filter also has a close
196 signature and this is characterised by an enrichment in iron and copper, not coming from the plume of
197 the Notre Dame fire. It is therefore difficult to demonstrate the influence of the plume on the isotopy
198 of the more distant filters, but it cannot be excluded, particularly for the Normandy (Paluel and Penly)
199 and Villeneuve d'Ascq filters.

200 A previous study has suggested lead contamination of Parisian honeys from the Notre Dame fire.²⁰
201 2019 Honey collected downwind of the fire has high Pb concentrations compared to other honeys from
202 central Paris and Paris before the fire. However, the isotopic signature of contaminated honey is not
203 unique and sufficiently distinct. It is possible that Parisian bees were subjected to contamination from
204 larger local particles such as those measured by Glorennec et al.¹⁸ in the vicinity of the site rather than
205 from lead present in the form of very fine aerosols that were transported over several tens of
206 kilometers as shown in our study. Another possibility is that the lead contamination of honey in
207 western Paris in 2019 is coming from other sources. Multi-year monitoring of honey and other
208 bioindicators could be useful to assess their effectiveness in monitoring acute contamination. Our SEM
209 analyses show that the lead-laden particles are at the micron scale in size (Figure 2) showing also the
210 intensity and temperature reached to probably render lead emission in gaseous forms and then during
211 transport in the plume integrate it as fine Pb oxide aerosols (Figure 2). The micron size confirms the
212 potential for transport of lead-laden particles from a building fire via air masses over long distances.¹⁷

213

214 Conclusions and perspectives

215

216 Our results show the interest of a sampling network of daily filters to better understand the plumes of
217 particulate pollutants in cities or industrial sites. These daily filters make it possible for a more accurate
218 determination of the intensity of atmospheric contamination and to validate atmospheric models. It is
219 possible to carry out this monitoring by using devices that compile several daily filters into weekly or
220 monthly filters in order to carry out the regulatory monitoring while allowing in case of an unexpected
221 event a return to the high-frequency samples. This strategy already exists for radionuclide monitoring
222 and is implemented under the Comprehensive Nuclear-Test-Ban Treaty. Some metals, including lead,
223 can also be monitored continuously by XRF detectors coupled with particle analysers. It seems
224 essential to us to densify the network of these two types of monitoring to allow retrospective
225 modelling of atmospheric pollutants.

226 The legacy of lead in major European cities is still a clear environmental threat. Building fires are a
227 potential additional source of lead to be considered in urban environmental risks.

228

229 Supporting information

230 Details on hysplit simulation; Hysplit plume trajectory and filter locations (Figure S1); hourly Pb
231 concentrations modelled with Hysplit at Le Vésinet and Limay (Figure S2); Filter quality control
232 (Figure S3 and Table S1); Pb isotopes diagram with Fe and Cu concentrations (Figure S4); 2019 Pb air
233 concentrations from french air monitoring network (Figure S5); Pb concentrations and Pb isotopes
234 data (xls spreadsheet S1)

235 Acknowledgements

236 This research was funded by the CNRS chantier Notre Dame de Paris. We thank the 3 reviewers for
237 their extremely constructive comments.

238

239 References

- 240 (1) Nriagu, J. O. *Lead and Lead Poisoning in Antiquity*; p.487, John Wiley: New York and Chichester,
241 1983.
- 242 (2) Settle, D. M.; Patterson, C. C. Lead in Albacore: Guide to Lead Pollution in Americans. *Science*
243 **1980**, *207* (4436), 1167–1176. <https://doi.org/10.1126/science.6986654>.
- 244 (3) De Vera, J.; Chandan, P.; Pinedo-González, P.; John, S. G.; Jackson, S. L.; Cullen, J. T.; Colombo,
245 M.; Orians, K. J.; Bergquist, B. A. Anthropogenic Lead Pervasive in Canadian Arctic Seawater.
246 *Proceedings of the National Academy of Sciences* **2021**, *118* (24), e2100023118.
247 <https://doi.org/10.1073/pnas.2100023118>.
- 248 (4) Syvilay, D.; Azema, A.; Coste, R.; Guillaume, S.; L'Héritier, M. Casting, Joining and Recycling. Use
249 of Lead and Plumbers Practices in Notre-Dame Building Yard (12th-19th c.) through Historical,
250 Archaeological and Chemical Investigations. *Journal of Cultural Heritage*. sous presse.

- 251 (5) AIRPARIF. *Surveillance de l'air suite à l'incendie de Notre-Dame #3 | Airparif*.
252 <https://www.airparif.asso.fr>. <https://www.airparif.asso.fr/actualite/2019/surveillance-de-lair->
253 [suite-lincendie-de-notre-dame-3](https://www.airparif.asso.fr/actualite/2019/surveillance-de-lair-suite-lincendie-de-notre-dame-3) (accessed 2022-12-19).
- 254 (6) Masson, O.; Romanenko, O.; Saunier, O.; Kirieiev, S.; Protsak, V.; Laptev, G.; Voitsekhovych, O.;
255 Durand, V.; Coppin, F.; Steinhäuser, G.; de Vismes Ott, A.; Renaud, P.; Didier, D.; Boulet, B.;
256 Morin, M.; Hýža, M.; Camps, J.; Belyaeva, O.; Dalheimer, A.; Eleftheriadis, K.; Gascó-Leonarte,
257 C.; Ioannidou, A.; Isajenko, K.; Karhunen, T.; Kastlander, J.; Katzlberger, C.; Kierepko, R.;
258 Knetsch, G.-J.; Kónyi, J. K.; Mietelski, J. W.; Mirsch, M.; Møller, B.; Nikolić, J. K.; Povinec, P. P.;
259 Rusconi, R.; Samsonov, V.; Sýkora, I.; Simion, E.; Steinmann, P.; Stoulos, S.; Suarez-Navarro, J.
260 A.; Wershofen, H.; Zapata-García, D.; Zorko, B. Europe-Wide Atmospheric Radionuclide
261 Dispersion by Unprecedented Wildfires in the Chernobyl Exclusion Zone, April 2020. *Environ.*
262 *Sci. Technol.* **2021**, *55* (20), 13834–13848. <https://doi.org/10.1021/acs.est.1c03314>.
- 263 (7) Air Resources Laboratory. Hysplit on the READY Website.
264 <https://www.ready.noaa.gov/HYSPLIT.php>.
- 265 (8) van Geen, A.; Yao, Y.; Ellis, T.; Gelman, A. Fallout of Lead Over Paris From the 2019 Notre-Dame
266 Cathedral Fire. *GeoHealth* **2020**, *4* (8), e2020GH000279.
267 <https://doi.org/10.1029/2020GH000279>.
- 268 (9) Manhès, G.; Allègre, C. J.; Dupré, B.; Hamelin, B. Lead Isotope Study of Basic-Ultrabasic Layered
269 Complexes: Speculations about the Age of the Earth and Primitive Mantle Characteristics. *Earth*
270 *and Planetary Science Letters* **1980**, *47* (3), 370–382. <https://doi.org/10.1016/0012->
271 [821X\(80\)90024-2](https://doi.org/10.1016/0012-821X(80)90024-2).
- 272 (10) Thirlwall, M. F. Multicollector ICP-MS Analysis of Pb Isotopes Using a 207pb-204pb Double
273 Spike Demonstrates up to 400 Ppm/Amu Systematic Errors in Tl-Normalization. *Chemical*
274 *Geology* **2002**, *184* (3), 255–279. [https://doi.org/10.1016/S0009-2541\(01\)00365-5](https://doi.org/10.1016/S0009-2541(01)00365-5).
- 275 (11) Stein, A. F.; Draxler, R. R.; Rolph, G. D.; Stunder, B. J. B.; Cohen, M. D.; Ngan, F. NOAA's HYSPLIT
276 Atmospheric Transport and Dispersion Modeling System. *Bulletin of the American*
277 *Meteorological Society* **2015**, *96* (12), 2059–2077. <https://doi.org/10.1175/BAMS-D-14->
278 [00110.1](https://doi.org/10.1175/BAMS-D-14-00110.1).
- 279 (12) Tognet, F.; Truchot, B. *Modélisation de la dispersion des particules de plomb du panache de*
280 *l'incendie de Notre Dame - Rapport Technique | Ineris*; rapport technique 200480-879062-v2.0;
281 INERIS; p 29. <https://www.ineris.fr/fr/modelisation-dispersion-particules-plomb-panache->
282 [incendie-dame-rapport-technique](https://www.ineris.fr/fr/modelisation-dispersion-particules-plomb-panache-incendie-dame-rapport-technique) (accessed 2022-05-19).
- 283 (13) Monna, F.; Lancelot, J.; Croudace, I. W.; Cundy, A. B.; Lewis, J. T. Pb Isotopic Composition of
284 Airborne Particulate Material from France and the Southern United Kingdom: Implications for
285 Pb Pollution Sources in Urban Areas. *Environ. Sci. Technol.* **1997**, *31* (8), 2277–2286.
286 <https://doi.org/10.1021/es960870+>.
- 287 (14) AIRPARIF. Synthèse des connaissances sur les particules en suspension dans l'air et des travaux
288 d'Airparif sur ces polluants. **2008**.
- 289 (15) McNamee, M.; Marlair, G.; Truchot, B. *Research Roadmap: Environmental Impact of Fires in the*
290 *Built Environment*; Fire Protection Research Foundation, 2020; p 81.
- 291 (16) Ferreira, T. M. Notre Dame Cathedral: Another Case in a Growing List of Heritage Landmarks
292 Destroyed by Fire. *Fire* **2019**, *2* (2), 20. <https://doi.org/10.3390/fire2020020>.
- 293 (17) Le Roux, G.; De Vleeschouwer, F.; Weiss, D.; Masson, O.; Pinelli, E.; Shotyk, W. Learning from
294 the Past: Fires, Architecture, and Environmental Lead Emissions. *Environ. Sci. Technol.* **2019**, *53*
295 (15), 8482–8484. <https://doi.org/10.1021/acs.est.9b03869>.
- 296 (18) Glorennec, P.; Azema, A.; Durand, S.; Ayrault, S.; Le Bot, B. The Isotopic Signature of Lead
297 Emanations during the Fire at Notre Dame Cathedral in Paris, France. *International Journal of*
298 *Environmental Research and Public Health* **2021**, *18* (10), 5420.
299 <https://doi.org/10.3390/ijerph18105420>.
- 300 (19) Briard, J.; Ayrault, S.; Roy-Barman, M.; Bordier, L.; L'Héritier, M.; Azéma, A.; Syvilay, D.; Baron,
301 S. Determining the Geochemical Fingerprint of the Lead Fallout from the Notre-Dame de Paris

- 302 Fire: Lessons for a Better Discrimination of Chemical Signatures. *Science of The Total*
303 *Environment* **2023**, *864*, 160676. <https://doi.org/10.1016/j.scitotenv.2022.160676>.
- 304 (20) Smith, K. E.; Weis, D.; Chauvel, C.; Moulin, S. Honey Maps the Pb Fallout from the 2019 Fire at
305 Notre-Dame Cathedral, Paris: A Geochemical Perspective. *Environ. Sci. Technol. Lett.* **2020**,
306 *acs.estlett.0c00485*. <https://doi.org/10.1021/acs.estlett.0c00485>.
- 307 (21) Hansson, S. V.; Claustres, A.; Probst, A.; De Vleeschouwer, F.; Baron, S.; Galop, D.; Mazier, F.; Le
308 Roux, G. Atmospheric and Terrigenous Metal Accumulation over 3000 Years in a French
309 Mountain Catchment: Local vs Distal Influences. *Anthropocene* **2017**, *19*, 45–54.
310 <https://doi.org/10.1016/j.ancene.2017.09.002>.
- 311 (22) Forel, B.; Monna, F.; Petit, C.; Bruguier, O.; Losno, R.; Fluck, P.; Begeot, C.; Richard, H.; Bichet,
312 V.; Chateau, C. Historical Mining and Smelting in the Vosges Mountains (France) Recorded in
313 Two Ombrotrophic Peat Bogs. *Journal of Geochemical Exploration* **2010**, *107* (1), 9–20.
314 <https://doi.org/10.1016/j.gexplo.2010.05.004>.
- 315 (23) Shotyk, W.; Weiss, D.; Appleby, P. G.; Cheburkin, A. K.; Frei, R.; Gloor, M.; Kramers, J. D.; Reese,
316 S.; Van Der Knaap, W. O. History of Atmospheric Lead Deposition Since 12,370 14C Yr BP from a
317 Peat Bog, Jura Mountains, Switzerland. *Science* **1998**, *281* (5383), 1635–1640.
318 <https://doi.org/10.1126/science.281.5383.1635>.
- 319 (24) Allan, M.; Le Roux, G.; De Vleeschouwer, F.; Bindler, R.; Blaauw, M.; Piotrowska, N.; Sikorski, J.;
320 Fagel, N. High-Resolution Reconstruction of Atmospheric Deposition of Trace Metals and
321 Metalloids since AD 1400 Recorded by Ombrotrophic Peat Cores in Hautes-Fagnes, Belgium.
322 *Environmental Pollution* **2013**, *178*, 381–394. <https://doi.org/10.1016/j.envpol.2013.03.018>.
- 323

324

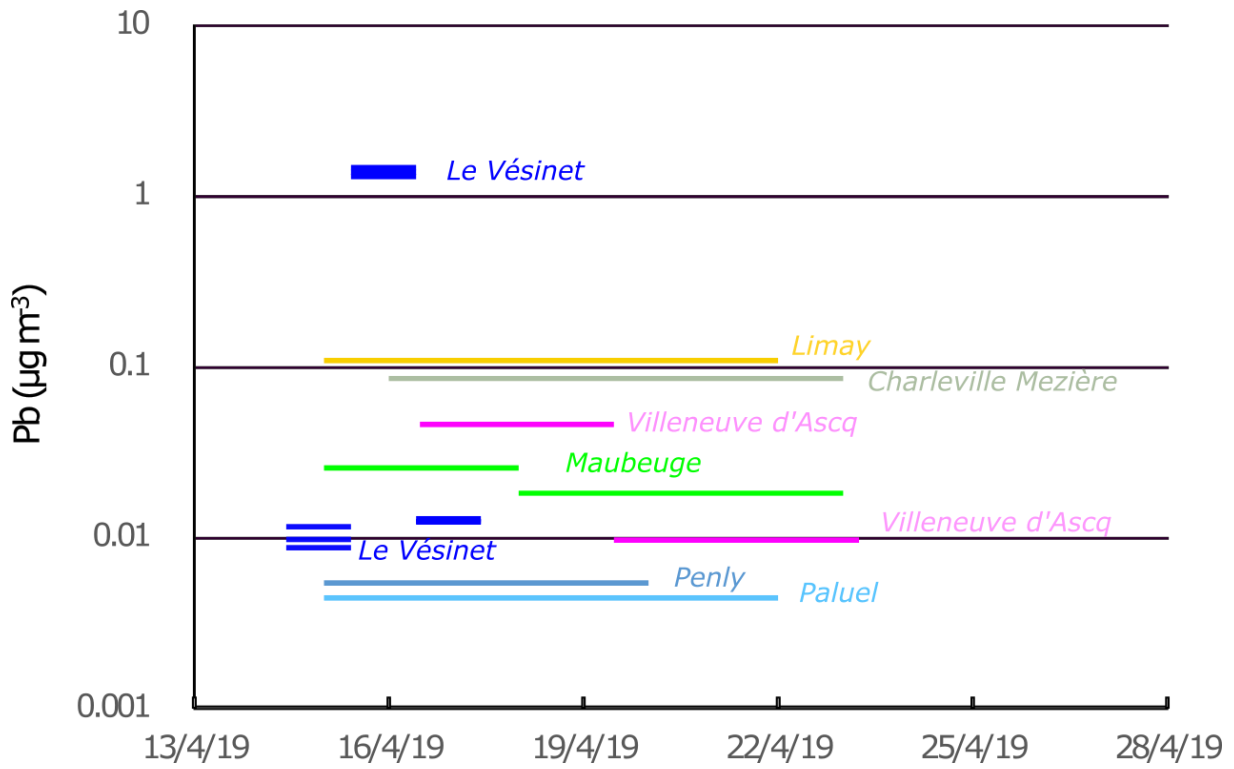
325

326

327 FIGURES

328

329 **Figure 1**



330

331

332 *Figure 1: Pb concentrations ($\mu\text{g m}^{-3}$) measured in air filters from OPERA monitoring program. For le Vésinet, triplicates are*
333 *indicated by different bars. Limay data is from AirParif*

334

335

336

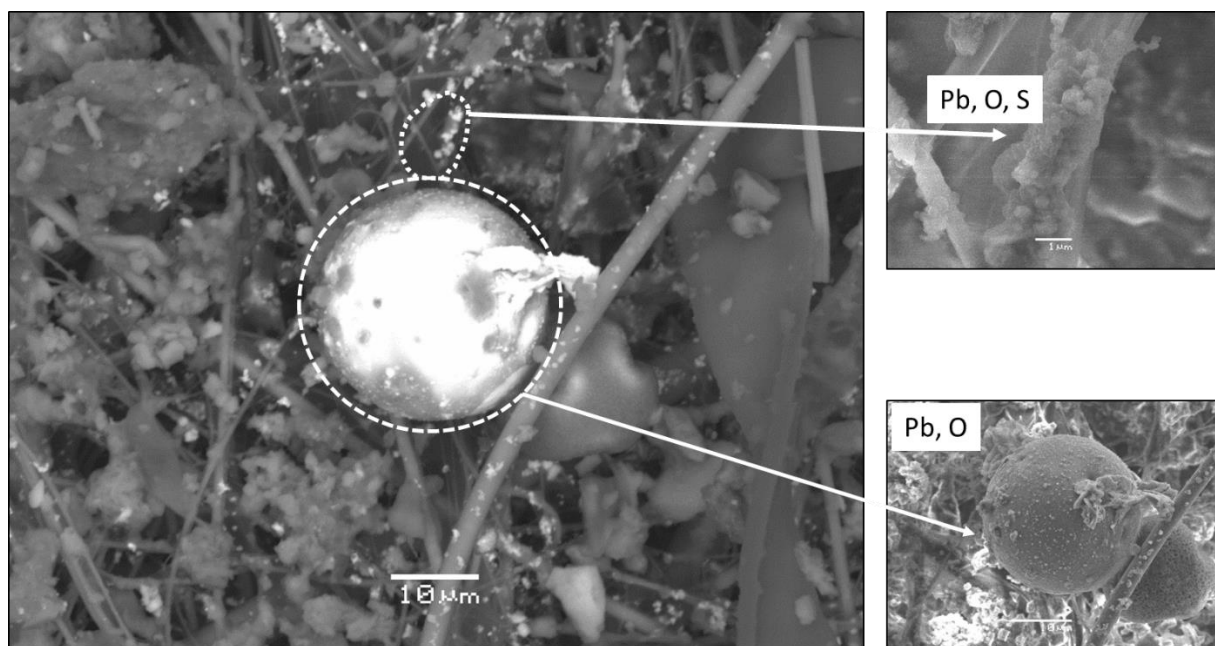
337

338

339

340

341 **Figure 2**



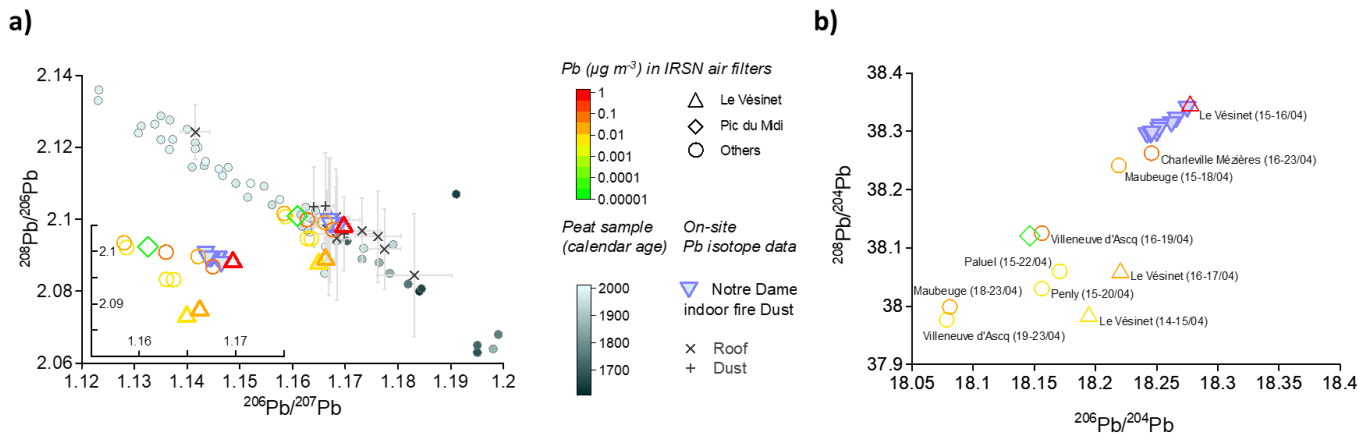
342

343 *Figure 2: Examples of Pb particles found in Le Vésinet filter sampled from 15 April 2019 to 16 April 2019 identified with SEM.*
344 *Backscattered electron image on the left, secondary electron images on the right.*

345

346

347 **Figure 3**



348

349 *Figure 3 : Pb isotopes measured in air filters from OPERA network following Notre Dame fire. Uncertainties from our*
 350 *measurements are smaller than the symbol sizes. A) $^{206}\text{Pb}/^{207}\text{Pb}$ vs. $^{208}\text{Pb}/^{206}\text{Pb}$ diagram: Pb isotopes measured by*
 351 *Gloennec et al.¹⁸ in Notre Dame dusts and melted Pb are shown in purple (2 “outliers” are not figured), Notre Dame indoor*
 352 *fire dust are from Briard et al.¹⁹ Background Pb isotope chronological evolution of French atmosphere is based on dated peat*
 353 *layers from eastern and south France, Belgium and Swiss Jura (original data Hansson et al.²¹, Forel et al.²², Shotyk et al.²³,*
 354 *Allan et al.²⁴). The inner panel is a zoom on the filter data and the data from Briard et al. B) $^{206}\text{Pb}/^{204}\text{Pb}$ vs. $^{208}\text{Pb}/^{204}\text{Pb}$*
 355 *diagram with indoor dust data from Briard et al.¹⁹*

356

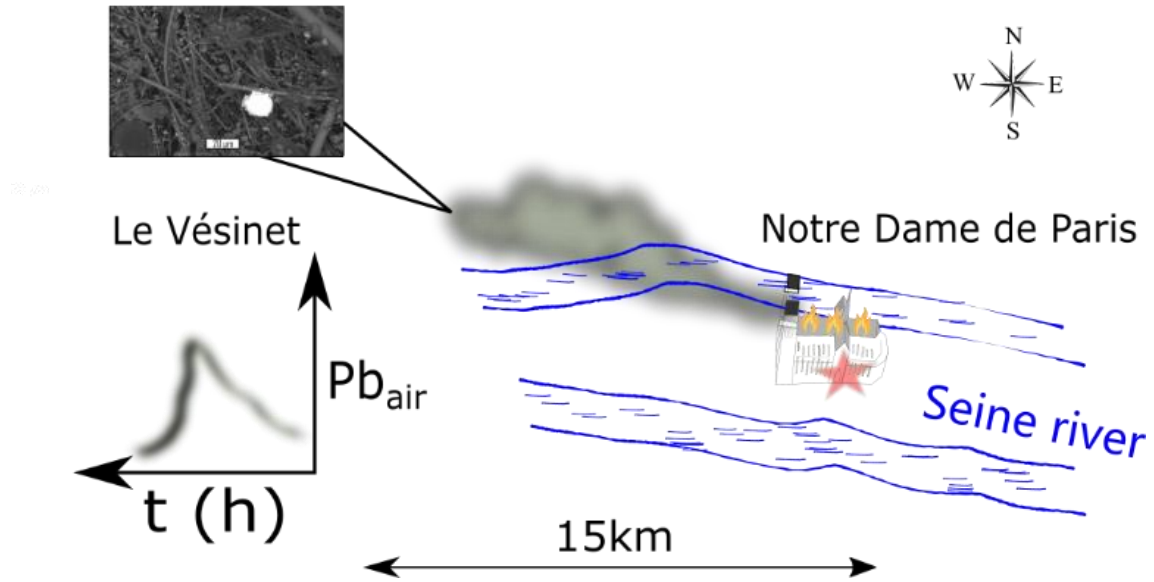
357

358

360 For TOC Only"

361

362



363

364

365

366 **Lead from Notre Dame fire plume caught 15 km from Paris**

367 **Supplementary information**

368 Gaël Le Roux¹, Olivier Masson², Christophe Cloquet³, Marie-José Tavella¹, Magali Beguin-Leprieur⁴,

369 Olivier Saunier⁵, David Baqué¹, Thierry Camboulive¹, Julien Berger⁶, Thierry Aigouy⁶, Françoise

370 Maube⁶, Sandrine Baron⁷, Sophie Ayrault⁸, Maxime L'Héritier⁹

371 **1. Laboratoire d'Ecologie Fonctionnelle et Environnement CNRS-Université de Toulouse**

372 Campus Ensat - Avenue de l'Agrobiopole- BP 32607 Auzeville tolosane
373 31326 Castanet-Tolosan, France

374 **2. Institut de Radioprotection et de Sécurité Nucléaire- PSE-ENV/SEREN**

375 Cadarache BP3 13115 Saint-Paul-Lez-Durance Cedex, France

376 **3. CRPG, UMR 7358, Université de Lorraine-CNRS, Nancy**

377 15 rue Notre Dame des Pauvres, 54500 Vandœuvre les Nancy – France

378 **4. Institut de Radioprotection et de Sécurité Nucléaire- PSE-ENV/SIRSE**

379 31 rue de l'écluse. BP 40035 78116 Le Vésinet Cedex, France

380 **5. Institut de Radioprotection et de Sécurité Nucléaire -PSE-SANTÉ/SESUC BMCA**

381 31 Av. de la Division Leclerc, 92260 Fontenay-aux-Roses, France

382 **6. GET-OMP CNRS-IRD-Université de Toulouse**

383 14 Av. Edouard Belin, 31400 Toulouse, France

384 **7. TRACES, CNRS UMR 5608, Université de Toulouse**

385 Université Toulouse Jean Jaurès, Maison de la Recherche

386 5, allées Antonio Machado F 31058 Toulouse cedex 9, France

387 **8. Laboratoire des Sciences du Climat et de l'Environnement (LSCE-IPSL), UMR 8212 (CEA/CNRS/UVSQ), Université**
388 **Paris-Saclay, Gif-sur-Yvette, France**

389 CEA Saclay, Bat 714, Site de l'Orme des Merisiers

390 Chemin de Saint Aubin - RD 128 F-91191 Gif sur Yvette Cedex, France

391 **9. Université Paris 8, ArScAn UMR 7041.**

392 Maison de l'Archéologie et de l'Ethnologie

393 21 allée de l'Université

394 F-92023 Nanterre Cedex, France

395

396

397 Corresponding author :

398 Gaël Le Roux

399 Tel : 33 (0) 5.34.32.37.79

400 fax : 33 (0) 5.34.32.39.01

401

402

403 9 pages

404 5 supplementary figures

405 1 supplementary table included in this text file

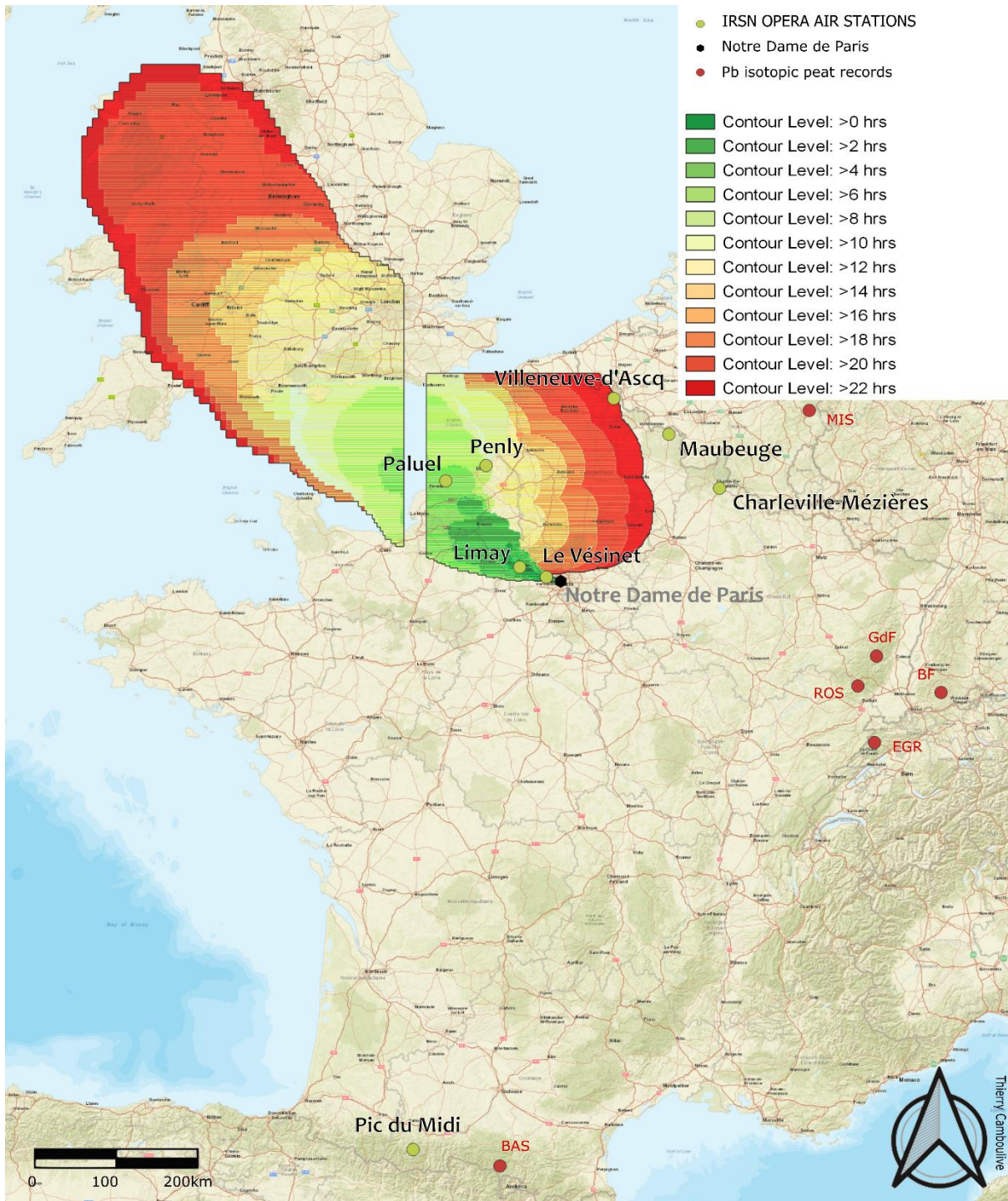
406 1 supplementary XLS spreadsheet

407 Online and offline hysplit simulations:

408

409 An online simulation was run on hysplit the day after the fire to assess the potential of OPERA IRSN
410 filters to track the particulate plume from the Notre Dame fire.

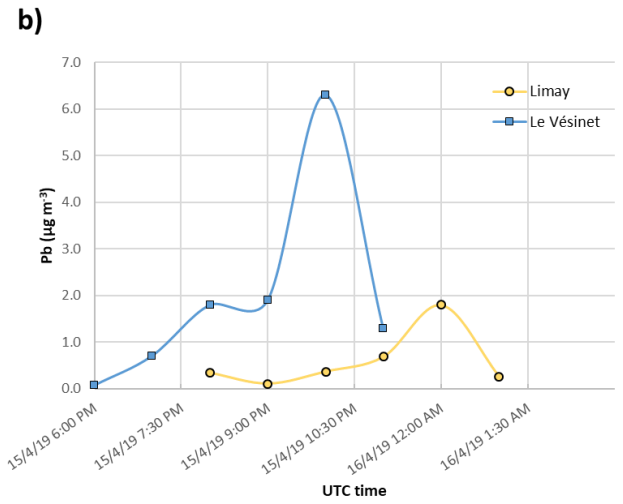
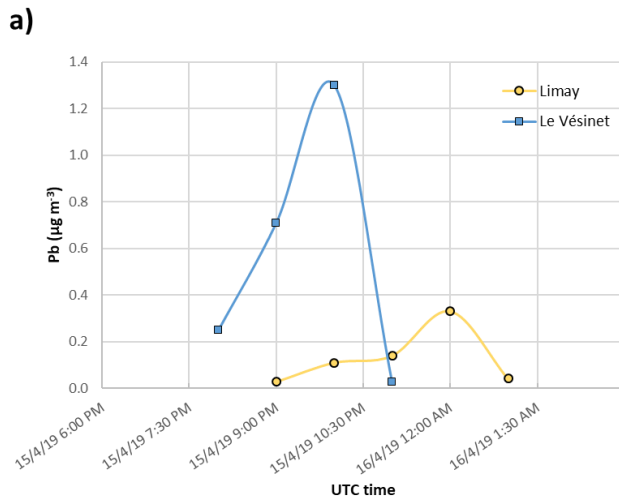
411 In addition, 2 offline rough simulations have been produced as examples. The first is a simulation of a
412 lead emission of 40kgs/h of Pb from 50m to 100m height one hour after the start of the fire. The
413 second simulation considers an emission of 125kgs/h of Pb for 4 hours one hour after the start of the
414 fire. To obtain the potential hourly air concentrations in Limay and Le Vésinet, we used the offline
415 version of Hysplit (particle mode concentration model). For the 1st simulation, no wet deposition
416 forcing was included. For the 2nd simulation, we used the default wet deposition parameters similar
417 to ¹³⁷Cs considering a micron size for the aerosols.



418

419
420

Figure S1 : Hysplit plume trajectory with OPERA AIR STATIONS and peatland locations used as Pb isotopic records – contour levels indicate time of arrival of the plume



421

422 *Figure S2 : a) offline hysplit Pb simulation at Le Vésinet (blue squares) and Limay (orange circles) locations using an emission*
 423 *of 40 kg of Pb emitted during 3 hours; b) offline hysplit Pb simulation using 125 kg/h rate of Pb emission during 4 hours at*
 424 *50m agl and with a similar behaviour as ¹³⁷Cs bearing particles*

425

426 **Filter quality control :**

427 OPERA filters are made of several filter layers ((polypropylene filter (3 M™). To evaluate potential
428 variability between the different layers, we measured triplicates made of subsamples of three layers.
429 A known surface of each filter was subsampled 3 times, digested using 3 ml of ultrapure 69% HNO₃
430 (Optima™, for Ultra Trace Elemental Analysis, Fisher Chemical™) at 105°C overnight in 50ml PP
431 DigiTUBEs (SCP Science, Villebon sur Yvette-France), then adequately diluted with ultrapure water
432 (Veolia Purelab, Toulouse-France) and measured with a ICP-OES (AMETEK SPECTRO ARCOS FHX22).
433 For Pb isotope measurements, we selected an intermediate filter.



434

435 *Figure S3: : Example of triplicates le Vésinet filter (16/04/2019 to 17/04/2019)*

436 The iron blank was negligible that is to say 10 times lower than the lowest concentration filter apart
437 the Pic du Midi filter. The lead blank was under the detection limit of the ICP-OES instrument that is
438 20 times lower than the lowest Pb concentration measured in OPERA filters. The blank for the
439 isotope separation chemistry is typically in the sub-nanogram range and usually around 300 pg of
440 lead.

441

442 To test the validity of our protocol, we put approximately 50mg of WQB-1 reference material (NWRI
443 CRM-Lake Ontario Sediment) on a blank filter and digest it similarly overnight as with the other
444 filters.

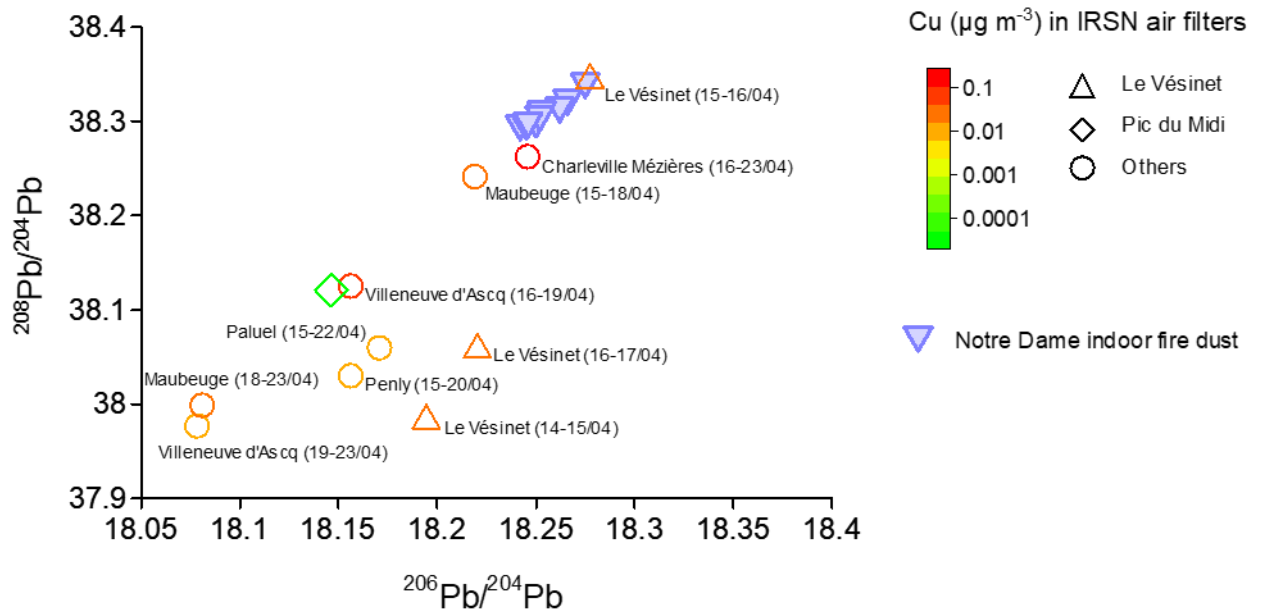
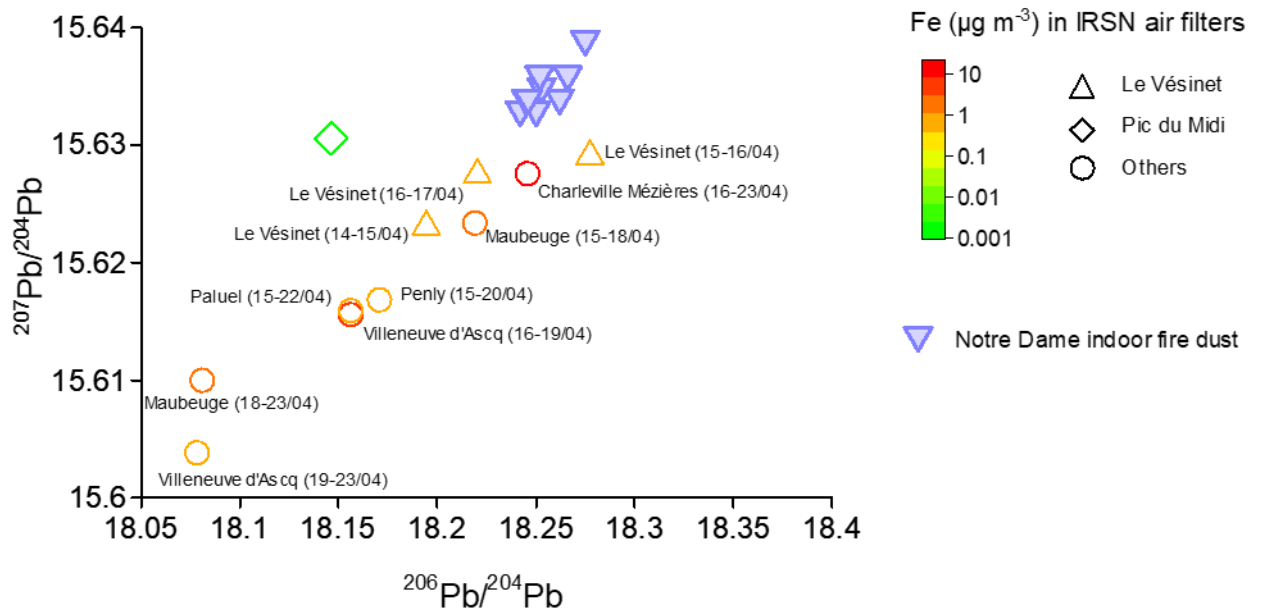
WQB-1 lake sediment	Measured	Certified
Pb (mg kg ⁻¹)	85	83.7
Fe (mg kg ⁻¹)	40020	47358
Cu (mg kg ⁻¹)	82	79.6

445 *Table S1 : Comparison of measured and certified values for WQB-1 reference material*

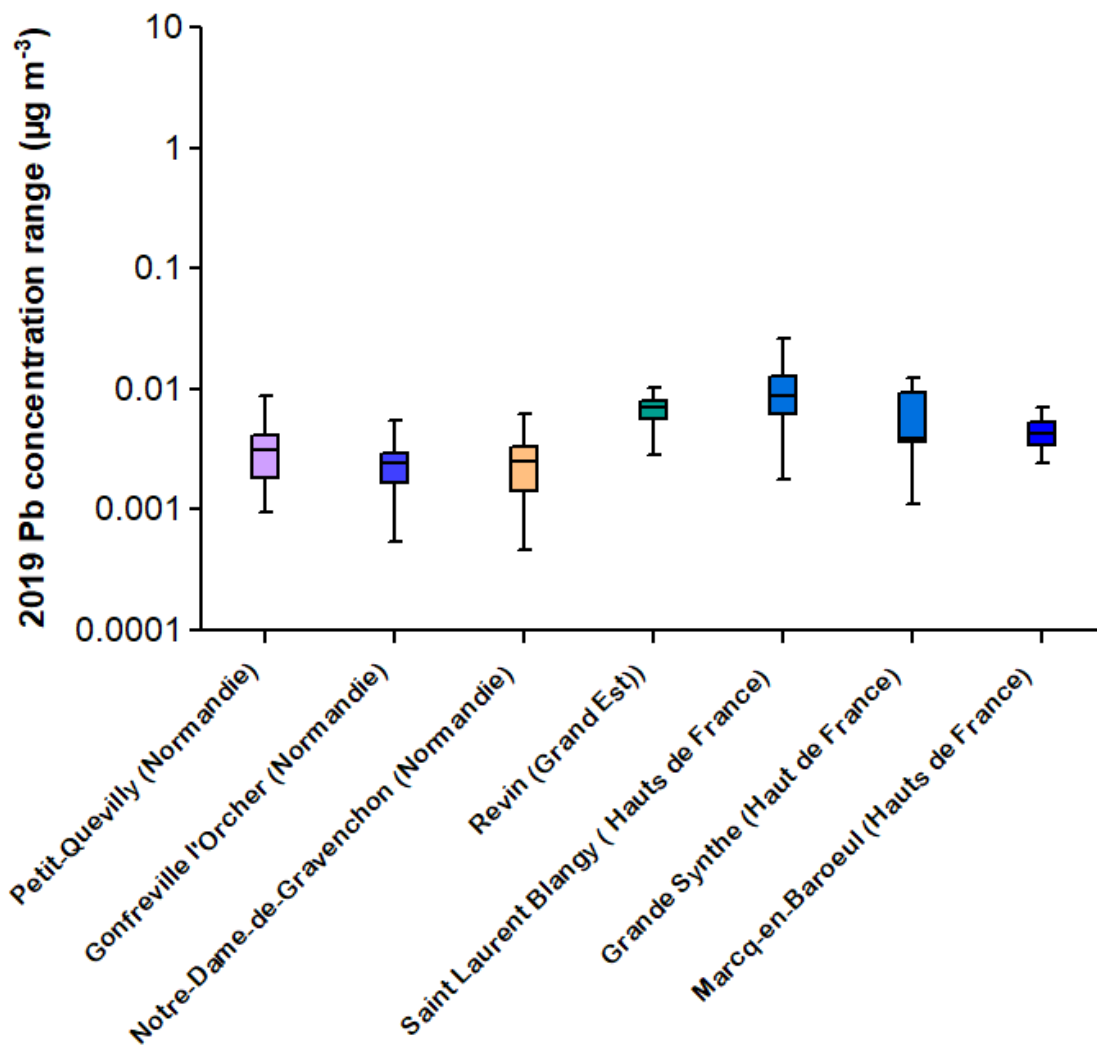
446

447

448



451 Figure S4 : Pb isotope ratio diagrams and Fe and Cu concentrations for IRSN OPERA filters



452

453 *Figure S5: Lead concentration ranges in filters collected a) every other week by Atmo Normandie in 2019 on three locations.*
 454 *The week of the fire, no filter was measured by Air Normandie. (data courtesy from Atmo Normandie) ; b) by Atmo*
 455 *Grand'Est in Revin from jan. 2019 to April. 2019 every ~2weeks ; c) by Atmo Hauts de France in 3 sites in winter and summer*
 456 *2019*

457

458 [XLS spreadsheet S1 : Pb concentrations and Pb isotopes data \(xls\)](#)

459

460