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1	Voltammetric study of the affinity of divalent heavy
2	metals for guanine functionalized iron oxide nanoparticles

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9 Abstract In this study, a novel nanobiomaterial based on (3-10 11 aminopropyl)triethoxysilane (APTES) coated iron oxide (Fe_3O_4) nanoparticles functionalized with newly synthesized guanine hydrazide 12 (GH) was elaborated. A boron-doped diamond electrode coated with GH-13 APTES-Fe₃O₄ nanoparticles was used to assess the interaction of heavy 14 15 metal ions with guanine hydrazide. The adsorption isotherms were 16 electrochemically investigated and it was shown that the adsorption capacity 17 of the nanoparticles towards heavy metals decreased in the following order: 18 $Cu^{2+}>Pb^{2+}>Cd^{2+}$. From the calibration curves, the sensitivities of detection 19 were as follows: 171.6 μ A/ μ M for Cu (II), 156 μ A/ μ M for Pb (II) and 101.4 20 $\mu A/\mu M$ for Cd (II).

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1	Ke	ywords	Nucleosides	•	Heavy	Metals	•	Nanostructures •
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1 Introduction

2 Despite the many uses of heavy metals, excessive exposure, either 3 externally or by ingestion, can be harmful to humans, animals and plants on 4 several levels. They are nonbiodegradable and thus have the tendency to 5 accumulate in living organisms. Besides being considered as environmental 6 pollutants [1], some heavy metals can cause skin, liver, kidney, bone and 7 neurological diseases [2]. Along with natural sources, heavy metals can be 8 released from industrial sources such as fertilizers, household waste, 9 cosmetics, paint, batteries and plastics [3]. The range of toxicity of heavy 10 metals depends on several factors including their charges, concentrations, 11 way of exposure...

12 Although the main function of nucleic acids is encoding genetic 13 information in human beings, they have been widely explored in the fields 14 of catalysis and ligand binding [4]. Nucleic acids are known to have different 15 metal binding sites, binding capacity and affinity for heavy metals [5]. For 16 years, the interaction of metal cations with nucleic acids has been 17 investigated. It was noted that metal ions impact differently the stability and 18 structure of nucleic acids depending on the type and concentration of metal 19 ion and its relative affinity to the phosphate and nucleobases. It has been 20 hypothesized that bioavailable metal ion-induced toxicity inactivates vital 21 processes, inhibits metabolic pathways and directly or indirectly displaces

essential metals from the active sites of macromolecules and/or disrupts 1 2 depolymerization or repair of nucleotide bases with subsequent errors in 3 protein synthesis [6]. The study of the interaction between nucleic acids and heavy metal ions can help explain their biotoxicity mechanisms, and these 4 interactions can be used for the detection of toxic heavy metals in waters. 5 6 Recently, nucleic acids, nucleotides, nucleosides, nucleobases and their 7 derivatives have been gaining wide popularity in heavy metal sensing [7]. 8 They are biodegradable, highly selective and can be produced by natural or 9 synthetic methods. Many studies have relied on the stabilization or 10 destabilization of G-quadruplex structure of guanine-rich DNA, formation 11 of stable base pairs or catalytic cleavage of DNA substrates for the detection 12 of heavy metals [8].

13 Among the five natural nucleobases, guanine and its derivatives have received considerable interest [9]. Guanine is a purine derivative that 14 15 consists of a fused pyrimidine-imidazole ring system with conjugated double 16 bonds. It has been well established that metal ions bind to the carbonyl 17 oxygen O(6) and/or the imidazolyl nitrogen N(7) of guanine. Preference of 18 the binding site depends on the metal ion itself [10]. In the present work, the 19 voltametric behavior of guanine in the presence of divalent heavy metal ions 20 was studied. It is known that the redox moiety of guanine is the imidazolyl 21 nitrogen N(7) [11]. The interaction of N(7) with heavy metal ions can then

be monitored through the change of the electron transfer rate. For the 1 sensitive detection of the oxidation of guanine, synthesized guanine 2 3 hydrazide (GH) was immobilized on iron oxide nanoparticles that present a large specific surface area and a high electrical conductivity. Recently, 4 magnetic nanoparticles, particularly magnetite Fe₃O₄ nanoparticles, have 5 6 attracted a lot of interest owing to their unique properties of 7 superparamagnetism, low Curie temperature, high magnetic susceptibility, 8 low toxicity and cost effectiveness [12,13]. The exclusive magnetic and electrical properties are the result of the transfer between Fe^{2+} and Fe^{3+} [14]. 9 They have a wide variety of applications ranging from environmental to 10 11 biological and medical. Synthesis conditions are crucial in determining their size and physicochemical properties, and thus several processes have been 12 investigated for their production depending on the desired application. 13 14 Moreover, Fe₃O₄ nanoparticles have the tendency to aggregate due to 15 magnetic dipole-dipole interactions between particles [15], which is why in 16 most electrochemical studies they are either coated or functionalized [16]; 17 i.e. iron oxide core encapsulated in an organic or inorganic shell [17]. Owing 18 to their versatile properties, Fe₃O₄ nanoparticles were extensively used in 19 combination with biological recognition elements such as enzymes, 20 antibodies and nucleic acids for various applications [18]. The guanine 21 functionalized iron oxide nanoparticles were immobilized on the surface of

a boron doped diamond (BDD) electrode. BDD electrodes are extensively 1 2 investigated for electroanalytical applications, because of their 3 electrochemical properties, as a low background current and a wide potential window in aqueous solutions (\sim -1.35 to +2.3 V versus the normal hydrogen 4 5 electrode), corrosion stability in aggressive media and resistance to 6 biofouling [19]. The redox properties of guanine were detected in the 7 presence of copper, lead and cadmium ions. From these results, adsorption 8 of the three heavy metals on N(7) site of guanine was characterized and the 9 sensitive electrochemical detection of these heavy metal could then be 10 achieved. This study aims to show that the use of the proposed guanine functionalized nanoparticles associated with a voltammetric technique could 11 present a great potential in the physico-chemical characterization of the 12 13 guanine/metal interaction.

14 **Results and Discussion**



15 Guanine hydrazide preparation and characterization

Fig. 1 Schematic diagram showing the steps of guanine hydrazide synthesis

2 Figure 1 summarizes the overall steps involved in the synthesis of 3-3 (2-amino-6-oxo-1,6-dihydro-9H-purin-9-yl)propane hydrazide (6, C₈H₁₀N₇O₂) or guanine hydrazide. The process of guanine hydrazide 4 synthesis generated a white product with a yield of 98%. To prepare the 5 6 guanine hydrazide 6 shown in figure 1, The exocyclic amine of guanine 1 7 was first protected by acetylation according to literature procedure [21]. Then reaction of N^2 -acetylguanine **3** via a Michael addition with Ethyl 8 9 acrylate and sodium ethoxide afforded, as expected, a mixture of two 10 isomeric ethyl propionates resulting from guanine alkylation at either the N-11 9 (5) or N-7 (4) atom. These compounds were separated by Flash 12 chromatography and, after their structural assignment, the reaction of the N-13 9 ester with hydrazine hydrate in ethanol led to the formation of the desired 14 acyl hydrazide 6 according to an established literature procedure [22].

The synthesized guanine hydrazide was characterized using FTIR and ¹H and ¹³C NMR. The infrared spectrum shown in figure 2a revealed peaks at 1695 cm⁻¹ corresponding to C=O, 1657 and 3310 cm⁻¹ corresponding to NH and 1440 cm⁻¹ corresponding to C-N. Peaks at 2854 cm⁻¹ correspond to C-H stretching of CH₂ in the hydrazide function, along with a weak peak at 1563 corresponding to C=C stretching. ¹H NMR characterization (500 MHz, DMSO-d₆) showed that the two methylene group protons resonate at 3.36 (triplet, 2H) and 4.14 (triplet, 2H) while the proton from the methine group
resonates at 7.54. Protons from the primary amine groups at 4.22 – 4.19 and
6.49 whereas those from secondary amines appear at 9.04 and 10.49 (figure
2b). ¹³C NMR results in figure 2c are consistent with literature such that the
signals corresponding to the guanine carbons appear at 156.8, 153.5, 151.0,
137.4 and 116.5 while those for the hydrazide function appear at 168.7, 39.5
and 33.5 [23].









Fig. 2 (a) FTIR spectrum, (b) ¹H NMR, (c) ¹³C NMR and (d) HRMS of
guanine hydrazide

4 Preparation of GH-APTES functionalized Fe₃O₄ nanoparticles

The co-precipitation of ferrous and ferric ions in a basic medium is a 5 classical method for the preparation of Fe₃O₄. The size of the iron oxide 6 nanoparticles directly affects their magnetic properties and should therefore 7 be optimized based on the desired application [13]. Nanoparticles that are 8 9 too small (less than 20 nm) can't be manipulated by an external magnetic 10 field. Hence, the coprecipitation procedure was carried out at 0 °C to ensure 11 the synthesis of nanoparticles of adequate size. Upon the addition of NaOH, 12 the color of the solution changed gradually from orange to brown to black.

1 Precipitation of Fe_3O_4 is expected at pH > 7.5 [15]. The reactions involved

2 in the production of Fe_3O_4 are:

3
$$\operatorname{Fe}^{3+} + 3\operatorname{HO}^{-} \rightleftharpoons \operatorname{Fe}(\operatorname{OH})_{3}$$
 (1)

4
$$Fe(OH)_3 \rightleftharpoons FeO(OH) + H_2O$$
 (2)

5
$$\operatorname{Fe}^{2+} + 2\operatorname{HO}^{-} \rightleftharpoons \operatorname{Fe}(\operatorname{OH})_2$$
 (3)

6
$$\operatorname{Fe}(OH)_2 + 2\operatorname{FeO}(OH) \rightleftharpoons \operatorname{Fe}_3O_4 + H_2O$$
 (4)

The process of amino modification of Fe_3O_4 nanoparticles by direct immobilization of (3-aminopropyl)triethoxysilane on the surface of the nanoparticles can be expected to occur through a silanization reaction, leaving the NH₂ group of APTES exposed. This step is usually very complex and is influenced by several parameters including the reaction time, temperature and silane concentration.

Finally, the functionalization of the APTES-coated nanoparticles with guanine hydrazide was done using glutaraldehyde. The overall reaction for the synthesis of the functionalized electrode is shown in figure 3.



- 1 Fig. 3 Schematic diagram showing the process of synthesizing the GH-
- 2 APTES-Fe₃O₄ nanoparticles

3 Characterization of the functionalized nanoparticles

4 FTIR spectroscopy was used to study the functionalization of the bare 5 and coated nanoparticles, as shown in figure 4A. For the bare nanoparticles, 6 one peak at 550 cm⁻¹ is noticeable. This peak is the characteristic absorption 7 peak of Fe-O. This band can also be observed with the coated nanoparticles, 8 along with new bands confirming the adsorption of (3aminopropyl)triethoxysilane on the surface of Fe₃O₄. Bands at 3672 cm⁻¹ and 9 10 1624 cm⁻¹ can be attributed to the N-H stretching and NH₂ bending modes, respectively, of a free NH₂ group. Bands at 1063 cm⁻¹ and 1162 cm⁻¹ can be 11 12 attributed to Si-O-Si and SiO-H groups. The stretching of -CH₂ was confirmed with bands at 2902 cm⁻¹ and 2973 cm⁻¹ [13]. The absorption band 13 14 corresponding to the Fe-O-Si bond can't be seen since it overlaps with that of Fe-O vibration. However, the shift in the Fe-O band from 550 cm⁻¹ to 562 15 16 cm⁻¹ indicates that there is an Fe-O-Si bond [16]. FTIR measurement after 17 the functionalization of APTES coated nanoparticles with GH confirmed the 18 crosslinking of GH to the APTES by the appearance of new peaks not 19 characteristic of APTES. The shift in the peak corresponding to N-H stretching from 3672 cm⁻¹ to 3440 cm⁻¹ is consistent with N-H stretching of 20 21 an amide group along with an increased peak intensity. The -CH₂ stretching

peaks also increased in intensity due to the increased number of CH₂ groups
 from the hydrazide function.

Powder XRD patterns of the coated nanoparticles show diffraction peaks at 30.1°, 35.4°, 43.1°, 53.4°, 57.0°, and 62.6° corresponding to the characteristic pattern of pure magnetite Fe₃O₄ (figure 4B). The diffraction peaks can be assigned to the (220), (311) (400), (422), (511) and (440) diffraction planes, respectively [17]. This proves that the coating doesn't affect the core magnetite and doesn't lead to a phase change.

9 The morphology of the prepared nanoparticles capped with (3-10 aminopropyl)triethoxysilane was investigated using SEM as shown in figure 11 4C. The magnetic nanoparticles appear to be quasi-spherical, with uniform 12 distribution and an average size of 45 nm. Energy-dispersive X-ray 13 spectroscopy (EDX) was performed to study the elemental composition of 14 the APTES coated Fe₃O₄ nanoparticles (figure 4D). Peaks corresponding to 15 iron (Fe) at about 0.6, 6.3 and 7 keV and oxygen (O) confirm the formation 16 of iron oxide nanoparticles [18]. The presence of peaks corresponding to Si 17 and N confirm that the APTES coating was successful.





3 Fig. 4 (A) FTIR spectrum of (a) bare Fe_3O_4 NPs, (b) APTES- Fe_3O_4 NPs and

4 (c) GH-APTES-Fe₃O₄NPs, (B) XRD pattern, (C) SEM image and (D) EDX

5 spectrum of APTES-Fe₃O₄ nanoparticles

6 Voltammetric study of the GH-APTES functionalized nanoparticles in the

7 presence of divalent heavy metals

8 Prior to studying the voltammetric behavioral interaction of heavy 9 metals with the modified electrode, each of the different elements were 10 studied electrochemically to check for their respective signals. As shown in 11 figure 5a, no observable peaks were detected with square wave voltammetry in the range between -1 and 1.5 V for the BDD electrode coated with
APTES-Fe₃O₄ NPs. However, upon functionalization of the nanoparticles
with GH, a peak appears at 1.1 V as indicated in figure 5b. This peak is
characteristic of the oxidation of guanine at a BDD electrode [11, 24] and
not the hydrazide function which is oxidized at a higher potential (1.8 V)
[25]. The oxidation reaction of guanine in aqueous media is presented in Fig.



9 Fig. 5 SWV of (a) APTES-Fe₃O₄ nanoparticles and (b) APTES-Fe₃O₄

10 nanoparticles functionalized with GH



11

7

6 [11].

- 12 Fig. 6 Oxidation reaction of guanine, in aqueous medium
- 13 The effect of Cd (II) ions concentration on the oxidation signal of GH
- 14 was first investigated after interaction in solution and using a bare BDD (data

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not shown). Square wave voltammetry was applied for a solution of GH in citrate buffer (pH 4) with 0.5 mol dm³ KCl at the following optimized conditions: frequency 50 Hz; amplitude 50 mV; step potential 10 mV. Before adding cadmium, the oxidation peak of GH appeared at 1.1 V. The peak current decreased upon the addition of increasing concentrations of Cd²⁺ implying that the complex formation of cadmium ions with GH hinder the redox reaction of the guanine moiety.

8 Using a BDD electrode modified with GH functionalized 9 nanoparticles, the same behavior was observed. The intensity of the peak 10 current was found to decrease when the Cd (II) concentration increased. The 11 same behavior was observed with all the studied heavy metal ions. The 12 square wave signals in the presence of the different divalent heavy metals 13 are shown in figure 7.



Fig. 7 SWV of (a) Cd (II) at the GH-APTES-Fe₃O₄ NP electrode over a concentration range of $0 - 2.5 \mu mol dm^{-3}$, (b) Pb (II) at the GH-APTES-Fe₃O₄ NP electrode over a concentration range of $0 - 1 \mu mol dm^{-3}$ and (c) Cu

(II) at the GH-APTES-Fe₃O₄ NP electrode over a concentration range of 0 –
 3.5 µmol dm⁻³

3 Adsorption isotherms of the divalent heavy metal ions on the GH-APTES

4 functionalized nanoparticles

5 In order to characterize the interaction between heavy metals and the adsorption isotherms 6 functionalized nanoparticles, modeled. were 7 Adsorption isotherms are mathematical models describing the distribution of 8 the adsorbate between the solution and the adsorbent. These models are 9 based on to the possibility of interaction between the adsorbate species, type 10 of coverage and the homogeneity/heterogeneity of the adsorbent surface 11 [26]. The variation of the peak maximum (ΔI) (mean value of triplicate) is 12 proportional to the adsorbed concentration of the heavy metals at the 13 modified electrode surface. The adsorption isotherm of the heavy metals was 14 then built from this value, the initial concentration of the heavy metals was 15 not highly modified, due to the small surface area of the modified electrode 16 in contact. Two models were tested: Langmuir and Freundlich.

17 *i. Langmuir isotherm*

18 The Langmuir model assumes that adsorption occurs on a structurally 19 homogeneous adsorbent in a monolayer, or that the adsorption occurs at 20 fixed sites that are identical and energetically equivalent. The linearized form 21 of the Langmuir isotherm can be written as follows [27]:

$$1 c_e/\Delta I = 1/Q_0 b + c_e/Q_0$$

where Q_0 is the adsorption capacity (µg g⁻¹) and b is the energy of 2 adsorption (L μ g⁻¹). Figure 8 shows the Langmuir plot $c_e/\Delta I$ versus c_e for the 3 4 different heavy metals, each in their respective concentration range. The plots indicate that the experimental data are well fitted with the model with 5 6 $r^2 > 0.999$. The values of the adsorption capacities were 400, 280.1 and 343.6 mg.g⁻¹ for Cu²⁺, Cd²⁺ and Pb²⁺, respectively. Moreover, the essential 7 8 characteristics of the Langmuir model can be expressed as the dimensionless 9 constant separation factor R_L as follows [27]:

$$R_L = 1/(1 + Q_0 b)$$

In the case of the three different metals, R_L falls in the range of 0 - 1indicating that the nature of the adsorption isotherm is favorable. The adsorption capacity of the functionalized nanoparticles towards the studied heavy metal ions decreases in the following order: $Cu^{2+} > Pb^{2+} > Cd^{2+}$.

15 *ii. Freundlich isotherm*

Freundlich isotherm is another form of the Langmuir model describing a heterogeneous system with multiple adsorption sites. The linearized logarithmic form of the Freundlich model can be expressed as follows [28]:

$$20 \qquad \log \Delta I = \log K_F + (1/n) \log c_e$$

1	where K_F is the Freundlich constant (µg g ⁻¹) and n is the Freundlich
2	exponent. The different isotherms for each heavy metal are shown in figure
3	8 where log ΔI is plotted versus log c_e . It should be noted that the
4	concentration range over which this model was linear was much smaller than
5	that with the Langmuir model for all the studied metal ions. The values of n
6	lie between 1 and 10, which implies that the adsorption is favorable and
7	heavy metals are easily extracted from the solution. This result agrees with
8	the values of R_L from Langmuir's isotherm. Langmuir and Freundlich model
9	parameters and linear regression correlations calculated from curve fittings
10	are shown in Table 1. Better regression correlations were calculated for
11	Langmuir model suggesting that the Langmuir model is more convenient in
12	describing the adsorption process. This signifies that the adsorption is of a
13	typical monomolecular layer form [29]. It assumes that the adsorption is
14	localized, all active sites have similar energies and no interaction between
15	the adsorbed molecules exists [30]. The different affinities and adsorption
16	capacities rely on both the adsorbent (mainly its guanine moiety) and the
17	adsorbate (ionic radius, electronegativity, charge) [29,31]. Consequently,
18	comparing the divalent metal ions, the orders of affinity and adsorption
19	capacity decrease with increasing ionic radii and electronegativity.
20	Table 1 Constant parameters and correlation coefficients calculated for

21 different adsorption models for Cu^{2+} , Cd^{2+} and Pb^{2+} adsorption

Isotherm	Lai	ngmuir			F	reundlich	l
Constants	$Q_{ heta}$	b	R_L	r ²	n	<i>k</i> _{<i>F</i>}	r ²
	(µg g ⁻¹)	(L µg ⁻¹)				(µg g ⁻¹)	
Cd ²⁺	280.1	0.087	0.0394	0.9999	6.06	118.2	0.9614
Cu ²⁺	400	0.0707	0.0341	0.9998	5.447	150.7	0.9796
Pb ²⁺	343.6	0.1046	0.0271	0.9999	7.776	174.3	0.9961



1 **Fig. 8** Langmuir adsorption isotherms for (a) Cd^{2+} , (c) Cu^{2+} and (e) Pb^{2+} in

2 the range of $13 - 202 \text{ mg dm}^{-3}$. Freundlich adsorption isotherms for (b) Cd²⁺,

3 (d) Cu^{2+} and (f) Pb^{2+} in the range 40 – 124 mg dm⁻³

4 According to Sigel et al., guanine contains imidazole and pyridine-5 type nitrogens well suited for divalent metal ions binding [5], and in general 6 the binding of metal cations is governed by the formation of energetically 7 favorable coordination complexes [6], the interaction of guanine or a guanine derivative G with a divalent metal ion M^{2+} yields $M(G)^{2+}$ complexes [32]. 8 9 Most investigations about the complexation of the nucleobases with metal 10 ions do not focus on the free nucleobases, but rather on DNA itself or DNA-11 resembling structures and related bio-molecules [4]. Nonetheless, it is 12 reported that amongst divalent metal ions that can bind both the phosphate 13 and the DNA base, copper (II) has the strongest affinity to the base [6], which 14 confirms our findings. It was shown that copper can be coordinated to 15 guanine through the N(7) and O(6) atoms, as well as a possible mono-16 coordination to the N(7) position only, depending on the guanine tautomer 17 [33]. Lead on the other hand is known to weakly interact with the N(7) and O(6) atoms of guanine [6] knowing that Pb^{2+} is an excellent oxidizer and 18 19 strong Lewis acid [34]. Cadmium is classified as a soft metal that binds 20 preferentially to DNA bases more than DNA phosphates. Notwithstanding the different studies dedicated to the interaction between Cd^{2+} and 21

nucleobases, the exact nature of binding remains uncertain. It was suggested
 that cadmium most likely binds to guanine at its N(7) center in a 1:1 ratio
 [35].

4 Comparing the behavior of the three divalent heavy metals with 5 guanine, it is evident that the highest adsorption capacity is observed for Cu 6 (II) followed by Pb (II) and Cd (II). This trend comes in accordance with a 7 study done by Hammud et al. on the interaction of some nucleobase 8 derivatives with divalent metals. They found that it follows the ionic radii of 9 the cations such that as the radius increases, the stability of the formed 10 complexes increases [36]. This order is also in agreement with a report by 11 Sharma et al. who potentiometrically studied the stability constant of a 12 pyrimidine derivative with divalent metal ions. According to their 13 observations, the trend depends on both the ionic radius and the second 14 ionization energy [37]. Based on a recent study, different factors affect the 15 stability of metal complexes including the nature of the central metal ion, 16 nature of the ligand, chelating effect, macrocyclic effect, resonance effect 17 and steric effect/hindrance [38]. Thus, this original and sensitive design can 18 further help clarify the guanine-heavy metal interaction, with a possible 19 application of these functionalized beads as advanced material for heavy 20 metal ions complexation, removal and detection from aqueous solutions.

The effect of other divalent ions such as Ca²⁺ was tested. The SWV
 signal decreased less than 10% for high concentrations (>100mg/L),
 showing a very weak interaction.

4

5 Voltammetric detection of divalent heavy metals with the GH-APTES6 Fe₃O₄ modified BDD electrode

Following the adsorption isotherms, the plots of each heavy metal 7 8 were assessed to study the analytical behavior of the guanine hydrazide 9 functionalized nanoparticles. Starting with cadmium in figure 9a, its plot can 10 be divided into 2 linear parts: the first linear range was found between 0.232 and 0.809 μ mol dm⁻³ with a correlation coefficient r^2 of 0.9274 and 11 sensitivity of 101.4 μ A μ mol⁻¹ dm³ while the second linear range was 12 observed between 0.925 and 1.815 μ mol dm⁻³ with $r^2 = 0.9775$ and sensitivity 13 14 of 13.4 µA µmol⁻¹ dm³. Considering the first linear range, the obtained 15 concentration range and limit of detection are comparable to reported values 16 using Fe₃O₄ nanoparticles for the detection of Cd (II) by square wave 17 voltammetry (Table 2).

Table 2 Comparison of the linear detection range and sensitivity with
previously reported values for the electrochemical detection of Cd (II)

Electrode	Method	Concentration	LOD	Sensitivity	Reference
modification		Range (µmol	(µmol	(µA µmol ⁻¹	
		dm ⁻³)	dm ⁻³)	dm ³)	
Polydopamine-Fe ₃ O ₄	SWASV	0.02 - 0.59	9.2×10 ⁻⁵	196	[39]
Reduced graphene	SWASV	0 - 0.8	0.056	14.82	[40]
oxide-Fe ₃ O ₄					
Chitosan-Fe ₃ O ₄	SWASV	1.2 – 1.7	0.0392	8.11	[1]
Fe ₃ O ₄	SWASV	0.3 – 1.3	0.154	3.18	[41]
Terephthalic acid-	SWASV	0.4 - 1.1	0.2	12.15	[42]
Fe ₃ O ₄					
GH-APTES-Fe ₃ O ₄	SWV	0.232 - 0.809	0.077	101.4	Current
					work

The same behavior was noticed upon the subsequent addition of Pb (II) and Cu (II) respectively. The measurements were carried out in a citrate buffer solution pH 4. As shown in Fig. 9b, the calibration graph obtained for Pb (II) detection exhibited two linear response ranges of 0.0483 - 0.4521 μ mol dm⁻³ ($r^2 = 0.9113$) and $0.550 - 1 \mu$ mol dm⁻³ ($r^2 = 0.98601$) with the sensitivities of 156.38 μ A μ mol⁻¹ dm³ and 28.25 μ A μ mol⁻¹ dm³, respectively. By comparing the results obtained with previously reported

- 1 electrochemical studies, a noticeable increase in sensitivity, similar to the
- 2 case of Cd (II) was noticed (Table 3).



4 Fig. 9 Linear ranges of (a) cadmium, (b) lead and (c) copper at the GH-

5 APTES-Fe₃O₄ NP electrode

6 **Table 3** Comparison of the current concentration range and sensitivity with

7 previously reported values for the electrochemical detection of Pb (II)

Electrode	Method	Concentration	LOD	Sensitivity	Reference
modification		Range (µmol	(µmol	(µA µmol ⁻¹	
		dm ⁻³)	dm ⁻³)	dm ³)	
Amine–Fe ₃ O ₄	SWASV	0.5 – 8	0.15	10.07	[43]
Polydopamine-	SWASV	$5 \times 10^{-3} - 0.6$	1.4×10 ⁻⁶	235	[39]
Fe ₃ O ₄					
Fe ₃ O ₄	SWASV	0.3 – 1.3	0.119	14.9	[41]
Chitosan-Fe ₃ O ₄	SWASV	0.1 – 1.3	0.0422	50.6	[1]
Terephthalic	SWASV	0.4 - 1.1	0.04	8.56	[42]
acid-Fe ₃ O ₄					

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GH-APTES-	SWV	0.0483 - 0.497	0.011	156	Current
Fe ₃ O ₄					work

In the case of Cu (II), the calibration plots were divided between two linear ranges: the first range from 0.209 to 1.03 µmol dm⁻³ with r^2 = 0.92098 and a sensitivity of 171.6 µA µmol⁻¹ dm³, and the second range from 1.44 to 3 µmol dm⁻³ with r^2 = 0.94262 and a sensitivity of 13.5 µA µmol⁻¹ dm³. It is worth mentioning that the sensitivity for copper is much higher (more than ten folds) than in all other reported works using iron oxide nanoparticles [3, 11, 1, 41] with a comparable limit of detection of 0.069 µmol dm⁻³.

8 Furthermore, the reproducibility among five independent GH 9 functionalized Fe_3O_4 electrodes was evaluated by comparing the linear 10 ranges with each of the studied heavy metals. The data in figure 9 show that 11 the electrochemical sensor demonstrates a good reproducibility over five 12 independent measurements such that the maximum relative standard 13 deviation RSD is 4% in the case of Cu (II), 5% for Pb (II) and 10% for Cd 14 (II).

15 Conclusion

In the present work, we report the development of a novel material,
GH-APTES functionalized iron oxide nanoparticles. These nanoparticles
were used to modify a BDD electrode in order to assess the interaction
between guanine hydrazide and divalent heavy metal ions. The adsorption

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1	isotherms were investigated and it was proven that the data are better
2	modelled with Langmuir isotherm, with adsorption capacities decreasing in
3	the order: $Cu^{2+}>Pb^{2+}>Cd^{2+}$. This design allowed us to evaluate these ions
4	such that the sensitivities and LODs were 171.61 μ A μ mol ⁻¹ dm ³ and 0.069
5	$\mu mol~dm^{-3}$ for Cu (II), 156 $\mu A~\mu mol^{-1}~dm^{3}$ and 0.0161 $\mu mol~dm^{-3}$ for Pb (II)
6	and 101.4 μ A μ mol ⁻¹ dm ³ for 0.077 μ mol dm ⁻³ Cd (II), respectively.
7	Moreover, the sensitivity obtained in the detection of copper was at least ten
8	times greater than all reported sensitivities using iron oxide. We anticipate
9	that the synthesized GH-APTES-Fe ₃ O ₄ nanoparticles are promising
10	materials that can be used rapidly and economically for heavy metal
11	complexation and removal for water decontamination. Further tests utilizing
12	the different hydrazide nucleobases derivatives for the detection of heavy
13	metals could be of interest and could help better elucidate the sensitivity and
14	the mechanism of interaction between nucleobases and heavy metal ions.

15 **Experimental**

16 Chemical Reagents

17 All the reagents used in this study were of analytical grade and used as 18 received without any further purification. Ammonium hydroxide (NH₄OH) 19 (25% in H₂O), (3-aminopropyl)triethoxysilane (APTES) 99%, 20 glutaraldehyde solution Grade II (25% in H₂O), potassium chloride (99 – 21 100.5%) and potassium citrate tribasic monohydrate (\geq 99%) were

purchased from Sigma-Aldrich. Dimethyl sulfoxide (DMSO) (\geq 99.9%) was 1 2 purchased from Riedel-de-Haën, sodium hydroxide (NaOH) (98.5 - 100%), 3 cadmium, copper and lead standards from Fluka and ethanol (\geq 99.8%) from 4 Honeywell. Ferric chloride hexahydrate FeCl₃.6H₂O (98%) and ferrous 5 chloride tetrahydrate FeCl₂.4H₂O (99%) were purchased from Acros 6 Organic. All the chemicals used in the synthesis and characterization of 7 guanine hydrazide were purchased from Sigma-Aldrich, Merck or Fluka 8 Chemika. Deionized water (with a resistivity of 18.2 M Ω .cm) was used in 9 the preparation of all solutions.

10 Synthesis and characterization of guanine hydrazide

11 A round bottom flask was charged with guanine (10 g, 66.22 mmol) N,N-dimethylacetamide (DMA, 30 cm³) and acetic anhydride (16.5 cm³, 168 12 13 mmol). The reaction mixture was stirring overnight at 165°C after which it 14 was cooled to room temperature. The precipitate was collected and triturated with water/ethanol 1:1 (30 cm³) at 80°C for 2-3 h. The precipitate was 15 16 collected and dried under reduced pressure at 70°C to yield a white solid. The solid was transferred to a 500 cm³ round bottom flask with 17 dimethylformamide (50 cm³). An ethanolic solution of sodium ethoxide 18 19 (NaOEt, 2 cm³, 25.9 mmol) was added and the mixture was stirred for 15 20 min. Ethyl acrylate (3.8 cm³, 34.96 mmol) was added and the mixture was 21 refluxed at 165°C for 1h. The solvent was evaporated and the residue was

dissolved in methylene chloride and washed with distilled water. The organic 1 2 layer was separated, dried with anhydrous MgSO₄, filtered and the solvent 3 evaporated under reduced pressure. The crude was then triturated with toluene (250 cm³) with stirring overnight at room temperature. The 4 5 precipitate was collected, dried under reduced pressure and purified by 6 column chromatography on silica gel (CH₂Cl₂/MeOH 98:2) to yield a white 7 solid (2.85 g). The white solid (7 g, 23.9 mmol) was mixed with ethanol (70 8 cm³) and N₂H₄.H₂O (4.1 cm³, 83.6 mmol) and the mixture was stirred and 9 heated at 80°C for 24 h. The precipitate was collected and dried at 70°C 10 under reduced pressure to yield a white solid. ¹H and ¹³C nuclear magnetic 11 resonance (NMR) spectra were recorded on a Bruker 300 MHz NMR 12 spectrometer in DMSO, CDCl₃ and D₂O or D₂O/D₂SO₄, using TMS and DSS 13 as references; chemical shifts are reported in ppm. Fourier Transform 14 Infrared (FTIR) spectra were collected using PerkinElmer Spectrum Two 15 spectrometer.

16 Fabrication of APTES-coated iron oxide nanoparticles

Magnetic nanoparticles were synthesized according to a modified Massart's method [20] based on coprecipitation of an aqueous mixture of ferric and ferrous salt hydrates followed by the addition of a base. Briefly, 2.7 g of FeCl₃.6H₂O and 1.2 g of FeCl₂.4H₂O were mixed with a molar ratio of 2:1 and dissolved in 90 cm³ of deionized water. Afterwards, 100 cm³ of 1

mol dm⁻³ NaOH solution was added dropwise under vigorous stirring and 1 nitrogen gas at 0°C. A change in color indicated the formation of iron oxide 2 3 nanoparticles. The solution was subjected to magnetic stirring for 4 hours to 4 ensure nucleation and growth of magnetite nanoparticles. The latter were 5 collected with an external magnet, washed several times with water followed 6 by ethanol, and finally dried under vacuum. The solution was subjected to 7 magnetic stirring for 60 minutes to ensure nucleation and growth of 8 magnetite particles. APTES coating was carried out afterwards. Iron oxide 9 nanoparticles were sonicated for 25 minutes before the addition of NH₄OH 10 and (3-aminopropyl)triethoxysilane under nitrogen atmosphere. The mixture 11 was allowed to age for 24 hours, after which the nanoparticles were again 12 washed with water followed by ethanol and dried for later use.

13 Characterization of iron oxide nanoparticles coated with APTES

FTIR spectra were collected at each step of the synthesis using PerkinElmer Spectrum Two spectrometer in order to verify the successful synthesis of the APTES coated magnetic nanoparticles. In order to identify the crystalline phase of the prepared nanoparticles, X-ray diffraction (XRD) was performed using a Bruker D4 ENDEAVOR with Cu/K-α radiation. The operating target voltage was 50 kV and the tube current was 50 mA. The morphology and size of the nanoparticles were investigated using Scanning

- 1 Electron Microscopy (SEM) done using Tescan, Vega 3 LMU with Oxford
- 2 EDX detector (Inca XmaW20).

3 Fabrication of the electrochemical sensor

4 Guanine hydrazide was immobilized on the APTES coated iron oxide 5 nanoparticles using glutaraldehyde (GA) as a crosslinking agent. Briefly, a 6 1 mg/mL solution of GH dissolved in DMSO was mixed with 100 mg dm⁻³ 7 of the APTES coated nanoparticles. 20 mm³ of this mixture were dropped on 8 a clean BDD electrode and left for 20 minutes in saturated glutaraldehyde 9 vapor. The electrode was left to dry at room temperature. Afterwards, it was 10 placed on the electrochemical cell where a 300 mT cylindrical magnet will 11 permit the paramagnetic nanoparticles to remain attracted to the electrode's 12 surface.

13 Electrochemical measurements

All electrochemical measurements were performed using a PalmSens4. A conventional three-electrode system was employed: a bare or modified boron doped diamond electrode as the working electrode, Platinum plate as a counter electrode and Ag/AgCl as a reference electrode. Citrate buffered solutions (0.1 mol dm⁻³, containing 0.5 mol dm⁻³ KCl, pH 4) were used as electrolytes for all electrochemical measurements. The technique used in all the measurements was square wave voltammetry with the

1	following optimized conditions: frequency 50 Hz; amplitude 50 mV; step
2	potential 10 mV; time of equilibration 2 minutes.
3	
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7	
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11 12 13 14 15 16 17 18 19 20 21 22 23	
24	Figure Captions
25	Fig. 1 Schematic diagram showing the steps of guanine hydrazide synthesis
26	Fig. 2 (a) FTIR spectrum, (b) 1 H NMR (c) 13 C NMR and (d) HRMS of

27 guanine hydrazide

- 1 Fig. 3 Schematic diagram showing the process of synthesizing the GH-
- 2 APTES-Fe₃O₄ nanoparticles
- 3 Fig. 4 (A) FTIR spectrum of (a) bare Fe_3O_4 NPs, (b) APTES- Fe_3O_4 NPs and
- 4 (c) GH-APTES-Fe₃O₄NPs, (B) XRD pattern, (C) SEM image and (D) EDX
- 5 spectrum of APTES-Fe₃O₄ nanoparticles
- 6 Fig. 5 SWV of (a) APTES-Fe₃O₄ nanoparticles and (b) APTES-Fe₃O₄
- 7 nanoparticles functionalized with GH
- 8 Fig. 6 Oxidation reaction of guanine, in aqueous medium

9 Fig. 7 SWV of (a) Cd (II) at the GH-APTES-Fe₃O₄ NP electrode over a

- 10 concentration range of $0 2.5 \ \mu mol \ dm^{-3}$, (b) Pb (II) at the GH-APTES-
- 11 Fe₃O₄ NP electrode over a concentration range of $0 1 \mu mol dm^{-3}$ and (c) Cu
- 12 (II) at the GH-APTES-Fe₃O₄ NP electrode over a concentration range of 0 -
- 13 3.5 μmol dm⁻³
- 14 **Fig. 8** Langmuir adsorption isotherms for (a) Cd^{2+} in the range of 13 265
- 15 mg dm⁻³, (c) Cu^{2+} in the range of 13 229 mg dm⁻³ and (e) Pb²⁺ in the range
- 16 of $5-202 \text{ mg dm}^{-3}$. Freundlich adsorption isotherms for (b) Cd²⁺ in the range
- $17 \quad 40 142 \text{ mg dm}^{-3}$, (d) Cu²⁺ in the range $26 130 \text{ mg dm}^{-3}$ and (f) Pb²⁺ in the
- 18 range 31 124 mg dm⁻³
- 19 Fig. 9 Linear ranges of (a) cadmium, (b) lead and (c) copper at the GH-
- 20 APTES-Fe₃O₄NP electrode
- 21

1 Table 1 Constant parameters and correlation coefficients calculated for

Isotherm	La	ngmuir			Freundlich		
Constants	$Q_{ heta}$	b	R_L	r ²	n	<i>k</i> _{<i>F</i>}	r ²
	(µg g ⁻¹)	(L µg ⁻¹)				(µg.g ⁻¹)	
Cd ²⁺	280.1	0.087	0.0394	0.9999	6.06	118.2	0.9614
Cu ²⁺	400	0.0707	0.0341	0.9998	5.447	150.7	0.9796
Pb ²⁺	343.6	0.1046	0.0271	0.9999	7.776	174.3	0.9961

2 different adsorption models for Cu^{2+} , Cd^{2+} and Pb^{2+} adsorption

3

4 Table 2 Comparison of the linear detection range and sensitivity with

5 previously reported values for the electrochemical detection of Cd (II)

Electrode	Method	Concentration	LOD	Sensitivity	Reference
modification		Range (µmol	(µmol	(µA µmol ⁻¹	
		dm ⁻³)	dm ⁻³)	dm ³)	
Polydopamine-Fe ₃ O ₄	SWASV	0.02 - 0.59	9.2×10 ⁻⁵	196	[39]
Reduced graphene	SWASV	0 - 0.8	0.056	14.82	[40]
oxide-Fe ₃ O ₄					
Chitosan-Fe ₃ O ₄	SWASV	1.2 - 1.7	0.0392	8.11	[1]
Fe ₃ O ₄	SWASV	0.3 – 1.3	0.154	3.18	[41]

Terephthalic acid-	SWASV	0.4 - 1.1	0.2	12.15	[42]
Fe ₃ O ₄					
GH-APTES-Fe ₃ O ₄	SWV	0.232 - 0.809	0.077	101.4	Current
					work

- 1
- 2

3 Table 3 Comparison of the current concentration range and sensitivity with

Electrode	Method	Concentration	LOD	Sensitivity	Reference
modification		Range (µmol	(µmol	(µA µmol ⁻¹	
		dm ⁻³)	dm ⁻³)	dm ³)	
Amine–Fe ₃ O ₄	SWASV	0.5 – 8	0.15	10.07	[43]
Polydopamine-	SWASV	5×10 ⁻³ – 0.6	1.4×10 ⁻⁶	235	[39]
Fe ₃ O ₄					
Fe ₃ O ₄	SWASV	0.3 – 1.3	0.119	14.9	[41]
Chitosan-Fe ₃ O ₄	SWASV	0.1 – 1.3	0.0422	50.6	[1]
Terephthalic	SWASV	0.4 - 1.1	0.04	8.56	[42]
acid-Fe ₃ O ₄					
GH-APTES-	SWV	0.0483 - 0.497	0.011	156	Current
Fe ₃ O ₄					work
5					

4 previously reported values for the electrochemical detection of Pb (II)

1 Figure 1





2













4

5 Figure 4







- 1 Full Scale 2200 cts Cursor: 0.000
- 2 Figure 5









1 Graphical abstract

