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# A New Reactive Absorption Model Using Extents of Reaction and Activities. I. Application to Alkaline-salts-CO<sub>2</sub> Systems

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## <u>Keywords</u>

Gas-liquid absorption, Kinetic model, Non ideality, Hydroxide solutions, CO<sub>2</sub> absorption

## <u>Abstract</u>

 $CO_2$  absorption into basic aqueous solutions is widely used for  $CO_2$  separation from gas streams (e.g., for natural gas purification).  $CO_2$  loading and ionic strength increase significantly along industrial columns. In absorption modelling, deviation from ideality should then be considered.

This study implements a general steady-state model for reactive gas-liquid absorption. Firstly, equilibrium relations, Nernst-Planck diffusion fluxes and reaction rates are written based on activities. Secondly, local fluxes are related by stochiometric constraints through extents of reaction.

In a first case study, the model, together with an appropriate thermodynamic representation, was applied with the stagnant film theory (Whitman, 1923) to alkaline salts-water-CO<sub>2</sub> systems. The following Arrhenius expression was found for the direct kinetic constant of reaction  $CO_2 + HO^- \leftrightarrow HCO_3^-$ : lnk (m<sup>3</sup>.mol<sup>-1</sup>.s<sup>-1</sup>) = 19.84 - 5248.8/T (K) – 12% overall AAD. This kinetic law can be used in any system involving this reaction (e.g., aqueous amine solutions). This part I paves the way to the further study of  $CO_2$  absorption into aqueous amine solutions.

# <u>Nomenclature</u>

1	А	Logarithm of Arrhenius pre-exponential factor
2	a <sub>i</sub>	Activity of species i (mol.m <sup>-3</sup> )
3	a <sub>1,0</sub>	Activity of absorbed gas at gas-liquid interface (mol.m <sup>-3</sup> )
4	Ci	Molar concentration of species i (mol.m <sup>-3</sup> )
5	C <sub>1,0</sub>	Concentration of absorbed gas at gas-liquid interface (mol.m <sup>-3</sup> )
6	C <sub>abs,tot</sub>	Total absorptive species concentration (mol.m <sup>-3</sup> )
7	Ci	Normalised concentration of species i (-)
8	C <sub>b,i</sub>	Liquid bulk concentration of species i (mol.m <sup>-3</sup> )
9	C <sub>N,i</sub>	Normalisation concentration of species i (mol.m <sup>-3</sup> )
10	Di	Diffusivity of species i in solution (mol.m <sup>-3</sup> )
11	E	Enhancement factor (-)
12	E <sub>A</sub>	Activation energy (J.mol <sup>-1</sup> )
13	f <sub>obj</sub>	Optimisation objective function (-)
14	G	Gibbs free energy (J)
15	H <sub>1</sub>	Henry's constant of absorbed gas at infinite dilution (Pa)
16	Ha <sub>j</sub>	Hatta number of reaction j (-)
17	j <sub>i</sub>	Flux of species i through liquid film (mol.m <sup>-2</sup> .s <sup>-1</sup> )
18	<b>ј</b> 1т,0	Total absorption flux of absorbed gas (mol.m <sup>-2</sup> .s <sup>-1</sup> )
19	k <sub>j</sub>	Reaction j kinetic constant
20	K <sub>j</sub>	Reaction j equilibrium constant
21	K <sub>N,j</sub>	Reaction j normalising equilibrium constant
22	$k_{L,1}$	Liquid-side mass transfer coefficient regarding the absorbed gas (m.s <sup>-1</sup> )
23	Μ	Matrix of the differential problem for concentration-related equations
24	n <sub>c</sub>	Number of solute constituents in liquid phase
25	n <sub>i</sub>	mole number of species i (mol)
26	n <sub>R,kin</sub>	Number of finite-rate reactions in the reaction mechanism
27	n <sub>R,eq</sub>	Number of reactions at equilibrium in the reaction mechanism
28	o <sup>D</sup> ij	Direct kinetic order of species i in reaction j (-)
29	0 <sup>R</sup> ij	Reverse kinetic order of species i in reaction j (-)
30	Р	Pressure (Pa)
31	p <sub>1,0</sub>	Pressure of absorbed gas at gas-liquid interface (Pa)
32	r <sub>j</sub>	Reaction rate of reaction j (mol.m <sup>-3</sup> .s <sup>-1</sup> )
33	т	Temperature (K)
34	x <sub>i</sub>	Mole fraction of species I (-)
35	X <sub>i</sub>	Vector of the differential problem for concentration-related equations
36	z	Distance from interface (m)
37	Z	Normalised distance from interface (z)
38	Zi	Number of charges of species I (-)
39	$Z_{Di}$	Diffusion-concentration ratio of species i (-)
40	Consta	nts
41	$\mathcal{F}$	Faraday constant (C.mol <sup>-1</sup> )

42 R Ideal gas constant (J.mol<sup>-1</sup>.K<sup>-1</sup>)

43	Greek	letters
44	γi	Activity coefficient of species i (-)
45	$\delta_{ij}$	Kronecker delta between species i and j
46	$\delta_{\text{L},1}$	Film thickness regarding the absorbed gas (m)
47	$\mu_i$	Chemical potential of species i (J.mol <sup>-1</sup> )
48	$v_{ij}$	Stochiometric coefficient of species i in reaction j (-)
49	ψ	Local electric potential (V)
50	ξj	Extent of reaction j (-)
51 52	Subsc. app	ripts and superscripts Apparent
53	calc	Calculated
54	exp	Experimental
55	REF	Reference state
56	1	Absorbed gas
57		Molality-based asymmetric thermodynamic convention (see Supplementary Material section
58	2)	
59	~c	Molar concentration-based symmetric thermodynamic convention

### 60 1. Introduction

61 Chemical absorption into basic solutions has been in use since the 1930s for CO<sub>2</sub> separation from raw 62 natural gas or from flue gases (Bottoms, 1931). Solvent screening and design are at the heart of the 63 research effort to increase both CO<sub>2</sub> absorption rate and capacity, and to decrease solvent 64 regeneration duty (e.g., aqueous amine blends). Industrial absorption column design with a given 65 solvent requires the identification of the dissolved gas reaction mechanism. To do so, reactions at 66 equilibrium (e.g., acid-base reactions) and finite-rate reactions need to be studied at pressure, 67 temperature and composition conditions encountered in the absorption column.

- Kinetic parameters are typically obtained from absorption or desorption flux measurements in laboratory gas-liquid contactors of known mass transfer coefficient and interfacial area. Most kinetic studies deal with absorption experiments into unloaded basic solutions. Only one irreversible reaction is considered (e.g., (Pohorecki & Moniuk, 1988), (Pani, et al., 1997), (Derks, et al., 2006)). Such studies disregard the increase in ionic strength and non-ideality when CO<sub>2</sub> is absorbed, due to bicarbonate, carbonate, and carbamate ion production. However, absorption models based on activities rather than molarities should lead to more relevant process simulation.
- CO<sub>2</sub> absorption modelling into aqueous amine solutions first requires the study of CO<sub>2</sub> reaction with HO<sup>-</sup> ion. Indeed, this reaction takes place in all aqueous systems with CO<sub>2</sub>. Furthermore, kinetic constants between CO<sub>2</sub> and tertiary amines can be expressed as a function of amine pKa (Couchaux, et al., 2014). Through molecular simulation, (Rozanska, et al., 2021) show that CO<sub>2</sub> kinetics in aqueous tertiary amine solutions can be written as a function of CO<sub>2</sub> and HO<sup>-</sup> concentrations only.</sup>
- 80 For CO<sub>2</sub> absorption into aqueous alkaline salts solutions, researchers choose a one-irreversible-81 reaction mechanism. In this case, non-ideality is usually included. In early studies such as (Pinsent, et 82 al., 1956), (Hikita, et al., 1976), (Augugliaro & Rizzuti, 1987), (Pohorecki & Moniuk, 1988), and (Kucka, 83 et al., 2002), empirical ionic-strength, alkaline-ion-dependent kinetic constant and CO<sub>2</sub> solubility 84 relations are developed. When studying CO<sub>2</sub> absorption into 30wt% K<sub>2</sub>CO<sub>3</sub> solutions, (Thee, et al., 85 2012) fit one Arrhenius-type expression to their kinetic data between 313 and 353 K. The expression 86 is found coherent with (Knuutila, et al., 2010) correlation when considering identical salt and 87 concentration range. Under relatively diluted  $Na_2CO_3$  and  $K_2CO_3$  concentrations (300 mol/m<sup>3</sup>), 88 (Penders-van Elk, et al., 2016) find good agreement of their CO<sub>2</sub> absorption data between 283 and 89 313 K with pKa-dependent Bronsted kinetic law for tertiary amines. Following (Kucka, et al., 2002), 90 (Haubrock, et al., 2005), (Haubrock, et al., 2007), (Knuutila, et al., 2010) and (Gondal, et al., 2016) 91 develop an activity-based kinetic rate expression, using Pitzer-Debye-Hückel model, empirical activity 92 coefficients and e-NRTL model, respectively. Thus, a unique Arrhenius-type expression applies to all 93 systems (with  $Li^+$ ,  $Na^+$  or  $K^+$ ).
- 94 (Sheng, et al., 2019) apply current knowledge on CO<sub>2</sub> absorption into aqueous NaOH solutions to 95 measure pilot effective mass transfer area. The influence of five literature sources is evaluated for 96 kinetic and physical properties estimation. According to this study, different Arrhenius parameters 97 result in large variations for the effective mass transfer area estimation. Additionally, inconsistencies 98 between kinetics and physical parameter sources should be avoided.
- Absorption involves reversible finite-rate reactions and acid-base reactions at equilibrium. Speciessimultaneously diffuse to and from gas-liquid interface and react. The diffusion driving force is the

101 molar concentration gradient in the widely used semi-empirical Fick diffusion law (e.g., (Versteeg, et al., 1989), (Bishnoi & Rochelle, 2002), (Derks, et al., 2006)). However, Maxwell-Stefan theory 102 103 (Ahmadi, et al., 2010), or Nernst-Planck equation in case of unknown binary diffusivities (e.g., 104 (Glasscock & Rochelle, 1989) and (Littel, et al., 1991)) are more suitable to highly non ideal 105 multiconstituent electrolytic systems. The diffusion driving force is then the chemical potential 106 gradient (Taylor & Krishna, 1993). (Thomas, et al., 2016) model CO<sub>2</sub> diffusion-reaction into stagnant 107 water and aqueous alkaline salt solutions with one irreversible reaction. Instead of individual ion 108 diffusion, salt diffusion across the liquid medium is represented. Thus, diffusivities are decoupled 109 without the need for an electric potential gradient to keep electroneutrality.

110 Mass transfer problem formulation with any transient or steady-state mass transfer theory results in 111 a set of local balances of each species. The differential problem is a boundary value problem. At the 112 gas-liquid interface, absorbed gas physical equilibrium is assumed and in the liquid bulk, chemical 113 equilibrium is assumed. In some models, acid-base reactions are represented as reversible reactions 114 with a very high reaction rate constant (e.g., (Glasscock & Rochelle, 1989), (Versteeg, et al., 1989), 115 (Ahmadi, et al., 2010)). The problem then includes a system of n<sub>c</sub> second order ordinary or partial 116 differential equations. In other models, diffusion-reaction flux coupling is explicitly considered. The 117 number of differential equations decreases, but this formulation lacks generality (e.g., (Rinker, et al., 1995), (Bishnoi & Rochelle, 2002), (Servia, et al., 2014)). This second formulation ensures the n<sub>eg</sub> acid-118 119 base equilibria by the addition of n<sub>eq</sub> algebraic equations: a differential-algebraic problem needs to 120 be solved.

121 The purpose of this article is to propose a general reactive absorption model entirely based on 122 activities, unlike previous reactive absorption models. Indeed, previous models are either based on 123 Nernst-Planck diffusion with a complete mass transfer model, but activity coefficients are assumed equal to unity or, like (Gondal, et al., 2016), use a simplified mass transfer model with Fick's diffusion 124 125 law and constant activity coefficients. Activity coefficients are then calculated at equilibrium with a 126 thermodynamic model. For steady-state absorption, extent-of-reaction variables are introduced to 127 take advantage of flux coupling (section 2). This has two advantages compared to usual model 128 formulations: (1) the number of equations is reduced while still maintaining a general formulation, 129 and (2) solving a system of algebraic equations and substituting the results into the differential 130 equations takes care of acid-base equilibria. Extents of reactions have yet to be implemented in 131 reactive absorption problems.

This model is then applied to  $CO_2$  absorption in aqueous alkaline salts solutions with a full reaction mechanism considering two reversible reactions. This third section aims at fitting a unique Arrheniustype expression for the direct kinetic constant of reaction  $CO_2 + HO^- \leftrightarrow HCO_3^-$  to literature  $CO_2$ absorption flux data with all three alkaline salt counter ions. This part I paves the way to the further study of  $CO_2$  absorption into aqueous amine solutions. Indeed, this reaction takes place in all aqueous solvents. It is necessary to characterise it before focusing on more complex systems.

138 2. Modelling framework

139 2.1. <u>Activity-based diffusion, fluxes, rates of reaction and equilibria</u>

Equation 1 gives the Nernst-Planck expression for the diffusion flux of a given species i (Taylor &Krishna, 1993).

$$j_i = -\frac{D_i c_i}{RT} (\nabla \mu_i + z_i \mathcal{F} \nabla \psi)$$
 Eq. (1)

142 In case of different diffusion coefficients of ionic species, the electric potential  $\psi$  is to ensure local 143 electroneutrality through the liquid film ( $\sum_{i \text{ species }} z_i j_i = 0$ ):

$$\nabla \psi = -\frac{\sum_{i \text{ species } z_i D_i c_i \nabla \mu_i}{F \sum_{i \text{ species } z_i^2 D_i c_i}}$$
 Eq. (2)

144 The chemical potential of a given species i,  $\mu_i$ , is defined as its partial molar Gibbs free energy 145 (Prausnitz, et al., 1998). A thermodynamic convention consists in defining a reference state, then the 146 activity  $a_i$  (eq. 3) and the activity coefficient (eq. SM16 in Supplementary Material) of the species i.

$$\mu_i(T, P, \mathbf{n}) = \left(\frac{\partial G}{\partial n_i}\right)_{T, P, n_{j \neq i}} = \mu_i^{REF}(T, P) + RT \ln(a_i)$$
Eq. (3)

147  $\mu_i^{\text{REF}}$  and  $a_i$  are expressed in the same chosen thermodynamic convention. The dimension of the 148 activity depends on the composition scale, such as molalities, molarities, mass fractions, etc. We 149 express reaction rates as functions of activities as well (eq. 4).

$$r_{j} = k_{j} \left( \prod_{i \text{ species}} a_{i}^{o_{ij}^{D}} - \frac{1}{K_{j}} \prod_{i \text{ species}} a_{i}^{o_{ij}^{R}} \right)$$
Eq. (4)

Kinetic constants k<sub>j</sub> and equilibrium constants K<sub>j</sub> also need to be expressed according to the chosen
 thermodynamic convention. For reactions assumed to be at equilibrium:

$$K_j = \prod_{i \text{ species}} a_i^{\nu_{ij}} \qquad \qquad \text{Eq. (5)}$$

152  $v_{ij}$  is the stoichiometric coefficient of species i in reaction j. Consistency between orders of reaction 153 and stoichiometric coefficients (eq. 4 and 5) yields  $o_{ij}^R - o_{ij}^D = v_{ij}$ .

Absorption problems are typically one dimension. At gas-liquid interface, the partial pressure and the
activity of the absorbed gas (subscript 1) are assumed at equilibrium. A Henry's relation is obtained
(Prausnitz, et al., 1998), assuming asymmetric thermodynamic convention for the absorbed gas (eq.
6).

$$H_1 = \frac{p_{1,0}}{a_{1,0}}$$
 Eq. (6)

158

## 159 2.2. Extents of reaction and resulting system of equations

For a steady-state mass transfer model such as the film model, a new algebraic variable is introduced: the extent of each reaction. The extent  $\xi_j$  of a given reaction j measures its progress in the liquid film from the interface to a distance z. The relation between fluxes and extents of reaction is dictated by stochiometric constraints. The flux of a given species i through the film, as a function of the extents of reaction, is given in eq. 7.

$$j_{i} = j_{1T,0} \left( \delta_{1i} + \sum_{j \text{ reactions}} v_{ij} \xi_{j} \right)$$
 Eq. (7)

165 Where  $j_{1T,0}$  is the total flux of gas absorbed. Due to the normalisation by  $j_{1T,0}$ ,  $\xi_j$  is dimensionless. This 166 expression sums up the contribution of extents both finite-rate reactions and reactions assumed at 167 equilibrium where species i is either a reactant or a product. At the gas-liquid interface (z=0), only 168 reactions at equilibrium may have a non-zero extent. For non-volatile species, the extents of all 169 finite-rate reactions must be zero. Eq. 7 then becomes eq. 8 at the gas-liquid interface.

$$j_i(z=0) = j_{1T,0} \left( \delta_{1i} + \sum_{k \text{ reactions}} v_{ik} \xi_k(z=0) \right)$$
 Eq. (8)

170 Where index k refers to a reaction assumed at equilibrium. As a result, the flux  $j_1(z=0)$  is not only the 171 flux of 1 crossing the interface,  $j_{1T,0}$ . It is modified by all reactions at equilibrium with a non-zero 172 extent of reaction at the gas-liquid interface (eq. 9).

$$j_1(z=0) = j_{1T,0} \left( 1 + \sum_{k \text{ reactions}} \nu_{1k} \xi_k(z=0) \right)$$
 Eq. (9)

The local variation of the extent of a finite-rate reaction, is by definition proportional to its local rate (eq. 10).

175 Combining equations 3 and 7 yields eq. 11.

$$j_{i} = j_{1T,0} \left( \delta_{1i} + \sum_{j \text{ reactions}} \nu_{ij} \xi_{j} \right) = -\frac{D_{i} c_{i}}{RT} (\nabla \mu_{i} + z_{i} \mathcal{P} \nabla \psi)$$
Eq. (11)

176 Differentiation of equilibrium relation (eq. 5) of reaction k generates an algebraic relation between 177 its extent of reaction and the extents of other reactions. Substituting the expression of  $\nabla \mu_i$  from eq. 178 11:

$$0 = \sum_{i \text{ species}} v_{ik} \left( j_{1T,0} \frac{\left( \delta_{1i} + \sum_{j \text{ reactions}} v_{ij} \xi_j \right)}{D_i c_i} - z_i \mathcal{F} \nabla \psi \right)$$
Eq. (12)

179 Where index k refers to a given reaction at equilibrium. The resulting system of  $n_{R,eq}$  algebraic 180 equations can be solved separately. The resulting extents  $\xi_k$  of reactions at equilibrium are 181 substituted back into the differential equations.

### 182 2.3. <u>Thermodynamic framework and mass transfer and diffusion assumptions</u>

The mass transfer framework used in this work is the stagnant film theory (Whitman, 1923). In aqueous solution, the asymmetric convention is more relevant for solute species. The selected thermodynamic model is described in the Supplementary Material section 2. Ionic diffusivities are known only in water at infinite dilution at 25°C. As a first application, equal diffusivities are assumed for all ions in solution (as did e.g., (Derks, et al., 2006), (Ahmadi, et al., 2010), (Luo, et al., 2015)). The electric potential term is thus neglected. Since the problem is spatially one-dimensional (in z), eq. 13 is obtained.

$$j_{i} = -D_{i} \sum_{k \text{ species}} \left( \delta_{ik} + c_{i} \frac{\partial \ln \ddot{\gamma}_{i}}{\partial c_{k}} \right) \frac{dc_{k}}{dz}$$
 Eq. (13)

- 190 Where  $\ddot{\gamma}_i$  is the activity coefficient of species i in the molality-based asymmetric convention, and  $\delta_{ik}$  is
- 191 the Kronecker delta (1 if i=k, 0 otherwise). Reactions rates can be expressed as:

$$r_{j} = \tilde{k}_{j}^{c} \left( \prod_{i \text{ species}} (\tilde{\gamma}_{i}c_{i})^{o_{ij}^{D}} - \frac{1}{\tilde{K}_{j}^{c}} \prod_{i \text{ species}} (\tilde{\gamma}_{i}c_{i})^{o_{ij}^{R}} \right)$$
Eq. (14)

192 Where  $\tilde{\gamma}_i = \frac{\tilde{\gamma}_i}{x_{H_20}}$  is the activity coefficient of species i in the molarity-based asymmetric convention. 193  $\tilde{k}_j^c(T)$  is the forward kinetic constant of reaction j  $((m^3.mol^{-1})^{\sum o_{ij}^D}.s^{-1})$  and  $\tilde{K}_j^c(T)$  is the 194 equilibrium constant of reaction j  $((m^3.mol^{-1})^{\sum v_{ij}})$ . Both are written in the same convention.

195 Reaction equilibrium equations are then written as follows.

$$\widetilde{K}_{j}^{c} = \prod_{i \text{ species}} (\widetilde{\gamma}_{i} c_{i})^{\nu_{ij}}$$
 Eq. (15)

196 Combining equations 3 and 14 yields eq. 16.

$$j_{i} = j_{1T,0} \left( \delta_{1i} + \sum_{j \text{ reactions}} v_{ij} \xi_{j} \right) = -D_{i} \sum_{k \text{ species}} \left( \delta_{ik} + c_{i} \frac{\partial \ln \ddot{\gamma}_{i}}{\partial c_{k}} \right) \frac{dc_{k}}{dz}$$
Eq. (16)

197 Without electric potential gradient, the system of equations defined by eq. 12 becomes linear198 regarding the extents of reactions at equilibrium:

$$0 = \frac{\nu_{1k}}{c_1 D_1} + \sum_{i \text{ species}} \frac{\nu_{ik}}{c_i D_i} \sum_{j \text{ reactions}} \nu_{ij} \xi_j$$
 Eq. (17)

199 In the chosen thermodynamic convention, Henry's relation for the absorbed gas at the interface200 (z=0) is expressed as follows:

$$p_{1,0} = \frac{H_1(T, P, \mathbf{n})}{c_{H_2 0}} \ddot{\gamma}_1 c_{1,0}$$
 Eq. (18)

- As experimental pressure is lower than 1 bar in the data considered in section 3, gas-phase is assumed ideal.
- 203 2.4. <u>Solution</u>

Before writing the differential problem to solve, dimensionless variables (eq. 19) and parameters (eq.20) are introduced.

$$Z = \frac{z}{\delta_{L,1}} = \frac{k_{L,1}}{D_1} z$$

$$C_i = \frac{c_i}{c_{N,i}}$$
Eq. (19)

206 Where  $c_{N,i}$  is the normalising concentration, such as if i=1,  $c_{N,i} = c_{1,0}$ , and  $c_{N,i} = c_{abs,tot}$  otherwise.  $c_{abs,tot}$ 207 is the total absorptive species concentration (mol.m<sup>-3</sup>).

$$Z_{D,i} = \frac{c_{N,i}D_i}{c_{1,0}D_1}$$

$$E = \frac{j_{1T,0}}{k_{L,1}c_{1,0}}$$

$$Ha_j^2 = \frac{D_1 \tilde{k}_j^c \left(\prod_i (c_{N,i})^{o_{ij}^D}\right)}{k_{L,1}^2 c_{1,0}}$$

$$K_{N,j} = \prod_i (c_{N,i})^{v_{ij}}$$
Eq. (20)

208 Therefore, the flux of species i is expressed as a matrix product (eq. 21).

$$j_{i} = j_{1T,0} \left( \delta_{1i} + \sum_{j \text{ reactions}} v_{ij} \xi_{j} \right) = -j_{1T,0} \frac{Z_{D,i}}{E} \sum_{k \text{ species}} M_{ik} \frac{dC_{k}}{dZ}$$

$$M_{ik} = \left( \delta_{ik} + C_{i} \frac{\partial \ln \ddot{\gamma}_{i}}{\partial C_{k}} \right)$$
Eq. (21)

209 Square matrix M inversion yields:

$$\frac{dC_k}{dZ} = -E\left(\sum_{i \text{ species}} (M^{-1})_{ki} X_i\right) \qquad k = 1..n_C$$

$$\frac{d\xi_j}{dZ} = \frac{\text{Ha}_j^2}{E}\left(\prod_{i \text{ species}} (\tilde{\gamma}_i C_i)^{o_{ij}^D} - \frac{K_{N,j}}{\tilde{K}_j^C} \prod_{i \text{ species}} (\tilde{\gamma}_i C_i)^{o_{ij}^R}\right) \qquad j = 1..n_{\text{R,kin}}$$
Eq. (22)

210 Where  $M^{-1}$  is the inverse of matrix M and vector X is defined as  $X = \left(\frac{\delta_{1i} + \sum_{j \text{ reactions }} v_{ij}\xi_j}{Z_{Di}}\right)_i$ .

At each resolution step, an algebraic system of equations is solved independently to obtain the extents of acid-base reactions (eq. 23). These values are used in the differential equations.

$$0 = \frac{v_{1k}}{C_1} + \sum_{i \text{ species}} \frac{v_{ik}}{Z_{D,i}C_i} \sum_{j \text{ reactions}} v_{ij}\xi_j$$
 Eq. (23)

213 Boundary conditions are presented in Table 1.

#### 214 Table 1 Film model boundary conditions with dimensionless variables

$Z = 0 \ (z = 0)$	$Z = 1 \ (z = \delta_{L,1})$
$C_1 = 1$	$C_i = \frac{c_{b,i}}{c_{N,i}}, \forall i \text{ species in solution}$
$\xi_j = 0, \forall j$ finite-rate reaction	-

215

216 This formulation with Nernst-Planck equation for molecular diffusion and activity-based rate 217 expressions consists in a system of n<sub>C</sub>+n<sub>R,kin</sub> first-order ordinary differential equations, thanks to steady-state extents of reaction, whereas writing local balances per species leads to a system of nc 218 219 second-order differential equations. Extents of reactions at equilibrium are calculated from a set of 220 n<sub>R,eq</sub> algebraic equations (eq. 23). Equilibrium relations, Nernst-Planck diffusion fluxes and reaction 221 rates are activity-based. Therefore, this model reconciles kinetics and thermodynamics especially 222 close to equilibrium. Simulations are performed in Matlab R2013a. The bvp4c function available in 223 the software is used to solve the differential problem (Kierzenka & Shampine, 2001).

#### 224 3. Application to alkaline salts-water-CO<sub>2</sub> systems

Four species are considered:  $CO_2$ ,  $HCO_3^-$ ,  $CO_3^{2-}$  and  $HO^-$  (n<sub>c</sub>=4). The alkaline ion does not take part in the reaction scheme, so that its concentration is constant through the film because the electric potential is disregarded. However, this ion contributes to non-ideality. The mechanism consists in two reversible reactions (n<sub>R</sub>=2):

$$CO_2 + HO^2 \leftrightarrow HCO_3^2$$
 (R-I)

$$HCO_3^{-} + HO^{-} = CO_3^{-2^{-}} + H_2O$$
 (R-II)

229 Reaction R-I is reversible and finite-rate  $(n_{R,kin}=1)$ :

$$r_{\rm I} = r_{\rm (CO_2,HO^-)} = \tilde{k}_{\rm (CO_2,HO^-)}^c \left( \tilde{\gamma}_{\rm CO_2} \tilde{\gamma}_{\rm (HO^-)} c_{\rm CO_2} c_{\rm (HO^-)} - \frac{\tilde{\gamma}_{\rm (HCO_3^-)} c_{\rm (HO_3^-)}}{\tilde{K}_{\rm (CO_2,HO^-)}^c} \right)$$
Eq. (24)

230 Where  $\tilde{k}_{(CO_2,HO^-)}^c$  is its forward kinetic constant – to determine – (m<sup>3</sup>.mol<sup>-1</sup>.s<sup>-1</sup>) and  $\tilde{K}_{(CO_2,HO^-)}^c$  its 231 equilibrium constant (m<sup>3</sup>.mol<sup>-1</sup>).

In addition, the acid-base equilibrium between  $HCO_3^{-1}$  and  $CO_3^{-2^{-1}}$  is considered (reaction R-II;  $n_{R,eq}=1$ ).

The equilibrium constant parameters are taken from (Kamps, et al., 2001). Solution density is taken from (Laliberté & Cooper, 2004). Liquid phase CO<sub>2</sub> diffusivity is calculated with the Stokes-Einstein relation (Versteeg & Van Swaaij, 1988). (Knuutila, et al., 2010) use this relation as well in combination with CO<sub>2</sub> diffusivity in pure water from (Danckwerts & Lannus, 1970). Viscosity is taken from (Laliberté, 2007). Hydroxide ion diffusivity is set to  $1.7 \times D_{CO2}$  based on the ratio proposed by (Hikita, et al., 1976), and  $D_{(HCO3-)}$  and  $D_{(CO32-)}$  are supposed equal to  $D_{(HO-)}$  (see Supplementary Material section 1 for detailed correlations).

As a base case, the mass transfer model was tested with no deviation from ideality ( $\gamma$ =1) and (Weisenberger & Schumpe, 1996) CO<sub>2</sub> apparent solubility ("concentration-based model"). Secondly, the activity-based model was implemented. Few  $CO_2$  absorption flux measurements are available in these systems. (Kucka, et al., 2002), (Pohorecki & Moniuk, 1988) and (Knuutila, et al., 2010) do not publish raw experimental data, such as  $CO_2$  partial pressure. (Gondal, et al., 2016) use raw absorption flux data from (Gondal, et al., 2015) in alkaline hydroxide solutions and from (Knuutila, et al., 2010) in alkaline carbonate solutions. Note that data from (Knuutila, et al., 2010) is not available in open literature. The only other raw dataset on  $CO_2$  absorption in sodium hydroxide and carbonate solutions was found in (Hikita, et al., 1976)

- 249 (see Table 2).
- 250 Table 2 Available CO<sub>2</sub> absorption flux data in alkaline hydroxide and carbonate aqueous solutions

Ref, apparatus	$Alk^+$	x(Alk⁺,app)	x(HO <sup>-</sup> ,app)	x(CO <sub>2</sub> ,app)	Т (К)	P <sub>CO2</sub> (kPa)	n <sub>data</sub>
(Gondal, et al., 2015), SDC	Li⁺	1.8.10 <sup>-4</sup> -0.035	1.8.10 <sup>-4</sup> -0.035	0-3.5.10 <sup>-4</sup>	298-338	0.17-0.28	30
(Hikita, et al., 1976), Jet & WWC	Na⁺	2.3.10 <sup>-3</sup> -0.025	0-9.0.10 <sup>-3</sup>	2.3.10 <sup>-7</sup> -0.017	298-303	100	157
(Gondal, et al., 2015), SDC	Na⁺	1.8.10 <sup>-4</sup> -0.036	$1.8.10^{-4}$ -0.036	0-3.6.10 <sup>-4</sup>	298-333	0.18-0.32	38
(Gondal, et al., 2015), SDC	K⁺	1.6.10 <sup>-4</sup> -0.033	1.6.10 <sup>-4</sup> -0.033	0-3.6.10 <sup>-4</sup>	298-337	0.16-0.29	35

SDC: "String of Discs Column", Jet: "Laminar Jet Absorber", WWC: "Wetted-Wall Column", Alk<sup>+</sup>: alkaline ion.

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For kinetic constant regression, (Gondal, et al., 2015) data was used, as it encompasses all three alkaline ions between 298 and 338 K. To keep datasets with each counter ion approximately to the same size, (Hikita, et al., 1976) data for sodium hydroxide and carbonate solutions between 298 and 303 K was used as a validation dataset. (Hikita, et al., 1976) data therefore also serves as a model prediction test. Modelling each flux measurement requires the value of the liquid-side mass transfer coefficient, which is taken from each study.

Kinetic parameter optimisation of reaction R-I direct kinetic constant is achieved using the objectivefunction specified in eq. 25.

$$f_{\rm obj} = \sum_{i \, \rm data} \frac{1}{2} \left( \left( \frac{j_i^{\rm calc} - j_i^{\rm exp}}{j_i^{\rm exp} + \varepsilon} \right)^2 + \left( \frac{j_i^{\rm calc} - j_i^{\rm exp}}{j_i^{\rm calc} + \varepsilon} \right)^2 \right)$$
Eq. (25)

260  $\epsilon$  is a constant set to  $1.10^{-4}$  mol.m<sup>-2</sup>.s<sup>-1</sup> to mitigate the error value for very small absorption fluxes. 261 The two terms in the objective function (eq. 25) give this function a symmetrical behaviour for flux 262 underestimation and overestimation. Optimisation function fminsearch from Matlab R2013a was 263 used to fit the Arrhenius parameters of the kinetic constant (eq. 26). In fminsearch, the simplex 264 algorithm by (Lagarias, et al., 1998) is used.

$$\ln k_{(\rm CO_2, \rm HO^-)} = A + \frac{E_A}{RT}$$
 Eq. (26)

The fitted Arrhenius parameters as well as model performance is presented in Table 3: the average absolute deviation (AAD, eq. 27) and the bias (eq. 28) of regression (Gondal, et al., 2015) and validation (Hikita, et al., 1976) datasets. Model performance is also compared to (Gondal, et al., 2016).

$$AAD = \frac{1}{n_{data}} \sum_{i \text{ data}} \left| \frac{j_i^{\text{calc}} - j_i^{\text{exp}}}{j_i^{\text{exp}}} \right| \times 100$$
 Eq. (27)

bias = 
$$\frac{1}{n_{\text{data}}} \sum_{i \text{ data}} \frac{j_i^{\text{calc}} - j_i^{\text{exp}}}{j_i^{\text{exp}}} \times 100$$
 Eq. (28)

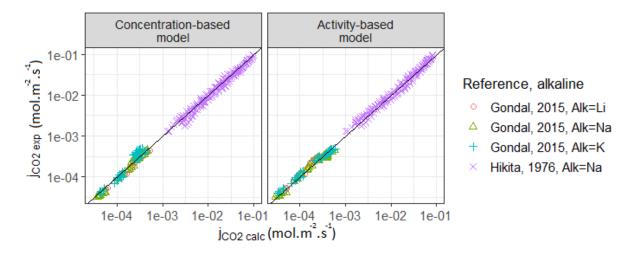
- Parameter standard errors are calculated by variance-covariance analysis, as described in the Supplementary Material section 3.
- 271 Table 3 Optimised Arrhenius parameters of reaction R-I direct kinetic constant over (Gondal, et al.,
- 272 2015) CO<sub>2</sub> absorption flux data in alkaline hydroxide solutions and performance

Model		Concentration-based	Activity-based	(Gondal, et al., 2016)
	А	19.387 ± 0.004	19.841 ± 0.003	19.53
Reference	$E_A/R$ (K)	-5020.99 ± 0.07	-5248.77 ± 0.05	-5111.2
(Gondal, et al.,	AAD (%)	14	10	15
2015) (Li⁺, Na⁺, K⁺)	Bias (%)	6.2	-2.1	-
(Hikita, et al.,	AAD (%)	17	14	-
1976) (Na⁺)	Bias (%)	12	-0.23	-

Remark: (Gondal, et al., 2016) do not provide parameter standard deviations.

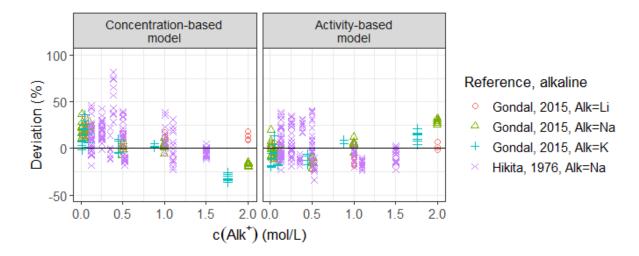
273

# 274 Parity and deviation plots are shown in Figure 1 and Figure 2, respectively.



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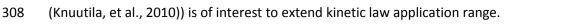
278 Figure 2 Reactive absorption model deviations for alkaline salts-water-CO<sub>2</sub> systems

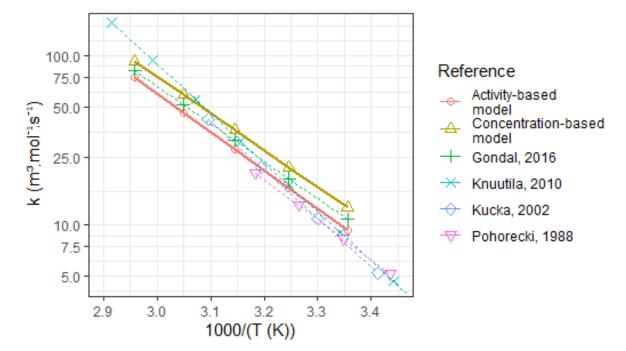
279 Considering solution non-ideality improves model performance for all three types of salts. The only exception is the data of (Gondal, et al., 2015) with NaOH data at 2 mol.L<sup>-1</sup>. All data from (Gondal, et 280 281 al., 2015) for all three salts (Li, Na and K-based) and most of the data from (Hikita, et al., 1976) are 282 represented within a +/-30% relative deviation by the activity-based model. Only 13 of all 157 283 validation data points have a relative deviation exceeding +/-30%. AAD over validation dataset is 284 13,7% and bias -0,2%, which is on par with the model performance of (Gondal et al., 2016). 285 Concentration-based model performance is on par with that of (Gondal, et al., 2016) when 286 considering the data of (Gondal, et al., 2015). Ideal solution is assumed in the concentration-based 287 model for reaction rate, diffusion flux, and chemical equilibria, but uses a correlation for CO<sub>2</sub> physical 288 solubility at gas-liquid interface. Salting-out effect on CO<sub>2</sub> solubility due to increasing ionic strength is 289 represented in both models: in the concentration-based model, by ion-concentration-dependent 290 (Weisenberger & Schumpe, 1996) correlation, and in the activity-based model, by the developed 291 thermodynamic model (see Supplementary Material section 2). Finally, only the activity-based model 292 unifies the representation of all equilibria with the chosen thermodynamic model, whereas the 293 concentration-based model is correlation-dependent.

The activity-based model is better at predicting (Hikita, et al., 1976) measurements in NaOH and Na<sub>2</sub>CO<sub>3</sub> aqueous solutions than the concentration-based model. This suggests the activity-based model can more easily be extrapolated outside regression area than the concentration-based model.

Arrhenius parameters are compared to previous studies in Figure 3, specifically (Pohorecki & Moniuk, 1988), (Kucka, et al., 2002), (Knuutila, et al., 2010) and (Gondal, et al., 2016). All these studies consider the solution non ideality in the reaction rate expression. Both concentration-based and activity-based models lead to close kinetic constants compared to (Gondal, et al., 2016). The Arrhenius law obtained with the activity-based model displays a -14% to -9% difference in the regression temperature range.

Fitted Arrhenius law extrapolation to higher temperatures leads to significant underestimation compared to the kinetic constant obtained by (Knuutila, et al., 2010) with their own experimental data and methodology – 94 vs. 157 m<sup>3</sup>.mol<sup>-1</sup>.s<sup>-1</sup> at 343 K. The value remains close to that of (Gondal, et al., 2016) – 102 m<sup>3</sup>.mol<sup>-1</sup>.s<sup>-1</sup>. Temperatures of 343 K and higher occur in industrial absorption 307 columns. Therefore, having access to raw absorption flux data at higher temperatures (e.g., found in





309

310 Figure 3 Arrhenius plot and comparison to previous studies

Activity-based model results at specific conditions are then examined. Theses conditions are listed in in Table 4. The four selected points cover the examined data experimental conditions (type and concentration of salt, temperature,  $CO_2$  partial pressure). As a reminder,  $CO_2$  concentration is normalised by  $c_{1,0}$ , its concentration at the interface, while all other concentrations are normalised by  $c_{abs,tot} = c_{(Alk+)}$ .

316 **Table 4** Selected points for concentration, extent of reaction and activity coefficient profile 317 simulation: system and conditions

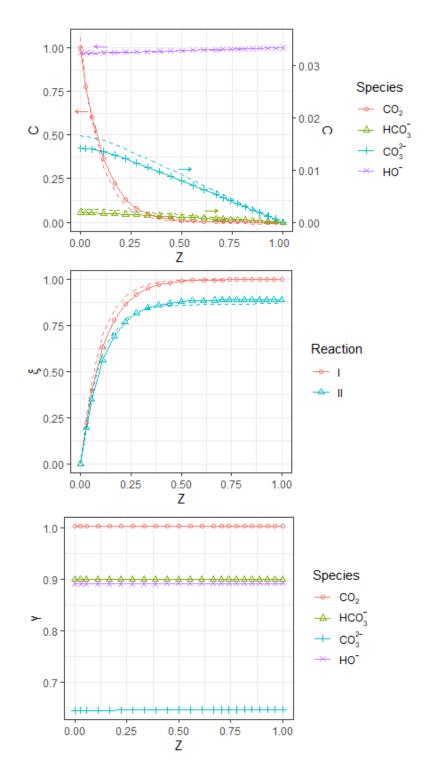
Figure	Reference	Salt	C <sub>salt</sub>	Т	P <sub>CO2</sub>	k <sub>L,1</sub>	C <sub>1,0</sub>
			(mol.L <sup>-1</sup> )	(K)	(kPa)	(m.s <sup>-1</sup> )	(mol.L⁻¹)*
Figure 4	(Gondal, et al., 2015)	LiOH	0.01	336	0.2	$1.7.10^{-4}$	3.1.10 <sup>-5</sup>
Figure 5	(Gondal, et al., 2015)	КОН	1.76	317	0.2	$1.1.10^{-4}$	2.6.10 <sup>-5</sup>
Figure 6	(Hikita, et al., 1976)	$Na_2CO_3$	0.06	298	100	5.0.10 <sup>-5</sup>	32.0.10 <sup>-3</sup>
Figure 7	(Hikita, et al., 1976)	NaOH	0.13	303	100	4.5.10 <sup>-5</sup>	28.2.10 <sup>-3</sup>

\*Calculated variable (eq. 18), in the activity-based model

318

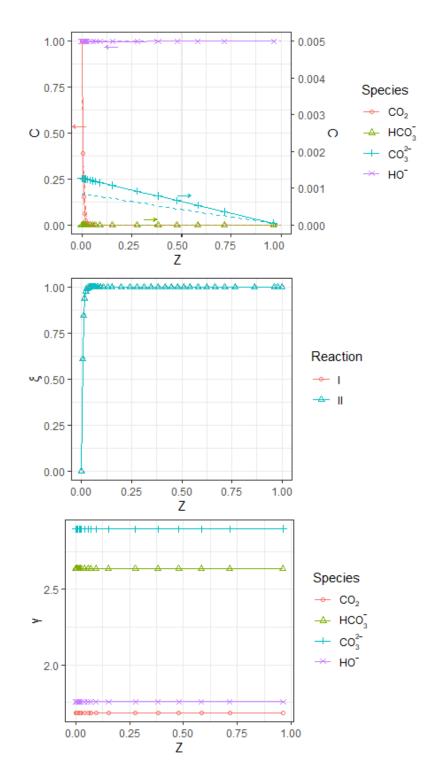
Concentration and extent of reaction profiles calculated with non-ideality consideration (solid lines and symbols in Figure 4 to Figure 7) or with ideal solution hypothesis (dashed lines in Figure 4 to Figure 7) display similar trends. CO<sub>2</sub> concentration profiles are especially close in both models. Consequently, resulting Arrhenius parameters are very close in both models as well. As expected, in these systems and at these conditions, calculated activity coefficient profiles are mostly constant but largely different from unity. Other more complex systems, such as acid gas absorption into aqueous amine solutions, could lead to higher non-ideality influence.

(Gondal, et al., 2015) experiments are compatible with pseudo-first order absorption regime 326 assumption (Figure 4 and Figure 5). Concentration profiles are compatible with this assumption: HO 327 concentration barely drops at the interface. The main reaction product is not  $HCO_3^{-}$ , but  $CO_3^{2-}$ , and 328 329 extents of both reactions are close along the liquid film. They are even the equal in Figure 5. This is 330 because  $CO_3^{2-}$  (pKb(25°C) = 3,7) is more basic than  $HCO_3^{-}$  (pKb(25°C) = 7,3). At these conditions, 331 activity coefficients are constant, as the overall solution composition remains close to bulk 332 composition everywhere. Still, activity coefficients differ significantly from unity and non ideality 333 needs to be considered. Therefore, according to the model, (Gondal, et al., 2015) absorption flux 334 measurements can be interpreted with constant activity coefficients (as done by (Gondal, et al., 335 2015) and (Gondal, et al., 2016)). The same behaviour, but amplified, is displayed in Figure 5. HO<sup>-</sup> excess is much more important, therefore, reaction R-II equilibrium shifts toward  $CO_3^{2-}$  formation. 336 Activity coefficient divergence from unity is significantly higher, due to very high ionic strength. 337



**Figure 4** Simulated absorption of CO<sub>2</sub> in LiOH aqueous solution at 0.01 mol.L<sup>-1</sup>, 336 K and P<sub>CO2</sub> = 0.2 kPa measured by (Gondal, et al., 2015). C: normalised concentration,  $\xi$ : extent of reaction,  $\gamma$ : activity coefficient. Solid line & symbol: activity-based model, dashed line: concentration-based model (C and

342 ξ). γ<sub>Li+</sub> = 0.88

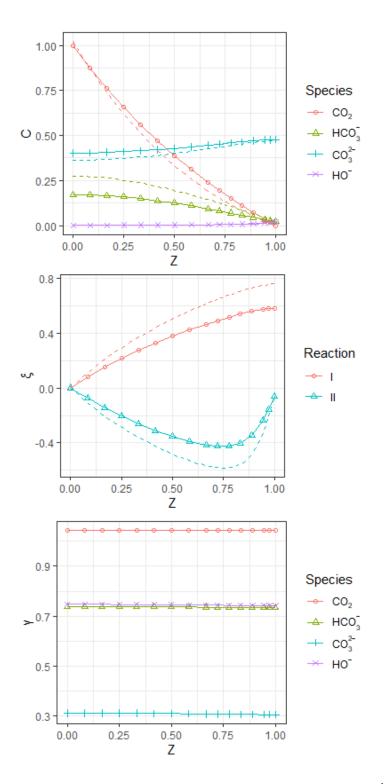


**Figure 5** Simulated absorption of CO<sub>2</sub> in KOH aqueous solution at 1.76 mol.L<sup>-1</sup>, 317 K and P<sub>CO2</sub> = 0.2 kPa measured by (Gondal, et al., 2015). C: normalised concentration,  $\xi$ : extent of reaction,  $\gamma$ : activity coefficient. Solid line & symbol: activity-based model, dashed line: concentration-based model (C and  $\xi$ ).  $\gamma_{K+}$  = 1.48

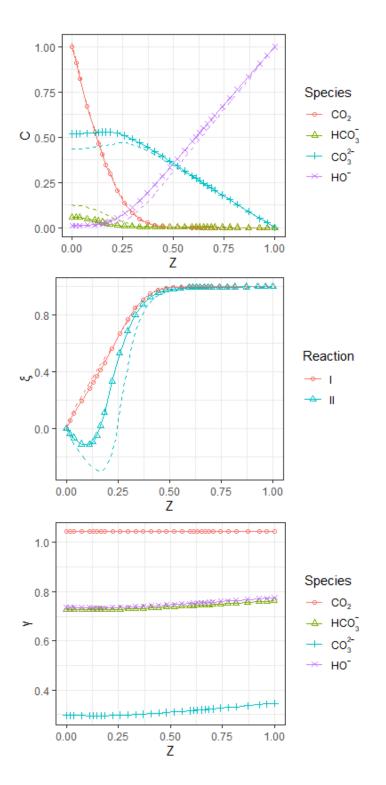
In Figure 6, resulting  $CO_2$  absorption profiles are shown in aqueous  $Na_2CO_3$  at conditions researched by (Hikita, et al., 1976). In this case, the main reaction product is  $HCO_3^{-1}$  from  $CO_3^{-2}$  conversion. This is also coherent with a negative extent of reaction R-II. Here, reversibility needs to be considered,

which the model does, unlike models from the literature (e.g., (Knuutila, et al., 2010), (Gondal, et al.,2016)).

Activity coefficients display a gradient along the liquid film in Figure 7, with a maximum of 14% 353 decrease from liquid bulk to interface for  $\gamma_{(CO32-)}$ . In (Hikita, et al., 1976), experimental CO<sub>2</sub> partial 354 pressure is of several orders of magnitude higher than in (Gondal, et al., 2015) – 100 vs. 0.2 kPa. HO<sup>-</sup> 355 356 concentration drops significantly along the liquid film, more so close to the interface, in a similar behaviour to the instantaneous absorption regime.  $CO_3^{2-}$  is consumed in the vicinity of the interface. 357 At these experimental conditions, it is important to consider the reaction mechanism with both 358 359 reactions and varying activity coefficients along the liquid film. The high absorption driving force 360 results in ionic strength variation. These conditions are closer to those expected in an industrial setting. This shows how the model could accurately represent mass transfer phenomena at industrial 361 362 conditions.



**Figure 6** Simulated absorption of  $CO_2$  in  $Na_2CO_3$  aqueous solution at 0.06 mol.L<sup>-1</sup>, 298 K and  $P_{CO2}$  = 100 kPa measured by (Hikita, et al., 1976). C: normalised concentration,  $\xi$ : extent of reaction,  $\gamma$ : activity coefficient. Solid line & symbol: activity-based model, dashed line: concentration-based model (C and  $\xi$ ).  $\gamma_{Na+}$  = 0.68



**Figure 7** Simulated absorption of CO<sub>2</sub> in NaOH aqueous solution at 0.13 mol.L<sup>-1</sup>, 303 K and P<sub>CO2</sub> = 100 kPa measured by (Hikita, et al., 1976). C: normalised concentration,  $\xi$ : extent of reaction,  $\gamma$ : activity coefficient. Solid line & symbol: activity-based model, dashed line: concentration-based model (C and  $\xi$ ).  $\gamma_{Na+} = 0.76$ 

## 373 4. <u>Conclusion</u>

This study implements a steady-state reactive absorption model applied to the stagnant film theory and focusing on solution non-ideality representation in equilibrium relations, Nernst-Planck diffusion fluxes and reaction rates. This approach is more complete than literature models where activity 377 coefficients are assumed constant, and most often equal to unity. Extents of reactions, a new feature 378 for reactive absorption models, reduces the number of differential problem variables while 379 maintaining a general formulation. This formulation with Nernst-Planck equation for molecular 380 diffusion and activity-based rate expressions consists in a system of  $n_c+n_{R,kin}$  *first*-order ordinary 381 differential equations, whereas writing local balances per species leads to a system of  $n_c$  *second*-382 order differential equations. The relation between fluxes and extents of reactions is dictated by 383 stochiometric constraints.

384 This model is first applied to alkaline salts-water-CO<sub>2</sub> systems and the CO<sub>2</sub> absorption flux data from 385 the literature. Arrhenius parameters for  $CO_2 + HO \leftrightarrow HCO_3$  direct kinetic constant are fitted. This is 386 a necessary step to study amine-based systems in the same modelling framework. This progressive 387 approach is made possible by the kinetic data availability in alkaline salts-water-CO<sub>2</sub> systems. The 388 obtained Arrhenius parameters are coherent to previous studies in the regression temperature 389 interval. Overall AAD is 12%. Besides, the consideration of activity coefficient variation along the 390 liquid film is shown to be relevant when experimental conditions diverge from the simple pseudo-391 first order absorption regime. These conditions are closer to the industrial setting.

392 Model improvement perspectives include extending mass transfer representation to steady-state393 eddy-diffusivity theory, and assessing uncoupled ion diffusivities influence.

394 With this model, reaction mechanism investigation in any reactive absorption system is possible. Our 395 further goal is to apply the model to  $CO_2$  absorption in aqueous amines solutions (namely MDEA), 396 especially for increasing  $CO_2$  partial pressure and loading. In these solutions, the more complex 397 reaction scheme could lead to higher variations in ionic strength and activity coefficients through the 398 liquid film.

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