

A universal dynamical metabolic model representing mixotrophic growth of Chlorella sp. on wastes

Bruno Assis Pessi, Caroline Baroukh, Anais Bacquet, Olivier Bernard

▶ To cite this version:

Bruno Assis Pessi, Caroline Baroukh, Anais Bacquet, Olivier Bernard. A universal dynamical metabolic model representing mixotrophic growth of Chlorella sp. on wastes. Water Research, 2023, 229, pp.119388. 10.1016/j.watres.2022.119388 . hal-03920752

HAL Id: hal-03920752 https://hal.science/hal-03920752v1

Submitted on 3 Jan 2023

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers. L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

A universal dynamical metabolic model representing mixotrophic growth of *Chlorella* sp. on wastes

Bruno Assis Pessi^a, Caroline Baroukh^b, Anais Bacquet^c, Olivier Bernard^{a,c}

^a Biocore, INRIA, Université Côte d'Azur, Sophia Antipolis, France

^b LIPME, Université de Toulouse, INRAE, CNRS, Castanet Tolosan, France

^cLOV, UMR 7093, Sorbonne university, CNRS, Villefranche-sur-mer

7 Abstract

1

2

3

4

5

6

An emerging idea is to couple wastewater treatment and biofuel production using microalgae to achieve higher productivities and lower costs. This paper proposes a metabolic modelling of *Chlorella sp.* growing on fermentation wastes (blend of acetate, butyrate and other acids) in mixotrophic conditions, accounting also for the possible inhibitory substrates. This model extends previous works by modifying the metabolic network to include the consumption of glycerol and glucose by *Chlorella sp.*, with the goal to test the addition of these substrates in order to overcome butyrate inhibition. The metabolic model was built using the DRUM framework and consists of 188 reactions and 173 metabolites. After a calibration phase, the model was successfully challenged with data from 122 experiments collected from scientific literature in autotrophic, heterotrophic and mixotrophic conditions. The optimal feeding strategy estimated with the model reduces the time to consume the volatile fatty acids from 16 days to 2 days. The high prediction capability of this model opens new routes for enhancing design and operation in waste valorisation using microalgae.

* Keywords: Chlorella, metabolic modelling, heterotrophy, mixotrophy, diauxic

⁹ growth, dynamical modelling

10 1. Introduction

Microalgae have been extensively studied during the past decade. Some species 11 are capable of producing lipids or carbohydrates that can in turn be converted into 12 biofuel (Sajjadi et al., 2018). Microalgae use light energy, via photosynthesis, to 13 fix carbon dioxide. Not only, their growth rate is much faster than that of higher 14 plants, but they can also be cultivated in wastewater, reducing the environmental 15 impact of bioproducts (Morales et al., 2019, Arashiro et al., 2022). An emerging idea 16 suggests using wastewater treatment to obtain biofuels as a co-product (Barsanti and 17 Gualtieri, 2018). Even though the production efficiency appears to be attractive, 18

Preprint submitted to Elsevier

January 3, 2023

many optimization steps still need to be carried out for this process to become
sufficiently cost-effective and environmentally-friendly (Tan et al., 2018).

There is already an extensive list of works about wastewater treatment with 21 microalgae. Nevertheless there is no unified framework for the modeling of this 22 process (Shoener et al., 2019). Most of these models are variants of active sludge 23 models (ASM) and anaerobic digestion models (ADM), which were originally de-24 signed for bacterial systems (Casagli et al., 2021b, Wágner et al., 2016). In addition 25 mixotrophy is rarely considered in these models and microalgae are assumed to grow 26 autotrophically, neglecting the simultaneous use of organic compounds. In contrast 27 to these models, metabolic models can acquire the reaction yields, including pho-28 tosynthesis and organic carbon uptake by using the knowledge of the biochemical 29 reactions taking place in the organism which is reconstructed from genomic data. 30 Despite this, less than 5% of models of water recover facilities have metabolic re-31 constructions and usually rely on empirical yield coefficients (Shoener et al., 2019). 32 Metabolic models also have the advantage of estimating internal metabolic fluxes, 33 providing valuable information for future strain improvement via metabolic engi-34 neering. Nevertheless, the applicability of metabolic models to predict dynamical 35 systems is constrained by the size of the metabolic network - which in genome scale 36 models usually consists of thousands of reactions and metabolites. Consequently, to 37 embed mixotrophic microalgal models in larger frameworks representing the mass 38 fluxes within the wastewater process, it is necessary to use techniques to reduce 39 the metabolic models to a reasonable size, while still keeping the phenotype of the 40 original network. 41

The development of techniques to reduce the size of metabolic networks started 42 more than 20 years ago (Singh and Lercher, 2020). Since then new methods have 43 been developed relying on a variety of different approaches, such as linear program-44 ming, graph-based search and elementary flux modes (Singh and Lercher, 2020). 45 Our increasing understanding of genomic information led to the construction of 46 more complex genome scale networks. This has motivated the field of bioinformat-47 ics to research new ways to reduce and analyze metabolic models, consequently new 48 methods are still being developed (Küken et al., 2021, Hameri et al., 2021). The 49 choice of the reduction method will depend on the intentions of the modeler, the 50 wished degree of flexibility of the phenotype prediction and the final size of the 51 model. In this work, we will rely on the DRUM framework (Baroukh et al., 2014, 52 2016), which although requires input from the modeler and prior knowledge about 53 the network, it is able to reproduce behaviour outside the steady state, while greatly 54 reducing the number of metabolites and reactions in the final model. On top of this, 55 such dynamic model can support a control strategy to enhance process efficiency. 56

⁵⁷ During dark fermentation of organic wastes, anaerobic bacteria and archaea con-⁵⁸ vert complex and non-assimilable compounds into Volatile Fatty Acids (VFAs) us-

able by microalgae. Indeed, VFAs can support heterotrophic growth of microalgae, 59 while they use ammonium and phosphate in the wastewater as source for nitrogen 60 and phosphorus (Baroukh et al., 2017, Turon et al., 2015b,a, Gao et al., 2022). In 61 this perspective, *Chlorella sp.* was selected for its potential in associating biofuel 62 production with effluent treatment (Casagli et al., 2021b, Gao et al., 2022, Wágner 63 et al., 2016). Indeed, this species can accumulate up to 50% of its dry weight in 64 lipids, essentially in the form of triacylglycerol (TAG). In addition, its capacity to 65 grow in the main dark fermentation effluents has been demonstrated, under het-66 erotrophic or mixotrophic conditions (Turon et al., 2015a, Lacroux et al., 2022). 67 The VFA mixture resulting from dark fermentation is typically made of a blend 68 composed of about 30% acetate and 70% butyrate. Other organic acids can also 69 be found (Rafrafi et al., 2013, Turon et al., 2015a), among which lactate, which is 70 not consumed and does not affect the growth of *Chlorella* (Turon et al., 2015a). 71 Turon et al. (2015a) first proposed a kinetic model of the consumption of butyrate 72 and acetate by *Chlorella*. Later, an extended metabolic model was proposed using 73 the DRUM framework (Baroukh et al., 2017). Here, we extend further this model 74 by including the consumption of glucose and glycerol by *Chlorella* and propose a 75 universal multi-substrate dynamic reduced metabolic model. 76

This model is the cornerstone for tackling the major issue due to the high concen-77 tration of butyrate which is slowly consumed by microalgae and inhibited by acetate 78 (Hu et al., 2012, Turon et al., 2015b, Lacroux et al., 2022). To address this problem, 79 another organic substrate can be added to lever the inhibition effect of butyrate and 80 eventually accelerate growth in dark fermentation effluents. Microalgae will first 81 use a more efficient carbon substrate to reach a higher biomass concentration. It 82 follows that acetate will be faster consumed, and in the end butyrate. Glucose could 83 ideally play this role, but its cost is not compatible with process economics. Acetate 84 addition is more reasonable but would generate large variations in pH. Glycerol is 85 a by-product of biodiesel synthesis by transesterification. Its low cost is likely to be 86 compensated by the enhanced microalgal productivity. Developing a mixotrophic 87 multi-substrate metabolic model is the main objective of this paper, and such ap-88 proach can be used to identify strategies to more efficiently use dark fermentation 89 products and optimize their conversion into algal biomass. 90

The model developed here represents, in detail, the growth under different au-91 totrophic, heterotrophic and mixotrophic conditions and for four organic substrates. 92 To our knowledge, this is the first model including such a large range of potential 93 substrates in mixotophic regimes. The model was validated in different cultivation 94 conditions using the abundant literature available on autotrophic, heterotrophic 95 or mixotrophic growth of *Chlorella*. To this end, data from 122 experiments was 96 extracted from 15 publications, amounting to more than 2600 concentration data 97 points (see Table 1), thus reaching an unprecedented level of validation. This model 98

Substrate	# exp.	Species	References	
Glycerol	12(9)	C. sorokiniana	León-Vaz et al. (2019)	
		C. sp.	Sen and Martin (2018)	
		C. protothecoides	Chen and Walker (2011), O'Grady and Morgan (2011)	
		C. vulgaris	Ma et al. (2016)	
Glucose	22 (11)	C. sorokiniana	León-Vaz et al. (2019), Li et al. (2013, 2014)	
		C. protothecoides	Espinosa-Gonzalez et al. (2014), Chen and Walker (2011)	
		C. pyrenoidosa	Ogbonna et al. (1997)	
Glucose/Glycerol	2(2)	C. protothecoides	O'Grady and Morgan (2011)	
Acetate	40 (25)	C. sorokiniana	León-Vaz et al. (2019), Turon et al. (2015a,b), Chen et al. (2017a,b), Xie et al. (2020a)	
		C. sacchrarophila	Xie et al. $(2020b)$	
Butyrate	10(4)	C. sorokiniana	Turon et al. $(2015b,a)$	
Acetate/Butyrate	23~(6)	C. sorokiniana	Turon et al. (2015b,a)	
Autotrophic	13	C. sorokiniana	Li et al. (2014) , León-Vaz et al. (2019) , Turon et al. $(2015b,a)$	
		<i>C. sp.</i>	Sen and Martin (2018)	

Table 1: Considered experiments for each substrate. In parenthesis, the number of experiments in mixotrophic conditions out of the total experiments.

is shown to support a strategy to enhance the bioconversion of VFA into microalgal
 biomass by managing the way the different substrates are supplied.

¹⁰¹ 2. Materials and methods

¹⁰² 2.1. General Principles of the DRUM approach

¹⁰³ The dynamic model development follows the DRUM (Dynamic Reduction of ¹⁰⁴ Unbalanced Metabolism) approach. The full description and complete explanation ¹⁰⁵ of the approach is available in Baroukh et al. (2014).

¹⁰⁶ Briefly, the metabolism of a microorganism can be described by its metabolic ¹⁰⁷ network composed of a set of n_r biochemical reactions (here $n_r = 188$) involving n_m metabolites (here $n_m = 173$) and represented by the stoichiometric matrix $K \in \mathcal{R}^{n_m \times n_r}$ (see Appendix for the full list of reactions and metabolites). The biomass B is produced from a set of substrates S and excretes a set of products P. In a perfectly mixed reactor with a constant volume, the system can be described by the following set of ordinary differential equations:

$$\frac{dM}{dt} = \frac{d}{dt} \begin{pmatrix} S \\ C \\ P \\ B \end{pmatrix} = \begin{pmatrix} K_S \\ K_C \\ K_P \\ K_B \end{pmatrix} .v(M).B - DM + DM_{in} = K.v(M).B - DM + DM_{in}$$
(1)

where M represents the vector of the concentrations of metabolites composed of 113 substrate (S), intracellular metabolites (C), excreted products (P) and biomass (B). 114 M_{in} is the influent concentration of these quantities. The dilution rate of the reactor 115 (ratio of influent flow rate over the reactor volume) is D (D = 0 for a batch process). 116 All the concentrations are expressed as total concentrations in the solution. $v \in \mathcal{R}^{n_r}$ 117 is the reaction kinetic vector, while the matrices $K_S \in \mathcal{R}^{n_S \times n_r}, K_C \in \mathcal{R}^{n_C \times n_r}$ 118 $K_P \in \mathcal{R}^{n_P \times n_r}$ and $K_B \in \mathcal{R}^{1 \times n_r}$ correspond, respectively, to the stoichiometric 119 matrices of substrates S, products P, intracellular metabolites C and biomass B 120 $(n_S + n_C + n_P + 1 = n_m).$ 121

In metabolic models, intracellular metabolites are generally assumed to be quasi-122 stationary $\left(\frac{dC}{dt} = K \cdot v = 0\right)$, i.e. they are assumed to be consumed as soon as they 123 have been synthesised. However, in the case of microalgae, this hypothesis has 124 proven to be false for certain of its metabolites (denoted A) during mixotrophic or 125 autotrophic growth (Baroukh et al., 2014, 2017). The DRUM method (Baroukh 126 et al., 2014), consists in dividing the metabolic network into n quasi-stationary 127 subnetworks $(K = (K_{SN_1}, \ldots, K_{SN_n}), K_{SN_i} \cdot v_{SN_i} = 0$ for $i \in 1, \ldots, n$. These are 128 linked by the A metabolites that are, in contrast, non-stationary, can accumulate 129 and be later consumed. This division into subnetworks is justified by the presence of 130 metabolic pathways that correspond to metabolic functions, to reaction groups that 131 are regulated simultaneously and to the presence of compartments within the cell. 132 Cellular mechanisms are therefore employed for assessing the subnetwork. Hence, 133 the A metabolites can either be found at the junction of several metabolic pathways, 134 or they can be transported from one compartment to another, or they can be final 135 products that accumulate in the cell. The system of ordinary differential equations 136 (1) therefore becomes: 137

$$\frac{dM'}{dt} = \frac{d}{dt} \begin{pmatrix} S\\A\\P\\B \end{pmatrix} = \begin{pmatrix} K_S\\K_A\\K_P\\K_B \end{pmatrix} . \alpha . B = K' . \alpha . B - DM' + DM'_{in}$$
(2)

With K' the stoichiometric matrix of macroscopic reactions obtained through the analysis of elementary modes (Schuster et al., 1999) on the subnetworks, and α the kinetic vector associated to these macroscopic reactions. B now represents the structural biomass, i.e. the fraction of biomass that does not contain the inert compartments of reserve A. The total biomass can be deduced using a mass balance of the elemental compounds (C, N, P, O, ...).

144 2.2. Construction of the model

The core of the metabolic network from Baroukh et al. (2017) has been used and 145 modified in order to add the glucose and glycerol consumption pathways (Figure 146 1). This network contains the central autotrophic, mixotrophic and heterotrophic 147 metabolic pathways including photosynthesis, glycolysis, the pentose phosphate 148 pathway, the Krebs cycle, oxidative phosphorylation and the synthesis of chloro-149 phyll, carbohydrates (e.g. starch), amino-acids and nucleotides. The synthesis 150 pathways of macromolecules such as proteins, lipids, starch, DNA, RNA as well 151 as the functional biomass are represented through macroscopic reactions. 152

The DRUM method requires the partitioning of the metabolism into subnet-153 works as well as the identification of the metabolites, in between the subnetworks, 154 which can accumulate. The subnetworks are defined by their metabolic function 155 and/or their affiliation to a cellular compartment. Different partitions among the 156 188 reactions have been tested, with a view to select the one which minimized the 157 number of parameters to assess while providing a correct representation of the ex-158 perimental data. The best result was obtained when the network is divided into four 159 subnetworks (Figure 1), corresponding to, 1) the glyoxysome, 2) the chloroplast, 3) 160 the absorption of glycerol and 4) the synthesis of biomass. The glyoxysome and 161 chloroplast subnetworks remain unchanged in comparison with the initial Baroukh 162 et al. (2017) model. 163

The macroscopic reactions associated to each subnetwork are deduced from the elementary mode analysis (Klamt and Stelling, 2003). The Matlab "efmtool" was run to calculate the Elementary Flux Modes (EFMs) (Terzer and Stelling, 2008). In total, 86, 142 EFMs including 3, 310 associated to futile cycles (dissipation of carbon substrate in the form of CO_2) have been achieved. These macroscopic reactions are further used to determine the mass fluxes in the different parts of the network by assembling reactions belonging to the same kinetics.

171 2.3. Analysis of the sub-networks

172 2.3.1. Motivations

In this section, we present the macroscopic reactions which result from the reduction of each subnetwork by the computation of the EFMs. As recommended by Baroukh et al. (2014), the reaction kinetics must be mathematically represented



Figure 1: Considered metabolic subnetworks to represent growth of *Chlorella* on a mixture of glycerol, glucose, acetate and butyrate

	Subnetwork	Macroscopic reaction
MR_1	Glyoxysome	$2 \text{ ACE} + 3.5 \text{ H} + 0.5 O_2 \rightarrow SUC + 0.5 H_2O$
MR_2		$BUT + 7H + 1.5O_2 \rightarrow SUC + 5H_2O$
MR_3	Chloroplast	$Light + 3CO_2 + 2H_2O + Pi \rightarrow GAP + 3O_2$
MR_4	Glycerol pathway	$GLY + Pi \rightarrow GAP + H_2O$
$MR_5(NH_4)$	Biomaga gynthogia	$4.15 \; GAP + 2.54O_2 + 0.99NH_4 + 0.02SO_4 + 0.01Mg_2 \rightarrow$
	Diomass synthesis	$B + 0.99H + 2.90H_2O + 3.92CO_2 + 4.02Pi$
MP(NO)		$4.64GAP + 2.04O_2 + 0.99NO_3 + 0.98H + 0.02SO_4 +$
$MR_5(NO_3)$		$0.01Mg_2 \to B + 5.39CO_2 + 2.90H_2O + 4.51Pi$
$MR_6(NH_4)$	Biomass synthesis	$4.15 SUC + 7.30H + 4.61O_2 + 0.99NH_4 + 0.12Pi +$
		$0.02SO_4 + 0.01Mg_2 \rightarrow B + 7.04H_2O + 8.06CO_2$
MP(NO)		$4.90SUC5.28O_2 + 0.99NO_3 + 0.12Pi + 10.78H +$
$MR_6(NO_3)$		$0.02SO_4 + 0.01Mg_2 \rightarrow B + 11.07CO_2 + 8.31H_2O$
$MR_7(NH_4)$	Biomass synthesis	$2.07 \ GLC + 2.54O_2 + 0.99NH_4 + 0.12Pi + 0.02SO_4 +$
		$0.01Mg_2 \to B + 3.91CO_2 + 7.04H_2O + 0.99H$
MP(NO)		$2.34GLC + 2.14O_2 + 0.99NO_3 + 0.12Pi + 0.98H +$
$\frac{M}{n_7}(NO_3)$		$0.02SO_4 + 0.01Mg_2 \rightarrow B + 5.49CO_2 + 7.63H_2O$

Table 2: List of the macroscopic reactions in each respective subnetwork. For biomass production, the stoichiometric values differ if the nitrogen source is nitrate or ammonium.

using minimal hypotheses, and when possible applying a mass action law. A list of all sub-networks and the macroscopic reactions can be found at Table 2.

178 2.3.2. Glyoxysome subnetwork

The glyoxysome pathway consists of 26 reactions, from which 8 are exchange 179 reactions. The glyoxysome is the peroxysome compartment where the glyoxylate 180 cycle occurs. Here carbon compounds are converted to succinate, also allowing the 181 production of glucose from lipids. In this compartment, fatty acids can be used as 182 a source of energy and carbon for growth is produced when no photosynthesis takes 183 place. Two EFMs have been achieved for this subnetwork (MR1 and MR2). In the 184 glyoxysome, butyrate and acetate are converted into acetyl-CoA, which is in turn 185 converted, via the glyoxylate cycle, into succinate. The succinate then enters the 186 cytosol and is injected into the Krebs cycle, thus producing the different metabolites 187 necessary for the synthesis of biomass. 188

¹⁸⁹ Butyrate is known for inhibiting algal growth under heterotrophic and mixotrophic ¹⁹⁰ conditions (Turon et al., 2015a). Furthermore, acetate inhibits the absorption of ¹⁹¹ butyrate, thus leading to diauxic growth (Turon et al., 2015a). Thereby, Michaelis-¹⁹² Menten kinetics have been proposed to describe the absorption of acetate (α_{MR1}).

$$\alpha_{MR_1} = \frac{k_{MR1}.ACE}{KS_{MR1} + ACE} \tag{3}$$

For butyrate (α_{MR2}) Haldane kinetics have been chosen with an inhibition by acetate term.

$$\alpha_{MR2} = \frac{k_{MR2}.BUT}{BUT + \frac{k_{MR2}}{\beta_{MR2}}.(\frac{BUT}{Sopt_{MR2}} - 1)^2} \frac{k_d}{(ACE + k_d)}$$
(4)

195 2.3.3. Chloroplast subnetwork

The chloroplast subnetwork is composed of 21 reactions, from which 7 are exchange reactions. The glycerate-3-phosphate produced by photosynthesis is assumed to be transferred from the chloroplast towards the cytosol where it can be converted by glycolysis into glucose-6-phosphate or pyruvate. These metabolites are essential for the synthesis of functional biomass.

For autotrophic growth, light drives the reaction rate. When algae are grow-201 ing on a turbid medium like wastewater, the average light intensity stays low and 202 the local photoinhibition impact can be neglected (Martínez et al., 2018). Also un-203 der mixotrophic conditions with elevated concentration of carbon substrates, large 204 biomass densities can be reached. Dense microalgal cultures strongly attenuate 205 light. In line with Baroukh et al. (2017), photosynthesis rate is assumed to be lin-206 early depending upon the average light intensity I_{μ} in the culture (see Equation 207 6).208

Moreover, light attenuation in the culture medium is expected to follow the Beer-Lambert law. The light intensity at depth z depends on the incident light I_0 and the extinction coefficient α due to the biomass (the turbidity of the medium without algae is negligible):

$$I\left(z\right) = I_0.e^{-\alpha.z.B} \tag{5}$$

The average light intensity of the culture medium in the reactor of depth L is given as follows (with $\beta_{MR3} = \alpha.L$):

$$I_{\mu} = \frac{I_0}{L} \int_0^L e^{-\alpha . B . z} dz = \frac{I_0 (1 - e^{-\beta_{MR3} . B)}}{\beta_{MR3} . B}$$
(6)

²¹⁵ The kinetics in the chloroplast subnetwork is finally given by:

$$\alpha_{MR3} = k_{MR3}.I_{\mu} \tag{7}$$

216 2.3.4. Glycerol absorption subnetwork

The glycerol pathway subnetwork consists of 5 core reactions, plus the exchange 217 reactions. Only one EFM was found for the glycerol absorption subnetwork (MR4). 218 The glycerol in the medium is transferred to the cytosol. Within three steps, 219 it is transformed into glycerate-3-phosphate. During glycolysis, this glycerate-3-220 phosphate is then used for the synthesis of precursor metabolites that are in turn 221 required for the synthesis of functional biomass. Since inhibition has been observed 222 for glycerol assimilation (Chen and Walker, 2011, Ma et al., 2016, Liang et al., 2009), 223 a Haldane reaction kinetics with inhibition was chosen (α_{MR4}) : 224

$$\alpha_{MR4} = \frac{k_{MR4}.GLY}{GLY + \frac{k_{MR4}}{\beta_{MR4}}.(\frac{GLY}{Sopt_{MR4}} - 1)^2}$$
(8)

225 2.3.5. Functional biomass synthesis subnetwork

The reactions for the synthesis of lipids, proteins, DNA, RNA, chlorophyll and carbohydrates are all lumped together in the functional biomass synthesis subnetwork. This subnetwork includes glycolysis, the Krebs cycle, oxidative phosphorylation, the pentose phosphate pathway, carbohydrate, lipid, amino-acid and nucleotide synthesis, as well as the assimilation of nitrogen, sulphur and glucose. In total there are 141 reactions in the functional biomass subnetwork.

This subnetwork generated 86, 167 EFMs, including 3310 that did not produce biomass. Nearly all of the calculated EFMs are part of the biomass synthesis network. They can be sorted by using a similar method to the FBA (Flux Balance Analysis). The standard hypothesis supporting FBA is that evolution has selected metabolisms maximising biomass growth on each substrate (Orth et al., 2010), or equivalently, minimising the loss of carbon as CO_2 . Therefore, for each substrate, the EFM presenting the highest GAP/BIOM, SUC/BIOM and GLC/BIOM yields were selected. In this way, the use of GAP, SUC or GLC for the synthesis of biomass resulted from three macroscopic reactions (MR5, MR6 and MR7). The yield of biomass on the carbon substrate depends on the nitrogen source (NH_4 or NO_3). Table 1 shows the resulting macroscopic reactions for both cases.

²⁴³ Glycerate-3-phosphate originates from the chloroplast and from the assimila-²⁴⁴ tion of glycerol. It is injected into the glycolysis so as to produce the necessary ²⁴⁵ metabolites for growth (α_{MR5}). The kinetics is supposed to be linear with respect ²⁴⁶ to glycerate-3-phosphate (*GAP*):

$$\alpha_{ME5} = k_{MR5}.GAP \tag{9}$$

Succinate originates from the glyoxysome. It enters the Krebs cycle, thus also leading to the production of metabolites required for growth (α_{MR6}). Assuming a linear kinetics we get:

$$\alpha_{MR6} = k_{MR6}.SUC \tag{10}$$

²⁵⁰ Glucose in the medium is transferred to the cytosol where it enables the produc-²⁵¹ tion of biomass. Glucose is inhibiting at high concentration (osmotic stress), and ²⁵² its consumption is assumed to follow a Haldane kinetics (α_{MR7}) (Azma et al., 2011, ²⁵³ Wu and Shi, 2007, Liang et al., 2009). MR7:

$$\alpha_{MR7} = \frac{k_{MR7}.GLC}{GLC + \frac{k_{MR7}}{\beta_{MR7}}.(\frac{GLC}{Sopt_{MR7}} - 1)^2}$$
(11)

254 2.4. Global dynamics of the network

Finally, the dynamical evolution of the metabolic fluxes associated to the 188 considered metabolic core reactions can be derived from a system with 17 ordinary differential equations comprising 17 metabolites and 7 macroscopic reactions:

$$\frac{dM^{"}}{dt} = \frac{d\begin{pmatrix} S\\A\\B \end{pmatrix}}{dt} = K^{"}.\alpha.B - DM^{"} + DM^{"}_{in}$$
(12)

²⁵⁸ Where M" is the metabolite vector (17x1) comprising the substrates S, the ²⁵⁹ metabolites that can accumulate A (SUC and GAP) and the functional biomass ²⁶⁰ B. K' is the stoichiometric matrix (17x7) of the macroscopic reactions and α the as-²⁶¹ sociated kinetics vector (7x1). It is worth noting that, even if the model results from ²⁶² a reduction process through the DRUM approach, it can still predict the evolution ²⁶³ of the 188 metabolic fluxes in the cell: Moreover, the total biomass comprising the functional biomass and the metabolites A can be described as follows:

$$X_{z}(t) = \sum_{A} Z_{A}.A(t) + Z_{B}.B(t)$$
(13)

Where Z is a chemical element $(Z \in \{C; N; O; H; P; ...\})$, Z_A and Z_B are the number of chemical elements Z per mole of metabolites A and biomass B, A(t) and B(t) are the concentrations of A and B at time t $X_Z(t)$ is the concentration of the chemical element in the total biomass X at time t.

Finally, the metabolic fluxes within the whole network can be derived from the α kinetics and the elementary modes associated to the E_{SN_i} , $i \in 1, 2, 3$ subnetworks:

$$v = \begin{pmatrix} v_{SN_1} \\ \dots \\ v_{SN_k} \end{pmatrix} = \begin{pmatrix} E_{SN_1} & \alpha_{SN_1} \\ \dots \\ E_{SN_k} & \alpha_{SN_k} \end{pmatrix}$$
(14)

272 2.5. Sensitivity analysis

We calculate the sensitivity coefficient for the model parameters as defined in Bernard et al. (2001):

$$\sigma_y^{\Delta p} = \frac{1}{t_f} \int_0^{t_f} \frac{y(p + \Delta p, x_0, u, \tau) - y(p, x_0, u, \tau)}{y(p, x_0, u, \tau)} d\tau$$
(15)

where y is the simulated output at time τ with parameter set p, initial condition 275 x_0 and input variables u (e.g. light intensity and dilution rate). We calculate 276 the global sensitivity using Morris's sampling method implemented in the SALib 277 Python toolbox (Morris, 1991, Herman and Usher, 2017), replacing the standard 278 elementary effect by the sensitivity coefficient defined above. We analyse the region 279 between $\pm 20\%$ of the calibrated values of the parameters. The analysis is conducted 280 separately for each of the four carbon substrates in mixotrophic conditions at a light 281 intensity of 500 $\mu mol/(m^2.s)$. $\sigma_y^{\Delta p}$ is calculated as the average of the simulations 282 considering a cultivation of 6 hours and three different initial concentration of the 283 considered carbon substrate (0.1, 1 and 10 g/L). Table 3 shows the result of the 284 sensibility analysis for all the parameters. 285

286 2.6. Reduced model calibration

In order to calibrate, and then validate the model, a large set of experiments from the literature have been used. In total 122 selected experiments (see Table 1) gather data on growth i) under autotrophic conditions, without any organic carbon input and submitted to light intensities ranging from 30 to 540 $\mu E.m^{-2}.s^{-1}$ ii) under heterotrophic conditions, without any light, and with varying concentrations

Parameter	Value	Unit	$\sigma_y^{\Delta p}$	SD $\sigma_y^{\Delta p}$	Ref.
k_{MR1}	$3.79 \ 10^{-1}$	$M.h^{-1}.MB^{-1}$	4.5910^{-03}	2.5210^{-03}	В
KS_{MR1}	$5.50 \ 10^{-5}$	M	9.2010^{-05}	5.2410^{-03}	B
k_{MR2}	$3.60 \ 10^{-2}$	$Mh^{-1}.MB^{-1}$	9.9610^{-05}	1.4010^{-03}	B
$Sopt_{MR2}$	$1.90 10^{-5}$	M	2.4810^{-04}	1.4110^{-03}	B
β_{MR2}	$2.58 10^5$	$h^{-1}.MB^{-1}$	1.8910^{-04}	1.4010^{-03}	B
k_{MB3}	$1.90 \ 10^{-3}$	$Mh^{-1}.MB^{-1}.(m^2.s.\mu mol^{-1})$	9.2110^{-04}	5.7010^{-04}	B
β_{MB3}	$2.48 \ 10^3$	MB^{-1}	-5.3810^{-04}	8.8210^{-05}	B
k_{MB4}	1.01	$M.h^{-1}.MB^{-1}$	5.5410^{-04}	1.8610^{-03}	A
$Sopt_{MR4}$	$7.00 \ 10^{-2}$	M	9.2710^{-04}	2.2510^{-03}	A
β_{MR4}	9.05	$h^{-1}.MB^{-1}$	1.1010^{-04}	2.4310^{-03}	A
k_{MR5}	$2.82 \ 10^1$	$h^{-1}.MB^{-1}$	9.7310^{-04}	1.1610^{-03}	B
k_D	$5.39 \ 10^{-10}$	M	-1.1010^{-04}	1.4310^{-03}	B
k_{MB6}	$2.37 10^5$	$h^{-1}.MB^{-1}$	3.0710^{-06}	3.3310^{-03}	B
k_{MR7}	$6.81 \ 10^{-2}$	$M.h^{-1}.MB^{-1}$	2.3910^{-03}	2.1610^{-03}	A
$Sopt_{MB7}$	$5.07 \ 10^{-2}$	M	9.5410^{-05}	3.2210^{-03}	A
β_{MR7}	$2.95 \ 10^{+1}$	$h^{-1}.MB^{-1}$	4.2410^{-04}	3.2410^{-03}	A

Table 3: Kinetic parameters obtained after model calibration. Results of sensibility analysis ($\sigma_y^{\Delta p}$ and standard deviation (SD) of $\sigma_y^{\Delta p}$. Parameters calibrated in this work ^A and in Baroukh et al. (2017) ^B. MB: mole of biomass

in acetate, butyrate, glucose and glycerol, pure or combined iii) under mixotrophic conditions, with light and varying concentrations in acetate, butyrate, glucose and glycerol. Depending on the studies, different combinations of these substrates were tested.

Only the parameters for glucose and glycerol consumption were calibrated. All 296 the other kinetic parameters for the macro reactions are taken from Baroukh et al. 297 (2017). The calibration was done following a two-step process. First a stochastic 298 global optimizer, Differential Evolution algorithm (Storn and Price, 1997), calculates 299 the set of parameters minimizing the relative error between model and experimental 300 data of biomass and substrate concentration over time. This parameter set is, then, 301 used as initial point in a Markov Chain Monte Carlo sampler, which returns the 302 parameters set inside a confidence interval (Foreman-Mackey et al., 2013). Glucose 303 kinetics parameters were calibrated using concentration data from 5 experiments of 304 Li et al. (2013), while glycerol parameters were fitted using data from 6 experiments 305 of Ma et al. (2016). Table 3 shows the value of calibrated parameters and Figure 2 306 shows the simulation of the calibrated model together with the experimental data 307 used for calibration. 308



Figure 2: Comparison of the model against experimental data used for the calibration of glucose and glycerol kinetic parameters. Lines represent the average of 100 simulations and the colored region represents ± 1 standard deviation of estimated concentrations using Markov Chain Monte Carlo method.

309 2.7. Optimization of butyrate consumption

We consider the optimal control problem (Harmand et al., 2019), whose objective is to minimize the time t_f , where the chemical oxygen demand of the remaining waste substrates falls below the regulation threshold. The control variable is the concentration of the substrate to be added to the dark fermentation effluent (glucose or glycerol).

$$\min_{S_0} t_f : S(t_f) \le \bar{S} \tag{16}$$

To solve the minimization problem we use a Nelder-Mead algorithm (Gao and Han, 2012). The output function simulates the metabolic model for a given S_0 returning t_f , the time required to reach the regulation threshold.

318 3. Results and discussion

319 3.1. Validation of the model

The experimental data not used during the calibration stage were used to validate the model. A coherent set of experiments, representing various experimental



Figure 3: Model simulation and experimental data with two substrates. From left to right: glycerol and glucose; butyrate and acetate. The line represents the average of 100 simulations and the colored region represents ± 1 standard deviation of estimated concentrations using Markov Chain Monte Carlo method.

conditions, was kept for this validation stage (see Table 1). We used a Markov Chain Monte Carlo method to select parameters (Foreman-Mackey et al., 2013). We considered a $\pm 20\%$ uncertainty in the initial concentration of substrates. In Figure 3, the results of the model simulations are compared with the experimental data with more than one substrate.

As illustrated in Figure 2 for single substrates, the model efficiently predicts the 327 production of biomass and the consumption of substrates, whether in autotrophic, 328 heterotrophic or in mixotrophic conditions. Furthermore, the model is still accurate 329 when there are two substrates (Figure 3). More generally, the predictive performance 330 of the model is summarized in the Taylor diagram in Figure 4 for the whole data 331 set. This diagram represents at the same time the standard deviation of the biomass 332 prediction error and the Pearson correlation coefficient (Taylor, 2001). It illustrates 333 both the centered and reduced quadratic errors between the experimental data sets 334 and the associated simulations, as well as the correlation between the model and 335 the data. It thus summarizes the degree of resemblance between the data and the 336 simulations for the vast range of considered data. Indeed, the closer a data point 337 is to (0;1), the better the model reproduces the experimental data (Taylor, 2001). 338 Figure 5 represents all the data points versus model prediction, also demonstrating 339 the goodness of fit of the model. 340

The results of the sensibility analysis is also show in Table 3. Most parameters have the same order of sensibility (10⁻⁴) showing that all have an importance in the model. The high standard deviation of the sensibility coefficient demonstrates the intertwined influence of parameters, for example maximum uptake rate and the optimal concentration of the substrate, also and the dependence of the actual substrate and biomass concentration in the dynamics of the system. This demonstrates the necessity of calibrating the model in a wide range of conditions.



Figure 4: Taylor Diagram where each point represents the Pearson correlation coefficient and a normalized standard deviation of one experiment and model simulation. The semi circles centered at standard deviation 1.0 show the root-mean-square error.



Figure 5: Validation of model predictions based on experimental data of butyrate, glycerol, glucose, acetate and biomass. All p-values for the regression are below 10^{-3} . R^2 for the lines are, respectively, 0.97, 0.71, 0.95, 0.97 and 0.64. Darker colors represent a concentration of data points.



Figure 6: Time to reach the regulation threshold of a typical dark fermentation effluent using different conditions of cultivation - Batch (typical waste effluent), Batch GLC/GLY (typical waste with addition of an optimal concentration of glucose and glycerol), Fed-batch GLC, GLY, GLC/GLY (typical waste feeding, respectively, glucose, glycerol and a mixture of both.)

3.2. Optimization of microalgae growing on a mixture of dark fermentation products 348 Acetate and butyrate are the main volatile fatty acids (VFA) products of dark 349 fermentation. Butyrate will inherently lead to growth inhibition, and a strategy 350 must be found to unblock this inhibition. Here, we propose a strategy to enhance 351 batch and fed-batch cultivation of microalgae from dark fermentation wastes, by 352 adding glucose and glycerol. We first study mixotrophic conditions, considering 353 a typical effluent from dark fermentation with 3.5g/L of butyrate and 1.7g/L of 354 acetate (Lacroux et al., 2020, Ghimire et al., 2015) and a continuous light intensity 355 of 500 $\mu mol/(m^2.s)$. The objective was that the chemical oxygen demand (COD) of 356 the effluent at the end of the effluent treatment must be below \overline{S} (here $\overline{S} = 125 mg/L$) 357 so that it can satisfy the state policies for discharge in the environment. We call this 358 objective the regulation threshold. We solve the resulting optimal control problem 359 in minimal time, with the objective that the COD of the remaining waste substrates 360 falls below the regulation threshold. The control variable is the concentration of 361 the substrate to be added to the dark fermentation effluent (glucose or glycerol). 362 In the initial situation, without addition of organic carbon substrates, 16 days are 363 necessary to reach the the regulation threshold. 364

³⁶⁵ Considering that glucose can be added, it turns out that the addition 98g/L of ³⁶⁶ glucose reduces the time to reach the COD threshold to 4.0 days (see Figures 6 and ³⁶⁷ 7). Using only glycerol takes longer: 5.6 days with an addition of 39.5g/L glycerol. ³⁶⁸ Using a mixture of both substrates, the optimal starting concentration for the batch



Figure 7: Concentrations of butyrate and acetate over time in different cultivation conditions. Optimal fed-batch with glycerol (blue), optimal fed-batch with glucose (red), batch with optimal addition of glycerol and glucose (orange), batch with only butyrate and acetate as substrates (green).

is 115g/L of glucose and 19g/L of glycerol, reducing the time to consume the VFAs to only 2.7 days.

Considering now a fed-batch cultivation systems instead of a batch one, the minimisation problem can be rewritten as following:

$$\min_{D} t_f : S(t_f) \le \bar{S} \tag{17}$$

where D is the dilution rate of the inflow containing only the additional substrates, 373 at a high concentration, so that the volume of the reactor does not change. t_f is the 374 time to reach the regulation threshold. The optimal strategy can be approximated 375 into a sub-optimal strategy, which would maximize the reaction rates for glucose 376 and glycerol consumption. The strategy thus consists in computing the dilution 377 rate such that glucose and glycerol concentrations stay constant close to the values 378 for which the consumption of glucose and glycerol is maximum. The control problem 379 is then reduced to finding the optimal final time of the inlet flux in the cases of a 380 glucose, glycerol, and mixture inlet. In the case of the mixture, a fraction of 0.21 of 381 glucose and 0.79 of glycerol is obtained, keeping glycerol at an optimal concentration. 382 Using this control strategy, the final times for fed-batch cultivation are: 3.2 days for 383 glucose, 1.4 days for glycerol and 1.2 days for a mixture of glucose and glycerol (see 384 Figure 6), to be compared to the 16 days whithout inorganic carbon addition. 385

386 3.3. Analysis of metabolic maps

One of the advantages of metabolic model is that the main metabolic fluxes can be estimated. They are represented in Figure 8 considering the exponential growth phase. During autotrophic growth, a strong activity takes place in the chloroplast subnetwork where photosynthesis occurs. The greatest fluxes are associated to the fixation of CO_2 by RUBISCO, the conversion to 3PG (3-phosphoglycerate) and the

conversion of 3PG to glycerate-3-phosphate (GAP). GAP is then mainly transported 392 toward the cytoplasm, where it is injected into the glycolysis. From then on, it 393 enables the synthesis of precursor metabolites that are necessary for the synthesis of 394 functional biomass composed of proteins, DNA, RNA, chlorophyll, carbohydrates, 395 and lipids. During pure heterotrophic growth, in the dark, no reaction occurs in 396 the chloroplast subnetwork. Considering a mix of acetate or butyrate, the largest 397 fluxes are concentrated in the glyoxysome subnetwork where carbon substrates are 398 converted into succinate. The synthesised succinate is exported from the glyoxysome 399 and injected into the Krebs cycle. The latter then enables the synthesis of precursor 400 metabolites for the production of functional biomass. To ascend the glycolysis, the 401 anaplerotic pathways are in an upward direction. 402

No reaction takes place in the glyoxysome, neither in the glycerol utilization pathway for growth on glucose only. Glucose carried from the medium into the cytoplasm is directly injected into the upper glycolysis for the production of precursor metabolites necessary for the synthesis of functional biomass. Glycolysis is therefore entirely in a downward direction, as the anaplerotic pathways that enable the synthesis of oxaloacetate (OXA) fueling the Krebs cycle.

For growth on glycerol only, the greatest fluxes are located in the glycerol utilization pathway subnetwork, where the uptaken glycerol is converted to glycerate-3phosphate. The glycerate-3-phosphate is then injected into the middle of glycolysis, this time in an upward direction for the high glycolysis and in a downward direction for the lower glycolysis. The remnants of the fluxes are similar to those observed for glucose.

In mixotrophic conditions, when all carbon substrates are present in the medium, all metabolic reactions are activated. First, diauxic growth occurs, and acetate is consumed instead of butyrate. Glycerol and glucose are also used, but the flux remains after the acetate is depleted. Secondly, once all acetate has disappeared, butyrate is in turn consumed.

Figure 8 shows the metabolic fluxes after 30 hours of batch culture with opti-420 mal initial conditions (110 g/l of glucose and 20 g/L of glycerol). At this point, 421 the acetate in the culture is practically depleted, but the concentration of butyrate 422 remains similar to the initial condition. The rate of consumption of butyrate then 423 remains stable for 2 days, and the rate rapidly increases until the concentration 424 threshold is reached. Since the optimal concentration is about 40 times lower than 425 the regulation threshold, the butyrate consumption rate does not reach the max-426 imum value. Glycerol has a much lower concentration, while the flux of glucose 427 concentration is still high. As a result of the higher concentration of biomass and 428 the lower availability of light per cell, the metabolic fluxes in the chloroplast are 429 reduced. 430



Figure 8: Metabolic charts showing the fluxes of an optimized batch. Above, A, at initial concentration of acetate and butyrate with optimal conditions of glucose and glycerol. Below, B, 30 hours after the beginning of the batch. The widths are linearly proportional to the calculated flux.

431 3.4. Model limitations and perspectives

The reduced metabolic model efficiently represents the microalgal growth un-432 der various substrates in heterotrophic or mixotrophic conditions. More accurate 433 predictions could probably be obtained by expanding the model to include other 434 factors such as pH, CO_2 , O_2 and temperature which were differing among the large 435 experimental data set considered in our study. According to Lacroux et al. (2020), 436 pH fluctuations when algae consume VFA can strongly impact growth and should 437 now be included in the model. Associating a pH model that accounts for the various 438 chemical species and their speciation, as proposed by Casagli et al. (2021a), would 439 allow the calculation of the pH and the concentration of the undissociated form of 440 the acids, which is the one actually taken up by the microalgae. 441

The effect of mixotrophic growth, here considered as the sum of autotrophic and 442 heterotrophic conditions, can be more subtle in some cases. There is no consen-443 sus in the literature for a general model that fits all cases. For example, according 444 to Martínez and Orús (1991), mixotrophic growth is greater than the sum of au-445 totrophic and heterotrophic conditions when the concentration of CO_2 is limiting, 446 since CO_2 produced by respiration can be recycled for the photosynthesis pathway. 447 Recently, circular use of CO_2 and O_2 and preferential consumption of inorganic car-448 bon has also been evidenced for *Chlorella vulgaris*, nonetheless heterotrophic growth 449 was promoted at high 0_2 concentrations (Manhaeghe et al., 2020). Furthermore, 450 under certain conditions an inverse relationship between light intensity and glucose 451 consumption has been recorded (Patel et al., 2019, Wan et al., 2011). 452

Temperature is one of the most important factors affecting microalgae growth, 453 even if only a minority of the models account for it (Shoener et al., 2019). The 454 experimental data considered here were carried out at various temperatures and 455 accounting for it would increase model accuracy. In the case of metabolic models, 456 including temperature is challenging, and will involve a large number of parameters 457 to characterize the influence of temperature on each individual reaction. The influ-458 ence of short variations of temperature could be well estimated by the Arrhenius 459 equation, but large variations of temperature typically occurring in outdoor cultiva-460 tion require more advanced models (Casagli and Bernard, 2022, Pessi et al., 2022). 461 For these reasons, calibrating the temperature effect would require a large amount 462 of dedicated experiments. 463

It is likely that other factors were influencing the experimental outcomes in the considered data base, especially for the experiments carried out with high concentrations of glucose and glycerol. It seems that another substrate was sometimes limiting growth, most probably nitrogen according to the mass balance from the medium initial composition.

Extending the metabolic model to account for all these mechanisms is beyond the scope of this paper. It will require a large number of experiments to further

calibrate and validate the model. Our objective was primarily to validate the model 471 across a wide range of conditions, demonstrating a strong foundation for future 472 improvements in the model, knowing that most models of water resource recovery 473 facilities are calibrated, but not validated (Shoener et al., 2019). Overall, despite 474 the simplicity of the model in its present form, it is already very efficient. Especially 475 when accounting for the diversity of strain and experimental conditions considered 476 through the 15 studied papers. The model can then, already in its present form, 477 can be used as a tool for optimizing microalgae growth on a mix of substrates. 478

Better predictions will be achieved if biochemical and also cellular level processes 479 (e.g. metabolic reactions) are considered in the next generation of water treatment 480 models (Batstone et al., 2019). Metabolic models are able to accurately predict 481 the specific VFAs produced by a mixed-culture depending on the components of 482 the input medium (Regueira et al., 2020). We could envisage the coupling of these 483 models - VFA production by mixed-culture and treatment by microalgae - to opti-484 mize production of hydrogen constrained by the capacities of the waste treatment 485 system, or to predict in advance the necessity of adding another carbon substrate 486 such as glycerol. Metabolic knowledge could also give a strong foundation for esti-487 mating possible medium conditions take could lead to undesirable emissions, such 488 as N_2O , and which are not well modeled through the current approaches (Casagli 489 et al., 2021a). 490

491 4. Conclusion

The metabolic model developed in this work for *Chlorella* accurately predicts 492 growth under autotrophic, heterotrophic and mixotrophic conditions with acetate, 493 butyrate, glucose and glycerol for more than a hundred experiments from the liter-494 ature. Covering a large range of conditions, strains and substrates, the predictive 495 capacity of this reduced metabolic model remains remarkable. Moreover, to lift the 496 inhibition exerted by the presence of butyrate, the optimal addition of different sub-497 strates in the medium has been predicted by the model. Optimising the cultivation 498 conditions reduce the time to reach the regulation threshold from 16 to less than 2 499 days. 500

Thanks to this approach, it will become possible to streamline two-stage water treatment strategies, and to recycle, carbon, nitrogen and phosphorus into the microalgal biomass. In particular, this approach highlights noteworthy synergies between waste molecules, and the important role of of the metabolic network for future models. The anticipated management of these molecules could improve productivity significantly.

507 Acknowledgment

The authors acknowledge the support of the ControlAB ANR project (ANR-20-CE45-0014) and of the BIOMSA Ademe project. The UCAJEDI and EUR DS4H investments in the Future projects managed by the National Research Agency (ANR) with the reference numbers ANR-15-IDEX-0001.

512 **References**

Arashiro, L.T., Josa, I., Ferrer, I., Van Hulle, S.W.H., Rousseau, D.P.L., Garfí,
 M., 2022. Life cycle assessment of microalgae systems for wastewater treatment
 and bioproducts recovery: Natural pigments, biofertilizer and biogas. Sci. Total
 Environ. 847, 157615. doi:10.1016/j.scitotenv.2022.157615.

Azma, M., Mohamed, M.S., Mohamad, R., Rahim, R.A., Ariff, A.B., 2011. Improvement of medium composition for heterotrophic cultivation of green microalgae, Tetraselmis suecica, using response surface methodology. Biochem. Eng. J. 53, 187–195. doi:10.1016/j.bej.2010.10.010.

Baroukh, C., Muñoz-Tamayo, R., Bernard, O., Steyer, J.P., 2016. Reply
to the Comment on "Mathematical modeling of unicellular microalgae and
cyanobacteria metabolism for biofuel production" by Baroukh et al. [Curr.
Opin. Biotechnol. 2015, 33:198–205]. Curr. Opin. Biotechnol. 38, 200–202.
doi:10.1016/j.copbio.2016.02.018.

Baroukh, C., Muñoz-Tamayo, R., Steyer, J.P., Bernard, O., 2014. DRUM: A
new framework for metabolic modeling under non-balanced growth. Application to the carbon metabolism of unicellular microalgae. PLOS ONE 9, 1–15.
doi:10.1371/journal.pone.0104499.

Baroukh, C., Turon, V., Bernard, O., 2017. Dynamic metabolic modeling of heterotrophic and mixotrophic microalgal growth on fermentative wastes. PLoS Comput. Biol. 13, 1–18. doi:10.1371/journal.pcbi.1005590.

Barsanti, L., Gualtieri, P., 2018. Is exploitation of microalgae economically and
 energetically sustainable? Algal Res. 31, 107–115. doi:10.1016/j.algal.2018.02.001.

Batstone, D., Hülsen, T., Oehmen, A., 2019. Metabolic modelling of mixed
culture anaerobic microbial processes. Curr. Opin. Biotechnol. 57, 137–144.
doi:10.1016/j.copbio.2019.03.014.

Bernard, O., Hadj-Sadok, Z., Dochain, D., Genovesi, A., Steyer, J.P., 2001. Dynamical model development and parameter identification for an anaerobic wastewater
treatment process. Biotechnol. Bioeng. 75, 424–438. doi:10.1002/bit.10036.

Casagli, F., Bernard, O., 2022. Simulating biotechnological processes affected by
meteorology: Application to algae–bacteria systems. J. Cleaner Prod. 377, 134190.
doi:10.1016/j.jclepro.2022.134190.

Casagli, F., Rossi, S., Steyer, J.P., Bernard, O., Ficara, E., 2021a. Balancing Microalgae and Nitrifiers for Wastewater Treatment: Can Inorganic Carbon Limitation Cause an Environmental Threat? Environ Sci Technol 55, 3940–3955.
doi:10.1021/acs.est.0c05264.

Casagli, F., Zuccaro, G., Bernard, O., Steyer, J.P., Ficara, E., 2021b. ALBA: A
comprehensive growth model to optimize algae-bacteria wastewater treatment in
raceway ponds. Water Res. 190, 116734. doi:10.1016/J.WATRES.2020.116734.

⁵⁵¹ Chen, C.Y., Ho, S.H., Liu, C.C., Chang, J.S., 2017a. Enhancing lutein produc⁵⁵² tion with Chlorella sorokiniana Mb-1 by optimizing acetate and nitrate con⁵⁵³ centrations under mixotrophic growth. J. Taiwan Inst. Chem. Eng. 79, 88–96.
⁵⁵⁴ doi:10.1016/j.jtice.2017.04.020.

⁵⁵⁵ Chen, J.H., Chen, C.Y., Chang, J.S., 2017b. Lutein production with wild-type
⁵⁵⁶ and mutant strains of Chlorella sorokiniana MB-1 under mixotrophic growth. J.
⁵⁵⁷ Taiwan Inst. Chem. Eng. 79, 66–73. doi:10.1016/j.jtice.2017.04.022.

⁵⁵⁸ Chen, Y.H., Walker, T.H., 2011. Biomass and lipid production of heterotrophic
⁵⁵⁹ microalgae Chlorella protothecoides by using biodiesel-derived crude glycerol.
⁵⁶⁰ Biotechnol. Lett. 2011 3310 33, 1973–1983. doi:10.1007/S10529-011-0672-Y.

Espinosa-Gonzalez, I., Parashar, A., Bressler, D.C., 2014. Heterotrophic growth
and lipid accumulation of Chlorella protothecoides in whey permeate, a dairy
by-product stream, for biofuel production. Bioresour. Technol. 155, 170–176.
doi:10.1016/J.BIORTECH.2013.12.028.

Foreman-Mackey, D., Hogg, D.W., Lang, D., Goodman, J., 2013. Emcee : The
 MCMC Hammer. Publ. Astron. Soc. Pac. 125, 306–312. doi:10.1086/670067.

Gao, F., Han, L., 2012. Implementing the Nelder-Mead simplex algorithm with adaptive parameters. Comput Optim Appl 51, 259–277. doi:10.1007/s10589-010-93293.

Gao, Y., Guo, L., Jin, C., Zhao, Y., Gao, M., She, Z., Wang, G., 2022. Metagenomics
and network analysis elucidating the coordination between fermentative bacteria
and microalgae in a novel bacterial-algal coupling reactor (BACR) for mariculture
wastewater treatment. Water Res. 215, 118256. doi:10.1016/j.watres.2022.118256.

Ghimire, A., Frunzo, L., Pirozzi, F., Trably, E., Escudie, R., Lens, P.N., Esposito,
G., 2015. A review on dark fermentative biohydrogen production from organic
biomass: Process parameters and use of by-products. Appl. Energy 144, 73–95.
doi:10.1016/j.apenergy.2015.01.045.

Hameri, T., Fengos, G., Hatzimanikatis, V., 2021. The effects of model complexity
and size on metabolic flux distribution and control: Case study in Escherichia
coli. BMC Bioinf. 22, 1–25. doi:10.1186/s12859-021-04066-y.

Harmand, J., Lobry, C., Rapaport, A., Sari, T., 2019. Optimal Control in Bio processes: Pontryagin's Maximum Principle in Practice. volume 3 of *Chemical Engineering Series - Chemostat and Bioprocesses*. Wiley.

Herman, J., Usher, W., 2017. SALib: An open-source Python library for Sensitivity
 Analysis. J Open Source Softw 2. doi:10.21105/joss.00097.

Hu, B., Min, M., Zhou, W., Du, Z., Mohr, M., Chen, P., Zhu, J., Cheng, Y., Liu, Y.,
Ruan, R., 2012. Enhanced mixotrophic growth of microalga Chlorella sp. on pretreated swine manure for simultaneous biofuel feedstock production and nutrient
Bioresour. Technol. 126, 71–79. doi:10.1016/j.biortech.2012.09.031.

Klamt, S., Stelling, J., 2003. Two approaches for metabolic pathway analysis?
Trends Biotechnol. 21, 64–69. doi:10.1016/S0167-7799(02)00034-3.

Küken, A., Wendering, P., Langary, D., Nikoloski, Z., 2021. A structural property
for reduction of biochemical networks. Sci Rep 11, 17415. doi:10.1038/s41598021-96835-1.

Lacroux, J., Jouannais, P., Atteia, A., Bonnafous, A., Trably, E., Steyer, J.P., van
Lis, R., 2022. Microalgae screening for heterotrophic and mixotrophic growth on
butyrate. Algal Res 67, 102843. doi:10.1016/j.algal.2022.102843.

Lacroux, J., Trably, E., Bernet, N., Steyer, J.P., van Lis, R., 2020. Mixotrophic
growth of microalgae on volatile fatty acids is determined by their undissociated
form. Algal Res. 47, 101870. doi:10.1016/j.algal.2020.101870.

León-Vaz, A., Léon, R., Díaz-Santos, E., Vigara, J., Raposo, S., 2019.
Using agro-industrial wastes for mixotrophic growth and lipids production
by the green microalga Chlorella sorokiniana. N. Biotechnol. 51, 31–38.
doi:10.1016/j.nbt.2019.02.001.

Li, T., Zheng, Y., Yu, L., Chen, S., 2013. High productivity cultivation of a heatresistant microalga Chlorella sorokiniana for biofuel production. Bioresour. Technol. 131, 60–67. doi:10.1016/j.biortech.2012.11.121. Li, T., Zheng, Y., Yu, L., Chen, S., 2014. Mixotrophic cultivation of a Chlorella
sorokiniana strain for enhanced biomass and lipid production. Biomass Bioenergy
66, 204–213. doi:10.1016/j.biombioe.2014.04.010.

Liang, Y., Sarkany, N., Cui, Y., 2009. Biomass and lipid productivities of Chlorella
vulgaris under autotrophic, heterotrophic and mixotrophic growth conditions.
Biotechnol. Lett. 2009 317 31, 1043–1049. doi:10.1007/S10529-009-9975-7.

Ma, X., Zheng, H., Addy, M., Anderson, E., Liu, Y., Chen, P., Ruan, R., 2016.
Cultivation of Chlorella vulgaris in wastewater with waste glycerol: Strategies for
improving nutrients removal and enhancing lipid production. Bioresour. Technol.
207, 252–261. doi:10.1016/j.biortech.2016.02.013.

Manhaeghe, D., Blomme, T., Van Hulle, S.W., Rousseau, D.P., 2020. Experimental
assessment and mathematical modelling of the growth of Chlorella vulgaris under
photoautotrophic, heterotrophic and mixotrophic conditions. Water Res. 184,
116152. doi:10.1016/J.WATRES.2020.116152.

Martínez, C., Mairet, F., Bernard, O., 2018. Theory of turbid microalgae cultures.
 J. Theor. Biol. 456, 190–200. doi:10.1016/j.jtbi.2018.07.016.

Martínez, F., Orús, M.I., 1991. Interactions between Glucose and Inorganic Carbon
 Metabolism in Chlorella vulgaris Strain UAM 101. Plant Physiol. 95, 1150–1155.
 doi:10.1104/PP.95.4.1150.

Morales, M., Collet, P., Lardon, L., Hélias, A., Steyer, J.P., Bernard, O., 2019. Chapter 20 - Life-cycle assessment of microalgal-based biofuel, in: Pandey, A., Chang,
J.S., Soccol, C.R., Lee, D.J., Chisti, Y. (Eds.), Biofuels from Algae (Second Edition). Elsevier. Biomass, Biofuels, Biochemicals, pp. 507–550. doi:10.1016/B978-0-444-64192-2.00020-2.

Morris, M.D., 1991. Factorial Sampling Plans for Preliminary Computational Experiments. Technometrics 33, 161–174. doi:10.2307/1269043.

Ogbonna, J.C., Masui, H., Tanaka, H., 1997. Sequential heterotrophic/autotrophic
cultivation—an efficient method of producing Chlorella biomass for health food
and animal feed. J. Appl. Phycol. 9, 359–366. doi:10.1023/A:1007981930676.

O'Grady, J., Morgan, J.A., 2011. Heterotrophic growth and lipid production
of Chlorella protothecoides on glycerol. Bioprocess Biosyst. Eng. 34, 121–125.
doi:10.1007/s00449-010-0474-y.

Orth, J.D., Thiele, I., Palsson, B.Ø., 2010. What is flux balance analysis? Nat.
Biotechnol. 28, 245–248. doi:10.1038/nbt.1614.

Patel, A.K., Joun, J.M., Hong, M.E., Sim, S.J., 2019. Effect of light conditions on
mixotrophic cultivation of green microalgae. Bioresour. Technol. 282, 245–253.
doi:10.1016/J.BIORTECH.2019.03.024.

Pessi, B.A., Pruvost, E., Talec, A., Sciandra, A., Bernard, O., 2022. Does temperature shift justify microalgae production under greenhouse? Algal Res. 61, 102579.
doi:10.1016/j.algal.2021.102579.

Rafrafi, Y., Trably, E., Hamelin, J., Latrille, E., Meynial-Salles, I., Benomar, S.,
Giudici-Orticoni, M.T., Steyer, J.P., 2013. Sub-dominant bacteria as keystone
species in microbial communities producing bio-hydrogen. Int. J. Hydrogen Energy 38, 4975–4985. doi:10.1016/j.ijhydene.2013.02.008.

Regueira, A., Bevilacqua, R., Lema, J.M., Carballa, M., Mauricio-Iglesias, M.,
2020. A metabolic model for targeted volatile fatty acids production by cofermentation of carbohydrates and proteins. Bioresour. Technol. 298, 122535.
doi:10.1016/j.biortech.2019.122535.

Sajjadi, B., Chen, W.Y., Raman, A.A.A., Ibrahim, S., 2018. Microalgae lipid and
biomass for biofuel production: A comprehensive review on lipid enhancement
strategies and their effects on fatty acid composition. Renewable Sustainable
Energy Rev. 97, 200–232. doi:10.1016/j.rser.2018.07.050.

Schuster, S., Dandekar, T., Fell, D.A., 1999. Detection of elementary flux modes
in biochemical networks: A promising tool for pathway analysis and metabolic
engineering. Trends Biotechnol. 17, 53–60. doi:10.1016/S0167-7799(98)01290-6.

Sen, R., Martin, G.J.O., 2018. Glycerol and nitrate utilisation by marine
microalgae Nannochloropsis salina and Chlorella sp. and associated bacteria during mixotrophic and heterotrophic growth. Algal Res. 33, 298–309.
doi:10.1016/j.algal.2018.06.002.

Shoener, B.D., Schramm, S.M., Béline, F., Bernard, O., Martínez, C., Plósz, B.G.,
Snowling, S., Steyer, J.P., Valverde-Pérez, B., Wágner, D., Guest, J.S., 2019.
Microalgae and cyanobacteria modeling in water resource recovery facilities: A
critical review. Water Res. X 2, 100024. doi:10.1016/j.wroa.2018.100024.

Singh, D., Lercher, M.J., 2020. Network reduction methods for genome-scale
 metabolic models. Cell. Mol. Life Sci. 77, 481–488. doi:10.1007/s00018-019-03383 Z.

Storn, R., Price, K., 1997. Differential evolution-a simple and efficient heuristic
for global optimization over continuous spaces. J. Global Optim. 11, 341–359.
doi:10.1023/A:1008202821328.

Tan, X.B., Lam, M.K., Uemura, Y., Lim, J.W., Wong, C.Y., Lee, K.T.,
2018. Cultivation of microalgae for biodiesel production: A review on upstream and downstream processing. Chinese J. Chem. Eng. 26, 17–30.
doi:10.1016/j.cjche.2017.08.010.

Taylor, K.E., 2001. Summarizing multiple aspects of model performance in a single
 diagram. J. Geophys. Res. Atmos. 106, 7183–7192. doi:10.1029/2000JD900719.

Terzer, M., Stelling, J., 2008. Large-scale computation of elementary flux modes with bit pattern trees. Bioinformatics 24, 2229–2235.
doi:10.1093/bioinformatics/btn401.

Turon, V., Baroukh, C., Trably, E., Latrille, E., Fouilland, E., Steyer, J.P., 2015a.
Use of fermentative metabolites for heterotrophic microalgae growth: Yields and kinetics. Bioresour. Technol. 175, 342–349. doi:10.1016/j.biortech.2014.10.114.

Turon, V., Trably, E., Fouilland, E., Steyer, J.P., 2015b. Growth of Chlorella
sorokiniana on a mixture of volatile fatty acids: The effects of light and temperature. Bioresour. Technol. 198, 852–860. doi:10.1016/j.biortech.2015.10.001.

Wágner, D.S., Valverde-Pérez, B., Sæbø, M., Bregua de la Sotilla, M., Van Wagenen, J., Smets, B.F., Plósz, B.G., 2016. Towards a consensus-based biokinetic model for green microalgae – The ASM-A. Water Res. 103, 485–499. doi:10.1016/j.watres.2016.07.026.

Wan, M., Liu, P., Xia, J., Rosenberg, J.N., Oyler, G.A., Betenbaugh, M.J., Nie, Z.,
Qiu, G., 2011. The effect of mixotrophy on microalgal growth, lipid content, and
expression levels of three pathway genes in Chlorella sorokiniana. Appl. Microbiol.
Biotechnol. 91, 835–844. doi:10.1007/s00253-011-3399-8.

Wu, Z., Shi, X., 2007. Optimization for high-density cultivation of heterotrophic
Chlorella based on a hybrid neural network model. Lett. Appl. Microbiol. 44,
13–18. doi:10.1111/j.1472-765X.2006.02038.x.

Xie, Y., Li, J., Ho, S.H., Ma, R., Shi, X., Liu, L., Chen, J., 2020a. Pilot-scale
cultivation of Chlorella sorokiniana FZU60 with a mixotrophy/photoautotrophy
two-stage strategy for efficient lutein production. Bioresour. Technol. 314, 123767.
doi:10.1016/j.biortech.2020.123767.

Xie, W., Liu, J., Luo, J., 2020b. Mixotrophic cultivation Ζ., Lin, 707 of Chlorella for biomass production by using pH-stat culture medium: 708 Glucose-Acetate-Phosphorus (GAP). Bioresour. Technol. 313, 123506. 709 doi:10.1016/j.biortech.2020.123506. 710