

Combatting global grassland degradation

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1 **Combatting global grassland degradation**

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Abstract 26

27 Grasslands are under severe threat from on-going degradation, undermining their capacity to support biodiversity, ecosystem services, and human wellbeing. Yet grasslands are largely 28 29 ignored in sustainable development agendas. In this Perspective, we examine the current state 30 of global grasslands and explore the extent and dominant drivers of their degradation. Socio-31 ecological solutions are needed to combat degradation and promote restoration. Important 32 strategies include: increasing recognition of grasslands in global policy; developing standardised indicators of degradation; using scientific innovation for effective restoration at 33 34 regional and landscape scales; and enhancing knowledge transfer and data sharing on restoration experiences. Stakeholder needs can be balanced through standardised assessment 35 36 and shared understanding of the potential ecosystem service trade-offs in degraded and restored 37 grasslands. The integration of these actions into sustainability policy will aid in halting

degradation and enhancing restoration success, and protect the socio-economic, cultural and
 ecological benefits that grasslands provide.

40

41 **Table of Contents Summary**

Grasslands provide key ecosystem services, but their protection is often ignored in sustainable
policy. This Perspective describes grassland degradation and sets out the steps needed to
protect these systems and promote their restoration.

45

46 [H1] Introduction

47 Grasslands, comprising open grassland, grassy shrublands and savannas, cover about 48 40% of the Earth's surface and 69% of Earth's agricultural land area¹⁻³. Grasslands serve as an 49 important global reservoir of biodiversity, including many iconic and endemic species, but also 50 provide a wide range of material and non-material benefits to humans. These benefits include 51 a wide range of ecosystem services, such as food production, water supply and regulation, carbon storage and climate mitigation, pollination, and a host of cultural services¹⁻³. Despite 52 53 the importance of grassland, its degradation is widespread and accelerating in many parts of the world⁴⁻⁶, with as much as 49% of grassland area globally having been degraded to some 54 extent^{5,7,8}. 55

56 Grassland degradation poses an enormous threat to the hundreds of millions of people 57 who rely on grasslands for food, fuel, fibre and medicinal products, in addition to cultural values^{9,10}. The global cost of grassland degradation on livestock production has been estimated 58 at \$6.8 billion over the period 2001-2011¹¹, with the resulting impact on human welfare being 59 60 particularly severe in regions where most the population is below the poverty line. Grassland 61 degradation also creates major environmental problems, as grasslands play a critical role in 62 biodiversity conservation, climate and water regulation, and global biogeochemical cycles^{2,4}. For example, the conversion of tropical grassy biomes to arable cropland poses a threat to 63 biodiversity, given grasslands have vertebrate species richness comparable to forests¹². The 64 65 expansion of croplands in United States has caused widespread conversion of prairie grasslands, with considerable cost to wildlife⁶. Moreover, the conversion of grasslands to 66 67 arable cropland and disturbance through overgrazing, fire and invasive species can lead to substantial soil carbon loss¹³. 68

Because of these problems, grassland degradation represents a major global challenge
that must be addressed to achieve the key targets of biodiversity agendas, such as the Aichi
Biodiversity Targets of the Convention on Biological Diversity (CBD) and the United Nations

Sustainable Development Goals (SDGs), including restoration and sustainable use of terrestrial ecosystems, hunger and poverty alleviation, and climate change mitigation. Combating degradation is also central to the UN Decade on Ecosystem Restoration (2021-2030), which puts a firm focus on the urgent need for restoration strategies for degraded ecosystems, including grasslands.

77 There are multiple, co-occurring drivers of grassland degradation, including over-78 grazing, eutrophication, land conversion to forestry and crops, land abandonment, invasive 79 species, climate extremes and altered fire regimes (Fig. 1). These drivers are also often closely 80 linked to socio-economic drivers, such as spatial expansion of humans, the economics of land 81 use, changes in affluence and dietary preferences, and in land tenure and a lack of enforcement 82 of land use rights⁹. Climate change is further exacerbating grassland degradation, especially through more frequent and intense fires and droughts ⁸. Despite the importance of grasslands, 83 84 there has been little progress in finding solutions to halting and reversing global grassland 85 degradation, which is compromising sustainable development and the ecosystem services 86 grasslands provide.

87 In this Perspective, we consider the current state of and ecosystems service provided by 88 grasslands (open grassland, grassy shrublands, and savanna). The known extent and dominant 89 causes of grassland degradation are reviewed. We then consider the challenges that grassland 90 restoration efforts face, from societal attitudes to practical efforts, and identify actions that are 91 critical to the development of socio-ecological solutions to combat degradation and promote 92 restoration of global grasslands. We argue that progress can only be made through increasing 93 recognition of grasslands in regional and global policy, developing standardised indicators of 94 degradation and restoration, introducing ecological innovation into restoration, and enhancing 95 knowledge transfer and data sharing regarding restoration experiences.

96

97 [H1] Defining grasslands

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A key issue in the poor recognition of grassland in policy is a lack of clarity in defining the grassland biome. Indeed, there are many definitions of "grassland" and the boundaries between grasslands and other ecosystems are often poorly delineated, which contributes to variable estimates of their global extent³. For example, some definitions of grassland have a relatively high tree cover of 30% before they become classed as forest¹⁴ and in many regions there is no clear delineation of grasslands, forests and their distinct transition zones, which in Eurasia are called forest-steppes¹⁵. This Perspective uses the common definition of grasslands
as areas dominated by grasses or other graminoids, with a low abundance of trees or shrubs^{3,16}.

107 The definition here includes both temperate and tropical grasslands and considers broad 108 grassland types based on a gradient anthropogenic influence, ranging from 'natural grasslands' 109 that are climatically determined with minimal human interference, to 'semi-natural grasslands' that have been shaped by centuries of human activity, and 'intensively managed grasslands' 110 111 that have been substantially modified for agriculture³. As agriculture has become increasingly 112 intensive over the last century, 'intensively managed grasslands' of low plant species diversity 113 now dominate in some parts of the world, characterised by the use of inorganic fertilisers, high 114 stocking densities of productive livestock breeds, and often frequent re-sowing with highvielding plant varieties. While intensively managed grasslands are valued by some stakeholder 115 116 groups, here we consider them potentially degraded given that they are fundamentally different 117 from natural and semi-natural grasslands in terms of their biodiversity and the ecosystem services they provide¹⁷. 118

119 Natural grasslands are often considered the most important grassland type. They are 120 found in both the tropical (namely as savannas, which are extensive in Africa and South 121 America) and temperate (North American prairies, Eurasian steppes and South American 122 pampas) regions of the world. Their extent is largely climatically determined, occurring where 123 the climate and/or soils cannot support woody plant growth, such as above the tree-line or in 124 drylands. However, climatic factors often interact with natural disturbances, such as by fire or 125 wild herbivores, which constrain growth of woody plants and maintain the grasses and associated low-growing plants¹⁸. Indeed, changes in the balance of these interacting factors can 126 trigger radical change. For example, the encroachment of native woody plants into natural 127 128 grasslands is occurring in many parts of Africa (and elsewhere), which is linked to a decline in burned area and herbivore densities along with warmer and wetter climates¹⁹. In contrast, the 129 130 Holocene transition from extensive grasslands into the Sahara Desert was likely due to the onset of drier conditions at the end of the African Humid Period²⁰. 131

Semi-natural grasslands, also referred to as secondary grasslands¹⁶, have been created by clearing natural vegetation such as forest and by modifying natural grasslands. They are typically 'early successional' systems and would not exist without some human influence³. This type of grassland is of particularly high concern for conservation in Europe, due to their importance as reservoirs for biodiversity²¹ and their historical role in the landscape²². Indeed, semi-natural grasslands have been produced by centuries of human activity, generally by traditional, low-intensity agriculture involving livestock grazing, hay-cutting, and sometimes management with fire, and are often valued for their aesthetics and are related to a sense of
 identity, place and heritage²².

141 The vegetation of both natural and semi-natural grasslands, while typically dominated by grasses and graminoids, can have a large forb component²³. In terms of species richness, 142 forbs dominate grasslands worldwide, and are often the most species-rich functional group²³. 143 Natural grasslands are grazed by wild herbivores, but large areas are also used for domestic 144 145 livestock. As such, meat, dairy and other livestock products, such as wool, comprise the most prominent ecosystem services provided by both natural and semi-natural grasslands. Indeed, 146 147 the livelihoods and cultures of some traditional pastoralist communities revolve almost entirely around their livestock systems²⁴, which in turn maintains the diversity of livestock breeds²⁵. 148 149 Natural grasslands also support rich biodiversity, hosting numerous species not found in forests^{12,26} and distinct assemblages of native large herbivores and their predators, many of 150 which are threatened by extinction²⁷. Many of these species are charismatic and globally 151 152 recognized, such as bison, wolves, rhinos and lions, and hence provide cultural services, such as $ecotourism^{28}$. 153

154 It is less acknowledged, however, that both natural and semi-natural grasslands provide 155 other ecosystem services, including holding flood waters, providing clean water, regulating 156 soil erosion, and providing wild food and medicines¹⁰. These services and their importance 157 vary geographically in relation to local environments and the demands from people¹⁰. Cultural 158 services include aesthetic value, hunting, heritage values and resources for tourism and 159 recreation, especially where grasslands define the ecology and culture of an area²⁹.

The role of grasslands in carbon storage is often overlooked, but could contain 30% of 160 the world's soil carbon stock³⁰, with natural and sparsely grazed grasslands potentially acting 161 as a notable global carbon sink and providing an important climate cooling service³¹. However, 162 163 there is a global trend of grasslands transitioning towards a net warming effect on climate due 164 to increased greenhouse gas emission associated with higher livestock numbers and management intensification³¹. Natural and sparsely grazed grasslands contain "irrecoverable 165 166 carbon" that is vulnerable to land use conversion; once lost, this carbon is not recoverable over timescales relevant to climate mitigation³². Nevertheless, there is high potential for increasing 167 168 soil carbon sequestration in grasslands via improved grazing and by arresting grassland conversion and degradation³³. 169

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171 [H1] Extent and drivers of degradation

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Globally, estimates suggest that as much as 49% of the total grassland^{5,7,8} and about 173 half of the natural grassland area has been degraded to some extent³⁴. However, specific 174 estimates of the extent of grassland degradation vary greatly. For instance, about 90% of the 175 176 UK's semi-natural species-rich grasslands have been degraded as result of intensive agriculture and land conversion since the 1940s¹⁷. Up to 90% of the vast grasslands of the Qinghai-Tibetan 177 Plateau have been degraded due to human activities and climate change³⁵. Over 60% of the 178 179 former grassland area of southern Brazil has been lost to unsuitable management and land use change³⁶. 180

181 A key issue when considering grassland degradation is clarifying what it constitutes, 182 especially since definitions of both grassland and degradation vary. Degradation is broadly 183 defined as a decline in land condition, or more specifically the human-caused processes that 184 drive a persistent decline or loss in biodiversity, ecosystem functions or ecosystem services³⁴. 185 Such a broad definition of grassland degradation can help to inform policy and practice about the level of grassland degradation in a certain region. However, pinpointing what degradation 186 constitutes in specific settings can be less clear because it has both ecological and 187 socioeconomic dimensions³⁷, and different stakeholder groups prioritise different 188 combinations of ecosystems services^{38,39}. As such, the definition of grassland degradation can 189 190 depend upon the stakeholder and needs to be tailored accordingly.

191 Grassland degradation is defined here from a socio-ecological perspective, where 192 grassland is considered degraded if the supply of multiple ecosystem services falls short of that demanded by grassland stakeholders⁴⁰. In some cases, grassland degradation is apparent to all 193 grassland stakeholders, for instance when overgrazing leads to loss of vegetation, declines in 194 soil organic matter and consequent soil erosion⁴¹, or when natural or semi-natural grassland is 195 196 converted to another land use. However, in many cases, perceptions of degradation could 197 differ: the increase in plant production but accompanying loss of plant species diversity 198 resulting from fertiliser use might be considered an improvement by pastoralists (due to 199 increased forage production), but degradation by conservationists concerned with biodiversity 200 protection⁴². Hence, defining grassland degradation from a socio-ecological perspective 201 recognises that it alters the supply of multiple ecosystem services from grassland relative to their human demand by different stakeholders⁴⁰. Moreover, a socio-ecological approach 202 provides a framework for guiding the restoration of degraded grasslands as it considers the 203 204 need to enhance the co-supply of multiple ecosystem services in efforts to meet the needs of all stakeholders^{37,40,43}. 205

206 Although threats to natural and semi-natural grasslands are present globally (Fig. 1), the tropics face particularly acute threats¹⁸, while degradation of European semi-natural 207 grassland has largely slowed⁴⁴. Human activities are the principal drivers of grassland 208 209 degradation. For example, increased disturbance from over-grazing or a heightened fire 210 frequency has reduced vegetation cover, increasing susceptibility to soil erosion and 211 desertification⁴⁵. Conversely, the cessation of livestock grazing associated with extensive land abandonment of semi-natural grasslands in Europe during the 20th century has led to grassland 212 degradation due to the expansion of scrub^{46.} Nutrient enrichment of natural and semi-natural 213 214 grasslands from fertiliser use and/or atmospheric nitrogen deposition has led to widespread declines in biodiversity and other ecosystem services^{17,47}. These impacts are especially notable 215 216 when combined with the sowing of productive cultivars to support heavy grazing and/or silage 217 production in intensively managed grasslands³. The conversion of natural and semi-natural 218 grasslands to other land uses, such as arable farming, built infrastructure and forestry, is also a major driver of grassland degradation worldwide^{36,48}. One immediate land use threat is the 219 220 planting of trees on natural grasslands, ostensibly to meet afforestation targets for climate 221 change mitigaton⁴⁹.

222 Ongoing climate change also poses a threat to all grasslands, as it causes grassland 223 degradation. Projected future climate change will likely combine with human activities to cause increased woody plant encroachment in some areas and desertification in others^{50,51}. For 224 225 example, natural grasslands in the Americas, Australia and Africa are being degraded due to 226 woody plant encroachment, with the major causes thought to be a combination of higher 227 atmospheric CO₂ concentrations, warmer and wetter conditions, and changes in grazing 228 intensity and timing relative to fire, which is key to episodic tree recruitment⁵². Conversely, in many parts of the world, such as Central Asia and Africa, over-grazing combined with more 229 230 intense and frequent droughts is exacerbating grassland desertification and degradation⁵³.

231

232 [H1] Promoting restoration

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A socio-ecological approach is also key to identifying how restoration of degraded grasslands might be achieved^{37,43}. Five actions are proposed here that should be deployed to develop effective socio-ecological solutions to the halting the degradation and promoting the restoration of global grasslands.

[H2] Policy recognition of grasslands. National and global policy are needed to recognize the

role of grasslands in addressing climate mitigation, food security, biodiversity conservation and poverty alleviation. A major factor contributing to the degradation of grasslands is the lack of representation of grassland in national and international policy. In fact, grasslands are often underappreciated, with the scientific and political focus of global sustainability policy being on ecosystems such as oceans, freshwaters, forests and croplands. Indeed, despite suffering some of the highest rates of degradation of any biome, only 8% of grassland and savanna biomes are protected⁵⁴.

The reasons for this neglect of grasslands in sustainability policy are unclear, but it 246 247 likely relates to historical factors and misconceptions about grasslands. For example, natural 248 grasslands are often erroneously considered to be degraded lands or wastelands, a result of 249 human action, or early successional stages of forest, whereas forests are perceived as more productive, pristine ecosystems that are rich in diversity^{18,55}. Moreover, natural grasslands 250 251 primarily occur in areas where climatic and soil factors only allow for relatively low economic 252 benefits to land users, whereas their economic value for the provision of ecosystem services has only been fully recognised over the last twenty years ⁵⁶. Another reason for this bias might 253 254 be that worldwide, forests were the first lands to be converted to agriculture, which has 255 contributed to greater awareness of the consequences of deforestation for biodiversity and 256 society. In contrast, most natural and semi-natural grasslands continued to be grazed at low 257 densities by wild herbivores and livestock until technological barriers hindering agricultural 258 improvement of grasslands and land conversion were overcome. As such, the consequences of 259 large-scale grassland degradation, which are also less obvious than cutting down trees, are 260 newly appreciated⁵⁷.

261 Although grasslands are recognised in some national policies on environment and 262 sustainability, such as China's Grassland Law, numerous prominent examples exist where they 263 have been ignored. For example, halting and reversing ecosystem degradation is a central goal 264 of the SDGs (Goal 15), the UN Framework Convention on Climate Change (UNFCCC), and 265 the Convention on Biological Diversity (CBD), but there is no explicit mention of grasslands 266 in any of them. The SDG targets also fail to mention grasslands, whereas forests, oceans and 267 coastal ecosystems are repeatedly referenced in targets and indicators. Further, while around 268 half the world's 234 Centres of Plant Diversity-sites of global botanical importance based on 269 endemism and species richness-are contained in grassland regions, they were not mentioned 270 in the Aichi Biodiversity Targets of the CBD. The European Biodiversity Strategy to 2020 also 271 makes little mention of grasslands relative to forests, which feature heavily. Although the EU 272 Common Agricultural Policy (CAP) does aim to protect grasslands, there has been widespread criticism about the effectiveness of the grassland measures for biodiversity conservation in this
 policy⁵⁸.

275 Pressures on grasslands have also been accentuated through the unforeseen 276 consequences of global and national environmental and socio-economic policies. In particular, 277 natural grasslands are increasingly targeted in carbon sequestration programs that emphasize 278 tree planting, such as REDD+ and China's Grain-to-Green Project, which often leads to their 279 degradation. The Chinese government, for example, has invested heavily in tree planting, with 280 tree coverage of previously non-forested areas increasing at an average rate of 0.15 million hectares per vear over the last 25 years⁵⁹. However, there is major concern that tree planting in 281 arid and semi-arid grasslands with water-demanding trees is exacerbating grassland 282 283 degradation, reducing plant diversity, damaging soils, and increasing water shortages⁶⁰⁻⁶².

284 Similarly, large tracts of natural grassland in Brazil have been identified as targets for tree planting, posing a major threat to these ancient and highly diverse ecosystems²⁶. An 285 286 estimated 7.41 million ha of eucalypt and 2.07 million ha of pine plantation already occupy 287 mostly former savannas and grassland, at a rapid rate of 0.4 million ha per year from 2013-2017 (Ref⁶³). Indeed, it has been calculated that meeting land restoration and protection targets 288 would increase global tree cover by 4 million km², often at the expense of grasslands⁶⁴. 289 290 Furthermore, models of global forest restoration potential have identified natural grasslands as 291 sites for restoration using tree planting, with potential to contribute to carbon sequestration and 292 climate mitigation⁶⁵. The logic of such proposals for climate mitigation, however, is 293 increasingly been challenged on the basis of their damaging impact on natural grasslands and because their carbon sequestration potential is considered inflated^{49,66.} 294

Evidence is mounting that grasslands can be more resilient carbon sinks than forests⁶⁷, 295 and afforestation can cause soil carbon loss^{68,69}, soil acidification and nutrient-depletion⁷⁰, 296 especially when trees are planted in natural grasslands⁶⁸, which can make them prone to carbon 297 298 loss from fires⁶⁶. Large-scale afforestation also leads to changes in surface albedo, given that 299 forests absorb more short-wave radiation than grasslands, thereby creating a warming effect⁷¹. 300 As such, changes in albedo resulting from afforestation can reduce or even negate benefits of increased carbon capture, potentially leading to a net warming effect of tree planting⁷¹. Another 301 302 issue is that policies such as REDD+ focus primarily on carbon sequestration in aboveground 303 tree biomass, while healthy and restored grasslands can store comparable amounts of organic carbon as forests, but mainly belowground^{67,72}. Grasslands have also been shown to be more 304 effective than forests in providing soil erosion control and water protection in semi-arid 305 ecosystems⁷³, and in some situations the conversion of grassland to forest, either through 306

natural regeneration or afforestation, can be highly detrimental to people who depend on
 grasslands for forage, game habitat, water reserves, and cultural services^{66,74,75.}

309 The lack of emphasis on grasslands in international and national policies has a long 310 history. Grasslands in India have been historically undervalued in national policies, and still 311 today they are widely considered to be unproductive wastelands of limited value, leaving them vulnerable to land conversion^{76,77}. In Brazil, major progress has been made in the conservation 312 313 of forest ecosystems, but non-forest biomes-including the Cerrado and Pampas grasslands-314 have been largely neglected, despite being among the most species rich grasslands in the 315 world⁷⁸. Similarly, and despite their ancient origin, Asian savannas have been considered 316 degraded forest since colonial times, which has led to inappropriate management and policies 317 that promote savanna degradation and loss^{77,79}. China's grassland law now includes policies to 318 protect healthy and restore degraded grasslands, and progress being made in certain areas⁸⁰, 319 but was not incorporated in the Constitution of the People's Republic of China until 1985.

320 If grasslands are to be valued and managed sustainably, then both global and regional policy must change to recognize the value of grasslands as providers of multiple ecosystem 321 322 services and hotspots for biodiversity, and to establish targets for their protection and 323 sustainable management. Approaches could comprise specific inclusion of natural and semi-324 natural grasslands in conservation laws and policies and ensuring conservation policies do not 325 have perverse consequences for grasslands, such as promoting fire suppression or afforestation. 326 New policies to promote and fund appropriate management and restoration of grasslands are 327 also needed, including penalties for degrading activities such as overgrazing.

328

329 [H2] Standardised assessment. New approaches are needed to enable standardised assessment 330 of grassland condition globally under different situations. These approaches should assess the 331 severity of grassland degradation and its consequences for biodiversity and ecosystem services 332 and evaluate the success of restoration schemes. Currently, restoration efforts are disjointed across regions and carried out by a wide range of organisations, which often leads to 333 334 incompatible and inaccessible datasets, and a lack of communication about successful methods⁸¹. The diversity of grasslands across and within regions, and the many drivers of 335 336 degradation, mean a fully uniform set of guidelines is impractical. Yet, standardised 337 approaches are needed to underpin effective decision-making.

Progress can be made by defining grassland degradation from a socio-ecological perspective, where the severity of degradation and the relative benefits of different grassland restoration practices can be assessed by combining measures of ecosystem service supply and 341 stakeholder priorities and demands⁴⁰. Such measures can show how alternative restoration 342 options differ in their impacts on ecosystem service supply and in their resulting benefits for 343 different stakeholder groups. We recommend that this goal be achieved via a general five-step 344 approach for standardising the assessment of grassland degradation and restoration (Fig. 2). 345 The results of this approach can also be used to support negotiation over restoration 346 management.

347 The first step is to assess the demand for, and relative priority of, different ecosystem services by multiple stakeholder groups, for instance via social surveys of many stakeholders, 348 or representative community leaders^{38,82} (Fig. 2). It is vital that all major stakeholder groups 349 are represented and the full range of relevant ecosystem services considered, including non-350 351 material (cultural) benefits. Otherwise, bad management decisions could result and potential conflict⁸³. Moreover, it is important that potential power imbalances among stakeholders are 352 353 addressed, for instance by strengthening and enforcing land use rights of pastoralists and agro-354 pastoralists.

355 Once ecosystem service demand is determined, indicators need to be identified and 356 measured for each of the services (Fig. 2). These should fall into two classes: general indicators, 357 which measure ecosystem services for all stakeholder groups (measured in step two); and 358 specific indicators for services used by a subset of the stakeholder community (measured in 359 step three). The general indicators measured in step two should be inexpensive and easy to 360 measure over large scales, and relate to key regulating functions that underpin all other 361 ecosystem services. Such properties could include the cover and type of vegetation present, 362 aboveground biomass, and soil properties such as organic matter, nutrient content, and pH 363 (Table 1).

364 While the general indicators proposed above provide broadly comparable information on grassland condition, perceptions of grassland degradation differ between stakeholder 365 366 groups⁸⁴. Therefore, in addition to general indicators, other variables should be identified locally for these context-specific ecosystem services via stakeholder engagement, and 367 368 developed according to local conditions (step three). In the French Alps, for example, surveys 369 identified that aesthetic value of mountain grasslands was essential to local farmers and 370 tourists, and related not only to flower diversity, but also to the absence of a build-up of plant litter⁸⁵. As another example, the cover of woody species can be used as an indicator of the vital 371 372 regulating service of shade provision in East African grasslands⁸⁶. Visually estimable indicators can also be employed in such assessments to ensure both relevance and cost-373 374 effectiveness.

375 In the fourth step, general and specific measures are combined to calculate an integrated 376 index of grassland condition that can be related to local environmental conditions, management 377 factors, global change drivers and restoration management. The creation of such indices can 378 be achieved by adapting approaches developed for measuring the co-delivery of multiple ecosystem services based on stakeholder preferences^{40,87}, and other participatory multi-criteria 379 analyses of ecosystem services⁸⁸. Both approaches can also be used to weigh the measures 380 381 included in the calculation of the index, for example by encouraging representatives of different 382 stakeholder groups to assign scores to the ecosystem services considered⁸⁶.

383 In the final step, conclusions from this standardised approach are used to inform 384 management decisions regarding which restoration options should be employed and where. 385 Deciding on which restoration options are employed and where is best resolved in participatory 386 approaches, so that stakeholders share understanding, and conflict is minimised (Fig. 2). The 387 best options for whole communities can be estimated by weighting stakeholder groups equally 388 in a community level metric, or by evaluating which restoration options minimise trade-offs 389 between groups. As a single restoration practice might not benefit all stakeholder groups, the 390 compartmentalisation of the landscape into different restoration options should also be 391 considered (Fig. 2). The approach presented here would not only allow for more detailed 392 examination of grassland condition with regards to local needs, but also a better means of 393 assessing the severity of grassland degradation and restoration success at both local and larger 394 scales (Figures 3 and 4; see Supplementary Material for details on the source of these 395 estimates).

396

397 [H2] Shared understanding and trade-offs. When assessing grassland restoration options that 398 best suit different stakeholder groups, there will likely be differences in preferred options (Figs. 399 3 & 4), which might cause disagreement and conflict. To address these differences, there is a 400 need for shared understanding, based on best knowledge and practice and operational 401 constraints, of the potential for different grassland restoration options to supply different 402 ecosystem services^{43,89}, which can be quantified using the same tools as described here for 403 assessing degradation. Such an approach could help identify which restoration option might 404 best deliver the needs of all stakeholder groups, thereby supporting resolution of disagreements 405 and conflict.

To illustrate these trade-offs and how they could be resolved, two examples are considered (Figs. 3 and 4). In both, fundamental trade-offs caused by degradation and restoration options prevent the co-supply of all ecosystem services. The first example is taken 409 from permanent agricultural grasslands in Europe (Fig. 3), where extensive management 410 promotes functions related to water and soil carbon storage, which support ecosystem services 411 linked to water and climate regulation. In contrast, intensive management with fertiliser use 412 and high grazing pressures leads to faster rates of nutrient cycling and higher plant productivity, thereby promoting provisioning services^{42,90}. However, intensive management also reduces 413 plant and soil biodiversity and causes soil compaction, which diminishes the supply of several 414 415 ecosystem services demanded by stakeholders, leading to reduced overall benefits to stakeholder groups (Fig. 3) 90 . To reverse these impacts, restoration to high diversity grassland 416 417 can enhance the supply of multiple ecosystem services, and therefore provides substantial 418 benefits to all stakeholder groups. Despite these benefits, the abandonment of agricultural 419 management altogether, which leads to woody plant encroachment, could be the best option to 420 provide some of the benefits prioritized by national decision makers, namely water quality and 421 climate regulation. Given these various trade-offs, and depending on European or national objectives, and on regional conditions⁹¹, collective choice would likely favour a mosaic of all 422 423 three grassland types with the restoration of high diversity grassland, accompanied by the 424 segregation of intensive production and climate regulation into specific regions and landscapes⁹². 425

426 The second example is taken from natural grasslands in arid and semi-arid regions of 427 Eastern Africa (Fig. 4). Here, nomadic pastoralists favour provisioning services of forage 428 production and natural products from grasslands, while charcoal producers derive their 429 livelihoods from native woody vegetation and, increasingly, from invasive woody species such as Prosopis juliflora. Because the invasion of P. juliflora transforms grasslands into scrubland 430 with bare soil⁹³, the supply of most ecosystem services demanded by stakeholders is greatly 431 reduced, but fuel wood and charcoal production are promoted⁸⁶. The supply of other services, 432 433 such as soil carbon storage, is context-dependent; soil conditions are improved when P. 434 juliflora invades grasslands degraded by overgrazing and other forms of mismanagement. However, if undegraded grassland is invaded, the supply of these services is reduced⁹⁴ and all 435 436 stakeholders, apart from charcoal producers, lose ecosystem service benefits and so perceive 437 the invasion as degradation. This example also clearly demonstrates that successful restoration 438 depends on stakeholder perspectives as all restoration options benefit the priority services of 439 some groups, while diminishing those of others. These constraints preclude a single win-win-440 win restoration option. Therefore, regional-scale restoration that establishes a landscape 441 mosaic, including livestock grazing and patches of conserved or abandoned land, is likely to 442 be required for all three stakeholder groups to be satisfied. It should be noted, however, that such a landscape mosaic would require continuous management of *P. juliflora*, otherwise it
will re-invade grasslands and conserved land.

445 Identifying the presence and cause of trade-offs is a first step towards management and 446 policy solutions to grassland degradation. We suggest that the approaches presented here could be combined as a decision support tool in ecosystem management⁹⁵. This tool can be used as 447 part of a 'landscape approach' to management, which emphasises close collaboration and 448 449 shared understanding between stakeholder groups to identify and mitigate land management 450 conflicts. Depending on context, this approach could allow for the allocation of different 451 restoration options to different sections of the landscape, as in semi-arid grasslands of Eastern 452 Africa (Figure 4), or the identification of win-win options where a single restoration practice 453 might suit multiple stakeholder groups, such as in European grasslands (Figure 3).

454 These approaches are already considered in existing grassland management 455 approaches, such as the Participatory Rangeland Management (PRM) approach, which has 456 been developed to improve the sustainable management of rangeland resources and their security of access for local rangeland users in arid and semi-arid grasslands⁹⁶. This approach 457 458 includes a step where stakeholder groups, through extensive meetings, discussions and 459 negotiations, develop a rangeland management agreement that considers the needs of all 460 stakeholder groups involved. This agreement should ultimately become a binding contract 461 document between a representative rangeland management institution and the appropriate 462 government authority. The PRM was initially developed by a consortium including the 463 International Livestock Research Institute in Ethiopia and has been applied in Kenya and Tanzania⁹⁷. A similar approach called Participatory Rangeland Management Planning (PRMP) 464 465 has been adopted by the International Union for Conservation of Nature $(IUCN)^{98}$.

466 Complex trade-offs might prevent the formulation of simple solutions, and in such situations multi-criteria decision analysis (MCDA) can be used to support deliberation among 467 468 groups and help steer them towards a compromise where local communities have a voice and multiple perspectives of grassland degradation are accommodated^{99,100}. Advanced MCDA 469 470 approaches also allow for the incorporation of socio-economic factors, such as access rights 471 and ease, and power relationships between stakeholder groups across scales, which will 472 determine the ultimate benefits of ecosystem services to different groups¹⁰¹. However, to ensure better outcomes from decision tools such as MCDA it is also important to consider 473 474 optimal allocation and prioritisation of limiting resources to actions¹⁰², especially since funds for grassland restoration are often limited. 475

476

477 [H2] Knowledge sharing in grassland restoration. Evidence of success or failure of grassland restoration programmes is scant^{103,104}. Accordingly, there is a need for standardised and 478 479 accessible reporting of restoration successes and failures, and ongoing monitoring of grassland 480 restoration programmes and their outcomes in different parts of the world. Restoration 481 knowledge must be shared to achieve this goal, which requires the formation of networks of 482 scientists and practitioners that exchange knowledge on successful, but also unsuccessful, 483 restoration efforts. Examples of platforms that provide information on successes and failures 484 of grassland restoration programmes are beginning to emerge. The Society for Ecological 485 Restoration (SER) launched their Restoration Resource Centre in 2017, an online, publicly accessible platform for exchanging knowledge and experience through ecological restoration 486 487 projects, publications, and other resources from around the world. The World Overview of 488 Conservation Approaches and Technologies (WOCAT) database provides a format and a 489 platform for knowledge exchange regarding Sustainable Land Management (SLM) 490 technologies and approaches and is recognized as the primary recommended database by 491 UNCCD. The Global Landscape Forum (GLF), led by the Centre for International Forestry 492 Research (CIFOR) and supported by the UN and World Bank, provides a mobile platform for 493 discussions leading to action on sustainable land use, including land restoration.

494 While grasslands are currently underrepresented within these databases, which are still 495 under development and continually being expanded, they demonstrate the potential for user-496 driven platforms to provide valuable insight into effective restoration practices in different 497 parts of the world (Box 1 Fig. 1). To be effective, such platforms should include a wide range 498 of restoration programmes, both local and regional, and small and large scale. They should also 499 include less successful programmes and identify why they have not reached their objectives. 500 Moreover, grassland restoration projects should identify clear aims of restoration and an 501 inventory before restoration starts to be able to evaluate restoration success¹⁰⁵. By comparing 502 case studies, it is possible to understand how the effectiveness of specific restoration methods 503 differ geographically and depending on the degree of degradation and socio-ecological context, 504 the first step towards finding general rules, both social and ecological, for successful 505 restoration. They could also illustrate the successful translation of research into improved 506 restoration and management, which should include the co-creation and co-implementation of 507 technical solutions to grassland restoration, and the political and socio-economic conditions 508 that made them possible.

509 Case studies reporting technical solutions to grassland degradation at scales of hundreds 510 to thousands of hectares are particularly scarce, especially for developing countries in semi511 arid and arid regions of the world. However, previous efforts indicate that grassland restoration 512 is feasible, at least starting from low levels of degradation, and that technical and socio-513 ecological solutions often need to be combined to achieve long-term management goals (Box 514 1). For example, steppe grasslands in Inner Mongolia are challenged with a diverse set of 515 stakeholder demands, including the need for environmental services, such as increased soil carbon storage, the local herders' wish to retain traditional livelihoods while increasing their 516 517 income, and an increasing demand for red meat by the wider population in China and 518 elsewhere⁸⁰. To address these demands of various stakeholder groups, a grassland management 519 system was developed that built on optimizing income from livestock production per unit area, 520 which was achieved by increasing animal growth rates (and thus reaching marketable size 521 earlier) by decreasing livestock numbers by approximately 50%⁸⁰. Thus, the system created a 522 win-win situation by improving local herders' income and enabling the recovery of steppe and 523 associated ecosystem services (Box 1.

524 In northern Ethiopia, several years of excluding livestock from degraded communal grazing lands led to the recovery of plant species richness, forage and wood production, and 525 526 soil health, and thus helped to improve the supply of services prioritised by local stakeholders^{106,107}. However, the uptake of technical solutions for restoring degraded 527 grasslands depends on socio-economic factors, including reconciliation with traditional land-528 529 tenure systems, well-established communication pathways, and supporting institutional 530 settings and policies (Box 1. For example, implementation of sustainable grassland 531 management in Inner Mongolia benefits from efforts of the Chinese government to emphasize sustainable grassland management in agricultural policies¹⁰⁸. The success of the grazing 532 533 exclosure programme in northern Ethiopia will depend on successful communication and 534 negotiations between local stakeholder groups and policy, and among local stakeholders to 535 ensure land use enforcement and sharing of management responsibilities and harvesting 536 benefits¹⁰⁹.

537 538

539 [H2] Research innovation: Innovations in environmental research need to be developed and 540 employed to improve the assessment of grassland degradation and provide novel solutions to 541 restoration and sustainable grassland management. At a global and regional scale, 542 developments in remote sensing offer potential to evaluate both the extent and status of 543 grasslands and inform the spatial targeting of large-scale restoration efforts. For example, maps 544 of general degradation indicators, such as primary productivity, standing biomass, soil 545 moisture, phenology, soil organic carbon¹¹⁰⁻¹¹², could be combined with spatial information on 546 climate, edaphic and socio-economic data to identify national and global patterns of grassland 547 degradation and pin-point locations where restoration efforts could have the greatest impact. 548 Although current global maps of primary productivity and biomass carbon still need 549 improvement, they offer a means to monitor the dynamics of grassland aboveground biomass 550 and evaluate the degradation of global grasslands^{113,114}.

551 There has been a boom in research directed at grassland management for sustainable agroecosystems since the turn of the 21st Century, leading to improved grazing and fertiliser 552 553 management, and reduced reliance on external inputs, nutrient losses and emissions of 554 greenhouse gases¹¹⁵ However, new ecological understanding could also facilitate grassland 555 restoration. For example, new research could enable landscape and regional scale restoration 556 of inter-connected grasslands in regions where semi-natural grassland remnants are fragmented 557 and persist within a matrix of forested or intensive agriculture land uses. Indeed, there is 558 increasing evidence that grassland dynamics and function are enhanced by connectivity that encourages flows of species and biogeochemical processes^{116,117}. In theory, such spatial 559 networks have improved resilience to perturbations, such as climate extremes, and the ability 560 to adapt to a changing environment¹¹⁸. These emerging concepts on landscape design could 561 also be applied to the restoration of ecological flows of species and processes within 562 landscapes^{119,120}, and at a practical level, there is already considerable opportunity to trial these 563 approaches in large-scale restoration programmes¹²¹. Such approaches also require improved 564 565 understanding of the tele-coupling of different landscape elements, whereby human-induced 566 processes in one area impact on distant areas. For example, pastoralists in arid regions depend on grazing areas along major water bodies, which are also of interest to other stakeholders for 567 crop production or biodiversity conservation¹²². Loss of such grazing areas forces pastoralists 568 569 to overgraze rain-fed grasslands elsewhere, leading grassland degradation¹²³.

570 Another fast-developing area of ecological research that could assist grassland restoration programmes concerns the application of knowledge of aboveground-belowground 571 interactions^{124,125}. A wealth of studies show that high grassland plant diversity can enhance 572 573 multiple ecosystem functions, both above- and belowground, and increase the resistance of plant production to environmental perturbations, such as climate extremes¹²⁶⁻¹²⁸; such 574 knowledge could be applied to degraded grassland to restore ecosystem functions and their 575 resilience to environmental change^{129,130}. Benefits can also be realised through informed 576 selection of plant species based on functional traits, especially root traits, which has been 577 shown to be an effective way of enhancing grassland multifunctionality¹³¹ and the physical 578

properties of degraded soils¹³². Research points to the potential of soil inoculation and plant translocation from donor sites to enhance the recovery of degraded ecosystems and steer plant community development^{133,134}. Further, there is increasing awareness that diversifying livestock promotes plant diversity and grassland multifunctionality¹³⁵, especially at moderate levels of grazing¹³⁶, and that interventions can facilitate robust multi-trophic interactions in restored ecosystems could benefit restoration¹³⁷.

585 A potential hurdle to the adoption of new research innovation is that the majority of 586 scientific knowledge underpinning grassland restoration comes from studies done at a local 587 scale and in a handful of regions, especially North America, Australasia and Europe. As such, 588 there is a need for new, long-term research to underpin approaches to large-scale restoration of 589 degraded grasslands, especially in Asia, Africa and South America where grassland degradation is widespread¹³⁸. One solution to these geographic gaps might be the use of meta-590 level analyses¹³⁹⁻¹⁴⁰ and data collection using standardised methodology from globally 591 distributed grassland studies such as the Nutrient Network¹⁴¹ and HerbDivNet¹⁴². These can 592 593 help to identify generalisable impacts of biotic and environmental drivers and disturbances, 594 such as grazing and nutrient enrichment, on grassland diversity and function, that could be 595 transferred to regions where data is lacking. However, that large gaps in environmental and 596 management conditions still need to be filled, and the ecological context will affect the specific applicability of any generalisations¹⁴³. At all scales, new technology and practices need to be 597 598 embedded within socio-ecological approaches to grassland restoration to ensure empowerment 599 of local people and best outcomes for their quality of life.

600

601 [H1] Summary and future perspectives

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603 Despite progress in understanding the causes and consequences of grassland degradation, and 604 in developing restoration techniques, there remain many barriers to halting grassland 605 degradation and effective restoration. Here, we have argued that overcoming these barriers 606 requires an integrated socio-ecological approach to grassland degradation and restoration, which not only demands greater recognition of grasslands in global policy, but standardised 607 608 approaches for assessing grassland degradation and restoration success. It also requires the 609 adoption of innovations in environmental science for detecting grassland degradation and 610 enhancing restoration success, along with enhanced knowledge transfer and data sharing 611 regarding restoration experiences.

612 Many examples exist of national and international sustainability policies that have ignored grasslands, including the SDGs and CBD. Giving due attention in sustainability policy 613 614 to grasslands and the ecosystem services they provide, on a par with other biomes such as 615 forests, is therefore essential for the future protection of healthy grasslands and restoration of 616 those that are degraded. The UN Decade on Ecosystem Restoration (2021-2030) and CBD 617 post-2020 global biodiversity framework provides an opportunity to set future targets for the 618 protection, sustainable management and restoration of grasslands, and prevent damaging practices such as the planting of trees in natural grasslands¹⁴⁴. Put simply, if grasslands are to 619 620 be managed sustainably, then both global and regional policy must be revised to recognize the 621 value of grasslands for multiple ecosystem services and establish targets for their protection, 622 restoration and sustainable management.

623 There is hope in some initiatives, though. The Worldwide Fund for Nature launched a 624 'Global Grassland & Savannah Dialogue Platform' in 2020, with the aim to "develop" 625 consensus around human and biological importance of these ecosystems". The UN Decade on 626 Ecosystem Restoration (2021-2030) has added 'grasslands, shrubland and savannahs' to their 627 set of focal ecosystems. Furthermore, while the Aichi targets of the CBD largely ignored 628 grasslands, and the draft of its post-2020 framework lacks specificity, the most recent synthesis 629 of the proposals of the parties includes mentions of grasslands (albeit other ecosystems such as 630 forest and wetlands are discussed more frequently).

631 Whilst we demonstrate how our five-step approach can be applied using specific case 632 studies, future research is needed to test this approach in different contexts and at local and 633 larger scales. Research is also needed to better understand different societal perceptions of 634 grasslands and the reasons why they have been neglected in sustainability policy, to develop 635 and test promising new ways of assessing grassland degradation and restoration, and to harness 636 ecological knowledge for restoration success. This aim will require ambitious, interdisciplinary 637 national and international research programmes of the kind that could be facilitated by major 638 research and innovation schemes targeted at achieving Sustainable Development Goals. We 639 hope the approach we present provides a basis for such future research aimed at the assessment 640 of grassland degradation and restoration in the context of stakeholder needs.

The actions we raise for developing effective socio-ecological solutions to the degradation and restoration of global grasslands are not exhaustive and many challenges remain. Nevertheless, we hope that they provide a guide to future research and policy needs for halting grassland degradation and achieving restoration success. We also hope that they serve to raise awareness of the plight of global grasslands and the need for urgent action to halt

| 646 | grassla | and degradation and enhance restoration success, thereby conserving the many socio- |
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| 647 | econor | mic, cultural and ecological benefits that grasslands provide. |
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- 1028

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1039 Author contributions

- 1040 R.D.B. conceived the idea and gained funding with input from N.O. R.D.B wrote the paper
- 1041 with significant input from J.M.B, P.M., U.S. and S.L. G.L.P and P.M. designed the figures.
- 1042 All authors contributed to the development of ideas and writing of the paper.
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1051

Table 1. Common ecosystem services of grasslands and indicators of associated ecosystem structure and function parameters

 1055 1056

| Ecosystem services | Ecosystem structure / function | Common indicators | Scale and means of measurement |
|-------------------------------------|-------------------------------------------------------------------------------------------------|----------------------------------------------------------------|------------------------------------------------------------------|
| Forage production (quantity) | Annual aboveground biomass production | Net primary productivity (NPP) | NDVI (remote sensing) at a landscape scale |
| | | Aboveground biomass | Direct harvesting at a local scale |
| Forage production (quality) | Protein content and digestibility | Leaf N content Presence of species | Remote sensing at a landscape scale |
| | | of nutritional importance | Direct measures at a local scale |
| Forage reliability | Interannual variation in aboveground | NPP | Remote sensing at a landscape scale |
| Other grassland | biomass production Species of particular | Species composition Species presence | Field survey at a |
| products (medicinal, food, hunting) | interest | and/or abundance | local scale |
| Biofuels | Woody species of interest | Cover or biomass of species of interest | Direct field measures or remote sensing |
| | Grass species of particular interest (including flammable or high yielding species) | | |
| Species of cultural value | Presence of species of cultural interest | Species presence | Local direct measures and records |
| | | | Regional surveys |
| Aesthetic value | Plant community composition and phenology | Flower and flower colour, and presence of 'unattractive' | Direct surveys Species list and |
| | | species Flowering | external databases for phenology |
| | | phenology | |
| Biodiversity | Plant and animal | Presence and | Global Biodiversity |
| conservation value | (vertebrate and invertebrate) species | abundance of species of conservation value | Information Facility (GBIF) and nationa monitoring schemes |

| | | | at larger scales |
|---------------------------------------------------------------------|------------------------------------------------------|-----------------------------------------------------------------------------|-----------------------------------------------------------------------------------|
| | | | Direct surveys at a local scale |
| Regulation of invasive exotics and other undesired species | Invasive exotic species Species of negative | Presence and abundance of undesired species | GBIF and national monitoring schemes at a larger scale |
| | pastoral or cultural value | | Direct surveys at a local scale |
| Global climate regulation | Carbon stocks and carbon cycling processes | Soil respiration and carbon stocks | Remote sensing for aboveground stocks |
| | | Woody species biomass and vegetation carbon stocks | Soil sampling and interpolation for belowground carbon stocks and fluxes |
| | | Litter mass and depth | |
| Maintenance of soil fertility | Nutrient stocks and nutrient cycling processes | Soil nutrient and carbon content | Soil sampling and interpolation to required scales |
| | 1 | Litter mass | |
| | | Soil enzyme activities | |
| Maintenance of soil stability and regulation of erosion | Soil stability in the root profile | Evidence for erosion, bare ground cover and soil | Field observations or remote sensing to detect erosion |
| C | Erosive flows | organic matter (SOM), and measures of soil loss | Direct sampling and measurement of |
| | | and erosive flows | multiple indicators |
| | | Soil aggregate stability, bulk density, and water holding capacity | |
| | | Plant rooting profile | |
| Regulation of hydrological flows | Soil water retention and flows | Soil texture and bulk density, and SOM content | Direct survey and interpolation |
| | | Soil electrical and hydraulic conductivity) | Rapid assessment methods |

| | Regulation of water quality | Retention and transformation of pollutants in soil | Soil properties including texture, pH, CEC, salinity and water table depth | Direct sampling of soil and water |
|----------------------------------------------|-----------------------------------------------------------------------------------------------------|----------------------------------------------------------|----------------------------------------------------------------------------------------|-----------------------------------|
| | | | SOM content and available water capacity. | |
| | | | Nutrient and pollutant concentrations in freshwater bodies | |
| 1057 1058 1059 1060 1061 1062 | | | inesitwater boules | |
| 1063 | Figure legends | | | |
| 1064 1065 | Figure 1. Degraded g | rasslands. The extent | of degraded grasslands v | worldwide with |
| 1066 | examples of paired non-degraded (left) and degraded (right) grasslands. Grassland | | | |
| 1067 | classification follows the UN FAO Land Cover Classification System (LCCS) (data | | | |
| 1068 | downloaded at <u>https://lcviewer.vito.be/2015</u> ¹⁴⁵ with tundra ecosystems excluded). | | | |
| 1069 | Degradation is measur | ed as greenness change | es, as measured by rain u | use efficiency (RUE) |
| 1070 | adjusted Sum Normalized Differential Vegetation Index (NDVI) between (1981-2015) ¹⁴⁶ | | | |
| 1071 | with regions showing a reduction in greenness of 0.01 being classed as degraded. Therefore, | | | |
| 1072 | much degradation invo | olving vegetation chang | ge is not shown. Degrada | ation is caused by many |
| 1073 | factors, including over | grazing, fertilization, t | ree planting, and invasiv | ve species. Image credits: |
| 1074 | United States (L. Brud | vig), United Kingdom | (L. Hulmes), India (S.K | . Chengappa), and |
| 1075 | Australia (S. Prober). I | Map © Copernicus Ser | vice Information year of | publication [2015] |
| 1076 1077 1078 | Figure 2. Standardise | ed assessment of grass | land degradation and 1 | restoration. Included are |
| 1079 | the steps required to a | ssess grassland conditi | on and select restoratio | n options. These include |
| 1080 | | | | akeholders (step 1) and |
| 1081 | identification of unive | rsal (step 2) and speci | fic (step 3) ecosystem s | ervice, the evaluation of |
| 1082 | grassland restoration a | nd identification of res | toration options (step 4) | , and the shared selection |
| 1083 | and implementation of | Frestoration strategy (s | tep 5). The relationsh | ip between this approach |
| | | | | |

and the other strategies described in this Perspective are shown on the side panel. For specific
and quantitative examples of steps 1 and 4, see Figures 3 and 4. For steps 2 and 3, which relate
to identifying indicators, see Table 1.

1087

1088 Figure 3. Assessing ecosystem service trade-offs in degraded and restored European 1089 grasslands. Steps 1 and 4 from Figure 2 are shown in greater detail so that the quantitative 1090 methods can be demonstrated. In this system, degradation caused by intensive agriculture 1091 reduces plant and soil biodiversity and causes soil compaction. This degradation diminishes 1092 the supply (SD) of ecosystem services demanded (D, represented for each group as D followed 1093 by a number) by stakeholders, resulting in a low multifunctionality (MF) for all stakeholder 1094 groups (D x SD). Restoration to high diversity grassland (SR1: Conservation) enhances the 1095 supply of multiple ecosystem services, though only moderately for fodder production and 1096 climate regulation through carbon storage. Thus, it provides the greatest multifunctionality to 1097 all stakeholder groups among restoration options. Restored high diversity grassland also 1098 provides the best option for priority ecosystem services (PR) for conservationists (see the D1 1099 x SR1 circle), and for local farmers given their demand for organically-maintained soil fertility 1100 and cultural identity (D2 x SR2), which could compensate for lower fodder production than 1101 intensive grassland (as in SR2: Livestock grazing). However, abandonment of agricultural 1102 management (SR3: Abandoned grassland) could be seen as favourable by some national 1103 decision makers due to their prioritization of water quality and climate regulation (D3 x SR3 1104 circle). All MF and PR scores are scaled between zero and one to provide a comparable metric 1105 and check marks (+, ++) indicate the restoration option with the highest MF or PR.

- 1106
- 1107
- 1108
- 1109

1110 Figure 4. Assessing ecosystem service trade-offs in degraded and restored grasslands in 1111 arid and semi-arid Eastern Africa. Steps 1 and 4 from Figure 2 are shown in greater detail 1112 so that the quantitative methods can be demonstrated. In these regions, invasion of alien woody 1113 plants such as Prosopis juliflora transforms grasslands into scrubland with bare soil. While the 1114 supply (SD) of most ecosystem services (ES) demanded (D) by stakeholders is greatly reduced 1115 by this invasion, fuel wood and charcoal production are promoted (as in D2). The supply of 1116 other services, such as soil carbon storage, is context-dependent; if grassland previously 1117 degraded by overgrazing and other forms of mismanagement is invaded, then soil conditions

| 1118 | are improved. However, if healthy grassland is invaded, the supply of these services is reduced. |
|--------------|-------------------------------------------------------------------------------------------------------------------|
| 1119 | All stakeholders, apart from charcoal producers, lose ecosystem service benefits (D2 x SD) if |
| 1120 | undegraded grasslands are invaded by P. juliflora, and so perceive it as degradation. |
| 1121 | Restoration options have differing impacts on stakeholder groups. Sustainable livestock |
| 1122 | grazing (SR2) requires fencing off or access restriction but promotes ES supply more than |
| 1123 | conservation (SR1) and supports the greatest multifunctionality (MF) from the perspective of |
| 1124 | all the considered groups. Abandonment does not restore the supply of any ES and even incurs |
| 1125 | ES losses (SR3), because <i>P. juliflora</i> consumes substantial amounts of groundwater. However, |
| | |
| 1126 | considering economic priorities towards only the priority ES (PR) of fodder for pastoralists, |
| 1127 | wood products for charcoal producers and ecotourism for tourists would lead these |
| 1128 | stakeholders to respectively favour livestock production, abandonment or conservation as best |
| 1129 | restoration options (highest respective PR scores). All MF and PR scores are scaled between |
| 1130 | zero and one to provide a comparable metric and check marks (+, ++) indicate the restoration |
| 1131 | option with the highest MF or PR. |
| 1132 | |
| 1133 | |
| 1134 | |
| 1135 | |
| 1136 1137 | |
| 1137 | |
| 1139 | Box 1 Case studies of successful grassland restoration. |
| 1140 | |
| 1141 1142 | [bH1] Ethiopia (Tigray & Amhara region) |
| 1142 | Ref ¹⁴⁷⁻¹⁴⁸ |
| 1144 | |
| 1145 | [bH2] Restoration - Needs and Actions |
| 1146 1147 | [b1] Drivers of degradation: overgrazing [b1] People's needs: fodder, wood and water |
| 1147 | [b1] Restoration goal: promote soil fertility, herbaceous layer and tree growth, increase water |
| 1140 | retention |
| 1150 | [b1] Actions: time series of area exclosures (5-20 years) protected by community bylaws and |
| 1151 | /or hired guards. |
| 1152 | (H2) Outcomes(+/) |
| 1153 1154 | [bH2] Outcomes (+/-) [b1] 2 to 3-fold increase in plant species richness, compared to areas outside exclosures |
| 1154 | [b1] 2 to 3-fold increase in plant species fremess, compared to areas outside exclosures |
| 1156 | [b1] Increased soil nitrogen, available phosphorus and cation exchange capacity |
| 1157 | [b1] Erosion and flood perceived to be highly reduced and soil moisture increased |
| 1158 | [b1] Improved provision of fodder for stall-fed livestock |
| | |

[b1] No published information on the economic benefits for local livelihoods.

1161 [bH2] Opportunities and challenges

- [b1] Opportunities: government- and NGO-provided incentives including food for work andpayments for the protection of exclosures
- 1165 [b1] Challenges: government has stopped paying incentives in some areas; membership to
- 1166 exclosures and distribution of benefits unclear; decision processes and forms of collaboration
- 1167 need clarification; limited participation by women in decision processes.
- 11681169 [bH1] Brazil (Cerrado Region)
- 1170 Refs ^{134,149}
- 1171

1160

1172 [bH2] Restoration - Needs and Actions

- 1173 [b1] Drivers of degradation: pine plantations (afforestation) and invasions
- 1174 [b1] People's needs: water
- 1175 [b1] Restoration goal: recovery of herbaceous layer; increase habitat for grassland fauna
- 1176 threatened by land conversion and woody encroachment
- 1177 [b1] Actions: cutting pines, prescribed burns, transplanting native grasses, controlling re-
- 1178 invasion by pines and exotic grasses

1179 [bH2] Outcomes (+/-)

- 1180 [b1] Within two years, recovery of 70% the ground cover and 55% plant species relative to
- 1181 reference ecosystem.
- 1182 [b1] 65 plant species recorded after two years
- 1183 [b1] Increase of grassland area in the protected area by 5%
- 1184 [b1] Rise in the water table in a cleared watershed
- 1185 [b1] Economic benefits not expected nor quantified
- 1186

1187 [bH2] Opportunities and challenges

- 1188
- 1189 [b1] Opportunities: eradication of pines mandatory in protected areas; companies obliged to
- 1190 mitigate environmental damage by carrying out restoration work experiments provide 1191 scientific support for the practices applied
- 1192 [b1] Challenges: tendency for woody encroachment and re-invasion by pines from the
- neighbourhood; fire management would be an important tool but there are legal constraints in
- 1194 restored ecosystems; invasion by exotic grasses is a permanent threat.
- 1195

1196 [bH1] China (Inner Mongolia)

- 1197 Ref^{80,150,151,152,153,154}
- 1198

1199 [bH2] Restoration - Needs and Actions

- 1200 [b1] Drivers of degradation: overgrazing
- 1201 [b1] People's needs: increased revenue from livestock
- [b1] Restoration goal: mutually beneficial outcome for environmental health and householdincome
- 1204 [b1] Actions: reduced stocking rates since 2004; optimization of management practices (such
- 1205 as the increased use of feed supplement in winter and spring, and retaining the most
- 1206 productive animals) to increase efficiency.
- 1207 [bH2] Outcomes (+/-)

- 1208 [b1] Reduction in stocking rate increased aboveground and below-ground biomass, species
- 1209 diversity, C sequestration and storage, and net sustainable livestock carrying capacities of 1210 desort grassland in China
- 1210 desert grassland in China
- 1211 [b1] Light and moderate grazing led to greater ecosystem stability
- 1212 [b1] Low stocking rate, reducing number of old animals and other optimization practices
- 1213 increased net household income and alleviated grazing pressure.

1214 [bH2] Opportunities and challenges

- 1215 [b1] adoption of policies for poverty alleviation and grassland restoration, including
- 1216 environmental payment schemes, hiring of supervisors to protect grassland ecosystems, or
- 1217 funding of warm sheds; adoption of an energy-balance/market-based approach to reduce
- 1218 livestock numbers; increased demand for meat in China
- 1219 [b1] Challenges: uptake of practices of reducing animal numbers and optimizing livestock
- 1220 management by pastoralists.
- 1221
- 1222

1223 [bH1] China (Qinghai Tibetan Plateau)

1224 Ref¹⁵⁵⁻¹⁵⁸

1225 [bH2] Restoration - Needs and Actions

- 1226 [b1] Drivers of degradation: human disturbance (overgrazing, excessive exploitation), climate
- 1227 change, rodent damage as secondary cause
- 1228 [b1] People's needs: reasonable income, sustainable grassland management, water
- 1229 conservation
- 1230 [b1] Restoration goal: promote herbaceous layer; restore soil health; reduce area of 'black
- 1231 soil' degraded grassland
- 1232 [b1] Actions: planting artificial or semi-artificial grasslands; reducing grazing pressure;
- 1233 establishing protected areas.

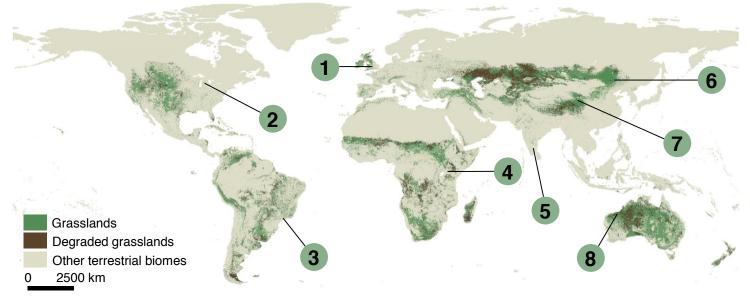
1234 *[bH2] Outcomes (+/-)*

- 1235 [b1] Continued degradation in grassland has initially been stopped
- 1236 [b1] In a 17-year chronosequence, grassland restoration increased plant biodiversity and
- 1237 vegetation cover; soil total nitrogen sequestration improved by 25%–40%
- 1238 [b1] Water supply capacity of watersheds has increased
- 1239 [b1] No significant change in economic profit along restoration chronosequence

1240 [bH2] Opportunities and challenges

- 1241 [b1] Opportunities: increased awareness that grassland degradation is not only a threat to
- 1242 pastoral livelihoods, but also to ecological stability of the region and to people living in other
- 1243 regions implementation of environmental payment schemes; government provided funding for
- 1244 an ecological conservation and restoration project in the Three-River Source Region
- 1245 [b1] Challenges: could affect livelihoods because it restricts access to grazing areas; change in
- 1246 grazing system can require high input costs (stalls, new feed sources); collaboration among
- 1247 farmers, local officials and extension workers required for technology transfer and policy
- 1248 implementation.

Global grassland degradation





United Kingdom

Ploughing, reseeding and fertilisation has transformed species-rich chalk grasslands into 'improved' grasslands with higher fertility, lower species richness and lower levels of many cultural ecosystem services.



United States

Reseeding and fertilisation of tallgrass prairie has transformed species-rich prairie grasslands with a mixture of native C3 and C4 grasses, sedges and forbs, to species-poor grassland dominated by Eurasian C3 grasses.



Brazil

Soil tillage and fertilisation for cultivation of Cerrado grasslands destroy underground plant structures and soil aggregates, severely lowering resilience of these ancient and highly diverse ecosystems.





Decades of overgrazing by cattle and sheep of lowland grassland in Kenya has caused excessive soil erosion, and loss of biodiversity and ecosystem services.





India

Invasion by exotic woody species, primarily *Acacia Mearnsii*, poses a major threat to ancient shola-grassland mosaics in the upper reaches of the Nilgiri Biosphere Reserve, Western Ghats.



Inner Mongolia, China

Planting of pine trees in semi-arid grasslands can reduce plant diversity, damages soils, and increase water shortages.





Qinghai-Tibetan Plateau, China

Extensive overgrazing and the increase in rodent population have caused widespread degradation of alpine grasslands, causing soil erosion and loss of biodiversity and ecosystem services.

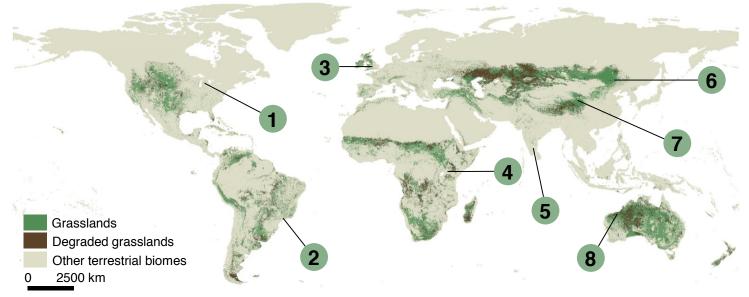




Australia

In many parts of Australia, livestock grazing can quickly break down *Themeda* swards and promote invasion by exotic species.

Global grassland degradation





United States

Reseeding and fertilisation of tallgrass prairie has transformed species-rich prairie grasslands with a mixture of native C3 and C4 grasses, sedges and forbs, to species-poor grassland dominated by Eurasian C3 grasses.



Brazil

Cultivation and abandonment of Cerrado grasslands poses a major threat to these ancient and highly diverse ecosystems.



United Kingdom

Ploughing, reseeding and fertilisation has transformed species-rich chalk grasslands into 'improved' grasslands with higher fertility, lower species richness and lower levels of many cultural ecosystem services.



4 Kenya

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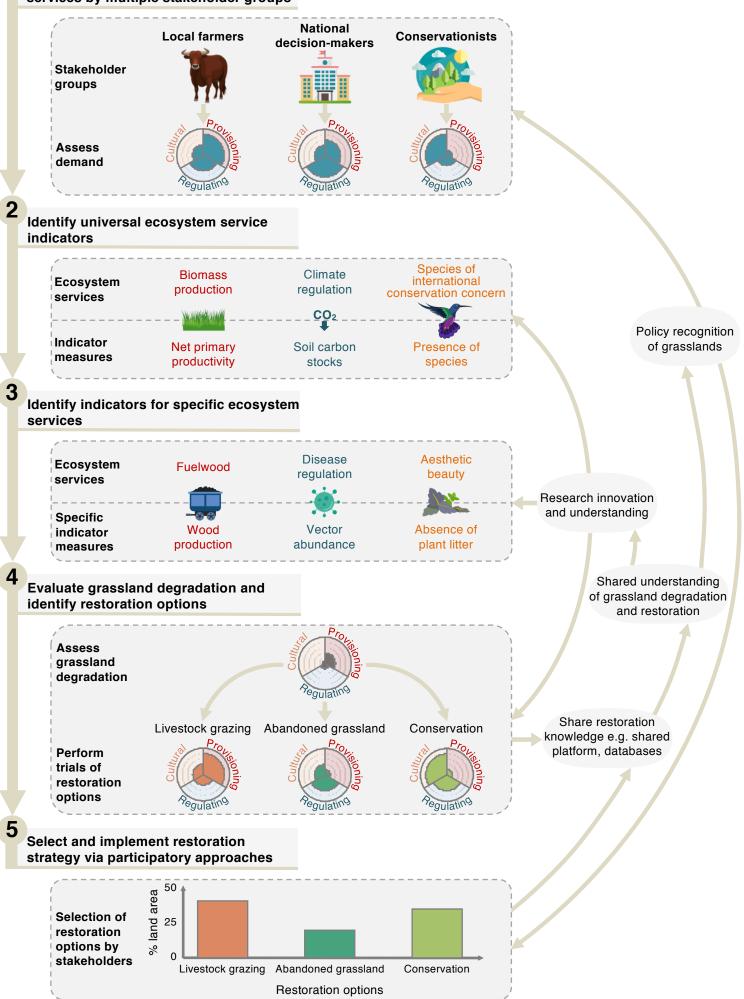
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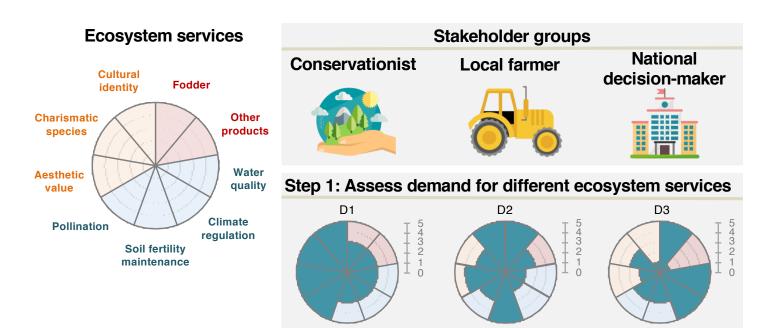


Australia

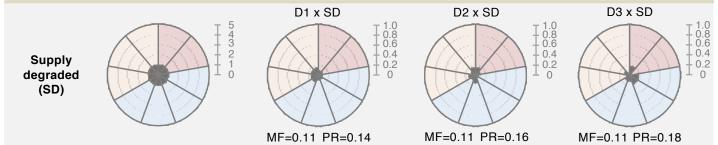
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Assess demand for different ecosystem services by multiple stakeholder groups



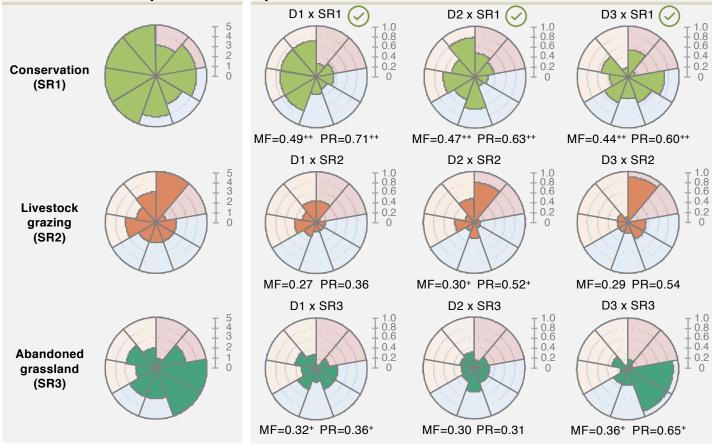


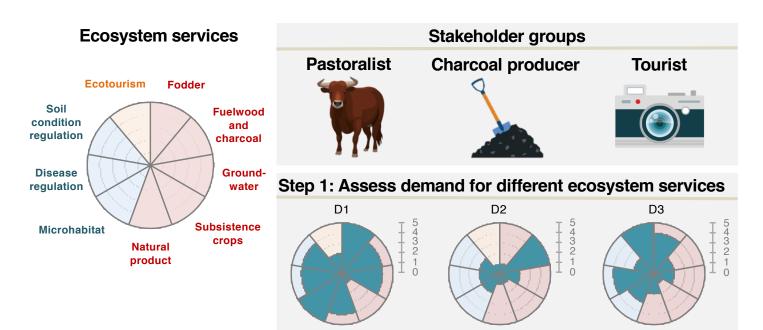




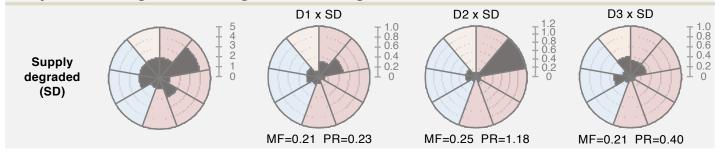
Step 3: Measure supply gain from restoration options

Step 4: Assess stakeholder benefits from restoration options





Step 2: Assess grassland degradation using identified indicators



Step 3: Measure supply gain from restoration options

Step 4: Assess stakeholder benefits from restoration options

