

# Prepared for submission to npj Materials Degradation as an invited paper Weak impact of microorganisms on Ca, Mg-bearing silicate weathering

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Oleg Pokrovsky, Liudmila Shirokova, Svetlana Zabelina, Guntram Jordan, Pascale Bénézeth. Prepared for submission to npj Materials Degradation as an invited paper Weak impact of microorganisms on Ca, Mg-bearing silicate weathering. npj Materials Degradation, 2021, 5 (1), pp.51. 10.1038/s41529-021-00199-w. hal-03872212

# HAL Id: hal-03872212 https://hal.science/hal-03872212v1

Submitted on 25 Nov 2022

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3	Weak impact of microorganisms on Ca, Mg-bearing silicate weathering
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18	
19	Abstract
20	Assessment of the minute is investign minute discolution is social for a mediative or denotes discover the

20 Assessment of the microbial impact on mineral dissolution is crucial for a predictive understanding of basic 21 (Ca, Mg-bearing) silicate weathering and the associated CO<sub>2</sub> consumption, bioerosion and CO<sub>2</sub> storage in basaltic rocks. However, there are controversies about the mechanism of microbial effect, which ranges from 22 inhibiting via nil to accelerating. Here we studied diopside interaction with the heterotrophic bacterium 23 24 Pseudomonas reactants and the soil fungus Chaetomium brasiliense using a combination of mixed-flow and 25 batch reactors and in-situ (AFM) and ex-situ (SEM) microscopy. The results provide new nano-level insights into the degree to which microorganisms modify silicate dissolution. Taking into account negligible effects 26 of organic ligands on diopside dissolution as reported earlier, we conclude that the microbial effect on Ca-27 Mg silicates is weak and the acceleration of dissolution of "basic" silicate rocks in the presence of soil biota 28 29 is solely due to pH decrease in porewaters.

30

#### 31 INTRODUCTION

32 Chemical weathering of basic (Ca- and Mg-bearing) silicates is a primary factor regulating the climate of our

planet via a long-term (million year scale)  $CO_2$  uptake from the atmosphere<sup>1</sup>. Both direct biological

34 (enzymatic action of bacteria, fungi, and plant roots) and indirect physico-chemical (soil porewater pH,

35 extracellular organic ligands and biomass degradation products) mechanisms have been identified for

36 chemical weathering of primary silicate minerals in soil environments<sup>2-4</sup>. Among the "aqueous" factors,

37 controlling the intensity of this process, solution pH, pCO<sub>2</sub> and organic compounds produced during

enzymatic degradation of vegetation litter or the exudates of soil bacteria and plant roots are believed to be

39 the most important<sup>5</sup>. The organic compounds capable to affect mineral dissolution rate include (but are not

40 limited to) extracellular acid and neutral polysaccharides, simple carboxylic acids, uronic acids (galacturonic,

41 guluronic), peptides and aminoacids<sup>6,7</sup>, as well as various lichen polyphenolic compounds<sup>8</sup>.

Positive effects of bacteria and organic ligands on alumosilicate dissolution are fairly well established<sup>9-15</sup>. 42 43 These effects are related to the capacity of biota to facilitate the Al-O-Si and Fe-O-Si rate-limiting bond 44 breaking, due mostly to strong complexation of bacterially-originated organic moieties with rate-controlling 45 aqueous  $Al^{3+}(aq)$  cation<sup>16</sup>, surface >AlOH<sub>2</sub><sup>+</sup> and >FeOH<sub>2</sub><sup>+</sup> sites<sup>17</sup>, or Fe<sup>3+</sup>(aq) mobilization in solution<sup>18-20</sup>. The latter is known to prevent passivation of mineral surfaces by newly-formed Fe(III) hydroxides<sup>21,22</sup>. 46 47 However, all the above mentioned mechanisms are not likely to operate for Fe- and Al-free, Ca and Mg dominant silicates. In fact, several experimental works performed over the past two decades unambiguously 48 49 demonstrated the lack of any sizable impact of CO<sub>2</sub> and organic ligands (at otherwise constant pH) on Ca-50 and Mg-bearing mineral dissolution at typical soil water environments. The reason for this is rather weak complexation of organic ligands with  $>CaOH_2^+$  and  $>MgOH_2^+$  surface sites responsible for the rate-51 controlling Mg-O and Ca-O bond breaking in these minerals<sup>23-30</sup>. According to these studies, unreasonably 52 high concentrations of naturally-relevant organic ligands (0.01 - 0.1 M) are necessary to significantly 53 54 enhance or inhibit mineral dissolution. As a result, the effect of extracellular organic products and in general 55 soil biota on the weathering rate of Mg- and Ca-bearing minerals in natural settings is expected to be weak. Further, work with heterotrophic bacteria and basaltic glass<sup>31</sup>, wollastonite<sup>28</sup> and olivine<sup>32</sup> demonstrated that 56 even unrealistically high concentration of live biomass (several gram per liter, regardless of the presence of 57 58 nutrients) are not capable of substantially accelerating the Ca, Mg or Si release rate from the mineral. As a 59 result, despite the dominant paradigm on the importance of microbes in chemical weathering of Ca-Mgsilicates in soil environments, the impact of biota on dissolution rates of these silicates might be lower than 60 61 generally thought. This, in turn, may require revisiting the current understanding of biotic control on feedbacks in atmospheric CO<sub>2</sub> consumption vs. silicate weathering processes. To further address the role of 62 63 soil biota on mineral weathering process and to extend the previous results to other alkaline-earth bearing minerals (e.g., crystalline mixed calcium and magnesium silicates), in the present study we measured the 64 65 dissolution rate of typical basic silicate diopside ( $CaMgSi_2O_6$ ) in the presence of aquatic and soil microorganisms. For these experiments, we were able to separate the effect of pH from that of bacterial 66 67 presence using a mixed-flow and batch reactor systems coupled with an in-situ and ex-situ microscopic

68 examination of reacted surfaces.

69 To provide a better understanding of the bacteria-mineral interaction in the context of chemical weathering in soils and CO<sub>2</sub> geological storage, we selected a common soil and underground bacterium, *Pseudomonas* 70 reactans. Bacterial strains of the genera Pseudomonas, thanks to their relatively easy culturing, have been 71 frequently used as a model organisms in studying mineral weathering in the past<sup>33,34</sup> and recently<sup>15,28,35</sup>. 72 73 Diopside was chosen as a typical model mineral for basic silicates, frequently used as a representative of 74 silicate dissolution at the Earth surface<sup>36</sup>. Indeed, because hydrolysis of Ca-O and Mg-O bonds is often the 75 rate-limiting step for Ca- and Mg-bearing mineral dissolution<sup>37</sup>, understanding of microorganism interaction 76 with diopside may help to predict the reactivity of other basic silicates in the presence of microbes. In the 77 present study, the experiments were designed in batch and flow-through reactors under constant biomass 78 concentration, in the presence of both live and dead cells in both nutrient-free and nutrient-bearing media. 79 We measured forward far from equilibrium dissolution rates and we microscopically examined reacted 80 surfaces with respect to microorganism attachment, etch pits and corrosion features. The ensemble of 81 obtained results demonstrated rather weak impact of both bacterium and fungus on diopside reactivity, 82 regardless of the presence of nutrients and live status of microorganisms. Moreover, the passivation of mineral surfaces via formation of an exopolysaccharide layer (in case of bacteria) may act as protection 83 mechanism in natural settings. Globally, these new findings challenge the dominant paradigm on dominant 84 biotic control of basic silicates weathering at the Earth surface conditions. This may require revising the 85 86 quantitative modeling of chemical weathering as main factor of biotically driven long term CO<sub>2</sub> consumption 87 (coupled to Ca and Mg release from silicates). Further, the results call for a need of *in-situ* assessment of soil porewater pH at high spatial and temporal resolution as a sole factor controlling the rate of Ca, and Mg 88 89 bearing silicate weathering.

90

### 91 **RESULTS**

### 92 Diopside interaction with *P. reactans*

93 In most experiments performed at pH of 6 to 9, stoichiometric release of Si, Ca and Mg was observed. In bacteria-free experiments, diopside dissolution rates decreased by ca. 0.5-1 order of magnitude 94 95 during the pH decrease from 6 to 9 (Fig. 1, Table 1). The DOC concentration remained fairly stable in the 96 course of given bacterial experiment and generally ranged from 7 to 25 mg L<sup>-1</sup> being the highest at highest 97 biomass concentration. Considering certain scatter in the rates obtained in different experiments, the presence of bacteria in either nutrient-free or nutrient-rich media did not exhibit any significant (at p < 0.05) 98 99 effect on Si release rate from diopside, or even led to a decrease of dissolution rates (Fig. 2). In fact, the effect of bacterial presence did not exceed 0.2 log  $R_{\rm Si}$  units, whereas the reproducibility of rates measured in 100 individual reactors at otherwise similar aqueous solution composition ranged from 0.1 to 0.3 log  $R_{\rm Si}$  units. 101 102 Thus it can be concluded that, within the resolution of our experiments, the impact of bacteria on diopside Si 103 dissolution rate could not be quantified or it was below 0.2 log R unit. Furthermore, there was a net 104 inhibiting effect of bacteria on Ca and Mg release rate from the diopside in nutrient solution at pH = 6.1 (Fig. 2 A). In more alkaline solutions (pH = 8.5 to 9.0), the presence of bacteria (with or without nutrients) did not 105 considerably (> 0.2 log R unit) impact the release rate of Si, Ca and Mg (Fig. 2 B). 106

107 The second remarkable result of BMFR experiments is that live bacteria with (D2-2, D2-3, D4-2, 108 D4-3) and without (D1-2, D1-3, D3-2, D3-3) nutrients exhibited a negligible effect of biomass concentration 109 (between 0.54 and 3.85  $g_{wet}/L$ ) on Mg, Ca and Si release rates because there was no substantial (> 0.2 log R 110 unit) change in rate during biomass increase by a factor of 7 (Table 1).

111 The third important observation is that there was a significant, by a factor of 3 to 5, decrease in Ca, Mg, Si concentrations and release rates between the beginning (exp. D1-1, D2-1, D3-1 and D4-1) and the 112 113 end of bacteria-free experiments (exp. D1-4, D2-4, D3-4 and D4-4). This is illustrated in Supplementary 114 Figure 1, where the Si concentration in the outlet solution corresponding to the sequence of inlet solutions 115 "bacteria-free 0.1 M NaCl  $\rightarrow$  0.54 g<sub>wet</sub>/L bacteria  $\rightarrow$  3.85 g<sub>wet</sub>/L bacteria  $\rightarrow$  bacteria-free 0.1 M NaCl", at a constant flow rate, demonstrated a 3- to 5-fold decrease between the initial and the final bacteria-free 116 experiment. Such a decrease could be related to passivation of mineral surface by bacterial cells and their 117 118 exometabolites as evidenced from in-situ microscopic observations (see section on the AFM observations 119 below).

120

#### 121 Ca, Mg and Si release in the presence of fungus

122 Results of Ca, Mg and Si release from diopside in 0.1 M NaCl and in nutrient media in the presence of live and dead Chaetomium brasiliense demonstrated that concentrations of all mineral constituents 123 124 increase quasi-linearly with time in all types of experiments (illustrated for Si in Supplementary Figure 2). 125 The linear regression coefficients varied from 0.9 to 0.99 for Si and from 0.8 to 0.95 for Ca and Mg. The pH 126 remained generally constant ( $\pm 0.2$  unit) or slightly increased (within 0.2 - 0.3 units) over the 19 days of 127 exposure. In batch experiments with fungus, the DOC increased from 10 to 30 mg/L over the first 100 h of 128 reaction and then remained constant over the following 300 h. Taking into account that 1) all solutions are strongly undersaturated with respect to diopside, all other possible Ca and Mg silicates and amorphous silica 129 130 and 2) there is no effect of aqueous Ca, Mg and Si concentration on diopside forward dissolution rates in neutral solution<sup>24,38,39</sup>, one can quantify the far-from equilibrium dissolution rate (R) for each individual 131 experiment using the following equation which can be applied over the whole duration of the experiment: 132

133 
$$R = (d[\operatorname{Ca}, \operatorname{Mg}, \operatorname{Si}/2]_{\operatorname{tot}} / dt) / s$$

where t (s) designates the elapsed time, [Ca, Mg, Si]<sub>tot</sub> (mol/L) stands for the concentration of calcium,

(1)

135 magnesium or silica released from the solid, and s  $(cm^2/L)$  is the powder B.E.T. surface area. Results are

136 listed in Table 2. The uncertainties on these rates stem from duplicate experiments and range between  $\pm 10$ 

- and 30%. The rates measured in the presence of dead or live fungi generally agree (within  $\pm 0.2 \log \text{ units}$ )
- with those of abiotic and biotic experiments in mixed-flow reactor (Fig. 1). The Si release rates were a factor
- of 1.5 higher in experiments with live fungus (with or without nutrients) compared to experiments with deadfungus (Fig. 3). However, there was no difference in Mg release rate between different treatments, whereas

the Ca release rate, although more variable among replicates, exhibited a maximum in nutrient-free

- experiments with live fungus (Fig. 3). The element release rate in the presence of fungus was mostly
- 143 stoichiometric (Table 2). The disagreement between Si and Mg release rate ranged from 0.1 to 0.05 log R
- 144 unit except in one experiment with dead biomass (0.2 units). Calcium release rate was in agreement with that
- of Si within 0.01-0.19 log R units, except in one experiment with live biomass (0.24 log R lower).
- 146

# 147 AFM observations in the presence of *P. reactans*

148 Freshly reacted diopside surfaces (2-5 h in sterile NaCl solution) largely preserved their cleavage 149 plane morphology (Fig. 4 A). Occasionally, some etch pits and local surface roughness were observable (Fig. 4 B). After 36 h reaction in bacteria-free nutrient solution, rounding of edges occurred (Fig. 4 C). After 1 to 2 150 days of reaction with nutrient solution at pH = 8.5 or 6.5, we have observed a strong colonization of diopside 151 surfaces by a biofilm-like bacterial assemblage (Fig. 4 D). This extracellular polymeric substances (EPS) 152 153 coverage included individual elongated spheres and generally convex forms, less than 0.5 µm in longest axis, although their size distribution was not uniform (Figs. 4 D, E, F). Groups of objects occasionally exhibited a 154 155 distinct internally preferred orientation, which was, however, quite variable within a relatively small area 156 (~50 µm<sup>2</sup>). The obtained images are consistent with earlier reported bacterial EPS via AFM<sup>40,41</sup>. Extensive 157 production of exopolymeric substances precluded imaging of individual bacterial cells within this thick biofilm coverage. In contrast, mineral interaction with bacteria in nutrient-free neutral electrolyte allowed 158 159 recognizing individual cells attached to the diopside surface (Fig. 4 G); however, no site of preferential 160 interactions such as steps or edges could be identified.

161

# 162 SEM imaging of diopside reacted with fungus

163 The diopside grains reacted with dead fungus in 0.1 M NaCl demonstrated sharp borders and no clear dissolution features (Fig. 5 A). After reaction with live fungus in nutrient-free media, there was an 164 appearance of specific dissolution features such as etching and rounding of edges (Fig. 5 B). The rounded 165 166 edges were also observed in experiments with live fungus; these features were distinctly different from cleavage (Fig. 5 C). The diopside crystals were also imaged in the presence of fungal mycelium: the hyphae 167 168 were strongly attached to the grain surfaces and not lost during sample preparation for SEM (Fig. 5 D). Some grains appeared significantly (> 50%) covered by or even embedded into network of fungal hyphae. There 169 were no distinct sites of hyphae location at the diopside surface, and overall, the interaction appeared to be 170 171 quite inert, non-specific and site-indifferent. With increasing the resolution, some organic spherical entities,  $2 \text{ to } 3 \text{ } \mu\text{m}$  in diameter, were identified at the extremities of the hyphae (Fig. 5 D). 172

173

# 174 DISCUSSION

175 Results of the present study demonstrated a weak (or non-measurable) impact of bacterium and
176 fungus on Ca, Mg and Si release rate from diopside. Biologically accelerated weathering of silicates is
177 thought to be controlled by microbial production of extracellular enzymes, chelates, simple and complex
178 organic acids and exopolysaccharides<sup>5,11,42,43,44</sup>. However, there are serious controversies, whether such an
179 effect is largely direct (metabolically-driven nutrient mobilization and specific ligand production) or indirect

(passive pH lowering and cation complexation via EPS excretion), given that it is largely dependent on the 180 nature of mineral and the identity of microorganism<sup>9,45-48</sup>. Furthermore, considering basic silicates, some 181 studies reported nil or inhibiting effect of bacteria on olivine<sup>32,49,50</sup> or wollastonite<sup>28</sup> dissolution, whereas 182 other claimed strong accelerating effect of fungi on olivine<sup>21,22,51</sup>. However, in the latter works, all the effects 183 of microorganisms were limited to their ability to sequester rate-inhibiting Fe(III) thus preventing the 184 185 slowdown of dissolution, rather than directly promoting the breaking of structural bonds in the mineral lattice. Overall, in contrast to large number of works on microbial impact on alumosilicate dissolution<sup>52-56</sup>, 186 187 the interaction of aquatic microorganisms and relevant organic ligands with Ca or Mg-bearing silicates 188 remained insufficiently characterized.

In the present study, the global rate uncertainty in the studied pH range (6 to 9) was between 20 and 189 190 50%. This uncertainty stems from both the pH variation ( $\pm 0.2$ -0.3 units) during the course of an experiment 191 and the reproducibility of steady state concentrations of diopside constituents in mixed-flow reactors with 192 bacteria-free systems or among duplicates in batch reactors with dead fungus ( $\pm 20\%$ ). Therefore, we 193 conclude that the difference between abiotic (bacteria-free or dead cells) and live bacteria experiments is 194 within the experimental uncertainty. As such, the effect of selected live microorganisms on Ca, Mg and Si 195 release rate from diopside is not significant, or cannot be resolved using the experimental design employed in 196 the present study.

197 In this regard, the present result corroborates the former works of our group which demonstrated weak inhibiting or negligible impact of cultivable heterotrophic bacteria P. aureofaciens on wollastonite<sup>28</sup> 198 and *P. reactans* on basaltic glass<sup>31</sup> and olivine<sup>32</sup>, in agreement with other studies<sup>15,47,49</sup>. The broad meaning of 199 these results is that microbial impact on basic silicate mineral dissolution is limited by *i*) pH decrease, thus 200 201 accelerating proton-promoted dissolution, which is relevant to all basic silicates; *ii*) release of Fe(III)-202 complexing organic ligands including siderophores. The latter is pertinent solely to minerals whose dissolution is limited by Fe(OH)<sub>3</sub> passivation of the surface. In all other cases of microbially-affected basic 203 204 silicate dissolution, the concentrations of excreted organic ligands are not sufficient to sizably complex Mg 205 and Ca at the surface or in the aqueous solution which is necessary for breaking of rate-controlling Mg-O or 206 Ca-O bonds at the mineral surface. For example, gluconic acid and its derivatives are believed to be the main exometabolite of heterotrophic bacteria that use glucose as a substrate<sup>10,34</sup>. Gluconic acid can favor olivine 207 dissolution in neutral waters<sup>32</sup>; however, this effect occurs at concentrations > 1 mM, totally irrelevant to 208 natural environments. Another example comes from a concerted study of ~40 various ligand impact on 209 210 wollastonite dissolution<sup>28</sup>. These authors calculated, based on the results in mixed-flow reactors, the 211 threshold concentration of ligands necessary to triple the rates of wollastonite at pH around 7. For most 212 ligands, very high concentration (2 to 3 orders of magnitude higher than those encountered in natural 213 settings) were necessary to appreciably affect the rates. Finally, Golubev and Pokrovsky  $(2006)^{24}$ demonstrated that very high concentrations (0.01-0.1 M) of organic ligands originated from enzymatic 214 215 degradation of organic matter or bacterial metabolic activity, are necessary to appreciably affect diopside 216 dissolution. Considering these results and by analogy with olivine and wollastonite, we hypothesize that 217 possible impacts of studied heterotrophic bacteria on the diopside dissolution via gluconic acid or other soluble exometabolite products are not detectable due to insufficient concentration of these compounds 218 219 generated via metabolism of *P. reactans* in experimental reactors.

220

Below we discuss the mechanisms of diopside interaction with microbial cells and natural implications.
Reactive surfaces in both bacterial and fungal experiments demonstrated lack of specific sites of dissolution,
etch pit formation, surface amorphisation or new mineral occurrences. Rather, our in-situ (AFM) and ex-situ
(SEM) microscopic observations suggested "passivation" of reacted surfaces by microbial metabolic

products. We therefore hypothesize that the biofilm coverage either limited aqueous solution contact with

226 mineral surface or merely did not produce sufficient amount of organic ligands capable to impact the

diopside dissolution.

Microbial biofilm formation, dominated by exopolysaccharides (EPS), is known to protect the 228 mineral surface from interaction with bulk fluid 42,46,47,57 thus inhibiting overall dissolution rate 2,3,6,7,35,58. In the 229 230 present experiments with diopside, the biofilm formation was also evidenced from indirect assessment of 231 dissolution rate: the rates decreased by a factor of 2 to 3 after ~500 h treatment with bacteria; this decrease persisted even after changing back the inlet solution to sterile bacteria-free background electrolyte (see 232 233 Supplementary Figure 1). Noteworthy that such an irreversible rate decrease was also observed after reaction 234 with live bacteria in nutrient-free media and suggested substantial and long-lasting interaction of live P. 235 *reactans* with diopside surface in both neutral (pH = 6) and slightly alkaline (pH = 9) solutions. At the same 236 time, no decrease in diopside dissolution rate was observed over 400-800 h of experiments in the abiotic inorganic systems<sup>23,24</sup>. 237

238 As a result of lack of measurable rate increase in the presence of either bacteria or fungus in this 239 range of pH, nutrient and biomass concentration, we suggest, in agreement with previous studies, that, regardless of nutrient regime, biomass concentration and vital status of microorganism, solely the solution 240 241 pH can act as the main governing factor of mineral dissolution in the presence of aquatic microorganisms. A 242 similar conclusion has been recently achieved for olivine dissolution in the presence of complex natural microbial communities<sup>59</sup>. Therefore, implementation of mineral-microorganism interaction in chemical 243 244 weathering codes requires solely the information on porewater solution chemistry that can be directly 245 assessed from macroscopic measurements of interstitial soil solutions.

Moreover, extremely high concentrations of organic ligands, necessary to affect the rates of basic 246 silicate dissolution in laboratory experiments<sup>24,28,32</sup>, together with high bacterial and fungi concentrations, 247 248 comparable with those used in the present study, can be encountered exclusively in surface (litter) soil 249 horizons, and unlikely to ever occur in deep (mineral) soil horizons. Exceptions are specific 250 microenvironments around fungal hyphae, or near decaying organic matter. For example, a strong acidification was reported with hyphae bound to the biotite surface<sup>60</sup>. To quantify these effects on a broader 251 scale, high-resolution in-situ quantitative analysis of chemical composition of solution, bacteria and organic 252 253 ligand concentration in soil pore water microenvironments are needed.

254 At the same time, the role of microorganisms under natural environment, especially in the non-255 saturated zone, might be more important than that in aqueous environment of mixed-flow experiments. It is generally understood that a lack of "wetted" mineral surface area, or mineral surfaces being exposed to the 256 reactive fluid, could be a partial explanation for slower reaction rates observed in the field compared to those 257 measured in the laboratory with unlimited water<sup>61,62</sup>. For example, via exerting hydrated exopolysaccharides, 258 microbial cells are capable of prolonging the degree of mineral surfaces hydration during natural 259 wetting/drying cycles. Therefore, although our experiments, performed under nutrient-free conditions with 260 261 starved cells, are most similar to natural environments under short-term scale, there is sizable uncertainty in 262 extrapolating these short-term laboratory results to long-term field conditions. The latter are known to exhibit one to three orders of magnitude lower weathering rates, due to different time scale of observation, diversity 263 264 of microorganisms, and nutritional conditions.

265 For more general natural implications of the obtained results, we consider below all Ca, Mg-bearing 'basic' silicates including Ca alumosilicates. Several lines of evidences allow us to suggest rather weak 266 267 effect of dissolved organics on (Ca, Mg, Al)-bearing silicate dissolution in natural settings. First, the concentration range of various carboxylic ligands (oxalic, citric, malonic, succinic, lactic, formic, acetic), 268 detected in soil solutions, spans from 10<sup>-6</sup> to 10<sup>-4</sup> M<sup>63</sup>. Laboratory experiments suggest that these 269 concentrations are still not sufficient to appreciably affect feldspars dissolution rates. Second, the effect of 270 organic ligands on Al release and thus on overall dissolution rates of feldspars and basaltic glass is absent or 271 becomes negligible at  $pH \ge 7^{16,64}$ . The pH of rivers and streams draining 'basic' silicate rocks is between 7 272 and  $8^{65,66}$  whereas the pH of soil solutions in organic rich soils typically ranges from 4 to  $7^{67}$ . However, 273 274 surficial acid- and organic-rich soil horizons rarely contain primary silicates, whereas in deep (C) soil horizons, the interstitial soil solution pH is always above 7 (see ref. 65). At these conditions, the effect of 275 organic ligands on Ca, Mg, Al and Si release rate from primary silicates should be negligible which 276

277 corroborates the conclusion of Drever  $(1994)^{68}$  that vegetation should cause an increase in weathering rate 278 through the pH effect only where the pH is below 4-5. As a results, the sole effect of soil organics on both 279 acidic and felsic rock weathering seems to be in decreasing the aqueous activity of Al<sup>3+</sup> thus preventing

280 formation of secondary minerals without significant modification of Ca and Mg ions release.

281 It has been widely argued that the presence of soluble minerals in trace amounts in the rocks (i.e.,

carbonates<sup>69</sup>; phosphates<sup>70</sup> may strongly affect the intensity of chemical weathering and export fluxes of

283 major ions. The effect of biological activity in this case is expected to be solely due to pH decrease in the

- vicinity of microorganisms or plant roots. Indeed, the data for dolomite<sup>25</sup>, calcite and magnesite<sup>27,29</sup>,
- brucite<sup>26</sup>, and calcite<sup>30</sup> dissolution in the presence of various organics demonstrate a negligible effect of
   organic ligands on Ca and Mg release from carbonates and oxides at environmentally-relevant conditions.
- The effect of soil and atmospheric  $pCO_2$  is always considered to be an important factor accelerating rock weathering at the Earth surface. In this regard, we would like to underline that this effect stems solely from the decrease of soil solution pH since there is no chemical effect of  $CO_2(aq)$  or dissolved H<sub>2</sub>CO<sub>3</sub> molecules on the main rock-forming silicates dissolution rate<sup>39</sup>, and in neutral and alkaline solutions the pCO<sub>2</sub> effect becomes even inhibiting<sup>71</sup>.

292 Overall, this work allows a better understanding of microbially-affected basic silicate dissolution and 293 provides microscopic insights for quantifying the degree to which the heterotrophic bacteria and fungi are 294 capable of modifying the dissolution rate of rock-forming silicates. Using bacterial-mixed flow and batch 295 reactors combined with in-situ atomic force microscopy of bacteria-diopside surfaces and experiments on 296 diopside dissolution in the presence of fungus combined with microscopic examination of reacted surfaces, 297 we demonstrated a rather weak impact of microorganisms on Ca, Mg and Si release rate. Regardless of the 298 presence or absence of nutrients, in a wide range of pH (from 6 to 9) and despite very high biomass 299 concentration (up to 4-5 gwet/L), the rates were not considerably affected by microbial presence. Moreover, there was a detectable (ca. by a factor of 3) decrease in reactivity of diopside after its prolonged contact with 300 301 live heterotrophic bacteria. This 'passivation' of the mineral surface was evidenced by extensive production 302 of exopolymeric substances yielding a biofilm-like coverage already after 1-2 days of interaction in nutrient-303 rich media. Despite tight contact between fungus and diopside surfaces, no site of preferential interactions 304 such as steps, kinks or edges, etch pits or surface amorphisation was detected. The obtained results are fully consistent with large body of evidences that any possible microbially produced organic ligands at 305 306 concentrations relevant to live biotic systems or natural settings are not capable of noticeably impacting the 307 reactivity of Mg-O and Ca-O sites on diopside surface. As such, this work challenges the dominant paradigm of an overwhelming role of biota on silicate mineral weathering in soils. At the same time, it is in line with a 308 309 controversial opinion of Sverdrup  $(2009)^{72}$  that physical and chemical conditions of forest soils do not allow 310 any significant direct surface actions on minerals by microorganisms or tree roots. While we do not contest 311 the importance of microbial organic matter in Fe and Al complexation and mobilization from silicate structure, the role of biota on basic (Ca and Mg) silicates, playing a pivotal role in atmospheric CO<sub>2</sub> 312 313 sequestration via chemical weathering, may be much lower than traditionally thought. In fact, only drastic 314 pH decrease (i.e., over several pH units), probably occurring in the microniches adjacent to plant roots, is capable of sizably (> a factor of 2 to 3) increase mineral dissolution rate. It is not excluded however that this 315 promoting effect will be partially attenuated due to biofilm formation and passivation of silicate surfaces by 316 317 microbial cells. Detecting such 'active' or 'passive' microniches within the complex plant-microbe-mineral 318 systems and more importantly, characterizing chemical environment of the fluids with spatial resolution on the level of µm to mm may turn out to be a bottleneck for quantitative mechanistic modeling of chemical 319 320 weathering. Currently, these gradients are not taken into account even in most sophisticated modeling 321 approaches<sup>73,74</sup>. Further experimental and field studies are thus needed to answer the question whether a quantitative modelling of basic silicate mineral weathering and related CO<sub>2</sub> consumption require an explicit 322 323 prediction for microbial activity or it can be merely approximated by the pH of bulk soil porewaters.

324

#### 325 METHODS

#### 326 Mineral, bacterial and fungal culture

327 Natural diopside (Ca<sub>0.99</sub>Mg<sub>0.98</sub>Fe<sub>0.02</sub>Cr<sub>0.01</sub>Si<sub>2</sub>O<sub>6</sub>) crystals characterized in previous works<sup>24,39</sup> were 328 used in this study. Ultrasonically-cleaned 100 to 200  $\mu$ m size fraction having a specific surface area of 1045 329  $\pm$  50 cm<sup>2</sup>/g was selected for dissolution rate measurements.

330 Pseudomonas reactans is a common rod-shape groundwater and soil bacteria averaging 2 µm in size 331 (see further description in ref. 31, 32). For experiments, cells were first cultured to a stationary stage on a shaker at 25°C in the darkness using 10% nutrient broth with the following composition: 0.1 g/L glucose, 1.5 332 g/L peptone, 0.6 g/L NaCl and 0.3 g/L veast extract. Cells were allowed starving for 2 days in nutrient-free 333 media, prior the macroscopic dissolution and microscopic (AFM) experiments. For this, freshly-grown cell 334 335 suspension (at the stationary stage) was rinsed by sterile  $0.6 \text{ g L}^{-1}$  NaCl solution via centrifugation, and then placed in 0.6 g L<sup>-1</sup> NaCl without nutrients for 2 days. Before the experiment, this bacterial suspension was 336 rinsed again with sterile 0.6 g L<sup>-1</sup> NaCl to remove, as possible, all cellular exometabolites. 337

Active bacteria number counts (colony forming-units, CFU/mL) were performed using Petri dish 338 339 inoculation on a nutrient agar (0.1, 0.2, and 0.5 mL of sampled solution in three replicates) in a laminar hood 340 box. Inoculation of blanks was routinely performed to assure the absence of external contamination. The biomass of live bacteria suspensions was also quantified by measuring wet (after it was centrifuged 15 min at 341 342 10,000 rpm) and freeze-dried weights in duplicates. Before the experiments, cells were rinsed twice in either the appropriate fresh culture media or a sterile 0.1 M NaCl solution using centrifugation with ~ 500 mL of 343 344 solution for 1 g of wet biomass, to remove adsorbed metals and cell exudates from the surface. The conversion ratio wet/freeze-dried weight of *P. reactans* was equal to 8.4±0.5. The conversion factor of 345 optical density (600 nm, 10 mm path) and wet biomass ( $g_{wet}/L$ ) to the cell number (CFU/mL) was equal to 346  $(5\pm1)\times10^8$  and  $(8.0\pm1.5)\times10^7$ , respectively, as determined by triplicate measurements. Typical live biomass 347 concentration during experiments was 0.54 and 3.85 gwet/L. 348

349 Experiments on fungus were performed with *Chaetomium brasilience*, belonging to the largest genus 350 of saprophytic ascomycetes, inhabiting soils worldwide. The environmental habitats of genus Chaetomium 351 include deserts, salterns, agricultural, mangrove, living plants and animals, air, decayed wood and stones<sup>75</sup>. 352 Chaetomium brasilience Bat. Et Pontual 1948, stumm F-3649 was obtained from All-Russian Collection of 353 Microorganisms - VKM (http://www.vkm.ru/Collections.htm). Fungal biomass was produced on liquid 354 Czapek medium (3 g/L NaNO<sub>3</sub>, 1 g/L K<sub>2</sub>HPO<sub>4</sub>, 0.5 g/L KCl; 0.5 g/L MgSO<sub>4</sub>×7H<sub>2</sub>O, 0.01 g/L FeSO<sub>4</sub>×7H<sub>2</sub>O, 355 30 g/L saccharose at pH = 6.0). The biomass was centrifuged at 5000 rpm during 15 min, rinsed 4 times in 356 0.1 M NaCl and allowed starving during 2 days in 0.1 M NaCl prior the experiment. Dead (heat-killed) cells were produced via autoclaving 1 h at 132°C the freshly grown fungal biomass followed by thorough rinsing 357 358 in sterile 0.1 M NaCl. Cell concentration in the sample was analyzed by measuring optical density (OD) at 359 580 nm using a spectrophotometer, with the culture medium used as a reference. Diopside dissolution 360 experiments were performed in duplicates, in 0.1 M NaCl with dead biomass, and live biomass with and 361 without 1% nutrient solution.

362

### 363 Bacterial mixed-flow reactor

The Bacterial Mixed-Flow Reactor (BMFR) system used in this work was similar to the one used for basaltic glass dissolution in the presence of bacteria<sup>31</sup>. The Teflon-made 25 mL reactor with a suspended Teflon-coated stirbar, immersed in a water bath at  $25 \pm 0.5$  °C, was fitted with 20 µm poresize Magna Millipore Nylon outlet filters (45 mm diameter) to allow bacteria to pass while retaining the 100-200 µm mineral powder in the reactor. This system therefore maintained a constant biomass concentration in the reactor, equal to that of the inlet fluid. The input fluids were kept in 1L polypropylene bottles closed with Biosilico® ventilated caps. These fluids were stirred continuously during the experiments and were changed

371 typically each 7 days. This allowed maintenance of a constant and stable stationary phase bacterial culture in

the bacteria-bearing inlet fluid. Bacteria concentration was verified by periodical sampling of inlet and outlet

fluids for cell optical density (total biomass) and live cell numbers (via agar plate counting). Prior to each

- experiment, all reactor system parts, including tubing, were sterilized at 130°C for 30 min and rinsed with
   sterile MilliQ water.
- 376 Steady-state dissolution rates ( $R_i$ , mol/cm<sup>2</sup>/s) were computed from measured solution composition 377 using:
- 378

$$R_i = -q \cdot 1/n \cdot \Delta[i(aq)]_{tot} / s$$
<sup>(2)</sup>

where q (L/s) designates the fluid flow rate,  $\Delta[i(aq)]_{tot}$  (mol/L) stands for the difference between the input and output solution concentration of *i*-th element, *n* is the stoichiometric coefficient of *i*-th element in diopside formula equaled to 2, 1, and 1 for Si, Mg and Ca, respectively, and *s* (cm<sup>2</sup>) refers to the total mineral surface area present in the reactor. The surface area used to calculate the rates was that measured on the fresh (unreacted) diopside powder. Note that during 600 h of reaction time, no systematic variation of the surface area for powder reacted at 5 < pH < 10 in bacteria-free solutions was observed<sup>24</sup>.

385 Experiments were performed in series at constant flow rate between 0.018 and 0.022 mL/min. This corresponded to a mechanical steady-state achievement (x 5 reactor volume renewal) after 4 to 5 days. At the 386 beginning of each series, a quantity of fresh diopside powder was added to the reactor system. For each 387 388 experimental series, a sterile bacteria-free 0.1 M NaCl with relevant buffer (10 mM NaMES (2-389 morpholinoethanesulfonic acid) or NaHCO<sub>3</sub>) was first prepared and injected into the reactor until a steadystate outlet fluid composition was attained (typically 2 weeks, D1 series). Once steady state was attained, live 390 391 bacteria (0.54  $g_{wet}/L$ ) with or without nutrients were added to this inlet solution (D2 series). The next series comprised higher concentration of live bacteria in the inlet fluid (3.85 gwet/L). The 2nd and 3rd series lasted 392 393 7-10 days each. After that, the inlet solution was replaced by fresh sterile bacteria-free fluid; this experiment, 394 which lasted > 3 weeks, was destined to assess the recovery of the mineral reactivity and to test the 395 consequences of passivation of mineral surface by a bacterial biofilm.

The inlet solutions were composed of 0.1 M NaCl and 0.01 M NaMES (D1 and D2 series) and 0.1 M NaCl with 0.01 M NaHCO<sub>3</sub> (D3 and D4 series). The 1:10 diluted Aldrich nutrient broth (0.1 g/L glucose, 1.5 g/L peptone, 0.6 g/L NaCl and 0.3 g/L yeast extract) was added to selected live bacteria experiments. These were compared to live bacteria experiments performed without nutrients. Overall, the experiments conducted in this study can be divided into three main categories: 1) bacteria- and nutrient-free, 2) live bacteria in nutrient-free media, and 3) live bacteria in 10% growth media.

All outlet sample fluids were additionally filtered through 0.45 μm Millipore acetate cellulose filters
 and acidified with ultrapure 2% HNO<sub>3</sub> before chemical analysis. The inlet fluids were routinely analyzed at
 the beginning, in the middle and at the end of each experiment and were filtered and processed using exactly
 the same protocol as the outlet fluids.

406 Mechanical steady-state was achieved after 24 h of reaction. For each steady-state condition, 407 minimum 4 independent measurements of Mg, Ca and Si concentrations, bacterial optical density, pH and 408 flow rate were performed (reproducibility of  $\pm$  10%) and used for calculating the dissolution rate. The cell 409 count (CFU) on nutrient agar plates was performed at the beginning, in the middle and at the end of each 410 experimental run and did not vary significantly throughout all experiments.

411

# 412 Batch experiments on diopside dissolution in the presence of fungus

The aim of these experiments was to compare, under identical solution conditions, the Ca, Mg and Si release rate from diopside placed in contact with dead and live fungus cultures. The batch reactors used to study diopside–fungus interaction consisted of sterile 250 mL polypropylene culture flasks with vented caps (Biosilico). All manipulations were conducted in a sterile laminar hood box (class 100). For experiments, 1.3 g of sterilized 100-200 µm diopside were placed in 250 mL sterile solution. We conducted six experiments in duplicates, four involving live cultures, with and without 1% nutrient solution, and two with dead

419 (autoclaved) biomass (in nutrient-free 0.1 M NaCl).

420 The reactors were inoculated with 4.2-5.6 gwet/L of fresh Chaetomium brasiliense culture or the 421 equivalent amount of dead biomass and placed on rotative shaker (120 rpm) at  $25 \pm 1^{\circ}$ C for 19 days. Sterile controls were routinely run both for nutrient media and 0.1 M NaCl solutions and did not demonstrate any 422 423 fungal contamination. Periodically, 10 mL aliquots of homogeneous mineral suspension + fungus were collected using sterile serological pipettes and transferred in polystyrene vials for Ca, Mg, Si, pH and cell 424 425 biomass measurements. The mineral/fluid ratio remained constant during experiments and the concentration 426 of fungus was not affected by the sampling. Calcium, Mg and Si concentrations were analyzed in 0.45 µm

427 filtrates and pH and cell biomass were measured in unfiltered samples.

# 428

#### 429 Analyses

430 All input and output solutions were analyzed for Mg, Ca, Si, optical density and pH as a function of 431 time. NIST buffers ( $pH = 4.008, 6.865, and 9.180 at 25^{\circ}C$ ) were used for calibration of a combination pHelectrode (Schott Geräte H62). Magnesium and Ca concentrations were measured by flame atomic 432 absorption spectroscopy with an uncertainty of 1 and 2 % and a detection limit of 0.1 and 0.2  $\mu$ M, 433 434 respectively. Total silica concentration was determined by a Technicon automatic analyzer using the 435 molybdate blue method with an uncertainty of 2% and a detection limit of 0.3  $\mu$ M. The concentration of soluble organic exometabolites of bacteria and fungi was measured as total dissolved organic carbon (DOC) 436 using a TOC Shimadzu VSCN analyzer with an uncertainty of 5% and a detection limit of 0.1 mg  $C_{org}$  L<sup>-1</sup>. 437 All experiments were performed at far from equilibrium conditions with respect to diopside as calculated via

- 438
- vMinteq program<sup>76</sup> with relevant solubility parameters<sup>77</sup>. 439

#### **Microscopic observations** 440

441 An AFM study using a biological instrument (JPK NanoWizard) allowed in-situ high-resolution 442 characterization of the status of diopside surfaces in the presence of live bacteria with and without nutrients. Diopside cleavage planes were obtained by cleaving optically transparent crystals 5-10 mm in size in a 443 444 laminar hood box (A100). Fresh cleavage surfaces were sterilized by alcohol and fixed within petri dishes for 445 imaging with the AFM. For long-term (> 5 h) exposure, diopside cleavages were reacted under sterile 446 conditions with 1 gwet/L of bacterial biomass in separate petri dishes. The petri dishes were mounted into the 447 AFM for in-situ inspection of the diopside surfaces directly immersed within the bacteria containing medium, removed from the microscope for further storage, and re-investigated with the microscope 448 whenever appropriate. Depending on the specific aim of the scanning process (attempt of potentially imaging 449 450 all adhering matter on the surface vs. attempt of removing matter adhering to the surface), images were 451 acquired with both intermittent-contact mode and contact mode using uncoated Si-cantilevers with integrated 452 tips.

453 The diopside grains reacted with fungus were examined by Scanning Electron Microscopy (SEM). Prior to characterization by SEM, samples were centrifuged (3000 g, 30 min) and the organic matter was 454 removed from the solid phase by rinsing them in nutrient-free sterile 0.1 M NaCl solution. The remaining 455 solids were rinsed three times with MilliQ water and freeze-dried at -60 °C. They were observed after carbon 456 457 film coated on the sample surface with a JEOL JSM 6360LV SEM coupled with a SDD PGT Sahara EDS 458 analyzer operating at 30 kV.

459

#### DATA AVAILABILITY 460

Supplementary material is available. All measured dissolution rates are presented in Tables 1 and 2. Other 461 462 relevant data are also available from the corresponding author of this paper upon a reasonable request.

#### 463 **ACKNOWLEDGEMENTS:**

464 Support from grants BIOCASTRO of MITI (CNRS) and CaBioCa of Défi (Action conjointe, MITI CNRS) 465 are acknowledged.

#### 466 AUTHOR CONTRIBUTIONS

467 O.S.P. developed the concept of the study. L.S.S., S.A.Z. and O.S.P. performed dissolution experiments. G.J.
 468 performed AFM study. P.B. participated in laboratory experiments and SEM study. O.S.P. wrote the paper,

469 with input from L.S.S., P.B., G.J. and S.A.Z. All the co-authors discussed the final version of the paper.

470

#### 471 Competing Interests

- 472 The authors declare that they have no known competing financial interests or personal relationships that
- 473 could have appeared to influence the work reported in this paper.
- 474

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- 650

# 651 Figure Legends

- 652
- **Figure 1.** Calcium (a), Mg (b) and Si (c) release rates from diopside measured in the mixed-flow
- reactor with and without bacteria and 10% nutrient or nutrient-free solution.

655

- **Figure 2.** Synthesis of obtained steady-state Ca, Mg and Si release rates (average  $\pm$  s.d.) from
- diopside at pH 6 (a) and 8.5-9.0 (b) with and without *Pseudomonas reactans* and in the presence or
- not of nutrients. The error bars represent average of 2 to 4 duplicates, each comprising 3 to 4
- 659 independent sampling at the steady state.

660

661	<b>Figure 3.</b> Synthesis of obtained Ca, Mg and Si release rates (average $\pm$ s.d., mol cm <sup>-2</sup> s <sup>-1</sup> ) from
662	diopside at pH 6.0-6.5 with and without Chaetomium brasiliense fungus (4.5-5.6gwet/L) and in the
663	presence or not of nutrients (1% Czapek medium). The error bars represent average of 2 reactors,
664	each comprising 9 to 10 independent sampling as a function of time over 19 days of exposure.
665	Figure 4. A-C: In-situ AFM images of diopside surfaces immersed within sterile electrolyte media
666	for up to 5 hours (A: 70 x 70 $\mu$ m <sup>2</sup> , deflection mode; B: 20 x 20 $\mu$ m <sup>2</sup> , deflection mode; C: 12 x 12
667	$\mu$ m <sup>2</sup> , height mode - with the color scale spanning a total of 238 nm). <b>D-G</b> : In-situ AFM images of
668	diopside surfaces immersed in nutrient media (D: $8.6 \times 8.6 \mu m^2$ , deflection mode; E: $10 \times 10 \mu m^2$ ,
669	deflection mode; <b>F</b> : 5.5 x 5.5 $\mu$ m <sup>2</sup> , deflection mode; <b>G</b> : 1.5 x 1.5 $\mu$ m <sup>2</sup> , deflection mode.
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671	Figure 5. SEM images of reacted diopside surfaces in the sterile electrolyte (a), live fungus in 0.1
672	M Na Cl (b, d) and in 1% nutrient media (c, d).
673	
674	<b>Table 1.</b> Results of Bacterial MFR experiment. $R$ units are mol cm <sup>-2</sup> s <sup>-1</sup> .
675	

**Table 2.** Results of fungus batch experiments (0.1 M NaCl). R units are mol cm<sup>-2</sup> s<sup>-1</sup>.