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- Spatial heterogeneity of interaction strength has contrasting
- effects on synchrony and stability in trophic metacommunities

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5 Abstract

Spatial heterogeneity is a fundamental feature of ecosystems, and ecologists have identified it as a factor promoting the stability of population dynamics. In particular, the asymmetry of interaction strength and resource supply between patches generates an asymmetry of biomass turnover with a fast and a slow patch. The coupling of these two energy channels by mobile predators has been identified to increase stability at different scales by promoting the asynchrony of population dynamics between each patch. Here, we demonstrate that asymmetry has a contrasting effect on the stability of asymmetric metacommunities receiving localised perturbations. We built a model of an asymmetric metacommunity with two patches linked by the dispersal of predators and in which prey receive stochastic perturbations only in one patch. Perturbing prey in the fast patch synchronises the dynamics of prey biomass between the two patches and destabilises predator dynamics by increasing their temporal variability. Conversely, perturbing prey in the slow patch desynchronises their dynamics and stabilises predator dynamics. This discrepancy between the responses is due to asymmetric transmission of perturbations caused by the asymmetric distribution of biomass between the fast and the slow patch. Consequently, the fast patch drives the dynamics of the metacommunity and imposes synchrony while the slow patch does not. Therefore, local perturbations can have opposite consequences at the regional scale depending on the characteristics of the perturbed patch. Our results have strong implications for conservation ecology and suggest reinforcing protection policies in fast patches to dampen the effects of perturbations and promote the stability of population dynamics at the regional scale.

Key words

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source-sink, stochastic perturbations, food chain, dispersal, asymmetry, conservation

■ Introduction

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tion networks, ecologists have been relentlessly looking for the mechanisms ensuring ecosystem stability. Spatial heterogeneity has long been identified as one of the main factors promoting the mechanisms 31 underlying the maintenance of biodiversity and the stability of ecosystems. For instance, spatial heterogeneity provides local favourable conditions to each species of the regional pool (Holt, 1984; Chesson, 2000; Amarasekare and Nisbet, 2001), which in turn ensures species persistence in less favourable patches by source-sink dynamics (Mouquet and Loreau, 2002, 2003; Loreau et al., 2003). The stability of the 35 temporal dynamics of species biomass is ensured by the asynchrony of the dynamics between patches, which leads to compensatory dynamics (Loreau et al., 2003; Loreau and de Mazancourt, 2008). Therefore, spatial heterogeneity has been accepted as an element of the common wisdom on stability but the models cited above relied on neutral or competitive communities and did not account for the feedbacks induced by trophic interactions. Inspired by the description of fast and slow energy channels by soil ecologists (i.e., in terms of biomass turnover), Rooney et al. (2006) noted the stabilising effect of the asymmetry of energy flows in ecosystems with a food web model consisting of one mobile predator feeding on two energy channels. In their model, the asymmetry of energy flow is generated by the asymmetry of interaction strength between predators and prey and the asymmetry of the consumption of a common resource by the two basal species, which in turn promotes the asynchrony of prey biomass dynamics in response to perturbations. However, Rooney et al. (2006) and subsequent studies (Goldwyn and Hastings, 2009; Ruokolainen et al., 2011) did not identify the mechanisms underlying this asynchrony and the dynamics of asymmetric food webs in general. Recent theoretical studies were able to accurately explain the synchrony patterns in metacommunities (Quévreux et al., 2021a,b), and we propose to consider the effects of asymmetry from the metacommunity 50 perspective to fill this gap.

Since May (1972) demonstrated that stability was not an inherent property of ecological interac-

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Metacommunities embody the spatial dimension of interaction networks: they consist of distant patches

connected by the dispersal of the organisms living in each patch (Leibold et al., 2004; Leibold and Chase,

observed in ecosystems because each community has its own characteristics such as biomass turnover. Following Rooney et al.'s (2006) model, many studies varied the asymmetry of interaction strength and/or resource supply to manipulate the difference in biomass turnover between the energy channels hosted by each patch (Goldwyn and Hastings, 2009; Ruokolainen et al., 2011; Anderson and Fahimipour, 2021). More generally, interaction strength is key in community dynamics because it governs food web structure, stability (Neutel et al., 2002) and biomass distribution (Barbier and Loreau, 2019) by simultaneously determining predator growth and prey mortality. Therefore, its significant variations observed in nature, often reported as predation risk by prey in field studies (Table 1), should lead to dramatic variations in community functioning across space.

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In addition to the asymmetry of interaction strength, Rooney et al. (2006) highlighted the importance of mobile predators coupling two different energy channels, a keystone role in ecosystem functioning largely reported by empirical studies (Schindler and Scheuerell, 2002; Vadeboncoeur et al., 2005; Olff et al., 2009; Dolson et al., 2009). Rooney et al. (2006) explained the increase in food web stability with 69 asymmetry by the asynchronous response of prey to the perturbation of the predator. Indeed, small spatial heterogeneity can lead to strong asynchrony in oscillating predator-prey systems (Goldwyn and Hastings, 2009), but Ruokolainen et al. (2011) identified a humped-shaped relationship in which only moderate asymmetry generates a maximum of stability (a lower coefficient of variation of biomass in 73 their study), which suggests that the effects of asymmetry on stability are not trivial. Several major gaps limit our understanding of the consequences of asymmetry on metacommunity stability. First, the mechanisms linked to asymmetry and promoting asynchrony are not identified. Second, the way perturbations propagate in heterogenous metacommunities remains unknown. Stability patterns are tightly linked to synchrony patterns since the asynchrony of local population dynamics leads to more stable dynamics (low biomass CV) at higher scales due to compensatory dynamics (Loreau et al., 2003; Gonzalez and Loreau, 2008; Loreau and de Mazancourt, 2013; Wilcox et al., 2017). However, Quévreux et al. (2021a) showed that the perturbation and dispersal of particular trophic levels govern synchrony and stability in sym-81 metric metacommunities. Therefore, we expect asymmetric effects of perturbations affecting particular patches.

Here, we expect that asymmetry is not a generic stabilising factor, as claimed by previous works

Table 1: Approximative relative increase in predation risk between low-risk and high-risk environments (equivalent to the asymmetry of interaction strength γ in Figure 1). See Gorini et al. (2012) for an extended review and more references.

Predator	Prey	γ	Reference	
American marten	Vole species	1.6	Andruskiw et al., 2008	
Wolf	Moose	14-100	Gervasi et al., 2013	
Wolf	Roe deer	2.5 - 8	Gervasi et al., 2013	
Wolf	Elk	10	Kauffman et al., 2007	
Savannah predators	Savannah ungulates	1.5 - 4.5	Thaker et al., 2011	
Artificial gecko	Australian predators	2.8	Hansen et al., 2019	
Lynx	Roe deer	2	Gehr et al., 2020	
Puma	Vicuña	1.6	Donadio and Buskirk, 2016	

(Rooney et al., 2006; Goldwyn and Hastings, 2009; Ruokolainen et al., 2011), but strongly depends on which patch is perturbed according to its characteristics. To explore this statement, we consider a simple metacommunity model of two patches hosting a predator-prey couple and with asymmetric interaction strength and resource supply. The stability of the metacommunity is assessed by the response at different scales when prey receive stochastic perturbations in one of the two patches. We show contrasting effects of asymmetry: perturbing prey in the fast patch (equivalent to the fast channel defined by Rooney and McCann, 2012) promotes prey synchrony and decreases predator stability at the metapopulation scale while perturbing the slow patch has the opposite effects.

94 Methods

95 Metacommunity model

We use the model proposed by Quévreux et al. (2021a) based on the food chain model developed by Barbier and Loreau (2019). The model consists of two patches that each sustain a food chain with Lotka-Volterra predator-prey interactions (equations (1a) and (1b)).

$$\frac{1}{D}\frac{dB_1^{(1)}}{dt} = B_1^{(1)}\left(\omega\frac{g}{D} - B_1^{(1)} - \gamma maB_2^{(1)}\right) \tag{1a}$$

$$\frac{1}{mD} \frac{dB_2^{(1)}}{dt} = \underbrace{B_2^{(1)} \left(-B_2^{(1)} + \gamma \epsilon a B_1^{(1)} \right)}_{\text{intra-patch dynamics}} + \underbrace{d_2 \left(B_2^{(2)} - B_2^{(1)} \right)}_{\text{dispersal}} \tag{1b}$$

Prey have a positive effect εa on predators (ε is the conversion efficiency and a is the interspecific interaction rate relative to intraspecific interactions), and predators have a negative effect ma on prey (m is the predator to prey metabolic rate ratio)(Figure 1A). The time scale of the system is rescaled by the

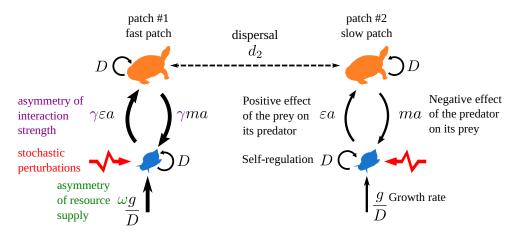


Figure 1: The metacommunity model consists of two patches, each sustaining a predator-prey couple linked by predators, which disperse at a very high scaled rate d_2 . Prey grow at a rate g/D and have a positive effect εa on predators, while predators have a negative effect ma on prey. Each species population is also limited by self-regulation D (negative intraspecific interactions). Spatial heterogeneity is embodied by the asymmetry of resource supply (green) and the interaction strength (purple), which are higher in patch #1 by factors ω and γ , respectively. Consistent with Rooney et al. (2006), patch #1 is called the fast patch, and patch #2 is called the slow patch. Prey receive stochastic perturbations either in patch #1 or in patch #2 (red arrows).

metabolic rate of prey, and biologic rates of each species i are rescaled by its intraspecific interaction rate

 D_i . Therefore, we obtain the relative growth rate g/D and the scaled dispersal rate d_i (see Appendix S1-1 for a detailed description of the rescaling). Considering scaled parameters and aggregated parameters 101 (εa) and (εa) enables us to explore a wide range of ecological situations. We do not describe the food web model in detail because it is not required to understand our results but a thorough description is 103 available in Appendix S1-1. Parameters and their values are summarised in Table 2. Predators disperse at a very high scaled rate d_2 and strongly couple the two patches. As in Rooney et 105 al. (2006), resource supply and interaction strength are asymmetric between patches since they are higher 106 in patch #1 by factors γ and ω respectively (Figure 1B). Patch #1 corresponds to the fast energy channel, 107 in which biomass has a high turnover, while patch #2 corresponds to the slow channel. Therefore, we 108 call patch #1 the fast patch and patch #2 the slow patch. We set $\gamma = \omega$ to ensure species persistence 109 over the entire range of parameters (see Figure S2-3 in the supporting information) but varying them 110 independently does not qualitatively change the results (see Figure S2-11 in the supporting information). 111 In the following, we only refer to γ for the sake of simplicity and only consider $\gamma \geq 1$ because $\gamma \leq 1$ just 112

Response to stochastic perturbations

swaps the roles of patches #1 and #2.

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We use the same methods as Quévreux et al. (2021a) to study the response of metacommunities to stochastic perturbations. Indeed, recent studies advocate for the use of the temporal variability of biomass (Haegeman et al., 2016; Arnoldi et al., 2018), which is measured by the coefficient of variation (CV), and can be easily measured experimentally. In addition, Wang and Loreau (2014, 2016), Wang et al. (2019), and Jarillo et al. (2022) showed that CVs scale up from local populations to community, regional and metacommunity levels, therefore providing a comparison of stability at different scales. Here, we provide only a brief description of the main concepts, but a thorough description is available in Appendix S1.

Prey in the fast or slow channel receive stochastic perturbations that are represented by equation (2).

$$dB_i = \underbrace{f_i(B_1, ..., B_S)dt}_{\text{Deterministic}} + \underbrace{\sigma_i \sqrt{B_i^*} dW_i}_{\text{Perturbation}}$$
(2)

 $f_i(B_1, ..., B_S)$ represents the deterministic part of the dynamics of species i, as described by equations (1a) and (1b)). Stochastic perturbations are defined by their standard deviation σ_i and dW_i , a white noise term with a mean of 0 and variance of 1. Perturbations also scale with the square root of the biomass at equilibrium B_i^* of the perturbed population. Such scaling makes the perturbations similar to demographic stochasticity (from birth-death processes) that evenly affect each species regardless of abundance (Arnoldi et al., 2019). In other words, the ratio of species biomass variance to perturbation variance is roughly independent of biomass, which disentangles the effect of asymmetry on perturbation transmission from its effect on species abundance.

In the following, we assess the temporal variability of each population after stochastic perturbations affect the metacommunity in the vicinity of equilibrium. Therefore, we linearise the system in the vicinity of equilibrium to obtain equation (3) where $X_i = B_i - B_i^*$ is the deviation from equilibrium.

$$\frac{d\overrightarrow{X}}{dt} = J\overrightarrow{X} + T\overrightarrow{E} \tag{3}$$

J is the Jacobian matrix, which represents the linearised direct effects between populations in the vicinity of equilibrium, and T defines how the perturbations $E_i = \sigma_i dW_i$ apply to the system (i.e., which species they affect and how they scale with biomass, where T is a diagonal matrix whose terms are $T_{ii} = \sqrt{B_i^*}$).

We then obtain the variance-covariance matrix C^* of species biomasses (variance-covariance matrix of \overrightarrow{X}) from the variance-covariance matrix of perturbations V_E (variance-covariance matrix of \overrightarrow{E}) by solving the Lyapunov equation equation (4) (Arnold, 1974; Wang et al., 2015; Arnoldi et al., 2016; Quévreux

Table 2: Table of parameters. σ_i is set very small to keep the system in the vicinity of equilibrium. More combinations of εa and ma are tested in the supporting information. d_2 is set very high to emphasise the high mobility of predators and their ability to couple prey populations. ω is set equal to γ .

parameter	interpretation	value
σ_i	standard deviation of stochastic noise	10^{-3}
g	net growth rate of prey	1
r	death rate of predators	0
D	self-regulation	1
ϵ	conversion efficiency	0.65
m	predator/prey metabolic rate ratio	0.65
a	attack rate	1.54,
εa	positive effect of prey on predators	1
ma	negative effect of predators on prey	1
d_2	scaled dispersal rate of predators	10^{6}
ω^-	asymmetry of resource supply	[1,10]
γ	asymmetry of interaction strength	[1,10]

et al., 2021a).

$$JC^* + C^*J^\top + TV_ET^\top = 0 (4)$$

The expressions for V_E and T and the method to solve the Lyapunov equation are detailed in Appendix S1-133 5. From the variance-covariance matrix C^* , we compute the coefficient of correlation of the biomass dynamics between the two populations of each species (see equation (20) in the supporting information) 135 and we measure the stability with the coefficient of variation (CV) of the biomass. In addition, biomass 136 CVs can be measured at different scales: population scale (e.g., biomass CV of prey in patch #1), 137 metapopulation scale (e.g., CV of the total biomass of prey) and metacommunity scale (e.g., CV of 138 the total biomass of predator and prey put together) to assess the effects of asymmetry at local and 139 regional scales (Figure 4A and see Appendix S1-6). Finally, we quantify the synchrony of the dynamics of the different populations with the coefficient of correlation, which is also computed from the variance-141 covariance matrix C^* (Appendix S1-6). 142

$\mathbf{Results}$

Source-sink effect

In the following, we consider a metacommunity in which only predators are able to disperse ($d_2 = 10^6$)
to be consistent with the common observations of predators coupling different energy channels, which can
differ in their energy turnover with a fast patch (patch #1 here) and a slow patch (patch #2, see Figure 1).

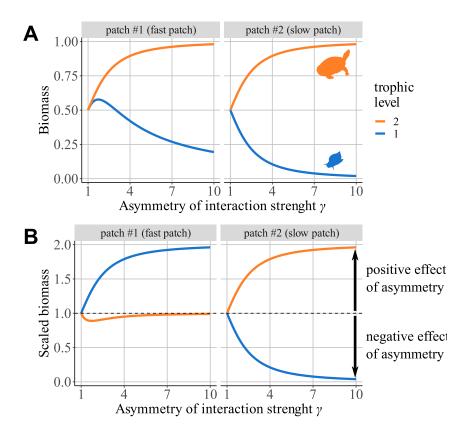


Figure 2: **A)** Distribution of the biomass of each species among patches depending on the asymmetry of interaction strength γ . **B)** Distribution of biomasses scaled by their value in a metacommunity without dispersal $(B_{scaled} = B_{d_2>0}/B_{d_2=0})$.

Varying the asymmetry of interaction strength γ is equivalent to varying the interaction strength a in patch #1 and has the same effects on species biomass: first increasing γ increases predator biomass by 149 increasing prey consumption, then it decreases predator biomass because of resource overexploitation 150 (Figure S2-2 in the supporting information). This leads to different biomass distributions in patches #1 151 and #2 (Figure 2A). Predator biomass increases with γ and is the same in both patches because their 152 high dispersal rate balances any difference. Prey biomass is higher in patch #1 than in patch #2, and 153 both decrease with γ , except in patch #1, where we first observe a small increase for $\gamma < 2$. This response 154 is due to source-sink effects: the increase in prey consumption in patch #1 increases predator biomass 155 (source) that spills over patch #2 (sink) due to dispersal (Figure 2B and Figure 6A). Therefore, predator 156 biomass is lower in patch #1 and higher in patch #2 compared to what we expect in the same food chains in isolation (i.e., without dispersal). This also prevents predators from overexploiting prey in patch #1 158 by spreading the increased predator biomass across the metacommunity, which explains why we do not observe a decrease in predator biomass for high values of γ , as shown in Figure S2-2. Conversely, the 160 distribution of prey biomass across the two patches is opposite (higher in patch #1 and lower in patch #2).162

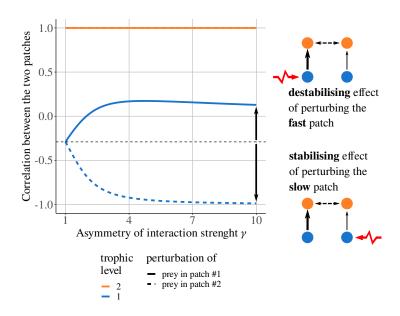


Figure 3: Spatial correlation between the populations of each species depending on asymmetry of interaction strength γ when predators disperse and prey are perturbed in patch #1 or #2. The dashed line emphasises the value of the correlation of prey populations without asymmetry ($\gamma = 1$).

Now, we describe how asymmetry of interaction strength γ shapes metacommunity stability at different

Effects on stability

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scales. Since predators have a very high scaled dispersal rate $(d_2 = 10^6)$, their populations are perfectly 165 correlated and display the same dynamics. Our main result is that prey become more correlated when 166 they are perturbed in patch #1 (fast channel in which $\gamma > 1$), while they become more anticorrelated when 167 they are perturbed in patch #2 (Figure 3). Increasing γ amplifies the difference in correlation between 168 these two scenarios, and this pattern qualitatively holds for various combinations of the physiological and 169 ecological parameters εa and ma (see Figure S2-6 in the supporting information). 170 Increasing the asymmetry of interaction strength γ has contrasting effects on biomass CV as well. At 171 the population scale, it increases the biomass CV of each population when prey are perturbed in the fast 172 patch (Figure 4B). When prey are perturbed in the slow patch, increasing γ slightly alters the biomass 173 CV of prey in patch #1, increases the biomass CV of prey in patch #2 and decreases the biomass CV 174 of predators. This discrepancy can be attributed to the strong effect of γ on prey biomass in patch #2 175 (Figure 2A): prey biomass strongly decreases with γ in patch #2, which increases their biomass CV. At the metapopulation scale, the asymmetry of interaction strength γ increases the biomass CV of 177 prey in both scenarios of perturbation (Figure 4C). However, this result is not true for all values of εa 178 and ma (Figure S2-7A in the supporting information) because of the various responses of prey biomass to 179

 γ among patches (Figure S2-5A in the supporting information). The biomass CV of predators is higher

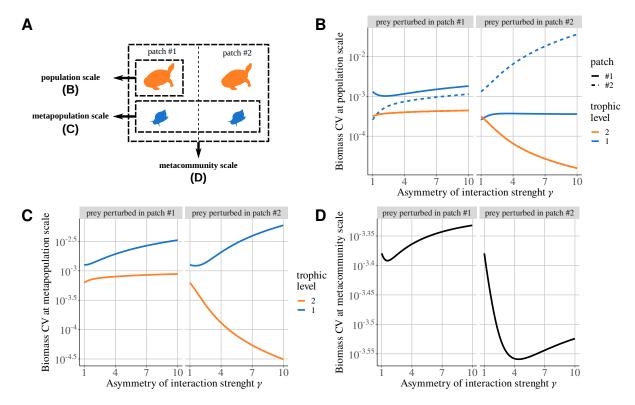


Figure 4: Stability at different scales depending on asymmetry of interaction strength γ when predators disperse and prey are perturbed in patch #1 or #2. **A)** The temporal variability in the metacommunity is assessed by the coefficient of variation (CV) of biomass at different scales: population scale (biomass CV of one species in one patch), metapopulation scale (CV of the total biomass of one species across patches) and metacommunity scale (CV of the total biomass of the entire metacommunity). **B)** Biomass CV at the population scale. **C)** Biomass CV at the metapopulation scale (CV of the total biomass of the metacommunity.

when prey are perturbed in the fast patch (patch #1) compared to the case in which prey are perturbed in the slow patch (#2) (Figure 4C), which is consistent for all values of εa and ma (see Figure S2-7A in the supporting information).

Finally, stability at the metacommunity scale depends on the distribution of biomass and CV among species. In our case ($\varepsilon a=1$ and ma=1), predators have the largest total biomass (Figure 2A) and drive the biomass CV at the metacommunity scale for low values of asymmetry of interaction strength γ (Figure 4D). For high values of γ , when prey are perturbed in patch #2, the CV of total biomass increases with γ because it is driven by prey in patch #2, whose biomass CV is much higher than the biomass CV of predators, which compensates for their lower biomass.

Underlying mechanisms

To unveil the mechanisms governing the stability of heterogeneous metacommunities, we look deeper into the dynamics after a pulse perturbation (Figure 5A) and explain them with the direct effects between species quantified by the Jacobian matrix (see equation (3)). When the perturbation of prey occurs in

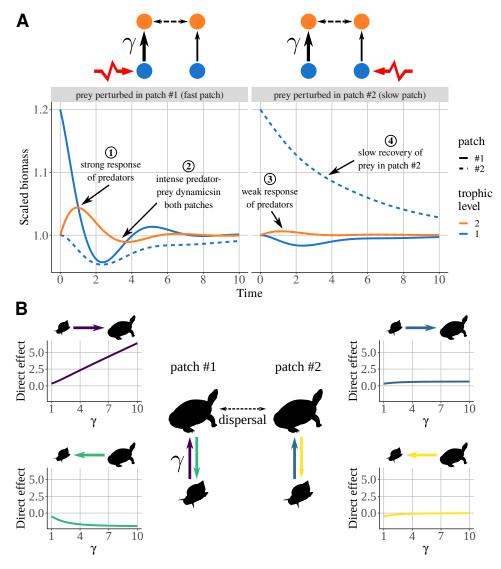


Figure 5: **A)** Time series of biomasses rescaled by their value at equilibrium after an increase of 20% in prey biomass in patch #1 (left panel) or patch #2 (right panel) for a value of interaction strength asymmetry $\gamma = 3$. **B)** Direct effect of prey on predator (and vice versa) depending on interaction strength asymmetry γ . Direct effects correspond to the terms of the Jacobian matrix.

patch #1, the strong direct effect of prey on predators (and vice versa) in patch #1 due to γ (Figure 5B) 194 leads to a strong response of predators (1), which in turn drives the response of the two prey populations (2). In detail, predator biomass in patch #1 first increases because of the abundance of prey. 196 Then, predators deplete prey biomass in both patches and correlate their dynamics, which explains why 197 asymmetry of interaction strength γ increases prey correlation when prey are perturbed in patch #1. 198 When the perturbation of prey occurs in patch #2, the weak direct effect of prey on predators (Figure 5B) leads to a small response of predators (3). In turn, the very low direct effect of predators on 200 prey in patch #2 does not allow perturbations to ripple back to patch #2 where prey slowly recover from the initial perturbation (4) (Figure 6B). This slow recovery is emphasised by the source-sink dynamics 202 in the metacommunity (Figure 6A), which leads to a lower biomass of prey in patch #2 compared to a metacommunity without dispersal, therefore decreasing biomass flows in patch #2 and its recovery speed. This difference in recovery speed between patches #1 and #2 leads to the anti-correlation of prey populations because it increases the time interval in which they have opposite variations: an increase in the biomass of prey in patch #1, which follows the initial decrease due to predation, and a slow decrease in prey biomass in patch #2.

In summary, the asymmetry of interaction strength alters the biomass distribution among trophic levels in each patch (Figure 6A), which leads to strong predator-prey interactions in the fast patch and weak interactions in the slow patch (Figure 6B). Therefore, if the fast patch is perturbed, it drives the

patch does not significantly affect the fast patch, and both independently recover from perturbations,

dynamics of the slow patch and synchronises prey dynamics. Conversely, the perturbation of the slow

214 which desynchronises prey dynamics.

Discussion

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We have shown that the asymmetry of interaction strength has contrasting effects on stability de-216 pending on which patch is perturbed. Perturbing prey in the fast patch (in which interaction strength is the highest) tends to synchronise the dynamics of prev biomass and increases the temporal variability of 218 predator dynamics at the metapopulation scale, while perturbing prey in the slow patch desynchronises 219 prey dynamics and decreases the temporal predator dynamics. This discrepancy between the responses 220 is due to asymmetric transmission of perturbations within each patch, itself caused by the alteration of biomass distribution and the asymmetry of interaction strength (Figure 6A). Perturbations are strongly 222 transmitted from the fast patch to the slow patch, while the reverse transmission is weak (Figure 6B). 223 Consequently, the fast patch drives the dynamics of the metacommunity and synchronises prey dynam-224 ics, while the slow patch does not, and the almost independent recovery from perturbation in each patch 225 desynchronises prey dynamics. 226

Stability in a heterogeneous world

Our results show that spatial heterogeneity, which is represented by the asymmetry of interaction strength and resource supply as in Rooney et al. (2006), generates mechanisms that alter local and regional dynamics, which deeply changes the synchrony of population dynamics and the stability of the metacommunity at different scales. In a homogeneous metacommunity, the spatial correlation of the

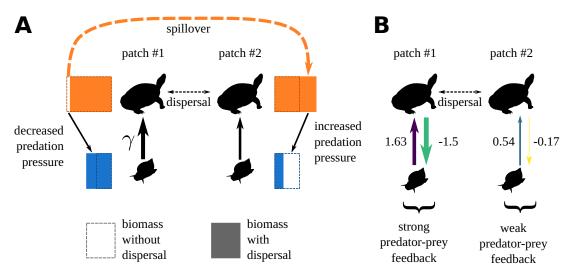


Figure 6: Mechanisms leading to the spatial synchrony of prey dynamics when prey are perturbed in the patch with the strongest interaction strength (patch #1 with $\gamma = 3$). A) Asymmetry in interaction strength alters the biomass distribution between the two patches. Increased interaction strength in patches #1 enhances the biomass production of predators that spill over patches from patch #1 to patch #2, therefore increasing prey biomass in patch #1 and decreasing it in patch #2. Plain rectangles represent the biomass of each population, while dashed rectangles represent the same population in a metacommunity without dispersal (no spillover effect). B) The asymmetry of biomass distribution and interaction strength alters the direct effects between predators and prey. High prey density and high interaction strength promote strong feedback in patch #1, which promotes the transmission of perturbations when prey are perturbed in patch #1. Low prey density and low interaction strength promote a weak feedback in patch #2, which limits the transmission of perturbations when prey are perturbed in patch #2. The discrepancy between the feedback strengths explains the synchrony of prey dynamics: when prey are perturbed in patch #1, the predator-prey couple in patch has an overwhelming impact on the metacommunity and drives the dynamics in patch #1, which leads to synchrony of prey populations. Conversely, perturbing prey in patch #2 has a minor impact on dynamics in patch #1, which decreases the synchrony of prey populations.

dynamics of the different populations of each species can be assessed from the correlation of the different species within each patch, the dispersing species making the link between the correlation patterns (see 233 Figure S2-25 for a summary of the results of Quévreux et al., 2021a). In other words, knowledge of the dynamics at the local scale is enough to completely understand the stability pattern at the meta-235 community scale. In a heterogeneous metacommunity, such a simple framework does not hold because each patch does not contribute equally to the dynamics due to the overwhelming influence of the fast 237 patch (see Figure S2-4 in the supporting information). Rooney et al. (2006) verbally explained that each patch has dynamics with different speeds: the fast channel (with higher interaction rates and resource 239 supplies) enables a quick recovery after a perturbation while the slow channel dampens the dynamics in the long term and prevents the system from overshooting. Here, we demonstrate that stability is 241 driven by the mechanisms responsible for the synchrony of prey populations. Perturbations are strongly transmitted from prey to predator in the fast patch, which enables strong transmission to the slow patch 243 and therefore synchronises the dynamics of prey populations (Figure 6B). Conversely, perturbations are poorly transmitted within the slow patch, which leads to the asynchrony of prey populations due to the

independent recovery in each patch. Undoubtedly, the dynamics at the metacommunity scale cannot be
assessed by the dynamics at the local scale, as in Quévreux et al. (2021a), and they are an emergent
property resulting from the tight interplay between the strength of perturbation transmission in each
patch.

Our description of the mechanisms underlying the apparent stabilising effects of spatial heterogeneity 250 should enlighten the results of previous theoretical studies. Goldwyn and Hastings (2009) and Ruokolainen et al. (2011) found that the asymmetry of interaction rate leads to asynchrony by generating 252 out-of-phase dynamics in a system with endogenous oscillations. In particular, Ruokolainen et al. (2011) found a U-shaped relationship: for moderate asymmetry, the spatial asynchrony of predator and prey 254 populations is maximal, which leads to optimum stability at the metacommunity scale. Our results suggest that moderate asymmetry would alter the phase of the oscillations in each patch while keeping 256 the amplitude of oscillations equivalent, therefore promoting asynchrony. Conversely, a strong asymmetry would increase the imbalance between oscillation amplitude and enable the fast patch to take 258 over the slow patch, which would bring back synchrony. However, their results rely on phase-locking (Jansen, 1999; Lloyd and May, 1999; Goldwyn and Hastings, 2008; Vasseur and Fox, 2009), which is 260 the coupling of the phase of oscillators embodied by predator-prey pairs in each patch. Although our results provide interesting insight into metacommunity dynamics, they cannot grasp the fine mechanisms 262 underlying nonlinear phenomena such as phase-locking and further studies are needed to identify these mechanisms. Finally, most of the theoretical studies consider the effect of spatial heterogeneity combined with variations in dispersal rate (Gravel et al., 2016; Anderson and Fahimipour, 2021) or habitat choice by predators (Rooney et al., 2006; Ruokolainen et al., 2011), therefore confounding the fundamental 266 effects of asymmetry. We have identified the mechanisms linked to asymmetry, but future work must now investigate their interactions with other processes such as density-dependent dispersal.

Generality of the effects of asymmetry on stability

Our main results is that asymmetry is stabilising when the slow patch is perturbed, while it is destabilising when the fast patch is perturbed. This result is strikingly robust to several deviations from the
original model we have described. First, we show that the described mechanisms are valid for a wide range
of ecological and physiological parameters leading to various distributions of biomass among predators
and prey (see Figures S2-5 and S2-6 in the supporting information). In addition, we observe the same
results for longer food chains as long as prey populations are directly coupled by the dispersing predator

(see Figures S2-15 and S2-16 in the supporting information). Currently, we do not identify a clear pattern for species lower in the food chain over a wide range of ecological and physiological parameters but further studies are needed to investigate the potential indirect effects propagating across the food chain. 278 Second, the mechanisms are not restricted to prey populations coupled by a mobile predator but also 279 apply to predator populations coupled by a mobile prey (see Figure S2-19 in the supporting information). 280 Therefore, we anticipate that mobile predators are not the only major drivers of synchrony and stability in ecosystems (Schindler and Scheuerell, 2002; Vadeboncoeur et al., 2005; Dolson et al., 2009; Olff et al., 282 2009; Rooney and McCann, 2012), and resource species may also have an equivalent impact. Taken 283 together, these two points strongly suggest that the mechanisms underlying stability and synchrony in 284 response to perturbations should be general to metacommunities regardless of the ecological parameters, biomass distribution and dispersal among species. 286 Spatial heterogeneity has often been presented as a generic condition generating mechanisms ensuring stability, but our results contradict this statement. The models focusing on the asymmetric feeding of 288 consumers on different energy channels or different patches showed that it promotes the existence of stable equilibria (McCann et al., 1998), greater asymptotic resilience (Rooney et al., 2006), asynchrony 290 of prey in response to predator perturbation (Rooney et al., 2006) and out-of-phase limit cycles (Goldwyn and Hastings, 2009; Ruokolainen et al., 2011). McCann et al. (1998) and Ruokolainen et al. (2011) 292 only observed the stabilising effect for moderate spatial heterogeneity, but it remains a stabilising factor nonetheless. All these studies considered measures of stability aiming to capture the general stability properties of metacommunities and miss the targeted effects of perturbations. Rooney et al. (2006) only considered the perturbation of the mobile predator and ignored the specific perturbation of the fast or slow channels. Although asymmetry does not necessarily promote stability, our results show that general mechanisms drive the response of metacommunities to localised perturbations, therefore providing a valuable framework to assess the response of ecosystems to localised perturbations due to human activity. Additionally, these mechanisms enable us to understand the effect of environmental perturbations affecting all patches. As demonstrated by Arnoldi et al. (2019), environmental perturbations affect abundant 301 populations the most, which is the prev population in the fast patch in our case (see Figure S2-22 in 302 the supporting information). Therefore, we anticipate that the fast patch will govern the dynamics of metacommunities in which all populations are perturbed (see Figure S2-24 in the supporting information).

₀₅ Implications for conservation

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The metacommunity framework has long been used in conservation ecology (Johnson et al., 2013; Patrick et al., 2021). Conservation efforts are usually concentrated on particular locations and useful 307 management must consider the ecological processes acting at the landscape scale (Van Teeffelen et al., 2012; Chase et al., 2020). For instance, spatial heterogeneity is key to ensuring species coexistence and 309 diversity at the regional scale, which ultimately provides important ecosystem services in agricultural 310 landscapes (Bennett et al., 2006). A large corpus of theoretical studies explored the local recovery of 311 communities in a landscape receiving perturbations (Mouquet et al., 2011; Economo, 2011; Holyoak et 312 al., 2020; Jacquet et al., 2022). However, these studies focus on extinction events recovered by dispersal 313 events in a patch dynamics framework, and little is known about the effect of moderate or small pertur-314 bations. In this context, the present study provides valuable insight into fine-scale dynamics in response 315 to perturbations. 316

Our results show that species interactions are a major driver of synchrony in heterogeneous metacom-317 munities. Even if the species of interest does not disperse significantly, the synchrony of the dynamics 318 of its different populations can strongly depend on the interactions with another species with a higher 319 dispersal across the landscape. For instance, Howeth and Leibold (2013) showed that predatory fish 320 promote the asynchrony of oscillating populations of zooplankton in a mesocosm experiment. Therefore, 321 species endorsing this role are called "mobile link organisms" (Lundberg and Moberg, 2003) and are par-322 ticularly targeted by conservation policy because they have major impacts on community dynamics and 323 ecosystem functioning (Soulé et al., 2005; Brodie et al., 2018). Such a species can be considered keystone 324 species (Mills and Doak, 1993) and must be clearly identified to properly manage the conservation of the other interaction species. However, our results show that mobile link organisms are not the only 326 driver of metacommunity stability, and the patch being perturbed also has a major impact. The concept 327 of a keystone community, defined by Mouquet et al. (2013) for communities whose destruction causes 328 species extinction or a decrease in biomass production, can be applied to better assess the stability of metacommunities. Therefore, identifying the communities living in fast and slow patches should be key 330 for conservation management aiming to mitigate the effects of perturbations.

According to our results, mitigating the effects of perturbations affecting the patch in which interaction strength is the highest is critical to avoid the synchrony of prey dynamics (Figure 3) and ensure predator stability (Figure 4C). Then, the patch in which the interaction strength between the species of interest

and the mobile link organism is the highest must be identified. Conservation policies usually target preserved areas because they are characterised by high species richness but identifying them as fast or slow patches is not trivial. Urban ecology is a relevant example because many species dwell in cities and 337 less anthropised ecosystems (e.g., agricultural and natural landscapes). Urban areas can be considered 338 fast patches because of the abundance of resources (parameters ω in our model) for opportunistic species, 339 but they can also be considered slow patches because of the reduced predation pressure (parameter γ in our model), cities acting as safe spaces (see Shochat et al. (2006) and Shochat et al. (2010) for review). 341 Typically, birds and rodents can find plenty of food due to human wastes, public parks and feeding while experiencing less predation (Rebolo-Ifrán et al., 2017). Therefore, focusing conservation efforts on urban areas to mitigate the perturbations affecting their ecosystem may be as important as protecting wild areas to protect species at the metapopulation scale. 345

346 Conclusion

Asymmetry of interaction strength, and spatial heterogeneity in general, is not stabilising factor per

se because perturbing prey in the fast patch leads to the synchrony of the dynamics of prey populations
and increases the temporal variability of the mobile predator linking the two patches. Therefore, the
response of metacommunities to perturbations is strongly context dependent, i.e., a good knowledge of
the characteristics of each patch relative to each other is required to assess stability at the metacommunity
scale. Based on our findings, we advocate for conservation efforts to target key patches not only according
to species richness or biomass density but also according to the distribution of interaction strength across
the metacommunity.

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359 Data accessibility

The R codes to reproduce the results and the figures are available on GitHub (https://github.com/
PierreQuevreux/model_metacommunity_spatial_heterogeneity).

References

```
Amarasekare, P., & Nisbet, R. M. (2001). Spatial heterogeneity, source-sink dynamics, and the local coexistence of competing
363
              species. The American Naturalist, 158(6), 572–584. https://doi.org/10.1086/323586
     Anderson, K. E., & Fahimipour, A. K. (2021). Body size dependent dispersal influences stability in heterogeneous meta-
              366
     Andruskiw, M., Fryxell, J. M., Thompson, I. D., & Baker, J. A. (2008). Habitat-mediated variation in predation risk by the
367
              american marten. Ecology, 89(8), 2273-2280. https://doi.org/10.1890/07-1428.1
368
     Arnold, L. (1974). Stochastic differential equations: Theory and applications. Wiley.
     Arnoldi, J.-F., Bideault, A., Loreau, M., & Haegeman, B. (2018). How ecosystems recover from pulse perturbations: A
370
              theory of short- to long-term responses. Journal of Theoretical Biology, 436, 79-92. https://doi.org/10.1016/j.
371
              jtbi.2017.10.003
372
     Arnoldi, J.-F., Loreau, M., & Haegeman, B. (2016). Resilience, reactivity and variability: A mathematical comparison of
373
              ecological stability measures. Journal of Theoretical Biology, 389, 47-59. https://doi.org/10.1016/j.jtbi.2015.10.
374
              012
375
     Arnoldi, J.-F., Loreau, M., & Haegeman, B. (2019). The inherent multidimensionality of temporal variability: How common
376
              and rare species shape stability patterns (J. Chase, Ed.). Ecology Letters, 22(10), 1557-1567. https://doi.org/10.
377
              1111/ele.13345
378
    Barbier, M., & Loreau, M. (2019). Pyramids and cascades: A synthesis of food chain functioning and stability. Ecology
379
              Letters, 22(2), 405-419. https://doi.org/10.1111/ele.13196
380
    Bennett, A. F., Radford, J. Q., & Haslem, A. (2006). Properties of land mosaics: Implications for nature conservation in
381
              agricultural environments. Biological Conservation, 133(2), 250-264. https://doi.org/10.1016/j.biocon.2006.06.
382
383
    Brodie, J. F., Redford, K. H., & Doak, D. F. (2018). Ecological function analysis: Incorporating species roles into conser-
              vation. Trends in Ecology & Evolution, 33(11), 840-850. https://doi.org/10.1016/j.tree.2018.08.013
    Chase, J. M., Jeliazkov, A., Ladouceur, E., & Viana, D. S. (2020). Biodiversity conservation through the lens of metacommu-
386
              nity ecology. Annals of the New York Academy of Sciences, 1469(1), 86-104. https://doi.org/10.1111/nyas.14378
387
     Chesson, P. (2000). Mechanisms of maintenance of species diversity. Annual Review of Ecology and Systematics, 31(1),
              343-366.\ https://doi.org/10.1146/annurev.ecolsys. 31.1.343
389
    Dolson, R., McCann, K., Rooney, N., & Ridgway, M. (2009). Lake morphometry predicts the degree of habitat coupling by
390
              a mobile predator. Oikos, 118(8), 1230–1238. https://doi.org/10.1111/j.1600-0706.2009.17351.x
391
    Donadio, E., & Buskirk, S. W. (2016). Linking predation risk, ungulate antipredator responses, and patterns of vegetation
              in the high Andes. Journal of Mammalogy, 97(3), 966-977. https://doi.org/10.1093/jmammal/gyw020
393
    Economo, E. P. (2011). Biodiversity conservation in metacommunity networks: Linking pattern and persistence. The Amer-
394
              ican Naturalist, 177(6), E167-E180. https://doi.org/10.1086/659946
395
    Gehr, B., Bonnot, N. C., Heurich, M., Cagnacci, F., Ciuti, S., Hewison, A. J. M., Gaillard, J.-M., Ranc, N., Premier, J.,
              Vogt, K., Hofer, E., Ryser, A., Vimercati, E., & Keller, L. (2020). Stay home, stay safe—Site familiarity reduces
397
              predation risk in a large herbivore in two contrasting study sites (L. Prugh, Ed.). Journal of Animal Ecology,
398
              89(6), 1329-1339. https://doi.org/10.1111/1365-2656.13202
399
```

```
Landscape structure and wolf behavior generate different predation patterns in two sympatric ungulates. Ecological
401
              Applications, 23(7), 1722-1734. https://doi.org/10.1890/12-1615.1
402
     Goldwyn, E. E., & Hastings, A. (2008). When can dispersal synchronize populations? Theoretical Population Biology, 73(3),
403
              395-402. https://doi.org/10.1016/j.tpb.2007.11.012
404
     Goldwyn, E. E., & Hastings, A. (2009). Small heterogeneity has large effects on synchronization of ecological oscillators.
405
              Bulletin of Mathematical Biology, 71(1), 130-144. https://doi.org/10.1007/s11538-008-9355-9
406
     Gonzalez, A., & Loreau, M. (2008). The causes and consequences of compensatory dynamics in ecological communities.
407
              Annual Review of Ecology, Evolution, and Systematics, 40(1), 393-414. https://doi.org/10.1146/annurev.ecolsys.
408
              39.110707.173349
409
    Gorini, L., Linnell, J. D. C., May, R., Panzacchi, M., Boitani, L., Odden, M., & Nilsen, E. B. (2012). Habitat heterogeneity
410
              and mammalian predator-prey interactions: Predator-prey interactions in a spatial world. Mammal Review, 42(1),
411
412
              55-77. https://doi.org/10.1111/j.1365-2907.2011.00189.x
     Gravel, D., Massol, F., & Leibold, M. A. (2016). Stability and complexity in model meta-ecosystems. Nature Communica-
413
              tions, 7, 12457. https://doi.org/10.1038/ncomms12457
414
     Haegeman, B., Arnoldi, J.-F., Wang, S., de Mazancourt, C., Montoya, J. M., & Loreau, M. (2016). Resilience, invariability,
415
              and ecological stability across levels of organization. bioRxiv. https://doi.org/10.1101/085852
416
     Hansen, N. A., Sato, C. F., Michael, D. R., Lindenmayer, D. B., & Driscoll, D. A. (2019). Predation risk for reptiles is
417
              highest at remnant edges in agricultural landscapes (T. M. Lee, Ed.). Journal of Applied Ecology, 56(1), 31-43.
418
              https://doi.org/10.1111/1365-2664.13269
419
    Holt, R. D. (1984). Spatial heterogeneity, indirect interactions, and the coexistence of prey species. The American Naturalist,
420
              124(3), 377-406. https://doi.org/10.1086/284280
421
     Holyoak, M., Caspi, T., & Redosh, L. W. (2020). Integrating disturbance, seasonality, multi-year temporal dynamics, and
422
              dormancy into the dynamics and conservation of metacommunities. Frontiers in Ecology and Evolution, 8, 571130.
423
              https://doi.org/10.3389/fevo.2020.571130
     Howeth, J. G., & Leibold, M. A. (2013). Predation inhibits the positive effect of dispersal on intraspecific and interspecific
425
              synchrony in pond metacommunities. Ecology, 94(10), 2220-2228. https://doi.org/10.1890/12-2066.1
426
     Jacquet, C., Munoz, F., Bonada, N., Datry, T., Heino, J., & Jabot, F. (2022). Temporal variation of patch connectivity
427
              determines biodiversity recovery from recurrent disturbances. bioRxiv. https://doi.org/10.1101/2022.01.02.474736
428
     Jansen, V. A. A. (1999). Phase locking: Another cause of synchronicity in predator-prey systems. Trends in Ecology &
429
              Evolution, 14(7), 278-279. https://doi.org/10.1016/S0169-5347(99)01654-7
430
     Jarillo, J., Cao-García, F. J., & De Laender, F. (2022). Spatial and ecological scaling of stability in spatial community
431
432
              networks. Frontiers in Ecology and Evolution, 10, 861537. https://doi.org/10.3389/fevo.2022.861537
     Johnson, P. T. J., Hoverman, J. T., McKenzie, V. J., Blaustein, A. R., & Richgels, K. L. D. (2013). Urbanization and wetland
433
              communities: Applying metacommunity theory to understand the local and landscape effects (M. Cadotte, Ed.).
434
              Journal of Applied Ecology, 50(1), 34-42. https://doi.org/10.1111/1365-2664.12022
435
```

Gervasi, V., Sand, H., Zimmermann, B., Mattisson, J., Wabakken, P., & Linnell, J. D. C. (2013). Decomposing risk:

400

```
shapes predation in a newly restored predator-prey system. Ecology Letters, 10(8), 690-700. https://doi.org/10.
437
              1111/j.1461-0248.2007.01059.x
438
    Leibold, M. A., & Chase, J. M. (2017). Metacommunity Ecology (Vol. 59). Princeton University Press. https://doi.org/10.
439
              2307/j.ctt1wf4d24
440
    Leibold, M. A., Holyoak, M., Mouquet, N., Amarasekare, P., Chase, J. M., Hoopes, M. F., Holt, R. D., Shurin, J. B., Law,
441
              R., Tilman, D., Loreau, M., & Gonzalez, A. (2004). The metacommunity concept: A framework for multi-scale
442
             community ecology. Ecology Letters, 7(7), 601-613. https://doi.org/10.1111/j.1461-0248.2004.00608.x
443
    Lloyd, A. L., & May, R. M. (1999). Synchronicity, chaos and population cycles: Spatial coherence in an uncertain world.
              445
    Loreau, M., Mouquet, N., & Gonzalez, A. (2003). Biodiversity as spatial insurance in heterogeneous landscapes. Proceedings
446
              of the National Academy of Sciences, 100(22), 12765-12770. https://doi.org/10.1073/pnas.2235465100
447
    Loreau, M., & de Mazancourt, C. (2008). Species synchrony and its drivers: Neutral and nonneutral community dynamics
448
             in fluctuating environments. The American Naturalist, 172(2), E48-E66. https://doi.org/10.1086/589746
449
    Loreau, M., & de Mazancourt, C. (2013). Biodiversity and ecosystem stability: A synthesis of underlying mechanisms.
450
              Ecology Letters, 16, 106–115. https://doi.org/10.1111/ele.12073
451
    Lundberg, J., & Moberg, F. (2003). Mobile link organisms and ecosystem functioning: Implications for ecosystem resilience
452
             and management. Ecosystems, 6(1), 0087-0098. https://doi.org/10.1007/s10021-002-0150-4
453
    May, R. M. (1972). Will a large complex system be stable? Nature, 238(5364), 413-414. https://doi.org/10.1038/238413a0
454
    McCann, K. S., Hastings, A., & Huxel, G. R. (1998). Weak trophic interactions and the balance of nature. Nature, 395 (6704),
455
              794-798. https://doi.org/10.1038/27427
    Mills, L. S., & Doak, D. F. (1993). The keystone-species con-cept in ecology and conservation. BioScience, 43(4), 219–224.
457
             \rm https://doi.org/10.2307/1312122
458
    Mouquet, N., Gravel, D., Massol, F., & Calcagno, V. (2013). Extending the concept of keystone species to communities and
459
             ecosystems (B. Blasius, Ed.). Ecology Letters, 16(1), 1-8. https://doi.org/10.1111/ele.12014
460
    Mouquet, N., & Loreau, M. (2002). Coexistence in metacommunities: The regional similarity hypothesis. The American
461
              Naturalist, 159(4), 420-426. https://doi.org/10.1086/338996
462
    Mouquet, N., & Loreau, M. (2003). Community patterns in source-sink metacommunities. The American Naturalist, 162(5),
463
             544-557. https://doi.org/10.1086/378857
    Mouquet, N., Matthiessen, B., Miller, T., & Gonzalez, A. (2011). Extinction debt in source-sink metacommunities (T.
465
             Romanuk, Ed.). PLoS ONE, 6(3), e17567. https://doi.org/10.1371/journal.pone.0017567
466
    Neutel, A.-M., Heesterbeek, J. A. P., & De Ruiter, P. C. (2002). Stability in real food webs: Weak links in long loops.
467
              Science, 296(5570), 1120-1123. https://doi.org/10.1126/science.1068326
    Olff, H., Alonso, D., Berg, M. P., Eriksson, B. K., Loreau, M., Piersma, T., & Rooney, N. (2009). Parallel ecological
469
              networks in ecosystems. Philosophical Transactions of the Royal Society B: Biological Sciences, 364 (1524), 1755-
470
              1779. https://doi.org/10.1098/rstb.2008.0222
471
```

Kauffman, M. J., Varley, N., Smith, D. W., Stahler, D. R., MacNulty, D. R., & Boyce, M. S. (2007). Landscape heterogeneity

436

```
Patrick, C. J., Anderson, K. E., Brown, B. L., Hawkins, C. P., Metcalfe, A., Saffarinia, P., Siqueira, T., Swan, C. M.,
              Tonkin, J. D., & Yuan, L. L. (2021). The application of metacommunity theory to the management of riverine
473
              ecosystems. WIREs Water, 8(6). https://doi.org/10.1002/wat2.1557
474
     Quévreux, P., Barbier, M., & Loreau, M. (2021a). Synchrony and perturbation transmission in trophic metacommunities.
475
              The American Naturalist, 714131. https://doi.org/10.1086/714131
476
     Quévreux, P., Pigeault, R., & Loreau, M. (2021b). Predator avoidance and foraging for food shape synchrony and response
477
              to perturbations in trophic metacommunities. Journal of Theoretical Biology, 528, 110836. https://doi.org/10.
478
              1016/j.jtbi.2021.110836
479
     Rebolo-Ifrán, N., Tella, J. L., & Carrete, M. (2017). Urban conservation hotspots: Predation release allows the grassland-
480
              specialist burrowing owl to perform better in the city. Scientific Reports, 7(1), 3527. https://doi.org/10.1038/
481
              s41598-017-03853-z
482
     Rooney, N., & McCann, K. S. (2012). Integrating food web diversity, structure and stability. Trends in Ecology & Evolution,
483
              27(1), 40-46. https://doi.org/10.1016/j.tree.2011.09.001
484
     Rooney, N., McCann, K. S., Gellner, G., & Moore, J. C. (2006). Structural asymmetry and the stability of diverse food
485
              webs. Nature, 442 (7100), 265-269. https://doi.org/10.1038/nature04887
486
     Ruokolainen, L., Abrams, P. A., McCann, K. S., & Shuter, B. J. (2011). The roles of spatial heterogeneity and adaptive
487
              movement in stabilizing (or destabilizing) simple metacommunities. Journal of Theoretical Biology, 291, 76-87.
488
              https://doi.org/10.1016/j.jtbi.2011.09.004
489
     Schindler, D. E., & Scheuerell, M. D. (2002). Habitat coupling in lake ecosystems. Oikos, 98(2), 177-189. https://doi.org/
490
              10.1034/j.1600-0706.2002.980201.x
491
     Shochat, E., Warren, P., Faeth, S., Mcintyre, N., & Hope, D. (2006). From patterns to emerging processes in mechanistic
492
              urban ecology. Trends in Ecology & Evolution, 21(4), 186-191. https://doi.org/10.1016/j.tree.2005.11.019
493
     Shochat, E., Lerman, S. B., Anderies, J. M., Warren, P. S., Faeth, S. H., & Nilon, C. H. (2010). Invasion, competition, and
494
              biodiversity loss in urban ecosystems. BioScience, 60(3), 199-208. https://doi.org/10.1525/bio.2010.60.3.6
495
     Soulé, M. E., Estes, J. A., Miller, B., & Honnold, D. L. (2005). Strongly interacting species: Conservation policy, manage-
              ment, and ethics. BioScience, 55(2), 168. https://doi.org/10.1641/0006-3568(2005)055[0168:SISCPM]2.0.CO;2
497
     Thaker, M., Vanak, A. T., Owen, C. R., Ogden, M. B., Niemann, S. M., & Slotow, R. (2011). Minimizing predation risk in a
498
              landscape of multiple predators: Effects on the spatial distribution of African ungulates. Ecology, 92(2), 398-407.
499
              https://doi.org/10.1890/10-0126.1
500
     Vadeboncoeur, Y., McCann, K. S., Zanden, M. J. V., & Rasmussen, J. B. (2005). Effects of multi-chain omnivory on the
501
              strength of trophic control in lakes. Ecosystems, 8(6), 682-693. https://doi.org/10.1007/s10021-003-0149-5
502
     Van Teeffelen, A. J., Vos, C. C., & Opdam, P. (2012). Species in a dynamic world: Consequences of habitat network dynamics
503
              on conservation planning. Biological Conservation, 153, 239-253. https://doi.org/10.1016/j.biocon.2012.05.001
     Vasseur, D. A., & Fox, J. W. (2009). Phase-locking and environmental fluctuations generate synchrony in a predator-prey
505
              community. Nature, 460 (7258), 1007-1010. https://doi.org/10.1038/nature08208
506
     Wang, S., Haegeman, B., & Loreau, M. (2015). Dispersal and metapopulation stability. PeerJ, 3, e1295. https://doi.org/
507
              10.7717/peerj.1295
```

472

```
Wang, S., Lamy, T., Hallett, L. M., & Loreau, M. (2019). Stability and synchrony across ecological hierarchies in heteroge-
509
              neous metacommunities: Linking theory to data. Ecography, 0(0). https://doi.org/10.1111/ecog.04290
510
     Wang, S., & Loreau, M. (2014). Ecosystem stability in space: \textalpha, \textbeta and \textgamma variability. Ecology
511
              Letters, 17(8), 891-901. https://doi.org/10.1111/ele.12292
512
     Wang, S., & Loreau, M. (2016). Biodiversity and ecosystem stability across scales in metacommunities (F. Jordan, Ed.).
513
              Ecology\ Letters,\ 19(5),\ 510-518.\ https://doi.org/10.1111/ele.12582
514
     Wilcox, K. R., Tredennick, A. T., Koerner, S. E., Grman, E., Hallett, L. M., Avolio, M. L., Pierre, K. J. L., Houseman, G. R.,
515
              Isbell, F., Johnson, D. S., Alatalo, J. M., Baldwin, A. H., Bork, E. W., Boughton, E. H., Bowman, W. D., Britton,
516
              A. J., Cahill, J. F., Collins, S. L., Du, G., ... Zhang, Y. (2017). Asynchrony among local communities stabilises
517
```

518

ecosystem function of metacommunities. $Ecology\ Letters,\ 20(12),\ 1534-1545.\ https://doi.org/10.1111/ele.12861$

S1 Complementary material and methods

520 S1-1 Model description

The model has been originally developed by Barbier and Loreau (2019), who considered a food chain model with a simple metabolic parametrisation. Their model corresponds to the "intra-patch dynamics" part of equations (5a) and (5b) to which we graft a dispersal term to consider a metacommunity with two patches.

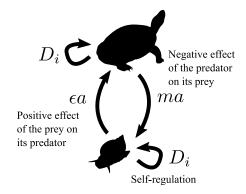
$$\frac{dB_1^{(1)}}{dt} = B_1^{(1)}(\omega_1 g_1 - D_1 B_1^{(1)} - \gamma_1 \alpha_{2,1} B_2^{(1)}) + \delta_1 (B_1^{(2)} - B_1^{(1)})$$
(5a)

$$\frac{dB_{i}^{(1)}}{dt} = \underbrace{B_{i}^{(1)}(-r_{i} - D_{i}B_{i}^{(1)} + \gamma_{1}\epsilon\alpha_{i,i-1}B_{i-1}^{(1)} - \gamma_{1}\alpha_{i+1,i}B_{i+1}^{(1)})}_{\text{intra-patch dynamics}} + \underbrace{\delta_{i}(B_{i}^{(2)} - B_{i}^{(1)})}_{\text{dispersal}} \tag{5b}$$

 $B_i^{(1)}$ is the biomass of trophic level i in the patch #1, ϵ is the biomass conversion efficiency and $\alpha_{i,j}$ is the interaction strength between consumer i and prey j. Species i disperses between the two patches at rate δ_i . The density independent net growth rate of primary producers g_i in equations (5a), the mortality rate of consumers r_i in equations (5b) and the density dependent mortality rate D_i scale with species metabolic rates m_i as biological rates are linked to energy expenditure.

$$g_1 = m_1 g \qquad r_i = m_i r \qquad D_i = m_i D \tag{6}$$

In order to get a broad range of possible responses, we assume the predator-prey metabolic rate ratio m and the interaction strength to self-regulation ratio a to be constant. These ratios capture the relations between parameters and trophic levels. This enables us to consider contrasting situations while keeping the model as simple as possible.



$$m = \frac{m_{i+1}}{m_i} \qquad a = \frac{\alpha_{i,i-1}}{D_i} \qquad d_i = \frac{\delta_i}{D_i}$$
 (7)

Varying m leads to food chains where predators have faster or slower biomass dynamics than their prey and varying a leads to food chains where interspecific interactions prevail or not compared with intraspecific interactions. As all biological rates are rescaled by D_i , we also define d_i , the dispersal rate relative to self-regulation (referred as scaled dispersal rate in the rest of the study), in order to keep the

values of the dispersal rate relative to the other biological rates consistent across trophic levels. Finally, the time scale of the system is defined by setting the metabolic rate of the primary producer m_1 to unity. Thus, we can transform equations (5a) and (5b) into:

$$\frac{1}{D}\frac{dB_1^{(1)}}{dt} = B_1^{(1)}(\omega \frac{g}{D} - B_1^{(1)} - \gamma maB_2^{(1)}) + d_1(B_1^{(2)} - B_1^{(1)})$$
(8a)

$$\frac{1}{m^{i-1}D} \frac{dB_i^{(1)}}{dt} = \underbrace{B_i^{(1)} \left(-\frac{r}{D} - B_i^{(1)} + \gamma \epsilon a B_{i-1}^{(1)} - \gamma m a B_{i+1}^{(1)} \right)}_{\text{intra-patch dynamics}} + \underbrace{d_i \left(B_i^{(2)} - B_i^{(1)} \right)}_{\text{dispersal}} \tag{8b}$$

Thus, ϵa and ma defines the positive effect of the prey on its predator and the negative effect of the predator on its prey, respectively. These two synthetic parameters define the overall behaviour of the food chain and will be varied over the interval [0.1, 10] to consider a broad range of possible responses.

S1-2 Biomass at equilibrium when top predators populations are perfectly coupled

The system can be easily solved if we consider the total population of top predator instead of two populations connected by dispersal. Since the two populations are perfectly coupled by dispersal, top predator i biomass is constant across patches and we have $B_i^{(1)*} = B_i^{(2)*} = 0.5B_i^{tot*}$. Thus we have the following system at equilibrium for the two top species (the equations for the other species are the same as the symmetric case):

$$0 = -\frac{r}{D} - B_i^{*tot} + \varepsilon a \left(\gamma B_{i-1}^{*(1)} + B_{i-1}^{*(2)} \right)$$
(9a)

$$0 = -\frac{r}{D} - B_{i-1}^{*(1)} + \gamma \varepsilon a B_{i-2}^{*(1)} - \gamma m a \frac{B_i^{*tot}}{2}$$
(9b)

$$0 = -\frac{r}{D} - B_{i-1}^{*(2)} + \varepsilon a B_{i-2}^{*(2)} - ma \frac{B_i^{*tot}}{2}$$
(9c)

Which translates into the following matrix equation:

$$\begin{pmatrix} -1 & -\gamma ma & & & & & \\ \gamma \varepsilon a & -1 & -\gamma ma & & & & \\ & \ddots & \ddots & \ddots & & & & & \\ & & \gamma \varepsilon a & -1 & 0 & & & -\gamma ma/2 \\ & & & 0 & -1 & -ma & & & \\ & & & \varepsilon a & -1 & -ma & & \\ & & & & \varepsilon a & -1 & -ma/2 \\ & & & & \gamma \varepsilon a & & \varepsilon a & -1 & -ma/2 \\ \end{pmatrix} \begin{pmatrix} B_1^{*(1)} \\ B_2^{*(1)} \\ B_2^{*(1)} \\ \vdots \\ B_{i-1}^{*(1)} \\ B_i^{*(1)} \\ B_i^{*(1)} \\ \vdots \\ B_{i-1}^{*(1)} \\ B_i^{*(2)} \\ B_i^{*(2)} \\ \vdots \\ B_{i-1}^{*(2)} \\ B_i^{*(2)} \\ C_i \\ C_i$$

S1-3 Biomass at equilibrium when basal species populations are perfectly coupled

In the same way, we have:

$$0 = \frac{g}{D} - B_1^{*tot} - ma \left(\gamma B_2^{*(1)} + B_2^{*(2)} \right)$$
(11a)

$$0 = -\frac{r}{D} - B_2^{*(1)} + \gamma \varepsilon a \frac{B_1^{*tot}}{2} - \gamma m a B_3^{*(1)}$$
(11b)

$$0 = -\frac{r}{D} - B_2^{*(2)} + \varepsilon a \frac{B_1^{*tot}}{2} - maB_3^{*(2)}$$
(11c)

Which translates into the following matrix equation:

$$\begin{pmatrix}
-1 & -\gamma ma & -ma \\
\gamma \varepsilon a/2 & -1 & -\gamma ma & (0) \\
& \ddots & \ddots & \ddots \\
& \gamma \varepsilon a & -1 & 0 \\
\varepsilon a/2 & 0 & -1 & -ma \\
& (0) & \ddots & \ddots & \ddots \\
& \varepsilon a & -1
\end{pmatrix}\begin{pmatrix}
B_1^{*tot} \\
B_2^{*(1)} \\
\vdots \\
B_i^{*(1)} \\
B_i^{*(2)} \\
\vdots \\
B_i^{*(2)} \\
\vdots \\
B_i^{*(2)} \\
\vdots \\
B_i^{*(2)}
\end{pmatrix} = 0$$

$$(12)$$

534 S1-4 Linearisation of the system

The system of equations (1a) and (1b) can be linearised in the vicinity of equilibrium:

$$\frac{dB_i}{dt} = \underbrace{f_i(B_1^*, ..., B_S^*)}_{=0} + \sum_{j=1}^{S} \left(\frac{\partial f_i}{\partial B_j} \Big|_{B^*} (B_j - B_j^*) \right)$$
(13)

Thus, by setting $X_i = B_i - B_i^*$ the deviation from equilibrium, we have:

$$\frac{dX_i}{dt} = \sum_{j=1}^{S} J_{ij} X_j \tag{14}$$

Then, we can consider small perturbations defined by \overrightarrow{E} whose effects on \overrightarrow{X} are defined by the matrix T (Arnoldi et al., 2016). We get the linearised version of equation (2):

$$\frac{d\overrightarrow{X}}{dt} = J\overrightarrow{X} + T\overrightarrow{E} \tag{15}$$

The elements of \overrightarrow{E} are defined by stochastic perturbations $E_i = \sigma_i dW_i$ with σ_i their standard deviation and dW_i a white noise term with mean 0 and variance 1. In our model, each species i in each patch k can receive demographic perturbations scaling with the square root of their biomass at equilibrium. Thus, \overrightarrow{E} contains the white noise term $\sigma_i^{(k)} dW_i^{(k)}$ for each population of each species, T is a diagonal matrix whose terms are $\sqrt{B_i^{*(k)}}$ and the matrix product $T\overrightarrow{E}$ results in the product of the white noise and the biomass scaling as in equation (2) in the main text.

S1-5 Resolution of the Lyapunov equation

In the vicinity of equilibrium, the Lyapunov equation links the variance-covariance matrix V_E of the perturbation vector \overrightarrow{E} to the variance-covariance matrix C^* of species biomasses (see the appendix of Wang et al. (2015) for more details on the Lyaponov equation).

$$JC^* + C^*J^{\top} + TV_ET^{\top} = 0 {16}$$

The diagonal elements of V_E are equal to σ_i^2 (variance of the white noises) and the non-diagonal elements are equal to zero because perturbations are independent. \top is the transpose operator. C^* can be calculated using a Kronecker product (Nip et al., 2013). The Kronecker product of an $m \times n$ matrix A

and a $p \times q$ matrix B denoted $A \otimes B$ is the $mp \times nq$ block matrix given by:

$$A \otimes B = \begin{pmatrix} a_{11}B & \cdots & a_{1n}B \\ \vdots & \ddots & \vdots \\ a_{m1}B & \cdots & a_{mn}B \end{pmatrix}$$

We define C_s^* and $(TV_ET^\top)_s$ the vectors stacking the columns of C^* and TV_ET^\top respectively. Thus, equation (16) can be rewrite as:

$$(J \otimes I + I \otimes J)C_s^* = -(TV_E T^\top)_s$$

$$C_s^* = -(J \otimes I + I \otimes J)^{-1}(TV_E T^\top)_s$$
(17)

542 S1-6 Coefficient of variation and correlation

Our different metrics of stability can be easily computed from the elements of the variance-covariance matrix C^* defined by elements $w_{i^{(k)}j^{(\ell)}}$ that are the covariance between species i in patch k and species j in patch ℓ .

$$C^* = \begin{pmatrix} w_{1(1)1(1)} & \cdots & w_{1(1)S(1)} & \cdots & w_{1(n)1(n)} & \cdots & w_{1(n)S(n)} \\ \vdots & \ddots & \vdots & \ddots & \vdots & \ddots & \vdots \\ w_{S(1)1(1)} & \cdots & w_{S(1)S(1)} & \cdots & w_{S(n)1(n)} & \cdots & w_{S(n)S(n)} \\ \vdots & \vdots & \vdots & \ddots & \vdots & \vdots & \vdots \\ w_{1(n)1(1)} & \cdots & w_{1(n)S(1)} & \cdots & w_{1(n)1(n)} & \cdots & w_{1(n)S(n)} \\ \vdots & \ddots & \vdots & \ddots & \vdots & \ddots & \vdots \\ w_{S(n)1(1)} & \cdots & w_{S(n)S(1)} & \cdots & w_{S(n)1(n)} & \cdots & w_{S(n)S(n)} \end{pmatrix}$$

$$(18)$$

The temporal variability of the metacommunity is assessed with the coefficient of variation (CV) of biomass at different scales: **population scale** $CV_i^{(k)}$, which is the biomass CV of species i in patch k, **metapopulation scale** CV_i , which is the biomass CV of the total biomass of species i across patches and **metacommunity scale** CV_{MC} , which is the total biomass of the entire metacommunity (Wang and Loreau, 2014; Wang et al., 2019; Jarillo et al., 2022).

$$CV_{i}^{(k)} = \frac{\sqrt{w_{i^{(k)}i^{(k)}}}}{\mu_{i}^{(k)}} \qquad CV_{i} = \frac{\sqrt{\sum_{k\ell} w_{i^{(k)}j^{(\ell)}}}}{\sum_{k} B_{i}^{*(k)}} \qquad CV_{MC} = \frac{\sqrt{\sum_{ijk\ell} w_{i^{(k)}j^{(\ell)}}}}{\sum_{ik} B_{i}^{*(k)}}$$
(19)

$$C^* = \begin{bmatrix} w_{1^{(1)}1^{(1)}} \cdots w_{1^{(1)}S^{(1)}} \cdots w_{1^{(n)}1^{(n)}} \cdots w_{1^{(n)}S^{(n)}} \\ \vdots & \ddots & \vdots & \ddots & \vdots \\ w_{S^{(1)}1^{(1)}} \cdots w_{S^{(1)}S^{(1)}} \cdots w_{S^{(n)}1^{(n)}} \cdots w_{S^{(n)}S^{(n)}} \\ \vdots & \vdots & \ddots & \vdots & \vdots \\ w_{1^{(n)}1^{(1)}} \cdots w_{1^{(n)}S^{(1)}} \cdots w_{1^{(n)}1^{(n)}} \cdots w_{1^{(n)}S^{(n)}} \\ \vdots & \ddots & \vdots & \ddots & \vdots \\ w_{S^{(n)}1^{(1)}} \cdots w_{S^{(n)}S^{(1)}} \cdots w_{S^{(n)}1^{(n)}} \cdots w_{S^{(n)}S^{(n)}} \end{bmatrix} CV_{MC}$$

Figure S1-1: Elements of the variance-covariance matrix C^* used to compute the biomass CV at different scales defined in equation (19).

The correlation matrix R^* of the system, whose elements $\rho_{i^{(k)}j^{(\ell)}}$ are defined by:

$$\rho_{i^{(k)}j^{(\ell)}} = \frac{w_{i^{(k)}j^{(\ell)}}}{\sqrt{w_{i^{(k)}i^{(k)}}}\sqrt{w_{j^{(\ell)}j^{(\ell)}}}}$$
(20)

543 S1-7 Asymptotic resilience

In addition to the response to stochastic perturbations, we consider asymptotic resilience to measure
the long term return time of the metacommunity. Asymptotic resilience is measured by the opposite of
the real part of the dominant eigenvalue λ_{dom} of Jacobian matrix J ($\Re(\lambda_{dom})$). Since the dominant
eigenvalue is the eigenvalue with the largest real part and we only consider ecosystems at equilibrium
(i.e. all eigenvalues have negative real parts), the lower the real part of the dominant eigenvalue, the
faster the long term return time.

Moreover, we can assess the influence of each species on asymptotic resilience by comparing the absolute
value of the real part of each element e_i of the dominant eigenvector (E_{dom}). Because e_i is the contribution
of species i to E_{dom} , $|e_i|/\sum_{j=1}^n |e_j|$ is the relative weight of species i in the dynamics of long term return
to equilibrium (with n the number of populations in the metacommunity).

References

Arnoldi, J.-F., Loreau, M., & Haegeman, B. (2016). Resilience, reactivity and variability: A mathematical comparison of
ecological stability measures. *Journal of Theoretical Biology*, 389, 47–59. https://doi.org/10.1016/j.jtbi.2015.10.
012
Barbier, M., & Loreau, M. (2019). Pyramids and cascades: A synthesis of food chain functioning and stability. *Ecology*Letters, 22(2), 405–419. https://doi.org/10.1111/ele.13196

```
Jarillo, J., Cao-García, F. J., & De Laender, F. (2022). Spatial and ecological scaling of stability in spatial community
networks. Frontiers in Ecology and Evolution, 10, 861537. https://doi.org/10.3389/fevo.2022.861537
```

- Nip, M., Hespanha, J. P., & Khammash, M. (2013). Direct numerical solution of algebraic Lyapunov equations for largescale systems using Quantized Tensor Trains. 52nd IEEE Conference on Decision and Control, 1950–1957. https://doi.org/10.1109/CDC.2013.6760167
- Wang, S., Haegeman, B., & Loreau, M. (2015). Dispersal and metapopulation stability. PeerJ, 3, e1295. https://doi.org/
 10.7717/peerj.1295
- Wang, S., Lamy, T., Hallett, L. M., & Loreau, M. (2019). Stability and synchrony across ecological hierarchies in heterogeneous metacommunities: Linking theory to data. *Ecography*, 0(0). https://doi.org/10.1111/ecog.04290
- Wang, S., & Loreau, M. (2014). Ecosystem stability in space: \textalpha, \textbeta and \textgamma variability. Ecology
 Letters, 17(8), 891–901. https://doi.org/10.1111/ele.12292

S2 Complementary results

S2-1 General description of parameters

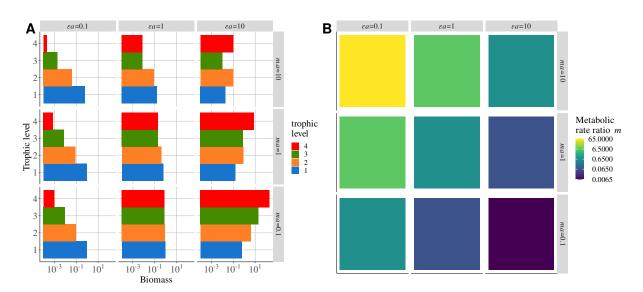


Figure S2-1: Distribution of parameters and their effects on an isolated food chain. A) Biomass distribution depending on the positive effect of prey on predator ϵa and the negative effect of predator on prey ma. B) Value of the ratio of predator to prey metabolic rate $m = m_{i+1}/m_i$ for each combination of ϵa and ma.

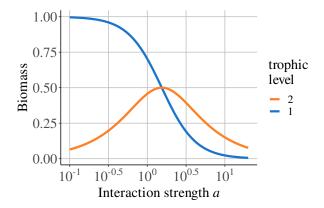


Figure S2-2: Distribution of biomass in an isolated predator-prey system (without dispersal) depending on interaction strength a relative to self-regulation. Increasing the asymmetry of the interaction strength γ is equivalent to increasing a (m = 0.65).

Increasing the interaction strength a relative to self-regulation decreases the biomass of prey because of the increased mortality due to predation (Figure S2-2). However, the biomass of predators follows a hump-shaped relationship: it first increases due to the increased resource consumption and then decreases because of prey overexploitation.

S2-2 Dispersal of predators and perturbation of prey

S2-2-1 Conditions of coexistence

Asymmetry and dispersal lead to competition, apparent competition and source-sink dynamics that can rescue or drive local populations to extinction. Therefore, we consider limit cases in which dispersal is infinite (well mixed populations across the metacommunity) to analytically calculate biomasses at equilibrium and determine the range of values of ω and γ enabling the coexistence of all populations of each species.

We consider the total biomass of predators $B_2^{tot} = B_2^{(1)} + B_2^{(2)}$ and because the very high dispersal of predators equally distributes its biomass among the two patches, we have $B_2^{*tot} = 2 \times B_2^{*(1)}$. Then, we can define the system:

$$\frac{dB_1^{(1)}}{dt} = DB_1^{(1)} \left(\frac{\omega g}{D} - B_1^{(1)} - \gamma ma \frac{B_2^{tot}}{2} \right)$$
 (21a)

$$\frac{dB_1^{(2)}}{dt} = DB_1^{(2)} \left(\frac{g}{D} - B_1^{(2)} - ma\frac{B_2^{tot}}{2}\right) \tag{21b}$$

$$\frac{dB_2^{tot}}{dt} = mD \frac{B_2^{tot}}{2} \left(-\frac{r}{D} - B_2^{tot} + \varepsilon a(\gamma B_1^{(1)} + B_1^{(2)}) \right)$$
(21c)

Since r = 0, we remove it from the equations for the sake of simplicity. We define $\lambda = \varepsilon ma^2$, which is the intensity of top-down control defined by Barbier and Loreau (2019). At equilibrium, we obtain:

$$B_1^{(1)*} = \frac{g}{D} \left(\frac{2\omega + \omega\lambda - \gamma\lambda}{2 + \lambda(\gamma^2 + 1)} \right) \tag{22a}$$

$$B_1^{(2)*} = \frac{g}{D} \left(\frac{\lambda \gamma^2 - \omega \lambda \gamma + 2}{2 + \lambda (\gamma^2 + 1)} \right)$$
 (22b)

$$B_2^{tot*} = \frac{2\varepsilon ag(1+\omega\gamma)}{D(2+\varepsilon a^2 m(\gamma^2+1))}$$
 (22c)

Prey biomass in patch #1 $B_1^{*(1)}$ is positive only if:

$$\gamma < \frac{\omega(2+\lambda)}{\lambda} \xrightarrow[\lambda \to \infty]{} \omega \tag{23}$$

Prey biomass in patch #2 $B_1^{*(2)}$ is positive if $f(\gamma) = \lambda \gamma^2 - \omega \lambda \gamma + 2 > 0$. f opens upwards: thus, if

 $\omega < \sqrt{8/\lambda}$, f has no roots and is always positive. Otherwise, $B_1^{(2)}$ is positive if:

$$\gamma > \frac{\lambda \omega + \sqrt{\lambda(\lambda \omega^2 - 8)}}{2\lambda} \xrightarrow[\lambda \to \infty]{} \omega \quad \text{if} \quad \omega > \sqrt{\frac{8}{\lambda}} \quad \text{or}$$
 (24a)

$$\gamma < \frac{\lambda\omega - \sqrt{\lambda(\lambda\omega^2 - 8)}}{2\lambda} \xrightarrow[\lambda \to \infty]{} 0 \quad \text{if} \quad \omega > \sqrt{\frac{8}{\lambda}}$$
 (24b)

Predators B_2 thrive in each patch for all values of ω and γ (Figure S2-3C). Hence, coexistence is ensured for all values of top-down control λ , asymmetry of resource supply ω and asymmetry of interaction strength γ only if $\gamma = \omega$. In the main text, we always consider $\gamma = \omega$ (Figure S2-3D), but their independent effects are detailed in the following.

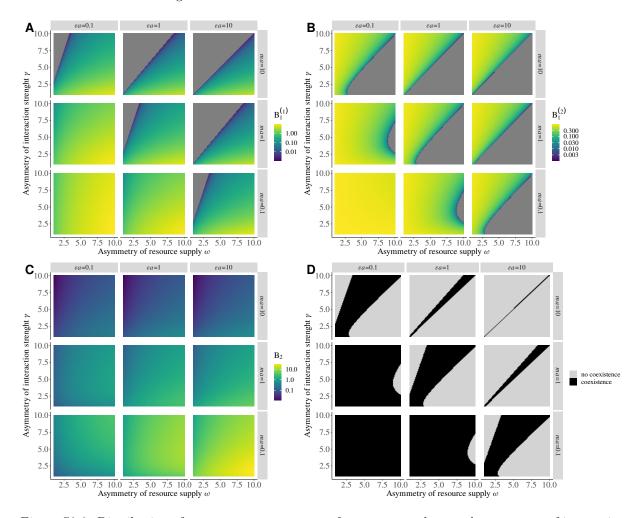


Figure S2-3: Distribution of parameters, asymmetry of resource supply ω and asymmetry of interaction strength γ , leading to the coexistence of predator and prey in each patch. Only predators are able to disperse at an infinite rate (well-mixed predator populations). This distribution is assessed for different values of the positive effect of prey on predator ϵa and negative effect of predator on prey ma. A) Biomass of prey in patch #1 $B_1^{*(1)}$ and B) in patch #2 $B_1^{*(2)}$. C) Biomass of predator in patches #1 and #2 ($B_2^{*(1)} = B_2^{*(2)}$ because predator populations are well mixed). D) Coexistence of predator and prey in each patch.

S88 S2-2-2 Nontransitivity of correlation

To explain the correlation between prey populations, we can track the transmission of perturbations in
the metacommunity. Increasing the asymmetry of interaction strength γ tends to decorrelate predator and
prey dynamics within each patch (Figure S2-4A). When prey are perturbed in patch #1, the dynamics
of predator and prey biomass are correlated in patch #1 and anticorrelated in patch #2 due to the
bottom-up and top-down transmissions of perturbations, respectively. Although we would expect the
two populations of prey to be anti-correlated according to the mechanism described by Quévreux et al.
(2021) (see Figure S2-25 in the following), we actually observe a weak correlation of these two populations
(Figure S2-4B). In the same way, the intermediate correlation and anti-correlation of prey populations
when prey are perturbed in patch #2 do not explain the strong anti-correlation of prey populations
(Figure S2-4C). Therefore, other mechanisms are acting in our system.

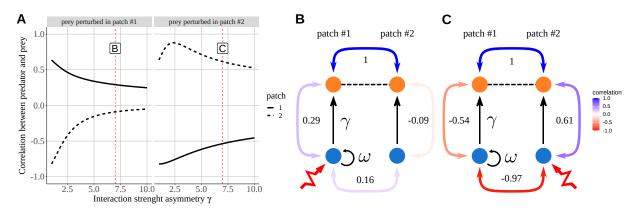


Figure S2-4: Correlation between predator and prey with each patch ($\varepsilon a=1,\ ma=1,\ d_2=10^6,\ \omega=\gamma$). A) Correlation in each patch depending on asymmetry of interaction strength γ when prey in patch #1 (left panel) or in patch #2 (right panel) are perturbed. Labels and vertical dashed lines represent the correlation values used in panels B and C. B) Schematic representation of correlations when $\gamma=7$. Coloured double arrows and their associated number (see also the colour scale) represent the correlations between populations. Prey are perturbed in patch #1. C) Prey are perturbed in patch #2.

S2-2-3 Complete effects of εa and ma

The stability patterns observed in Figures 3 to 5 in the main text are also observed for a wide range of ecological and physiological parameters aggregated into the positive effect of prey on predators εa and the negative effect on predators on prey ma. Therefore, our results are robust and the identified mechanisms are specific to a particular combination of parameters.

The response of the asymptotic resilience to the asymmetry of interaction strength (Figure S2-8A) is not similar to the results of Rooney et al. (2006). Indeed, we do not observe minimum of resilience for $\gamma = 1$ for all combinations of εa and ma. The variations in asymptotic resilience depend on the

relative contribution of each population of each species (Figure S2-8B), which is governed by the biomass distribution of each species among patches (Figure S2-5A) and the ratio of predator to prey metabolic rate ratio m (Figure S2-1B). As demonstrated by Haegeman et al. (2016) and Arnoldi et al. (2018), rare species control the long term recovery to perturbations of the metacommunity (*i.e.*, the asymptotic resilience), as well as species with a slow pace of life (*i.e.*, a slow metabolism).

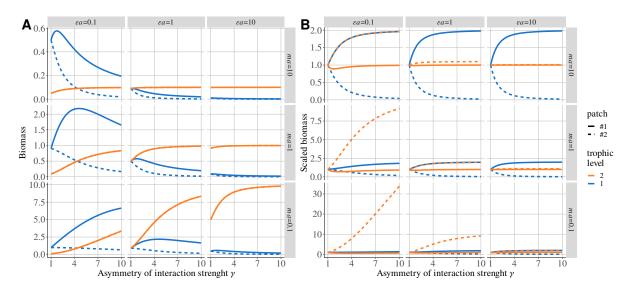


Figure S2-5: Biomass distribution of each species in each patch depending on asymmetry of interaction strength γ , positive effect of prey on predator εa and negative effect of predator on prey ma. A) Biomass distribution. B) Biomass scaled by the biomass in the metacommunity without dispersal $(d_2 = 0)$.

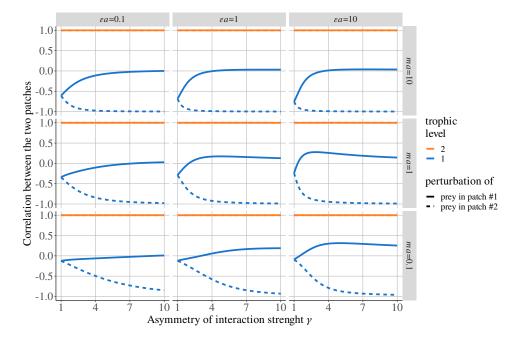


Figure S2-6: Correlation between populations depending on asymmetry in interaction strength γ when predators disperse and prey are perturbed in patch #1 or #2. Predators have a high scaled dispersal rate $(d_2 = 10^6)$, which strongly couples their two populations $(\gamma = \omega)$.

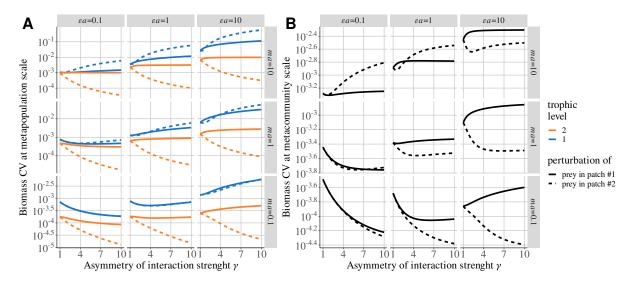


Figure S2-7: Biomass CV at different scales depending on asymmetry of interaction strength γ , positive effect of prey on predator εa and negative effect of predator on prey ma ($d_2 = 10^6$ and $\omega = \gamma$). A) Biomass CV of the population of each species in each patch. B) CV of the total biomass of each species.

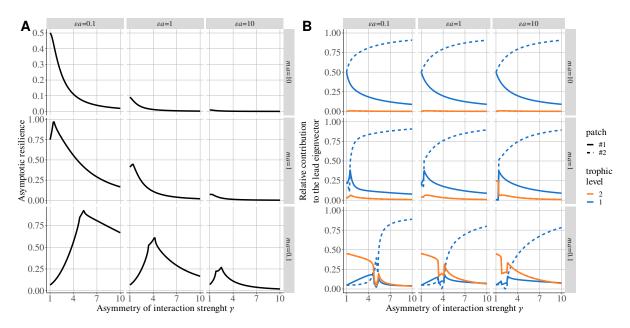


Figure S2-8: Linear stability depending on asymmetry of interaction strength γ , positive effect of prey on predator εa and negative effect of predator on prey ma. A) Asymptotic resilience (real part of the dominant eigenvalue of the Jacobian matrix) B) Contribution of the populations of each species to the dominant eigenvector.

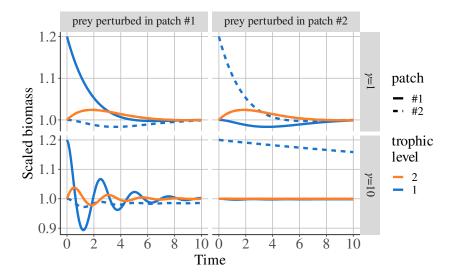


Figure S2-9: Time series of biomasses rescaled by their value at equilibrium after an increase in prey biomass by 20% in patch #1 (left panel) or patch #2 (right panel) for two values of interaction strength asymmetry ($\gamma = 1$ or $\gamma = 10$, $\varepsilon a = 1$, ma = 1, $d_2 = 10^6$ and $\omega = \gamma$).

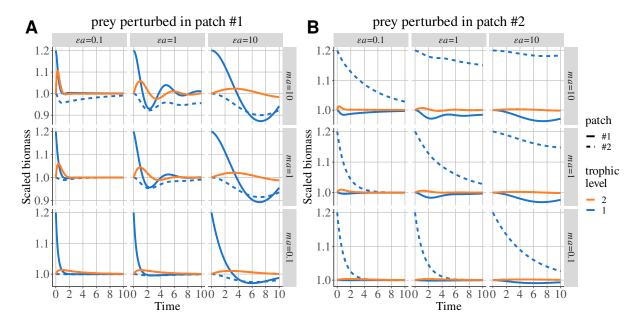


Figure S2-10: Time series of biomasses rescaled by their value at equilibrium after an increase in prey biomass by 20% in patch #1 (left panel) or patch #2 (right panel) depending on asymmetry of interaction strength γ , positive effect of prey on predator ϵa and negative effect of predator on prey ma ($\gamma = 3$, $d_2 = 10^6$ and $\omega = \gamma$).

S2-2-4 Effect of asymmetry of resource supply ω

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According to Figure S2-3, we set $\omega = \gamma$ to ensure the coexistence of prey and predators in each patch for all combinations of εa and ma. Varying the asymmetry of resource supply ω does not qualitatively alter the response of biomass (Figure S2-11), correlation (Figure S2-12A) and biomass CV (Figure S2-12B) to the variations in the asymmetry of interaction strength γ .

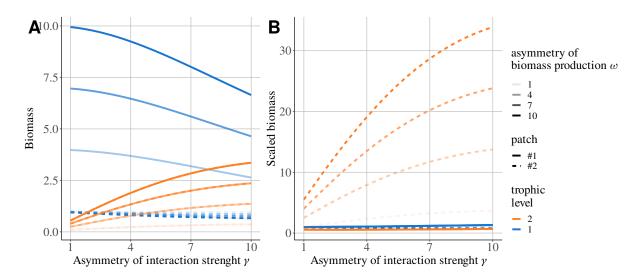


Figure S2-11: Biomass distribution of each species in each patch depending on asymmetry of interaction strength γ and biomass production ω ($\epsilon a = 0.1$, ma = 0.1 and $d_2 = 10^6$). A) Biomass distribution. B) Biomass scaled by the biomass in the metacommunity without dispersal ($d_2 = 0$).

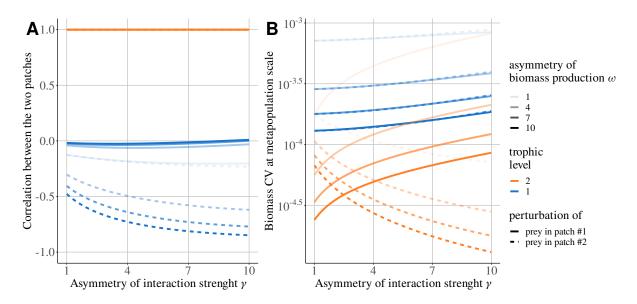


Figure S2-12: Stability depending on asymmetry of interaction strength γ and biomass production ω ($\epsilon a = 0.1$, ma = 0.1 and $d_2 = 10^6$). A) Correlation between populations. B) Biomass CV of the population of each species in each patch.

S2-2-5 Effect of perturbation of predators

The perturbation of predators leads to the same response regardless of the perturbed patch because
the very high dispersal of predators perfectly synchronises their population dynamics. The asymmetry
of interaction strength leads to different dynamics in each patch that decreases the correlation of prey
dynamics (Figure S2-13) and stabilises predator dynamics by decreasing their biomass CV (Figure S214A).

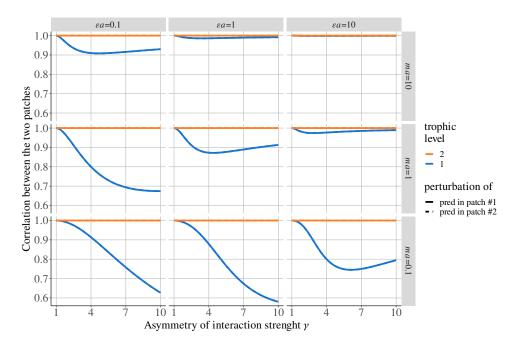


Figure S2-13: Correlation between populations depending on asymmetry of interaction strength γ , positive effect of prey on predator εa and negative effect of predator on prey ma. Predators disperse and are perturbed in patch #1 or patch #2 ($d_2 = 10^6$ and $\omega = \gamma$).

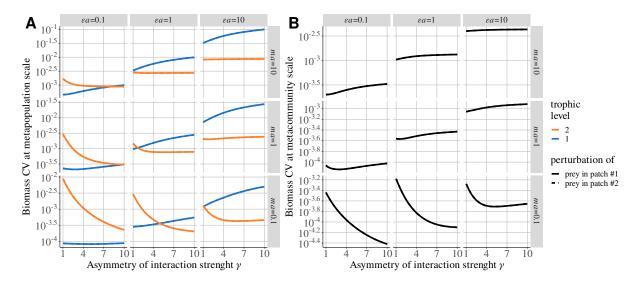


Figure S2-14: Biomass CV at different scales depending on asymmetry of interaction strength γ , positive effect of prey on predator εa and negative effect of predator on prey ma. Predators disperse and are perturbed in patch #1 or patch #2 ($d_2 = 10^6$ and $\omega = \gamma$). A) CV of the total biomass of each species. B) CV of the total biomass of the metacommunity.

S2-2-6 Effect of food chain length

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Here, we consider three trophic levels to extend our results to metacommunities with longer food chain lengths. In this setup, only top predators (species 3) are able to disperse, and basal species (species 1) receive stochastic perturbations. γ also has the same value across trophic levels. We observe a similar response to the case with two trophic levels for the correlations of the dynamics of the biomass of species 2 and 3 (Figure S2-15) as well as for biomass CV (Figure S2-16). However, the response of species 1

 $_{629}$ in completely different. Therefore, the mechanisms described in the main text are only acting for the

dispersing species and the species directly interacting with it.

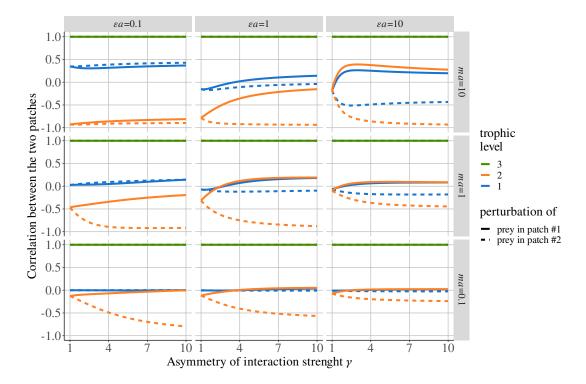


Figure S2-15: Correlation between populations in a three trophic level food chain depending on the asymmetry of interaction strength γ , positive effect of prey on predator εa and negative effect of predator on prey ma ($d_3=10^6$ and $\omega=1$). Top predators disperse, and the basal species is perturbed in patch #1 or #2.

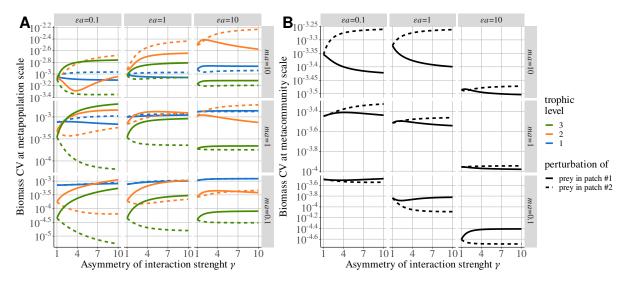


Figure S2-16: Biomass CV in a three trophic level food chain at different scales depending on asymmetry of interaction strength γ , positive effect of prey on predator εa and negative effect of predator on prey $ma~(d_3=10^6~{\rm and}~\omega=1)$. **A)** Biomass CV of the population of each species in each patch. **B)** CV of the total biomass of each species.

S2-3 Dispersal of prey and perturbation of predators

In this section, we consider a setup mirroring the metacommunity model described in the main text. Here, only prey are able to disperse at a very high rate $(d_1 = 10^6)$, and predators receive stochastic perturbations. We also set $\gamma = \omega$ to be consistent with the results in the main text. In the following, we find the same responses to the asymmetry of interaction strength γ , which demonstrates that the mechanisms described in the main text are not conditioned by the trophic position of the dispersing species.

638 S2-3-1 Conditions of coexistence

When only prey are able to disperse, all populations of each species have positive biomasses for all values of ω , γ , εa and ma (Figure S2-17).

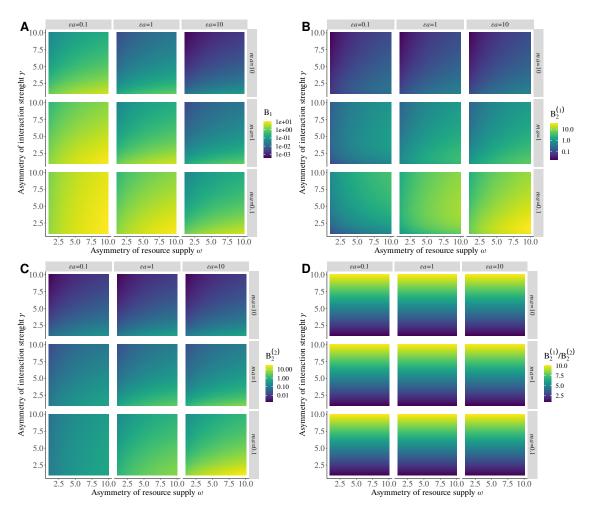


Figure S2-17: Distribution of parameters, asymmetry of resource supply ω and asymmetry of interaction strength γ , leading to the coexistence of predator and prey in each patch. Only prey are able to disperse at an infinite rate (well mixed prey populations). This distribution is assessed for different values of the positive effect of prey on predator εa and the negative effect of predator on prey ma. The product of εa and ma is the strength of top-down control λ ($\lambda = \varepsilon a^2 m$, see Barbier and Loreau (2019)). A) Biomass of prey in patches #1 and #2 ($B_1^{*(1)} = B_1^{*(2)}$ because prey populations are well mixed). B) Biomass of predator in patch #1 ($B_2^{*(1)}$ C) in patch #2 $B_2^{*(2)}$.

S2-3-2 Complete effects of εa and ma

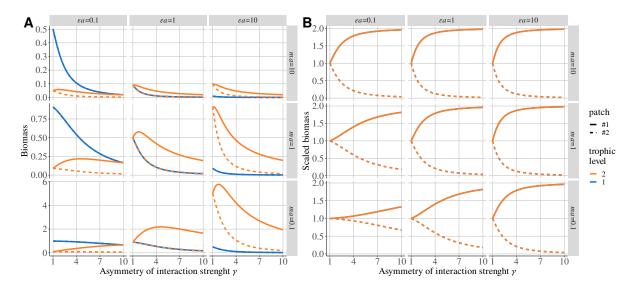


Figure S2-18: Biomass distribution of each species in each patch depending on asymmetry of interaction strength γ , positive effect of prey on predator εa and negative effect of predator on prey ma. A) Biomass distribution. B) Biomass scaled by the biomass in the metacommunity without dispersal $(d_1 = 0)$. Prey and predator curves perfectly overlap

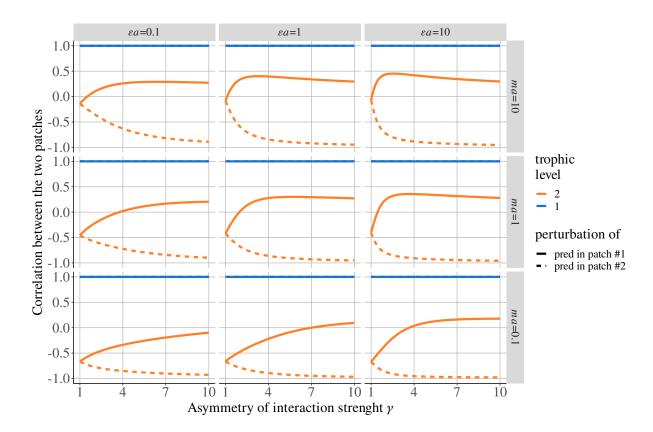


Figure S2-19: Correlation of population when prey disperse and predators are perturbed in patch #1 or in patch #2 depending on asymmetry of interaction strength γ , positive effect of prey on predator εa and negative effect of predator on prey ma.

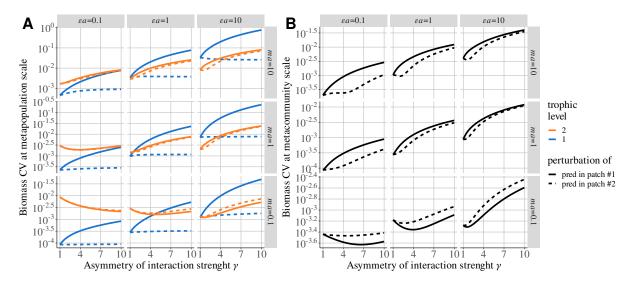


Figure S2-20: Biomass CV at different scales depending on asymmetry of interaction strength γ , positive effect of prey on predator εa and negative effect of predator on prey ma ($d_1 = 10^6$ and $\omega = \gamma$). A) Biomass CV of the population of each species in each patch. B) CV of the total biomass of each species.

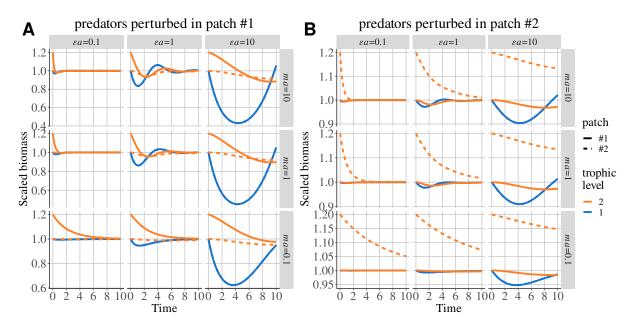


Figure S2-21: Time series after pulse perturbation of predators in patch #1 or in patch #2. Biomasses are scaled by their value at equilibrium, and dispersal is high ($\gamma = 3$ and $d = 10^6$).

S2-4 Correlated environmental perturbations

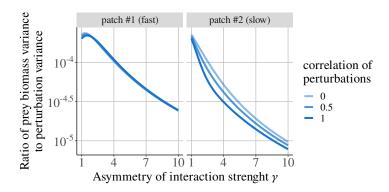


Figure S2-22: Ratio of the variance of prey biomass to the variance of environmental perturbations $(\text{Var}(B_i)/\sigma_i)$ in each patch depending on asymmetry of interaction strength γ . Each prey population receives spatially correlated environmental perturbation (colour gradient scale) scaling with equilibrium biomass B_i^* .

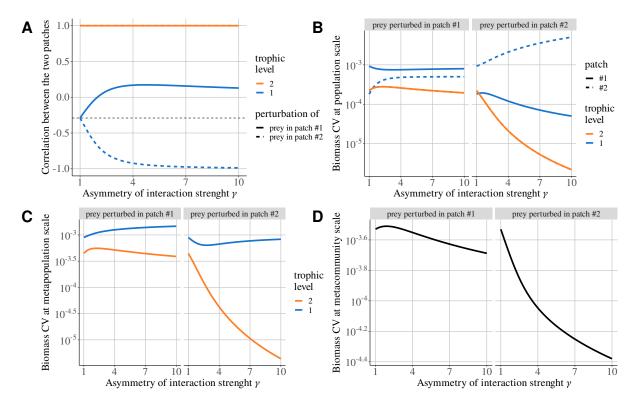


Figure S2-23: Stability at different scales depending on asymmetry of interaction strength γ when predators disperse and prey are perturbed in patch #1 or #2 with environmental perturbations ($\varepsilon a=1$, $ma=1, \omega=\gamma$). A) Spatial correlation between the populations of each species. B) Biomass CV at the population scale. C) Biomass CV at the metapopulation scale (CV of the total biomass of each species). D) Biomass CV at the metacommunity scale (CV of the total biomass of the metacommunity.

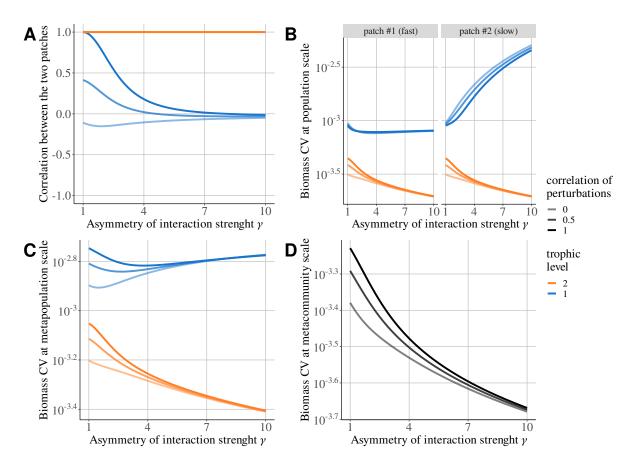


Figure S2-24: Stability at different scales depending on asymmetry of interaction strength γ when predators disperse and prey are perturbed by a spatially correlated environmental perturbations (colour gradient scale) ($\varepsilon a = 1, \ ma = 1, \ \omega = \gamma$). A) Spatial correlation between the populations of each species. B) Biomass CV at the population scale. C) Biomass CV at the metapopulation scale (CV of the total biomass of each species). D) Biomass CV at the metacommunity scale (CV of the total biomass of the metacommunity.

Here, we consider the same metacommunity as in the main text (see Figure 1 in the main text), but prey receive spatially correlated environmental perturbations. Environmental perturbations correspond to the synchronous response of all individuals of the same population to an environmental factor (e.g., drought), and they scale with equilibrium biomass B_i^* (see the supporting information of Quévreux et al. (2021) for the demonstration). In our metacommunity, we also consider that environmental perturbations are spatially correlated since it is reasonable to assume that different populations of the same species will respond in a similar way to environmental perturbations.

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The effect of a perturbation on a population can be assessed by the ratio of the variance of biomass to the variance of perturbations $Var(B_i)/\sigma_i$. As demonstrated by Arnoldi et al. (2019), environmental perturbations affect abundant populations the most, which is the prey population in the fast patch in our case (Figure S2-22). Therefore, we can approximate the effect of environmental perturbations by the effect of the perturbation of prey in the fast patch. Perturbing a single population with demographic or environmental perturbations leads to exactly the same qualitative results (Figure S2-23 and Figure 3 and Figure 4 in the main text), and only the CV values change because of the different biomass scaling.

Increasing the correlation of perturbations increases the correlation of the dynamics of prey populations (Figure S2-24A) because of the Moran effect (Moran, 1953). The increase in synchrony explains the 658 increase in the biomass CV observed at each scale for all species (Figure S2-24B-D), except for prey in 659 the slow patch (Figure S2-24B). The Moran effect is particularly strong at low asymmetry ($\gamma < 4$), but 660 once asymmetry is high enough, two mechanisms disrupt the Moran effect. First, when asymmetry is high, the dynamics in each patch become so different that correlated perturbations are not able to generate 662 similar responses. Second, because of the discrepancy in the distribution of prey biomass among the two patches, environmental perturbations mostly affected prey in the fast patch (Figure S2-22). Therefore, with increasing asymmetry of interaction strength γ , the response of the metacommunity to correlated environmental perturbations converges towards the response of a metacommunity in which only prey in 666 the fast patch are perturbed.

S2-5 Reminder of the symmetric case

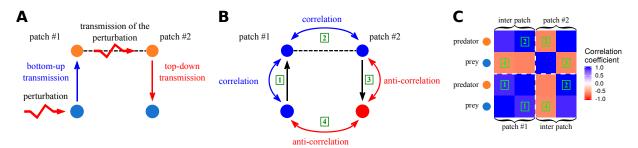


Figure S2-25: Summary of the main results from Quévreux et al. (2021), who considered a two patch predator-prey metacommunity with passive dispersal. In the setup presented in $\bf A$), prey are perturbed in patch #1 and only predators are able to disperse. Thus, perturbations have a bottom-up transmission in patch #1 (*i.e.* transmission from lower to upper trophic levels). This leads to the temporal correlation of the biomass dynamics of predators and prey in patch #1 showed in $\bf B$)(1) because if a perturbation increases the biomass of prey, it also increases the biomass of predators due to the vertical transfer of biomass. The passive dispersal of predators transmits the perturbations and spatially correlate their populations as shown in $\bf B$)(2). Then, perturbations have a top-down transmission in patch #2 (*i.e.* transmission from upper to lower trophic levels). This leads to the temporal anticorrelation (negative coefficient of correlation) of the biomass dynamics of predators and prey in patch #2 showed in $\bf B$)(3) because if a perturbation increases the biomass of predators, it decreases the biomass of prey due to the negative effect of predators on prey. Eventually, prey populations are spatially anticorrelated, as shown in $\bf B$)(4). Hence, by knowing which species is perturbed, which species disperses and how perturbations propagate within a food chain, Quévreux et al. (2021) were able to explain the spatial synchrony of the various populations of a metacommunity, summariesed by the correlation matrix in $\bf C$).

References

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Arnoldi, J.-F., Bideault, A., Loreau, M., & Haegeman, B. (2018). How ecosystems recover from pulse perturbations: A
670
              theory of short- to long-term responses. Journal of Theoretical Biology, 436, 79-92. https://doi.org/10.1016/j.
671
              jtbi.2017.10.003
     Arnoldi, J.-F., Loreau, M., & Haegeman, B. (2019). The inherent multidimensionality of temporal variability: How common
673
              and rare species shape stability patterns (J. Chase, Ed.). Ecology Letters, 22(10), 1557-1567. https://doi.org/10.
674
              1111/ele.13345
675
     Barbier, M., & Loreau, M. (2019). Pyramids and cascades: A synthesis of food chain functioning and stability. Ecology
              Letters, 22(2), 405-419. https://doi.org/10.1111/ele.13196
677
     Haegeman, B., Arnoldi, J.-F., Wang, S., de Mazancourt, C., Montoya, J. M., & Loreau, M. (2016). Resilience, invariability,
678
              and ecological stability across levels of organization. bioRxiv. https://doi.org/10.1101/085852
679
     Moran, P. (1953). The statistical analysis of the Canadian Lynx cycle. Australian Journal of Zoology, 1(3), 291. https:
680
              //doi.org/10.1071/ZO9530291
681
     Quévreux, P., Barbier, M., & Loreau, M. (2021). Synchrony and perturbation transmission in trophic metacommunities.
682
              The American Naturalist, 714131. https://doi.org/10.1086/714131
683
     Rooney, N., McCann, K. S., Gellner, G., & Moore, J. C. (2006). Structural asymmetry and the stability of diverse food
              webs. Nature, 442 (7100), 265-269. https://doi.org/10.1038/nature04887
685
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