



**HAL**  
open science

## **Determinants of Legacy Persistent Organic Pollutant Levels in the European Pond Turtle (*Emys orbicularis*) in the Camargue Wetland, France**

Louisiane Burkart, Anthony Olivier, Olivier Lourdais, Marion Vittecoq, Gabriel Blouin-demers, Fabrice Alliot, Clément Le Gac, Nicolas Martin, Aurélie Goutte

### ► To cite this version:

Louisiane Burkart, Anthony Olivier, Olivier Lourdais, Marion Vittecoq, Gabriel Blouin-demers, et al.. Determinants of Legacy Persistent Organic Pollutant Levels in the European Pond Turtle (*Emys orbicularis*) in the Camargue Wetland, France. *Environmental Toxicology and Chemistry*, 2021, 40 (8), pp.2261 - 2268. 10.1002/etc.5077 . hal-03822164

**HAL Id: hal-03822164**

**<https://hal.science/hal-03822164v1>**

Submitted on 25 Oct 2022

**HAL** is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

1 Running head: Low levels of POPs in *Emys orbicularis* in the Camargue

2

3 **Determinants of legacy POPs levels in the European**  
4 **pond turtle (*Emys orbicularis*) in the Camargue**  
5 **wetland, France**

6

7 LOUISIANE BURKART<sup>1,2</sup>, ANTHONY OLIVIER<sup>1</sup>, OLIVIER LOURDAIS<sup>3,4</sup>, MARION  
8 VITTECOQ<sup>1,5</sup>, GABRIEL BLOUIN-DEMERS<sup>6</sup>, FABRICE ALLIOT<sup>2</sup>, CLÉMENT LE  
9 GAC<sup>1</sup>, NICOLAS MARTIN<sup>1</sup>, AURÉLIE GOUTTE <sup>2</sup>

10

11 <sup>1</sup> La Tour du Valat, Research institute for the conservation of Mediterranean Wetlands,  
12 Arles, France

13 <sup>2</sup> École Pratique des Hautes Études (EPHE), PSL Research University, UMR 7619  
14 Sorbonne University, Paris, France

15 <sup>3</sup> Centre d'études biologiques de Chizé, CNRS, F-79360, Villiers en Bois, France

16 <sup>4</sup> School of Life Sciences, Arizona State University, Tempe, AZ 85287-4501, USA

17 <sup>5</sup> MIGEVEC Laboratoire maladies infectieuses & vecteurs: écologie, génétique  
18 évolution et contrôle, UMR (CNRS/IRD 224/UM), Montpellier, France

19 <sup>6</sup> Department of Biology, University of Ottawa, Ottawa, Ontario, Canada

20

21 Correspondence:

22

23 E-mail: aurelie.goutte@ephe.psl.eu

24

25

26 **Highlights**

27

28 • We analyzed POPs levels in 174 European pond turtles in the Camargue  
29 wetland.

30 • Only 3 POPs were detected in more than 10% of individual turtles.

31 • Sex, date, and mass did not explain the occurrence and levels of POPs in  
32 turtles.

33 • The occurrence and levels of PCB 153 increased with turtle age.

34 • Turtles from the two study sites differed in POP burdens, possibly due to  
35 different hydraulic system.

36

37 **Abstract**

38 Many banned persistent organic pollutants (POPs) remain for decades in the aquatic  
39 environment and can have harmful effects on long-lived predators because of their  
40 high bioaccumulation and biomagnification potentials. We investigated the  
41 occurrence and levels of 18 polychlorinated biphenyls (PCBs) and 16 organochlorine  
42 pesticides (OCPs) in European pond turtles (n = 174) from April to July 2018 in the  
43 Camargue wetland, France. Although the Camargue was highly contaminated in the  
44 last decades, plasma occurrence and levels of POPs were very low: we detected only  
45 3 of the 34 compounds we analyzed in more than 10% of the turtles. POP burdens  
46 did not differ between males and females and was uncorrelated with sampling date  
47 and body mass. We observed differences in POP burdens between turtles from the  
48 two sampling sites. One possible explanation is the agricultural hydraulic system:  
49 plasma occurrence and levels were higher for PCB-52 and HCB in turtles captured in  
50 drainage channels, but lower for PCB-153 at the site that receives irrigation. Finally,  
51 the occurrence and levels of PCB 153 in turtles increased with age, likely because of  
52 bioaccumulation and much higher exposure 20-30 years ago than now.

53

54 Key words: reptiles, PCB, organochlorine pesticides, delta wetlands

55

56

## 57 **1. Introduction**

58 In Europe, the production and use of persistent organic pollutants (POPs) are  
59 prohibited or severely restricted since 2004 through the Stockholm Convention  
60 (<http://www.pops.int/>), but POPs still cause environmental concerns.  
61 Polychlorinated biphenyls (PCBs) and organochlorine pesticides (OCPs) remain in  
62 the environment for decades because of their slow biodegradation and improper  
63 handling of contaminated wastes. Levels of POPs in animals increase with age  
64 (Vives et al., 2005; Binnington & Wania, 2014) and trophic level (Goutte *et al.*, 2020).  
65 Exposure to POPs causes a wide range of adverse effects, including neurotoxicity,  
66 endocrine disruption, immune dysfunction, reproductive impairment, and  
67 developmental abnormalities, which may ultimately compromise survival and  
68 reproductive output and lead to population declines of wild vertebrates (Goutte *et*  
69 *al.*, 2014, 2015; Salice *et al.*, 2014).

70 Pollution, and especially industrial and agricultural discharge, is one of the  
71 major threats to freshwater ecosystems (Holt, 2000; Dudgeon *et al.*, 2006). Wetlands  
72 support an extremely rich biodiversity, but are among the most transformed and  
73 threatened ecosystems of the world (Revenga *et al.*, 2005), experiencing rates of  
74 population declines and species extinctions far higher than those in forests,  
75 grasslands, and coastal ecosystems (Dudgeon *et al.*, 2006). Deltas are often exposed to  
76 high water pollution from intensive agriculture (Kuenzer & Renaud, 2012) and from  
77 discharge of contaminants in upstream waters. The Camargue, in the Rhône River  
78 delta, is the largest wetland in France and is of international importance under the  
79 Ramsar Convention (<https://www.ramsar.org/>), but it is heavily impacted by

80 human activities (Cheiron *et al.*, 2018). The contamination of water bodies by  
81 agricultural, industrial, and urban discharges has been well studied over the past  
82 decades (Comoretto *et al.*, 2007). In particular, levels of PCBs in sediments of the  
83 Rhône River increase from upstream to downstream, reaching 417 µg/kg dw for the  
84 sum of 7 PCB congeners (28, 52, 101, 118, 138, 153, and 180, Mourier *et al.* 2014).  
85 Mean flux over the 2011–2016 period averaged 14 kg/year for PCB 180 in suspended  
86 particulate matter at the outlet of the Rhône River basin (Poulier *et al.*, 2019).  
87 Previous studies conducted in the Camargue confirmed that birds (Berny *et al.*, 2002)  
88 and fish (Roche *et al.*, 2002, 2003, 2009a) are exposed to OCPs and PCBs.

89         Freshwater turtles remain poorly studied in ecotoxicology although they can  
90 provide useful information on local contamination because of their longevity  
91 (Campbell & Campbell, 2002 ; EL Hassani *et al.*, 2019 ; Gaus *et al.*, 2019), high trophic-  
92 level, and low dispersal capacity (Ming-ch'eng Adams *et al.*, 2016). Moreover, as  
93 ectothermic vertebrates, they have a lower ability to metabolize pollutants than  
94 endothermic mammals and birds (de Solla, 2015). The European pond turtle, *Emys*  
95 *orbicularis*, a long-lived (> 40-80 years) opportunistic predator (fish, amphibian,  
96 crayfish) and scavenger (Ottonello *et al.* 2005; Ficetola & De Bernardi, 2006), is facing  
97 significant population declines due to multiple environmental alterations, including  
98 water pollution (Cheylan 1998). This species is listed as “near threatened” (NT) on  
99 the UICN Red List of threatened species in France (REF). Pollution by trace metal  
100 elements has been studied in *E. orbicularis* (Namroodi *et al.*, 2017 ; Guillot *et al.*, 2018 ;  
101 Beau *et al.*, 2019), but data are lacking on burdens of POPs.

102 We studied contamination levels in two populations of *E. orbicularis* at the  
103 Natural Reserve of the Tour du Valat, in the Camargue, France. Several individuals  
104 were of known-age due to capture-mark-recapture (CMR) program, starting more  
105 than 20 years ago (Olivier *et al.*, 2010 ; Ficheux *et al.*, 2014 ; Arsovski *et al.*, 2018). The  
106 hydraulic system combines irrigation canals originating from the Rhône river and  
107 drainage canals, which are exposed to various environmental contaminants from the  
108 Rhône river and agricultural plots, especially rice fields (Chauvelon, 1996). In the  
109 present study, we assessed recent (2018) levels of OCPs and PCBs in plasma of the  
110 European pond turtle (n = 174). We tested the effects of individual traits (mass, sex,  
111 and age) on POP burdens. We expected that older and larger individuals should  
112 have higher levels of POPs because of bioaccumulation and length of exposure.  
113 Moreover, we compared POP burdens of turtles from two populations in different  
114 locations in the agricultural hydraulic system (irrigation vs drainage). We expected  
115 high levels of OCPs in turtles from the drainage site, as a possible consequence of the  
116 remobilization of pesticides from agricultural soils, and high levels of PCBs in turtles  
117 from the irrigation channel, as a possible consequence of the historical  
118 contamination.

119

## 120 **2. Material and methods**

### 121 **2.1. Sampling sites and capture**

122 We conducted the study in the Natural Reserve of the Tour du Valat (43°30' N, 4°40'  
123 E, [Fig. 1](#)) in France. We captured European pond turtles (n = 174) in canals and  
124 marshes from 24 April to 26 July 2018 by hand or with funnel traps (Olivier *et al.*,

125 2010; Ficheux *et al.*, 2014). We captured turtles at two sites: (1) irrigation canals and  
126 their associated marshland (site of Esquineau, n = 126) and (2) drainage canals of the  
127 Fumemorte basin and marshes filled with water by these canals (site of Faïsses, n =  
128 48). The sex ratio was the same between the two sites (58.3% females at Faïsses and  
129 60.3% females at Esquineau, Pearson's Chi-squared test,  $p = 0.95$ ).

130 Individual turtles were identified as part of a long-term CMR program with  
131 shallow notches on the marginal and nuchal scales in unique combinations (Olivier,  
132 2002 ; Olivier *et al.*, 2010). Individuals were weighed and sexed by visual observation  
133 of the secondary sexual characters. The year of birth can be determined if the first  
134 capture occurred within the first 5 years of life by counting the number of growth  
135 streaks (Castanet, 1988). This long-term survey starts in 1997, although a few tens of  
136 individuals were also marked as adults from 1976. In our data set, 66 of the 174  
137 individuals were of known-aged (5 to 26 year-old, corresponding to a first capture  
138 between 1997 and 2018) and some were first captured as adults in 1976, thus being  
139 considered as more than 44 year-old. When turtles were captured after their first 5  
140 years of life (N = 108 in the present study), we determined a stage of growth  
141 according to growth streaks located on the plastron (Olivier, 2002). By combining  
142 estimated age and stage of growth, we then attributed an age class to each turtle: 4 to  
143 8 year-old (N = 20), 9 to 13 year-old (N = 46), 14 to 22 year-old (N = 54), and 23 to 44+  
144 year-old (N = 54). Turtles were released on the same day at their capture site.

145

## 146 **2.2. Blood sampling**

147 Plasma is a good matrix to determine individual burdens of POPs because blood  
148 sampling is minimally invasive and because circulating levels of POPs in blood are



149 significantly correlated to concentrations of POPs stored in fat (Keller *et al.*, 2004 ;  
150 Dabrowska *et al.*, 2006), liver, kidney, and muscle tissues (Van de Merwe *et al.*, 2010).  
151 We collected blood samples (2 ml, always less than 1% of the turtle body mass) from  
152 the dorsal coccygeal vein (Innis & Knotek, 2020; Keller *et al.*, 2004) with a Terumo  
153 Company® syringe (Somerset, USA) pre-impregnated with heparin to prevent blood  
154 clotting during collection and equipped with a 25G needle (Franklin Lakes, USA).  
155 We then centrifuged the samples to separate the plasma from the red blood cells. We  
156 collected blood samples from 174 individuals (126 from Esquineau and 48 from  
157 Faïsses), 66 individuals were of known age ranging from 5 to 31 years old (32 females  
158 and 19 males from Esquineau; 9 females and 6 males from Faïsses). Samples were  
159 stored at  $-18\text{ }^{\circ}\text{C}$  until analysis in the UMR 7619 METIS of Sorbonne Université,  
160 France.

161

### 162 **2.3. Chemical analyses**

163 We determined the levels of 17 OCPs (p,p'-DDT and metabolites (p,p'-DDE and p,p'-  
164 DDD); pentachlorobenzene (PeCB); hexachlorobenzene (HCB);  
165 pentachloronitrobenzene (Quintozene); four isomers of hexachlorocyclohexane ( $\alpha$ -,  
166  $\beta$ -,  $\gamma$ - (lindane),  $\delta$ -HCH); aldrin; endrin; isodrin; telodrin (Isonbenzan); heptachlor;  
167 heptachlor epoxide; heptachlor endo-epoxide), 7 marker PCBs (M-PCBs, IUPAC  
168 numbers # 28, 52, 101, 118, 138, 153 and 180), and 12 dioxin-like PCBs (DL-PCBs,  
169 IUPAC numbers # 77, 81, 105, 114, 118, 123, 126, 156, 157, 167, 169 and 189) in the 174  
170 plasma samples.

171 Samples were processed through Solid Phase Extraction (SPE), using hexane:  
172 dichloromethane, 9:1, using a validated protocol (Tapie *et al.* 2011) that was adapted

173 for plasma samples (see Supplementary Material for a detailed description of the  
174 procedure and method validations). PCBs and OCPs were analyzed using an Agilent  
175 7890 A gas chromatograph (GC) coupled to a 7000 B triple quadrupole mass  
176 spectrometer system (MS/MS) (Agilent Technologies, Les Ulis, France). Recovery  
177 rates (% RR) of compounds were assessed on replicate plasma samples of European  
178 pond turtles (n = 4) with spiked solutions (100 ng of each compound). The  
179 repeatability of the method was assessed in terms of relative standard deviation (%  
180 RSD) of the recovery. RR and RSD are provided in Table S2 (Supplementary  
181 Material). Recovery rates were not satisfactory (< 75% or > 125%) for isodrin, endrin,  
182 heptachlor epoxid, heptachlor-endo-epoxide, p,p'-DDD, PCB-77, PCB-189 (Table S2)  
183 and they were not further considered.

184

#### 185 **2.4. Statistical analyses**

186 We performed statistical analyses with R software version 3.3.2 (R Core Team, 2016).  
187 Only POPs detected (i.e. > LOQ) in at least 10% of the samples were included in the  
188 statistical analyses. Since age and weight were positively correlated ( $t = 3.65$ ,  $df = 64$ ,  
189  $p = 0.0005$ ) and females were significantly heavier than males ( $W = 474$ ,  $p < 0.0001$ ),  
190 body mass, sex and age were tested separately..

191 We used generalized linear models (GLM) with a binomial distribution and a  
192 logit link function to test for the effects of sex, site, the interaction sex \* site, sampling  
193 date, and age (or body mass) on contaminant occurrence. For each model, we used a  
194 backward elimination to progressively remove non-significant terms ( $p > 0.05$ ). For  
195 all analyses, model specification and validation were based on residual analysis.

196 The effects of sex, sampling site, date, and age (or body mass) on pollutant  
197 concentrations were tested by considering left-censored data, i.e. pollutant levels  
198 below the LOQ. To do so, group comparisons and linear regressions were performed  
199 using Peto-Prentice tests and tobit models respectively, with the function `cendiff` of  
200 the NADA package and the function `tobit` of the AER R-package.

201

## 202 **3. Results**

203

### 204 **3.1. POP levels and occurrences**

205 In 37 of the 174 samples, all contaminant levels were below the limit of  
206 quantification, which represented 10.4% of turtles sampled in Faïsses and 24.4% of  
207 turtles sampled in Esquineau. Eleven compounds were never detected in the plasma  
208 samples: PeCB, Lindane, Quintozene, Heptachlor, and the dioxin-like PCBs: PCB-  
209 118, -105, -114, -126, -156, -157 and -167. The following seven compounds were  
210 detected at least in one sample from Esquineau, but not in samples from Faïsses:  
211 alpha-HCH, 44'-DDT and PCBs 28, 101, 123, 169 (Table 1). Overall, the contaminant  
212 concentrations in the plasma samples were very low (Table 1).

213

### 214 **3.2. Influence of individual traits, sampling site, and date on POP occurrences and 215 levels**

216 Only three POPs (HCB, PCB-52 and -153) had substantial occurrence, being detected  
217 in more than 10% of individuals (Table 1). Plasma occurrences of PCB-52 and HCB  
218 were significantly higher in turtles captured at Faïsses compared to Esquineau, while

219 the occurrence of PCB-153 was higher in turtles captured at Esquineau (Tables 1 and  
220 2). HCB occurrence slightly increased with sampling date (Table 2). Sex, mass, and  
221 age class did not explain POP occurrence in turtles, except for PCB-153 whose  
222 detection frequency was higher in older individuals (Table 2).

223         Levels of HCB and PCB-52 were significantly higher in turtles from Faïsses,  
224 while PCB-153 levels tended to be higher in turtles from Esquineau (Tables 1 and 3).  
225 Moreover, levels of PCB-153 were lower in the youngest turtles (Table 3, Fig. 2). Sex,  
226 mass, and date did not explain variation in levels of POPs among turtles (Table 3).

227

## 228 **4. Discussion**

229

230 The aim of this study was to characterize legacy levels of POPs (OCPs and PCBs) in  
231 the plasma of European pond turtles in the Camargue, France, and to determine  
232 whether these levels were a function of individual traits and habitat types. Plasma  
233 levels of POPs were low and often below limits of quantification (from 0.2 to 3.1  
234 ng/mL). We found significant differences in burdens of POPs between turtles from  
235 the two sampling sites and these differences could be attributed to the hydraulic  
236 system (drainage/irrigation). The occurrence and levels of PCB-153 were higher in  
237 older turtles.

238         The Rhône River has been historically contaminated, leading to an  
239 accumulation of POPs in sediments downstream (Mourier et al. 2014). Previous  
240 studies conducted on several animal taxa in the Camargue confirmed high exposure  
241 during the last decades, with high concentrations of PCBs in muscles of eels (*Anguilla*  
242 *anguilla*) fished between 1997 and 2000 (Roche et al., 2004), but also in eggs of little

243 egrets (*Egretta garzetta*) collected in 1996 (Berny *et al.*, 2002). We detected low levels of  
244 POPs in the pond turtle in Camargue, which could indicate a decline in legacy POP  
245 exposure for wild species in the Camargue. Although a longitudinal study of the  
246 same species at the same locations would be required to test this hypothesis,  
247 decreasing PCB concentrations have been observed in the Rhône River sediments  
248 over the last decades (Liber *et al.*, 2019) and water analyses conducted by the Nature  
249 Protection National Society (SNPN) have not detected PCBs and OCPs in the  
250 Fumemorte canal since 2011 (Cheiron, 2019). Moreover, organochlorine  
251 contamination across food webs tended to diminish in the Vaccarès lagoon (Roche *et*  
252 *al.*, 2009a), the largest water body of the Camargue Biosphere Reserve. Exotic red  
253 swamp crayfish (*Procambarus clarkii*), the main prey of European pond turtles in the  
254 Camargue (Ottonello *et al.*, 2005), were not contaminated by POPs in 2019 (i.e. <  
255 LOQ, with LOQ ranging from 0.9. to 1.98 ng/g dw, unpublished data).

256 One of the first studies looking at the plasma concentration of POPs in turtles  
257 documented high PCBs and OCPs levels in snapping turtles (*Chelydra serpentina*), in  
258 Ontario, Canada and sampled in 2001–2004 (Letcher *et al.*, 2015): plasma  
259 concentrations of OCPs (sum of 17 contaminants) ranged from 0.2 to 236 ng/g w.w.  
260 and the most abundant pesticide was p,p'-DDE (mean  $\pm$  SE:  $27 \pm 6$  ng/g w.w.). In our  
261 study, plasma concentrations of POPs (sum of 16 contaminants) were much lower,  
262 ranging from 0 to 2 ng/g w.w and p,p'-DDE levels didn't exceed 2.3 ng/g ww.  
263 Snapping turtles are freshwater turtles with similar feeding habits than European  
264 pond turtle: they also consume plant and animal matter, including aquatic  
265 invertebrates, fish, frogs, reptiles (Ernst *et al.* 1994). Other studies of turtles

266 documented plasma concentrations of POPs comparable to those in our study; for  
267 instance, for PCB 153 and HCB in another freshwater species, the Western pond  
268 turtle (*Actinemys marmorata*) in Sequoia National Park, USA, in 2012 (Meyer *et al.*,  
269 2016), as well as in marine turtle, such as Loggerhead sea turtles (*Caretta caretta*) in  
270 the Eastern Atlantic Ocean in 2011-2012 (Bucchia *et al.*, 2015), and in Green turtles  
271 (*Chelonia mydas*) and Hawksbill turtles (*Eretmochelys imbricata*) in Cape Verde  
272 between 2009 and 2011 (Camacho *et al.*, 2014).

273 In terms of occurrence of POPs, our results in European pond turtles were low  
274 compared to other studies (Bucchia *et al.*, 2015 ; Meyer *et al.*, 2016). The occurrences  
275 of dioxin-like PCBs in 8% of European pond turtles, as well as p,p'-DDE in 2% of  
276 individuals were much lower than those found in plasma of Loggerhead sea turtles  
277 that ranged from 63% in the Atlantic Ocean to 100% in the Adriatic Sea for dioxin-  
278 like PCBs and 100 % for p,p'-DDE (Bucchia *et al.*, 2015).

279 Despite low detection frequencies and levels, we found differences in POP  
280 burdens between turtles from the two sampling sites. Turtles in Faïsses, a site  
281 receiving water from drainage canals, exhibited significantly higher occurrence and  
282 levels of HCB compared to turtles from Esquineau, likely because of the  
283 remobilisation of trapped OCPs in soils. On the other hand, the occurrence and levels  
284 of PCB 153 were higher in turtles from Esquineau, a site with irrigation channels. A  
285 previous study has documented higher concentrations of PCBs in the aquatic fauna  
286 in irrigation canals compared to the drainage canals downstream of rice fields (Roche  
287 *et al.*, 2009b). We also found that concentrations of PCB-52 were higher in turtles at  
288 Faïsses, however, which may be due to a lower sedimentation rate of PCB-52

289 associated with its lower molecular weight and a greater solubility due to a lower  
290 level of chlorination compared to PCB-153 (Gong *et al.*, 1998 ; Alkhatib & Weigand,  
291 2002). The processes of deposition and release of PCBs from the irrigation part to the  
292 drainage part of the canals need to be further studied. In addition, other populations  
293 of pond turtles in the two types of hydraulic systems should be studied to confirm  
294 this pattern.

295         The occurrence and levels of PCB-153 were lower in the youngest individuals  
296 (4 to 8 year old), which could be attributed to a shorter exposure period. PCB-153 is a  
297 recalcitrant chemical with a high hydrophobicity (high octanol-water partition  
298 coefficient), high resistance to metabolic transformation, and slow respiratory  
299 elimination through air-exhalation (high octanol-air partition coefficient). PCB-153 is  
300 also more prone to bioaccumulation than many other PCB congeners (Kelly *et al.*  
301 2007), especially in long-lived air-breathing predatory species (Rowe, 2008). It is also  
302 possible that there is a dietary shift during growth in pond turtles leading to the  
303 consumption of higher trophic-level prey with age, and these higher trophic level  
304 prey would be more PCB contaminated due to biomagnification. This hypothesis,  
305 however, is not supported by dietary studies of *E. orbicularis* in Camargue), showing  
306 a shift to a more herbivorous diet with age, using prey items identification in faecal  
307 samples (Ottonello *et al.*, 2005) or no difference in the proportion of plants,  
308 invertebrates and vertebrates, using metabarcoding analyses (Duccotterd *et al.*  
309 2020).)

310         We did not detect differences in POP loads between male and female pond  
311 turtles, despite males often being more contaminated than females in other turtle

312 species (Guirlet *et al.*, 2010 ; Bangma *et al.*, 2019; Lambiase *et al.*, 2021) due to a  
313 transfer of pollutants from the mother to the eggs through vitellogenesis (Moss *et al.*,  
314 2009 ; de Solla, 2010).

315

316



## 317 **Declaration of competing interest**

318 The authors declare that they have no known competing financial interests or  
319 personal relationships that could have appeared to influence the work reported in  
320 this paper.

321

## 322 **Acknowledgements**

323 We are indebted to the many people who participated in the turtle captures since  
324 1997.

325 We thank the French Ministry of Environment for giving us the permission to  
326 capture European pond turtles.

327

## 328 **Funding**

329 This work was funded by the Water agency Rhône Méditerranée Corse and the  
330 Fondation de la Tour du Valat.

331

## 332 **Author contributions**

333 AO, LB, CLG and NM conducted the field work and collected the blood samples. LB  
334 and FA conducted the laboratory work to prepare plasma samples for analysis with  
335 gas chromatography and tandem mass spectrometry. LB, OL, AG, AO, GBD, and MV  
336 conceived and coordinated the study, participated in data analysis, and in writing  
337 the manuscript. All authors gave final approval for publication and agree to be held  
338 accountable for the work performed therein.

339

340 **Ethical consideration**

341 This study was performed in accordance with laws relative to capture, transport, and  
342 experiments on *E. orbicularis* (DREAL permit # CERFA\_13616-01) and all procedures  
343 were approved by an independent ethical committee (APAFIS#17899-  
344 201812022345423 v2).

345 **Literature cited**

- 346 Alkhatib E. & Weigand C. 2002. Parameters affecting partitioning of 6 PCB congeners in natural  
347 sediments. *Environmental Monitoring and Assessment*. 78(1), p. 1- 17.
- 348 Arsovski D., Olivier A., Bonnet X., Drilholle S., Tomović L., Béchet A., Golubović A., & Besnard A. 2018.  
349 Covariates streamline age-specific early life survival estimates of two chelonian species.  
350 *Journal of Zoology*. 306(4), p. 223- 234.
- 351 Bangma J.T., Ragland J.M., Rainwater T.R., Bowden J.A., Gibbons J.W., & Reiner J.L. 2019.  
352 Perfluoroalkyl substances in diamondback terrapins (*Malaclemys terrapin*) in coastal South  
353 Carolina. *Chemosphere*. 215, p. 305- 312.
- 354 Beau F., Bustamante P., Michaud B., & Brischoux F. 2019. Environmental causes and reproductive  
355 correlates of mercury contamination in European pond turtles (*Emys orbicularis*).  
356 *Environmental Research*. 172, p. 338- 344.
- 357 Berny P., Sadoul N., Dol S., Videman B., Kayser Y., & Hafner H. 2002. Impact of local agricultural and  
358 industrial practices on organic contamination of little egret (*Egretta garzetta*) eggs in the  
359 Rhône Delta, Southern France. *Environmental Toxicology and Chemistry*. 21(3), p. 520- 526.
- 360 Binnington M.J. & Wania F. 2014. Clarifying relationships between persistent organic pollutant  
361 concentrations and age in wildlife biomonitoring: individuals, cross-sections, and the roles of  
362 lifespan and sex: Clarifying POP concentration-age trends in biomonitoring. *Environmental*  
363 *Toxicology and Chemistry*. 33(6), p. 1415- 1426.
- 364 Bucchia M., Camacho M., Santos M.R.D., Boada L.D., Roncada P., Mateo R., Ortiz-Santaliestra M.E.,  
365 Rodríguez-Estival J., Zumbado M., Orós J., Henríquez-Hernández L.A., García-Álvarez N., &  
366 Luzardo O.P. 2015. Plasma levels of pollutants are much higher in loggerhead turtle  
367 populations from the Adriatic Sea than in those from open waters (Eastern Atlantic Ocean).  
368 *Science of The Total Environment*. 523, p. 161- 169.
- 369 Camacho M., Boada L.D., Orós J., López P., Zumbado M., Almeida-González M., & Luzardo O.P. 2014.  
370 Monitoring organic and inorganic pollutants in juvenile live sea turtles: Results from a study  
371 of *Chelonia mydas* and *Eretmochelys imbricata* in Cape Verde. *Science of The Total*  
372 *Environment*. 481, p. 303- 310.
- 373 Campbell K.R. & Campbell T.S. 2002. A logical starting point for developing priorities for lizard and  
374 snake ecotoxicology: A review of available data. *Environmental Toxicology and Chemistry*.  
375 21(5), p. 894- 898.
- 376 Castanet J. 1988. Les méthodes d'estimation de l'âge chez les chéloniens. *Mesogée*. 48, p. 21–28.
- 377 Chauvelon P. 1996. *Hydrologie quantitative d'une zone humide méditerranéenne aménagée : le*  
378 *bassin de Fumemorte en Grande Camargue, delta du Rhône*. thesis. Montpellier 2
- 379 Cheiron A. 2019. Rapport d'activité 2018. Société nationale de protection de la nature – Réserve  
380 Naturelle Nationale de Camargue. , p. 216.
- 381 Cheiron A., Befeld S., & Bricault B. 2018. *Rapport d'activité 2017 de la Réserve Naturelle Nationale de*  
382 *Camargue. Société nationale de protection de la nature – Réserve Naturelle Nationale de*  
383 *Camargue, 242 p.*

- 384 Cheylan M. 1998. Evolution of the distribution of the European pond turtle in the French  
385 Mediterranean area since the post-glacial. In : Fritz U., Joger U., Podloucky R. & Servan J.  
386 (eds). Proceeding of the EMYS Symposium. Dresden. Mertensiella. 10, p. 47-65.
- 387 Comoretto L., Arfib B., & Chiron S. 2007. Pesticides in the Rhône river delta (France): Basic data for a  
388 field-based exposure assessment. *Science of The Total Environment*. 380(1), p. 124- 132.
- 389 Dabrowska H., Fisher S.W., Estenik J., Kidekhel R., & Stromberg P. 2006. Polychlorinated Biphenyl  
390 Concentrations, Congener Profiles, and Ratios in the Fat Tissue, Eggs, and Plasma of Snapping  
391 Turtles (*Chelydra s. serpentina*) from the Ohio Basin of Lake Erie, USA. *Archives of*  
392 *Environmental Contamination and Toxicology*. 51(2), p. 270- 286.
- 393 Ducotterd, C., Crovadore, J., Lefort, F., Guisan, A., Ursenbacher, S., & Rubin, J. F. (2020). The feeding  
394 behaviour of the European pond turtle (*Emys orbicularis*, L. 1758) is not a threat for other  
395 endangered species. *Global Ecology and Conservation*, 23, e01133.
- 396 Dudgeon D., Arthington A.H., Gessner M.O., Kawabata Z.-I., Knowler D.J., Lévêque C., Naiman R.J.,  
397 Prieur-Richard A.-H., Soto D., Stiassny M.L.J., & Sullivan C.A. 2006. Freshwater biodiversity:  
398 importance, threats, status and conservation challenges. *Biological Reviews*. 81(02), p. 163.
- 399 EL Hassani M.S., El Hassan E.M., Slimani T., & Bonnet X. 2019. Morphological and physiological  
400 assessments reveal that freshwater turtle (*Mauremys leprosa*) can flourish under extremely  
401 degraded-polluted conditions. *Chemosphere*. 220, p. 432- 441.
- 402 Ernst CH, Lovich JE, Barbour RW. 1994. Turtles of the United States and Canada. Smithsonian  
403 Institute Press, Washington, DC.
- 404 Ficetola G.F. & De Bernardi F. 2006. Is the European « pond » turtle *Emys orbicularis* strictly aquatic  
405 and carnivorous? *Amphibia-Reptilia*. 27(3), p. 445- 447.
- 406 Ficheux S., Olivier A., Fay R., Crivelli A., Besnard A., & Béchet A. 2014. Rapid response of a long-lived  
407 species to improved water and grazing management: The case of the European pond turtle  
408 (*Emys orbicularis*) in the Camargue, France. *Journal for Nature Conservation*. 22(4),  
409 p. 342- 348.
- 410 Gaus C., Villa C.A., Dogruer G., Heffernan A., Vijayasathy S., Lin C.-Y., Flint M., Hof C.M., & Bell I.  
411 2019. Evaluating internal exposure of sea turtles as model species for identifying regional  
412 chemical threats in nearshore habitats of the Great Barrier Reef. *Science of The Total*  
413 *Environment*. 658, p. 732- 743.
- 414 Gong Y., Depinto J.V., Rhee G.-Y., & Liu X. 1998. Desorption rates of two PCB congeners from  
415 suspended sediments---I. experimental results. *Water Research*. 32(8), p. 2507- 2517.
- 416 Goutte A., Barbraud C., Meillère A., Carravieri A., Bustamante P., Labadie P., Budzinski, H., Delord K.,  
417 Chérel Y., Weimerskirch H., Chastel O. (2014) Demographic consequences of heavy metals  
418 and persistent organic pollutants in a vulnerable long-lived bird, the wandering albatross.  
419 *Proceedings of the Royal Society B* 281, 20133313
- 420 Goutte A., Barbraud C., Herzke D., Bustamante P., Angelier F., Tartu S., Clément-Chastel C., Moe B.,  
421 Bech C., Gabrielsen G.W., Bustnes J.O., Chastel O. 2015 Survival rate and breeding outputs in  
422 a high Arctic seabird exposed to legacy persistent organic pollutants and mercury.  
423 *Environmental Pollution* 200, 1-9.

- 424 Goutte A., Alliot F., Budzinski H., Simonnet-Laprade C., Santos R., Lachaux V., Maciejewski K., Le  
425 Menach K., Labadie P. 2020. Trophic transfer of micropollutants and their metabolites in an  
426 urban riverine food web. *Environmental Science and Technology* 54(13), 8043-8050.
- 427 Guillot H., Bonnet X., Bustamante P., Churlaud C., Trotignon J., & Brischoux F. 2018. Trace Element  
428 Concentrations in European Pond Turtles (*Emys orbicularis*) from Brenne Natural Park,  
429 France. *Bulletin of Environmental Contamination and Toxicology*. 101(3), p. 300- 304.
- 430 Guirlet E., Das K., Thomé J.-P., & Girondot M. 2010. Maternal transfer of chlorinated contaminants in  
431 the leatherback turtles, *Dermochelys coriacea*, nesting in French Guiana. *Chemosphere*.  
432 79(7), p. 720- 726.
- 433 Holt M.S. 2000. Sources of chemical contaminants and routes into the freshwater environment. *Food  
434 and Chemical Toxicology*. 38, p. S21- S27.
- 435 Innis C. & Knotek Z. 2020. Tortoises and Freshwater Turtles. Dans : Heatley J.J., Russell K.E. (éd.).  
436 *Exotic Animal Laboratory Diagnosis*. Wiley, p. 255- 289.
- 437 Keller J.M., Kucklick J.R., Harms C.A., & McClellan-Green P.D. 2004. Organochlorine contaminants in  
438 sea turtles: Correlations between whole blood and fat. *Environmental Toxicology and  
439 Chemistry*. 23(3), p. 726- 738.
- 440 Kelly, B. C., Ikonomou, M. G., Blair, J. D., Morin, A. E., & Gobas, F. A. 2007. Food web-specific  
441 biomagnification of persistent organic pollutants. *science*, 317(5835), 236-239.
- 442 Kuenzer C. & Renaud F.G. 2012. Climate and Environmental Change in River Deltas Globally: Expected  
443 Impacts, Resilience, and Adaptation. Dans : Renaud F.G., Kuenzer C. (éd.). *The Mekong Delta  
444 System*. Dordrecht : Springer Netherlands, p. 7- 46.
- 445 Lambiase, S., Serpe, F. P., Pilia, M., Fiorito, F., Iaccarino, D., Gallo, P., & Esposito, M. (2021).  
446 Polychlorinated organic pollutants (PCDD/Fs and DL-PCBs) in loggerhead (*Caretta caretta*)  
447 and green (*Chelonia mydas*) turtles from Central-Southern Tyrrhenian Sea. *Chemosphere*,  
448 128226.
- 449 Letcher R.J., Lu Z., de Solla S.R., Sandau C.D., & Fernie K.J. 2015. Snapping Turtles (*Chelydra  
450 serpentina*) from Canadian Areas of Concern across the southern Laurentian Great Lakes:  
451 Chlorinated and brominated hydrocarbon contaminants and metabolites in relation to  
452 circulating concentrations of thyroxine and vitamin A. *Environmental Research*. 143,  
453 p. 266- 278.
- 454 Liber Y., Mourier B., Marchand P., Bichon E., Perrodin Y., & Bedell J.-P. 2019. Past and recent state of  
455 sediment contamination by persistent organic pollutants (POPs) in the Rhône River:  
456 Overview of ecotoxicological implications. *Science of The Total Environment*. 646,  
457 p. 1037- 1046.
- 458 Meyer E., Eskew E.A., Chibwe L., Schrlau J., Massey Simonich S.L., & Todd B.D. 2016. Organic  
459 contaminants in western pond turtles in remote habitat in California. *Chemosphere*. 154,  
460 p. 326- 334.
- 461 Ming-ch'eng Adams C.I., Baker J.E., & Kjellerup B.V. 2016. Toxicological effects of polychlorinated  
462 biphenyls (PCBs) on freshwater turtles in the United States. *Chemosphere*. 154, p. 148- 154.

- 463 Mourier B., Desmet M., Van Metre P. C., Mahler B. J., Perrodin Y., Roux G., Bedell J.P, Lefèvre I.&  
464 Babut M. (2014). Historical records, sources, and spatial trends of PCBs along the Rhône  
465 River (France). *Science of the total environment*, 476, 568-576.
- 466 Moss S., Keller J.M., Richards S., & Wilson T.P. 2009. Concentrations of persistent organic pollutants  
467 in plasma from two species of turtle from the Tennessee River Gorge. *Chemosphere*. 76(2),  
468 p. 194- 204.
- 469 Namroodi S., Zaccaroni A., Rezaei H., & Hosseini S.M. 2017. European pond turtle (*Emys orbicularis*  
470 *persica*) as a biomarker of environmental pollution in Golestan and Mazandaran provinces,  
471 Iran. *Veterinary Research Forum*. 8(4), p. 333- 339.
- 472 Olivier A. 2002. *Ecologie, traits d'histoire de vie et conservation d'une population de Cistude d'Europe*  
473 *Emys orbicularis en Camargue*. Ecole Pratique des Hautes Etudes, 165 p.
- 474 Olivier A., Barbraud C., Rosecchi E., Germain C., & Cheylan M. 2010. Assessing spatial and temporal  
475 population dynamics of cryptic species: an example with the European pond turtle.  
476 *Ecological Applications*. 20(4), p. 993- 1004.
- 477 Ottonello D., Salvidio S., & Rosecchi E. 2005. Feeding habits of the European pond terrapin *Emys*  
478 *orbicularis* in Camargue (Rhône delta, Southern France). *Amphibia-Reptilia*. 26(4),  
479 p. 562- 565.
- 480 Poulhier G., Launay M., Le Bescond C., Thollet F., Coquery M., & Le Coz J. 2019. Combining flux  
481 monitoring and data reconstruction to establish annual budgets of suspended particulate  
482 matter, mercury and PCB in the Rhône River from Lake Geneva to the Mediterranean Sea.  
483 *Science of The Total Environment*. 658, p. 457- 473.
- 484 Revenga C., Campbell I., Abell R., de Villiers P., & Bryer M. 2005. Prospects for monitoring freshwater  
485 ecosystems towards the 2010 targets. *Philosophical Transactions of the Royal Society B:*  
486 *Biological Sciences*. 360(1454), p. 397- 413.
- 487 Roche H., Buet A., & Ramade F. 2002. Relationships between persistent organic chemicals residues  
488 and biochemical constituents in fish from a protected area: the French National Nature  
489 Reserve of Camargue. *Comparative Biochemistry and Physiology Part C: Toxicology &*  
490 *Pharmacology*. 133(3), p. 393- 410.
- 491 Roche H., Buet A., & Ramade F. 2003. Mise en évidence et validation de biomarqueurs  
492 écotoxicologiques dans la population d'anguilles d'un étang de la réserve naturelle nationale  
493 de Camargue, le Vaccarès, exposée à des polluants organiques persistants. *Revue d'Écologie*.  
494 (1), p. 77- 102.
- 495 Roche H., Buet A., & Ramade F. 2004. Recherches écotoxicologiques sur la contamination des  
496 peuplements de poissons de la réserve naturelle nationale de Camargue par les polluants  
497 organiques persistants. *Revue d'écologie*. 59(1- 2), p. 101- 111.
- 498 Roche H., Vollaire Y., Persic A., Buet A., Oliveira-Ribeiro C., Coulet E., Banas D., & Ramade F. 2009a.  
499 Organochlorines in the Vaccarès Lagoon trophic web (Biosphere Reserve of Camargue,  
500 France). *Environmental Pollution*. 157(8- 9), p. 2493- 2506.
- 501 Roche H., Vollaire Y., Martin E., Rouer C., Coulet E., Grillas P., & Banas D. 2009b. Rice fields regulate  
502 organochlorine pesticides and PCBs in lagoons of the Nature Reserve of Camargue.  
503 *Chemosphere*. 75(4), p. 526- 533.

- 504 Rowe C.L. 2008. « The Calamity of So Long Life »: Life Histories, Contaminants, and Potential  
505 Emerging Threats to Long-lived Vertebrates. *BioScience*. 58(7), p. 623- 631.
- 506 Salice C. J., Rowe C. L., & Eisenreich K. M. (2014). Integrative demographic modeling reveals  
507 population level impacts of PCB toxicity to juvenile snapping turtles. *Environmental pollution*,  
508 184, 154-160.
- 509 de Solla S.R. 2010. Organic contaminants in reptiles. Dans : *Sparling, D.W., Linder,G., Bishop, C.A.,*  
510 *Crest, S.K.1 (Eds.), Ecotoxicology of Amphibians and Reptiles*. 2nd ed. SETACPress, Pensacola,  
511 FL, p. 289- 324.
- 512 de Solla S.R. 2015. Exposure, Bioaccumulation, Metabolism and Monitoring of Persistent Organic  
513 Pollutants in Terrestrial Wildlife. In Alae M. (éd.). *Dioxin and Related Compounds*. Cham :  
514 Springer International Publishing, p. 203- 252.
- 515 Tapie, N., Le Menach, K., Pasquaud, S., Elie, P., Devier, M.H., Budzinski, H., 2011. PBDE and PCB  
516 contamination of eels from the Gironde estuary: from glass eelsto silver eels. *Chemosphere*  
517 83, 175–185
- 518 Van de Merwe J.P., Hodge M., Olszowy H.A., Whittier J.M., & Lee S.Y. 2010. Using blood samples to  
519 estimate persistent organic pollutants and metals in green sea turtles (*Chelonia mydas*).  
520 *Marine Pollution Bulletin*. 60(4), p. 579- 588.
- 521 Vives, I., Grimalt, J. O., Ventura, M., Catalan, J., & Rosseland, B.O. 2005. Age dependence of the  
522 accumulation of organochlorine pollutants in brown trout (*Salmo trutta*) from a remote high  
523 mountain lake (Redó, Pyrenees). *Environmental Pollution* 133, 343–350.
- 524

525 **Figure captions**

526 **Figure 1:** Sampling locations for European pond turtles in the regional nature reserve  
527 of Tour du Valat, Camargue, France. The two capture locations are Esquineau (in  
528 blue), irrigated with water pumped from the Rhône River, and Faïsses (in orange),  
529 consisting of drainage canals of agricultural parcels.

530

531 **Figure 2:** Increase in plasma PCB 153 (ng/mL) levels in European pond turtles with  
532 age class (4 to 8, 9 to 13, 14 to 22, 23 to higher than 44 years old). Each point  
533 represents an individual. Red and green squares correspond to mean and median  
534 levels for age class, respectively. Age class was estimated according to growth  
535 streaks located on the plastron (see methods).