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#### Impact of nickel mining in New Caledonia on marbled eels Anguilla marmorata

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14

# 15 Abstract

16 New Caledonia is particularly affected by nickel open pit mining activities because of the presence of 17 ultramafic soils rich in metal trace elements. The nickel particles dispersed by atmospheric transport 18 and soil erosion during the excavation of nickel will end up by deposition or leaching in aquatic 19 ecosystems where they may be bioaccumulated by living organisms in the rivers downstream the 20 mines. Despite alarming freshwater metals concentrations, no study investigated the level of their 21 bioaccumulation in eels living downstream mining sites, and if high bioaccumulation levels occur, the 22 potential consequences on eel's health. The aim of this study was to determine how eels Anguilla 23 marmorata are impacted in situ by metal concentrations issued from mining activity by measuring: morphometric parameters; metal concentrations in tissues and organs and transcription levels of 24 25 target genes encoding proteins involved in several metabolic key functions. Among organs, liver was 26 found to be the most affected by mining with average nickel concentrations of 5.14 mg/kg versus 27 1.63 mg/kg for eel liver away from mines leading to dysregulation of numerous genes involved in 28 oxidative stress, DNA repair, apoptosis, reproduction and both lipid and mitochondrial metabolisms. 29 This study should allow us to define in an integrated way if metals released by mining activities 30 influence metals bioaccumulation in eels and induce biological effects.

31 Keywords: New Caledonia; Metals; Eels; Bioaccumulation; Gene transcription

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#### 33 **1. Introduction**

Aquatic ecosystems are particularly impacted by anthropogenic pollutants since many years, especially by metals coming from industrial, urban or agricultural sources, such as cadmium (Cd), mercury (Hg), zinc (Zn) or nickel (Ni) for instance, causing harmful impacts on species living in these environments. Some studies have already highlighted the toxic effects of water metal pollution on fish such as the yellow perch *Perca flavescens*, the spiny eel *Mastecembelus armatus*, or the European eel *Anguilla anguilla* [1-3].

40 Among the human activities responsible of this pollution, today, mining activities represent the second source of heavy metal soils contamination in the world [4]. In fact, soils are subjected to 41 42 significant metal leaching due to the high concentration of heavy metals and the low capacity of the 43 soil to retain water. Metals can then persist a long time in aquatic environment [5-9]. This type of 44 contamination is particularly important in New Caledonia, one of the largest Ni producer in the world 45 with resources concentrated in ultramafic formations covering 30 % of the subsoils [10]. These 46 formations are particularly rich in metal trace elements including Ni, chromium (Cr), cobalt (Co), and manganese (Mn) [11]. Natural erosion, due to the laterite soils fragility and instability, especially 47 48 under tropical climate conditions characterized by heavy rains [12], leads to mobilization of particles 49 rich in metals from soils to freshwater [13-16]. Anthropic activities that destroy vegetation and 50 reshape the soils like mining, increase this process by increasing soil leaching and exposing metals 51 rich deep horizons. Runoff water is the main pathway of particles dispersion from soils to freshwater. 52 Atmospheric dust issued from this soils can also contribute to a lesser extend to freshwater 53 contamination. The dust concentration of Ni and Cr in Noumea are particularly elevated in 54 comparison with other countries [17]. These particles finally end up by deposition or leaching in 55 aquatic environment, and may affect living organisms in freshwaters downstream the mines.

56 In rivers under mining influence, high Ni concentrations are reported, around 96 µg/L, compared to 57 those away from mining on the same type of lateritic soil, around 7  $\mu$ g/L [15], confirming the high-58 water pollution level in this country. Indeed, these values are largely above the European regulation 59 recommendations (< 4  $\mu$ g/L). The increased use of Ni in industry may contribute to increase 60 anthropic and natural emission of particles and, consequently, increase the risk of fish 61 contamination. Some authors have demonstrated natural ecosystem and lagoon degradations due to Ni exploitations in New Caledonia [16, 18, 19] and recent studies have revealed that the 62 63 contamination of aquatic organisms is directly linked to the enrichment of soils by Ni [20]. A 64 significant bioaccumulation of Ni in gills and livers of five species (carp Khulia rupestris, white and 65 black lochon Awaous quamensis and Sicyopterus lagocephalus, tilapia Sarotherodon occidentalis and 66 marbled eel Anquilla marmorata) living in rivers under mining influence in the north of New 67 Caledonia have been highlighted too (DMML project). Among these species, marbled eels, Anguilla 68 marmorata, are particularly interesting and not yet studied with regards to metals contamination in New Caledonia. Eels, as top predator, are good bio-indicators of water metal contamination [21-23] 69 70 as they accumulate metals previously consumed by other species. Furthermore, with a complex 71 biological life cycle, characterized by different stages of metamorphoses as previously described by 72 [24], eels are particularly sensitive to water pollution. Indeed, during their juvenile growth stage, 73 eels, living in freshwaters during several years, must accumulate energy reserves before starting their 74 migration to reproduce themselves in open sea, tens of thousands kilometers away. During this 75 migration, eels swim for several months, stop to feed and mature their gonads. Therefore, the 76 quality of life in freshwaters is essential to ensure the success of migration and reproduction [25, 26] 77 and prevent the marbled eel decline.

It is well known that other eel's species, like European eels, see their populations decrease due to metal contamination for instance by mercury [27]. This decline could be explained by morphometric alteration and metabolic parameters dysfunction especially due to oxidative stress, mitochondria disruption and carcinogenic factors activation [28-30].

82 Regarding the metals accumulated in eel's organs, previous studies suggested that metal 83 concentration in organs can be used as a first proxy to assess metal toxicity, and the physiological 84 consequences in each organ. Liver, kidneys and gills seems to be the main metals accumulating 85 organs in European eels. It seems essential to know these parameters on marbled eels to clarify the 86 consequences of water contamination in New Caledonia. Furthermore, determining if Ni and other 87 associated metals accumulate in marbled eels muscles is essential to estimate the risk for human's consumption [31]. In New Caledonia, marbled eels living downstream mining sites seem to 88 89 accumulate important Ni and other associated metals in their organs, but the consequences of this 90 accumulation are unknown yet. The aim of this study was thus to determine the consequences of Ni and associated metals contamination in Anguilla marmorata by studying: (i) morphometric 91 92 parameters; (ii) metal concentrations in liver, gills, kidneys, muscle and hearts; and (iii) transcription 93 level of genes encoding proteins involved in lipid metabolism, oxidative stress, detoxification, cell cycle regulation, DNA repair and immunity in six different organs or tissues: liver, gills, kidneys, 94 95 muscle, brain and spleen.

- 96 **2. Materials and methods**
- 97

#### 2.1. Animals and biometric measurements

Twenty-one immature yellow marbled eels, *Anguilla marmorata*, were captured by electric fishing in
May 2016 and July 2021 near the Koniambo Mount of New Caledonia in four different sites (with 4 to
6 individuals per site). Three sites were used for their high metal concentration levels due to the soil

101 composition (ultramafic soils) and the proximity of an open pit mine (Koniambo Nickel SAS, KNS) 102 operating since 1997 along the Taléa river: left Taléa (LT) directly located downstream the mine; right 103 Taléa (RT) not influenced by the mine; and downstream Taléa (DT), after the confluence between the 104 two arms of the Taléa river (Fig. 1). One site was used as a reference located on a volcano-105 sedimentary soil, the downstream of the Tivoli river (Tiv). These sites are located on an intertropical 106 climate zone, with an average of 1071.5 mm of water each year (MétéoNC), leading to important soil 107 erosion and leaching in freshwaters.

Anguilla marmorata total body lengths and weights were first recorded to determine the Fulton condition factor [weight of eel (g) / (total length of the fish)<sup>3</sup> (cm<sup>3</sup>)]. Eels were then euthanized and kidneys, brain, muscle, liver, gills, heart and spleen were recovered for further analyses. The total weight of liver was measured to calculate the hepato-somatic index (weight of the liver / total weight of the fish x 100). Samples for RT-qPCR analyses were immediately fixed in RNA later before to be stored at -20°C until analyses. Samples for metal analyses were stored as soon as possible at -20°C until analyses.



125

Figure 2 : Map of fishing sites, Koniambo Pilot Site, New-Caledonia

## 126 **2.2. Metal concentration determination**

Biological samples: kidneys, muscles, gills, liver and heart, were oven dried (72 h at 50 °C), grounded into a finely homogenized powder using an agate mortar, weighed and then digested with 1.5 mL of nitric acid 20 % at 200°C for 15 min (Microwave Acid Digestion - MARS 6, CEM, "Fish tissue" method). 130 After the mineralization, 1.5 mL of ultra-pure water (Milli-Q,Bedford, MA, USA) was added to each 131 sample. Six metals were analyzed simultaneously by Inductively Coupled Plasma Optical Emission 132 Spectrometer (700 Series ICP-OES, Agilent): Ni, Co, Cu, Cr and Mn. Certified reference materials DOLT-5 (dogfish (Squalus acanthias liver), TORT-3 (Lobster Hepatopancreas), and IAEA-413 (Algae) 133 were analyzed using the same methodology as the eel samples. Their recovery rates, limit of 134 135 detection and quantification are presented in table 1. Metal concentrations are expressed in mg.kg<sup>-1</sup>. 136 For values below the instrument detection limit, theoretical minimum concentration values were calculated (the detection limit of the instrument (in mg.kg<sup>-1</sup>) multiplied by the volume of the sample 137 138 (in L) divided by the sample weight (in g)).

**Table 1**: Limit of detection and quantification with the recovery rates obtained from the analyses of certifiedreference materials (TORT-3, DOLT-5 and IAEA-413).

	Ni	Cr	Со	Mn
Limit of Detection ( $\mu$ g/L ± SD)	$1.91 \pm 0.70$	0.46 ± 0.09	0.97 ± 0.31	0.045 ± 0.018
Limit of Quantification at 10 % ( $\mu$ g/L ± SD)	9.55 ± 3.30	2.32 ± 0.46	4.89 ± 1.53	0.23 ± 0.09
DOLT-5 recovery rate (% SD)				81.5 ± 3.10
TORT-3 recovery rate (% SD)	73.1 ± 3.30			78.5 ± 3.80
IAEA-413 recovery rate (% SD)	91.1 ± 1.30	87.9 ± 0.90	100.3 ± 5.10	87.4 ± 0.80

141

# 142 **2.3. Gene transcription**

143 Quantitative reverse transcription PCR (RT-qPCR) was used to determine mRNA transcription levels in 144 fish sampled in 2021. In order to obtain the coding sequences of target genes, cDNA sequences from 145 the assembled transcriptome of A. anguilla [32, 33] were aligned with the assembled genome of A. 146 marmorata (GCA 901111315.1) using the Blast program. Sequences with high homologies were then 147 used to design specific primers for A. marmorata by means of the primer3plus software (Table 1; 148 [34] ). Total mRNA extraction was performed from liver, kidneys, gills, brain, spleen and muscles 149 samples using a commercially available NucleoZOL (RNA isolation, nucleoZOL, Marcherey-nagel®, 150 USA), according to the manufacturer's recommendations. Reverse transcription was done by M-MLV 151 (Moloney Murine Leukemia Virus) Reverse Transcriptase kit (Promega). Finally, PCR reactions were 152 performed using a LC480 (Roche), 95°C for 2 min, followed by 45 cycles of 95°C for 15 sec and 60°C 153 for 1 min. Each 15 μl reaction contained 7.5 μl GoTag qPCR Master mix (Promega), 5 μl template and 154 the specific primer pairs at a final concentration of 200 nM each. The reaction specificity was 155 determined for each reaction from the dissociation curve of the PCR product. The ribosomal protein 156 L7 (rpl7) and  $\beta$ -actin ( $\beta$ -act) genes were used as reference genes. Relative quantification of each gene expression was normalized according to the reference genes and generated using the  $2^{-\Delta\Delta Ct}$ 157 158 method described by [35]. Amplification efficiencies for all primer sets were calculated; all values

# 159 proved to be sufficient to allow direct comparison of amplification plots according to the $2^{-\Delta\Delta Ct}$

160 method.

#### 161 **Table 2**: Specific primer pairs used for RT-qPCR

Gene name Gene code		Function	Forward primer	Reverse primer	Efficiency (mean ± SEM)		
ribosomal protein I7	rpl7	Ribosomal components (reference gene)	5 <sup>′</sup> -GCATCTGGGCAATTACCATT -3′	5 <sup>′</sup> -CAAGGAACTGGCTACCAAGC -3 <sup>′</sup>	98.2 ± 8 %		
β-actin	β-actin	Cell structure (reference gene)	5 <sup>-</sup> -CTCTATCGTCCACCGCAAAT -3'	5 <sup>'</sup> - CACCTTCACCGTTTTCCAGT-3 <sup>'</sup>	93.4 ± 5 %		
clathrin b	cltb	Internalization	5 <sup>-</sup> GCAGAAGTCCCATAGGTCCA -3'	5 - AACACCGGAAGCGTAATGTC-3	91.2 ± 8 %		
superoxide dismutase zinc/copper	sod <sub>zn/cu</sub>	Oxidative stress (copper/zinc)	5 <sup>′</sup> -TCAAGGACAGAATGCTCACG -3 <sup>′</sup>	5 <sup>′</sup> -TGCTTTTGGGACACCCTTAC -3 <sup>′</sup>	90.8 ± 5 %		
superoxide dismutase manganese	sod <sub>mn</sub>	Oxidative stress (manganese)	5'- CGCCACATATGTCAACAACC-3'	5'- TAGGGGACAGGTTTGTCCAG-3'	106.2 ± 8 %		
metallothioneine	mt	Metal homeostasis detoxification	5'-CGAGCTGCTGTCAGTGAAGA -3'	5'- GCTCTGCATGGATGACAAAA-3'	88.7 ± 4 %		
ribosomal rna 12s	125	Mitochondrial metabolism	5 <sup>'</sup> -CTCACCATCCCTTGCCTAAA -3 <sup>'</sup>	5 <sup>-</sup> CTCAGAGCCGGTTTCAAAAG-3 <sup>-</sup>	91.7 ± 1 %		
Cytochrome c oxidase subunit 1	cox1	Mitochondrial metabolism	5 <sup>-</sup> -CCAGATGCATACACCCTGTG -3'	5 <sup>′</sup> -CGAATGTGTGGTATGGTGGA -3 <sup>′</sup>	91.3 ± 3 %		
atp synthase membrane subunit 6-8	atp6-8	Mitochondrial metabolism	5 <sup>-</sup> TGCAGTCCCAATATGACTCG-3'	5- ATGGCCTGCTGTTAGGTTTG-3	90 ± 4 %		
acetyl-coa carboxylase	асс	Lipid metabolism	5 <sup>-</sup> -GAGCAGTACAAGCCCGACAT -3'	5 <sup>-</sup> AGCATGGTCACCAGGATGTT -3 <sup>-</sup>	97.4 ± 8 %		
triglyceride lipase	tgl	Lipid metabolism	5 <sup>'</sup> -GACGCTGTTGTTCCAGATGA -3'	5 <sup>′</sup> - GGGAACACGTACGAAGCCTA-3 <sup>′</sup>	95.8 ± 10 %		
glyceraldehyde- 3-phosphate dehydrogenase	gapdh	Lipid metabolism	5 <sup>-</sup> TCTTTTGGGGGGATCAAACTG -3'	5-TGGCCTCAAAGGAGTAGAGC -3	99.3 ± 9 %		
vitellogenin	vtg	precursor of egg yolk	5 - CAGCTGCTTCTCAGCACAAG-3'	5 - AGCCTGAAGAACTGGCAGAA-3	91.7 ± 2 %		
bcl-2—associated x protein	bax	Cell cycle arrest/apoptosis	5 <sup>-</sup> - GCGACCCTAACCACAAGAAG-3'	5 <sup>′</sup> - GTCAAGCTCGTCTCCGATTT-3 <sup>′</sup>	91.2 ± 11 %		
tumor protein 53	p53	Cell cycle arrest/apoptosis	5 <sup>-</sup> GAGCCCAAGAAAGGGAAAAG -3'	5 - GACATGGAACTGACCCCAAC-3	98.9 ± 9 %		
growth arrest and dna damage- inducible 45	gadd45	DNA repair	5 <sup>-</sup> - CCGTTGTTGTCCACTGTGTC-3'	5 <sup>′</sup> -GATGAGGAAGACGAGGACGA -3 <sup>′</sup>	98.4 ± 12 %		
nlr family card domain containing 3	nlrc3	Immune system	5 <sup>-</sup> CCTCCAGAGTCCGTGTTAGC-3'	5 <sup>-</sup> TGCGCTTTCTGTCTGGTATG-3	98.7 ± 13 %		

162

# 163 **2.4. Statistical analysis**

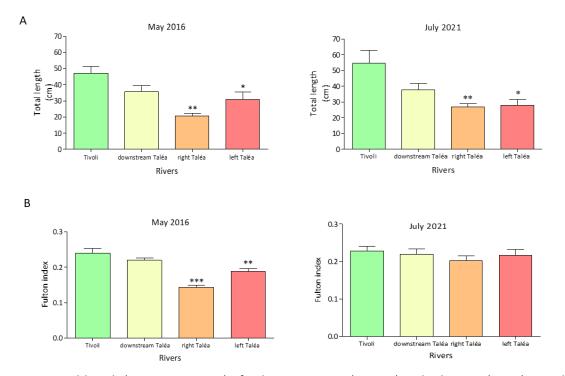
Data are expressed as mean ± Standard Error of the Mean (SEM) for n independent samples with five and four sampled eels in Left Taléa in 2016 and 2021 respectively, three in 2016 and five in 2021 in Right Taléa, five in 2016 and six in 2021 in Downstream Taléa and five in Tivoli in 2016 and 2021. Statistical analyses were performed using analysis of variance: one-way ANOVA followed by Tukey's post-hoc test for multiple comparisons, and t-test to compare each sites to reference site, Tivoli (\*P <0.05, \*\*P < 0.01 and \*\*\*P < 0.001) with GraphPad statistics software [36]. When assumptions necessary for using a parametric test were not met (normality, independence and homoscedasticity of the error term), we used non-parametric test. P values < 0.05 were considered statistically significant. Correlations between measured parameters on livers were revealed by principal component analysis using Statistica 13.3 software.

174 **3. Results** 

175

#### 176 **3.1. Morphometric parameters**

177 The analyses of A. marmorata morphometric parameters showed that eels living in the rivers downstream the mines are significantly smaller (\*\*P < 0.01 and \*P < 0.05 respectively for RT and LT 178 179 rivers) than eels from the reference site (Tiv), and this for the two periods (Fig.2A). Furthermore, eels 180 taken from the same rivers had approximately the same size in 2016 and 2021, for example in LT 181 they measured  $31.02 \pm 4.58$  cm in 2016 and  $27.75 \pm 3.1$  cm in 2021. The Fulton condition index did 182 not vary among sites in July 2021, but in May 2016, eels living in Taléa rivers had a lower Fulton index as compared to those from Tivoli (\*\*P < 0.01). However, the hepato-somatic index varied in the same 183 184 way for the two periods, with a significant lower value for eels living in Taléa rivers in 2016 (\*\*P < 0.01) and 2021 (\*P < 0.05) compared to those collected in Tivoli (Fig. 3). 185



187 Figure 2: A. Total length (cm, mean ± SEM) of eels in May 2016 (n = 4-6) and July 2021 (n = 5). B. Fulton

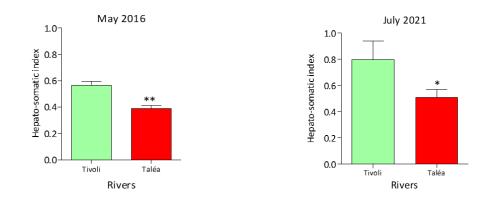
condition index (mean  $\pm$  SEM) of eels in May 2016 (n = 4-6) and July 2021 (n = 5). Means designated with star

189 (\*) are significantly different as compared to the reference Tivoli site (one-way analysis ANOVA followed by

190 Tukey's post-test, \* P < 0.05; \*\* P < 0.01).

191

188

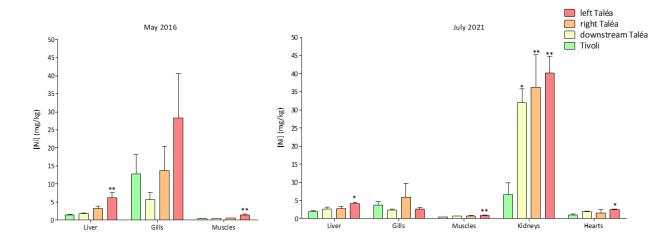


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Figure 3: Hepato-somatic index (mean ± SEM) of eels sampled in May 2016 (n = 5-13) and July 2021 (n = 5-9).
Means designated with star (\* or \*\*) are significantly different as compared to the reference Tivoli site (t-test,
\* P < 0.05; \*\* P < 0.01).</li>

## **3.2.** Metals bioaccumulation in fish organs

197 Metals concentrations were determined for Ni, Cr, Co and Mn in liver, gills and muscle at both 198 periods, 2016 and 2021. Kidneys and hearts were only analyzed in 2021. For Ni, the results showed hepatic concentrations significantly higher for eels living in Taléa river, particularly in LT, than those living in Tivoli, either in 2016 or 2021 (Fig. 4). For muscle in 2016 and 2021, and for kidneys and hearts in 2021, the same patterns were observed, with highest concentrations measured in eels from LT. The same tendencies were thus observed on both periods except for gills, for which Ni concentrations were largely higher in 2016 than in 2021 (28.34 mg/kg and only 2.58 mg/kg in LT respectively for 2016 and 2021).



205

Figure 4: Ni concentrations (mg/kg) in liver, gills, muscles, kidneys and hearts of eels sampled in Koniambo Mount in New Caledonia (mean  $\pm$  SEM, n = 3-5 in 2016 and n = 4-6 in 2021). Means designated with star (\*) are significantly different as compared to the reference Tivoli site (one-way analysis ANOVA followed by Tukey's post-test, \* P < 0.05; \*\* P < 0.01).

Eels collected from Taléa rivers accumulated also more Cr in their liver than eels living in Tivoli (p = 0.003 in 2016 and p = 0.024 in 2021) for the two periods of sampling (Fig. 5). Gills and muscle in 2016

bioaccumulated Cr to a greater extent in LT and kidneys in 2021 in all Taléa sites compared to Tiv.

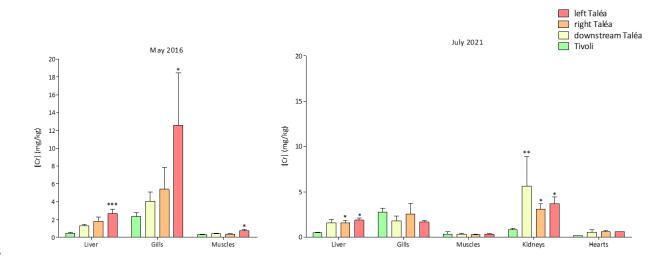
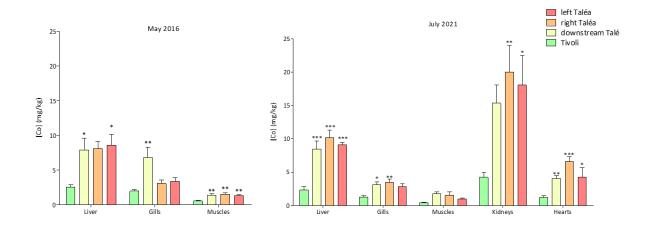


Figure 5: Cr concentrations (mg/kg) in liver, gills, muscle, kidneys and heart of eels sampled in Koniambo Mount in New Caledonia (mean  $\pm$  SEM, n = 3-5 in 2016 and n = 4-6 in 2021). Means designated with star (\*) are significantly different as compared to the reference Tivoli site (one-way analysis ANOVA followed by Tukey's post-test, \* P < 0.05; \*\* P < 0.01).

For Co, results showed concentrations significantly higher for eels living in Taléa rivers than those living in Tivoli for liver in 2016 and 2021 and for kidneys and heart in 2021 (Fig. 6). In 2016, the Co concentrations in muscle were higher in Taléa compared to Tivoli (p = 0.0034). For gills, only the concentrations measured in DT were significantly higher than the other sites in 2016, but in 2021, also those from RT bioaccumulated more Co than eels from Tiv.



223

Figure 6: Co concentrations (mg/kg) in liver, gills, muscle, kidneys and heart of eels sampled in Koniambo Mount in New Caledonia (mean ± SEM, n = 3-5 in 2016 and n = 4-6 in 2021). Means designated with star (\*) are significantly different as compared to reference site Tivoli (one-way analysis ANOVA followed by Tukey's posttest, \* P < 0.05; \*\* P < 0.01).

For Mn in 2016, no significant difference in concentrations was observed among sites, but in 2021 for heart and gills, Mn concentrations were higher in LT or RT respectively as compared to Tiv (p = 0.02 and p = 0.05 respectively) (Fig. 7).

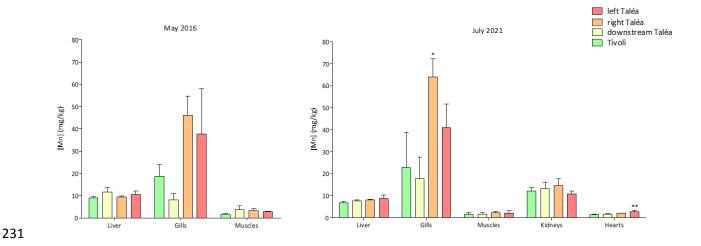


Figure 7: Mn concentrations (mg/kg) in liver, gills, muscle, kidneys and heart of eels sampled in Koniambo Mount in New Caledonia (mean  $\pm$  SEM, n = 3-5 in 2016 and n = 4-6 in 2021). Means designated with star (\*) are significantly different as compared to the reference Tivoli site (one-way analysis ANOVA followed by Tukey's post-test, \* P < 0.05; \*\* P < 0.01).

We can observe that the metal that is the most concentrated in the organs is Mn, then Ni. For Cr and
Co the concentrations are of the same order of magnitude in the organs, whereas the concentration
of Cr in the freshwaters is always higher than that of Co.

**3.3**.

#### . Gene transcription levels

240 To assess the impact of Ni and other associated metals from mining activity in organs of eels, RT-241 qPCR analyses were carried out on six organs: liver, kidneys, gills, brain, spleen and muscle from eels 242 sampled in 2021. Fifteen genes were analyzed to explore the internalization (cltb); the oxidative 243 stress (sod<sub>zn/cu</sub> sod<sub>mn</sub>); the metal detoxification (mt); the mitochondrial metabolism (12s; cox1 and atp6-8); the lipid metabolism (acc; tgl and gapdh); the reproductive function (vtg); the apoptosis 244 245 (p53 and bax); the DNA repair (gadd45) and the immune system (nlrc3). The results showed that liver 246 was the most impacted organ, as the transcription level of almost 12 genes on the 16 studied was 247 differentially regulated in eels from one or several Taléa sites compared to Tivoli (up-regulation for 248 oxidative stress and lipid metabolism and down-regulation for mitochondrial metabolism, DNA 249 repair, apoptosis, reproduction and immunity) (Table 3). **Table 3**: Fold changes in gene transcription levels in eels collected from Taléa river sites (n = 4-6 per site) as compared to eels from the reference Tivoli site. Only significant changes are reported (one-way analysis ANOVA followed by Tukey's post-test, \* P < 0.05; \*\* P < 0.01 as compared to Tivoli) (*e.g.* red = down regulation and green = up regulation).

		Genes and function															
Ultramafics rivers Or	Organs	Oxidative stress		Metal detoxification	Mitochondrial metabolism		Lipid metabolism			DNA repair	r Apoptosis		Endocytosis	Reproduction	Immunity		
		sodcu	sodmn	cat	mt	125	сох	atp6-8	gapdh	tg/	AcCoa	gadd45	p53	bax	cltb	Vtg	nlrc3
Taléa downstream		982.0* ± 243.5	14.27 <b>*</b> ± 4.11	/	/	/	/		16.85* ± 3.55	/	/	/	0.33* ±0.22	/	/	0.34* ±0.2	0.2* ±0.15
Taléa right	Liver	1189* ± 752.5	21.98* ± 8.62	/	/	0.20** ± 0.06	/	0.14** ± 0.053	24.4* ± 6.22	/	0.096* ±0.03	0.34 <b>*</b> ± 0.14	0.10° ±0.045	0.15* ± 0.044	/	0.093*±0.04	0.052*±0.03
Taléa left		1636*±513.6	20.95* ±4.1	/	/	0.18* ± 0.045	/	0.22* ± 0.06	36.38* ±11.4	62.02* ±42.4	0.046* ±0.016	0.14• ± 0.03	0.078* ± 0.035	0.13* ± 0.032	/	0.074* ± 0.052	0.03* ±0.018
Taléa downstream		1	1	1	,	1	1		1	1	,	1	,		1	,	1
	R. J.	/	/	/	/	/	/	/	/	/	/	/	/	/	/	/	/
Taléa right Taléa left	Kidney	/	/	/	/	/	/	/	/	/	/	/	/	/	/	/	2.052.10.45
Taleatert		/	/	/	/	/	/	/	/	3.03* ± 0.97	3.32*±0.9	5.92* ±1.72	/	2.59* ± 0.68	0.29° ± 0.85	/	2.06* ±0.46
Taléa downstream		/	/	1	/	/	/	/	/	/	/	/	/	0.32* ±0.13	/	/	/
Taléa right	Gills	/	/	/	/	/	/	/	/	/	/	/	/	0.45* ±0.12	/	/	/
Taléa left		24.15* ± 4.89	9.97 <b>*</b> ± 1.46	/	3.35* ± 1.43	/	/	/	/	/	/	/	0.36* ± 0.07	0.29* ± 0.035	/	/	/
Taléa downstream		/	/	/	/	/	/	/	/	/	/	/	/	/	/	/	/
Taléa right	Brain	/	/	/	/	/	/	/	/	/	/	3.8° ± 1.12	/	/	/	/	/
Taléa left		/	/	/	/	/	/	/	/	/	/	5.27* ±1.3	/	/	/	/	/
Taléa downstream	Spleen	/	/	/	/	/	/	/	/	/	/	/	/	/	/	/	/
Taléa right		/	/	/	/	/	/	0.36* ±0.12	/	/	/	/	/	/	/	/	/
Taléa left		/	/	/	/	/	/	0.52* ± 0.09	/	/	/	/	/	/	/	/	/

The kidneys showed only up-regulated genes involved in lipid metabolism, DNA repair, apoptosis,endocytosis and immunity.

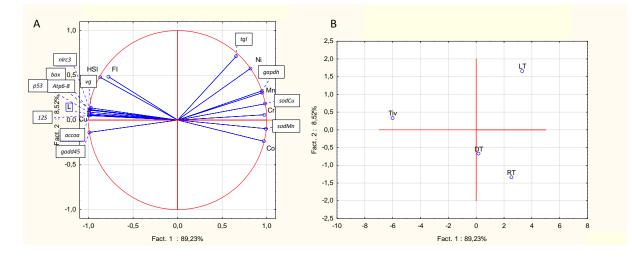
The gills were on the one hand up-regulated for genes involved in oxidative stress and metal detoxification and on the other hand down-regulated for genes involved in apoptosis.

Finally, brain and spleen are the less affected organs with only one gene that was differentially regulated (down-regulation of *atp6-8* for spleen and up-regulation of *gadd45* for brain).

257 No difference in gene transcription level was observed in muscles.

## 258 **3.4.** Multivariate analysis on the liver

259 As the liver appeared as the most impacted organ in eels living on ultramafic soils impacted by mining activities, a Principal Component Analysis (PCA) was done between the several studied 260 261 parameters - morphometric parameters, metals bioaccumulation and transcriptomic responses (Fig. 262 8). The PCA explained a total of 97.48% of the variability of the data, with the axis 1 the more 263 discriminant (89.23%) and corresponding to the most contaminated and disruptive stations on the 264 right (especially Left Taléa) and the less contaminated one on the left (Tivoli). Significant correlations 265 are observed between Ni, Mn, Cr and to a lesser extent Co with genes implicated in oxidative stress 266 and lipid metabolism on the most contaminated site (LT). On the contrary, morphological parameters 267 are correlated to genes implied in reproduction, in mitochondrial metabolism, in immunity, in cell 268 cycle regulation and correspond to the reference site Tiv.



270

Figure 8: Principal component analysis of morphometric parameters, metals bioaccumulation and gene
 transcription levels in liver of *A. marmorata*, after collection in July 2021 on the 4 sites: Tiv (Tivoli), LT (Left
 Taléa), RT (Right Taléa) and DT (Downstream Taléa). A. Correlation circle of the different studied parameters. B.
 Illustration of the sites.

#### 275 **4. Discussion**

276 This study aimed to assess the consequences for Anguilla marmorata to live in freshwaters 277 downstream Ni mines in New Caledonia, region particularly affected by mining activities due to the 278 composition of soils particularly rich in Ni. Eels are known to be sensitive to water metal 279 contamination [23] that could affect their migration and reproduction leading to a decline of their 280 populations [29, 30]. Until now, no study was conducted in New Caledonia on this species, whereas 281 eels are present in freshwaters as top fish predators and are potentially collected by local 282 populations to be eaten. Thus in this project, we collected eels in four sites (three on ultramafic soils 283 along the Taléa river, with one arm directly under mining influence, and one as a reference site, Tivoli 284 river) to assess the bioaccumulation of metals and their toxicity. The fish were sampled at two 285 different periods, in May 2016 and July 2021, in order to assess the influence of time on parameters 286 of interest and to verify whether the observed effects could be maintained over time.

287

#### 4.1. Morphometric parameters

288 First of all, we noticed that eels living in rivers downstream the mines were smaller than eels living in 289 the reference site, with a lower hepato-somatic index at both different periods, as it has been 290 already observed in eels exposed to other metals such as Hg and Cd [22]. These results seemed to 291 indicate that eels, in particular their liver function, are probably impacted by physico-chemical or 292 nutrient conditions occurring in Taléa river compared to Tivoli, whatever the period and season. 293 Interestingly, the Fulton index differed between the two periods. Indeed, no differences were 294 observed in 2021, whereas, in 2016, eels living in Taléa presented a smaller Fulton index compared to 295 Tivoli. In 2021, the sampling campaign was done after a long episode of rain which could explain at 296 least in part the low dissolved metal concentration in the rivers. For example, the concentrations of 297 dissolved Ni was 48  $\mu$ g/L in 2016 compared to 8  $\mu$ g/L few weeks before the fisheries in 2021. Indeed, 298 in 2016 the cumulative rainfall in the 6 months preceding the fisheries (December 2015 to May 2016) 299 was 930 mm of rain (an average of 5.2 mm of rain per day) but 2910 mm of rain (an average of 16.2 300 mm of rain per day) in 2021 (February 2021 to July 2021) (MeteoNC, 2016 ; MeteoNC, 2021). The higher dissolved metal levels in 2016, could further affect the food chain and eel nutrient quality, but 301 302 also eels themselves. As the Fulton index reflects the physiological state of the fish [37], it is possible 303 that in 2016, rivers under mining influence offered poor condition for growth and that eels inhabiting 304 these sites were more impacted by metals, and therefore, had a lower condition index.

305 4.2. Metals bioaccumulation

Due to the composition of laterite soils [12, 14], four metals were analyzed in eel organs: Ni, Cr, Co and Mn in 2016 and 2021. This kind of research has never been conducted in New Caledonia on *Anguilla marmorata*. At first, for all metals in 2016 and 2021, a dose-dependent contamination was observed according to the sites under mining influence or not, with eels from Taléa being significantly more contaminated than those from Tivoli. Those results are in line with the metal concentrations in the freshwater, which are much higher in Taléa river compared to Tivoli river.

312 The same trends were noticed in 2016 and 2021 but some variations especially for metal 313 concentrations in livers and gills were nevertheless observed. In fact, for Ni and Cr, in 2016, eels 314 appeared to be mainly contaminated by the direct route of exposure (*i.e.* dissolved metals in water) 315 whereas in 2021 the main route of contamination of eels in Taléa seemed to be the trophic route. 316 This was highlighted by higher metal concentrations in the gills compared to the livers for all sites in 317 2016 in contrast to 2021 where Ni and Cr concentrations in gills were lower than in livers. Indeed, 318 metals bioaccumulation in the gills is expected to be higher following waterborne exposure, while 319 higher bioaccumulation levels in the liver is followed by absorption of metals from the gut [38]. It 320 seems that few dissolved metals were assimilated by the gills in 2021, that could also be explained by 321 the moment of the fisheries, which were realized after heavy rain episodes in 2021. As already 322 evoked, these hydrologic conditions in 2021 would have probably diluted the metals in the 323 freshwaters by high leaching of soils during several months before sampling and modified their 324 chemical speciation, with a highest proportion of them adsorbed to particulate matter in 2021 325 compared to 2016 [39, 40]. All these findings could suggest that the contamination of eels is 326 dependent on the weather conditions and the fishing season. However, the livers are in all cases 327 impacted by higher concentrations of metals measured in eels living under mining influence 328 compared to Tivoli.

329 In contrast, for Co and Mn the same trends were observed for the two seasons, with contamination 330 mainly via the trophic way for Co and by the direct route for Mn. In 2016, the eels seemed to less 331 accumulate Mn than in 2021. Furthermore, in 2021, two other organs were analyzed: heart and kidneys. Kidneys are the organs that accumulate the highest levels of metals, as they represent the 332 333 main site of detoxification and excretion [41, 42]. Moreover, numerous studies have shown that for 334 the majority of metals, the highest concentrations in fish are found in the liver and kidneys [43-45]. 335 These observations are also reported in our study because for Ni, Co and Cr, kidneys are the organs 336 that bioaccumulated the most. On average, metal concentrations in the kidneys of Taléa eels were 12 337 to 16 times higher than those in Tivoli and more important than in gills, confirming the hypothesis of 338 trophic way of contamination. In order to ensure a good migration, eels need an optimal heart 339 capacity, thus, any cardiotoxicity could affect their migration [46, 47]. In 2021, we observed that 340 hearts were also affected, to a lesser extent because only eels of the LT site had significantly higher 341 Ni and Mn concentrations than eels living in the reference site, except for Co, for which all eels from 342 Taléa were impacted. Finally, concerning muscle, the accumulations remain very low compared to 343 the other organs, with generally higher values on the most impacted site (LT). These results are in 344 line with the literature which demonstrated that metals, including Ni, accumulates poorly in the fish 345 muscle [48]. Also in muscle, the concentrations are under the European recommendation threshold 346 for food, so the consumption for humans of Anguilla marmorata in New Caledonia seems totally 347 safe, at least for the metals analyzed in the present study.

348

## 4.3. Genes transcription

349 A clear link has been established between metals bioaccumulation in fish organs and alterations in gene transcription levels, especially those involved in the fight against oxidative stress [49]. Oxidative 350 351 stress is known to play a key role in the physiopathology of organs, notably liver, by altering the 352 calcium homeostasis and the mitochondrial function of hepatocytes [50, 51]. The qPCR analysis was 353 conducted only on the 2021 samples. The results suggest that livers and gills could be submitted to 354 an important oxidative stress (sod<sub>Cu</sub> and sod<sub>Mn</sub> inductions), biomarkers reflecting pollution of 355 freshwater ecosystems [52]. Oxidative stress could then be responsible of the mitochondrial 356 dysfunction observed with an alteration of the mitochondrial respiratory chain (MRC) in liver and 357 spleen (12S and atp6-8 depletions). These results are highlighted by the PCA analysis conducted on 358 the liver (Fig. 8), which demonstrated a clear link between oxidative stress, lipid metabolism and 359 bioaccumulation of Ni, Mn, Cr and to a lesser extent Co. In New Caledonia, just one study noted an 360 alteration of mitochondrial metabolism on goldfish after Ni exposition [53], such as in Pacific eels in the present study. These effects were associated to a dysregulation of lipid metabolism [54, 55], as 361 362 can be illustrated in our study by an induction of *qapdh* and *tql* in livers and kidneys, markers of an 363 increased utilization of glucose and triglycerides, which has previously been demonstrated in the 364 European eel Anguilla anguilla after Cd exposure [28]. This alteration could compromise the 365 efficiency of lipid storage necessary for their migration, and then, their reproductive success. 366 Furthermore, studies also suggested that triglyceride catabolism is a biomarker of oxidative stress in 367 eels [56] and could explain the lower hepato-somatic index of eels living under mining influence [57]. 368 In addition, we observed a decrease in the precursor of egg yolk, vtg, in the liver, which could be a 369 consequence of the lipid metabolism disruption and the oxidative stress observed [58]. Reproductive 370 impairment in eels living in polluted waters has already been reported in previous studies [26]. We 371 have to note that despite our sampling efforts, we failed to obtain fish of similar sizes among 372 sampling sites. This can be explained either by an effect of metals on fish growth rate or by variations 373 in fish age among sites [59]. However, even if we cannot exclude an effect of age or sex, we observed a significant link between metal concentrations in liver, hepato-somatic index and *vtg* transcription
levels, which suggests that contamination has probably an impact on reproduction and/or sex
determination [60, 61].

Another important fact is that eels living in contaminated freshwaters seem to have an important inflammation with *nlrc3* depletion in livers. On the contrary, an induction of *nlrc3* observed in kidneys indicate a decrease of pro-inflammatory response leading to a decrease of response if eels are in contact with virus, parasites or bacteria with potentially serious consequences about their health condition [62, 63].

382 Finally, the transcriptomic studies showed in liver and gills, a potential initiation of carcinogenic 383 pathways by down-regulating the transcription levels of genes involved in apoptosis and DNA repair (gadd45, p53 and bax). Indeed, the carcinogenic effects of metals in fish liver were already reported 384 385 in Solea vulgaris, Anguilla anguilla and Liza aurata [64, 65]. In contrast, the apoptosis pathway 386 seemed activated in kidneys (bax induction) with DNA repair, also reported in brain (gadd45 387 induction). The induction of apoptosis can remove potentially harmful cells, thereby blocking tumor 388 growth [66]. The kidneys may have activated this cell death process to protect against toxics effects 389 of metal contamination. Muscle were less impacted than other organs, accordingly to 390 bioaccumulation observations.

In New Caledonia, the mining activities seem then to alter the quality of life of eels living in the rivers downstream the mines, by organ function alteration, especially the liver. Due to the importance of liver function for eels, metallic contamination could contribute to hepatic dysfunctions leading to impairment of lipid metabolism and could then alter the success of their migration and maturation and, potentially their reproduction.

#### **5. Conclusion**

397 The current study shows that metal concentrations observed in freshwater downstream mining sites 398 may induce alterations of Anguilla marmorata organs in New Caledonia. Disruption of Fulton and 399 hepato-somatic index might be explained by the high metals levels found in kidneys and livers of 400 eels. In addition, metals triggered an important liver, kidneys and gills toxicity through the induction 401 of mitochondrial dysfunction, oxidative stress, inflammation, immunotoxicity and apoptosis or 402 carcinogenic effects. Therefore, the exposition of Anguilla marmorata to metals such as Ni, Cr, Co 403 and Mn in New Caledonia because of open pit mining activities on ultramafic basin catchments may 404 put the fish at risk of developing impaired migration and reproduction.

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#### 410 Statement of contributions

- 411 Germande Ophélie: Conceptualization, Formal Analysis, Methodology, Visualization, Writing -
- 412 original draft
- 413 Gunkel-Grillon Peggy: Methodology, Writing review & editing
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- 417 Dassié Emilie: Methodology, Writing review & editing
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- 419 Baudrimont Isabelle: Project administration, Conceptualization, Funding acquisition, Methodology,
- 420 Supervision, Validation, Writing review & editing
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