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Microbiome composition is shaped by geography and population structure in the parasitic wasp *Asobara japonica*, but not in the presence of the endosymbiont *Wolbachia*.

Running title: *Wolbachia* determines microbiome composition.

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Abstract

The microbial community composition is crucial for diverse life-history traits in many organisms. However, we still lack a sufficient understanding of how the host microbiome is acquired and maintained, a pressing issue in times of global environmental change. Here we investigated to what extent host genotype, environmental conditions, and the endosymbiont *Wolbachia* influence the bacterial communities in the parasitic wasp *Asobara japonica*. We sampled multiple wasp populations across ten locations in their natural distribution range in Japan and sequenced the host genome (whole genome sequencing) and microbiome (16S rRNA gene). We compared the host population structure and bacterial community composition of wasps that reproduce sexually and are uninfected with *Wolbachia* with wasps that reproduce asexually and carry *Wolbachia*. The bacterial communities in asexual wasps were highly similar due to a strong effect of *Wolbachia* rather than host genomic structure. In contrast, in sexual wasps, bacterial communities appear primarily shaped by a combination of population structure and environmental conditions. Our research highlights that multiple factors shape the bacterial communities of an organism and that the presence of a single endosymbiont can strongly alter their compositions. This information is crucial to understanding how organisms and their associated microbiome will react in the face of environmental change.

Keywords: Host microbiome interaction; Population structure; Environment microbe interaction; Reproductive mode; Geographical variation; Parasitoid wasp

Introduction

Microbes play essential roles in the ecology and evolution of their hosts (McFall-Ngai et al., 2013) by acting as modulators of host phenotypes (Brinker, Fontaine, Beukeboom, & Falcao Salles, 2019; Corbin, Heyworth, Ferrari, & Hurst, 2017; Moran, McCutcheon, & Nakabachi, 2008). The community of host-associated microbes, the microbiome, shape host phenotypes, and impact host fitness (Koch & Schmid-Hempel, 2012; Shropshire & Bordenstein, 2016; Wong, Dobson, & Douglas, 2014). Moreover, variation in microbiome composition, as well as symbiont presence and prevalence, has been linked to host disease susceptibility (Plottel & Blaser, 2011), metabolism (Ridley, Wong, Westmiller, & Douglas, 2012), behaviour (Sharon et al., 2010), and reproduction (Moran et al., 2008). Despite the importance of microorganisms for eukaryotic host functioning, we still lack a good understanding of how various factors, alone or in interaction, influence and shape host-microbiome composition. Understanding these factors in diverse systems will help develop a general framework on how microbial communities are shaped. This is crucial for developing predictive models on how a microbial community will react to future environmental changes, such as global warming, urbanisation, or land-use change, with potentially strong percolating effects on host fitness (Moran et al., 2008; Plottel & Blaser, 2011; Ridley et al., 2012; Sharon et al., 2010).

The host microbiome is shaped by factors associated with the host genotype (Jaenike, 2009; Kolasa et al., 2019; Rennison, Rudman, & Schluter, 2019), the environment (Colman, Toolson, & Takacs-Vesbach, 2012; de Vries, Jacobs, Sabelis, Menken, & Breeuwer, 2004; Duan et al., 2020; Ferguson et al., 2018; Ochman et al., 2010; Russell & Moran, 2006) and their interactions (van Veelen, Falcão Salles, Matson, van der Velde, & Tieleman, 2020). For instance, microbial communities of closely related host species are often similar and display

a host phylogenetic signal, i.e., host-microbe community compositions correlate with host phylogenies (Kolasa et al., 2019; Lim & Bordenstein, 2020). Moreover, population-specific microbiomes were found for terrestrial isopods (Bouchon, Zimmer, & Dittmer, 2016; Dittmer et al., 2014), fruit flies (Rudman et al., 2019), and humans (Falony et al., 2016). In addition, environmental factors influence the host microbiome, such as the geographical location of an individual, food availability, as well as temperature, and seasonality (Uren Webster et al., 2020; van Veelen, Falcão Salles, & Tieleman, 2017; Yun et al., 2014). For example, the population-specific microbiome of terrestrial isopods was only maintained in the wild but not laboratory-reared populations (Bouchon et al., 2016; Dittmer et al., 2014). Additionally, free-living microbes, present in water, soil, or air, as well as microbes on plants or other organisms, influence the community composition of a host (van Veelen et al., 2020). Finally, the presence of specific symbionts can influence the host microbiome, as reported for Lepidoptera (Johnston & Rolff, 2015), *Drosophila melanogaster* (Simhadri et al., 2017; Ye, Seleznev, Flores, Woolfit, & McGraw, 2017), and the mosquito *Aedes aegypti* (Audsley, Ye, & McGraw, 2017), potentially acting as a stabiliser of the microbial community (Herren & McMahon, 2018). These symbionts are, in turn, affected by environmental conditions. For example, an increase in temperature can reduce or eliminate the endosymbiont *Wolbachia* in fruit flies, mites, and other invertebrates (Corbin et al., 2017; Hurst, Johnson, Schulenburg, & Fuyama, 2000; Sumi, Miura, & Miyatake, 2017; van Opijnen & Breeuwer, 1999). Moreover, seasonal fluctuations in *Wolbachia* density have been documented in Lepidoptera (Sumi et al., 2017), mosquitoes (Novakova et al., 2017), and other blood-sucking arthropods (Cohen, Toh, Munro, Dong, & Hawlena, 2015). These studies point toward a complex interaction between symbionts, the other members of the microbial community, and the environment (Dittmer & Bouchon, 2018; Easson, Chaves-Fonnegra, Thacker, & Lopez, 2020; Jehrke, Stewart, Droste, & Beller, 2018).

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Here we investigate whether and to what extent host population genomic structure, environmental conditions (i.e., geographic location with differing abiotic and biotic factors), and the presence of a specific endosymbiont influence the bacterial community in the parasitic wasp *Asobara japonica*. *A. japonica* (Hymenoptera, Braconidae: Alysiinae) is a solitary endoparasitoid of *Drosophila* larvae and related genera, native to Japan and Southeast Asia. In Japan, two types of populations of this wasp exist. Wasps on the northern main islands carry the endosymbiont *Wolbachia*, a common bacterium in invertebrates, particularly insects (Zug & Hammerstein, 2012). *Wolbachia* infection causes asexual reproduction (thelytokous parthenogenesis) and all-female offspring in *A. japonica* (Kremer et al., 2009). *Asobara japonica* wasps have a haplodiploid reproductive system, like all Hymenoptera. In the absence of *Wolbachia*, unfertilised haploid eggs develop into males and fertilised diploid eggs into females. In contrast, *Wolbachia* infected females lay only unfertilised eggs, which develop into (infected) diploid daughters rather than haploid sons (Ma et al. 2015). On the southern islands of Japan, with a more subtropical climate, populations are sexual and uninfected with *Wolbachia*, having both males and females (Mitsui, van Achterberg, Nordlander, & Kimura, 2007). This allows us to assess the importance of geographic distribution, environmental conditions, population structure, and symbiont presence in shaping the host microbiome within a single species with two reproductive modes. For this study, we collected wasps of both reproductive modes from several localities spanning the distribution range of *A. japonica* in Japan. We determined population genomic structure using low-coverage whole-genome sequencing (lcWGS) and bacterial community composition using 16S rRNA metabarcoding. Given the documented strong influence of *Wolbachia* on host bacterial communities in multiple species (Simhadri et al., 2017; Ye et al.,

2017), we expected to find different bacterial communities in infected (asexual) and uninfected (sexual) wasps. Moreover, we predicted that the effect of host genetic background on the bacterial community would be more visible in the sexual wasps, given their higher genetic variance compared to asexual reproducing wasps.

Material & Methods

Wasp collection

Asobara japonica was collected from 10 locations in Japan, spanning a 4,000 km north-south gradient in June 2017 (Fig. 1). Asexually reproducing wasps infected with *Wolbachia* were collected from seven locations at the northern main islands Kyushu and Honshu (from south to north: Kagoshima, Fukuoka, Kobe, Kyoto, Kanazawa, Tokyo, Sendai). Sexual reproducing wasps uninfected with *Wolbachia* were collected from three islands in the south of Japan, Iriomote, Okinawa, and Amami. Collection locations were grouped into seven climatic zones based on climatic areas defined by the Japan Meteorological Agency (JMA; <https://www.jma.go.jp/jma/indexe.html>; Table S1, Fig. 1). The northern main island locations have a temperate humid climate, with cold, wet winters in Sendai, Kanazawa, Tokyo, Kobe, and Kyoto and dry winters in Fukuoka and Kagoshima. The sampled southern islands have a subtropical climate. Vegetation ranges from an evergreen broadleaved forest zone on the main islands to subtropical hardwood forests in the south (Miyawaki, 1984).

Asobara japonica is a parasitoid of fruit flies (*Drosophila spp.*). Adult wasps lay their eggs into 2nd instar *Drosophila* larvae, in which the wasps develop until hatching after approximately 19 days at 25°C. Therefore, the sampling of *A. japonica* was performed as follows (Fig. 2). Traps containing banana pieces and approximately three to four gram yeast were placed in bushes and trees at six points for each location to attract fruit flies (*Drosophila spp.*). Traps were separated by at least 300 to 600 meters at each collection site. Attracted *Drosophila* flies laid eggs in the baits, which developed into larvae after 2-3 days. *Asobara japonica* and other parasitoids were then attracted and laid eggs into these *Drosophila* larvae. Baits were collected after 6-7 days, developing *Drosophila* larvae were separated from the bait material

and placed in tubes containing a layer of agar. Tubes were brought to the laboratory in the Netherlands and stored in an incubator at 25°C with a light-dark cycle of 16 hours light and 8 hours darkness (LD16:08) until wasps hatched (Fig. 2). Emerging wasps were identified to species level based on morphological features (Guerrieri, Giorgini, Cascone, Carpenito, & van Achterberg, 2016; Mitsui et al., 2007).

Emerged *A. japonica* wasps (G0) were cultured on *D. melanogaster* ww-strain (Leiden, the Netherlands) at 25°C under LD16:08 as follows. Hatched wasps were fed honey and allowed to mature for three days, after which individual females were placed in agar bottles coated with a thin layer of yeast (AB Mauri S.p.A., Italy) containing 2nd instar *D. melanogaster* larvae for oviposition. For sexual lines, an additional 2-3 males of the same line were added to assure that females mated. G0 females were collected after 48 hours of parasitisation and stored at -80°C for population genomic analysis. Offspring emerging after parasitisation (G1) were collected in glass bottles with agar and fed honey (Fig. 2). After three days, three to four individuals were used for hosting a new generation, and the remaining wasps were stored at -80°C for the microbiome analyses.

Host DNA extraction, sequencing, and bioinformatic analysis

The genetic diversity and population structure of the sampled wasps were determined using low coverage whole genome sequencing (lcWGS). Genomic DNA from individual wasps hatched from the collected field material (generation G0, 3 sexual locations, n=13 per location; 7 asexual locations, n=8 per location; total n=95) was extracted using a CTAB protocol (Chen, Rangasamy, Tan, Wang, & Siegfried, 2010). DNA was converted into genomic fragment libraries for Illumina 150 bp paired-end lcWGS (4x read depth) by SNPsaurus, LLC.

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First, DNA was fragmented with Nextera reagent (Illumina, Inc), which ligates short adapter sequences to the fragment ends. The Nextera reaction was scaled for fragmenting 10 ng of genomic DNA, which was then amplified for 12 cycles. Libraries were paired-end sequenced on one lane of an Illumina HiSeq 4000 (2x150bp, University of Oregon). Read cleaning, mapping, and SNP calling were conducted by SNPsaurus LLC using BMAP tools v. 38.08 (<http://sourceforge.net/projects/bbmap/>; see supplementary methods for the used parameters). All reads were mapped to the reference genome of the asexual KG strain of *A. japonica* (Ma et al., 2021; BioProject: PRJNA661661, accession number: JADHZF000000000) with an alignment identity threshold of 0.95 using *bbmap* (BMAP tools).

We used *ANGSD* v.0.93 (Korneliussen, Albrechtsen, & Nielsen, 2014) for further processing and cleaning the population genomic dataset. Single nucleotide polymorphisms (SNPs) were called in *ANGSD* (Korneliussen et al., 2014) using the GATK model (GL 2), taking genotype uncertainty into account by calculating genotype likelihoods. Sites with a phred base quality score lower than 15 were excluded to remove potential erroneous SNP calls (-minMapQ) if not stated otherwise. Then, SNPs with a mean read depth lower than 2X and over 30X were excluded (-setMinDepth 190; i.e., an average of 2 per individual; setMaxDepth 2850; i.e., an average of 30 per individual). A minor-allele frequency (MAF) threshold was set to 10% (-minMaf 0.1) to remove rare variants and putative additional sequencing errors. We removed triallelic SNPs (-skipTriallelic) and only included high-confidence variants by setting the -option -SNP_pval to 1e-6. Additionally, only SNPs occurring in a minimum of 70% of the individuals (i.e., sites occurring in 67 individuals out of the 95) were kept (-minInd). Finally, the Neighbour-joining tree and the population structure (PCA, PCA admix) analyses required a linkage disequilibrium (LD-) pruned SNP dataset, keeping only unlinked SNPs. First, we

inspected LD-decay among SNPs, considering SNPs that are at most 50kb distant from each other (Fig. S1) using *ngsLD* (Fox, Wright, Fumagalli, & Vieira, 2019). Then, we generated an LD-pruned SNPs dataset considering an r^2 threshold of 0.6 between every pair of SNPs up to a distance of 20 kb using *ngsLD* (Fox et al., 2019). The resulting filtered high-quality SNP data set included 88,485 out of 130,571 SNPs in a total of 95 samples and was used for further population genetic analyses.

Host population genetic diversity and structure

Genome wide heterozygosity was calculated per individual using the folded 1d-SFS in *ANGSD* (Korneliussen et al., 2014) specifying the following filter values: -minQ 20 - minMapQ 20 - uniqueOnly 1 -baq 1 -remove_bads 1 -setMinDepth 2 -setMaxDepth 30. Per-individual inbreeding coefficients were estimated with *ngsF* (Nielsen, Korneliussen, Albrechtsen, Li, & Wang, 2012; Vieira, Fumagalli, Albrechtsen, & Nielsen, 2013) with the genotype likelihoods generated by *ANGSD* (Korneliussen et al., 2014) using the flag "-doGlf 3". *NgsF* was first run using the approximated EM algorithm to obtain an approximation of the parameters, which were then used as initial values for the main algorithm. The approximate and the main algorithms were run ten times with 5000 iterations. The flag "--min_epsilon" was set to 1e-10 as recommended for low coverage data (<https://github.com/fgvieira/ngsF>).

Population structure was first investigated by partitioning the genetic variance using a principal component analysis (PCA; Patterson, Price, & Reich, 2006) based on genotype likelihoods using *PCAngsd* (Meisner & Albrechtsen, 2018). The contribution of each principal component (PC) was calculated in R and plotted using *ggplot2* (Wickham, 2016). Based on genotype likelihoods, a pairwise genetic distances matrix was calculated using the *ngsDist*

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package (Vieira, Lassalle, Korneliussen, & Fumagalli, 2016) based on genotype likelihoods to create a mid-point rooted NJ tree. The NJ tree was inferred using *FASTME* v.2.1.5. (Lefort, Desper, & Gascuel, 2015) and midpoint rooted with the R package *phytools* v.0.7.47. Node support was estimated with 1,000 bootstrap resampling using *ngsDist*, and support values were placed on the main tree using *RaxML* v.8.2.11 (Stamatakis, 2014). Finally, the tree was plotted using the R package *ggtree* v1.4 (Yu et al., 2018).

Genetic structure and individual wasp ancestry proportions were further investigated based on the approach implemented in *PCAngsd* (Meisner & Albrechtsen, 2018), relying on the estimated individual allelic frequencies (Meisner & Albrechtsen, 2018). We ran the analysis using five eigenvalues, which capture 61.8% of the observed genetic variation. We set the maximum number of iterations parameter to 500 and alpha, the sparseness regularization parameter, to 50. We tested different numbers of plausible clusters (K) ranging from 1 to 8. The most likely number of genetic clusters was estimated by visually inspecting newly created clusters with increasing K values (Vercken et al., 2010).

The *realSFS* application (Nielsen et al., 2012) in *ANGSD* (Korneliussen et al., 2014) was used to calculate the weighted F_{ST} values between pairwise locations and between the found genetic cluster (Fumagalli et al., 2013). The site frequency spectrum (SFS) and genotype likelihoods were estimated using the GATK model in *ANGSD* (Korneliussen et al., 2014), keeping only sites occurring in a minimum of 80% of the individuals, with a minimum base and mapping quality of 20, and coverage between two and 30 per individual. This procedure was used to estimate the one-dimensional SFS for each location or cluster, to then estimate the two-dimensional SFS for all pairs of locations or clusters. The weighted F_{ST} (Reynolds,

Weir, & Cockerham, 1983) was then calculated in a sliding window of 10kb for each pair of populations directly from the folded 2d-SFS. We applied the "-whichFST 1" option because it performs better with small sample sizes (Bhatia, Patterson, Sankararaman, & Price, 2013). Finally, we estimated the 95% confidence interval using the R package *Rmisc* v1.5 (Ryan, 2013) and reported them together with the mean F_{ST} values observed across sliding windows.

Microbiome DNA extraction and bacterial community composition

To determine the bacterial community composition, ten female wasps hatched from the field material (G1) were pooled, as single wasps did not yield sufficient amounts of microbial DNA. This way, three replicates per location of asexual (*Wolbachia*-infected) wasps (7 locations: n=21) and seven replicates per location of sexual (uninfected) wasps (3 locations: n=21) were created. Before DNA extraction, pooled wasps were washed (1 min. in 70% ethanol and 3x in sterile water). Wasp microbial DNA was extracted and purified using the DNeasy Power Soil[®] DNA Isolation Kit, following the manufacturer's protocol (Power Soil[®], MoBio Laboratories Inc., CA, USA). In deviation from the protocol, wasps were snap-frozen in liquid nitrogen and crushed with a sterile pestle before being added to the homogenisation tubes. Additionally, wasp material was homogenised with a grinder (Kaiser, Germany) for 15 min and centrifuged (Minispin, Eppendorf AG, Hamburg, Germany). DNA was eluted in 100ul C6 solution (provided in the kit) and stored at -20°C for further analysis.

Samples were sent for sequencing (Illumina MiSeq, 2x300bp, 25million reads per run, 16S rRNA gene region V3-V4, Gatesoupe et al., 2016; Vastra, Salvin, & Roos, 2016) at the INRAE Toulouse Center (France). Demultiplexed paired-end reads were processed following the DADA2 pipeline v.1.18.0 (Callahan et al., 2016) in R v.4.0.2. First, reads were checked for

quality by visualizing the quality score at each base position for each sample for the forward and reverse reads separately. Positions with a lower mean quality score than 30 were in the following step removed. For this, reads were trimmed as follows, the first 15 nucleotides of forwards and reverse reads were clipped, the forward reads were truncated to 250 bases, and the reverse reads to 230 bases using the *filterAndTrim* function in *dada2*. Additionally, reads were truncated at the first instance of a quality score less than or equal to two (*truncQ=2*). Reads were further filtered using the same function. Expected errors, defined by the nominal definition of the quality score $EE = \sum(10^{-(Q/10)})$, were removed (*maxEE=2,2*). As well as reads matching to the phiX genome (*rm.phix=TRUE*) and sequences containing "n" (*maxN=0*). Error rates were estimated with a parametric error model on a subset of data using the *learnErrors* function and visually checked using the *plotErrors* function. During dereplication, we condense the data by collapsing together all reads that encode the same sequence. For this, identical sequencing reads were identified and grouped into "unique sequences" with a corresponding "abundance" for the following dereplication step using the *derepFastq* function. Estimated error rates and list of unique sequences were then used to dereplicate the data using the *dada* function. Forward and reverse reads were subsequently merged to obtain the full denoised sequences. Finally, chimeric sequences were removed using the *removeBimeraDenovo* function.

Taxonomy was assigned using the pre-trained classifier Silva 132 taxonomy classifier (clustered at 99% similarity; Callahan, 2018; Quast et al., 2013). The *DADA2* pipeline created an amplicon sequence variants (ASV) table and a table containing the assigned taxonomy, while an unrooted neighbour-joining (NJ) tree was created using the *phangorn* v.2.7.1 R package (Schliep, 2011; Schliep, Potts, Morrison, & Grimm, 2017). These output files, ASV

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table, taxa table, NJ tree, and a table containing metadata, such as sample location, reproduction, and environmental factors, were then used to create a phyloseq object (R package *phyloseq* v.1.34; McMurdie & Holmes, 2013). Sequences identified as chloroplast, mitochondria, or Archaea, and uncharacterised reads at the phylum level were removed from the phyloseq object. The dataset was rarefied to 11,993 reads based on visualisation by a rarefaction curve (*ggrare* in the *phyloseq* R package) and determination of the lowest number of reads in all samples ($\min(\text{sample_sums}(x))$) using the *phyloseq* function *rarefy_even_depth* before analysis. In addition, out of this rarefied dataset (*Wolbachia* +), a second phyloseq object was created, in which reads belonging to the genus *Wolbachia* were removed (*Wolbachia* -). Taxa abundance of the *phyloseq* object without *Wolbachia* (*Wolbachia* -) was normalised following the *edgeR* method (*microbiomeSeq* R package v.0.1; Ssekagiri, 2020) prior to ordination analyses. This second phyloseq object was used to check how the dominance of *Wolbachia* influences following analyses, as highly dominant species could influence (multivariate) analyses. All statistical analyses were analysed separately for each dataset (rarefied; *Wolbachia* +) and rarefied without reads from *Wolbachia* (*Wolbachia* -).

Normality and homogeneous variances of the ASV's and Shannon diversity were tested with the Shapiro Wilks test (Shapiro & Wilk, 1965) and fligner test (Conover, Johnson, & Johnson, 1981). The observed number of ASV's and Shannon diversity of the bacterial community of *A. japonica* were tested against reproductive mode (asexual, infected with *Wolbachia* or sexual, uninfected) in an ANOVA. Variation in community composition (beta diversity) among samples was visualised via a principal coordinates analysis (PCoA) based on Bray–Curtis distance dissimilarity matrices (*ordinate* function in *phyloseq*). Differences between the reproductive modes (asexual, sexual), the effect of location, and population genetic structure

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were tested by permutational multivariate analysis of variance using the function "*adonis2*" with 999 permutations (PERMANOVA, *vegan* R package v.2.5-7; Anderson & Willis, 2003; Oksanen et al., 2016). Moreover, environmental factors were fitted on the ordination with 999 permutations to investigate the influence of environmental conditions on the bacterial community composition using the "*envfit*" function (*vegan* R package v.2.5-7; Anderson & Willis, 2003; Oksanen et al., 2016). Average distances to the group centroid per explaining variable (population cluster, climatic zones, and reproductive mode) were plotted to establish which variable drives clustering patterns (*betadisper* function in *vegan*). Shared and unique taxa found between the reproductive modes and genetic cluster were displayed in a Venn diagram depicting taxa count and percentage (*ggVennDiagram* R package v.1.1.4; Gao, Yu, & Cai, 2021). We used the R package *DESeq2* version 1.36.0 (Love, Huber, & Anders, 2014) to identify differentially abundant taxa across the reproductive modes. We created a *deseq* object using the *phyloseq* function "*phyloseq_to_deseq2*" with reproductive mode as factor and the location Sendai removed due to the high number of zeros. Differential expressed taxa were detected using the function "*DESeq*" with "test=Wald", and the 20 most changed taxa at the genus level were visualised via a boxplot. Given the observed high number of zeros in the ASV table of Sendai, we decided to rerun all previously described analyses without this location. Removing samples from the location of Sendai did not change the observed patterns, so we decided to keep this location in our analyses.

Correlation of host genomic background with host-associated bacterial community

We correlated the pairwise genetic distance of the wasps hatched from the field material (G0) to the bacterial community of their respective offspring (G1) to analyse the influence of *A. japonica* genetic ancestry and population structure on the bacterial community composition.

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Only samples with the genetic data of the mother (G0) and the bacterial data of their offspring (G1) were included. This resulted in 42 samples, including seven replicates for each of the three locations harbouring sexual wasps (n=21) and three replicates for each of the seven locations harbouring asexual wasps (n=21). This subset was used to correlate pairwise genetic distance and the bacterial community. The pairwise genetic distance was calculated using the *ngsDist* package (Vieira et al., 2016). Community composition (Bray–Curtis dissimilarity) and host genetic distance, as well as Shannon diversity of the bacterial community and heterozygosity as genetic diversity measure, were plotted against each other, and a regression line was fitted (lm function of basic R package *stats*). Correlations were tested using a Spearman-rank correlation test (*cor.test* function of basic R package *stats*).

Results

Genetic diversity and population structure

Population genetic structure of the collected wasps across Japan was assessed using 88,485 high-quality SNPs. Overall samples showed low heterozygosity in the sexual (mean heterozygosity \pm S.E.M.: 0.004 ± 0.002) and even lower in the asexual wasps (0.001 ± 0.002 , see Fig. S2A for details per location), which is in line with the general low heterozygosity reported for haplodiploid insects (Menken, 1991; Metcalf, Marlin, & Whitt, 1975; Packer & Owen, 2001). A high inbreeding coefficient was observed in asexual (mean inbreeding coefficient \pm S.E.M.: 0.70 ± 0.01), which was not found in sexual wasps (0.28 ± 0.17 , see Fig. S2B for details per location).

The first two principal components (PCs) of the principal component analysis (PCA) captured 48.5% of the total variation and partitioned the samples into five distinct groups (Fig. 3A). The first PC discriminates the samples according to their reproductive mode (asexual, *Wolbachia*-

infected) vs. sexual (uninfected) and to a lesser degree population structure in sexual locations. Both clusters reveal additional sub-structuration along the first and third PCs (Fig. 3B). In particular, the population from Iriomote, the most distant island from the main island, separated from the other two populations (Amami and Okinawa), which also showed a separation along the first two PCs. The UPGMA tree revealed the five same clusters as the PCA (Fig. 3C). On the other side, the asexual populations differed between the northern (Sendai, Kanazawa, and Tokyo) and more southern (Kyoto, Fukuoka, and Kagoshima) locations with individuals from Kobe distributed among these two groups.

We further investigated the wasp population genetic structure by estimating the most likely number of genetic clusters and the individual ancestry proportions of each cluster using *PCAngsd* (Meisner & Albrechtsen, 2018), indicating five distinct genetic populations (Fig. 3D; see Fig. S3 for K2 to K5). The level of admixture was minimal among clusters, except for four individuals from the Kobe location displaying admixed ancestry between the northern and southern groups on the main islands of Japan and one sample of the southern island Okinawa. Thus, asexual reproduction induced by the endosymbiont *Wolbachia* on the main islands, and most likely the geographic isolation of the uninfected wasps in the southern islands, led to very low admixture among the distinct genetic populations. Average heterozygosity was low for all genetic clusters, especially clusters involving asexual populations (clusters with sexual locations: $K1=0.004\pm0.002$, $K2=0.005\pm0.002$, $K3=0.003\pm0.002$; clusters with asexual locations: $K4=0.0008\pm0.0005$, $K5=0.0002\pm0.0003$ SEM). The inbreeding coefficient, on the other hand, was high for all clusters, apart from the two sexual cluster K2 and K3 (clusters with sexual locations: $K1=0.48\pm0.1$, $K2=0.18\pm0.11$, $K3=0.18\pm0.07$; clusters with asexual locations: $K4=0.71\pm0.09$, $K5=0.69\pm0.01$). Subdivision into five genetic clusters was also supported by

high F_{ST} values between locations ranging between 0.3 to 0.4, with the extremes of 0.15 between Okinawa and Amami, two of the sexual locations, and up to 0.72 between the most southern sexual location and the asexual locations (Table S2). This was also reflected when looking at the clusters (Table S3).

Overall, all our analyses revealed five distinct genetic clusters (Fig. 3). The southern islands harbouring sexual reproducing wasps are separated into three genetic clusters, with cluster K1 for Iriomote, cluster K2 Okinawa, and K3 for Amami. The locations harbouring asexual wasps were grouped in two clusters separating the southern locations Kagoshima, Fukuoka, and Kobe (K4), from the northern locations Kyoto, Kanazawa, Tokyo and Sendai (K5).

Bacterial community composition

The bacterial community composition covering all locations (3 sexual locations, n=7; 7 asexual locations, n=3; total n=42) revealed 566 taxa in a total of 776,870 reads, with 461 taxa assigned to genus level (81.5%). After rarefaction to 11,993 reads, the minimum reads per sample, 535 taxa remained with 446 taxa assigned to genus level (83.4%). Removal of reads belonging to the genus *Wolbachia* from the rarefied dataset led to 523 taxa, 434 of which could be assigned to genus (82.9%) in 280,433 reads.

Overall, 56% of taxa (301 out of 535) were shared between the two reproductive modes (Fig. S4). However, the sexual wasps yielded a higher number of unique taxa (35%) than the asexual wasps, with only 9% percent unique taxa (Fig. S4A). The five genetic clusters distinguished by the population genetic analyses (Fig. 3) shared a total of 134 taxa (25%), with sexual populations having a higher number of unique taxa (7% for K1 and K3; 12% for K2)

compared to asexual populations (4% and 2% for K5 and K4, respectively) (Fig. S4B). Wasps from the three sexual clusters also shared fewer taxa (188, 4%; Fig. S4C) than wasps from the two asexual clusters (208, 6%; Fig. S4D). The 20 most changed taxa, on a genus level, are displayed in the supplement (Fig. S5).

Wolbachia infection significantly affected bacterial taxonomic biodiversity, leading to a lower richness and diversity in asexual individuals than in sexual ones (Fig. 4A, ANOVA: Observed: $F_{1,40}=41.67$, $p<10^{-6}$; Shannon: $F_{1,40}=125.2$, $p<10^{-6}$). After removing reads belonging to the genus *Wolbachia*, taxonomic richness was still higher in uninfected, sexual individuals (Fig. 4B, ANOVA: Observed: $F_{1,40}=50.89$, $p<10^{-6}$). However, removing the highly abundant symbiont led to a similar Shannon diversity between uninfected and infected individuals (Fig. 4B, Shannon: $F_{1,40}=0.26$, $p=0.6$).

A permutational multivariate analysis of variance revealed a strong effect of reproductive mode on bacterial community composition (Fig. 4C, ADONIS: $F_{1,41}=42.3$, $p=0.001$, $R^2=0.49$). This analysis also revealed that bacterial community composition is in addition to reproductive mode driven by genetic background (Fig. 4C, ADONIS: Cluster: $F_{3,41}=3.4$, $p=0.001$, $R^2=0.12$) and to a lesser degree climatic zones (Fig. 4C, ADONIS: Zones: $F_{5,41}=0.14$, $p=0.045$, $R^2=0.01$). Indeed, environmental factors significantly explained some of the found clustering, next to reproductive mode (Fig. S6A; Table S4). Visualisation of the bacterial community composition according to the genetic background in a PCoA-plot revealed no additional separation of clusters apart from reproductive mode (Fig. S7A), which was also not revealed when the reproductive modes were plotted separately (Fig. S8A). However, the effects of environmental conditions and population structure are difficult to disentangle

because reproductive mode and environment covary. The average distance to the group centroid per explanatory variable (reproduction: asexual/sexual; population identity: genetic cluster 1 to 5; climate zones 1-7) also revealed that genetic background and climatic zones aligned with reproductive mode (Fig. S9).

Principal Coordinates Analysis (PCoA) based on Bray-Curtis distances confirmed that reproductive mode is indeed a major factor determining bacterial community composition, with the first two PCoA axes explaining 63.5% of the variation (Fig. 4C). After removing *Wolbachia* reads, the distinct clusters of asexual samples became more diffuse, with the first two PCoA axes explaining 43.9% of the variation (Fig. 4D). However, a clear separation between asexual and sexual populations (Fig. 4D, ADONIS: Reproductive mode: $F_{1,41}=16.02$, $p=0.001$, $R^2=0.25$), clustering according to population structure (Fig. 4D, ADONIS: cluster: $F_{3,41}=2.7$, $p=0.001$, $R^2=0.12$), and climatic zones (Fig. 4D, ADONIS: Zones: $F_{5,41}=1.36$, $p=0.003$, $R^2=0.11$) was still found. This indicates that environmental factors still affected clustering, next to reproductive mode (Fig. S6B; Table S4). Visualizing the bacterial community composition according to the genetic background in a PCoA-plot revealed no additional separation of clusters apart from the reproductive mode (Fig. S7B), which was also not displayed when the reproductive modes were plotted separately (Fig. S8B). However, as mentioned above, the effects of environmental conditions and population structure are difficult to disentangle because the reproductive mode and environment covary. After removing *Wolbachia* reads, all factors showed reduced differences when comparing the average distance to the centroid. Differences between the climatic zones remained aligned with reproductive mode but not population genetic clusters (Fig. S10).

Interaction between bacterial community composition and host genomic background

The bacterial species composition dissimilarity (Bray-Curtis distance) was positively correlated with pairwise genetic distance between individual wasps, with a high dissimilarity being associated with a high genetic distance (Fig. S11A, Spearman's Rank Correlation: $\rho=0.255$, $p<0.001$). However, this pattern was likely driven by the strong effect of *Wolbachia* on bacterial diversity in asexual wasps, as the positive correlation was lost upon removal of reads belonging to *Wolbachia* (Fig. S11B, Spearman's Rank Correlation: $\rho=0.015$, $p=0.66$).

Bacterial Shannon diversity and host genetic diversity, estimated as the observed heterozygosity, were strongly correlated, with a high host genetic diversity being associated with high bacterial diversity (Fig. S12A, Spearman's Rank Correlation: $\rho=0.62$, $p=0.001$). However, this pattern is likely driven by the strong effect of *Wolbachia* on bacterial diversity in asexual wasps, as the positive correlation was lost upon removal of reads belonging to *Wolbachia* (Fig. S12B, Spearman's Rank Correlation: $\rho=0.01$, $p\text{-value}=0.95$).

Discussion

Host-associated microbes are important modulators of host phenotypes, but the factors shaping host microbial communities often remain unknown despite their relevance. Aspects such as geography, biotic and abiotic environment, host genome, and biotic interactions are expected to play a role. However, the relative importance of one factor over the other remains unclear. Using the parasitoid wasp *Asobara japonica* as model species, we investigated whether and to what extent its bacterial community composition is influenced by the wasp's genetic background, environmental conditions, and the presence of the endosymbiont *Wolbachia* causing asexual reproduction (Kremer et al., 2009). We found that the presence of *Wolbachia* strongly influences the bacterial community composition in asexual wasps, while population structure and environmental conditions appear to play, if at all, only a minor additional role. In contrast, in sexual wasps, bacterial community composition is shaped by population structure and environmental conditions.

Overall, sexual and asexual wasps shared 56% of the detected bacterial taxa indicating a species-specific core bacterial community in *A. japonica*. Although the wasp culturing methods may impact the core microbiome, this method allowed for consistent measurement and retrieval of wasps across seven climate zones and five distinct genetic populations. Species-specific microbiomes have been reported for many animals, including bees and ants (Ramalho et al., 2017) and higher vertebrates such as birds (Bodawatta et al., 2021). Despite the apparent existence of a species-specific bacterial community with a high number of shared taxa, we found that the bacterial community of *A. japonica* markedly differed between populations according to the presence of *Wolbachia*. Compared to sexual wasps, which lack *Wolbachia*, asexual wasps exhibited a lower bacterial diversity (observed taxa and alpha

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diversity) and a clear separation from sexual wasps in bacterial community composition and abundance. This is unlikely due to a technical artefact resulting from the overpowering detection of *Wolbachia* in 16S rRNA metabarcoding. Removal of *Wolbachia* reads from the data, albeit leading to an alignment of alpha diversity between asexual and sexual wasps, still allowed clear separation of sexual and asexual wasps in bacterial community composition and abundance. Future work will need to elucidate the effect of culturing methods on environmental and laboratory acclimated organisms.

The population genomic analyses also clearly separated the reproductive modes, with sexual wasps having a higher genetic diversity than asexual wasps. In addition, sexual wasps could be further divided into three genetically distinct clusters, whereas the asexual wasps could be divided into only two. This population subdivision is congruent with geography and climatic zones along a north-south gradient, potentially fostered by geographic distances and population isolation. Indeed, each sexual population is located on a separate island. The two asexual population clusters are located north and south of the main islands, showing admixture at the contact zone between the two. Four of the eight wasps sampled from the Kobe location displayed admixed genetic ancestry, potentially indicating sexual reproduction in a previous generation. Indeed, the production of males and successful fertilisation can occasionally occur in asexually reproducing *A. japonica* and other thelytokous wasps (Kraaijeveld, Franco, Reumer, & van Alphen, 2009; Ma & Schwander, 2017; Schneider et al., 2003; Wachi, Gau, Fujie, Fukano, & Maeto, 2021).

Asexual wasps will produce males when the *Wolbachia* density falls below a specific threshold (Ma et al., 2015). Changes in *Wolbachia* density might stem from environmental fluctuations.

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Host microbial communities or individual members, for example, symbionts, can strongly react to environmental changes (Cohen et al., 2015; Novakova et al., 2017; Sumi et al., 2017) and environmental stressors, with potentially far-reaching consequences for their host (Lapanje, Zrimec, Drobne, & Rupnik, 2010; Rudman et al., 2019; Wang, Dittmer, Douglas, Huang, & Brucker, 2021). Indeed, the bacterial community composition of samples displaying admixed ancestry differed from the other samples from Kobe, potentially indicating that the detected hybridisation was caused by an earlier disruption of the bacterial community. Such disruption could have been caused potentially by a strong fluctuation in temperature, as symbionts appear to be sensitive to increases in temperature (Corbin et al., 2017; Hurst et al., 2000; Sumi et al., 2017; van Opijnen & Breeuwer, 1999) and show the general geographical pattern of lower prevalence in warmer regions (Corbin et al., 2017). Moreover, experimental heat exposure was able to eliminate *Wolbachia* in two-spotted spider mites (van Opijnen & Breeuwer, 1999), and a green stinkbug (Kikuchi et al., 2016) and symbiont reduction in response to increasing temperatures was found in coral reefs (Goulet & Goulet, 2021). Although the underlying reasons for admixed ancestry in some Kobe samples remain speculative, our findings suggest that environmental factors and associated changes in the bacterial community composition may have played a role.

Although our population genomic analyses revealed a clear sub-structuring of sexual and asexual wasps, the bacterial community composition showed no apparent clustering according to population genetic structure apart from the separation of the reproductive modes. However, we found a correlation between host genetic diversity and bacterial community composition, mainly driven by the differences between the reproductive modes. Moreover, the bacterial community composition of sexual wasps revealed structuring

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according to geography and population identity, with wasps from the two most southern locations being very similar in their bacterial community composition but distinct from wasps of the more northern location. This indicates that for the sexual wasps, environmental and possibly host genetic background influence bacterial communities, a conclusion that was also drawn for Atlantic salmon (Uren Webster et al., 2020).

In contrast to the pattern observed for sexual wasps, the bacterial community composition of asexual wasps was highly similar over all locations with a large degree of shared taxa and without any apparent influence of host population structure nor geographic effect. In contrast to the sexual wasps, this indicates a strong influence of *Wolbachia* in shaping their host bacterial communities, a conclusion that is consistent with studies on fruit flies (Simhadri et al., 2017), the small brown planthopper (Duan et al., 2020), and artificially infected mosquito adults (Audsley et al., 2017). This indicates that *Wolbachia*, and potentially other endosymbionts, can strongly affect host bacterial community composition and potentially act as a community stabiliser (Herren & McMahon, 2018). However, this might depend on the particular endosymbiont in question. For example, a recent study on five bug species (Heteroptera) found that their microbiome diversity depended more on host phylogeny than on the abundance of endosymbionts (Kolasa et al., 2019).

In conclusion, multiple factors seem to shape the bacterial community of *A. japonica* populations, with the presence of the endosymbiont *Wolbachia* playing a prominent role. Disentangling the relative contribution of factors shaping the bacterial community composition in the presence of *Wolbachia* will require further investigations, especially experimental manipulations. Especially experiments transinjecting the endosymbiont to

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sexual reproducing wasp will help to disentangle the effects of host reproductive mode from the effects of the symbiont. Moreover, our findings suggest that organisms hosting a dominant endosymbiont may be more sensitive to environmental changes with far-reaching consequences for the host. Indeed recent research suggests that fast environmental changes due to new and/or less predictable abiotic and biotic stressors can affect symbiotic interactions with potential cascading effects on the population dynamics of host species and the communities in which they are embedded (Duplouy, Dotson, Nishiguchi, & Cárdenas, 2021; Greenspan et al., 2020; Kolodny & Schulenburg, 2020; Pita, Rix, Slaby, Franke, & Hentschel, 2018; Trevathan-Tackett et al., 2019). Our data add to these studies highlighting that future research needs to take a more holistic approach (Brinker et al. 2019), considering multiple factors such as bacterial community composition, symbiont presence, and population dynamics to predict the potential effects of global change on ecosystems.

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Data Accessibility:

Raw data, meta data and scripts are available on *DRYAD* (<https://doi.org/10.5061/dryad.h9w0vt4mp>). Additionally, short-read data with sample metadata have been deposited in NCBI's SRA archives under BioProject ID PRJNA876812 (accession numbers SAMN30673182– SAMN30673088).

Author Contributions

P.B, L.W.B., M.F.C. and J.F.S. designed this research. P.B. and F.C. performed the research.

P.B. and Y.B.C. analysed the data with the help of J.F.S. and M.F.C. P.B., F.C., L.W.B, M.C.F.

and J.F.S wrote the paper.

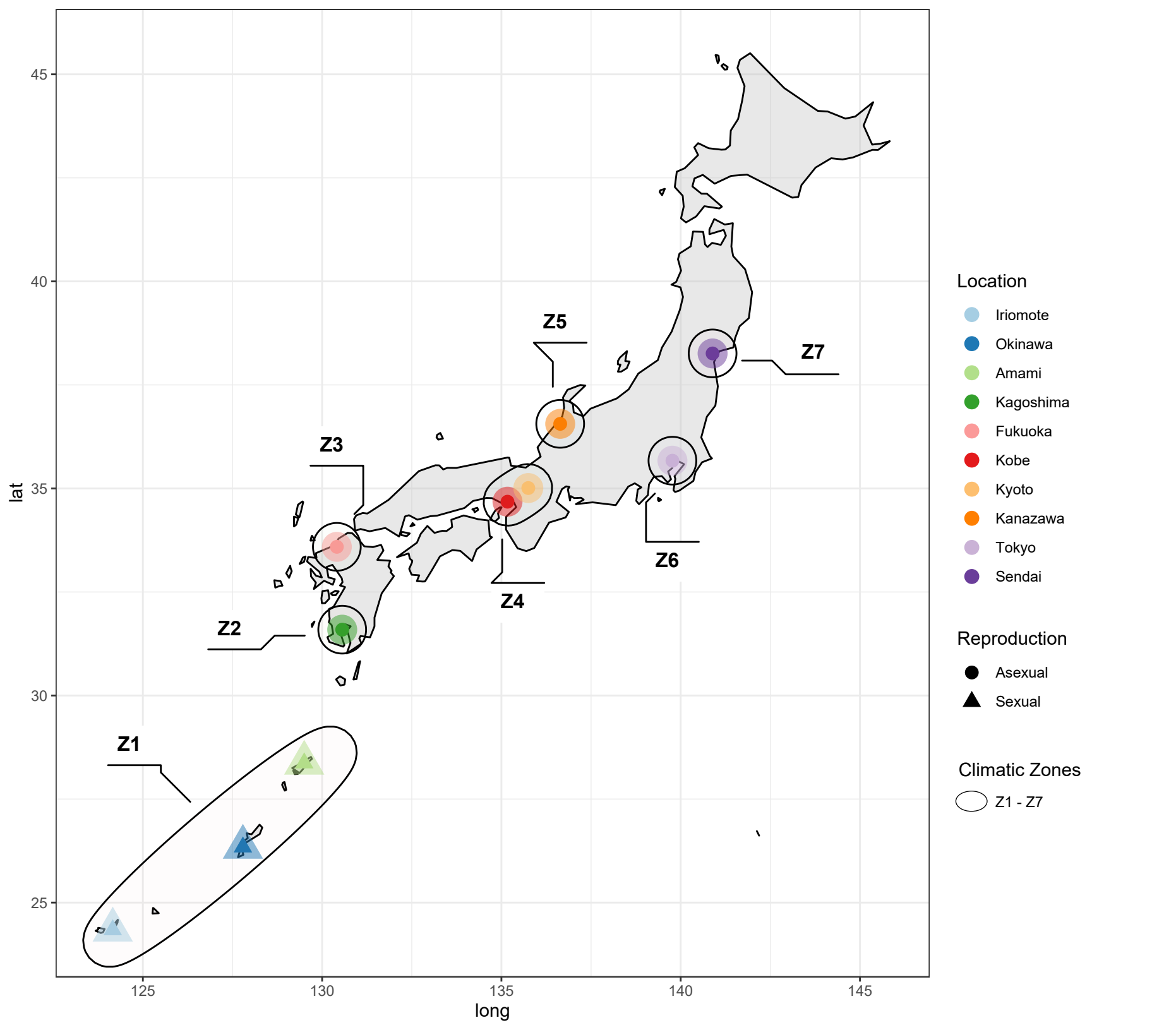
Figure legends:

Figure 1: Sample locations of the wasp, *Asobara japonica*, in Japan. Collection sites of *Asobara japonica* in Japan. Locations with sexual reproducing wasps are indicated with a triangle, locations with asexual reproducing wasps with a circle. Numbers indicate climatic zones. Six climatic zones were distinguished, ranging from subtropical (Z1, Z2) in the south to temperate conditions on the main island (Honshu) in the north (Z3, Z4, Z5, Z6, Z7). Ellipses depict locations with similar climatic conditions (see Table S1 for details).

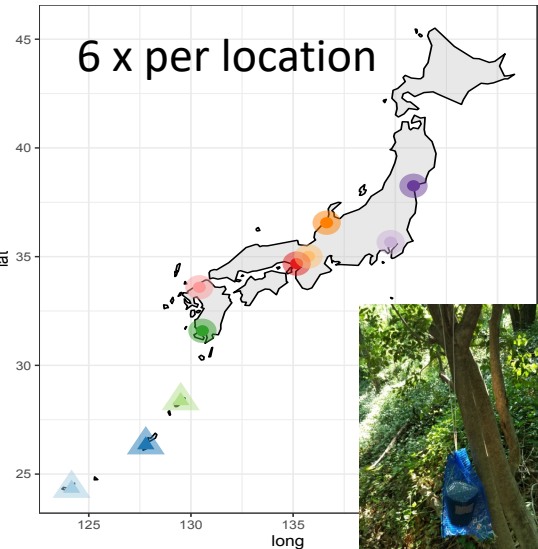
Figure 2: Schematic overview of sampling methodology. Panels visualize field work processes and sample collection. **1)** field bait trap distribution, **2)** fruitfly and parasitoid attraction, **3)** bait trap collection, and **4)** laboratory hosting of field collected *Asobara japonica*. Single female individuals for population genomic analysis were derived from females hatched from the field material (G0), whereas ten female wasps per sample were pooled derived from offspring (G1) of field collected wasps for bacterial community composition analyses.

Figure 3: Whole genome sequencing of *Asobara japonica* reveals five genetic clusters. Population genetic structure of *Asobara japonica* from ten locations, three uninfected sexual populations (Iriomote, Okinawa, Amami) from the three southern islands of Japan and seven with *Wolbachia*-infected, asexual populations (Kagoshima, Fukuoka, Kobe, Kyoto, Kanazawa, Tokyo, Sendai) on the main islands of Japan. Each sample represents one female wasp. Grouping of samples into genetic clusters via **A)** a principal component analysis of the first two principal components (PC1, PC2) and **B)** the third (PC1, PC3) showing locations in different colours with dots for asexuals and triangles for sexuals. **C)** Displays a mid-point rooted tree with bootstrap node support greater than 70% reported as grey circles. Samples clustered into five groups with a distinct separation between *Wolbachia*-infected asexual and uninfected sexual populations. **D)** Individual genetic ancestry to the five clusters estimated by *PCAngsd* analysis.

Figure 4: Bacterial diversity and community composition is driven by the endosymbiont *Wolbachia*. Bacterial diversity and community composition of the wasp *Asobara japonica*. Alpha diversity **A, B)** was measured as the observed number of species and expressed as Shannon diversity. Variance of bacterial community composition (beta diversity) **C, D)** was visualized via a Principal Coordinates Analysis (PCoA) based on Bray–Curtis distance dissimilarity matrices. Wasps from seven *Wolbachia*-infected, asexual populations (Kagoshima, Fukuoka, Kobe, Kyoto, Kanazawa, Tokyo, Sendai) and three uninfected, sexual populations (Iriomote, Okinawa, Amami) are included. Each sample contains a pool of ten female wasps. **A)** Alpha diversity of the rarefied dataset (*Wolbachia* +) and **B)** alpha diversity of the rarefied dataset excluding reads belonging to the genus *Wolbachia* (*Wolbachia* -). **C)** Beta diversity of the rarefied dataset (*Wolbachia* +) and **D)** beta diversity of the rarefied dataset excluding reads belonging to the genus *Wolbachia* (*Wolbachia* -). Circles represent asexual and triangles sexual populations. Alpha and beta diversity reveals a clear separation between infected and uninfected populations in both datasets.

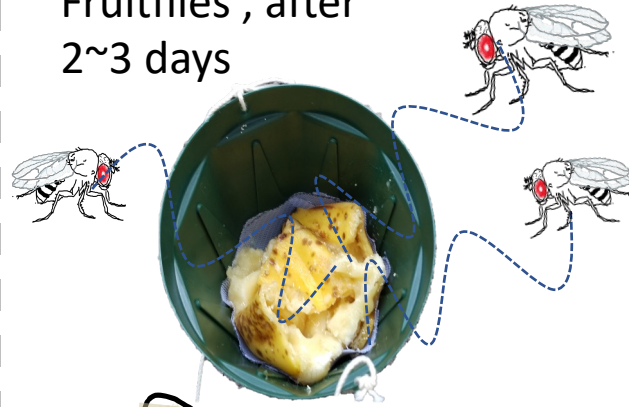


1. Sampling

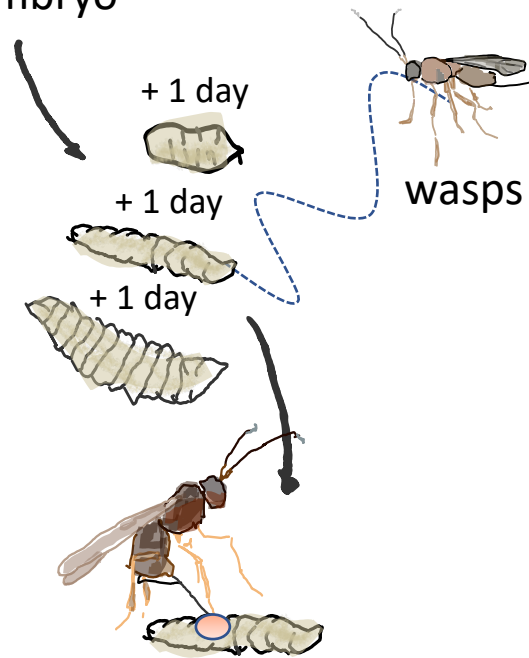


2. Attraction

Fruitflies, after 2~3 days



Embryo



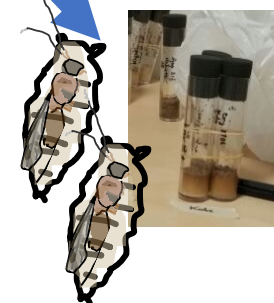
wasps

3. Collection

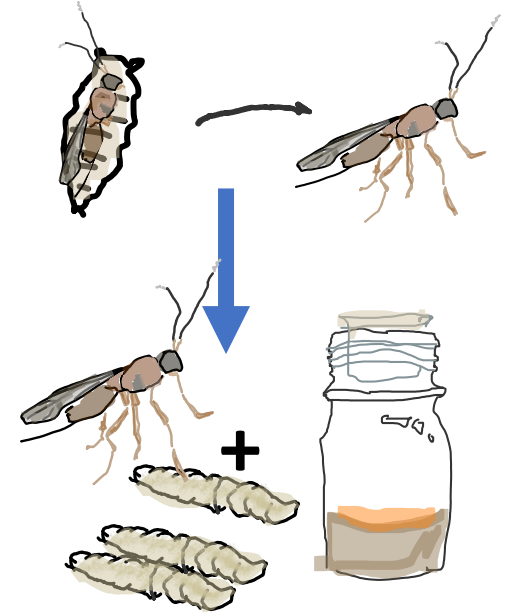
After 6~8 days
Trap material collected



+16 days



4. Hosting



+48 hours

G0

Pop. Gen

G1
+19 days



X10

G1

Bac. Com.



