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Can trees buffer the impact of climate change on pasture production and digestibility of Mediterranean dehesas?



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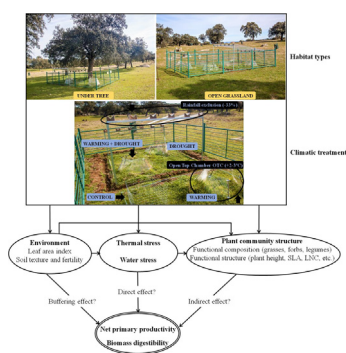
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HIGHLIGHTS

- Habitat type had a significant influence on biomass productivity and digestibility.
- Experimental warming increased by 33% biomass productivity, but reduced its quality.
- Reduced rainfall decreased by 13% biomass productivity.
- Changes in plant functional composition affected productivity and digestibility.
- Tree canopy played a buffering role in mitigating the impact of climate change.

GRAPHICAL ABSTRACT



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ABSTRACT

Sustainability and functioning of silvopastoral ecosystems are being threatened by the forecasted warmer and drier environments in the Mediterranean region. Scattered trees of these ecosystems could potentially mitigate the impact of climate change on herbaceous plant community but this issue has not yet tested experimentally. We carried out a field manipulative experiment of increased temperature (+2–3 °C) using Open Top Chambers and rainfall reduction (30%) through rain-exclusion shelters to evaluate how net primary productivity and digestibility respond to climate change over three consecutive years, and to test whether scattered trees could buffer the effects of higher aridity in Mediterranean dehesas. First, we observed that herbaceous communities located beneath tree canopy were less productive (351 g/m²) than in open grassland (493 g/m²) but had a higher digestibility (44% and 41%, respectively), likely promoted by tree shade and the higher soil fertility of this habitat. Second, both habitats responded similarly to climate change in terms of net primary productivity, with a 33% increase under warming and a 13% decrease under reduced rainfall. In contrast, biomass digestibility decreased under increased temperatures (–7.5%), since warming enhanced the fiber and lignin content and decreased the crude protein content of aerial biomass. This warming-induced effect on biomass digestibility only occurred in open grasslands, suggesting a buffering role of trees in mitigating the impact of climate change. Third, warming did not only affect these ecosystem processes in a direct way but also indirectly via

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changes in plant functional composition. Our findings suggest that climate change will alter both the quantity and quality of pasture production, with expected warmer conditions increasing net primary productivity but at the expense of reducing digestibility. This negative effect of warming on digestibility might be mitigated by scattered trees, highlighting the importance of implementing strategies and suitable management to control tree density in these ecosystems.

1. Introduction

Sustainability and functioning of silvopastoral ecosystems are being threatened by climate change (Campos et al., 2013). Thus, the warmer and drier conditions forecasted for the Mediterranean region might impact net primary productivity and biomass quality, which represent major services in this type of ecosystems (Fay, 2009; Martin et al., 2014), potentially influencing other ecosystem services such as cattle production (Dumont et al., 2015; Lee et al., 2017; Rojas-Downing et al., 2017). Previous studies have detected a differential impact of climate change depending on the identity and intensity of the climatic stressor (Ma et al., 2017; Wilcox et al., 2017). On the one hand, moderate increases in temperature tend to advance plant maturity and increase photosynthetic performance, resulting in improved nutrient use efficiency and greater plant biomass (Martinez et al., 2014; Song et al., 2019; Vicedo et al., 2019). Nevertheless, warming reduces the period when plant biomass shows optimal nutritional quality, increasing the fiber and lignin contents and reducing its digestibility (Dumont et al., 2015; Lee et al., 2017; Habermann et al., 2019b). In contrast, when temperature overpass an optimal threshold, excessive warming may negatively affect seed germination, photosynthesis and respiration, thus decreasing net primary productivity and quality (Fahad et al., 2017; D'Orangeville et al., 2018). On the other hand, a moderate water deficit slows plant maturation, although if it does not cause severe leaf loss, biomass quality can be maintained or even improved because it delays stem development, reduces the lignification and increases the digestibility (Buxton, 1996; Reddy et al., 2003). Nevertheless, severe droughts usually reduce net primary productivity and quality because water deficit constrains the growth and development of new organs, hastens leaf senescence, alters fiber and sugar content, and relocates different nutrients and carbohydrates from leaves to roots, decreasing thus biomass digestibility (Nandintsetseg and Shinoda, 2013; Dumont et al., 2015; Li et al., 2018). Despite the unquestionable value of these studies, understanding the isolated role of these two abiotic stressors only represents a part of the overall picture since the effect of one of them may be exacerbated or mitigated by the other. For instance, warming can trigger higher soil mineralization and higher nitrogen absorption by plants, while drought can reduce plant nutrient uptake (Matías et al., 2011; Song et al., 2019). Although the two abiotic stressors associated with climate change may impact net primary productivity and biomass quality differently, the combined effect of both factors remains poorly known, particularly in Mediterranean areas (Volaire et al., 2014). Therefore, an accurate prediction of climate change projections on ecosystem functioning requires the evaluation of the joint and interactive effects of temperature and precipitation.

The impact of climate change on these ecosystem properties (pasture productivity and quality) is expected to be spatially heterogeneous due to the mosaic of abiotic and biotic conditions caused by the typical scattered trees that characterize many silvopastoral ecosystems. Thus, trees reduce abiotic stress by attenuating air and ground temperature, decreasing plant evapotranspiration and reducing water stress for herbaceous plants growing under their canopies (Moreno, 2008; Abraham et al., 2014; Gargaglione et al., 2014; Gomes da Silva et al., 2021). In addition, trees favor hydraulic lift from the wettest soil layers (Ludwig et al., 2003, 2004a, 2004b; Yu and D'Odorico, 2015) and improve water infiltration (Joffre and Rambal, 1988; Ellison et al., 2017). Besides, tree litter decomposition provides nutrients and organic matter, which improves biomass quality under tree canopy (Ludwig et al., 2008; Cubera et al., 2009; Barnes et al., 2011). All these benefits of scattered trees on net primary productivity and biomass quality indicate that they could mitigate the impact of climate change on ecosystem functioning (Fay et al., 2011; Trambly et al., 2020;

IPCC, 2021). However, this issue has barely been experimentally tested, particularly in the Mediterranean region.

Trees might also attenuate the impact of climate change on ecosystem functioning indirectly via changes in the plant functional composition and diversity of pastures. The herbaceous species established under tree canopies usually develop morphological and physiological traits that allow them to tolerate the excess shade and the strong competition caused by trees, for example by having high values of height, leaf area and SLA that allow them to increase light uptake in this habitat where light is more limited (Scholes and Archer, 1997; Ludwig et al., 2004b; Dalke et al., 2018). This increases the heterogeneity of plant communities at the landscape level and contributes to spatial changes in community functional composition, diversity of pastures and, ultimately, in ecosystem functioning (Bruno et al., 2003; Hisano et al., 2018). However, the influence of tree canopy on different ecosystem properties seems to be highly context-dependent, with variable responses as a function of the characteristics of the dominant herbaceous species and the local climate (Maestre et al., 2009; Rivest et al., 2013). Therefore, it is essential to analyze the indirect effects of the forecasted rise in temperature and water stress on ecosystem functioning based on the interactions established between trees and the co-existing herbaceous species.

In Southwestern Spain, silvopastoral ecosystems termed 'dehesas' cover around 3.1 million ha (Moreno and Pulido, 2009). This ecosystem with scattered trees – mainly *Quercus* sp. – provide pasture for extensive cattle breeding (Moreno, 2008; Santamaría et al., 2014; López-Díaz et al., 2015; Iglesias et al., 2016). Due to their essential economic relevance and their high susceptibility to changes in climate, management and habitat structure (Moreno and Pulido, 2009), it is timely to investigate the potential role of trees to mitigate the impact of climate change on ecosystem functioning. To address this issue, we carried out a field experiment simulating rainfall reduction and increased temperature as predicted by climate change models during three consecutive years. The main objective was to evaluate how these silvopastoral ecosystems will respond to future climatic scenarios, and how scattered trees could modify net primary productivity and biomass digestibility and modulate these potential climate-induced changes. Specifically, we addressed the following questions: (i) Does climate change have an impact on net primary productivity and biomass digestibility? (ii) Could we differentiate the direct from the indirect effects of climate change on these two ecosystem properties?; and (iii) Would trees be able to buffer the impact that climate change has on net primary productivity and biomass digestibility? We hypothesize that net primary productivity and biomass digestibility will be altered under future scenarios of climate change. We expect to find stronger effects of warming than rainfall reduction on the variables analyzed due to the recognized tolerance of Mediterranean plant species to survive long periods of drought. We hypothesize that warmer conditions will cause an increase in net primary productivity at the expense of its digestibility. Moreover, we hypothesize that scattered trees will buffer the impact of abiotic stress (especially thermal stress) on net primary productivity and biomass digestibility. To our knowledge, this is the first study that provides an overall picture about the impact of climate change (including warming and drought) on pasture functioning in silvopastoral ecosystems, and evaluates the potential buffering role of scattered trees under experimental field conditions (but see Moreno, 2008, focused on watering effects).

2. Material and methods

2.1. Experimental design

The study was carried out at the Pedroches Valley (Córdoba, Southwestern Spain; 38° 22' 50.64"N, 4° 45' 27.69"W). The Mediterranean

climate of this area is characterized by cold, wet winters and hot, dry summers, showing a large intra- and interannual variability. The annual precipitation is 419 mm/yr and the mean annual temperature is 15.4 °C, being January the coldest month with an average temperature of 5.9 °C, and July the warmest month with an average temperature of 26.5 °C (IFAPA meteorological station, Hinojosa del Duque; records over the period 2009–2019; web reference at References section). The flora is characterized by a dense herbaceous layer ($\geq 80\%$ of cover) dominated by a diverse community of annual native species such as *Geranium dissectum* L., *Hordeum murinum* L. subsp. *leporinum* (Link) Arcangeli, *Calendula arvensis* L. subsp. *arvensis* and *Echium plantagineum* L. The mean species richness is 9.6 ± 0.3 species/m², and they coexist with scattered oak trees [*Quercus ilex* L. subsp. *ballota* (Desf.) Samp., with around 20% of canopy cover]. Tree density is 14.5 ± 1.3 trees/ha.

At this location, 36 experimental plots of 4 × 6 m were randomly established, placing half of them under the tree canopy and the other half in open grasslands (more than 10 m apart from the edge of tree canopies). The minimum distance between trees was 20 m. In each plot, four climatic treatments were set up to simulate future climate forecasts for the XXI century (2040–2070) in the Mediterranean area (SRES A-2 model by the IPCC, 2021). We delimited an area of 2.5 × 2.5 m to evaluate the impact of current climate conditions ('control', C, hereafter) on ecosystem functioning. We installed Open Top Chambers (OTC, Marion et al., 1997) to simulate the predicted annual temperature increase of 2–3 °C ('warming' treatment, W, hereafter). They were built of methacrylate sheets with a hexagonal

design and sloping sides of 40 × 50 × 32 cm (projected area in the soil of 0.65 m²) without UV-Filter to avoid modifying the light spectrum and to allow wavelength transmission between 280 and 750 nm (Faberplast, Madrid). A 'drought' treatment (D, hereafter) was implemented through the use of rain-exclusion shelters (2.5 × 2.5 × 1.5 m height) following the design of Matías et al. (2012). These shelters intercept 30% of the precipitation thanks to six methacrylate gutters of 0.14 m wide, and inclined at 20° (Yahdjian and Sala, 2002). Finally, to assess the impact of the combination of both sources of abiotic stress ('warming + drought' treatment, W + D, hereafter), OTCs were installed under the rain-exclusion shelters. This experimental design resulted in 144 experimental units (18 plots × 2 habitat types × 4 climatic treatments). To avoid interference with large herbivores, experimental plots were enclosed with cattle fences. The experimental design outline and main objectives of this paper are detailed in Fig. 1.

2.2. Environmental characterization of the experimental plots

2.2.1. Soil conditions

Soil texture was analyzed once in all 36 plots at the beginning of the experiment (2017) using samples from the first 10 cm soil layer (topsoil). These samples were collected by using a 3 cm diameter auger. Holes made with the auger were not filled afterwards to avoid using soil from an area situated outside our experimental units, which could alter the microbial community of the sampling points. Soil fertility was assessed each

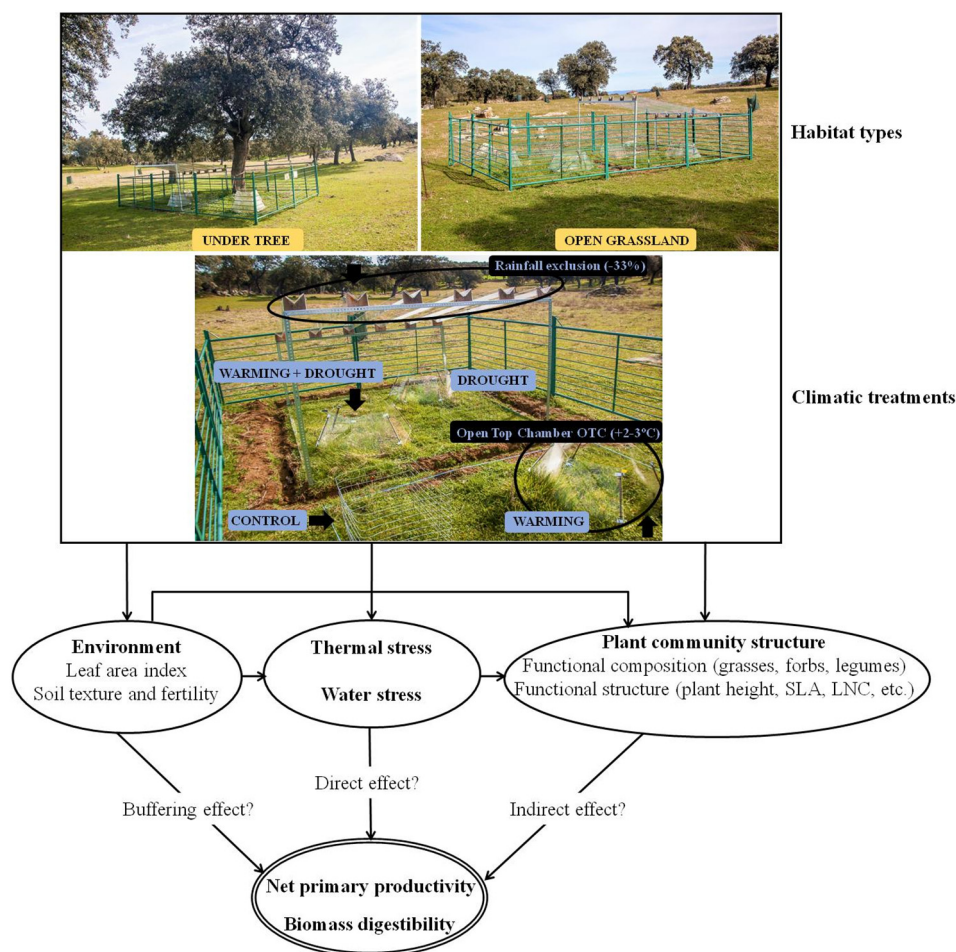


Fig. 1. Experimental design and the main variables analyzed in the study. We monitored 36 plots equally divided into two habitat types ('under tree' and 'open grassland') and four climatic treatments (control, warming, drought and warming + drought) installed in each one. We analyzed the direct, indirect and buffering effects of (i) abiotic factors such as tree shade (estimated by tree leaf area index), soil texture and fertility (ii) thermal and water stress indices and (iii) community functional composition (e.g. proportion in grasses, forbs and legumes) and structure (e.g. functional traits such as plant height, SLA or LNC) on net primary productivity and biomass digestibility.

year of study (2017–2019) using a composite top-soil sample collected at four different locations in each of the 144 experimental units. In the laboratory, samples were air-dried, crushed, and sieved, and the <2 mm fraction was analyzed for standard chemical properties (Sparks, 1996). Eight chemical properties were determined: acidity (measured in a 1:2.5 soil:water suspension), total organic carbon (through the oxidation of organic matter with H₂SO₄ and K₂Cr₂O₇), available P (following Watanabe and Olsen, 1965), NH₄⁺ and NO₃⁻ content (extracted with KCl 2 N and determined by distillation-titration), and available Ca⁺, Mg⁺ and K⁺ (extracted with NH₄CH₃CO₂ 1 N and determined by atomic absorption spectroscopy).

2.2.2. Microclimate

2.2.2.1. Tree cover. Tree cover was estimated in each of the 144 experimental units by measuring leaf area index [(LAI = leaf area (m²) / ground area (m²)] with hemispherical canopy photographs. They were taken in the middle of each of the 144 experimental units in spring 2018, before sunrise or after sunset. We used a horizontally-leveled digital camera (Coolpix 4500; Nikon, Tokyo, Japan) with a fish-eye lens of 180° field of view (FCE8; Nikon) situated at 0.5 m above ground level. Images were analyzed using Hemiview Canopy Analysis software version 2.1 (1999; Delta-T Devices Ltd., Cambridge, UK).

2.2.2.2. Thermal stress index. Air temperature was registered at ground level hourly with a resolution of 0.1 °C using HOBO MX2201 data loggers (Onset Computer, Bourne, MA, USA) in one sixth of the total experimental units (n = 24). A sensor was installed in each of the 24 randomly-chosen experimental units. According to some authors, plants generally suffer stress above 30 °C. From this temperature, photosynthesis experiences a rapid decline (Saini and Aspinall, 1982), which causes a reduction in biomass production (Benavides et al., 2009; Fahad et al., 2017). In addition, seed sterility appears (Saini and Aspinall, 1982) and root development is prevented (Ferris et al., 1998; Huang et al., 2012) when temperatures overpass 30 °C. Taking this information into account, cumulated thermal stress was calculated for mean temperature during January–July (covering the growth period of herbaceous plants in the study area) as in Eq. (1):

$$\text{Thermal stress} = \frac{\text{Accumulated temperature } (T_{\text{mean}} - 30)}{\text{Monitoring days}_{(\text{January} - \text{July})}} \quad (1)$$

where T_{mean} is the mean temperature, and monitoring days represent the number of days of the period sampled.

2.2.2.3. Water stress index. In half of the experimental units (n = 72), volumetric water content (% VWC) in the soil was periodically (i.e. weekly in spring and monthly in the rest of the year) assessed using a PR2 humidity probe (Delta-T Devices Ltd., Cambridge, UK) using permanent access tubes of 40 cm depth. It records moisture values at intervals of 10 cm. A permanent tube was installed in each of the 72 randomly-chosen experimental units. In order to determine the amount of water available for plants, the total transpirable soil water (% TTSW) was calculated for the average of the four depths as in Eq. (2):

$$\text{TTSW} = \text{VWC}_{\text{max}} - \text{VWC}_{\text{min}} \quad (2)$$

where VWC_{max} and VWC_{min} represent the maximum and minimum volumetric water content in the soil (%) for each experimental unit during January–July period, respectively.

TTSW shows the ability of plant communities to uptake water under particular soil conditions because it is based on the observed SWC dynamics, assuming that the minimum SWC in summer reflects the limit of plant water uptake. Assuming that 30% of the TTSW is the threshold under which plants suffer from water stress, i.e. when their accessible soil water

reserve is almost empty (e.g. Barkaoui et al., 2017), the water stress index was calculated during January–July using the following formula (Eq. (3)):

$$\text{Water stress index} = \frac{\text{VWC}_{\text{stress threshold}} - \text{VWC}_i}{\text{TTSW} \times 0,3} \quad (3)$$

where VWC_i is the volumetric water content in each monitoring data, and $\text{VWC}_{\text{stress threshold}}$ shows the result of the sum of 30% TTSW and VWC_{min} .

2.3. Plant measurements

2.3.1. Species composition

Plant species abundance and composition of herbaceous species were determined yearly from 2017 to 2019 during the peak of vegetative growth. Two annual censuses were carried out between March and May to record the maximum number of species, including the entire phenological spectrum. Four 21 × 21 cm quadrats (divided into 9 squares of 7 × 7 cm) were randomly placed in each experimental unit. Species frequencies were calculated in each of the 144 experimental units from the number of squares where each plant species was present. Species of each experimental unit were classified into three groups: ‘grasses’, ‘leguminous’ and ‘forbs’.

2.3.2. Trait measurements

The most dominant species (i.e. those making up at least the 70% of the total frequency of the community) were selected in each of the 144 experimental units. Five key functional traits were measured during the peak vegetation growth in spring of 2017 and 2018: plant height (cm), leaf area [LA (cm²)], leaf dry-matter content [LDMC = leaf dry weight (mg)/hydrated leaf weight (g)], specific leaf area [SLA = leaf surface (cm²)/leaf dry weight (g)], leaf nitrogen content [LNC (mg/g)] and leaf carbon content [LCC (mg/g)]. LNC and LCC were analyzed by the Analysis Service of the Institute of Natural Resources and Agrobiological of Seville. Plant height was measured in five individuals per species and experimental unit, whereas the five leaf traits were quantified in 10 individuals per species and experimental treatment combination (habitat type and climatic treatment) following standardized protocols (Cornelissen et al., 2003). Leaf size was quantified using an image analysis software (Image Pro-plus 4.5; Media Cybernetic Inc., Rockville, MD, USA).

Traits were measured at the species level during two successive years, with contrasting climate conditions. Thus, 2017 was a typically dry year (275.4 mm), while 2018 was particularly wet (533.6 mm). During 2019, which was also dry (228.5 mm), no functional trait measurements were taken, but they were estimated from the average traits quantified in the two previous years. Previously, we evaluated the interannual variability of species trait values between 2017 and 2018, with the aim of identifying which traits were the most variable and possibly less accurate to estimate for 2019 (Fig. A.1). Although the analyzed traits varied from one year to another (especially LA, LNC and LCC), no clear bias was identified since points were distributed around the 1:1 line (Fig. A.1).

2.3.3. Functional composition of plant communities

We employed the formula provided by Garnier et al. (2004) to calculate community weighted means (CWMs, hereafter) for each of the traits mentioned above (Eq. 4):

$$\text{CWM} = \sum_{i=1}^n p_i \times \text{trait}_i \quad (4)$$

where p_i is the relative contribution of the species i to the community, n is the number of most abundant species, and trait_i is the trait value of the species i .

2.4. Net primary productivity and biomass quality

2.4.1. Net primary productivity

At the end of the vegetative cycle (late June) of 2017, 2018 and 2019, aerial biomass produced in each of the 144 experimental units was

collected using 50 × 50 cm adjacent quadrats. Vegetation was cut at ground level, trying to collect the maximum possible amount of aerial biomass. Collected samples were cleaned of non-herbaceous material, and dry biomass was determined using a precision balance after drying it in a forced-air oven at 60 °C for 48 h.

2.4.2. Biomass digestibility

Samples of dry biomass were finely ground (<2 mm sieve) in an IKA mill. Enzymatic digestibility of organic matter (hereafter, digestibility) was estimated using the near-infrared reflectance spectroscopy technique (Vis-NIRS). Samples were scanned with the LabSpec 5000 spectrophotometer (350–2500 nm; ASD Inc., Boulder, Colorado, USA) using IndicoPro6.0 software (ASD Inc., Boulder, CO, USA). Four replicates of each biomass sample were scanned (each being an average of 50 internal scans). The final sample spectrum was obtained by averaging the four replicates. Statistic of Vis-NIRS equation used to predict digestibility is presented as an appendix (Table A.1).

2.5. Statistical analyses

We conducted a multiple factor analysis (MFA) to analyze the multivariate correlations among the abiotic and biotic factors quantified in this study and to identify the major axes of variation among the 144 experimental units. First, we analyzed differences between habitat types and their interaction with climatic treatments on thermal and water stress indices and CWMs using ANOVAs, followed by Tukey post-hoc tests. Second, we tested the individual and combined effects on net primary productivity and biomass digestibility of the following fixed factors: habitat type, climatic treatment, year, proportion of grasses, CWMs of plant height, LDMC, SLA, and LNC. For this purpose, we used Linear Mixed-effect Models (LMM), including the plot as a random factor, which was coded to affect only intercepts (not slopes). We compared a ‘full model’ based on the combination of all factors to a series of ‘reduced models’ in which factors were deleted one by one according to their significance (elimination of the least significant)

until reaching the ‘null model’. To accomplish this, we used the AICc (Akaike Information Criterion) with a greater penalty for an extra parameter, as recommended by Burnham and Anderson (2002). Models were considered different when $\Delta AICc < 2$. The best model with the lowest AICc was evaluated using maximum likelihood estimation. The proportion of variance explained by the fixed factors alone and by both the fixed and random factors was estimated by R^2_m (‘marginal R^2 ’) and by R^2_c (‘conditional R^2 ’), respectively (Nakagawa and Schielzeth, 2013). We derived the effect size of each factor using the estimates of the best model. Previously, normality of these variables was tested using Kolmogorov-Smirnov test, and variables were transformed in the cases in which it was necessary. Third, we then evaluated the relationships between stress indices and net primary productivity and biomass digestibility using bivariate linear regressions. Statistical analyses were performed with STATISTICA® Software and within the R environment v.4.0.0 using *tidyverse* (for data manipulation and visualization), *cowplot* (for multiple plots compilation), *Hmisc* (for graphical representation), *ggpubr* (for color and shapes), *FactoMineR* (for MFA computation), *lme4* and *lmerTest* (for running linear mixed models), *AICcmodavg* (for computing AICc and comparing models), *car* and *pbkrtest* (for testing the estimated parameters, *p-value*) and *MuMIn* (for determining R^2_m and R^2_c) packages (R Core Group, 2020).

3. Results

3.1. Descriptive analysis of the abiotic environment and plant community functional structure

3.1.1. Differences between habitat types

The 144 experimental units captured a large heterogeneity in both abiotic and biotic conditions, primarily promoted by differences associated to the two habitat types considered in our study, as summarized by the MFA (Fig. 2).

The MFA explained 54.80% of the total variability, with 33.40% accounted by dimension 1 (Dim-1, hereafter) and 21.40% by dimension 2

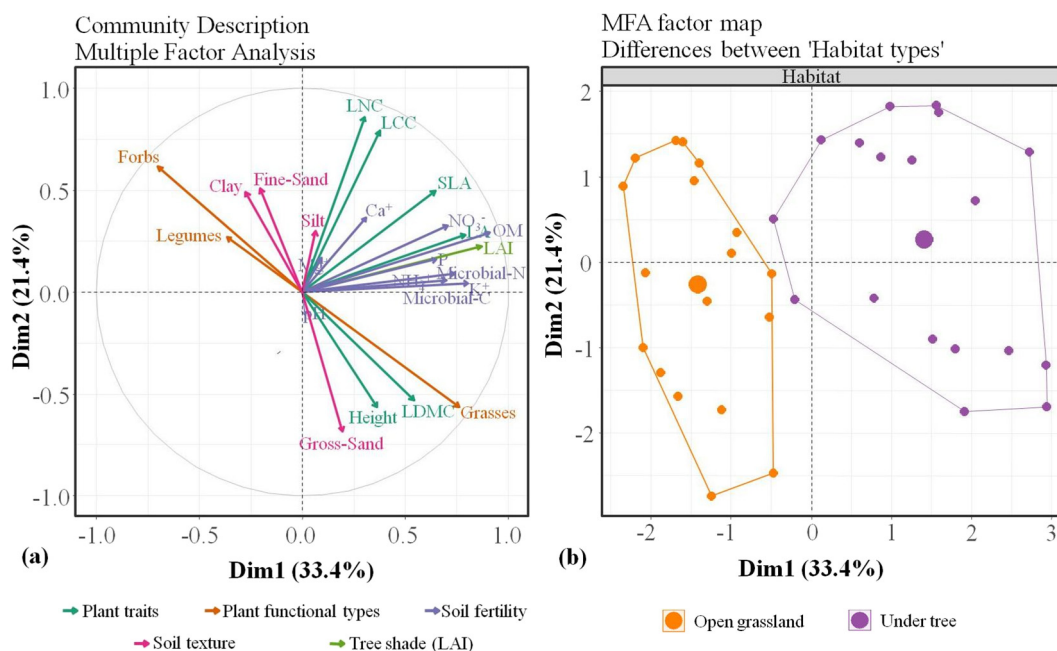


Fig. 2. Plot of the first and second axes of the multiple factor analysis (MFA) relating plant community functional structure, soil characteristics [soil fertility (organic matter, OM; C; N; P; NO₃⁻; NH₄⁺; Ca⁺; Mg⁺; K⁺), and soil texture (coarse-sand; fine-sand; clay; silt)] and tree shade (represented as leaf area index, LAI). Plant community functional structure was characterized in each of the 144 experimental units according to the following information: the proportion of different plant functional types (grasses, forbs and legumes), and the possession of some key functional traits [plant height (Height); leaf dry-matter content (LDMC); leaf area (LA); specific leaf area (SLA); leaf nitrogen content (LNC); leaf carbon content (LCC)] (panel a). Projection of the 36 plots on the plane defined by the two main MFA dimensions, with emphasis on differences due to habitat types (open grassland versus under tree). Three years of sampling (2017, 2018 and 2019) were represented (panel b).

(Dim-2, hereafter). Dim-1 was determined mainly by the level of tree shade (LAI) and, to a lesser extent by soil organic matter (OM), the proportion of grasses and some leaf morphological traits such as LA and SLA. Dim-2 was mainly determined by the two leaf chemical traits (LNC and LCC) and soil texture (Fig. 2a). Factor loadings are detailed in Fig. A.2. Habitat types differed strongly along Dim-1, with plots located under trees presenting higher OM and K^+ content in the soil, and their plant communities having a higher proportion of grasses, higher SLA and LNC than those located in open grasslands (Fig. 2b).

Microclimatic conditions also differed in the two habitat types. On the one hand, the annual mean air temperature was slightly lower under trees than in the open grasslands (15.5 ± 0.1 vs. 16.2 ± 0.1 °C). Thus, habitat located in open grasslands presented higher thermal stress than those situated under tree canopies, as evidenced by the negative relationship between LAI and thermal stress (Fig. A.3a; $p < 0.001$, $R^2 = 0.66$). No significant differences in water stress were detected when comparing both habitat types (Fig. A.3b; $p = 0.54$, $R^2 = -0.003$). On the other hand, soil moisture was slightly lower under trees ($p = 0.08$) when using an average of the 4 seasons of the year. These differences between habitat types in soil moisture were more pronounced in winter ($p < 0.001$) and spring ($p < 0.001$) than in the other two seasons ($p = 0.84$ in summer and $p = 0.78$ in autumn) (Table A.2).

3.1.2. Differences among climatic treatments

As expected, the climatic treatments exerted a significant influence on microclimatic conditions. First, the two treatments using Open Top Chambers (W and W + D) significantly increased air temperature for C or D treatments by about 2 °C (Fig. A.4). Consequently, these plots subjected to temperature increase exhibited the highest thermal stress values, but differences were only significant in open grasslands (Table A.2). Second, spring moisture in the top-soil was significantly higher in the C and W treatments than in the D and W + D treatments, but these differences were only significant in open grasslands (Table A.2 and Fig. A5). Water stress, however, did not differ significantly among climatic treatments (Table A.2).

Our experimental climatic treatments also induced significant changes in species composition and plant community functional structure. First, we detected a higher abundance of grasses in those units subjected to increased temperature. Second, plants subjected to warmer conditions were taller compared to those growing under C and D conditions, although these differences were only significant in open grasslands (Table A.3).

3.1.3. Differences among years

The three sampling years (2017, 2018 and 2019) did not differ significantly in thermal stress but did in water stress accumulated over the growing season (Table A.4). 2017 and 2019 were particularly dry (especially in spring), whereas 2018 was extremely rainy compared with the yearly mean precipitation registered for the last 20 years at the study area. These among-year differences in precipitation resulted in significant differences in SVWC and water stress. Specifically, 2019 presented higher values of accumulated water stress than 2018 and 2017 (Table A.4).

3.2. Factors affecting net primary productivity and biomass digestibility

Net primary productivity and biomass digestibility were affected by most of the fixed factors considered in this study, as evidenced by the inclusion of habitat type, climatic treatment, year and some variables related with plant community functional structure in the best linear mixed models (i.e. those with the lowest AICc values). The models presented a high goodness of fit in the two response-variables, with R^2 values ranging from 0.33 to 0.47. In general, the accuracy of the models only improved slightly when including the random effects of the experimental plots, as denoted by the small differences between R^2 c and R^2 m (Table 1).

3.2.1. Differences between habitat types

Significant differences in net primary productivity and biomass digestibility were found between habitat types (Tables 2 and 3), with lower

Table 1

Results from the linear mixed-effects models used to analyze the effects of climate change on the response variables considered in this study (net primary productivity and biomass digestibility). Habitat type (under tree and open grassland) and sampling year (2017, 2018 and 2019) were also included as fixed factors into the models to control the effects mediated by (habitat-associated) microclimatic and (inter-year) climatic differences, respectively. The best models were selected using the corrected Akaike Information Criterion (AICc). Differences between the best and the null model (Δ AICc) are indicated. Variability explained by fixed factors (R^2 marginal, R^2 m), by both fixed and random factors together (R^2 conditional, R^2 c).

Response variable	Best model	Δ AICc	R^2 m	R^2 c
Productivity	Habitat type \times Climatic treatment + Plant height + LNC + Grasses + Year + (1 Plot)	101.42	0.28	0.47
Digestibility	Habitat type \times Climatic treatment + Grasses + Year + (1 Plot)	86.43	0.27	0.33

values of net primary productivity and higher values of biomass digestibility under tree canopies than in open grasslands ($p < 0.001$; 351 g/m^2 vs. 493 g/m^2 for net primary productivity and 44% vs. 41% for biomass digestibility, respectively; Fig. 3).

3.2.2. Differences among climatic treatments

Climatic treatments influenced both net primary productivity and biomass digestibility, but with relevant differences depending on the response variable and the habitat type (Tables 2 and 3). In general, plant communities experimentally subjected to warmer conditions had higher net primary productivity (an increase of 33%; $p < 0.001$) (Fig. 3a) but lower biomass digestibility (a decrease of 7.5%; $p = 0.01$) (Fig. 3b). Both habitat types exhibited a similar trend, with enhanced net primary productivity in the W treatment (Fig. 3a). However, reduced biomass digestibility under warming conditions was significant in plots located in open grasslands but not in those under tree canopy (Fig. 3b). Conversely, reduced rainfall caused a significant decrease of net primary productivity by 13% ($p < 0.001$) with no significant impact on biomass digestibility ($p = 0.45$).

This above-reported variation in net primary productivity and biomass digestibility can be related to differences between habitat types and climatic treatments in abiotic stress. On the one hand, higher thermal stress

Table 2

Results from the analysis of variance evaluating the influence of fixed factors (habitat type, climatic treatment, plant height, LNC, grasses, year and the interaction habitat type \times climatic treatment) on net primary productivity.

Factor	SS	d.f.	F	p
Habitat type	615049	1	18.55	<0.001
Climatic treatment	2249541	3	22.62	<0.001
Plant height	157965	1	4.76	0.03
LNC	139633	1	4.21	0.04
Grasses	229839	1	6.93	0.01
Year	1315215	2	19.84	<0.001
Habitat type \times Climatic treatment	94572	3	0.95	0.42

Table 3

Results from the analysis of variance evaluating the influence of fixed factors (habitat type, climatic treatment, grasses, year and the interaction habitat type \times climatic treatment) on biomass digestibility.

Factor	SS	d.f.	F	p
Habitat type	1075.73	1	42.40	<0.001
Climatic treatment	294.49	3	3.87	0.01
Grasses	826.50	1	32.58	<0.001
Year	375.48	2	7.40	<0.001
Habitat type \times Climatic treatment	85.13	3	1.12	0.34

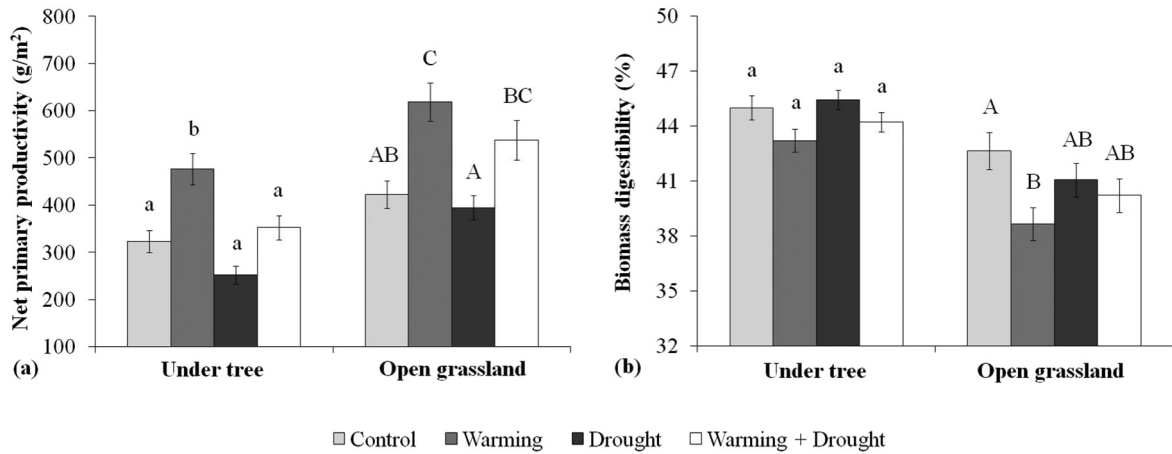


Fig. 3. Effects of climatic treatments (control, warming, drought and warming + drought) on net primary productivity (g/m²) (panel a) and biomass digestibility (%) (panel b) for the two habitat types considered in this study (under tree and open grassland). The mean of the three sampling years of (2017, 2018 and 2019) is represented. Letters (lowercase for habitat located under tree and uppercase for open grassland) are the result of factorial ANOVA, and subsequent post-hoc Tukey analyses were performed to compare the four climatic treatments within each habitat.

caused an increase in net primary productivity ($p < 0.001$, $R^2 = 0.17$) and a decrease in biomass digestibility ($p < 0.001$, $R^2 = 0.19$) (Fig. 4a and b, respectively). On the other hand, higher water stress caused a decrease in both net primary productivity ($p = 0.08$, $R^2 = 0.03$) and biomass digestibility ($p = 0.002$, $R^2 = 0.11$) (Fig. 4c and d, respectively).

3.2.3. Differences among years

Net primary productivity and biomass digestibility varied among the three sampling years, with 2019 presenting a significant reduction in these variables compared to 2017 and 2018 ($p < 0.001$). Specifically, net

primary productivity decreased from 459.5 g/m² in the most productive year (2017) to 356.5 g/m² in 2019 (22.4% reduction). Biomass digestibility decreased by 9% from 2018 (year with the highest values) to 2019.

3.2.4. Influence of plant community functional structure

Higher plant height and higher LNC were associated with greater net primary productivity (Fig. 5a; $p = 0.04$, $R^2 = 0.04$; and 5b; $p = 0.71$, $R^2 = -0.01$; respectively). A higher proportion of grasses was slightly and positively related with net primary productivity (Fig. 5c; $p = 0.29$, $R^2 = 0.001$). By contrast, biomass digestibility decreased strongly as the proportion

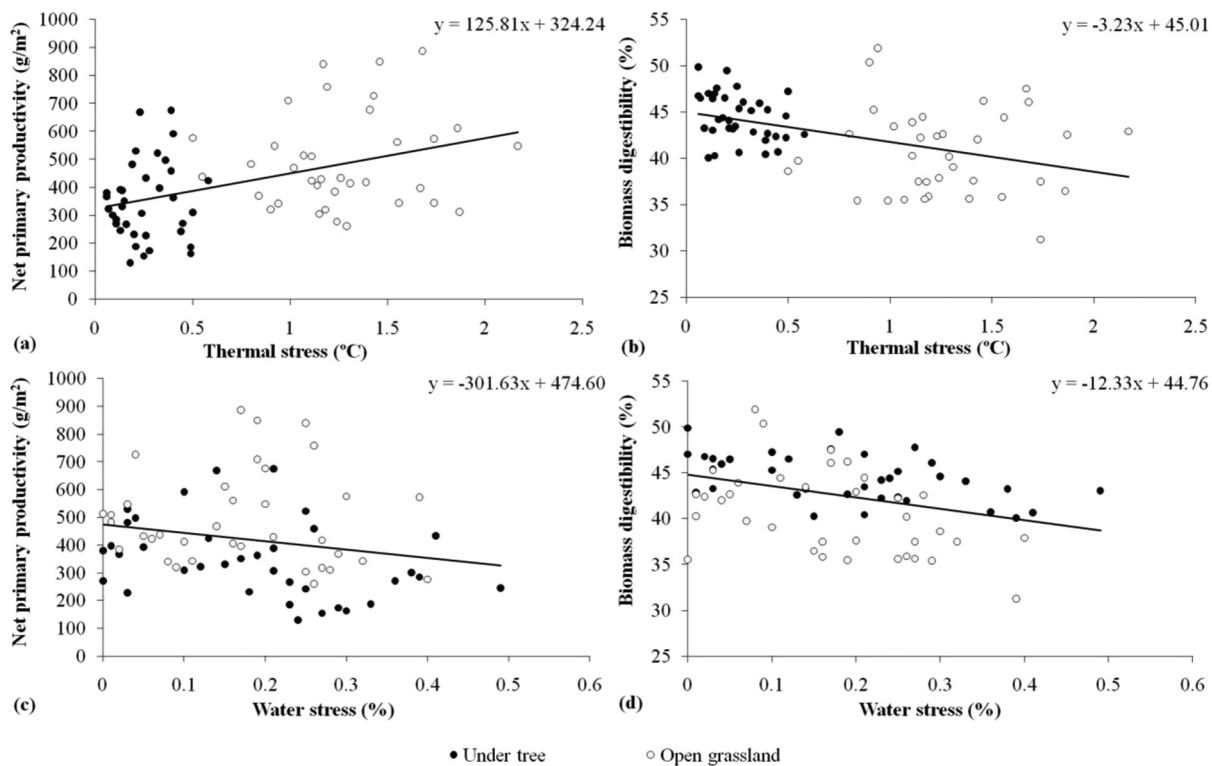


Fig. 4. Response of net primary productivity (g/m²) (panels a and c) and biomass digestibility (%) (panels b and d) to thermal stress (°C) and accumulated water stress (%), respectively. Open symbols represent the climatic treatments in open grassland habitat, whereas closed symbols represent the climatic treatments under tree habitat. The line is the result of the linear regression for the whole data set, whose equation is also indicated. Three years of sampling (2017, 2018 and 2019) are represented.

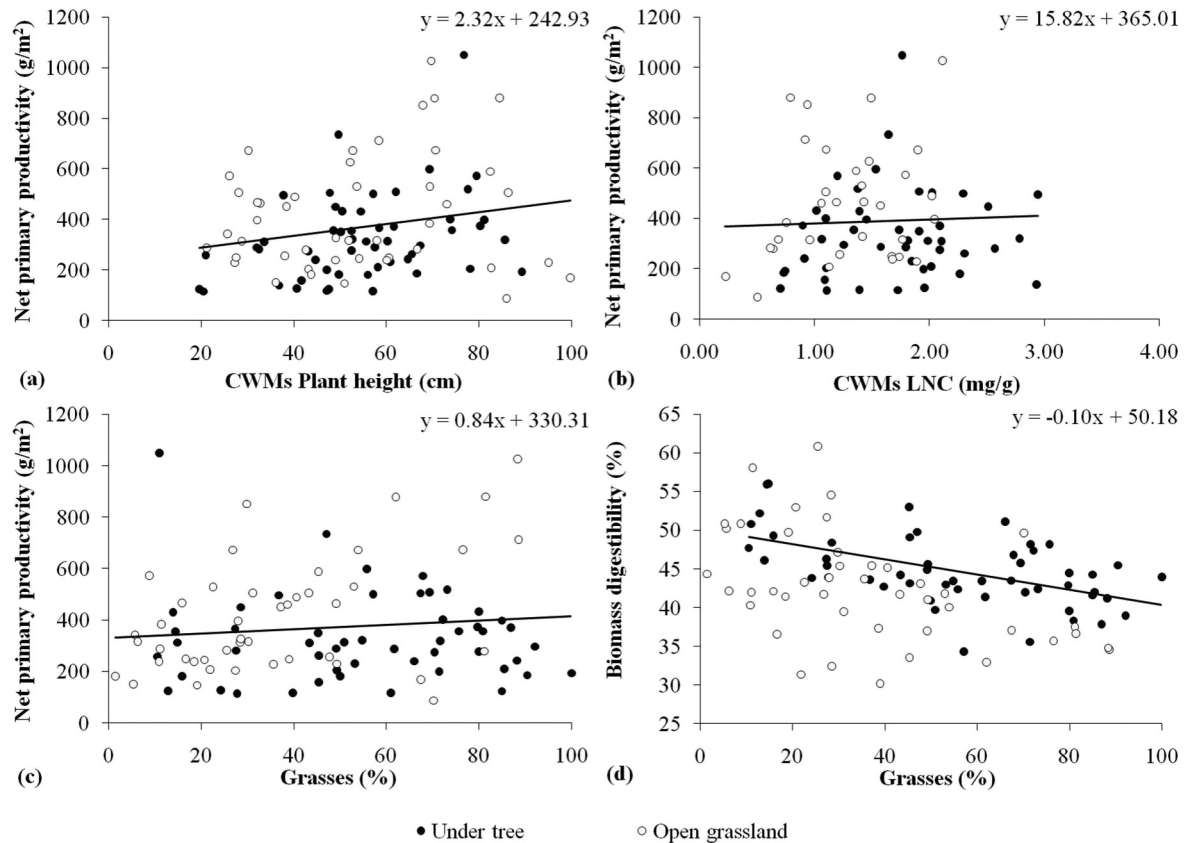


Fig. 5. Influence of plant functional traits [expressed in terms of CWMs of plant height (cm) and LNC (mg/g) (panels a and b, respectively)] and plant species composition [expressed in terms of percentage of grasses (panels c and d)] on net primary productivity (g/m^2) and biomass digestibility (%). Open symbols represent those located in open grassland, whereas closed symbols represent the experimental units located under tree habitat. The line is the result of the linear regression for the whole data set, whose equation is also indicated. Three years of sampling (2017, 2018 and 2019) are represented.

of grasses increased in the communities (Fig. 5d; $p < 0.001$, $R^2 = 0.14$). Detailed significance values are shown in Tables 2 and 3.

4. Discussion

We present a pioneering work on the buffering effect of trees on climate change and their impact on net primary productivity and biomass digestibility in semi-arid silvopastoral ecosystems of Southern Spain. We show how increased temperatures and reduced rainfall affect these ecosystem properties through direct (impact of abiotic stress) and indirect (changes in plant functional structure) effects. We discuss the role of trees in modifying aridity impacts on ecosystem properties of dehesa ecosystems.

4.1. Climate change had an impact on net primary productivity and biomass digestibility

Our experimental study of climate change simulation indicated that future climatic conditions could alter net primary productivity (both in quantity and quality), with contrasting effects of the two studied abiotic stressors.

4.1.1. Direct effects of climate change (abiotic factors)

On the one hand, increased temperature favored net primary productivity in the studied dehesas. These findings are in agreement with the results found by previous studies (Wu et al., 2011; de Assis et al., 2016; Yu et al., 2019), who showed that higher temperatures increased ecosystem respiration, causing an increase in plant productivity, as a consequence of an enhanced metabolism and growth (Lawlor, 1993). On the other hand, reduced rainfall caused a significant decrease in net primary productivity, probably because a lower soil water availability impeded seedling establishment and reduced plant growth (Espigares and Peco, 1993;

Dimitrakopoulos and Bemmerzouk, 2003). Moreover, a decrease in soil water content tends to reduce nutrient availability to plants because it limits soil microbial processes and ecosystem CO_2 fluxes, which translates into a decrease in plant productivity (De Dato et al., 2008; Sardans et al., 2008; Wu et al., 2011). This result is in agreement with previous works that reported a negative impact of water stress on vegetation diversity and net primary productivity (Fu et al., 2013, 2018; Fu and Shen, 2016; Wang et al., 2013; Sternberg et al., 2015). As a consequence of these contrasting effects of both climatic stressors, the amount of net primary productivity reached intermediate values in the combined W + D treatment, likely as a result of offsetting the positive effect of warming with the negative effect of drought. This lack of combined effects of the two climatic stressors contrasts with the results of other recent studies in regions with low rainfall (e.g. Suzuki et al., 2014; Zandalinas et al., 2018), where a moderate warming exacerbated the negative impacts of drought on plant growth.

Conversely to productivity, biomass digestibility was significantly lower under experimental warming, revealing that an increase in temperature negatively affects biomass quality (Čop et al., 2009; Bloor et al., 2010; Habermann et al., 2021). According to Wilson (1982), temperature is the most important environmental factor influencing biomass quality, reporting that for every 8°C that temperature increases, digestibility of grasses decreases $\sim 0.5\%$. Plant species, especially annuals, tend to advance their reproductive period in response to rising temperature, reaching earlier plant maturity and senescence (Whittington et al., 2015; Valencia et al., 2016; Moore and Lauenroth, 2017). In general, plant maturation increases fiber and lignin content and reduces crude protein content and digestibility, decreasing thus the nutritional value of plant biomass (Neel et al., 2008; Nordheim-Viken et al., 2009; Habermann et al., 2019b). In contrast, reduced rainfall did not significantly affect biomass digestibility. This lack of drought effects on biomass digestibility could be a

consequence of the lower productivity derived from reduced rainfall, which implies a reduction of the total transpiration surface and, therefore, water availability per plant biomass may be similar independently from water availability.

Our results suggest that the predicted temperature increase will decrease biomass digestibility in the dehesas of Southern Spain, with potential negative consequences for the cattle that feed on (lower nutritional intake) and for the farmers who obtain economic benefits from them (lower quality of milk, meat, etc.) (Pearce et al., 2010; Alothman et al., 2019). Additionally to the significant influence of climatic treatments on net primary productivity and biomass digestibility, we also detected a large variation between years, likely promoted by the inter-annual differences in water stress. Specifically, the lowest values of net primary productivity appeared in the driest year (2019), whereas the highest values of pasture digestibility were collected in the rainiest year (2018). This result is in agreement with those found in previous studies documenting greater inter-annual variation of plant production than that promoted by experimental climatic treatments (Yu et al., 2019).

4.1.2. Indirect effects of climate change

Climate change might also indirectly affect the ecosystem processes quantified in our study via significant changes in plant species composition and functional trait distribution within plant communities (Cowles et al., 2016; Napier et al., 2016; Yang et al., 2017; Li et al., 2018). On the one hand, our results revealed that grasses have a higher abundance under an experimental increased temperature (Shi et al., 2016). This enrichment in grasses under warmer conditions could partially explain the higher productivity and the lower biomass digestibility detected in the warming treatment. Similarly, Vázquez de Aldana et al. (2008) showed that an increase in precipitation over spring favored the presence of legumes, which provides a pasture with higher protein content and digestibility, while in drier springs the proportion of grasses increased and therefore the lignin content was higher.

On the other hand, plant communities exhibited higher plant heights in the warming treatment and the lowest ones under reduced rainfall. This could explain the enhanced productivity with warming detected in our study given the high correlation detected between this response variable and plant height at the community level. Previous studies have also found that plant height and productivity increase with higher temperature (Debouk et al., 2015; Zhang et al., 2015; Habermann et al., 2019a, 2019c) as a result of an acceleration of biochemical reactions (Copeland, 2000).

4.2. Tree canopy had an influence on net primary productivity and biomass digestibility

We observed different trends of net primary productivity and biomass digestibility according to the habitat type. Thus, plant communities located beneath tree canopy were less productive than those located in open grasslands but exhibited biomass of higher quality. The reduced productivity under tree canopy could be explained by the microclimatic alterations induced by tree shade. As stated by previous studies (Benavides et al., 2009; Hussain et al., 2009; Pang et al., 2019), the level of shade and its duration affect the physiological processes of plants, decreasing plant carbohydrate fabrication and net dry matter production. In contrast, other studies showed a positive effect of tree canopy on plant growth attributed to the reduced incoming solar irradiation and the temperature attenuation provided by the tree canopy (Frost and McDougald, 1989; Belsky, 1994; Moreno, 2008).

Some studies revealed that tree enhance water availability by hydraulic lift from the wettest soil layers and water infiltration (Joffre and Rambal, 1988; Ludwig et al., 2003, 2004a, 2004b), whereas other studies found the opposite pattern due to the large competition for resources between trees and herbaceous plants (Fay et al., 2003; Köchy et al., 2008). All these findings suggest that the impact of tree canopy on herbaceous productivity is highly context-dependent, being strongly influenced by the environmental characteristics of the ecosystem. Our results indicated that between-habitat type differences in productivity were due to an ameliorating effect of the tree on thermal stress since similar values of water stress

were detected in both habitat types. Therefore, these results suggest that trees exercise a buffer effect on temperature but not on available water, potentially reducing the impact of climate change on the ecosystem processes quantified in this study (as explained in detail below). Similarly, Mordelet and Menaut (1995) comparatively analyzed ten studies where the enhancement or suppression of biomass production by tree canopies was unrelated to annual rainfall.

Interestingly, we found the opposite trend for pasture digestibility, with higher values under tree habitat compared with those located in open grasslands. This enhanced pasture quality in more shaded microsites could be explained by the higher levels of soil fertility found under tree canopies, which provide soils large amounts of organic matter through leaf fall (Treydte et al., 2007, 2008; Ludwig et al., 2008; Cubera et al., 2009; Barnes et al., 2011). In addition, these microsites are usually enriched in nutrients due to the foraging patterns of large herbivores, which spend more time in this habitat to take shelter from the sun and consume acorns (Belsky et al., 1989; Belsky, 1994; Bardgett et al., 1998; Dijkstra et al., 2006).

4.3. Trees buffer the impact of climate change on biomass digestibility but not on net primary productivity

Our findings revealed that tree canopy could buffer the impact of climate change regarding only biomass digestibility. Both habitat types exhibited similar patterns in response to climate manipulation, with enhanced productivity under warming and decreased productivity under reduced rainfall. However, the decrease of biomass digestibility under warming conditions was only significant in open grasslands but not under tree canopy. This buffering effect of the tree canopy was likely promoted, at least partially, by an ameliorating effect of tree shade on thermal stress, which has been identified in our study as the main abiotic stressor that reduced biomass digestibility. Under these more favourable conditions, a slight increase of temperature was not likely enough to cause significant changes in the phenological cycle of the plant communities inhabiting this habitat, thus reducing the impact of warming on biomass digestibility. It has been proven that tree shade has a buffering effect on thermal stress and excess irradiance, with potential benefits on overall crop productivity and quality (Lin, 2007; Lott et al., 2009; Bayala et al., 2014; Sida et al., 2018).

Tree buffering effects on biomass digestibility could be also a consequence of the alterations of plant species composition and functional structure induced by tree shade. We detected a marked functional trait selection according to the microclimatic conditions generated under the tree canopies. On the one hand, we observed that plant communities located under trees developed higher SLA values, likely as an effective strategy to increase light uptake in this habitat where light is limited (Steinger et al., 2003; Curt et al., 2005). On the other hand, these plant communities growing beneath tree canopy exhibited a higher content in nitrogen, likely as a result of the mineralization of excess organic matter and nutrients through leaf fall (Gallardo, 2003; Cubera et al., 2009; Barnes et al., 2011) and the higher accumulation of herbivore droppings (Belsky et al., 1989; Kendall et al., 2006; Tucker et al., 2008). These functional attributes induced by tree canopy, which often confer higher nutritional value on plant biomass, probably attenuated the impact of simulated climate change on biomass digestibility. Further studies including a large array of functional traits are necessary to assess whether plant communities growing under tree canopies contain plant species inherently more adapted to climate change than those dominant in open grasslands.

5. Conclusions

This study provides an analysis of the major potentially influencing factors that can affect net primary productivity and biomass digestibility under different climatic scenarios. Although we detected an expected negative effect of reduced rainfall on net primary productivity, the strongest influence was exerted by warming. Higher temperatures caused an increase in net primary productivity and a decrease in biomass digestibility,

indicating a trade-off between both ecosystem processes. Interestingly, our findings highlight the relevant role of scattered trees in silvopastoral ecosystems such as Mediterranean dehesas due to their potential effect in mitigating the impact of climate change on one key ecosystem process (biomass quality). However, it is not easy to unravel the direct from the indirect effects that tree presence has on the studied ecosystem functions, because tree canopy can buffer the climate change impact directly through an alleviation of thermal stress, but also causes significant shifts in the functional structure of plant communities, with potential indirect repercussions on net primary productivity and biomass digestibility. These results could be used to develop strategies of controlling tree density in order to be applied as a useful tool in the management plans of agroforestry systems due to its potential ability to mitigate and adapt to the negative effects derived from predicted aridity increase in the Mediterranean area.

CRediT authorship contribution statement

Maria Dolores Hidalgo-Galvez: Data curation, Formal analysis, Investigation, Methodology, Writing – original draft, Writing – review & editing. **Karim Barkaoui:** Formal analysis, Writing – review & editing, Supervision. **Florence Volaire:** Formal analysis, Writing – review & editing, Supervision. **Luis Matías:** Formal analysis, Writing – review & editing, Supervision. **Jesús Cambrollé:** Writing – review & editing, Supervision. **Pilar**

Fernández-Rebollo: Formal analysis, Writing – review & editing. **Maria Dolores Carbonero:** Methodology, Writing – review & editing. **Ignacio Manuel Pérez-Ramos:** Funding acquisition, Project administration, Conceptualization, Investigation, Methodology, Formal analysis, Writing – review & editing, Supervision.

Declaration of competing interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

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Appendix

Note **Table A.1:** Calibration of Near Infrared Reflectance spectroscopy technique (NIRS): Before use it, the equipment calibration was carried out using 138 pasture samples collected in the same sampling area. These samples were analyzed in the Pasture and Forage laboratory of SERIDA (Asturias), using the standardized procedure PNTNANIR001, an accredited test under ISO-17025 standard (ENAC) law.

Calibration was performed using MPLS regression (modified partial least squares). The statistics that were used to select the best equations were: Standard Calibration Error (SCE), determination coefficient of calibration (R^2), Standard Cross-Validation Error (SCVE) and cross-validation determination coefficient (r^2). The statistic of calibration equations used are shown in **Table A.1**. Samples were scanned using two different devices, depending on the sample quantity. On the one hand, Muglight device for samples with more quantity with high intensity light (model A122100, ASD Inc.) and, on the other hand, for samples with low quantity, a reflectance probe of bifurcated fiber optic of 3 mm diameter.

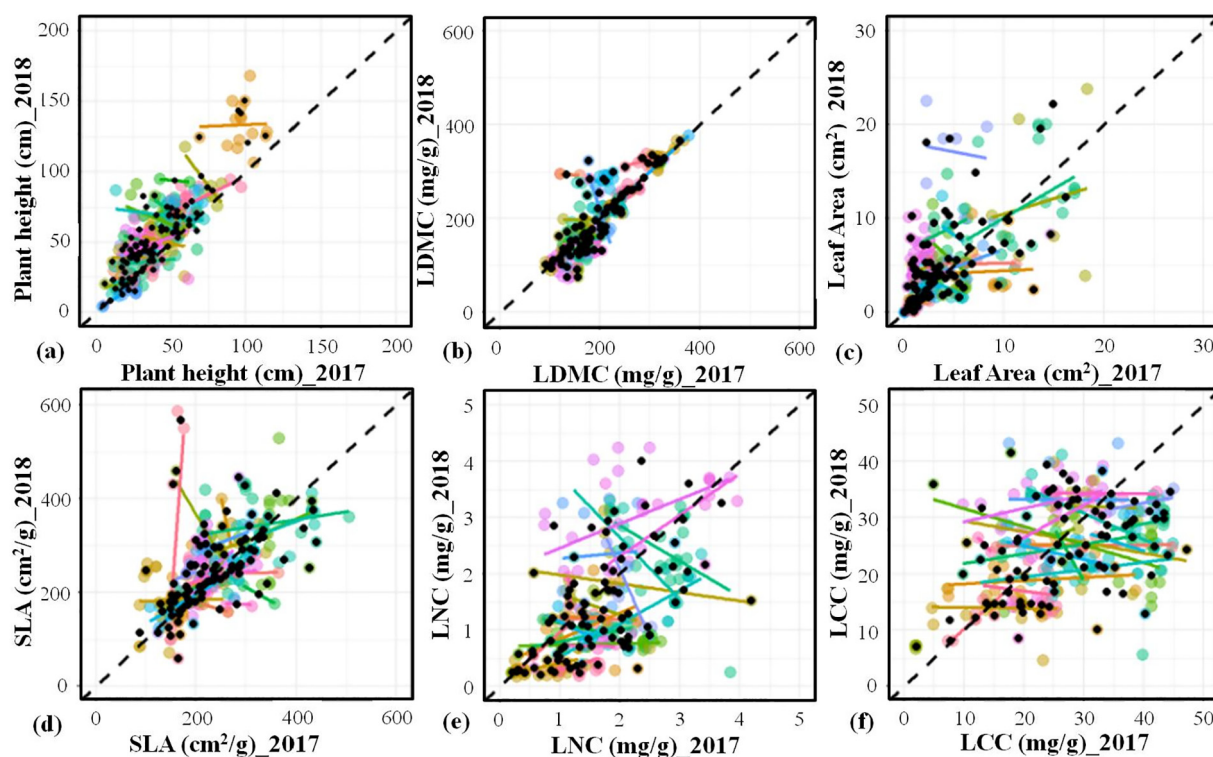


Fig. A.1. Relationship between trait values from 2017 and 2018. Plant height (cm) (panel a), LDMC (mg/g) (panel b), leaf area (cm^2) (panel c), SLA (cm^2/g) (panel d), LNC (mg/g) (panel e) and LCC (mg/g) (panel f). Colored data points represent individual species trait measurements; black points represent species means. Colored thin lines are linear regression for each species. Black dashed line is the 1:1 line.

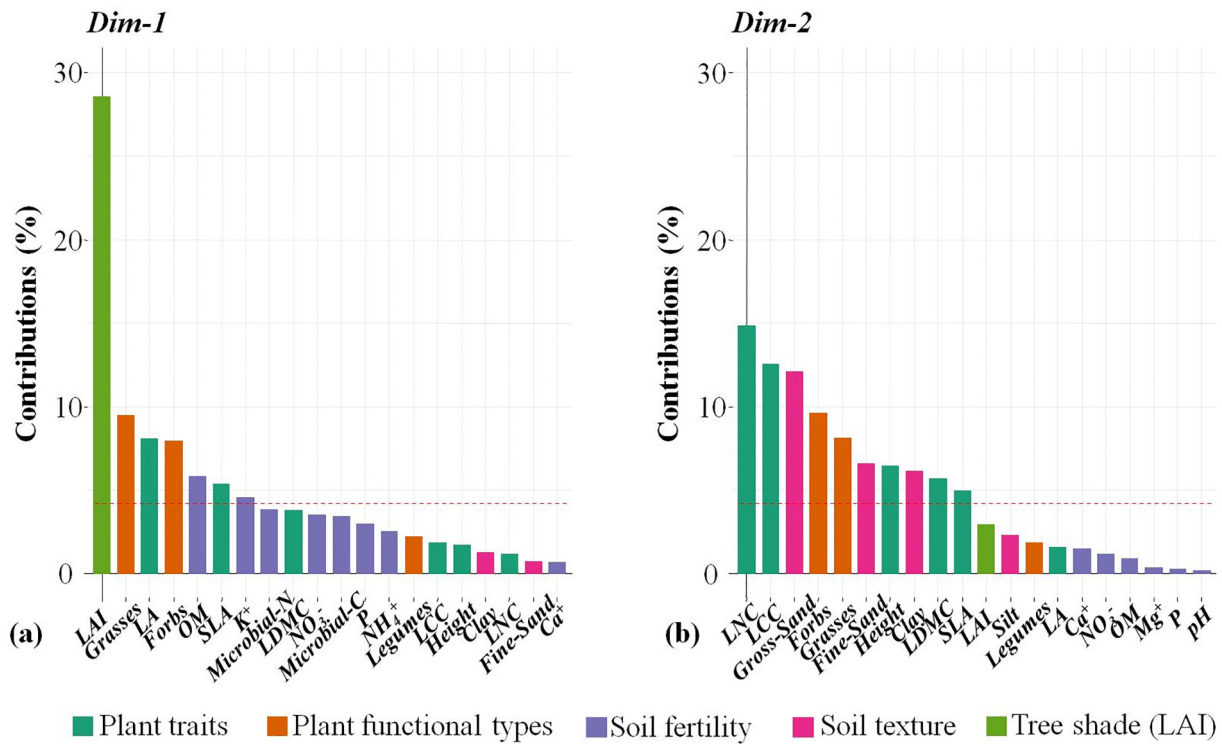


Fig. A.2. Factor loadings from the first (panel a) and second (panel b) dimensions of the MFA used to describe the 144 sampling plant communities according to their species composition, their plant functional structure and their environment (see details in Fig. 2 of the manuscript).

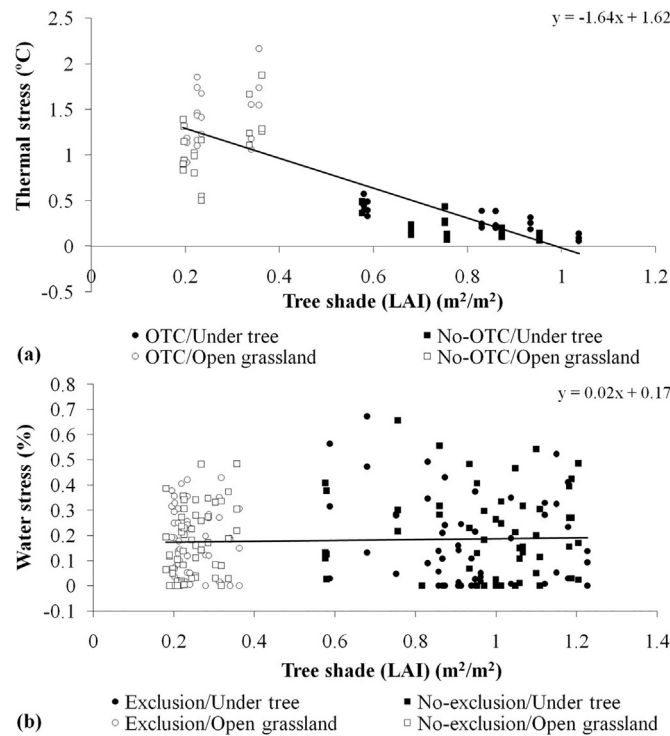


Fig. A.3. Impact of tree shade (represented as leaf area index, LAI) on thermal stress (°C) (panel a) and water stress (%) (panel b). Closed symbols represent plots located under tree habitat and opened symbols represent plots in open grassland habitat. In the panel a, circle symbols represent climatic treatments that increase the temperature by means of the use of open top chambers (warming and warming + drought), whereas square symbols represent climatic treatments without open top chambers (control and drought). In the panel b, circle symbols represent the climatic treatments that decrease the precipitation by means of the use of rainfall reduction structures (drought and warming + drought), whereas square symbols represent the climatic treatments without rainfall reduction structures (control and warming). The line is the result of the linear regression for the whole data set, whose equation is also indicated. Three years of sampling (2017, 2018 and 2019) are represented.

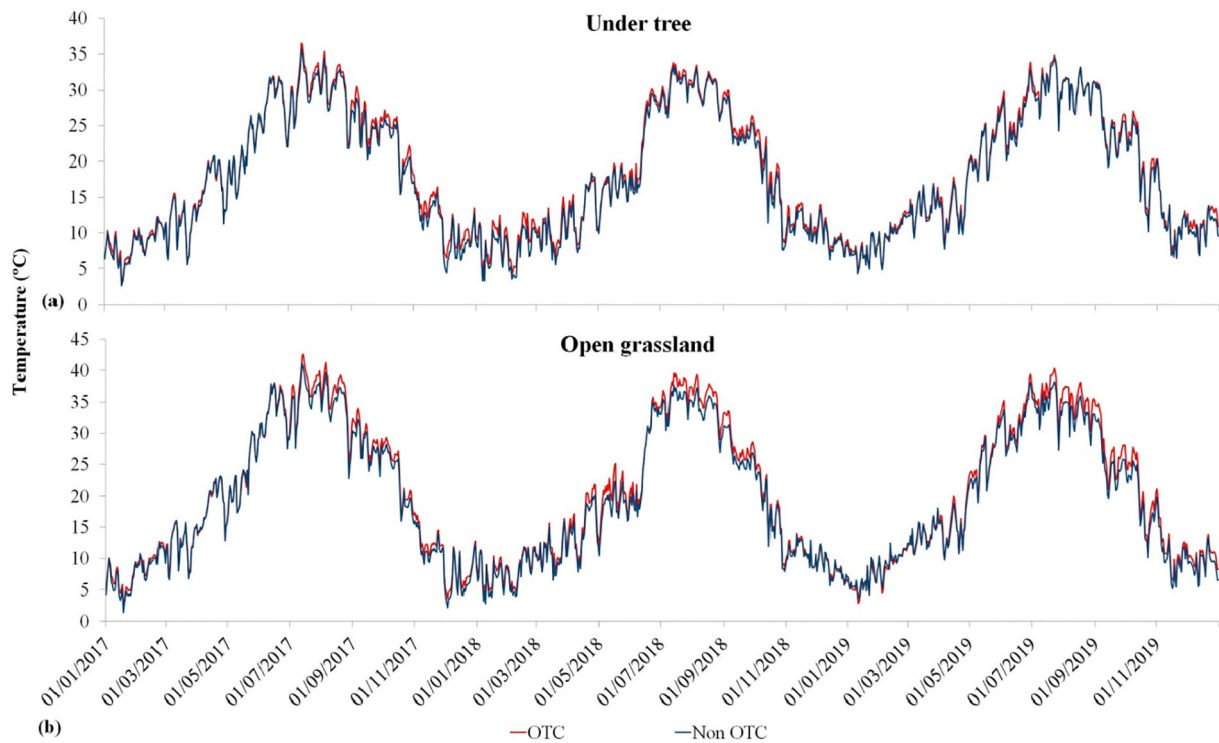


Fig. A.4. Average daily temperature (°C) during the study period (2017–2019) in the two habitat types: under tree (panel a) and open grassland (panel b). Red line (OTC) indicates the average of warming and warming + drought treatments, and blue line (non-OTC) indicates the average of control and drought treatments. Tukey's post-hoc test showed significant differences between temperature scenarios (OTC versus non-OTC) for both habitats.

Table A.1
Calibration equations with probe and muglight from dry pastures.

Parameter	Samples number		Average ^a		Rank		SD ^b		SCE ^c		R ^{2d}		SCVE ^e		r ^{2f}	
	Probe	Muglight	Probe	Muglight	Probe	Muglight	Probe	Muglight	Probe	Muglight	Probe	Muglight	Probe	Muglight	Probe	Muglight
Biomass digestibility	126	127	59.90	58.91	38.45–92.56	38.45–92.56	11.93	12.20	1.409	1.26	0.99	0.99	1.82	2.17	0.98	0.97

^a Average of calibration collective.
^b Standard deviation of calibration collective.
^c Standard calibration error.
^d Determination coefficient of calibration.
^e Standard cross-validation error.
^f Cross-validation determination coefficient.

Table A.2
Abiotic differences among climatic treatments for the two habitat types considered in this study. Thermal stress (°C) and water stress (%) is represented between January and July. The mean of 2017, 2018 and 2019 is shown, differentiating habitat types (under tree, open grassland) and climatic treatments (control, warming, drought and warming + drought). Superscript letters (lowercase for under tree habitat and uppercase for open grassland) are the result of factorial ANOVA for thermal and water stress, and repeated-measures ANOVA for soil volumetric water content (% SVWC) in winter, spring, summer and autumn, respectively, and subsequent post-hoc Tukey analyses were performed to compare the four climatic treatments within each habitat.

Habitat type	Climatic treatment	Thermal stress (°C)	Water stress (%)	SVWC (%)	SVWC (%)	SVWC (%)	SVWC (%)
		(January–July)	(January–July)	Winter	Spring	Summer	Autumn
Under tree	Control	0.22 ^a	0.20 ^a	17.59 ^a	15.35 ^a	10.37 ^a	15.30 ^a
	Warming	0.33 ^a	0.22 ^a	17.20 ^a	14.94 ^a	10.52 ^a	14.66 ^a
	Drought	0.22 ^a	0.16 ^a	17.92 ^a	15.46 ^a	11.54 ^a	15.24 ^a
	Warming + drought	0.26 ^a	0.18 ^a	16.08 ^a	14.91 ^a	11.23 ^a	14.74 ^a
Open grassland	Control	1.14 ^A	0.16 ^A	26.86 ^A	19.55 ^A	12.96 ^A	19.77 ^A
	Warming	1.58 ^B	0.19 ^A	23.75 ^A	17.03 ^A	10.77 ^A	16.97 ^A
	Drought	1.08 ^A	0.20 ^A	21.01 ^A	15.03 ^B	9.94 ^A	16.18 ^A
	Warming + drought	1.26 ^A	0.13 ^A	19.63 ^A	14.17 ^B	9.41 ^A	14.61 ^A

Table A.3

Mean values and standard deviation of CWMs of the main functional traits [plant height (cm), leaf dry-matter content (LDMC, mg/g), specific leaf area (SLA, cm²/g), leaf nitrogen content (LNC, mg/g) and leaf carbon content (LCC, mg/g)] and plant functional types (grasses, %) by habitat (under tree and open grassland) and climatic treatments (control, warming, drought and warming + drought). The mean of sampling years (2017, 2018 and 2019) is shown. Letters (lowercase for habitat located under tree and uppercase for open grassland) are the result of factorial ANOVA and subsequent post-hoc Tukey analyses were performed to compare the four climatic treatments within each habitat.

Habitat type	Climatic treatment	Plant height (cm)	LDMC (mg/g)	SLA (cm ² /g)	LNC (mg/g)	LCC (mg/g)	Grasses (%)
Under tree	Control	54.72 ± 16.97 ^{ac}	230.10 ± 52.36 ^a	267.30 ± 48.92 ^a	1.68 ± 0.59 ^a	27.40 ± 7.99 ^a	55.19 ± 25.23 ^a
Under tree	Warming	67.83 ± 19.32 ^b	224.88 ± 36.55 ^a	265.06 ± 36.49 ^a	1.35 ± 0.58 ^b	25.28 ± 5.93 ^a	62.68 ± 23.45 ^a
Under tree	Drought	48.70 ± 16.92 ^c	224.55 ± 41.00 ^a	268.56 ± 41.08 ^a	1.50 ± 0.62 ^{ab}	25.24 ± 5.77 ^a	58.38 ± 24.38 ^a
Under tree	Warming + drought	59.49 ± 23.81 ^{ab}	226.11 ± 42.64 ^a	248.72 ± 37.24 ^a	1.59 ± 0.57 ^{ab}	25.53 ± 6.21 ^a	57.13 ± 25.59 ^a
Open grassland	Control	52.70 ± 21.19 ^A	202.34 ± 36.66 ^{AB}	216.48 ± 46.52 ^A	1.29 ± 0.49 ^A	22.22 ± 5.31 ^A	36.17 ± 23.60 ^A
Open grassland	Warming	62.90 ± 20.18 ^B	194.41 ± 41.66 ^{AB}	219.82 ± 38.06 ^A	1.08 ± 0.62 ^A	22.06 ± 6.43 ^A	41.41 ± 24.58 ^A
Open grassland	Drought	52.97 ± 15.87 ^A	212.23 ± 37.84 ^A	204.41 ± 42.14 ^A	1.39 ± 0.75 ^A	23.55 ± 7.52 ^A	39.80 ± 22.79 ^A
Open grassland	Warming + drought	59.14 ± 18.28 ^{AB}	188.50 ± 51.39 ^B	217.82 ± 52.31 ^A	1.27 ± 0.74 ^A	22.60 ± 6.49 ^A	43.05 ± 24.75 ^A

Table A.4

Abiotic differences among the three sampling years (2017, 2018 and 2019) in thermal stress (°C) and water stress (%) between January and July. The mean of the 4 climatic treatments (control, warming, drought and warming + drought) is shown. Data on soil volumetric water content (% SVWC) for summer 2018 are not available due to methodological constraints. Superscript letters are the result of ANOVA-one way and subsequent post-hoc Tukey analysis where the three sampling years were compared.

Year	Thermal stress (°C) (Jan–Jul)	Water stress (%) (Jan–Jul)	SVWC (%) winter	SVWC (%) spring	SVWC (%) summer	SVWC (%) autumn
2017	0.93 ^a	0.19 ^a	21.68 ^a	15.38 ^a	11.22 ^a	12.05 ^a
2018	0.64 ^b	0.05 ^b	19.68 ^b	19.13 ^b		18.73 ^b
2019	0.73 ^a	0.30 ^c	18.65 ^b	12.91 ^c	10.46 ^a	17.02 ^c

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