

Pollination efficiency in farmland landscapes: exploring the relative roles of spillover, dilution and complementarity between habitats

Maxime Ragué, Vincent Bretagnolle, Olivier Martin, Thomas Perrot, Jean-Luc Gautier, Florence Carpentier, Sabrina Gaba

▶ To cite this version:

Maxime Ragué, Vincent Bretagnolle, Olivier Martin, Thomas Perrot, Jean-Luc Gautier, et al.. Pollination efficiency in farmland landscapes: exploring the relative roles of spillover, dilution and complementarity between habitats. Landscape Ecology, 2022, 37 (9), pp.2413-2433. 10.1007/s10980-022-01482-0. hal-03738600

HAL Id: hal-03738600 https://hal.science/hal-03738600v1

Submitted on 15 Nov 2023

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

Pollination efficiency in farmland landscapes: exploring the relative roles of spillover, dilution and

2 complementarity between habitats 3 4 Maxime Ragué^{1,2}, Vincent Bretagnolle^{1,3} Olivier Martin⁴, Thomas Perrot^{1,2}, Jean-Luc Gautier^{1,2}, Florence 5 Carpentier^{5,6}, Sabrina Gaba^{1,2,3} 6 7 ¹ Centre d'Etudes Biologiques de Chizé, UMR7372, CNRS & Université de La Rochelle, F-79360 Villiers-en-Bois, 8 9 ² USC 1339 Centre d'Etudes Biologiques de Chizé, INRAE, 76390 Villiers-en-Bois, France 10 ³LTSER « Zone Atelier Plaine & Val de Sèvre », 79360 Villiers-en-Bois, France 11 ⁴ BioSP, INRAE, 84914 Avignon, France 12 ⁵ Université Paris-Saclay, INRAE, UR MaIAGE, 78350, Jouy-en-Josas, France. 13 ⁶ AgroParisTech, 75005, Paris, France 14 15 16 Corresponding author: sabrina.gaba@inrae.fr (+33 5 49 09 96 01) 17 18 Maxime Ragué (0000-0001-9100-6923) 19 Vincent Bretagnolle (0000-0002-2320-7755) 20 Thomas Perrot (0000-0003-3881-1370) 21 Florence Carpentier (0000-0001-6204-2220) 22 Sabrina Gaba (0000-0002-7145-6713)

Abstract

Context Recognized as a critical ecosystem service in farmland, pollination is threatened by the decline of pollinators, notably due the homogenization of the landscape and the decline of floral resources. However, there is still a limited understanding of the interplay between landscape features and the pulses of floral resources provided by mass-flowering crops.

Objective The goals of this study were to (i) determine how pollination efficiency varies with the amount of floral resources at field and landscape scales through the oilseed rape (OSR) flowering period and (ii) quantify the magnitude of the pollination processes involved.

Methods Pollination efficiency (fruiting success) was measured using OSR plant phytometers placed in grasslands, cereals and OSR fields varying in quantity of floral resources at both field and landscape scales. The individual contributions of different processes to pollination were determined using a bagging experiment on plant phytometers.

Results Pollination efficiency was enhanced during both the temporal period and in landscapes with a high amount of OSR flowers, and semi-natural habitats as a result of higher pollinator presence. The bagging experiment also supported a complementarity between habitats for pollinators, as insect-pollination in grasslands and cereals was higher after OSR flowering, especially in OSR-rich landscapes, in regard to large-insect-pollination.

Conclusions The floral resource availability drives insect-pollination through attraction, spillover, and spatial and temporal complementarities between habitats. These results suggest that maximizing pollination efficiency in farmland landscapes partly consisting of OSR fields should include a combination of habitats that provide continuous floral resources.

Key words: Insect-pollination | Landscape composition | Semi-natural habitats | Temporal matching

Introduction

Agricultural landscapes are temporally and spatially highly dynamic mosaics of annual crops and seminatural elements. Landscape features such as crop diversity, or semi-natural habitat (SNH) surrounding fields, are increasingly recognized as key factors determining biodiversity and the associated ecosystem services (Fahrig et al. 2011, 2015; Martin et al. 2019; Dainese et al. 2019). However, despite the presumably strong effects of crop phenology or temporal changes in land use on the availability of resources and nesting habitats for species (Schellhorn et al. 2015), few studies have investigated the role of landscape features on biodiversity or ecosystem services at different times, either on an annual basis or at the crop rotation level (Rusch et al. 2011; Le Féon et al. 2013; Vasseur et al. 2013). Maintaining pollinating insects is critical because of their key role in agroecosystem functioning through the pollination of both crops and wild plants (Klein et al. 2007; Ollerton et al. 2011). Yet, these insects may be particularly sensitive to temporal resource limitation since they rely exclusively on nectar and pollen (Roulston and Goodell 2011), both available only for a few days in individual flowers (Rathcke and Lacey 1985; Elzinga et al. 2007).

Blooms of mass-flowering crops largely exceed the amount of floral resources provided by wild plants, and are therefore highly attractive to bees (Holzschuh et al. 2011). However, this high amount of floral resources occurs only during a limited period of time, a temporal pattern that may have cascading shortage effects on bees and pollination over the season in agricultural landscapes (Diekötter et al. 2014; Requier et al. 2017). Given that bees are optimal foragers and highly mobile organisms, temporal lags between flowering peaks either in a particular field or at the landscape scale may result in a mismatch between the local number of flowers to be pollinated and the number of available pollinating insects (Herbertsson et al. 2017; Fijen et al. 2019). Hence, sustaining pollinator populations in agricultural landscapes requires maintaining spatial and temporal continuity of floral resources as suggested by Mallinger et al. (2016). Ways to enhance the provision of complementary floral resources in space and time include increasing the amount of SNH (Rollin et al. 2013) and organically farmed (OF) fields that host higher density of wild plants (Requier et al. 2015; Wintermantel et al. 2019), or diversifying landscape composition (Meyer et al. 2007; Petersen and Nault 2014). However, increasing the diversity of flower availability may have considerable effects on pollinators and pollination. For instance, mass-flowering crops may attract pollinators from surrounding SNH and/or OF fields. These movements can negatively affect wild plant species that have overlapping flowering periods with crops, but at the same time may increase seed production in crops due to higher pollination efficiency (Holzschuh et al. 2011). A high density of mass-flowering crops can also have potential negative consequences on crop pollination (reduced seed production due to lower pollinator abundance) through the dilution of available pollinators during their flowering period (Holzschuh et al. 2011, 2016).

While a number of studies have explored the relationships between bees and landscape features, the efficiency of insect-pollination related to landscape elements while considering temporal variation in floral resource availability has been little studied. This study attempted to address this gap by exploring how pollination efficiency, in particular insect-pollination, varies with landscape composition both throughout the season, including the flowering period of a mass-flowering crop, and afterwards. Pollination efficiency was measured with a standardized metric, using the fruiting success of oilseed rape (*Brassica napus* L., OSR hereafter) plant phytometers as a proxy to allow comparisons in space and time. Several studies have used crop species as plant phytometers to estimate pollination efficiency: for example strawberry in SNH (Castle et al. 2019) or sunflower or radish in crop fields (Bennett and Isaacs 2014; Hass et al. 2018). In the present study, OSR plant phytometers

were placed in three different crop types (grassland, cereal and OSR fields), in fields varying in the amount of local floral resources, as well as in their surrounding landscapes, and at different times during OSR flowering periods. The focal fields thus differed both in the amount and duration of floral resource availability, e.g. high local floral resources in OSR fields during OSR blooming, and much lower after OSR flowering. Floral resources provided by wild plants were rather constant over time and moderate in grasslands, and lower in cereal and OSR fields (Frankl et al. 2005; Cole et al. 2017; Bourgeois et al. 2020).

The focal fields were distributed along three landscape gradients determined by the amount of SNH, OF and OSR crops. The amount of floral resources around fields thus differed, with a presumably high and constant availability of wildflowers in landscapes rich in OF (Hardman et al. 2016) and SNH (Cole et al. 2017), but a more heterogeneous availability of floral resources in landscapes with a high level of OSR, from a very high amount during OSR flowering to reduced resources outside the flowering period.

OSR is a mass-flowering crop that is mainly self-pollinated, however insect-pollination, involving both large insects such as honeybees (Lindström et al. 2016; Perrot et al. 2018) and smaller insects such as wild bees (Zou et al. 2017; Perrot et al. 2018) or to a lesser extent hoverflies (Jauker et al. 2012) also contributes to yield (Perrot et al. 2018). Therefore, to disentangle the contribution of the pollination processes involved (pollination via large or small insects, wind- or self-pollination), the phytometers were supplemented with a bagging experiment (using bags of different mesh sizes on selected phytometer branches). When placed in OSR fields, especially during peak OSR flowering, OSR phytometers were expected to show higher pollination efficiency because of higher OSR pollen availability and/or higher abundance of pollinators attracted to the flowering OSR field. The bagging experiment allowed differentiating between these two processes: no difference in fruiting success between control flowers compared to bagged flowers (excluding insects) would suggest higher levels of OSR pollen leading to higher pollination, while higher fruiting success in control flowers compared to bagged flowers would suggest a higher number of insect pollinators. A lower pollination efficiency of OSR phytometers was also predicted in landscapes with high amounts of OSR as compared to those with lower amounts of OSR, based on the hypothesis of the dilution of insect pollinators. Similarly, a higher pollination efficiency in landscapes with high amounts of SNH and OF was predicted as a result of higher pollinator abundance and attraction of insects to OSR. Finally, positing complementarity between habitats, a higher pollination efficiency was predicted for OSR phytometers placed in cereals and grasslands after the peak OSR flowering than during this peak, further predicting that the pattern would be enhanced in OSR-rich landscapes by a spillover effect.

Materials & methods

Study site

87

88

89

90 91

92

93

94

95

96 97

98

99

100

101

102

103

104

105

106

107

108

109

110111

112

113114

115

116

117

118

119

120121

122

123124

125

The study was carried out in spring from 2015 to 2019 in the Zone Atelier Plaine & Val de Sèvre (hereafter ZA-PVS; Fig. 1a), a research site of 435 km² in central-western France (46°14'N 0°28'W) that is part of the Long-Term Social-Ecological Research (LTSER) network. Land cover was monitored each year at the field scale by field workers and recorded in a GIS database (Bretagnolle et al. 2018). The study site was dominated by annual crops (almost 75% on average during the five years of study) that include more than 30 crop types. In the cultivated area, the most dominant crops during the study period were cereals as wheat or barley (54%) followed by sunflower (12%) and maize (12%) which flower in summer, then OSR (8%) and peas (4%). For this study, the amount of OF was calculated including all organically farmed fields. These covered between 6% (2015) to 9% (2019) of the

cultivated area. These OF fields consisted mainly of cereals (about 50%), followed by sunflower (10%) while less than 1% were OSR fields. Grasslands were excluded because these were included in the amount of SNH, together with hedgerows and fallow and set aside land. Grasslands encompassed meadows (permanent grasslands >6 years old) and alfalfa, clover, or mixed leguminous/grass (temporary grasslands, generally <5 years old).

129130131

132

133

134

135

136

137

138

139140

141

142

143

144

145

146

147

148

149

150

151152

153

154

155156

157

158

159

160

161

162

163

126

127

128

#Fig. 1 approximately here#

OSR plant phytometers

OSR plants were used as phytometers to measure pollination efficiency. Phytometers allow pollination efficiency to be studied independently of resource availability (e.g. water, nutrients) so that variation is only influenced by insect abundance and pollen availability (Woodcock et al. 2014; Castle et al. 2019). The fruiting success was used as a proxy of pollination efficiency. Following Jauker et al. (2012) fruiting success was measured by the ratio of pods and flowers of an OSR plant branch (Fig. 1b). Individual OSR plants were collected in cultivated fields at the end of March each year. Plants were transplanted in 3 L pots and placed in an insect-proof greenhouse to avoid any insect-pollination. Potted plants were watered as needed and no fertilizer was added. A total of 616 OSR plant phytometers were used over the study period (198 in 2015, 196 in 2016, 94 in 2017, 58 in 2018 and 70 in 2019).

From the OSR peak flowering (the beginning of April, though this depends on the year) to post-flowering (mid-May), phytometers were placed in OSR fields (124 fields in total over the five years), cereal fields (110 fields in four years) and grasslands (46 fields in three years; Fig. 1a) for four days, corresponding to the average flowering period of an individual OSR flower (Mesquida and Renard 1981). Two plant phytometers were placed per field, one in the field core and one at the edge (Fig. 1b), to account for variation in pollinator communities between these two field compartments (i.e. small pollinators are generally more abundant at field borders; Ricketts et al. 2008; Garratt et al. 2018), and therefore potential differences in the magnitude of small insect-pollination. On each OSR phytometer, four branches of similar size and flowering stage were selected, one of which became a control branch. On this branch, open flowers were tagged with two colour rings, delineating along the branch the portion in open flower stage. After four days, the number of flowers that further opened in the field were counted and identified using another colour mark on the branch (Fig. 1b). Plants were then placed back in the insect-proof greenhouse for pod ripening. When plant phytometers reached maturity (i.e. after seed maturation), all pods on each experimental branch were collected, either under treatment or control (see below and Appendix A for details on experimental protocols). Fruiting success (in %) was obtained by counting the number of flowers that produced pods on each branch between the marks and then calculating their proportion. 12.3% of the phytometers were excluded from the study because of missing data, e.g. dead plants, broken or dried branches that did not produce pods (7.9% of fields were without data).

Pollinator exclusion by bagging

For each OSR phytometer, the three other branches apart from the control branch were selected for exclusion treatments. Exclusion was used to assess the relative magnitude of large and small insect-, wind-, and self-pollination (Fig. 1b) and to separate from the effects of higher pollinator abundance from effects of higher amounts of OSR pollen. The exclusion treatments consisted of covering a branch with bags of different mesh sizes

depending on the study target (Fig. 1b): (i) an Osmolux bag allowing air but not pollen flow (Osmolux treatment: OS) thus permitting only self-pollination, (ii) a mesh of 0.6 mm allowing only wind- or self-pollination (small mesh treatment: SM), and (iii) a mesh of 3.0 mm allowing wind-, self-pollination and pollination by small insects (small wild bees and hoverflies), with an abdomen thinner than 3 mm (large mesh treatment: LM). The absence of a bag on the control branches allowed self-, wind- and insect-pollination by small or large insects (honeybees and bumblebees with an abdomen larger than 3 mm). Self-pollination was estimated using fruiting success (in %) in the OS treatment, and was obtained exactly as in control treatment, so both were easily compared to each other. Self-pollination was expected to be high, as it is the major pollination process in OSR (Perrot et al. 2018). Given the rather minor difference between control and self-pollination (see results below), contribution of other pollination processes, i.e. insect- and wind-pollination, to fruiting success were expected to be small. Dimensionless effect sizes (varying between -1 and 1) were therefore used to compare insect- and windpollination. They were computed between paired treatments as the ratio of the difference in fruiting success between treatments from the sum of the paired fruiting success rates. Paired treatments included control versus SM (complete exclusion of insects) to assess the magnitude of insect-pollination (Eq.1); control versus LM (exclusion of large insects) to assess the magnitude of large insect-pollination (Eq. 2); LM versus SM to assess the magnitude of small insect-pollination (Eq. 3); and SM versus OS (exclusion of wind-pollination) to assess the magnitude of wind-pollination (Eq. 4).

```
181 (Eq. 1) Insect – pollination = \frac{Fruiting\ success\ (C) - Fruiting\ success\ (SM)}{Fruiting\ success\ (C) + Fruiting\ success\ (SM)}

182 (Eq. 2) Large insect – pollination = \frac{Fruiting\ success\ (C) - Fruiting\ success\ (LM)}{Fruiting\ success\ (C) + Fruiting\ success\ (LM)}

183 (Eq. 3) Small insect – pollination = \frac{Fruiting\ success\ (LM) - Fruiting\ success\ (SM)}{Fruiting\ success\ (SM) - Fruiting\ success\ (SM)}

184 (Eq. 4) Wind – pollination = \frac{Fruiting\ success\ (SM) - Fruiting\ success\ (OS)}{Fruiting\ success\ (SM) + Fruiting\ success\ (OS)}
```

164

165

166167

168

169

170

171

172

173

174

175

176

177

178

179

180

185

186

187

188

189

190

191

192

193

194

195

196

197

198

199

200

201

Thus, for example using Eq. 2, a positive value of large insect-pollination indicated that large insects contributed to pollination, as fruiting success was higher in control treatment compared to treatment excluding large insects. A null value meant that large insects' contribution to pollination was negligible. Negative values were unexpected but were found in some situations, which may suggest that the contribution of large pollinators could be compensated by other processes (wind-, self-pollination and pollination by small pollinators) in the absence of large pollinators, and/or by pollen production increase.

Four measures of OSR floral seasonality and availability in focal fields

To account for spatial and temporal variation in floral resource quantity at both local and landscape scales, four measures of floral resources were derived from additional data (Fig. 1c).

(i) **Field Size.** Field size was used as a proxy of the quantity of available flowers at field scale, assuming that the amount of flowers was proportional to the field size, i.e. larger fields had more floral resources than smaller ones. Floral resources referred to OSR flowers in OSR fields, as wildflowers were minor compared to OSR flowers in this crop, and to wildflowers (mainly weed species) in grasslands and cereals. Previous studies in our site indicated a higher diversity and abundance of weeds in grasslands (average \pm sd: 22 ± 10 species and 308 ± 128 plants/m² respectively) than in cereals (12 ± 10 and 153 ± 176) or in OSR (17 ± 4 and 153 ± 83 ; Gaba et al. 2020; Bourgeois et al. 2020). By using field size as a proxy, the quantity of flowers was assumed to depend more on the field area than on the local flower density. Indeed, in the present study, field size varied from 1.1 to 27.6 ha for

OSR, from 0.7 to 33.7 ha for cereals and from 0.2 to 24.1 ha for grasslands. By contrast, flower density generally varies from 6.75 to 12.75 OSR plants/m² (Momoh and Zhou 2001), from 1% to 5% of wildflower density in cereal fields (Sidemo-Holm et al. 2021), and 1% to 13% in grasslands (Hegland and Boeke 2006), i.e. five to ten times less than the variation in field size.

202

203

204

205

206

207

208

209

210211

212

213214

215

216

217

218219

220221

222

223

224

225226

227

228229

230231

232

233234

235236

237

238239

240

- (ii) **Landscape Flower Quantity.** Floral resource quantity at the landscape scale varied with the amount of OSR, SNH and OF in the surrounding landscape at a given radius. Plant phytometers were placed along gradients of OSR flowers provided by OSR crops (%OSR), and wildflowers provided by SNH and OF (Cole et al. 2017; Sidemo-Holm et al. 2021). These were estimated in buffer zones excluding the focal fields (distances between focal field centroids and borders were 165.7 ± 58.9 m in OSR, 167.5 ± 59.4 m in cereals, and 116.7 ± 62.8 m in grasslands). For each crop type, fields were selected to maximize gradients of %OSR (min-max values in 250 m and 1000 m buffer radii: 0-55%; 0-25%), %SNH (0-72%; 0-38%) and %OF (0-76%; 0-53%). The landscape gradients were uncorrelated with crop type. The **Landscape Flower Quantity** provided by OSR, SNH and OF was estimated in buffer zones with various radii, specific for each type of pollination process and each crop type. At their respective buffer scales of maximum effect, %OSR, %SNH and %OF were then used as proxies of total floral resource availability.
- (iii) Phytometer-Regional Flowering Temporal Matching. Temporal matching between phytometer setting and regional OSR flowering in the study site that year was used to characterize the temporal lag between the flowering period of OSR and the deposition date of plant phytometers in the fields. This metric was calculated differently for plant phytometers placed in OSR fields and those placed in the other two crop types. First, generalized additive models (GAMs) was used to derive, separately for each year, the temporal dynamic of OSR flowering in the study site using % of OSR flower cover collected in OSR fields across the study site (from 0% when no OSR flowers were open in the field to 100% when OSR flowers covered the total area of the field; see Appendix B). For phytometers placed in OSR, from the fitted GAMs the period of OSR peak flowering was identified for each year as the range between the maximal estimated OSR flower cover and a relative 20% reduction of OSR flower cover. The choice of a 20% threshold was determined in order to balance the sampling dates with equal representation in both categories. The time period in days (ranging from 100 to 118 Julian days depending on the year) thus defined the peak flowering period. Phytometers placed in OSR fields during that period were considered to be placed at the peak flowering period, and those placed outside this range were considered to be placed after the peak flowering period. For phytometers placed in cereals and grasslands, a different threshold was used: a 60% reduction of OSR flower cover indicated the end of flowering (rather than 20% as in OSR fields). This difference in threshold accounted for the fact that the phytometers were placed later (on average 12 days) than in OSR fields, thus balancing sample sizes between the two categories required delaying the threshold value later in the season (Appendix B).
- (iv) **Focal Field-Regional Flowering Temporal Matching.** Temporal matching between OSR focal field flowering and OSR regional flowering in the study site for a given year referred to the temporal lag between the flowering of the focal OSR fields and the flowering of OSR at the site scale. This parameter was calculated only for plant phytometers placed in OSR fields. Here the % of OSR flower cover collected in OSR fields was used. Cumulating all data for a given year, a GAM was used to derive the deflowering pattern according to date and its intra-year variability. Using two quantiles (33% and 66%), for any given date three categories were defined based on % of flower coverage in a given focal field: OSR fields were either relatively early, i.e. below 33% deflowered

compared to average kinetics, relatively late (over 66%), or normal, i.e. within the range (33–66%; Appendix B).

Statistical analyses

242

243

244

245246

247

248249

250

251

252

253

254

255

256

257

258259

260

261

262

263

264

265

266

267

268269

270

271

272

273

274

275

276277

278279

280

281

In a first step, preliminary analyses were conducted to investigate the potential effect of the position of the plant phytometers in the field (core *versus* edge). As no significant differences were detected, the two values were averaged per treatment, field, and date (Appendix C). The sample size in our study was thus n = 308 fields. Differences in pollination efficiency were then explored between crop types by using a subset of the data in which all crops were investigated the same year. A linear model was used to test the effect of crop type (three levels) and year (three levels in common between crops) on phytometer fruiting success. For this first set of analyses, only data from the control branches were used.

In a second step, the effects of 'Phytometer-Regional Flowering Temporal Matching' and the amount of floral resources at both field and landscape scales were investigated on the fruiting success of plant phytometers. Again, only data from the control branches were used. Three models were built, one for each crop type, because (i) 'Phytometer-Regional Flowering Temporal Matching' was estimated differently for phytometers in OSR fields than for those in cereal fields and grasslands and (ii) the scales of influence of the landscape variables were assumed to vary with the focal crop type. Fruiting success of OSR phytometers was considered as the dependent variable, and 'Field Size', 'Landscape Flower Quantity' (%OSR, %SNH and %OF) and 'Phytometer-Regional Flowering Temporal Matching' as explanatory variables. The interaction of 'Field Size' and 'Landscape Flower Quantity' with 'Phytometer-Regional Flowering Temporal Matching' was also included (except in grasslands due to a smaller sample size) as a way to identify spillover and dilution processes. Year was included as a fixed effect factor with five, four and three levels respectively in OSR fields, cereal fields and grasslands to account for interannual dynamics of pollinating insects (Rollin et al. 2015; Perrot et al. 2018). For phytometers placed in OSR fields, the models also included 'Focal Field-Regional Flowering Temporal Matching' (a qualitative variable with three levels), in interaction with 'Field Size' and 'Phytometer-Regional Flowering Temporal Matching'. The models were estimated using the 'siland' method (Carpentier and Martin 2021) which enables the simultaneous estimation of the effect of all explanatory variables and the buffer radii of landscape variables i.e. radii of 'Landscape Flower Quantity' (%OSR, %SNH and %OF) without any prior assumption on the spatial scales of influence. The model residuals were assumed to be Gaussian (see details in Appendix D).

Finally, based on the bagging treatments, the variation of each pollination process (large and small insects-, wind- and self-pollination) with local, temporal and landscape factors for each crop type was investigated. A total of 15 linear models was built (three crop types with five pollination processes) to explore the effects of the amount of floral resources in the fields and in the landscape as well as 'Phytometer-Regional Flowering Temporal Matching'. Explanatory variables were the same as those used for the study of fruiting success of the control treatment. As the scale of effects of landscape features was expected to change with the pollination process, the 'siland' method was used.

All the analyses were performed using R software (version 3.6.2; R Core team 2019) with the 'siland' package (Carpentier and Martin 2020), 'car' for Type II analyses of variance (Fox et al. 2020), 'emmeans' (Lenth et al. 2020) for the estimations of effect sizes and 'BSDA' for unilateral sign tests (Arnholt and Evans 2017). For each model, residual diagnostics were looked to verify for normal hypothesis (quantile-quantile plots) and homoscedasticity (spread of residual did not change with predicted values; Kozak and Piepho 2018). The absence of spatial autocorrelation of residuals were also confirmed with variograms visualization (Dormann et al. 2007).

Finally, the landscape scales of influence were ensured to be well estimated, i.e. being the global rather than the local maximum of likelihood (profile likelihood visualization) and that the metrics of landscape composition did not correlate to each other at the estimated spatial scales (see Appendix E).

Results

282

283

284

285

286

287

288

289290

291

292

293

294295

296

297

298

299

300

301

302

303

304

305

306

307308

309

310

311

312

313

314

315316

317318

The fruiting success of the phytometer control branch (proportion of flowers producing a pod, ranging from 0% to 100%) varied with crop type (F = 10.23; df = 2; p < 0.001). Fruiting success was higher for phytometers in OSR fields (mean \pm sd: 49% \pm 28%, n = 135) than for phytometers in cereals (38% \pm 28%; n = 100) or grasslands (36% \pm 27%; n = 43). Year (df = 2; F = 13.74; p < 0.001) had also significant effects on fruiting success, but not its interaction with crop type (df = 4; F = 0.16; p = 0.96).

Factors affecting the fruiting success of OSR plant phytometers in OSR fields

Using 'Field Size' and 'Focal Field-Regional Flowering Temporal Matching', as two measures of floral resource availability at field scale, 'Landscape Flower Quantity' (measured by %SNH, %OF and %OSR) as a measure of landscape floral resource availability, and 'Phytometer-Regional Flowering Temporal Matching' as a measure of temporal deviation in flowering, as well as the interactions between these variables and the year (fixed effect factor), 44.3% of the variance in the fruiting success of phytometers placed in OSR fields was explained (Table 1a). Fruiting success strongly varied with year (15.2% of explained variance), 'Phytometer-Regional Flowering Temporal Matching' (14.0%), 'Landscape Flower Quantity' (8.7% cumulated for %OSR, %SNH and %OF), and the interaction between 'Phytometer-Regional Flowering Temporal Matching' and 'Field Size' (2.3%). 'Focal Field-Regional Flowering Temporal Matching' had no effect on fruiting success (Table 1a). Fruiting success was on average higher for phytometers in OSR fields during OSR peak flowering than after (Fig. 2a). The effect of 'Field Size' differed significantly between the two categories of 'Phytometer-Regional Flowering Temporal Matching' (Table 1a): during OSR peak flowering, fruiting success tended to decrease with 'Field Size', while post-peak it tended to increase (Appendix G). The fruiting success of phytometers was significantly modified by %OSR, %SNH and %OF (Table 1a): %OSR (Fig. 3a) and %SNH (Fig. 3d) had a positive effect on fruiting success while fruiting success decreased with %OF (Fig. 3g). The estimated scale of effect differed between landscape features, being small for %OSR (12 m outside the fields, meaning that considering the presence of an adjacent OSR field was enough to explain increased fruiting success in the focal field), high for %SNH (1772 m) and intermediate for %OF (210 m outside the fields).

#Fig. 2, 3 approximately here#

Fruiting success of OSR plant phytometers in cereal fields and grasslands

In cereal fields and grasslands, the statistical models explained respectively 35.8% and 54.6% of the variance in OSR phytometer fruiting success (Table 1b, c). As with phytometers in OSR fields, the fruiting success in cereal fields and grasslands varied with year, 'Phytometer-Regional Flowering Temporal Matching' and the amount of floral resources in the surrounding landscape; however, 'Field Size' had no effect (Table 1b, c). In cereals, the same pattern was observed as in OSR (Fig. 2b, Fig. 3b, e, h): fruiting success was higher during OSR flowering than after (Fig. 2b), increased with %OSR (Fig. 3b) and %SNH, but decreased with %OF, although non-

significantly for %SNH and %OF (Fig. 3e, h). A strikingly different pattern emerged with grasslands: as in OSR and cereals, the fruiting success of phytometers was higher during OSR flowering, though only slightly (Fig. 2c), increased with %OSR at landscape scale (Fig. 3c), but decreased with %SNH (Fig. 3f). The spatial extent of the effect of landscape differed between crop types, %OSR acted for example at higher spatial scale in grasslands (942 m) than in cereals (260 m) or in OSR (12 m; Fig. 3).

#Table 1 approximately here#

319

320

321322

323

324325

326

327

328

329

330331

332333

334335

336

337

338

339340

341

342

343

344345

346

347348

349350

351

352

353354

355

356

357

358

Pollination processes of OSR plant phytometers

The magnitude of the various pollination processes affecting the fruiting success of plant phytometers differed slightly between the three crop types. Self-pollination was the main pollination process of OSR phytometer whatever crop type where phytometers were set as fruiting success of the Osmolux treatment was close to fruiting success of the control treatment (Fig. 2 vs Fig. 4a). Self-pollination was however higher in OSR than in the two other crops, 41% average fruiting success compared to 49% in control (i.e., only 16% reduction compared to the control treatment). In grasslands and particularly in cereals, the reduction compared to the control was higher (Fig. 4a). Other pollination processes had therefore, overall, a small contribution, especially in OSR. Among them however, the second most dominant process of pollination was insect-pollination (Fig. 4c-e), being more important than wind-pollination, although in OSR (Fig. 4c), conversely to the other two crops (Fig. 4d, e), differences were minor. Of insects, large insects had the greatest effect in all crops, while the magnitude of small insects was greater than zero only in cereals (Fig. 4). Pollination processes also differed quite strongly according to the OSR flowering period during which the phytometer was placed, in addition to the crop in which it was placed (Fig. 4c-e). In OSR fields, higher fruiting success (Fig. 2a) in the peak flowering period resulted in significantly higher insectpollination (Fig. 4c). In contrast, insect-pollination dramatically increased in grasslands after the OSR flowering period (Table 2c; Fig. 4e). No such marked difference was detected for insect pollination in cereals, which was high in both periods (Fig. 4d). The relative importance of wind pollination versus insect pollination differed between crops, with wind pollination decreasing in grasslands after the OSR flowering period while remaining similar in OSR and cereals throughout the two periods (Table 2c; Fig. 4c, d, e). Self-pollination was higher during OSR flowering than after in all crops (Fig. 4b). Supplementary analyses also showed that self-pollination linearly decreased with the Julian date (Appendix H).

#Fig. 4 approximately here#

Insect-pollination was higher than wind pollination whatever the crop and season (though not necessarily significantly, e.g. in OSR), suggesting that the abundance of pollinators, rather than the abundance of pollen, was involved in the higher observed fruiting success (Fig. 4c–e). In addition, insect-pollination often responded in the same way as fruiting success to landscape features (Fig. 3, Fig. 5). In OSR fields, insect-pollination increased with %SNH (Fig. 5a) and decreased with %OF (Fig. 5d). There was an interaction between landscape composition and 'Phytometer-Regional Flowering Temporal Matching' on small and large insect-pollination (Table 2a): increasing %SNH improved small insect-pollination after OSR peak flowering (Fig. 5c), while large insect-pollination was negatively related to %OF only after OSR peak flowering (Fig. 5e). The increase in fruiting success with %OSR resulted from a joint increase of insect-, wind- and self-pollination (Appendix I). However, the increase in fruiting

success in cereals and grasslands resulted from an increase in large insect-pollination with %OSR, although this was the case only after OSR flowering in cereals (Fig. 5g, h; Table 2b, c). Small insect-pollination also increased with %SNH in grasslands (Fig. 5i), as opposed to the pattern observed for fruiting success (Fig. 3f) and wind-pollination diminished with %SNH (Table 2c; Fig. 5j).

The spatial extent of the effects of landscape varied with the pollination processes, being in general lower for wind than for insect pollination (e.g. 3 m and 747 m for %OSR in OSR crops). For insects, the spatial scales of landscape features were greater for large than for small insects (e.g. 321 m and 9 m for %SNH in OSR crops). Large insect-pollination was also shown to respond at different spatial scales depending on the focal crop type, as %OSR acted on large-insect pollination at a higher spatial scale in grasslands (1442 m) than in cereals (522 m; Fig. 5g, h). Self-pollination, which was expected to be constant across the conditions, was also modified by the amount of local and landscape floral resources in OSR, cereals and grasslands (Appendix H).

#Fig. 5 & Table 2 approximately here#

Discussion

The results of the present study showed that fruiting success of plant phytometers placed in OSR crop was higher in OSR peak flowering period, i.e. when OSR flower resources are highest, and was improved by neighbouring OSR fields. The pulse of OSR flowers was especially attractive for small insects that lead to an increase of fruiting success in OSR crops surrounded by a high amount of SNH. A temporal spillover occurred after the period of OSR flowering and involved large insects that dispersed from OSR to grasslands and cereal crops.

The phytometer approach revealed high heterogeneity in fruiting success rates (used as a proxy of pollination efficiency) even within crop types, suggesting that there may be a strong spatial and temporal variation in pollen limitation in our study site. Fruiting success of the OSR phytometers (49% on average) was within the range (from 31% to 75%) found in other studies on *Brassica* plant phytometers (Palmer and Zimmerman 1994; Hudewenz et al. 2014), though slightly lower than those estimated directly on OSR plants in crop fields (70%; Zou et al. 2017), including in our study site (65%; Perrot et al. 2018). Such pollination rates indicate a rather high abortion rate of flowers, i.e. from 51% in OSR fields to up to 62% and 64% in cereals and grasslands. Although other factors may also contribute to abortion, such as lack of water or nutrients that may physiologically limit plant phytometers (Marini et al. 2015), this is more likely to result, at least partly, from pollen limitation. Indeed, OSR pollen availability strongly varies between crop types (i.e. higher in OSR fields than in cereals and grassland), in the surrounding landscape of focal fields and throughout the flowering period of OSR. This lack of pollen in the air can reduce wind pollination (McCartney and Lacey 1991) at a low spatial scale, since wind-carried pollen rarely exceeds 50 m dispersion from OSR fields (Popławska et al. 2013).

Pollen limitation can also lower self-pollination later in the season. Lower airborne pollination may, however, be balanced out by higher insect-mediated pollination, as has been shown with radish or strawberry when used as plant phytometers in semi-natural habitats or crop fields (Hass et al. 2018; Castle et al. 2019). Pollinating insects carry pollen over long distances (Chifflet et al. 2011) and/or directly from opened flowers of plant phytometers by successively visiting different flowers (Rader et al. 2009). The bagging experiment performed in the present study confirmed that insect pollination was a major process in OSR pollination after self- and before

wind-pollination (Hudewenz et al. 2014; Perrot et al. 2018). The high insect-pollination rate found in cereals and grasslands suggests that this pollination process may compensate for low OSR pollen availability. As expected, a strong variation in the magnitude of insect-pollination was detected with the spatial and temporal conditions of the focal fields under study. In our site, pollinator richness ranged from 0 to 25 species (Rollin et al. 2015) and abundance from 0 to 68 individuals per 100 m² (Perrot et al. 2018). Moreover, pollinator richness and abundance strongly vary across years (Perrot et al. 2018), thus directly affecting the magnitude of insect-pollination.

In the present study, pollination efficiency in OSR fields was higher in periods and landscapes characterized by high availability of OSR flowers (Fig. 2a, 3a) in sharp contrast to predictions of the dilution hypothesis. This conflicts from most previous studies, which have detected such a dilution effect, finding that increasing the amount of mass-flowering crops in the landscape decreases large pollinator abundance, e.g. honeybees and bumblebees (Holzschuh et al. 2016; Bänsch et al. 2020), hence decreasing pollination efficiency in these crops (Shaw et al. 2020) or adjacent crops (Grab et al. 2017) and grasslands (Holzschuh et al. 2011). The gradients of %OSR used in the present study (from 0% to 26% in a buffer of 747 m outside OSR fields) were similar to those used in other studies (e.g. from 0% to 30% in a buffer of 1000 m; Holzschuh et al. 2011), thus other factors must explain this discrepancy. One possible reason may be the plant phytometer itself: dilution was found when plant phytometers used was not the blooming crop (e.g. strawberry plants with the blooming of apple flowers in Grab et al., 2017), or Primula veris with the blooming of OSR flowers (Holzschuh et al. 2011). Two other possible factors may also be involved in explaining the difference with our study. First, the rich and abundant pollinator community in our study site (Rollin et al. 2013; Bretagnolle et al. 2018) may avoid such a dilution effect more likely to be detected in depauperate pollinator communities (Tscharntke et al. 2012). Second, OSR accounted for about 8% of crops, and therefore a high load of OSR pollen in the air may have balanced out any dilution of pollinator insects (Waser and Price 2016; Cavallero et al. 2018). An increase in pollination efficiency with higher amount of OSR fields in the landscape was also found, but at a small scale (12 m), a distance consistent with the distance at which pollen disperses with wind (about 50 m; Popławska et al. 2013). Conversely, the spatial extent of the effect of %OSR on insect pollination was at much larger scale (e.g. 747 m for plant phytometers in OSR fields), underlying the ability of insect pollinators to forage over hundreds of metres (Greenleaf et al. 2007).

While the results of the present study did not support a dilution effect, patterns consistent with a reverse pattern was detected, in the form of an attractive effect of the bloom of OSR flowers. During peak OSR flowering, this was detected in the attraction of insects from SNH to OSR, as there was higher pollination efficiency of phytometers placed in OSR fields surrounded by a high amount of SNH (Fig. 3d), particularly for small pollinators (indicated by the bagging experiment; Fig. 5a, b, c). Such attraction operated at a larger scale for large pollinators (321m) than for small ones (9m), in accordance with the ability of large pollinators to forage at greater distances (Jackson and Fahrig 2012; Miguet et al. 2016). SNH are well-known sources of pollinators in agricultural landscapes, including bees (Steffan-Dewenter et al. 2002), hoverflies (Bommarco et al. 2012) and bumblebees (Hopfenmüller et al. 2014). The results of the present study thus confirm that SNH benefit wild pollinators in adjacent OSR fields (Bommarco et al. 2012), hence reducing pollen limitation (Cusser et al. 2016) and increasing pollination efficiency in insect-dependent crops (Bukovinszky et al. 2017; Raderschall et al. 2021) such as OSR (Bartomeus et al. 2014).

Finally, a temporal spillover mechanism was found to occur later in the flowering season when insects disperse from OSR fields after peak flowering to other crops and SNH, which become attractive (Fig. 4c, d, e; Fig.

5g). This process, already suggested, is related to the fact that pollinating insects follow the temporal pattern of floral resource availability in the landscape (Mandelik et al. 2012; Bretagnolle and Gaba 2015). At a seasonal scale, different habitats are therefore sequentially used, corroborating that the presence of OSR in a landscape benefits pollinator diversity (Kovács-Hostyánszki et al. 2013; Diekötter et al. 2014) and pollination of wild plants (Herbertsson et al. 2017) in adjacent grasslands. Such a spillover effect was also detected in cereal fields, as large insect-pollination was higher in cereals surrounded by a high amount of OSR, but only after peak OSR flowering. In contrast to the predictions however, a negative effect of organic farming was detected in small-bodied insects whatever the OSR flowering period, and in larger insects only outside OSR flowering. This effect may result from pollinator attraction to organic fields. Organically farmed cereal fields support a high diversity of wild plants (Henckel et al. 2015) which are visited by both wild pollinators (Sidemo-Holm et al. 2021) and honeybees (Wintermantel et al. 2019). The present study was inconclusive in regard to the possible effect of the amount of floral resources at the field scale possibly because of the crude measure of local availability of floral resources (field size was used as a proxy), which did not consider variation in the composition of floral resources.

Conclusion

The present study revealed complex and interacting mechanisms (i.e. attraction, spillover and temporal complementarity between habitats) through space and time, between crops, and between crops and semi-natural elements, depending on insect types. The amount of SNH in the landscape appeared to play a key role in pollination efficiency: it provided pollinating insects – especially small insects – to OSR crops, with a reverse spillover effect after OSR peak flowering. These results suggest that grasslands and cereals may act complementarily to OSR fields in supporting insects (and hence pollination) throughout the season in a landscape. Therefore, farmland landscapes that combine habitats providing nesting sites and floral resources after OSR flowering or, even better, continuous flowering resources across the seasons over the entire landscape, would contribute to supporting large pollinator populations and therefore improve pollination.

Acknowledgements

We would like to express our thanks to Marilyn Roncoroni, Alexis Saintilan, Quentin Van Hecke and Anthony Stoquaert for their help with the OSR phytometer experiment and pollinator trapping and identification (Marilyn Roncoroni). We also sincerely thank the farmers of the LTSER 'Zone Atelier Plaine & Val de Sèvre' for their involvement in our research programmes. MR was supported by a PhD grant from INRAE (Metaprogram SMACH) and the Nouvelle-Aquitaine region's Birdland project. SG, TP, OM, and FC are funded by INRAE, and VB by CNRS. The authors are grateful to the two anonymous reviewers whose comments greatly improved the manuscript.

Declarations

- **Funding.** This project was supported by the French Ministry of Ecology project (2017-2020 "Pollinisateurs"), the
- 471 ANR IMAGHO (ANR-18-CE32-0002) and the INRAE projects "ESPACE" and "INITIATE" (MP SumCrop).
- 472 This project also received funding from the European Union's Horizon 2020 research and innovation programme
- 473 under grant agreement SHOWCASE No 862480.

Conflicts of interest/Competing interests. The authors declare that they have no conflict of interest.

Availability of data and material. The datasets generated and used for the study are available (except confidential data such as landscape information) from the corresponding author on reasonable request.

Code availability. The code performed to analyse the dataset is available from the corresponding author on reasonable request.

Ethics approval. Not applicable

Consent to participate. Not applicable

Consent for publication. Not applicable

Authors' contributions. VB and SG conceived the study and were in charge of overall direction. VB and SG designed and planned the field work. JG worked in the field and in the laboratory for yield component quantification with the help of TP and MR. JG entered the data. MR analysed the data. SG, FC, OM, VB, and TP provided feedback and helped improve analyses. MR, SG, VB, FC, OM, and TP discussed and interpreted the results. MR, SG, VB, and FC wrote the paper. TP and OM commented and helped shape the manuscript. All authors approved the final version of the manuscript.

Table 1: Summary statistics of the models investigating the effects of 'Phytometer-Regional Flowering Temporal Matching' and floral resources in the fields (field size as a proxy) and 'Focal Field-Regional Flowering Temporal Matching' only for OSR fields) and in the surrounding landscapes (%OSR: % oilseed rape, %SNH: % semi-natural habitats and %OF: % organic farming) on fruiting success in the control treatment in OSR fields (a), cereal fields (b) and grasslands (c). We performed type II analyses of variance. The coefficients of determination of the three models (R²), the variance explained by each variable (%Var; sum of squares associated to a variable divided by the total sum of squares), the degrees of freedom (df), the F-value (F) and the p-value (p) are presented. Significant values are bold. Estimations of landscape effects and scale of effects were similar without local × landscape interactions (Appendix F).

	a. OSR					b. Cere	al		c. Grassland			
\mathbb{R}^2		44.3%		35.8%				54.6%				
	df	%Var	F	p	df	%Var	F	p	df	%Var	F	p
Year	4	15.2	7.13	< 0.001	3	18.1	6.34	< 0.001	2	18.4	7.08	< 0.01
Field size	1	0.0	0.07	0.79	1	0.3	0.27	0.60	1	0.5	0.35	0.56
Phytometer-Regional Flowering Temporal Matching	1	14.0	26.36	< 0.001	1	7.9	8.34	< 0.01	1	9.0	6.97	0.01
Focal Field-Regional Flowering Temporal Matching	2	1.1	1.00	0.37	-	-	-	-	-	-	-	-
%OSR (Landscape Flower Quantity)	1	2.9	5.45	0.02	1	4.6	4.82	0.03	1	5.6	4.31	0.05
%SNH (Landscape Flower Quantity)	1	3.0	5.65	0.02	1	0.4	0.46	0.50	1	19.1	14.73	< 0.001
%OF (Landscape Flower Quantity)	1	2.8	5.32	0.02	1	2.4	2.51	0.12	1	1.7	1.29	0.26
Field size x Phytometer-Regional Flow. Temp.	1	2.3	4.32	0.04	1	0.2	0.20	0.65	-	-	-	-
Matching												
Focal Field-Reg. Flow. Temp. Matching x Phytometer-	2	1.5	1.41	0.25	-	-	-	-	-	-	-	-
Reg. Flow. Temp. Matching												
Field size x Focal Field- Reg. Flow. Temp. Matching	2	0.8	0.71	0.50	-	-	-	-	-	-	-	-
%OSR x Phytometer-Reg. Flow. Temp. Matching	1	0.1	0.26	0.61	1	0.5	0.48	0.49	-	-	-	-
%SNH x Phytometer-Reg. Flow. Temp. Matching	1	0.7	1.23	0.27	1	1.0	1.07	0.30	-	-	-	-
%OF x Phytometer-Reg. Flow. Temp. Matching	1	0.0	0.02	0.89	1	0.4	0.39	0.53	-	-	-	-

Table 2: Effects of 'Phytometer-Regional Flowering Temporal Matching' and availability of floral resources in the fields and in the surrounding landscapes (%OSR: % oilseed rape, %SNH: % semi-natural habitats and %OF: % organic farming) on the magnitude of each pollination process in (a) OSR fields, (b) cereal fields and (c) grasslands. For each of 15 models, we applied type II analyses of variance. Coefficients of determination of the models (R²), degree of freedom (df), F-value (F) and p-value (p) for each explained variable are indicated. Significant values are in bold.

		Insects		Large insects		Small insects		Wind		Self	
	df	F	p	F	р	F	р	F	p	F	p
a. OSR		$R^2 = 32.0\%$		$R^2 = 22.3\%$		$R^2 = 16.7\%$		$R^2 = 28.4\%$		$R^2 = 47.8\%$	
Year	4	2.84	0.03	1.57	0.19	0.98	0.42	5.20	< 0.001	13.60	< 0.001
Field size	1	2.76	0.10	0.02	0.88	1.39	0.24	2.55	0.11	0.01	0.93
Phytometer-Regional Flowering Temporal Matching	1	4.89	0.03	0.49	0.49	1.12	0.29	0.17	0.68	12.96	< 0.001
Focal Field-Regional Flowering Temporal Matching	2	2.87	0.06	0.96	0.39	0.62	0.54	1.44	0.24	2.65	0.08
%OSR (Landscape Flower Quantity)	1	1.20	0.28	0.06	0.81	1.07	0.30	1.60	0.21	3.47	0.07
%SNH (Landscape Flower Quantity)	1	4.81	0.03	0.21	0.64	0.14	0.71	0.78	0.38	5.81	0.02
%OF (Landscape Flower Quantity)	1	6.65	0.01	0.85	0.36	3.71	0.06	1.50	0.22	0.36	0.55
Field size x Phytometer-Regional Flow. Temp.	1	0.69	0.41	1.44	0.23	0.03	0.86	1.53	0.22	7.58	0.01
Matching											
Focal Field-Reg. Flow. Temp. Matching x Phytometer-	2	3.35	0.04	0.16	0.85	0.82	0.44	1.20	0.31	0.53	0.59
Reg. Flow. Temp. Matching											
Field size x Focal Field- Reg. Flow. Temp. Matching	2	2.68	0.07	1.05	0.35	0.29	0.75	1.09	0.34	3.90	0.02
%OSR x Phytometer-Reg. Flow. Temp. Matching	1	3.02	0.09	1.87	0.17	3.10	0.08	0.41	0.52	1.18	0.28
%SNH x Phytometer-Reg. Flow. Temp. Matching	1	0.37	0.54	4.42	0.04	4.29	0.04	0.69	0.41	0.00	0.99
%OF x Phytometer-Reg. Flow. Temp. Matching	1	5.76	0.02	7.39	0.01	1.45	0.23	0.02	0.89	6.79	0.01
b. Cereal		$R^2 = 15.7\%$		$R^2 = 18.5\%$		$R^2 = 11.6\%$		$R^2 = 15.2\%$		$R^2 = 39.2\%$	
Year	3	1,24	0.30	0.51	0.68	0.20	0.90	0.61	0.61	10.13	< 0.00
Field size	1	0.04	0.84	0.01	0.91	0.07	0.79	2.77	0.10	0.95	0.33
Phytometer-Regional Flowering Temporal Matching	1	0.11	0.74	0.02	0.89	1.13	0.29	0.94	0.34	13.06	< 0.00
%OSR (Landscape Flower Quantity)	1	0.03	0.87	0.03	0.87	1.55	0.22	0.36	0.55	0.10	0.75
%SNH (Landscape Flower Quantity)	1	2.14	0.15	0.71	0.40	0.92	0.34	0.40	0.53	1.64	0.20
%OF (Landscape Flower Quantity)	1	0.17	0.68	2.69	0.11	1.95	0.17	0.20	0.65	1.10	0.30
Field size x Phytometer-Reg. Flow. Temp. Matching	1	0.09	0.77	1.09	0.30	0.16	0.69	0.30	0.58	0.02	0.88
%OSR x Phytometer-Reg. Flow. Temp. Matching	1	1.88	0.17	6.35	0.01	0.54	0.46	2.78	0.10	5.14	0.03
%SNH x Phytometer-Reg. Flow. Temp. Matching	1	1.47	0.23	1.54	0.22	0.73	0.39	1.37	0.25	2.80	0.10
%OF x Phytometer-Reg. Flow. Temp. Matching	1	1.64	0.20	0.11	0.74	0.22	0.64	3.29	0.07	3.68	0.06
c. Grassland		$R^2 = 57.4\%$		$R^2 = 56.5\%$		$R^2 = 51.4\%$		$R^2 = 42.7\%$		$R^2 = 38.4\%$	
Year	2	2.56	0.10	1.15	0.33	7.01	< 0.01	3.71	0.04	2.74	0.08
Field size	1	0.04	0.85	0.49	0.49	0.23	0.64	0.15	0.70	0.42	0.52
Phytometer-Regional Flowering Temporal Matching	1	25.17	< 0.001	26.54	< 0.001	3.89	0.06	10.13	< 0.01	0.40	0.53
%OSR (Landscape Flower Quantity)	1	1.72	0.20	10.34	< 0.01	1.64	0.21	2.76	0.11	9.35	< 0.01
%SNH (Landscape Flower Quantity)	1	10.38	< 0.01	1.48	0.23	6.35	0.02	7.90	0.01	1.33	0.26
%OF (Landscape Flower Quantity)	1	1.88	0.18	1.34	0.26	1.90	0.18	0.34	0.56	3.73	0.06

Fig. 1: (a) The experimental design was set up in the LTSER Zone Atelier Plaine & Val de Sèvre, a study area situated in central western France. Oilseed rape plant phytometers were placed in OSR fields (yellow), cereal fields (grey) and grasslands (blue) from 2015 to 2019. (b) In each farmer's field, two OSR phytometers (circles) were placed, one at the edge and one in the core of the field. Four branches were tested, one with a control treatment, one with a large mesh treatment, one with a small mesh treatment and one with an Osmolux treatment. In OSR, the flowering sequence goes from the base to the apex. Flowers opened during the experiment were delaminated from flowers opened before or after field experiment. Pollination processes involved in each experimental treatment include in the control treatment self-pollination, wind-pollination, small insect-pollination, and large insect-pollination. (c) Plant phytometers were deposited in OSR, cereals or grasslands, field size was used as a proxy of the amount of floral resources at the field scale. %OSR (oilseed rape), %SNH (semi-natural habitats) and %OF (organic farming) were used to estimate 'Landscape Flower Quantity'. Information about the amount and type of floral resources provided by the habitats (focal or landscape ones) were provided based on literature (Hegland and Boeke 2006; Hardman et al. 2016; Gaba et al. 2020; Bourgeois et al. 2020; Sidemo-Holm et al. 2021): +++: flower cover > 50%, ++: flower cover > 10%, +: flower cover > 0%, 0: flower cover = 0%. The amount of flowers provided by wild plants is rather constant (Cole et al. 2017). To account for the temporal variation of OSR flower cover we developed two variables: 'Phytometer-Regional Flowering Temporal Matching' described the period of OSR flowering during which phytometers were placed in the fields (during or after OSR flowering peak) and 'Focal Field-Regional Flowering Temporal Matching' defined the temporal deviation between flowering in the OSR focal field and OSR flowering in the study site and was thus designed only for OSR fields in contrast to the other three variables.

526527528

529

530531

532

533

534

535536

507

508

509

510

511

512513

514

515516

517

518519

520521

522

523

524

525

Fig. 2: Variation in OSR fruiting success with 'Phytometer–Regional Flowering Temporal Matching' in OSR (a), cereals (b) and grasslands (c). Thresholds discriminating 'Phytometer–Regional Flowering Temporal Matching' (i.e. during and after OSR flowering / flowering peak) were defined differently for phytometers in OSR fields and for phytometers placed in cereals and grasslands (see 'Methods' for details). Quantiles and means (black dots) of raw data are represented and numbers show the sample sizes. Significant differences between levels of 'Phytometer–Regional Flowering Temporal Matching' are indicated by asterisks (p-value: *** < 0.001; ** < 0.01 and * <0.05). Effect sizes: fruiting success was on average greater during OSR peak flowering than after in OSR (0.21 \pm 0.05 se; from Table 1a model), in cereals (0.15 \pm 0.05; from Table 1b model) and in grasslands (0.19 \pm 0.07; from Table 1c model).

537538

539

540541

542

543

544

545546

Fig. 3: Effect of 'Landscape Flower Quantity' on fruiting success of plant phytometers. The effects of %OSR (a, b, c), %SNH (d, e, f) and %OF (g, h, i) were quantified in buffer radii (r value in brackets) estimated by optimization of likelihood outside the OSR fields (a, d, g), cereal fields (b, e, h) and grasslands (c, f, i) in which the phytometers were placed. Dots of different shapes show raw data for each year. The lines show the adjusted relationships and the shaded area the standard error (based on models presented in Table 1). Lines in bold show significant effects (p-value<0.05) and dotted lines non-significant effects. Effect sizes: in OSR, fruiting success increased with %OSR (mean \pm se: 0.0030 \pm 0.0013 per %OSR) and %SNH (0.0126 \pm 0.0057 per %SNH) and decreased with %OF (-0.0034 \pm 0.0015 per %OF). Fruiting success increased with %OSR (0.0057 \pm 0.0026 per %OSR) in cereals and in grasslands (0.0134 \pm 0.0065 per %OSR), while it decreased with %SNH (-0.0305 \pm

0.0080 per %SNH) in grasslands.

Fig. 4: Fruiting success (in %) of self-pollination between crops (a, b) and effect sizes of other pollination processes in OSR fields (c), cereal fields (d) and grasslands (e) and the variation between categories of 'Phytometer-Regional Flowering Temporal Matching': during (orange) versus after (green) OSR flowering (peak). Significance of effect size are tested against 0, using unilateral sign tests (p-value: *** < 0.001; ** < 0.01 and * <0.05 into brackets). Asterisks (p-value: *** < 0.001; ** < 0.01 and * <0.05) also indicate significant differences between categories of 'Phytometer-Regional Flowering Temporal Matching' (see Table 2; c-e). Mean \pm se of raw data are presented, numbers indicate the sample sizes. Estimated effects: self-pollination was 0.11 ± 0.05 se more important during the OSR peak flowering than after in OSR fields, and 0.16 ± 0.04 (estimation from Table 2a model) greater during OSR flowering than after in CSR (from Table 2b model). Insect-pollination was 0.27 ± 0.18 and large insect-pollination (0.88 ± 0.17) were greater after OSR flowering than during, while wind-pollination was 0.73 ± 0.23 more important during OSR flowering than after in grasslands (from Table 2c model).

Fig. 5: Variation in the magnitude of pollination processes with 'Landscape Flower Quantity' (%OSR: % oilseed rape, %SNH: % semi-natural habitats and %OF: % organic farming) and 'Phytometer-Regional Flowering Temporal Matching' in OSR fields (a-f), cereals (g) and grasslands (h-j). The scale of effects was estimated by likelihood optimizations. Raw data is shown for each year, the predicted relationships are based on models presented in Table 2 (bold if a significant effect, or else dotted) and their standard errors. Black colour was used when landscape effect was independent of Phytometer-Regional Flowering Temporal Matching and green and orange colours were used for significant interaction terms.

References

- 570 Arnholt AT, Evans B (2017) BSDA: Basic Staffistics and Data Analysis. Version 1.2.0URL https://CRAN.R-project.org/package=BSDA
- Bänsch S, Tscharntke T, Gabriel D, Westphal C (2020) Crop pollination services: Complementary resource use by social vs solitary bees facing crops with contrasting flower supply. J Appl Ecol n/a: https://doi.org/10.1111/1365-2664.13777
- Bartomeus I, Potts SG, Steffan-Dewenter I, Vaissière B, Woyciechowski M, Krewenka K, Tscheulin T, Roberts S, Szentgyorgyi H, Westphal C, Bommarco R (2014) Contribution of insect pollinators to crop yield and quality varies with agricultural intensification. PeerJ 2:e328. https://doi.org/10.7717/peerj.328
- Bennett AB, Isaacs R (2014) Landscape composition influences pollinators and pollination services in perennial biofuel plantings. Agr Ecosyst Environ 193:1–8. https://doi.org/10.1016/j.agee.2014.04.016
- Bommarco R, Marini L, Vaissière BE (2012) Insect pollination enhances seed yield, quality, and market value in oilseed rape.
 Oecol 169:1025–1032. https://doi.org/10.1007/s00442-012-2271-6
- 581 Bourgeois B, Gaba S, Plumejeaud C, Bretagnolle V (2020) Weed diversity is driven by complex interplay between multi-scale dispersal and local filtering. Proc Biol Sci 287:20201118. https://doi.org/10.1098/rspb.2020.1118 583
- Bretagnolle V, Berthet E, Gross N, Gauffre B, Plumejeaud C, Houte S, Badenhausser I, Monceau K, Allier F, Monestiez P,
 Gaba S (2018) Towards sustainable and multifunctional agriculture in farmland landscapes: Lessons from the
 integrative approach of a French LTSER platform. Sci Total Environ 627:822–834.
 https://doi.org/10.1016/j.scitotenv.2018.01.142
- 588 Bretagnolle V, Gaba S (2015) Weeds for bees? A review. Agron Sustain Dev 35:891–909. https://doi.org/10.1007/s13593-015-589 0302-5
- Bukovinszky T, Verheijen J, Zwerver S, Klopa E, Biesmeijer J, Wäckers F, Prins H, Kleijn D (2017) Exploring the relationships
 between landscape complexity, wild bee species richness and reproduction, and pollination services along a
 complexity gradient in the Netherlands. Biol. Conserv 214:312–319. https://doi.org/10.1016/j.biocon.2017.08.027
- Carpentier, F., Martin, O., 2021. Siland a R package for estimating the spatial influence of landscape. Sci Rep 11, 7488. https://doi.org/10.1038/s41598-021-86900-0
- Carpentier F, Martin O (2020) siland: Spatial Influence of Landscape. Version 2.0URL https://CRAN.R-project.org/package=siland
- Castle D, Grass I, Westphal C (2019) Fruit quantity and quality of strawberries benefit from enhanced pollinator abundance at hedgerows in agricultural landscapes. Agr Ecosyst Environ 275:14–22. https://doi.org/10.1016/j.agee.2019.01.003
- Cavallero L, Morales CL, Montero-Castaño A, Gowda JH, Aizen MA (2018) Scale-dependent effects of conspecific flower availability on pollination quantity and quality in an invasive shrub. Oecol 188:501–513. https://doi.org/10.1007/s00442-018-4239-7
- 602 Chifflet R, Klein EK, Lavigne C, Le Féon V, Ricrich AE, Lecomte J, Vaissière BE (2011) Spatial scale of insect-mediated 603 pollen dispersal in oilseed rape in an open agricultural landscape. J Appl Ecol 48:689–696. 604 https://doi.org/10.1111/j.1365-2664.2010.01904.x
- Cole LJ, Brocklehurst S, Robertson D, Harrison W, McCracken D (2017) Exploring the interactions between resource availability and the utilisation of semi-natural habitats by insect pollinators in an intensive agricultural landscape.

 Agr Ecosyst Environ 246:157–167. https://doi.org/10.1016/j.agee.2017.05.007
- Cusser S, Neff JL, Jha S (2016) Natural land cover drives pollinator abundance and richness, leading to reductions in pollen limitation in cotton agroecosystems. Agr Ecosyst Environ 226:33–42. https://doi.org/10.1016/j.agee.2016.04.020
- Dainese M, Martin EA, Aizen MA, et al (2019) A global synthesis reveals biodiversity-mediated benefits for crop production. Sci Adv 5:eaax0121. https://doi.org/10.1126/sciadv.aax0121
- Diekötter T, Peter F, Jauker B, Wolters V, Jauker F (2014) Mass-flowering crops increase richness of cavity-nesting bees and wasps in modern agro-ecosystems. Glob Change Biol Bioenergy 6:219–226. https://doi.org/10.1111/gcbb.12080
- Dormann CF, McPherson JM, Araújo MB, et al (2007) Methods to account for spatial autocorrelation in the analysis of species

615 distributional data: a review. Ecography 30:609-628. https://doi.org/10.1111/j.2007.0906-7590.05171.x 616 Elzinga JA, Atlan A, Biere A, Gigord L, Weis A, Bernasconi G (2007) Time after time: flowering phenology and biotic 617 interactions. Trends in Ecol Evol 22:432–439. https://doi.org/10.1016/j.tree.2007.05.006 618 Fahrig L, Baudry J, Brotons L, Burel F, Crist T, Fuller R, Sirami C, Siriwardena C, Martin JL (2011) Functional landscape 619 heterogeneity and animal biodiversity in agricultural landscapes: Heterogeneity and biodiversity. Ecol Lett 14:101– 620 112. https://doi.org/10.1111/j.1461-0248.2010.01559.x 621 Fahrig L, Girard J, Duro D, Pasher J, Smith A, Javorek S, King D, Lindsay KF, Mitchell S, Tischendorf L (2015) Farmlands 622 with smaller crop fields have higher within-field biodiversity. Agr Ecosyst Environ 200:219-234. 623 https://doi.org/10.1016/j.agee.2014.11.018 624 Fijen TPM, Scheper JA, Boekelo B, Raemakers I, Kleteijn D (2019) Effects of landscape complexity on pollinators are 625 moderated by pollinators' association with mass-flowering crops. Proc Biol Sci 286:20190387. https://doi.org/10.1098/rspb.2019.0387 626 627 Fox J, Weisberg S, Price B et al (2020) car: Companion to Applied Regression. Version 3.0-8URL https://CRAN.R-628 project.org/package=car 629 Frankl R, Wanning S, Braun R (2005) Quantitative floral phenology at the landscape scale: Is a comparative spatio-temporal 630 description of "flowering landscapes" possible? Nat 631 https://doi.org/10.1016/j.jnc.2004.10.002 632 Gaba S, Cheviron N, Perrot T, Piutti S, Gautier JL, Bretagnolle V (2020) Weeds Enhance Multifunctionality in Arable Lands 633 in South-West of France. Front Sustain Food Syst 4: https://doi.org/10.3389/fsufs.2020.00071 634 Garratt MPD, Brown R, Hartfield C, Hart A, Potts SG (2018) Integrated crop pollination to buffer spatial and temporal 635 variability in pollinator activity. Basic Appl Ecol 32:77-85. https://doi.org/10.1016/j.baae.2018.06.005 636 Grab H, Blitzer EJ, Danforth B, Loeb G, Poveda K (2017) Temporally dependent pollinator competition and facilitation with 637 mass flowering crops affects yield in co-blooming crops. Sci Rep 7:45296. https://doi.org/10.1038/srep45296 638 Greenleaf SS, Williams NM, Winfree R, Kremen C (2007) Bee foraging ranges and their relationship to body size. Oecol 639 153:589-596. https://doi.org/10.1007/s00442-007-0752-9 640 Hardman CJ, Norris K, Nevard TD, Hughes B, Potts S (2016) Delivery of floral resources and pollination services on farmland 641 three different wildlife-friendly schemes. 220:142-151. Agr **Ecosyst** Environ 642 https://doi.org/10.1016/j.agee.2016.01.015 643 Hass AL, Kormann UG, Tscharntke T, et al (2018) Landscape configurational heterogeneity by small-scale agriculture, not 644 crop diversity, maintains pollinators and plant reproduction in western Europe. Proc R Soc B 285:20172242. 645 https://doi.org/10.1098/rspb.2017.2242 646 Hegland SJ, Boeke L (2006) Relationships between the density and diversity of floral resources and flower visitor activity in a temperate grassland community. Ecol Entomol 31:532-538. https://doi.org/10.1111/j.1365-2311.2006.00812.x 647 648 Henckel L, Börger L, Meiss H, Gaba S, Bretagnolle V (2015) Organic fields sustain weed metacommunity dynamics in 649 farmland landscapes. Proc Biol Sci 282:20150002. https://doi.org/10.1098/rspb.2015.0002 650 Herbertsson L, Rundlöf M, Smith HG (2017) The relation between oilseed rape and pollination of later flowering plants varies 651 across plant species and landscape contexts. Basic Appl Ecol 24:77-85. https://doi.org/10.1016/j.baae.2017.08.001 652 Holzschuh A, Dainese M, González-Varo JP, Mudri-Stojnic S, Riedinger V, Rundlof M, Scheper J, Wickens J, Wickens V, 653 Bommarco R, Kleijn D, Potts S, Roberts S, Smith H, Vilà M, Vujic A, Steffan-Dewenter I (2016) Mass-flowering 654 crops dilute pollinator abundance in agricultural landscapes across Europe. Ecol Lett 19:1228-1236. 655 https://doi.org/10.1111/ele.12657 656 Holzschuh A, Dormann CF, Tscharntke T, Steffan-Dewenter I (2011) Expansion of mass-flowering crops leads to transient 657 pollinator dilution and reduced wild plant pollination. Proc Biol Sci 278:3444-3451. 658 https://doi.org/10.1098/rspb.2011.0268 659 Hopfenmüller S, Steffan-Dewenter I, Holzschuh A (2014) Trait-Specific Responses of Wild Bee Communities to Landscape

Local

and

Factors.

PLoS

Configuration

Composition,

660

9:e104439.

661 https://doi.org/10.1371/journal.pone.0104439 662 Hudewenz A, Pufal G, Bögeholz A-L, Klein A-M (2014) Cross-pollination benefits differ among oilseed rape varieties. J Agric Sci 152:770-778. https://doi.org/10.1017/S0021859613000440 663 664 HB, Fahrig L (2012) What size is a biologically relevant landscape? Landsc Ecol 27:929-941. 665 https://doi.org/10.1007/s10980-012-9757-9 Jauker F, Bondarenko B, Becker HC, Steffan-Dewenter I (2012) Pollination efficiency of wild bees and hoverflies provided to 666 667 oilseed rape. Agri Fo Entomol 14:81–87. https://doi.org/10.1111/j.1461-9563.2011.00541.x 668 Klein A-M, Vaissière BE, Cane JH, Steffan-Dewenter I, Cunningham S, Kremen C, Tscharntke T (2007) Importance of 669 274:303-313. pollinators in changing landscapes for world crops. Proc Biol Sci 670 https://doi.org/10.1098/rspb.2006.3721 671 Knapp EE, Goedde MA, Rice KJ (2001) Pollen-limited reproduction in blue oak: implications for wind pollination in 672 fragmented populations. Oecol 128:48-55. https://doi.org/10.1007/s004420000623 673 Kovács-Hostvánszki A, Haenke S, Batáry P, Jauker B, Báldi A, Tscharntke T, Holzschuh A (2013) Contrasting effects of mass-674 flowering crops on bee pollination of hedge plants at different spatial and temporal scales. Ecol Appl 23:1938–1946. 675 https://doi.org/10.1890/12-2012.1 676 Kozak M, Piepho H-P (2018) What's normal anyway? Residual plots are more telling than significance tests when checking 677 ANOVA assumptions. J Agron Crop Sci 204:86–98. https://doi.org/10.1111/jac.12220 678 Kremen C, Williams NM, Aizen MA, Barbara Gemmill-Herren, LeBuhn G, Minckley R, Packer L, Potts S, Roulston T, 679 Steffan-Dewenter I, Vázquez D, Winfree R, Adams L, Crone E, Greenleaf S, Keitt T, Klein AM, Regetz J, Ricketts 680 T (2007) Pollination and other ecosystem services produced by mobile organisms: a conceptual framework for the 681 effects of land-use change. Ecol Lett 10:299-314. https://doi.org/10.1111/j.1461-0248.2007.01018.x 682 Le Féon V, Burel F, Chifflet R, Henry M, Ricroch A, Vaissière B, Baudry J (2013) Solitary bee abundance and species richness in dynamic agricultural landscapes. Agr Ecosyst Environ 166:94–101. https://doi.org/10.1016/j.agee.2011.06.020 683 Lenth R, Singmann H, Love J, Buerkner P, Herve M (2020) emmeans: Estimated Marginal Means, aka Least-Squares Means. 684 685 Version 1.4.6 URL https://CRAN.R-project.org/package=emmeans 686 Lindström SAM, Herbertsson L, Rundlöf M, Bommarco R, Smith H (2016) Experimental evidence that honeybees depress 687 wild insect densities in a flowering crop. Proc Biol Sci 283:20161641. https://doi.org/10.1098/rspb.2016.1641 688 Mallinger RE, Gibbs J, Gratton C (2016) Diverse landscapes have a higher abundance and species richness of spring wild bees 689 by providing complementary floral resources over bees' foraging periods. Landsc Ecol 31:1523-1535. 690 https://doi.org/10.1007/s10980-015-0332-z 691 Mandelik Y, Winfree R, Neeson T, Kremen C (2012) Complementary habitat use by wild bees in agro-natural landscapes. Ecol 692 Appl 22:1535–1546. https://doi.org/10.1890/11-1299.1 693 Marini L, Tamburini G, Petrucco-Toffolo E, Lindström SAM (2015) Crop management modifies the benefits of insect 694 pollination in oilseed rape. Agr Ecosyst Environ 207:61-66. https://doi.org/10.1016/j.agee.2015.03.027 695 Martin EA, Dainese M, Clough Y et al (2019) The interplay of landscape composition and configuration: new pathways to 696 manage functional biodiversity and agroecosystem services across Europe. Ecol Lett 22:1083-1094. 697 https://doi.org/10.1111/ele.13265 698 McCartney HA, Lacey ME (1991) Wind dispersal of pollen from crops of oilseed rape (Brassica napus L.). J Aerosol Sci 22: 699 467–477. https://doi.org/10.1016/0021-8502(91)90005-3 700 Mesquida J, Renard M (1981) Pollinisation du colza d'hiver mâle-fertile et mâle-stérile (Brassica napus L var. oleifera Metzger) 701 par l'abeille domestique (Apis M. mellifica L.). Effets sur la phénologie et le rendement. Apidologie 12:345-362. 702 https://doi.org/10.1051/apido:19810404 703 Meyer B, Gaebele V, Steffan-Dewenter I (2007) Patch Size and Landscape Effects on Pollinators and Seed Set of the Horseshoe Vetch, Hippocrepis comosa, in an Agricultural Landscape of Central Europe. Entomol Gen 173-185. 704 https://doi.org/10.1127/entom.gen/30/2007/173 705

- Miguet P, Jackson HB, Jackson ND, Martin AE, Fahrig L (2016) What determines the spatial extent of landscape effects on species? Landsc Ecol 31:1177–1194. https://doi.org/10.1007/s10980-015-0314-1
- Momoh EJJ, Zhou W (2001) Growth and Yield Responses to Plant Density and Stage of Transplanting in Winter Oilseed Rape (Brassica napus L.). J Agron Crop Sci 186:253–259. https://doi.org/10.1046/j.1439-037x.2001.00476.x
- 710 Ollerton J, Winfree R, Tarrant S (2011) How many flowering plants are pollinated by animals? Oikos 120:321–326. 711 https://doi.org/10.1111/j.1600-0706.2010.18644.x
- Palmer TM, Zimmerman M (1994) Pollen Competition and Sporophyte Fitness in Brassica campestris: Does Intense Pollen Competition Result in Individuals with Better Pollen? Oikos 69:80–86. https://doi.org/10.2307/3545286
- Perrot T, Gaba S, Roncoroni M, Gautier JL, Bretagnolle V (2018) Bees increase oilseed rape yield under real field conditions.

 Agr Ecosyst Environ 266:39–48. https://doi.org/10.1016/j.agee.2018.07.020
- Petersen JD, Nault BA (2014) Landscape diversity moderates the effects of bee visitation frequency to flowers on crop production. J Appl Ecol 51:1347–1356. https://doi.org/10.1111/1365-2664.12287
- Popławska W, Liersch A, Jędryczka M, Wielebski F, Bartkowiak-Broda I (2013) Wind-mediated pollen dispersal of oilseed rape an estimation using pollen traps. Geography
- R Core team (2019) A language and environment for statistical computing. Version 3.6.2. R Foundation for Statistical Computing, Vienna, Austria
- Rader R, Howlett BG, Cunningham SA, Westcott DA, Newstrom-Lloyd LE, Walker MK, Teulon D, Edwards W (2009)
 Alternative pollinator taxa are equally efficient but not as effective as the honeybee in a mass flowering crop. J Appl
 Ecol 46:1080–1087. https://doi.org/10.1111/j.1365-2664.2009.01700.x
- 725 Raderschall CA, Bommarco R, Lindström SAM, Lundin O (2021) Landscape crop diversity and semi-natural habitat affect
 726 crop pollinators, pollination benefit and yield. Agr Ecosyst Environ 306:107189.
 727 https://doi.org/10.1016/j.agee.2020.107189
- Rathcke B, Lacey EP (1985) Phenological Patterns of Terrestrial Plants. Annu. Rev Ecol Evol Syst 16:179–214. https://doi.org/10.1146/annurev.es.16.110185.001143
- Requier F, Odoux J-F, Henry M, Bretagnolle V (2017) The carry-over effects of pollen shortage decrease the survival of honeybee colonies in farmlands. J Appl Ecol 54:1161–1170. https://doi.org/10.1111/1365-2664.12836
- Requier F, Odoux J-F, Tamic T, Nathalie Moreau, Henry M, Decourtye A, Bretagnolle V (2015) Honey bee diet in intensive farmland habitats reveals an unexpectedly high flower richness and a major role of weeds. Ecol Appl 25:881–890. https://doi.org/10.1890/14-1011.1
- Ricketts TH, Regetz J, Steffan-Dewenter I, Cunningham S, Kremen C, Bogdanski A, Gemmill-Herren B, Greenleaf S, Klein AM, Mayfield M, Morandin L, Ochieng L, Viana B (2008) Landscape effects on crop pollination services: are there general patterns? Ecol Lett 11:499–515. https://doi.org/10.1111/j.1461-0248.2008.01157.x
- Rollin O, Bretagnolle V, Decourtye A, Aptel J, Michele N, Vaissière B, Henry M (2013) Differences of floral resource use between honey bees and wild bees in an intensive farming system. Agr Ecosyst Environ 179:78–86. https://doi.org/10.1016/j.agee.2013.07.007
- Rollin O, Bretagnolle V, Fortel L, Guilbaud L, Henry M (2015) Habitat, spatial and temporal drivers of diversity patterns in a wild bee assemblage. Biodivers Conserv 24:1195–1214. https://doi.org/10.1007/s10531-014-0852-x
- Roulston T, Goodell K (2011) The Role of Resources and Risks in Regulating Wild Bee Populations. Annu Rev Entomol 56:293–312. https://doi.org/10.1146/annurev-ento-120709-144802
- Rusch A, Valantin-Morison M, Sarthou J-P, Roger-Estrade J (2011) Multi-scale effects of landscape complexity and crop management on pollen beetle parasitism rate. Landsc Ecol 26:473–486. https://doi.org/10.1007/s10980-011-9573-7
- Schellhorn NA, Gagic V, Bommarco R (2015) Time will tell: resource continuity bolsters ecosystem services. Trends Ecol Evol 30:524–530. https://doi.org/10.1016/j.tree.2015.06.007
- Shaw RF, Phillips BB, Doyle T, Pell J, Redhead J, Savage J, Woodcock B, Bullock J, Osborne J (2020) Mass-flowering crops have a greater impact than semi-natural habitat on crop pollinators and pollen deposition. Landsc Ecol 35:513–527.

751	https://doi.org/10.1007/s10980-019-00962-0
752 753 754	Sidemo-Holm W, Carrié R, Ekroos J, Lindström SAM, Smith HG (2021) Reduced crop density increases floral resources to pollinators without affecting crop yield in organic and conventional fields. J Appl Ecol 1365-2664.13887. https://doi.org/10.1111/1365-2664.13887
755 756	Steffan-Dewenter I, Münzenberg U, Bürger C, Thies C, Tscharntke T (2002) Scale-Dependent Effects of Landscape Context on Three Pollinator Guilds. Ecol 83:1421–1432. https://doi.org/10.1890/0012-9658(2002)083
757 758	Tscharntke T, Tylianakis JM, Rand TA, et al (2012) Landscape moderation of biodiversity patterns and processes - eight hypotheses. Biol Rev 87:661–685. https://doi.org/10.1111/j.1469-185X.2011.00216.x
759 760 761	Vasseur C, Joannon A, Aviron S, Burel F, Meynard JM, Baudry J (2013) The cropping systems mosaic: How does the hidden heterogeneity of agricultural landscapes drive arthropod populations? Agr Ecosyst Environ 166:3–14. https://doi.org/10.1016/j.agee.2012.08.013
762 763	Waser NM, Price MV (2016) Drought, pollen and nectar availability, and pollination success. Ecology 97:1400–1409. https://doi.org/10.1890/15-1423.1
764 765	Wintermantel D, Odoux J, Chadœuf J, Bretagnolle V (2019) Organic farming positively affects honeybee colonies in a flower-poor period in agricultural landscapes. J Appl Ecol 1365-2664.13447. https://doi.org/10.1111/1365-2664.13447
766 767 768	Woodcock TS, Pekkola LJ, Dawson C, Gadallah F, Kevan P (2014) Development of a Pollination Service Measurement (PSM) method using potted plant phytometry. Environ Monit Assess 186:5041–5057. https://doi.org/10.1007/s10661-014-3758-x
769 770	Zou Y, Xiao H, Bianchi FJJA, Jauker F, Luo S, Van der Werf W (2017) Wild pollinators enhance oilseed rape yield in small-holder farming systems in China. BMC Ecol 17:6. https://doi.org/10.1186/s12898-017-0116-1