

# Metabolic syndrome and pesticides: A systematic review and meta-analysis

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## Metabolic syndrome and pesticides: a systematic review and meta-analysis. 1 Hugo Lamat<sup>1</sup>, Marie-Pierre Sauvant-Rochat<sup>3</sup>, Igor Tauveron<sup>2</sup>, Reza Bagheri<sup>4</sup>, Ukadike C 2 Ugbolue<sup>5</sup>, Salwan Maqdasi<sup>2</sup>, Valentin Navel<sup>6</sup>, Frédéric Dutheil<sup>7</sup> 3 4 5 <sup>1</sup> Université Clermont Auvergne, University Hospital of Clermont–Ferrand, CHU Clermont– Ferrand, Endocrinology-diabetology-nutrition, 63000 Clermont-Ferrand, France 6 <sup>2</sup> Université Clermont Auvergne, CNRS, GReD, Inserm, University Hospital of Clermont-7 Ferrand, CHU Clermont-Ferrand, Endocrinology, Clermont-Ferrand, France 8 9 <sup>3</sup> Université Clermont Auvergne, Clermont Auvergne INP, CNRS, Institut Pascal, F-63000 10 Clermont-Ferrand, France 11 <sup>4</sup> University of Isfahan, Exercise Physiology, Isfahan, Iran 12 <sup>5</sup> University of the West of Scotland, Health and Life Sciences, South Lanarkshire, Scotland, UK 13 <sup>6</sup> Université Clermont Auvergne, CNRS, INSERM, GReD, CHU Clermont-Ferrand, 14 15 University Hospital of Clermont-Ferrand, Ophthalmology, Clermont-Ferrand, France <sup>7</sup> Université Clermont Auvergne, CNRS, LaPSCo, Physiological and Psychosocial Stress, 16 University Hospital of Clermont-Ferrand, CHU Clermont-Ferrand, Occupational and 17 18 Environmental Medicine, WittyFit, F-63000 Clermont-Ferrand, France 19 20 \*Corresponding author: Valentin Navel, University Hospital of Clermont-Ferrand, CHU Clermont-Ferrand, Ophthalmology, Tel: +33 6 82 91 34 36, Fax: +33 4 73 75 14 61, 21 valentin.navel@hotmail.fr 22 23 **Competing Interests**: The authors have declared that no competing interests exist. **Running** title: pesticides 24 Metabolic syndrome and

## Abstract

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The relation between pesticides exposure and metabolic syndrome (MetS) has not been clearly identified. Performing a systematic review and meta-analysis, PubMed, Cochrane Library, Embase, and ScienceDirect were searched for studies reporting the risk of MetS following pesticides exposure and their contaminants. We included 12 studies for a total of 6789 participants, in which 1981 (29.1%) had a MetS. Overall exposure to pesticides and their contaminants increased the risk of MetS by 30% (95CI 22% to 37%). Overall organochlorine increased the risk of MetS by 23% (14 to 32%), as well as for most types of organochlorines: hexachlorocyclohexane increased the risk by 53% (28 to 78%), hexachlorobenzene by 40% (0.01 to 80%), dichlorodiphenyldichloroethylene by 22% (9 to 34%), dichlorodiphenyltrichloroethane by 28% (5 to 50%), oxychlordane by 24% (1 to 47%), and transnonchlor by 35% (19 to 52%). Sensitivity analyses confirmed that overall exposure to pesticides and their contaminants increased the risk by 46% (35 to 56%) using crude data or by 19% (10 to 29%) using fully-adjusted model. The risk for overall pesticides and types of pesticides was also significant with crude data but only for hexachlorocyclohexane (36% risk increase, 17 to 55%) and transnonchlor (25% risk increase, 3 to 48%) with fully-adjusted models. Metaregressions demonstrated that hexachlorocyclohexane increased the risk of MetS in comparison to most other pesticides. The risk increased for more recent periods (Coefficient = 0.28, 95CI 0.20 to 0.37, by year). We demonstrated an inverse relationship with body mass index and male gender. In conclusion, pesticides exposure is a major risk factor for MetS. Besides organochlorine exposure, data are lacking for other types of pesticides. The risk increased with time, reflecting a probable increase of the use of pesticides worldwide. The inverse relationship with body mass index may signify a stockage of pesticides and contaminants in fat tissue.

- **Keywords**: metabolic syndrome, pesticide, exposure, public health, systemic review, meta-
- analysis, meta-regression

# 51 Introduction

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In 2020, the global production of pesticide reached 3.5 million tons. Considering the damaging consequences on human health,2 many international measures have been taken to limit the production and use of pesticides,<sup>3</sup> but even those who have been prohibited still exist in the environment leaving us with important exposure.<sup>4,5</sup> Among the wide range of diseases enhanced by pesticides, they are an evident risk factor for cardiovascular diseases, but also for metabolic diseases such type 2 diabetes mellitus (T2DM)<sup>8</sup> and obesity.<sup>9</sup> Metabolic syndrome (MetS) is considered as a plurimetabolic condition, linked with obesity and T2DM pathophysiology. 10 However, if the level of proof is high regarding the increased risk of diabetes and obesity following pesticides exposure, few studies have focused on the relationship between exposure to pesticides and the risk of having MetS. Among hundreds of pesticides, the most used are organochlorine pesticides (OCP), organophosphates, carbamates, pyrethoids and triazines. Moreover, pesticides exposure is often accompanied by contaminants such as polychlorodibenzop-dioxins and polychlorodibenzofuranes (PCCD/PCDF) and polychlorobiphenyls (PCB), 11 with each of them having specific effects on human health.<sup>6</sup> Furthermore, simultaneous exposure to several pesticides and contaminants seems to increase the risk for multiple metabolic parameters such as T2DM, dyslipidemia and insulin resistance. 12,13 Such effects are yet to be demonstrated in MetS. Considering that pesticides production are increasing to meet worldwide food demand, 14,15 the risk of MetS following pesticides exposure may have also increased over the last years. From a physiological point of view, most pesticides are known to be lipophile compounds that accumulate preferably in adipose tissue.<sup>16</sup> Considering that MetS is closely related to visceral adiposity, the influence of adiposity can be a very important factor in effects of pesticides exposure on MetS.<sup>17</sup> Similarly, sociodemographic characteristics like age, sex and race have un evident impact on the risk of MetS, <sup>18,19</sup> as previously demonstrated for other pathologies like T2DM.<sup>20</sup>

Therefore, we aimed to conduct a systematic review and meta-analysis based on the risk of MetS following pesticides exposures. Secondary aims were to compare the risk depending on the type of pesticides, the effects of multiple exposures to pesticides and their contaminants and the effect of time of exposure and to assess the influence of characteristics of individuals (anthropometric, sociodemographic, and lifestyle).

# **Methods**

## Literature search

We reviewed all studies reporting the risk of having MetS in relation with pesticides exposure. The PubMed, Cochrane Library, ScienceDirect and Embase databases were searched until October 24<sup>th</sup> 2021, with the following keywords: "metabolic syndrome" and "pesticide" (details of the search strategy is available in Appendix 1). The search was not limited to specific years and no language restrictions were applied. To be included in the systematic review, articles needed to describe our primary outcome variables i.e. the risk of MetS following exposure to pesticides or reporting numbers to calculate this risk. Then, articles describing either an odds ratio, a relative risk, or a hazard ratio, or giving data to calculate this risk, were included in the meta-analysis. In addition, reference lists from all publications meeting the inclusion criteria were manually searched to identify any further studies that were not found with the electronic search. The search strategy is presented in (Figure 1). Two authors (H Lamat and S Maqdasy) conducted all literature searches, collated, and reviewed the abstracts and based on the selection criteria, decided the suitability of the articles for inclusion. A third author (F Dutheil) was asked to review the articles where consensus on suitability was debated. All authors then reviewed the eligible articles. We followed the guidelines outlined by PRISMA.

## **Data collection**

The data collected included first author's name, publication year, study design, periods of studies, aims, outcomes of included articles, sample size, mean age, sex (percentage of males), risk of MetS (depending on method of calculation or adjustment model such as crude or adjusted odd ratio, and types of risk such as odds ratio, relative risks, and prevalence ratio), characteristics of pesticide and contaminants exposure (types, multi-exposure, years of exposure), and characteristics of individuals (such as body mass index (BMI), age, sex, geographical zone, smoking habits, alcohol consumption).

# Risk of bias and certainty of evidence

The Newcastle-Ottawa Scale (NOS) was used to check the risk of bias of included articles.<sup>21</sup> This scale has an original version to evaluate case-control and cohort studies. Adapted version exists and widely used to evaluate cross-sectional studies.<sup>22–24</sup> Through different items, three dimensions are evaluated: selection, comparability, and exposure for case-control studies; and selection, comparability and outcome for cohort and cross-sectional studies. One point was given for each item. The maximum score is 9 for case-control and cohort studies and 10 for cross-sectional studies (Figure 2). The evidence synthesis for this review was conducted in accordance with the Grading of Recommendations Assessment, Development and Evaluation (GRADE) guidelines (Appendix 2).<sup>25</sup>

## **Statistical considerations**

Statistical analysis was conducted using Stata software (v15, StataCorp, College Station, US). <sup>26–33</sup>

Baseline characteristics were summarized for each study sample and reported as mean (standard-deviation) and number (%) for continuous and categorical variables respectively. Random effects

meta-analyses (DerSimonian and Laird approach) were conducted when data could be pooled.<sup>34</sup> We conducted meta-analysis on the risk of MetS following pesticides exposure using all odds ratio, relative risk, or hazard ratio. We stratified these meta-analyses on the type of pesticides (organochlorines and types of organochlorines, organophosphates, carbamates, pyrethroids, triazines) and contaminants (OCP, PCB and PCCD/PCDF). We described our results by calculating the hazard ratio of the risk of MetS after exposure to pesticides.<sup>34</sup> Hazard ratio were centered at one if the risk of MetS after exposure to pesticides did not differ from the risk of MetS without exposure to pesticides. Hazard ratio >1 denoted an increased risk of MetS, and hazard ratio <1 reflected a decreased risk. Heterogeneity in the study results was evaluated by examining forest plots, confidence intervals (CI) and using formal tests for homogeneity based on the Isquared (I<sup>2</sup>) statistic. I<sup>2</sup> is easily interpretable and the most common metric to measure the magnitude of between-study heterogeneity. I2 values range between 0% and 100% and are typically considered low for <25%, modest for 25-50%, and high for >50%. For rigor, funnel plots of these meta-analyses were also used to search for potential publication bias. In order to verify the strength of the results, further meta-analyses were then conducted after exclusion of outliers i.e. by excluding studies that were not evenly distributed around the base of the funnel,<sup>35</sup> as well as sensitivity analyses based on the type of risks (odds ratio, relative risk, or hazard ratio). Moreover, when a study reported simultaneously both crude and adjusted risks, we also computed first sensitivity analyses using only the most adjusted risks listed in included articles, and in a second time, only crude or less adjusted risks. Therefore, for each meta-analysis, we computed two models (a model using the most adjusted risks listed in included articles, and a model using only crude or less adjusted risks). Lastly, because of a possible time-dependent cumulative effect for the risk of MetS, we computed a separated subgroup sensitivity analyses by removing the studies that included individuals over 70 years old. When possible (sufficient sample size), metaregressions were proposed to study the relationship between the risk of MetS and pesticides

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exposure and clinically relevant parameters such as characteristics of exposure (type of pesticides, multi-exposure, years of exposure) or those of individuals (BMI, age, sex, lifestyle). Results were expressed as regression coefficients and 95% CI. P values less than 0.05 were considered statistically significant.

# **Results**

An initial search produced a possible 8595 articles (Figure 1). Removal of duplicates and use of the selection criteria reduced the number of articles reporting MetS as an outcome of pesticides exposure to 12 articles in the systematic review.<sup>36–47</sup> One study reported only  $\beta$  coefficients<sup>38</sup> and was not included in the meta-analyses.<sup>48</sup> Thus all 12 studies were included in the systematic review and only 11 studies were included in the meta-analyses. All articles were written in English.

# Risk of bias and certainty of evidence

The assessment of the quality of the 12 included studies was performed using the Newcastle-Ottawa quality scale to assess the risk of bias and GRADE to assess the certainty of evidence. All studies had a low risk of bias. Overall, case-control studies performed better in the comparability section and worse in the exposure section, while cohort and cross-sectional studies performed better in comparability section and worse in selection section (Figure 2). All studies mentioned ethical approval. The certainty of the evidence was rated as high for our main outcome and for most other meta-analytic outcomes. The certainty of the evidence was low for mirex, aldrin, PCDD, and PCDF (Appendix 2).

## Aims and outcomes of included articles

Eight studies aimed to evaluate the association between pesticides exposure and the incidence or the prevalence of MetS as a primary result<sup>36–39,42–44,47</sup> while four studies evaluated this association as a secondary result.<sup>40,41,45,46</sup> Ten studies evaluated only pesticides. The two other studies assessed the impact of other environmental pollutants.<sup>40,41</sup>

# **Study designs**

Our analysis was performed from six cross-sectional studies, three nested case-control studies, two cohort studies and one case-control study. Duration of recruitment in cross-sectional studies ranged from 6 months<sup>42</sup> to 5 years,<sup>38</sup> and duration of follow up in prospective studies ranged from 2.6 years<sup>44</sup> to 10 years.<sup>40</sup> Five studies were conducted in Asia (South Korea,<sup>39,42,44,45</sup> India<sup>36</sup>), four in Europe (Sweden,<sup>40,41</sup> Spain,<sup>46</sup> Norway<sup>47</sup>), and two in USA.<sup>37,43</sup> Seven studies were monocentric<sup>36,37,39–42,47</sup> and five multicentric.<sup>38,43–46</sup>

## Inclusions and exclusions criteria of included articles

Diagnosis of MetS was based on NCEP-ATPIII definition in all studies except one that used both NCEP-ATPIII and IDF definitions.<sup>46</sup> Some of the included articles targeted at the population at advanced age<sup>40,41</sup> or being overweight or obese,<sup>46,47</sup> and therefore predisposed to risk of MetS. Two studies included MetS patients.<sup>36,42</sup> Inclusion criteria for other articles were normal weight,<sup>45</sup> non-diabetic,<sup>37,39,43</sup> non-MetS at baseline,<sup>40,44</sup> or living in the same area.<sup>38</sup> All studies excluded participants with incomplete data on MetS or pesticides exposure. Other main exclusion criteria were diabetes mellitus,<sup>37,39,43,47</sup> underweight,<sup>46</sup> occupational exposure.<sup>43</sup>

## Recruitment

Six studies recruited participants from other ongoing surveys<sup>37,38,43–46</sup> whereas six studies specifically recruited participants for their studies. 14,19,35,45,49,50 Recruitment was from general population, <sup>37,39–41,43,46</sup> medical centres, <sup>36,47</sup> high exposed area, <sup>38</sup> professional farmers <sup>44</sup> or other targeted populations. 40-42,45

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# **Population**

- Sample size ranged from 100<sup>36,39,45</sup> to 1162.<sup>44</sup> We included 6789 individuals in total, among 203 which 1981 (29.1%) had MetS. Prevalence of MetS ranged from 20.4<sup>40</sup> to 64.0%.<sup>47</sup>
- **Age** was reported in all studies except one.<sup>36</sup> Mean age ranged from 18<sup>38,47</sup> to 78<sup>47</sup> years. 205
- **Gender** was reported in all studies except three. <sup>36,41,43</sup> Proportion of men ranged from 32.0% <sup>37</sup> to 206
- 50%.51 207
- Body mass index (BMI) and waist circumference were reported as mean±SD in all 208
- studies except five, <sup>36,40,43,44,51</sup> and ranged from 23.2±1.4<sup>45</sup> to 30.2±7.4<sup>37</sup> kg/m<sup>2</sup> for BMI, and from 209
- 75.3<sup>39</sup> to 99.9<sup>36</sup> cm for waist circumference. Only one study included only obese individuals' 210
- grade 2 and 3, with a BMI of 42±5.1 kg/m<sup>2</sup> and a waist circumference of 125±13 cm.<sup>47</sup> Three 211
- 212 studies only reported proportions of normal, overweight, and obese participants. <sup>37,44,46</sup>
- **Blood pressure (BP)** was reported in all studies except four. 41–43,46 Mean systolic BP ranged 213
- from 114.9<sup>39</sup> to 147.1 mmhg<sup>40</sup> while mean diastolic BP ranged from 70.6<sup>39</sup> to 85.5<sup>47</sup> mmHg. 214
- **Triglycerides and HDL** were reported in all studies except three. 41,43,47 Mean levels were 215
- reported in six studies, <sup>36,39,41,42,44,45</sup> and ranged from 76.8<sup>39</sup> to 211.7<sup>39</sup> mg/dl for triglycerides, and 216
- from 35.6<sup>39</sup> to 62.0<sup>45</sup> mg/dl for HDL. Proportion of individuals with high triglycerides levels 217
- 218 (≥150 mg/dl) or with low HDL level (≤40 in men and ≤50 mg/dl in women) were reported in two
- studies,<sup>37,46</sup> ranging from 13.5%<sup>46</sup> to 38.0%<sup>37</sup> for TG level, and from 34.9%<sup>46</sup> to 59%<sup>37</sup> for HDL 219
- 220 level.

- Fasting glucose level was reported in all studies except three. 40,41,43 Mean glucose levels
- 222 ranged from 81.4<sup>36</sup> to 122.7<sup>36</sup> mg/dl.
- 223 **Smoking status** was reported in all studies except four. <sup>36,39,43,46</sup> Proportion of smokers ranged
- 224 from 8.0 <sup>45</sup> to 17.6%.<sup>47</sup>
- 225 **Alcohol consumption** was reported in three studies.<sup>37,42,44</sup> Percentage of consumers ranged
- from  $31.0^{37}$  to 49.6%, <sup>44</sup> but studies used different cut-offs.
- Other descriptive variables were marital status,<sup>37,44</sup> education<sup>37,40,46</sup> or occupation,<sup>46</sup>
- race, 37 physical activity 40,42,45,46 or energy intake, 40 family history of diabetes 37,45 or heart
- disease, 37 or personal history of liver disease, 37 myocardial infarction 41 or stroke, 41
- 230 medications, 40,41 other biological parameters (HOMA-IR, 37,39,45,46 insulin, 39,45 total
- 231 cholesterol,<sup>37,42</sup> LDL,<sup>40</sup> CRP,<sup>46,47</sup> or ALAT<sup>47</sup>), and body composition.<sup>45</sup>

# Measurements of pesticides exposure and risk of MetS

All studies employed gas chromatography/isotope and high-resolution mass spectrometry using the isotope dilution method to measure the concentration of pesticides in plasma<sup>40,41</sup> or serum<sup>36,37,39,42,43,45-47</sup> blood sampling as exposure measurement except one study that used the intensity level of pesticides exposure and cumulative exposure index.<sup>44</sup> Similarly, all studies used either serum or plasma samples to measure the biological parameters of MetS (blood glucose levels, HDL cholesterol and triglycerides). Risk of MetS were expressed as odds ratio in all studies except three studies that used relative risks,<sup>44</sup> prevalence ratio,<sup>46</sup> or β-coefficient.<sup>38</sup> All studies reported measurements of pesticides and their contaminants except three where only pesticides were explored.<sup>36,39,44</sup> Contaminants were measured using the same methods.

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# Meta-analysis on the risk of MetS following pesticides exposure

Overall exposure to pesticides and their contaminants increased the risk of MetS by 30% (OR=1.30, 95CI 1.22 to 1.37). All studies included exposure to organochlorine pesticides. Organochlorine increased the risk of MetS by 23% (14 to 32%). Among organochlorine, most pesticides increased the risk of MetS: hexachlorocyclohexane increased the risk by 53% (OR=1.53, 95CI 1.28 to 1.78) (using data from six studies<sup>36,37,39,42,43,46</sup>), hexachlorobenzene by 40% (95CI 0 to 80%) (five studies<sup>37,39,42,46,51</sup>), dichlorodiphenyldichloroethylene by 22% (9 to 34%) (seven studies<sup>36,37,39,41-43,46</sup>), dichlorodiphenyltrichloroethane by 28% (5 to 50%) (five studies<sup>36,37,39,42,46</sup>), oxychlordane by 50% (14 to 85%) (four studies<sup>37,39,42,43</sup>), and transnonchlor by 56% (22 to 90%) (five studies<sup>37,39,41-43</sup>), with more than four studies for each type of pesticides. Besides organochlorine pesticides, there were no data for the following types of pesticides: organophosphates, carbamates, pyrethroids, triazines. Only one study reported a risk for concomitant exposure to several pesticides, precluding further analyses. Pesticides' contaminants increased the risk by 31% (19 to 42%), mainly represented by PCB that increased the risk of MetS by 32% (18 to 45%) (Figure 3).

# Sensitivity analyses

All mentioned meta-analysis gave similar results after exclusion of studies that were not evenly distributed around the base of the metafunnel (Figure 4), as well as sensitivity analyses based on the type of risks i.e. while removing the three studies that did not express the risk of MetS as odds ratio. We also performed a meta-analysis only using crude odds ratio (Figure 5 and Supplementary Figure 1) and only using fully adjusted odds ratio (Figure 5 and Supplementary Figure 2). The meta-analysis using only crude odds ratio still showed an increased risk of MetS after exposure to pesticides and their contaminants by 46% (35 to 56%), and for overall pesticides by 42% (29 to 55%). Nevertheless, meta-analysis using fully adjusted odds ratio were not significant for the risk

of MetS following pesticides exposure (OR=1.09, 95CI 0.98 to 1.19). However, the risk of MetS was significantly increased following exposure to polychlorobiphenyls and overall contaminants. Such an effect of contaminants was still apparent as a 19% (10 to 29%) increase of the fully adjusted odds ratio for the combination of pesticides and contaminants. The only single pesticides increasing risk of MetS in the fully-adjusted meta-analysis hexachlorocyclohexane at an increased risk of 36% (17 to 55%) and transnonchlor at 25% (3 to 48%). Lastly, the results were also similar after exclusion of the two studies that included individuals over 70 years old, 46,47 i.e. an increased risk of MetS after exposure to pesticides and their contaminants of 29% (22 to 36%), and for overall pesticides of 27% (18 to 37%).

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# Metaregressions

281 Hexachlorocyclohexane increased the risk of MetS in comparison to most other pesticides.

282 Simultaneous exposure to several pesticides showed no higher risk of MetS in comparison to

single exposure of pesticide, but there was only one study reporting multi-pesticides exposure.

The risk of MetS following pesticides exposure increased for most recent studies (Coefficient

0.28, 95CI 0.20 to 0.37, p<0.001, by year). Higher BMI and waist circumference decreased

the risk of MetS following pesticides exposure (-0.28, -0.38 to -0.18; and -0.11, -0.23 to -0.00,

respectively). The risk of MetS following pesticides exposure was lower in men (-0.14, -0.19

to -0.10, per %men) (Figure 6).

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# **Discussion**

The main finding was that pesticides exposure increases the risk of MetS, with data mainly for organochlorine pesticides. The risk has increased in recent years. More recent studies have shown an increased risk of MetS following pesticides exposure. Physiologically, the inverse relationship with BMI may signify a stockage of pesticides in fat tissue.

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# Pesticides exposure as a major risk of MetS

The exposure to pesticides has been proven to increase the risk for multiple cardiovascular and metabolic pathologies<sup>6</sup> such as myocardial infarction, <sup>7</sup> diabetes<sup>8</sup> and obesity. <sup>9</sup> MetS is considered a prediabetic stage and is also a major risk of cardiovascular diseases. 52-55 The onset of MetS is linked to visceral adiposity, which is itself linked to an excess of fast fat and sugar, and sedentary lifestyle. This visceral adiposity is responsible for activating the inflammatory pathways mediated by cytokines.<sup>56</sup> This mechanism leads to the production of nitric oxide and insulin resistance. The last two lead to a vasodilation defect and thus an alteration of the endothelium implicated in the cardiovascular complications of the MetS.<sup>57</sup> Major meta-analyses demonstrated the increased risk of T2DM, obesity and cardiovascular diseases following exposure to pesticides.<sup>7-9</sup> We demonstrated that pesticides exposure increases the risk for MetS which is closely related to cardiovascular and metabolic diseases.<sup>50</sup> Mechanistically, MetS is closely related to dysregulation of adipose tissue, i.e. visceral adiposity.<sup>58</sup> The increased risk of MetS following pesticides exposure may be directly linked to their effects on adipose tissue. Some pesticides, like DDE may enhance proliferative capacity of preadipocytes and thus the development of adipose tissue.<sup>59</sup> Others, like OCP, may disturb the function of estrogen receptors that play a role in adipogensis.<sup>60</sup> In addition to pesticides exposure, their contaminants also increased the risk of MetS. PCDF and PCDD are produced during pesticides manufacturing, burning of waste, and automobile emissions. 61,62 Their abundance in the environment may explain our findings as other pesticides are mainly used in agriculture. Despite OCP and PCB being banned since 1970s, 63 they are still a major risk of MetS which can be explained by their persistence in the environment. 5,64,65 For example, the unfortunately famous Chlordecone was an organochlorine used in the 1980s in the French West Indies. Despite its hazardous effect on humans, it was used until 1993 against a banana pest. Considered as one of the worst persistent organic pollutants (POP), for centuries chlordecone because of its very long remanence will continue to pollute Martinican rivers.<sup>66</sup>

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## Risk of MetS and types of pesticide

Organochlorines are the main types of pesticides explored in our study. Among organochlorines, hexachlorocyclohexane has the most important risk as it is found in soil, water and air; and as other organochlorine, it has a prolonged remanence.<sup>49</sup> It is produced in many countries and exported to countries that do not produce it anymore.<sup>49</sup> Although mechanism of pathological effects are not totally known, studies in animals showed alteration in cytochrome P450, glutathione-S-transferase and glucose-6-phosphate dehydrogenase activities indicating its implication in oxidative dysregulation.<sup>67</sup> Oxidative stress is now known to be increased in metabolic syndrome as it's associated with adiposity and insulin resistance.<sup>68</sup> Heptachlor-epoxide use is now restricted to control fire ants in power transformers.<sup>69</sup> We failed to show that heptachlor-epoxide increased a risk for MetS, which can be explained by its fast elimination from the organism,<sup>70</sup> and undetectable levels in population.<sup>71,72</sup> Several studies showed that transnonchlor and oxychlordane exposure increased the risk of diabetes and induced alteration of the metabolic profile, <sup>73</sup> and therefore our results on the risk of MetS is in line with the literature. A study on effects of oxychlordane in female rats showed a cluster of vacuolated hepatocytes between the hepatic triad and central vein.<sup>74</sup> These same changes are seen in nonalcoholic fatty liver disease<sup>75</sup> which is now known to have a close relation with MetS both pathophysiologically and by cardiovascular risk prediction.<sup>76</sup> Hexachlorobenzene is known to have pathological effects on the skin, liver and thyroid.<sup>77</sup> Despite a limited number of studies assessing its effects on the metabolic profile, we showed its risks on MetS. Dicholordiphenyldichloroethylene (DDE) and Dichlorodiphenyltrichloroethane (DDT) are related pesticides.<sup>77</sup> Even if DDE and DDT are mainly known for their side effects on liver and reproductive function, 77 we also demonstrated their potential to increase the risk of MetS. Surprisingly, mirex had an inverse relationship with the risk of MetS but scarcity of data prevailed as shown using the GRADE certainty of evidence. Although, no data are known on how it affects human health, studies on animals showed liver and kidney effects. Apart from organochlorines, results of our meta-analysis showed a lack of data for other types of pesticides (organophosphates, carbamates, pyrethroids, and triazines). Such data should be strongly needed considering their global use. We also showed that the risk was the highest for hexachlorocyclohexane and transnonchlor whatever the statistical model used, highlighting their side effect on the risk of MetS.

# Unique and multiple exposures in a time-effect manner

Multiple pesticides exposure is the most common exposure of populations, worldwide. 57,69 Literature showed that simultaneous exposure to pesticides and their contaminants increased the risk of variable pathologies more than single exposure. 13,40,41,79 We did not find such relation for the risk of MetS as most included studies examined only single pesticide exposure. Toxicological effects of low-dose pesticides mixture on human health are largely unknown and anticipating these effects is often difficult to prove, however synergism and potentiation in pesticides effects have been demonstrated in laboratory animals. Nonetheless, we demonstrated that even unique exposures increased the risk of MetS. Therefore, preventive strategy should also focus on handling single exposure as this may prevent multiple exposure. We demonstrated an increased risk with time as recent studies showed more important risk of MetS due to pesticides exposure. This may be explained by the increase production and use of pesticides worldwide to meet the growing global food demand, 14,15 but also by the persistence of banned pesticides in the environment. Furthermore, pesticides like other pollutants are well known to accumulate all along the food chain, with transfer across areas and over boarders; they even have been identified in areas where they have never been used. Our findings should trigger an alarm regarding limitations in the

steps taken to restrict the production of pesticides and development of better technologies for agriculture, food, and pesticide industries.<sup>62</sup>

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# **Anthropometric and sociodemographic effects**

MetS is closely related to visceral adiposity, <sup>56</sup> where pesticides mainly accumulate because of their lipophilic nature. 16 However, we demonstrated an inverse relationship between BMI or waist circumference, and the risk of MetS following pesticides exposure. Similarly as exposure to other lipophilic agents, such as POP for example, the adipose tissue may capture the pesticides and store them without provoking major detrimental effects. Sequestration of pesticide in adipose tissue, like other POPs, decrease their concentration in blood, limiting their availability to other tissues where they may have damaging effects.<sup>83</sup> This may add a protective function to adipose tissue regarding pesticides.<sup>84</sup> And paradoxically, lean individuals with few adipose tissues will have higher detrimental effects of POP because they will not be stored in the adipose tissue. Even if not all pesticides are POPs, they are lipophile, and thus the protective effect of adipose tissue should be similar. Despite the very strong inverse relationship between BMI and the risk of MetS following exposure to pesticides, two limitations should be noted. First, pesticides are present in higher concentration in adipose tissue than in blood, 16 but all included studies in our meta-analysis assessed blood pesticides levels and none assessed levels of pesticides within adipose tissue. However, assessing pesticides levels in adipose tissue need biopsies which is practically impossible to achieve on a large population scale. Second, a recent review of 17 studies showed significant increased blood levels of pesticides after weight-loss interventions such as hypocaloric diet, medications and bariatric surgery.<sup>85</sup> Studies included in our meta-analyses should have reported weight variation history of participants ahead of inclusion. The main other influencing variable is gender. While being a man is a risk factors for MetS, 18,86 we demonstrated that women were

more at risk of MetS following pesticides exposure. Our findings may be explained by differences in characteristics of adipose tissue and lifestyle patterns between men and women.<sup>87–92</sup> Even if most included studies enrolled more women than men, the unbalanced sex ratio should not have affected our results as we controlled this parameter in our analysis. Lastly, despite smoking and alcohol<sup>93–95</sup> are known to increase the risk of MetS, we showed that they have no modification on the risk of MetS following exposure to pesticides.

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## Limitations

Our study has some limitations. Only 11 studies were included in which some case-control studies had a limited number of participants, <sup>36,39,45</sup> and some studies included targeted population that did not represent the general population.<sup>36,40–42,44,45,47</sup> However, we included a large sample size of 6789 individuals, in which 1708 were included in case control studies of all ages and categories promoting the generalizability of our results. We also computed sensitivity analyses by excluding studies with the oldest individuals to limit age-related cofounders. Most studies had crosssectional or case-control designs, and prospective studies had also a short follow-up period, limiting the possibility to detect causality between pesticides exposure and MetS. All studies were in high-income and developed countries except one in India, where the use and exposure to pesticides is of greater concern. Having significant risk in developing countries where exposure to pesticides is of less concern in comparison to undeveloped countries may upgrade the quality of evidence of this meta-analysis. No study reported the duration of exposure allowing the analysis of a dose-effect relationship. However, the importance of duration of exposure has been illustrated in several pathologies including diabetes. 6,8,96-100 Some studies evaluated the association between pesticides exposure and MetS as a secondary aim which required extra statistical tests that might generate false-positive associations. 40,41,45,46 In all studies, exposure was assessed by measuring pesticides plasma or serum levels using gas chromatography/isotope and high-resolution mass

spectrometry using the isotope dilution method, except one study that assessed exposure using a questionnaire. 43 Although, not all studies offered details of measurement protocols, all measurements were done in laboratories attached to research facilities. We believe that differences in protocols could be negligible and would not affect the overall comparability across studies. Even though the major site of accumulation of pesticides is adipose tissue, blood pesticides levels are the only realistic large-scale assessments. 40 Most included studies did not adjust for lipid levels, even if such normalization is controversial. 12 There could be selection bias as one study used capillary blood instead of venous blood to define hyperglycemia.<sup>46</sup> There could also be an observation bias as only some selected pesticides were evaluated. However, our meta-analysis covered most common environmental pesticides. The fact that risk of MetS following pesticides exposure was less salient when using a fully adjusted model, could be related to differences across studies in their adjustment criteria, included populations, and observational designs. However, results were still significant for overall pesticides and contaminants, and for hexachlorocyclohexane and transnonchlor. All studies did not report characteristics of nonparticipants. Most included studies did not report lifestyle behavior such as pesticides using behavior, dietary habits, or physical activity. However, all included studies adjusted their risk measurement on anthropometric and sociodemographic variables such as age, sex, smoking, alcohol, education, or obesity. Overall, our meta-analysis inherited the limitations of the 11 individual studies of which it was composed and therefore was subjected to the bias of included studies such as bias of selection and bias of outcome assessment.

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# Conclusion

We demonstrated that pesticides exposure is a major risk of metabolic syndrome, increasing the risk by 42%. There were data mainly for organochlorine pesticides that globally increased the risk of MetS, as well as for most types of organochlorines, with a high level of proof for

hexachlorocyclohexane and transnonchlor. Besides organochlorine exposure, data are lacking for other types of pesticides. The risk increased in more recent studies which may be due to the increase of the worldwide use of pesticides. We also demonstrated an inverse relationship with body mass index, that may be linked with a storage of pesticides in fat tissue. Authorities should enforce measurements to limit the use and production of pesticides.

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#### 452 Competing Interests

No conflicting relationship exists for any author.

#### 454 **Author contributions**

- 455 Conceived and designed the study: FD. Conducted the systematic literature search: HL, SM. Final
- approval of eligible articles: all authors. Analyzed the data: FD. Wrote the first draft of the article:
- 457 HL, FD. Critical revision of the article: all authors. Integrity of the data analysis: HL, FD.

### 458 **Data availability statement**

459 All relevant data are within the paper.

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## References

- Sharma, A. et al. Worldwide pesticide usage and its impacts on ecosystem. *SN Appl. Sci. 2019 111* **1**, 1–16 (2019).
- Gangemi, S. et al. Occupational and environmental exposure to pesticides and cytokine pathways in chronic diseases (Review). *Int. J. Mol. Med.* **38**, 1012 (2016).
- 466 3. Lallas, P. L. et al. The Stockholm Convention on Persistent Organic Pollutants. *Am. J. Int. Law* **95**, 692–708 (2001).
- 468 4. Manz, M. et al. Persistent organic pollutants in agricultural soils of central Germany. 469 *Sci. Total Environ.* **277**, 187–198 (2001).
- Jiamo, F. et al. Persistent organic pollutants in environment of the Pearl River Delta, China: an overview. *Chemosphere* **52**, 1411–1422 (2003).
- 472 6. Kim, K. et al. Exposure to pesticides and the associated human health effects. *Sci.* 473 *Total Environ.* **575**, 525–535 (2017).
- 7. Zago, A. M. et al. Pesticide exposure and risk of cardiovascular disease: A systematic review. *Glob. Public Health* (2020) doi:10.1080/17441692.2020.1808693.

- 476 8. Evangelou, E. et al. Exposure to pesticides and diabetes: A systematic review and meta-analysis. *Environ. Int.* **91**, 60–68 (2016).
- 478 9. Cano-Sancho, G. et al. Association between Exposure to p,p'-DDT and Its Metabolite p,p'-DDE with Obesity: Integrated Systematic Review and Meta-Analysis. *Environ. Health Perspect.* **125**, (2017).
- 481 10. Leonetti, F. et al. Clinical, physiopathological and dietetic aspects of metabolic syndrome. *Dig. Liver Dis.* **34 Suppl 2**, S134–S139 (2002).
- 483 11. Eljarrat, E. et al. Priority lists for persistent organic pollutants and emerging contaminants based on their relative toxic potency in environmental samples. *TrAC* 485 *Trends Anal. Chem.* **22**, 655–665 (2003).
- 486 12. Gauthier, M. et al. The metabolically healthy but obese phenotype is associated with lower plasma levels of persistent organic pollutants as compared to the metabolically abnormal obese phenotype. *J. Clin. Endocrinol. Metab.* **99**, (2014).
- 489 13. Lee, D. et al. Low dose organochlorine pesticides and polychlorinated biphenyls predict obesity, dyslipidemia, and insulin resistance among people free of diabetes.

  491 PLoS One 6, e15977 (2011).
- 492 14. Popp, J. et al. Pesticide productivity and food security. A review. *Agron. Sustain. Dev.* 493 2012 331 33, 243–255 (2012).
- 494 15. Carvalho, F. P. et al. Pesticides, environment, and food safety. *Food Energy Secur.* **6**, 48–60 (2017).
- 496 16. Jackson, E. et al. Adipose tissue as a site of toxin accumulation. *Compr. Physiol.* **7**, 497 1085–1135 (2017).
- 498 17. Mustieles, V. et al. Human adipose tissue levels of persistent organic pollutants and metabolic syndrome components: Combining a cross-sectional with a 10-year longitudinal study using a multi-pollutant approach. *Environ. Int.* **104**, 48–57 (2017).
- 501 18. Moore, J. et al. Metabolic syndrome prevalence by race/ethnicity and sex in the United States, National Health and Nutrition Examination Survey, 1988-2012. *Prev. Chronic Dis.* **14**, (2017).
- 504 19. Li, Y. et al. Metabolic syndrome prevalence and its risk factors among adults in China: A nationally representative cross-sectional study. *PLoS One* **13**, (2018).
- 506 20. Park, S. et al. Exposure to pesticides and the prevalence of diabetes in a rural population in Korea. *Neurotoxicology* **70**, 12–18 (2019).
- 508 21. Stang, A. et al. Critical evaluation of the Newcastle-Ottawa scale for the assessment of the quality of nonrandomized studies in meta-analyses. *Eur. J. Epidemiol.* **25**, 603–605 (2010).
- 511 22. Alshabanat, A. et al. Asthma and COPD Overlap Syndrome (ACOS): a systematic review and meta analysis. *PLoS One* **10**, (2015).
- 513 23. Modesti, P. et al. Panethnic differences in blood pressure in europe: a systematic review and meta-analysis. *PLoS One* **11**, (2016).
- 515 24. Herzog, R. et al. Are healthcare workers' intentions to vaccinate related to their knowledge, beliefs and attitudes? a systematic review. *BMC Public Heal*. 2013 131 13, 1–17 (2013).
- 518 25. Guyatt, G. et al. GRADE guidelines: 1. Introduction-GRADE evidence profiles and summary of findings tables. *Journal of clinical epidemiology* **64**, 383–394 (2011).
- 520 26. Benichou, T. et al. Heart rate variability in type 2 diabetes mellitus: A systematic review and meta-analysis. *PLoS One* **13**, (2018).
- 522 27. Chaplais, E. et al. Effects of interventions with a physical activity component on bone 523 health in obese children and adolescents: a systematic review and meta-analysis. *J.*
- *Bone Miner. Metab.* **36**, 12–30 (2018).
- 525 28. Courtin, R. et al. Prevalence of dry eye disease in visual display terminal workers: a

- 526 systematic review and meta-analysis. *BMJ Open* **6**, (2016).
- 527 29. Benoist d'Azy, C. et al. Antibioprophylaxis in prevention of endophthalmitis in intravitreal injection: a systematic review and meta-analysis. *PLoS One* **11**, e0156431 (2016).
- 530 30. Benoist d'Azy, C. et al. Oxidative and anti-oxidative stress markers in chronic glaucoma: a systematic review and meta-analysis. *PLoS One* **11**, (2016).
- 532 31. Lanhers, C. et al. Creatine supplementation and upper limb strength performance: a systematic review and meta-analysis. *Sports Med.* **47**, 163–173 (2017).
- 534 32. Lanhers, C. et al. Creatine supplementation and lower limb strength performance: a systematic review and meta-analyses. *Sports Med.* **45**, 1285–1294 (2015).
- 536 33. Ollier, M. et al. Chest CT scan screening for lung cancer in asbestos occupational exposure: a systematic review and meta-analysis. *Chest* **145**, 1339–1346 (2014).
- 538 34. DerSimonian, R. et al. Meta-analysis in clinical trials revisited. *Contemp. Clin. Trials* 45, 139 (2015).
- 540 35. Russo, M. W. et al. How to review a meta-analysis. *Gastroenterol. Hepatol. (N. Y).* **3**, 637 (2007).
- 542 36. Tomar, L. R. et al. Serum organochlorine pesticide levels in patients with metabolic syndrome. *Indian J. Endocrinol. Metab.* **17**, S342 (2013).
- 544 37. Rosenbaum, P. et al. Metabolic syndrome is associated with exposure to organochlorine pesticides in Anniston, AL, United States. *Environ. Int.* **108**, 11–21 (2017).
- 547 38. Aminov, Z. et al. Serum concentrations of persistent organic pollutants and the metabolic syndrome in Akwesasne Mohawks, a Native American community. *Environ.* 549 *Pollut.* **260**, (2020).
- 550 39. Park, S. et al. Relationship between serum concentrations of organochlorine pesticides and metabolic syndrome among non-diabetic adults. *J. Prev. Med. Public Health* **43**, 552 1–18 (2010).
- 553 40. Lind, L. et al. Mixture effects of 30 environmental contaminants on incident metabolic syndrome-A prospective study. *Environ. Int.* **107**, 8–15 (2017).
- 555 41. Lind, P. et al. An environmental wide association study (EWAS) approach to the metabolic syndrome. *Environ. Int.* **55**, 1–8 (2013).
- Lee, Y. et al. Prospective associations between persistent organic pollutants and metabolic syndrome: a nested case-control study. *Sci. Total Environ.* **496**, 219–225 (2014).
- 560 43. Lee, D. et al. Relationship between serum concentrations of persistent organic 561 pollutants and the prevalence of metabolic syndrome among non-diabetic adults: 562 results from the National Health and Nutrition Examination Survey 1999–2002. 563 *Diabetol.* 2007 509 **50**, 1841–1851 (2007).
- Kim, S. et al. Pesticides as a risk factor for metabolic syndrome: population-based longitudinal study in Korea. *Mol. Cell. Toxicol.* **15**, 431–441 (2019).
- Ha, K. et al. Can persistent organic pollutants distinguish between two opposite metabolic phenotypes in lean Koreans? *Diabetes Metab.* **44**, 168–171 (2018).
- 568 46. Gasull, M. et al. Blood Concentrations of persistent organic pollutants and unhealthy metabolic phenotypes in normal-weight, overweight, and obese individuals. *Am. J. Epidemiol.* **187**, 494–506 (2018).
- 571 47. Dusanov, S. et al. Associations between persistent organic pollutants and metabolic syndrome in morbidly obese individuals. *Nutr. Metab. Cardiovasc. Dis.* **28**, 735–742 (2018).
- 574 48. Szumilas, M. et al. Explaining odds ratios. *J. Can. Acad. Child Adolesc. Psychiatry* **19**, 575 227 (2010).

- 576 49. Agency for Toxic Substances and Disease. Hexachlorocyclohexane (HCH) | Public Health Statement | ATSDR. Centers for Disease Control and Prevention CDC (2005).
- 578 50. Kim, J. et al. Reciprocal Relationships between insulin resistance and endothelial dysfunction. *Circulation* **113**, 1888–1904 (2006).
- 580 51. Lind, L. et al. Mixture effects of 30 environmental contaminants on incident metabolic syndrome—A prospective study. *Environ. Int.* **107**, 8–15 (2017).
- 582 52. Vinet, A. et al. Different modalities of exercise improve macrovascular function but not microvascular function in metabolic syndrome: The RESOLVE randomized trial. *Int. J. Cardiol.* **267**, 165–170 (2018).
- 585 53. Dutheil, F. et al. Different modalities of exercise to reduce visceral fat mass and cardiovascular risk in metabolic syndrome: the RESOLVE\* randomized trial. *Int. J. Cardiol.* **168**, 3634–3642 (2013).
- 588 54. Boudet, G. et al. Paradoxical dissociation between heart rate and heart rate variability following different modalities of exercise in individuals with metabolic syndrome: The RESOLVE study. *Eur. J. Prev. Cardiol.* **24**, 281–296 (2017).
- 591 55. Dutheil, F. et al. Cardiovascular risk of adipokines: a review. *J. Int. Med. Res.* **46**, 592 2082–2095 (2017).
- 593 56. Quyyumi, A. et al. Endothelial function in health and disease: new insights into the genesis of cardiovascular disease. *Am. J. Med.* **105**, 32S-39S (1998).
- 595 57. Galassi, A. et al. Metabolic syndrome and risk of cardiovascular disease: a meta-596 analysis. *Am. J. Med.* **119**, 812–819 (2006).
- 597 58. Walther, G. *et al.* Metabolic syndrome individuals with and without type 2 diabetes mellitus present generalized vascular dysfunction: cross-sectional study.

  599 *Arteriosclerosis, thrombosis, and vascular biology* **35**, 1022–1029 (2015).
- 600 59. Chapados, N. A. *et al.* Increased proliferative effect of organochlorine compounds on human preadipocytes. *Molecular and cellular biochemistry* **365**, 275–278 (2012).
- 602 60. Casals-Casas, C. & Desvergne, B. Endocrine disruptors: from endocrine to metabolic disruption. *Annual review of physiology* **73**, 135–162 (2011).
- 604 61. Anwar, W. A. et al. Biomarkers of human exposure to pesticides. *Environ. Health Perspect.* **105**, 801 (1997).
- 606 62. Damalas, C. et al. Pesticide exposure, safety issues, and risk assessment indicators. *Int. J. Environ. Res. Public Health* **8**, 1402–1419 (2011).
- 608 63. Liu, J. et al. A Critical review of polychlorinated biphenyls metabolism, metabolites, and their correlation with oxidative stress. *Chem. Res. Toxicol.* **33**, 2022–2042 (2020).
- 610 64. Vandermarken, T. et al. Dioxins, furans and dioxin-like PCBs in sediment samples and suspended particulate matter from the Scheldt estuary and the North Sea Coast:
  Comparison of CALUX concentration levels in historical and recent samples. *Sci.*1013 *Total Environ.* 626, 109–116 (2018).
- 614 65. Li, Y. et al. Quantitative analysis of 209 polychlorinated biphenyl (PCB) congeners in sediments of the Raritan River estuary, USA. *Water Supply* **20**, 2400–2414 (2020).
- 616 66. Baudry, T. *et al.* The invasive crayfish Cherax quadricarinatus facing chlordecone in Martinique: Bioaccumulation and depuration study. *Chemosphere* **286** (2022).
- 618 67. Willett, K. L., Ulrich, E. M. & Hites, R. A. Differential toxicity and environmental fates of hexachlorocyclohexane isomers. *Environmental Science and Technology* **32**, 2197–2207 (1998).
- 621 68. Roberts, C. K. & Sindhu, K. K. Oxidative stress and metabolic syndrome. *Life Sciences* **84**, 705–712 (2009).
- 623 69. Reed, N. R. et al. Heptachlor. in *Encyclopedia of Toxicology* 840–844 (Elsevier, 2014).
- Wagstaff, D. J. et al. Egg Binding in Broiler Chickens Fed Heptachlor. *Poult. Sci.* 60, 1578–1580 (1981).

- Mussalo-Rauhamaa, H. et al. Heptachlor, heptachlor epoxide, and other chlordane compounds in Finnish plywood workers. *Arch. Environ. Health* **46**, 340–346 (1991).
- 528 72. Stehr-Green, P. et al. Demographic and seasonal influences on human serum pesticide residue levels. *J. Toxicol. Environ. Health* **27**, 405–421 (1989).
- 630 73. Carpenter, D. et al. Environmental contaminants as risk factors for developing diabetes. 631 *Rev. Environ. Health* **23**, 59–74 (2008).
- 632 74. Bondy, G. *et al.* Toxicity of the chlordane metabolite oxychlordane in female rats: clinical and histopathological changes. *Food and chemical toxicology: an international journal published for the British Industrial Biological Research Association* **41**, 291–301 (2003).
- Kleiner, D. E. & Makhlouf, H. R. Histology of nonalcoholic fatty liver disease and nonalcoholic steatohepatitis in adults and children. *Clinics in liver disease* **20**, 293–312 (2016).
- 76. Yki-Järvinen, H. Non-alcoholic fatty liver disease as a cause and a consequence of metabolic syndrome. *The Lancet Diabetes & Endocrinology* **2**, 901–910 (2014).
- Reed, L. et al. Environmental toxicology and health effects associated with hexachlorobenzene exposure. *Rev. Environ. Health* **22**, 213–243 (2007).
- 643 78. Agency for Toxic Substances and Disease. Chlordecone | Public Health Statement | ATSDR. Centers for Disease Control and Prevention CDC (1995).
- 645 79. Margina, D. et al. Overview of the effects of chemical mixtures with endocrine 646 disrupting activity in the context of real-life risk simulation (RLRS): An integrative 647 approach (Review). *World Acad. Sci. J.* **1**, 157–164 (2019).
- 648 80. Hernández, A. F. *et al.* Toxic effects of pesticide mixtures at a molecular level: Their relevance to human health. *Toxicology* **307**, 136–145 (2013).
- Bro-Rasmussen, F. Contamination by persistent chemicals in food chain and human health. *Science of The Total Environment* **188**, S45–S60 (1996).
- 652 82. Zaynab, M. *et al.* Health and environmental effects of silent killers Organochlorine 653 pesticides and polychlorinated biphenyl. *Journal of King Saud University - Science* **33**, 654 101511 (2021).
- 655 83. La Merrill, M. *et al.* Toxicological function of adipose tissue: focus on persistent organic pollutants. *Environmental health perspectives* **121**, 162–169 (2013).
- 657 84. Jackson, E., Shoemaker, R., Larian, N. & Cassis, L. Adipose tissue as a site of toxin accumulation. *Comprehensive Physiology* **7**, 1085 (2017).
- Jansen, A., Lyche, J. L., Polder, A., Aaseth, J. & Skaug, M. A. Increased blood levels of persistent organic pollutants (POP) in obese individuals after weight loss-A review.
   *Journal of toxicology and environmental health. Part B, Critical reviews* 20, 22–37 (2017).
- 86. Beigh, S. H. et al. Prevalence of metabolic syndrome and gender differences. Bioinformation **8**, 613 (2012).
- Varì, R. et al. Gender-related differences in lifestyle may affect health status. *Ann. Ist.*Super. Sanita **52**, 158–166 (2016).
- 667 88. Study, F. A. et al. Lifestyle factors associated with age-related differences in body composition: the Florey Adelaide Male Aging Study. *Am. J. Clin. Nutr.* **88**, 95–104 (2008).
- 670 89. Karastergiou, K. et al. Sex differences in human adipose tissues the biology of pear shape. *Biol. Sex Differ. 2012 31* **3**, 1–12 (2012).
- 672 90. Goedecke, J. et al. The role of adipose tissue in insulin resistance in women of African ancestry. *J. Obes.* **2013**, (2013).
- Hoffman, D. et al. Comparison of visceral adipose tissue mass in adult african americans and whites\*\*. *Obes. Res.* **13**, 66–74 (2005).

- 676 92. Hill, J. et al. Racial differences in amounts of visceral adipose tissue in young adults: the CARDIA (Coronary Artery Risk Development in Young Adults) Study. *Am. J. Clin. Nutr.* **69**, 381–387 (1999).
- 679 93. Sun, K. et al. Active smoking and risk of metabolic syndrome: a meta-analysis of prospective studies. *PLoS One* **7**, e47791 (2012).
- 681 94. Fan, A. Z. et al. Patterns of alcohol consumption and the metabolic syndrome. *J. Clin. Endocrinol. Metab.* **93**, 3833–3838 (2008).
- 683 95. Baik, I. et al. Prospective study of alcohol consumption and metabolic syndrome. *Am. J. Clin. Nutr.* **87**, 1455–1463 (2008).
- Whorton, D. et al. Infertility in male pesticide workers. *Lancet (London, England)* **2**, 1259–1261 (1977).
- 687 97. Barthel, E. et al. Increased risk of lung cancer in pesticide-exposed male agricultural workers. *J. Toxicol. Environ. Health* **8**, 1027 (1981).
- 689 98. Merhi, M. et al. Occupational exposure to pesticides and risk of hematopoietic cancers: meta-analysis of case-control studies. *Cancer Causes Control* **18**, 1209–1226 (2007).
- 691 99. Settimi, L. et al. Prostate cancer and exposure to pesticides in agricultural settings. *Int.* 692 *J. cancer* **104**, 458–461 (2003).
- 693 100. Priyadarshi, A. et al. Environmental risk factors and parkinson's disease: A metaanalysis. *Environ. Res.* **86**, 122–127 (2001).
- 695 101. Sterne, J. A. C. et al. Systematic reviews in health care: Investigating and dealing with publication and other biases in meta-analysis. *BMJ Br. Med. J.* **323**, 101 (2001).
- 697 102. Martínez-Valenzuela, M. D. C. *et al.* Comparison of organochlorine pesticide levels 698 between human blood serum and adipose tissue. *Revista Internacional de* 699 *Contaminación Ambiental* 33, 393–401 (2017).

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**Table 1.** Characteristics of included studies

\*: not included in the meta-analysis. ATP3: National cholesterol education program, adult treatment panel III; BMI: Body mass index; IDF: International Diabetes Federation; OR: Odds ratio; PR: prevalence risk; RR: relative risk.

Charder		Period	Design	Population					Metabolic syndrome	0.1	
Study	Country			Total (n)	Men (%)	Age	Туре	Definition	Risk	Adjustment	Other variables listed
Aminov 2020*	USA	2016	cross-sectional	601	37.27	18-84	Local community	ATP3	β	Age, gender, smoking, BMI, serum total lipids	-
Dusanov 2018	Norway	2005-2010	cross-sectional	431	37.4	18-78	Clinic patients	ATP3	OR	Age, sex, BMI, smoking, alcohol, total cholesterol	CRP, ALAT
Gasull 2018	Spain	2002	cross-sectional	860	43.6	-	General population	ATP3 / IDF	PR	Age, sex, BMI, smoking, alcohol, physical activity	Social class, education, physical activity, CRP, HOMA-IR
Ha 2018	South Korea	2016	nested case- control	100	18.0	-	Targeted population	ATP3	OR	Smoking, family history of diabetes, HOMA-IR.	Total fat mass, lean mass, visceral fat mass, physical activity, family history of diabetes, fasting insulin
Kim 2019	South Korea	2005-2008	cohort	1162	56.6	39-72	Farmer	ATP3	RR	Age, smoking, alcohol, regular exercise, education, marital status	Marital status
Lee 2007	USA	1999-2002	cross-sectional	721	-	≥20	General population	ATP3	OR	Age, sex, race, poverty income ratio, smoking, serum cotinine, alcohol, exercise, BMI	-
Lee 2014	South Korea	2006-2010	nested case- control	246	32.9	≥40	Local community	ATP3	OR	Age, sex, smoking, alcohol, exercise, triglycerides, total cholesterol, BMI	Physically inactive, total cholesterol
Lind 2013	Sweden	2011	cross-sectional	1016	-	70	General population	ATP3	OR	Fatty acids, fatty acid factor, sex.	Myocardial infarction, stoke, diabetes, statin treatment, alcohol intake
Lind 2017	Sweden	2006-2017	cohort	452	50.0	70	General population	ATP3	OR	Lipid levels	Education, exercise habits, energy intake, fat intake, hypertension treatment, LDL, alcohol intake
Park 2010	South Korea	2006	nested case- control	100	28.0	≥40	General population	ATP3	OR	Age, sex, alcohol, smoking, BMI	Race, education, marital status, family history of diabetes / heart disease, liver disease, total lipids
Rosenbaum 2017	USA	2005-2007	cross-sectional	548	32.0	>18	General population	ATP3	OR	Age, family of heart disease/diabetes, liver disease, sex, race, alcohol, smoking, marital status, education, BMI, total lipids	Fasting insulin
Tomar 2013	India	2013	case-control	100	-	-	Clinic patients	ATP3	OR	Age, sex, smoking, alcohol	-

# **Figures Legend**

#### **Figure 1**. Search strategy.

We followed the Preferred Reporting Items for Systematic Reviews and Meta-Analyses (PRISMA) guidelines for the search strategy.

#### Figure 2. Risk of bias assessment – Newcastle Ottawa scale

We used the Newcastle Ottawa scale (NOS) criteria to assess the methodological quality of the included articles, both for case-control and cohort studies (9 items) and cross-sectional studies (10 items), with the dedicated evaluation grids. We presented an overall quality score based on three dimensions: selection, comparability, and exposure for case-control studies; and selection, comparability and outcome for cohort and cross-sectional studies.

#### **Figure 3**. Meta-analysis stratified by type of pesticides (using all risks)

Each overall summary of a meta-analysis is represented in the graph by a lozenge on a horizontal line and each stratification by a dot. Dots represent risk of MetS after exposure by individual pesticides or individual contaminants, grey lozenges represent risks by types of pesticides, blue lozenges for the overall pesticides and the overall contaminants, and the red lozenge for the overall combination of pesticides and contaminants. The lozenges / dots represent the overall pooled-effect estimate of individual meta-analyses (pooled effect size - ES), and the length of each horizontal line around the dots represent their 95% confidence interval (95C1). Shorter lines represent a narrower 95C1 thus higher precision around pooled-ES. Conversely, longer lines represent a wider 95Cl and less precision around pooled-ES. The black solid vertical line represents the null estimate (with a value of 1). Horizontal lines that cross the null vertical line represent a non-significant overall summary of the meta-analysis, for each efficacy outcome.

n studies (subgroups): number of studies and subgroups included for each meta-analysis; I-squared (%): percentage of heterogeneity between studies for each meta-analysis; Weight (%): Weight of each stratification (meta-analysis) in the overall meta-analysis.

#### **Figure 4**. Metafunnels

Each dot represents a single study, with its corresponding effect size (x axis) and its associated standard error of the effect estimate (y-axis). Large high-powered studies are placed towards the top, and smaller low-powered studies towards the bottom. The plot should ideally resemble a pyramid or inverted funnel, with scatter due to sampling variation. Studies outside funnel plot are likely to present bias.<sup>101</sup>

**Figure 5**. Sensitivity analysis: meta-analysis using crude odds ratio and fully adjusted model (see details in Supplementary Figure 1 and 2)

Each overall summary of a meta-analysis is represented in the graph by a lozenge on a horizontal line and each stratification by a dot. Dots represent risk of MetS after exposure by individual pesticides or individual contaminants, grey lozenges represent risks by types of pesticides, blue lozenges for the overall pesticides and the overall contaminants, and the red lozenge for the overall combination of pesticides and contaminants. The lozenges / dots represent the overall pooled-effect estimate of individual meta-analyses (pooled effect size - ES), and the length of each horizontal line around the dots represent their 95% confidence interval (95CI). The black solid vertical line represents the null estimate (with a value of 1). Horizontal lines that cross the null vertical line represent a non-significant overall summary of the meta-analysis, for each efficacy outcome.

**Figure 6**. Meta-regression i.e. factors influencing the risk of MetS following pesticides exposure

The effect of each variable on the outcome is represented in the forest-plot by a dot on a horizontal line. The dots represent the coefficient for each variable, and the length of each line around the dots represent their 95% confidence interval (95CI). The black solid vertical line represents the null estimate (with a value of 0). Horizontal lines that cross the null vertical line represent non-significant variables on the outcome.

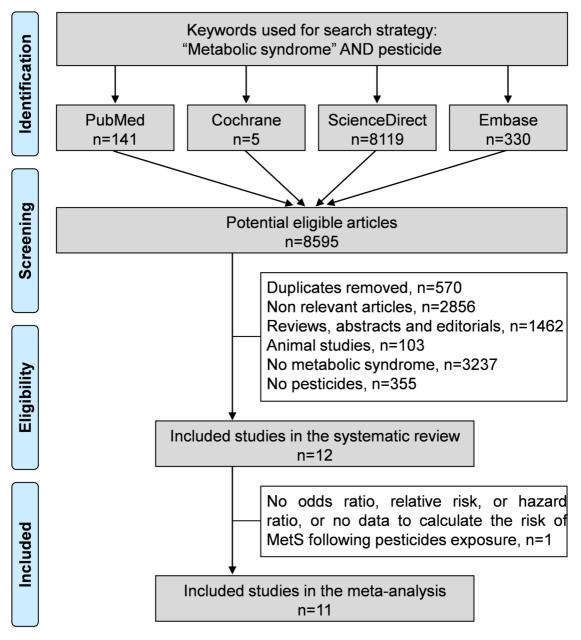
# **Supplementary materials**

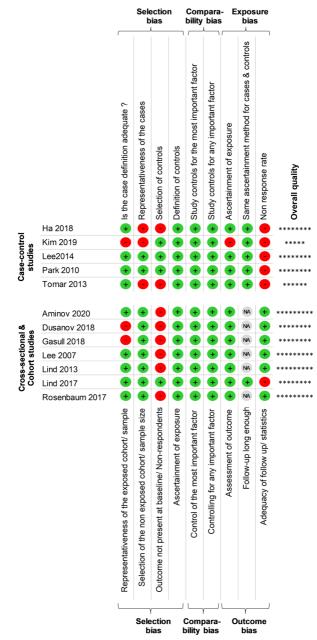
**Supplementary Figure 1**. Meta-analysis stratified by type of pesticide (crude model)

**Supplementary Figure 2**. Meta-analysis stratified by type of pesticide (fully adjusted model)

**Appendix 1.** Details for the search strategy used within each database

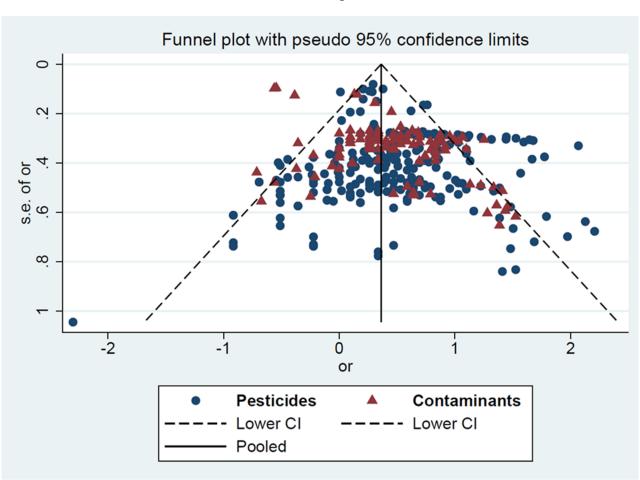
**Appendix 2.** Grading of Recommendation, Assessment, Development, and Evaluation (GRADE) instrument





	n studies	I-squared					Odds Ratio	Weight
								_
Bar California	(subgroups)	(%)					(95% CI)	(%)
Pesticides								
Organochlorine pesticides								
Hexachlorocyclohexane (HCH)	6 (28)	6.8		_		_	1.53 (1.28 to 1.78)	7.9
Hexachlorobenzene (HCB)	5 (28)	68.4				_	1.40 (1.00 to 1.80)	10.9
Dicholordiphenyldichloroethylene (DDE)	7 (33)	0.0		<u> </u>			1.22 (1.09 to 1.34)	16.5
Dichlorodiphenyltrichloroethane (DDT)	5 (24)	5.8					1.28 (1.05 to 1.50)	10.5
Oxychlordane (OC)	4 (26)	0.0			_		1.24 (1.01 to 1.47)	10.1
Transnonchlor (TNC)	5 (28)	0.0		_	_		1.35 (1.19 to 1.52)	10.6
Heptachlor epoxide (HCE)	2 (10)	0.0	<del></del>		_		0.99 (0.49 to 1.50)	2.2
Mirex	2 (22)	9.9	<b>—</b> ○—				0.73 (0.57 to 0.89)	16.1
Aldrin	1 (1)	-	ŀ		_		1.23 (0.99 to 1.48)	1.7
Mixed organochlorine	4 (32)	9.5	-		_		1.28 (1.00 to 1.56)	8.7
Overall organochlorine	11 (216)	32.4		<b>-</b> ◇-			1.23 (1.14 to 1.32)	95.1
Organophospates							no data	
Carbamates							no data	
Pyrethroids							no data	
Triazines							no data	
Mixed pesticides	1 (6)	0.6				$\diamond$ —	1.77 (1.47 to 2.07)	4.9
Overall pesticides	11 (222)	34.9		-			1.27 (1.18 to 1.36)	100.0
Contaminants	, ,						,	
Polychlorobiphenyls (PCB)	8 (90)	59.9		_	_		1.32 (1.18 to 1.45)	78.1
Polychlorodibenzo-p-dioxins (PCDD)	1 (9)	0.0	+		_		1.21 (0.95 to 1.46)	11.7
Polychlorodibenzofuranes (PCDF)	1 (9)	0.0					1.40 (1.11 to 1.68)	10.2
Overall	8 (108)	56.4		-	-		1.31 (1.19 to 1.42)	100
Pesticides + Contaminants	. ,						, , , , , , , , , , , , , , , , , , , ,	
Overall cumulative exposure	11 (330)	48.5		•			1.30 (1.22 to 1.37)	100
·	. ,	0.8	5 1		1.5	2	2.5	

# Overall exposure



	Sensitivity analysis				
	Crude odds ratio	Fully-adjusted model			
Pesticides					
Organochlorine pesticides	1				
Hexachlorocyclohexane (HCH)					
Hexachlorobenzene (HCB)	<del></del>	<del></del>			
Dicholordiphenyldichloroethylene (DDE)	<del></del>	+-			
Dichlorodiphenyltrichloroethane (DDT)	<del></del>	<del></del>			
Oxychlordane (OC)		<del></del>			
Transnonchlor (TNC)	<del></del>	<del></del>			
Heptachlor epoxide (HCE)					
Mirex	<del></del>	·			
Aldrin	<b>—</b>	<del>-</del>			
Mixed organochlorine	<u> </u>	<del></del>			
Overall organochlorine	<b>→</b>	<b>├</b> -			
Organophospates					
Carbamates					
Pyrethroids					
Triazines					
Mixed pesticides	+	<del>-</del>			
Overall pesticides	<del></del>				
Contaminants					
Polychlorobiphenyls (PCB)	<del>-</del>	<b>—</b>			
Polychlorodibenzo-p-dioxins (PCDD)	<del></del>	<del></del>			
Polychlorodibenzofuranes (PCDF)	<del></del>				
Overall contaminants	_ <del>_</del>	<b>—</b>			
Pesticides + Contaminants					
Overall cumulative exposure	<b>→</b>	<b>*</b>			
0.5	1 1.5 2	0.5 1 1.5 2			

Variables		Coefficient (95CI)	p- value
Characteristics of individuals	1		
Age, years	•	-0.07 (-0.07 to 0.03)	0.16
Sex, % men		-0.14 (-0.19 to -0.10)	<0.001
Body Mass Index, kg.m-2		-0.28 (-0.38 to -0.18)	<0.001
Waist circumference, cm	1	-0.11 (-0.23 to 0.00)	0.05
Smoking, % smokers	Ĵ.	0.00 (0.01 to 0.01)	0.97
Alcohol, % consummers Periode of exposure (most recent)	Ĭ.	-0.01 (-0.02 to 0.05) 0.28 (0.20 to 0.37)	0.20 <b>&lt;0.001</b>
Comparisons between exposures		0.26 (0.20 to 0.37)	<b>\0.001</b>
HCH vs HCB		0.60 (0.16 to 1.04)	0.01
" vs DDE	<u> </u>	0.47 (0.05 to 0.90)	0.03
" vs DDT	<u> </u>	0.47 (0.04 to 0.90)	0.03
" vs OC	<b>─</b> ─	0.46 (0.04 to 0.89)	0.03
" vs TNC	<b>⊢</b> ∽	0.37 (-0.06 to 0.79)	0.09
" vs HCE	⇡─	0.01 (-0.56 to 0.58)	0.97
" vs Mirex	<b>─</b>	0.98 (0.57 to 1.39)	<0.001
" vs Aldrin	$\vdash$	0.44 (-1.81 to 2.68)	0.70
" vs Mixed OCP	<u>†</u> ~	0.23 (-0.20 to 0.67)	0.29
" vs Mixed pesticides		0.01 (-1.03 to 1.05)	0.98
HCB vs DDE		0.13 (-0.28 to 0.54)	0.54
" vs DDT		0.13 (-0.29 to 0.55)	0.54
" vs TNC	<u> </u>	0.14 (-0.27 to 0.54)	
" vs HCE		0.24 (-0.18 to 0.65) 0.59 (0.03 to 1.15)	0.26 <b>0.04</b>
" vs Mirex —o-	4	-0.38 (-0.77 to 0.01)	0.04
" vs Aldrin	-	0.16 (-2.08 to 2.40)	0.89
" vs Mixed OCP	<b>├</b> ─	0.37 (-0.06 to 0.79)	0.09
" vs Mixed pesticides —		0.59 (-0.44 to 1.62)	0.26
DDE vs DDT	<del>}</del>	0.01 (-0.40 to 0.41)	0.98
" vs OC —	<b>←</b>	0.01 (-0.38 to 0.40)	0.96
" vs TNC -	<del>-</del>	0.11 (-0.29 to 0.50)	0.59
" vs HCE	<del>-</del> -	0.46 (-0.08 to 1.01)	0.09
vs Mirex ——		-0.51 (-0.88 to -0.13)	0.01
VS AIUIIII		0.04 (-2.20 to 2.28)	0.97
vs Mixed OCP		0.24 (-0.17 to 0.65)	0.24
" vs Mixed pesticides — DDT vs OC —		0.46 (-0.56 to 1.49) 0.00 (-0.40 to 0.40)	0.37
" vs TNC -	<u></u>	0.10 (-0.30 to 0.51)	0.61
" vs HCE	L.	0.46 (-0.10 to 1.01)	0.10
" vs Mirex ———		-0.51 (-0.90 to -0.13)	0.01
" vs Aldrin	<del></del>	0.03 (-2.21 to 2.27)	0.98
" vs Mixed OCP	<u> </u>	0.24 (-0.18 to 0.65)	0.27
" vs Mixed pesticides —	<del></del>	0.46 (-0.57 to 1.48)	0.38
OC vs TNC	-	0.10 (-0.30 to 0.50)	0.62
" vs HCE	+~-	0.45 (-0.09 to 0.99)	0.10
vs ivillex ———	1	-0.52 (-0.89 to -0.14)	0.01
vs Aldrin ————		0.03 (-2.21 to 2.27)	0.98
vs Mixed OCF		0.23 (-0.18 to 0.64)	0.26
" vs Mixed pesticides —— TNC vs HCE		0.46 (-0.57 to 1.48) 0.35 (-0.19 to 0.90)	0.38
" vs Mirex ——		-0.62 (-0.99 to -0.24)	<0.001
" vs Aldrin		-0.07 (-2.30 to 2.17)	0.95
" vs Mixed OCP	_	0.13 (-0.28 to 0.54)	0.53
" vs Mixed Mixed pesticides	<u> </u>	0.36 (-0.67 to 1.38)	0.49
HCE vs Mirex ——		-0.97 (-1.50 to -0.43)	<0.001
" vs Aldrin —o—		-0.43 (-0.70 to 1.84)	0.71
" vs Mixed OCP	-	-0.22 (-0.33 to 0.79)	0.43
" vs Mixed pesticides ————	<del></del>	0.00 (-1.09 to 1.09)	0.10
Mirex vs Aldrin	$\vdash$ ${\sim}$	0.54 (-1.70 to 2.78)	0.63
" vs Mixed OCP		0.75 (0.36 to 1.14)	<0.001
" vs Mixed Mixed pesticides	_	0.97 (-0.05 to 1.99)	0.06
Aldrin vs Mixed OCP		0.20 (-2.04 to 2.45)	0.86
" vs Mixed pesticides  Mixed OCP vs Mixed pesticides  ———		0.43 (-2.00 to 2.86)	0.73
Overall pesticides vs overall contaminants	<u> </u>	0.22 (-0.81 to 1.25) 0.13 (-0.06 to 0.31)	0.67
	1-		0.20
-1.5	0 1.5	i	

## The risk of MetS following pesticides exposure

	n studies	Odds Ratio
	(subgroups)	(95% CI)
Pesticides		
Organochlorine pesticides		
Hexachlorocyclohexane	6 (28)	
Hexachlorobenzene	5 (28)	
Dicholordiphenyldichloroethylene	7 (33)	<del>_</del>
Dichlorodiphenyltrichloroethane	5 (24)	
Oxychlordane	4 (26)	
Transnonchlor	5 (28)	<del></del>
Overall organochlorine	11 (216)	<b>-</b> ◇ <b>-</b>
Organophospates	` ′	
Carbamates		
Pyrethroids		
Triazines		
Overall pesticides	11 (222)	<del></del>
Contaminants	8 (108)	<b>—</b>
Pesticides + Contaminants	11 (330)	



0.5

1.5