

# Food limitation of seabirds in the Benguela ecosystem and management of their prey base

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1	Food limitation of seabirds in the Benguela ecosystem and management of their prey base
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20	Abstract
21	Four of seven seabirds that are endemic to the Benguela ecosystem (African Penguin Spheniscus
22	demersus, Cape Gannet Morus capensis, Cape Cormorant Phalacrocorax capensis, Bank Cormorant P.
23	neglectus) compete with fisheries for prey and have an IUCN classification of Endangered. Prey
24	depletion and food resource limitations have been major drivers of recent large population decreases of
25	each of these species. As populations decrease, colony sizes also dwindle rendering them susceptible to
26	Allee effects and higher probabilities of extinction. Therefore, it is necessary to maintain colonies at sizes
27	that minimise their probability of extinction. Means to ensure an adequate availability of food to
28	achieve this goal include closing important seabird foraging areas (often adjacent to key colonies) to
29	relevant fishing, implementing ecosystem thresholds below which such fishing is disallowed (which are
30	also expected to benefit forage resources) and, should there be an altered distribution of prey,
31	attempting to establish seabird colonies close to the new location of forage resources.
32	

### **1. Introduction**

This paper provides a brief review of food limitation of four Endangered (EN) seabirds that breed in the Benguela ecosystem, of the viability of seabird colonies and Allee effects, of forage and consumption thresholds and of potential management interventions that may enhance food availability. The locations of some of the system's important seabird colonies are shown in Figure 1.

38

#### 39 **2. Status of the Benguela's seabirds**

Fifteen seabird species breed in the cool Benguela ecosystem, which extends from about Benguela in 40 41 Angola to Woody Cape in South Africa, being bounded in the north and southeast by the warm Angola 42 and Agulhas Currents, respectively (Figure 1). Seven of these are endemic to the ecosystem (Hockey et 43 al. 2005). Of the endemic species, in terms of criteria of the International Union for Conservation of 44 Nature (IUCN) four are EN (African Penguin Spheniscus demersus, Cape Gannet Morus capensis, Cape 45 Cormorant Phalacrocorax capensis, Bank Cormorant P. neglectus), one is Vulnerable (VU, Damara Tern 46 Sternula balaenarum), one is Near-Threatened (NT, Crowned Cormorant Microcarbo coronatus) and 47 Hartlaub's Gull Larus hartlaubii is Least Concern (LC, IUCN 2020). The non-endemic Leach's Storm Petrel 48 Hydrobates leucorous (or Oceanodroma leucorhoa) is VU, whereas the other seven non-endemics are LC 49 (IUCN 2020). Although substantial numbers of Leach's Storm Petrel migrate from the Northern 50 Hemisphere into the Benguela system (Pollet et al. 2019), the breeding population of this species in 51 South Africa is Critically Endangered (CR, Taylor et al. 2015). In South Africa small, isolated populations 52 of Great White Pelican Pelecanus onocrotalus, Caspian Tern Hydroprogne caspia and Roseate Tern 53 Sterna dougallii are regarded as regionally VU (Taylor et al. 2015). Leach's Storm Petrel and Roseate 54 Tern do not breed In Namibia, where the Cape Gannet is considered CR (Kemper 2015), the Great White 55 Pelican and Caspian Tern VU (Simmons 2015a, b) and the Damara Tern NT (Simmons et al. 2015).

56

#### 57 **3. Importance and evidence of food limitation**

58 For a seabird population to be self-sustaining, mortality from its mature component must be 59 balanced by recruitment or immigration. Recruitment is influenced by age at first breeding, breeding participation (proportion of mature birds electing to breed), breeding frequency (number of clutches 60 61 produced annually), breeding success (chicks fledged per clutch) and survival from fledging to breeding. 62 Insufficient food resources may negatively affect these demographic parameters leading to population 63 decreases, as is described below for the Benguela's four EN seabird species, which all feed mainly on three forage resources that are heavily exploited by fisheries when these are readily available. African 64 65 Penguin, Cape Gannet and Cape Cormorant feed preferentially on sardine Sardinops sagax and anchovy Engraulis encrasicolus (Hockey et al. 2005). In South Africa, the intensively fished West Coast rock 66 67 lobster Jasus lalandii is an important prey of Bank Cormorant (Dyer et al. 2019), whereas in Namibia the lightly-exploited but nutritionally-deficient bearded goby Sufflogobius birarbatus is one of its staple 68 69 foods (Ludynia et al. 2010a). After the collapse of sardine in Namibia in the 1970s, bearded goby also 70 replaced it as the dominant prey item for African Penguins breeding in Namibia (Ludynia et al. 2010b).

71

#### 72 African Penguin

Large decreases of African Penguins at colonies between Lüderitz in Namibia and Dassen Island off west South Africa from the 1960s to the 1980s were recorded following the collapse of sardine stocks off Namibia and South Africa. North of Lüderitz, increases at Mercury and Ichaboe islands after the 1960s probably resulted from an increased local abundance of bearded goby (Crawford and Shelton 1978, Shelton et al. 1984, Crawford et al. 1985, 2001a). However, this did not prevent an overall decrease of 90% in penguins breeding in Namibia between 1956 and the early 2000s (Figure 2, Crawford 2007).

80 By contrast, an increase between the 1950s and 1970s and a later decrease in numbers of penguins 81 at Dyer Island, southwest South Africa, coincided with an increase and subsequent fluctuations in

82 anchovy abundance (Crawford 1998). From 1989–2009 numbers of penguins breeding along west South 83 Africa were significantly correlated with the overall biomass of sardine and anchovy in South Africa, as also was the case from 1999–2009 in Algoa Bay (Figure 2, Crawford et al. 2011). The formation of new 84 85 penguin colonies at Stony Point and Boulders on South Africa's mainland and at Robben Island, and 86 increases at these colonies and at Dassen Island, took place during a period of recovery of South Africa's 87 sardine in the 1980s and 1990s and an increase in the biomass of anchovy at the start of the 21<sup>st</sup> 88 century. This led to an overall increase in the numbers of penguins breeding in west South Africa 89 (Underhill et al. 2006).

90 After 2004, large decreases in numbers of penguins breeding in west South Africa were associated 91 with a shift to the south and east of adult anchovy and sardine and a collapse of sardine (Crawford et al. 92 2011). Concomitant with the altered distribution of prey there were decreases in penguins at all South 93 African colonies north of Cape Town (the northernmost colony at Lambert's Bay became extinct in 2006; 94 Crawford et al. 2013). By contrast, in the south numbers increased or stabilized at Boulders and Stony 95 Point and penguins attempted to form a new colony at De Hoop Nature Reserve (Underhill et al. 2006, 96 Crawford et al. 2011). Although trends in numbers of African Penguins at colonies are often regionally 97 coherent (Underhill et al. 2006), at Dyer Island, also in the south, penguins decreased in the 2000s. This 98 may have resulted from heavy exploitation of sardine in its vicinity: 70,000 tonnes of sardine were 99 caught within 30 km of Dyer Island in 2004 (Ludynia et al. 2014) and the exploitation rate of sardine 100 west of Cape Agulhas rose to 30-44% in 2002 and from 2005-2007 (Coetzee et al. 2008). When the 101 colony size at Dyer Island was > 3,500 pairs, numbers of penguins breeding there were negatively 102 related to sardine catches made within 20 nautical miles of the island (Ludynia et al. 2014), suggesting 103 an impact of catches on local food availability.

Local or regional abundances or availability of sardine and anchovy have been related to numbers of African Penguins breeding (Crawford 2007, Crawford et al. 2011, Sherley et al. 2013), to adult and

106 immature survival (Crawford et al. 2011, 2019, Sherley et al. 2014, Robinson et al. 2015), to age at first 107 breeding (Whittington et al. 2005), to breeding participation (Crawford et al. 1999b, Durant et al. 2010), 108 to breeding success (Adams et al. 1992, Crawford and Dyer 1995, Crawford et al. 1999b, 2006a, Cury et 109 al. 2011, Sherley et al. 2013) and to foraging performance (Pichegru et al. 2010, Campbell et al. 2019, 110 McInnes et al. 2019). Additionally, onset of breeding was probably delayed during a period of food 111 scarcity (Crawford and Dyer 1995) and the chick fledging period was negatively correlated with prey 112 abundance (Sherley et al. 2013). Delayed breeding and longer fledging periods may reduce the breeding 113 frequency and cause abandonment of unfledged chicks when parents are constrained to leave colonies 114 for pre-moult fattening (Crawford et al. 2018).

115

#### 116 Cape Gannet

In Namibia, the Cape Gannet population decreased by 95% after the collapse of that country's sardine stock in the 1970s (Figure 2, Crawford 2007). However, at the same time the numbers of gannets in South Africa increased. Trends in the proportions of forage fish (sardine and anchovy) and gannets found in Namibia and in South Africa were similar (Figure 3) suggesting an impact of food. However, increases of gannets in South Africa did not offset losses in Namibia so that the overall abundance of Cape Gannets deceased (Crawford et al. 2007a).

123 In the 21<sup>st</sup> century, the movement of Cape Gannets to the south and east continued with a decrease 124 in numbers breeding off west South Africa and an increase in Algoa Bay, matching a shift in the 125 distribution of their prey (Crawford et al. 2015, Sherley et al. 2019). In the west, foraging effort of Cape 126 Gannets increased and nest attendance decreased with reduced consumption of sardine and anchovy, 127 and adult body condition was negatively impacted by increases in the dietary proportion of hake 128 *Merluccius* sp. discarded by fisheries (Cohen et al. 2014). At Malgas Island, Cape Gannets had 129 substantially reduced breeding success when they fed their chicks mainly low-quality fish discards (Grémillet et al. 2008). Between 2011 and 2014 there was a net deficit in energy expenditure during
foraging for most Cape Gannets tracked from Malgas Island (Grémillet et al. 2016).

132

#### 133 Cape Cormorant

134 In Namibia, numbers of Cape Cormorants decreased by 62% after the collapse of sardine in the 1970s 135 (Figure 2). It was a lesser decrease and occurred later than those of African Penguins and Cape Gannets, 136 possibly owing to Cape Cormorants benefitting from increased breeding space at Ichaboe Island and 137 making use of alternative prey in the form of bearded goby (Crawford 2007, Crawford et al. 2007b, 138 Erasmus et al. 2021). Additionally artificial nesting platforms constructed in central Namibia provided 139 breeding habitat for Cape Cormorants closer to the reduced range of the sardine (Crawford 2007). By 140 contrast, gannets and most penguins in Namibia bred in the south and suffered food shortages as the 141 sardine's range contracted northwards (Crawford 2007). Cape Cormorants colonised Ilha dos Tigres in 142 southern Angola in the 2000s and > 16,000 pairs bred there in 2017 (Mendelsohn and Haraes 2018).

143 Unlike in Namibia, numbers of Cape Cormorants breeding in west South Africa decreased in the mid-144 1990s prior to recent decreases in that region of African Penguins and Cape Gannets (Crawford et al. 145 2007b). There were substantial mortalities of adult Cape Cormorants at several colonies in west South 146 Africa between 1991 and 2006 caused by avian cholera Pasteurella multocida, outbreaks of which may 147 be triggered by stressful events such as food limitation (Crawford et al. 1992a, Ward and Williams 2004, 148 Waller and Underhill 2007). Its initial occurrence at Dassen Island may have been precipitated by a 149 scarcity of forage fish during 1989 and 1990 (Crawford et al. 1992a). Conforming to the recent shift in 150 the distribution of sardine and anchovy, Cape Cormorants decreased off northwest South Africa but 151 remained stable in the south, where new colonies formed (including at Robben Island and Stony Point) 152 and some others increased (Crawford et al. 2016).

153 When food is scarce, as well as experiencing increases in adult mortality, Cape Cormorants, like some 154 other cormorant species, often skip or abandon breeding or delay their first breeding attempt (Berry 155 1976, Crawford et al. 2001b, 2016, Crawford and Dyer 1995). They may also suffer extensive breeding 156 failure (Duffy et al. 1984, Crawford et al. 1992b). In November 2008, Cape Cormorants brooding small 157 chicks expended significantly greater foraging effort at Malgas and Dassen islands than at Dyer Island 158 farther south (Hamann et al. 2012). During May and June 2014 Cape Cormorants north of Cape Town 159 attempted to take bait from hand-lines used to fish for snoek *Thyrsites atun*, a behaviour not previously 160 observed and that suggested poor food availability at that time (Crawford et al. 2016).

161

#### 162 Bank Cormorant

In Namibia, Bank Cormorants decreased from 7,166 pairs in 1978–1980 to 3,735 pairs in 1995–1997; 163 164 a severe reduction of bearded goby in central Namibia in 1994 was a probable cause of the decrease 165 (Crawford et al. 1999a). In South Africa, Bank Cormorants decreased from 1,506 pairs in 1978–1980 to 166 846 pairs in 2010–2013 (Crawford et al. 2015). Total numbers breeding at 11 readily accessible localities 167 between Lambert's Bay and Dyer Island fell from above 500 pairs during 1978–1987 to 350 pairs from 168 1995 to 2006. The most northern colony (Lambert's Bay) was extinct by 1999 and there were substantial 169 decreases at the two largest colonies, Malgas and Dassen islands, which are north of Cape Town. There 170 was fluctuating growth at Robben Island, near Cape Town, where birds were oiled by spills in 1994 171 (Apollo Sea) and 2001 (Treasure). Farther east numbers at two colonies increased. These trends are 172 consistent with a reduced abundance of rock lobsters in the north and an expansion of this resource to 173 the east (Crawford et al. 2008). As with other endangered species, fishing in the immediate vicinity of 174 Malgas and Dassen islands may have reduced densities of lobsters below levels needed to sustain the 175 larger colonies of Bank Cormorants. Where rock lobsters had not been severely depleted, Bank 176 Cormorants showed a positive response to their local availability and modelling suggested that areas

with no-take of lobsters at 20–30 km around Bank Cormorant colonies would benefit the conservation
of this seabird (Sherley et al. 2017).

179

#### 180 4. Colony viability and Allee effects

Traditionally, the Benguela's four Endangered seabirds bred in large colonies and those three that feed mostly on sardine and anchovy have often foraged in large multi-species flocks (Hockey et al. 2005). However, as populations have fallen, breeding colonies have decreased in size and on occasion fragmented into smaller units (e.g. Cordes et al. 1999, Crawford et al. 2016, Sherley et al. 2019).

185 The probabilities of extinctions of different-sized colonies of African Penguins over 40 years were 186 obtained from empirical observations on the performance of 41 discrete colonies in southern Africa 187 from 1956–1996 (Crawford et al. 2001a). Only one (< 4 %) of 28 colonies that in 1956 had  $\leq$  50 pairs was 188 extant in 1996, compared to 26% of those having 251-1,000 pairs, 67% of those having 1,001-5,000 189 pairs and 100% of those with > 5,000 pairs (Figure 4). Whereas three of the 30 breeding colonies extant 190 in 2004 held > 5,000 pairs and, in terms of these probabilities, had no likelihood of extinction within 40 191 years, by 2019 none of the 25 remaining colonies had > 5,000 pairs so that all had some chance of 192 extinction during such a period. In 2019, fifteen colonies had < 250 pairs and hence a 96% chance of 193 extinction by 2059 (Figure 4). In 2019, only seven colonies held > 1,000 pairs and had a 67% probability 194 of surviving to 2059: Mercury Island in Namibia; Dassen, Robben, Dyer, St Croix and Bird (Algoa Bay) 195 islands and Stony Point in South Africa.

Fragmented seabird populations may suffer from Allee effects, i.e. reduced fitness at low population size increasing their likelihood of extinction (Ryan et al. 2012). African Penguins that feed in groups have a higher catch of prey per unit effort than solitary birds (McInnes et al. 2017), but colonies may become too small for sufficient foraging groups to form (Ryan et al. 2012). For example, in 2012 more penguins were documented foraging in a single flock of at least 158 birds (Ryan et al. 2012) than were breeding in

201 2019 at 10 of the 25 presumed extant African Penguin colonies. For plunge-diving Cape Gannets, highly-202 synchronised prey pursuits, which are likely contingent on their ability to locate an adequate number of 203 conspecifics at sea, benefitted from improved foraging efficiency when feeding on sardine schools 204 (Thiebault et al. 2016). A lower number of individuals could also decrease the role of colonies as 205 information centres, or the possibility to use conspecifics as cues to locate prey (e.g. Thiebault et al. 206 2016). Again smaller colonies have a larger proportion of birds nesting at colony edges, where eggs and 207 chicks are at a greater risk to predation (e.g. Cordes et al. 1999), and sex-biased mortality (e.g. Pichegru 208 and Parsons 2014) may decrease productivity of monogamous breeders at small colonies. The higher 209 risks of extinction for small colonies make it imperative to maintain colony sizes above thresholds at 210 which Allee effects are manifest. Furthermore, small numbers of seabirds that have a late age at 211 maturity may take substantial periods to recover, even with fishery closures (e.g. Robinson et al. 2015).

212 In 2010, Mercury Island held 72% of the global population of Bank Cormorants (Roux and Kemper 213 2015) and 73% of Namibia's African Penguins (Crawford et al. 2013). In 2018, c. 70% of the overall Cape 214 Gannet population was at Bird Island, Algoa Bay at the eastern boundary of the Benguela ecosystem 215 (Sherley et al. 2019). In the 2010s, 54% of South Africa's Cape Cormorants bred at Dyer Island (Crawford 216 et al. 2016). Such congregations of large proportions of a species at a single locality may offset Allee 217 effects but render the species highly susceptible to local catastrophic events such as oil spills (e.g. 218 Crawford et al. 2000), disease outbreaks (e.g. Waller and Underhill 2007), severe weather events (e.g. 219 Randall et al. 1986) and reduced availability of resources.

220

#### 221 **5. Forage thresholds**

#### 222 Forage abundance thresholds

In a meta-analysis of long-term studies in seven marine ecosystems from around the globe, Cury etal. (2011) demonstrated key thresholds in prey (fish and krill) abundance, below which seabirds

dependent on that food type suffered reduced and more variable breeding success. The threshold was equal to the long-term average biomass of prey and equivalent to one-third of the maximum observed prey biomass. Similar relationships were established at the species level, namely for 13 of the 14 seabirds studied, including African Penguin and Cape Gannet from the Benguela ecosystem. The socalled "1/3 for the birds" rule provides a benchmark for management of forage fish fisheries at a level that would sustain seabird productivity over the long-term (Cury et al. 2011).

231 In a related study, Robinson et al. (2015) showed that annual mortality of adult African Penguins at 232 Robben Island increased markedly when the biomass of sardine aged 1 y or older off west South Africa 233 fell below 25% of its maximum value. The lower forage threshold observed for adult mortality (25%) 234 than for breeding success (33%) confirmed the theoretical prediction of Cairns (1987) that decreases in 235 forage abundance would influence seabird breeding success ahead of survival. When data for African 236 Penguins at Dassen and Robben islands were grouped, a generalised additive model (GAM) showed that 237 adult survival decreased when standardised (i.e. centred and scaled) sardine abundance off west South 238 Africa fell below a threshold of zero (Figure 5).

It is noteworthy that off west South Africa, Cape Gannets were better able to buffer effects of reduced prey abundance than African Penguins (Distiller et al. 2012, Sherley et al. 2014, Robinson et al. 2015). After an eastward displacement of the main forage resources and the collapse of sardine (Coetzee et al. 2008), adult gannets, due to their movement capabilities, were able to feed on alternative food, such as saury *Scomberesox saury* and hake offal (Grémillet et al. 2008, 2019, Crawford et al. 2014, 2019). Consequently, their survival was unrelated to standardised estimates of the overall abundance of sardine and anchovy (Figure 5).

246

#### 247 Forage consumption thresholds

248 Saraux et al. (2021) investigated thresholds for the consumption of forage fish stocks by seabirds in 249 five marine ecosystems: off Norway, South Africa, Peru, in the Baltic Sea (Sweden) and at Shetland 250 (Scotland). In each of these systems, the predation pressure, estimated as the proportion of a forage fish 251 stock consumed by seabirds, was generally small (median = 1% of biomass) but increased sharply when 252 prey biomass decreased below a threshold of 15 to 18% of its maximum recorded value. A threshold of 253 18% was considered a limit not to be reached in order to avoid a sharp rise in the natural mortality rate 254 of forage fish resources, and below which extra cautious management of fisheries would be required 255 (Saraux et al. 2021). Similarly, Essington et al. (2015) advised that a risk-based management scheme that 256 reduces fishing when populations become scarce would protect both forage fish and their predators 257 from collapse, with little effect on long-term average catches.

258

#### 259 6. Other Forage Considerations

#### 260 Forage quality

261 Condition of prey may influence its ability to sustain predator populations in addition to its 262 abundance (e.g. Grémillet et al. 2008). For example, poor food quality likely caused unprecedented 263 breeding failures at many seabird colonies on the east coast of Great Britain in 2004 (Wanless et al. 264 2005), decreased mass at fledging of Common Guillemots (Murres) Uria aalge in the Baltic Sea and 265 brought about breeding failures and mass mortality of this species in the Northeast Pacific in 2015–2016 266 (Osterblom et al. 2006, Piatt et al. 2020). In South Africa, the condition of sardine deteriorated in the 267 2000s (Ndjaula et al. 2013), possibly contributing to some of the seabird population decreases. As also 268 mentioned above, reliance on fish offal as a food source reduced Cape Gannet chick survival (Grémillet 269 et al. 2008).

270

#### 271 Fisheries mortality

272 In addition to contributing to collapses of forage resources (e.g. Essington et al. 2015), fishing may 273 cause localised depletion of prey; as such, the fisheries mortality rate (F) can be employed as another 274 management threshold. For example, the performance of three species of *Pyqoscelis* penguins breeding 275 near the Antarctic Peninsula was reduced when local harvest rates of Antarctic krill Euphausia superba 276 were  $\geq$  10% of the estimated biomass (Watters et al. 2020). By comparison, the exploitation rate of 277 South Africa's sardine west of Cape Agulhas increased substantially after 1999 and reached 44% in 2006 278 (Coetzee et al. 2008). Spatially explicit thresholds in fishing mortality need to be carefully defined to 279 assist in management of fish stocks so as to support viable seabird populations in the Benguela system.

280

#### 281 Forage availability

282 It is not only the absolute abundance of prey but also its availability that impacts energy acquisition, 283 demography and population trends of seabirds. For example, Crawford et al. (2019) used long-term 284 information on the diet of Cape Gannets off west South Africa to derive an index of the availability of 285 anchovy and sardine per unit biomass to seabirds in that region, terming it a Forage Availability Index 286 (FAI). The study demonstrated a substantial decrease in the availability of anchovy and sardine to 287 seabirds in the 2000s compared to the 1990s. Breeding numbers of Cape Gannets and Cape Cormorants 288 and survival of adult African Penguins at both Dassen and Robben islands were significantly related to 289 the FAI. For gannets, the numbers breeding increased when the FAI, which increases as anchovy and 290 sardine contribute more to the diet, was > c. -1 and for cormorants when it was > c. 1. The higher 291 threshold for Cape Cormorants than Cape Gannets at which numbers breeding were predicted to 292 increase probably resulted from the fact that, unlike Cape Gannets but similarly to African Penguins, 293 Cape Cormorants are mostly unable to access alternative food such as saury and hake offal (Crawford et 294 al. 2019). Notably, frequent aggregations of anchovy close to the seabed were thought to have reduced

their availability to purse-seiners, and were suggested as a reason for a large under-catch of the total
allowable catch (TAC) of anchovy in recent years (DAFF 2016).

The importance of prey depth was highlighted in a modelling study from the Humboldt upwelling system off western South America, where it primarily determined foraging success of Peruvian Boobies *Sula variegata* and Guanay Cormorants *P. bougainvilliorum*, which feed mainly on Peruvian anchoveta *Engraulis ringens* (Boyd et al. 2017). Peruvian Boobies and Guanay Cormorants are the ecological equivalents in that system of Cape Gannets and Cape Cormorants in the Benguela system (Crawford et al. 2006b).

African Penguins and other diving piscivores may herd prey shoals upwards, thereby facilitating their availability to surface-feeding predators (e.g. McInnes and Pistorius 2019). Hence their loss will influence ecosystem functioning.

306

#### **7. Possible interventions to enhance food availability**

308 Various international agreements require participating parties to implement an ecosystem-based 309 approach to fisheries management (or ecosystem approach to fisheries, EAF). For example, the Code of 310 Conduct for Responsible Fisheries of the United Nations' Food and Agriculture Organisation, adopted in 311 1995, stated that Management measures should not only ensure the conservation of target species but 312 also of species belonging to the same ecosystem or associated with or dependent upon the target species 313 (http://www.fao.org/3/v9878e/v9878e00.htm). The sixth Aichi Target of the Convention on Biological 314 Diversity's Strategic Plan for Biodiversity 2011–2020 reads: By 2020 all fish and invertebrate stocks 315 and aquatic plants are managed and harvested sustainably, legally and applying ecosystem based 316 approaches, so that overfishing is avoided, recovery plans and measures are in place for all 317 depleted species, fisheries have no significant adverse impacts on threatened species and 318 vulnerable ecosystems and the impacts of fisheries on stocks, species and ecosystems are within

safe ecological limits (<u>https://www.cbd.int/convention/text/default.shtml</u>). In South Africa, the Marine
 Living Resources Act of 1988 attempts to provide legislative support for the EAF but is in urgent need of
 updating and further amendment better to achieve this (Cochrane et al. 2015).

322 One tool available to manage fisheries so as to avoid harmful impacts on associated or dependent 323 species (ADS) is marine spatial planning, for example the implementation of fishing closures around 324 important seabird colonies (e.g. Ludynia et al. 2012, Sherley et al. 2015, 2018). In northeast Scotland, 325 Black-legged Kittiwakes Rissa tridactyla benefitted from closure of fishing for sandeels Ammodytes 326 marinus, an important prey for kittiwakes, around breeding colonies (Daunt et al. 2008, Furness 2020). 327 In the Californian upwelling ecosystem off west North America, a precautionary ecosystem approach 328 that avoids localised depletion of forage resources (including anchovy and sardine) near predator 329 concentrations is deemed necessary. In 1983, nearshore (within 3 to 6 km of the coast) fishing for 330 anchovy was prohibited, the main objectives being to protect pre-recruit anchovy, live bait and 331 recreational fisheries, and predator forage (Pacific Fisheries Management Council 1983, Thayer and 332 Sydeman 2020). At the time, the anchovy-dependent Brown Pelican Pelecanus occidentalis was listed 333 as Federally Endangered in USA (Anderson et al. 1982). In the Humboldt upwelling ecosystem off west South America, fishing for Peruvian anchoveta close to a Peruvian Booby colony increased the birds' 334 335 foraging effort. The more the fishery reduced the quantity of anchoveta in the area, the farther the 336 breeding boobies needed to forage from the colony to find food (Bertrand et al. 2012). In the Benguela 337 system, in which similarly to the California and Humboldt systems sardine and anchovy are key forage 338 resources, short-term fishery closures around African Penguin colonies, even though not matched to the 339 late age at recruitment and high mate and site fidelity of the penguins (Crawford et al. 2013), have 340 increased breeding productivity and deceased parental foraging costs (Pichegru et al. 2010, Sherley et al. 341 2015, 2018, Sydeman et al. 2021).

342 A second means to account for the foraging requirements of marine predators is the identification 343 and implementation of ecosystem thresholds, below which fishing would be prohibited. Ecosystem 344 thresholds have already been widely applied to minimise by-catch mortality of ADS in fisheries (e.g. 345 Rollinson et al. 2017, Da Rocha et al. 2021) and are increasingly proposed as means to ensure sufficient 346 quantities of prey for predators, such as seabirds (e.g. Cury et al. 2011, Smith et al. 2011, Sydeman et al. 347 2017, Hill et al. 2020). In the Californian upwelling ecosystem a harvest rule disallows fishing of sardine 348 when its biomass is below 150,000 metric tonnes (Pacific Fisheries Management Council 2003, Thayer 349 and Sydeman 2020).

A third mechanism to achieve adequate food for seabirds is to offset mismatches in the distributions of the breeding localities and the prey of seabirds through establishment of colonies nearer to the food supply. Guano platforms in Namibia served this purpose for Cape Cormorants (Crawford 2007) and an attempt to establish a new African Penguin colony in South Africa closer to the present distributions of sardine and anchovy has been initiated (BirdLife South Africa 2016).

355

#### 356 *Recommendations from the 2020* AEWA Benguela Current Forage Fish Workshop

In November 2020, a *Benguela Current Forage Fish (BCFF) Workshop*, organized by the African-Eurasian Migratory Waterbird Agreement (AEWA), in collaboration with the Benguela Current Convention (BCC) and BirdLife South Africa, recommended actions to be undertaken as a matter of urgency under the auspices of BCC, AEWA and the AEWA Benguela Coastal Seabird International Working Group, as well as by the national governments of Angola, Namibia and South Africa. These included:

developing tools to increase the availability of sufficient forage [fish] for threatened endemic
 Benguela seabird species, such as setting ecosystem thresholds (i.e. sizes of forage resource
 populations below which a range of precautionary measures relating to fishing would be

implemented at various spatial scales) and closing key foraging areas to fishing, adjacent to major
seabird colonies;

ensuring the existence or creation of suitable seabird breeding habitat within the contracted or
 altered distributions of forage fish species to partially alleviate the impact of an altered
 distribution of prey on affected seabird species;

and facilitating and prioritising the recovery of seabird colonies to sufficient sizes to minimise
 known and potential Allee effects, thus reducing the probability of colony extinctions (AEWA
 2020).

374

#### 375 **8. Conclusion**

In conclusion, it is apparent that food limitation has been a major driver of adverse trends in the population sizes and conservation status of four Benguela seabirds that compete with fisheries for food. Hence it will be important to manage the impact of fisheries on the availability of prey to these seabirds, including through the use of measures that were successfully applied elsewhere or that have empirical support for use in the Benguela ecosystem.

381

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387

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650 Figure 1. A map of the Benguela Current ecosystem showing locations of some of the region's important





С 1,0 Proportion of maximum 0,8 0,6 0,4 0,2 0,0 1989 19911993 1995 1999 1997 Sardine & anchovy 

652

656 Figure 2. Comparison of trends in standardised estimates (maximum = 1) of the combined biomass of 657 sardine and anchovy and breeding populations of seabirds that feed on these forage species for (a) 658 Namibia 1956/60-2001/05 (redrawn from information in Crawford 2007), (b) west South Africa 1989-659 2010 and (c) Algoa Bay 1999–2009 (redrawn from information in Crawford et al. 2011). For west South 660 Africa and Algoa Bay fish estimates are for the whole of South Africa.

661

655



1984/85 1988/89 1992/93 1996/97 2000/01 2004/05

Cape gannet

1972/73 1976/77 1980/81

Sardine & anchovy

0,8

0,6

0,2

0,0

1960/61 1964/65 1968/69

1956/57

Proportion 0,4

662



- **Figure 3.** Trends in the proportions of the Benguela ecosystem's sardine and anchovy and Cape Gannets
- found in a) Namibia and b) South Africa, 1956/1957–2005/2006 (redrawn from information in Crawford
- 667 et al 2007a).







679 Figure 5. Estimated adult survival off west South Africa of (top) African Penguins at Dassen and Robben islands and (bottom) Cape Gannets at Lambert's Bay and Malgas Island compared with the standardised 680 681 abundance of prey off west South Africa (i.e. centred and scaled by dividing by the standard deviation). 682 Estimates of survival were from Sherley et al. (2014) for African Penguins and Distiller et al. (2012) for 683 Cape Gannets. Estimates of prey abundance were updated from Cury et al. (2011). The prediction (and 684 95% confidence interval) of the GAM is indicated with the black curve (and grey polygon) for African 685 Penguins. The vertical dashed line represents the threshold below which adult survival decreased 686 markedly with decreasing sardine abundance. For Cape Gannets, no significant relationship was found 687 between adult survival and prey (sum of anchovy and sardine) abundance.