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Testing various food-industry wastes for electricity production in microbial fuel cell Bibiana Cercado-Quezada, Marie-Line Delia, Alain Bergel*

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ABSTRACT

Three food-industry wastes: fermented apple juice (FAJ), wine lees and yogurt waste (YW) were evaluated in combination with two sources of inoculum, anaerobic sludge and garden compost, to produce electricity in microbial fuel cells. Preliminary potentiostatic studies suggested that YW was the best candidate, able to provide up to 250 mA/m² at poised potential +0.3 V/SCE. Experiments conducted with two-chamber MFCs confirmed that wine lees were definitely not suitable. FAJ was not able to start an MFC by means of its endogenous microflora, while YW was.

Both FAJ and YW were suitable fuels when anaerobic sludge or compost leachate was used as inoculum source. Sludge-MFCs had better performance using YW (54 mW/m² at 232 mA/m²). In contrast, compost-leachate MFCs showed higher power density with FAJ (78 mW/m² at 209 mA/m²) than with YW (37 mW/m² at 144 mA/m²) but YW gave more stable production. Under optimized operating conditions, compost-leachate MFCs fueled with YW gave up to 92 mW/m² at 404 mA/m² and 44 mW/m² in stable conditions.

1. Introduction

Microbial fuel cells (MFCs) are electrochemical devices that use the metabolic activity of microorganisms to oxidize fuels, generating electric current by direct or mediated electron transfer to electrodes (Rabaey et al., 2007; Schröder, 2007). In the anodic compartment, organic matter is oxidized by microbial metabolism, which transfers the electrons to the anode. In the cathodic compartment, oxygen or oxidized compounds are reduced either via an abiotic process or by microbially mediated reduction (He and Angenent, 2006).

Early studies on fuels in MFCs dealt with low molecular weight substrates, carbohydrates like glucose, fructose, xylose, sucrose, maltose and trehalose (Chauduri and Lovley, 2003; Kim et al., 2000), organic acids like acetate, propionate, butyrate, lactate, succinate and malate (Bond and Lovley, 2005; Holmes et al., 2004; Min and Logan, 2004), alcohols like ethanol and methanol (Kim et al., 2007) and inorganic compounds like sulfate (Rabaey et al., 2006). Later, the interest in complex substrates led to tests on starch, cellulose, dextran, molasses, chitin and pectin (Melhuish et al., 2006; Niessen et al., 2005, 2006; Rezaei et al., 2007). In parallel, the use of domestic wastewater as fuel and microbial source was largely reported (Liu et al., 2004; Rabaey et al., 2005). Research on domestic wastewater was extended to a large variety of industrial wastewaters, e.g. from starch (Gil et al., 2003) and wastewaters coming from the meat packing industry (Heilmann and Logan, 2006), swine farms (Min et al., 2005) and cereal- (Oh and Logan, 2005) and potato-producing units (Rabaey et al., 2005). Solid agricultural wastes such as corn stover (Zuo et al., 2006) and manure (Scott and Murano, 2007) have also been tested as fuel after being pretreated. Numerous variants of the experimental conditions affected the results but the power density obtained in these studies was generally in the range of 1–3600 mW/m², with most values lying between 10 and 1000 mW/m².

The choice of the inoculum source is a key parameter in MFC design. The most common sources of electroactive microorganisms have been domestic wastewater, activated and anaerobic sludge and marine sediments (Erable et al., 2009; Kim et al., 2007; Liu et al., 2004; Rezaei et al., 2007). Alternative sources have been reported: heat treated soils (Niessen et al., 2006), garden compost (Parot et al., 2008a), manure (Scott and Murano, 2007) and rumen (Rismani-Yazid et al., 2007). However, very little research has investigated electroactive native microflora in agro-industrial wastes are rich in mixed populations.

The present work tested three kinds of waste coming from different sectors of the agricultural and food industries: fermented apple juice (FAJ), wine lees (WL) and yogurt wastewater (YW). Each waste was tested for its possible capacity to form electroactive biofilms from its endogenous flora. The same wastes were then tested as fuels to feed bioanodes that were formed with two different sources of inoculum: anaerobic sludge and garden compost. Anaerobic sludge (actually aerobic sludge driven to anaerobic conditions during an acclimating period) was chosen because of its common use as MFC inoculum. Bioanodes formed in garden

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compost have shown efficient electroactive properties with acetate (Parot et al., 2008b) and have revealed a rich microbial flora with remarkable electroactive capabilities (Parot et al., 2009). It was consequently expected that such bioanodes might be appropriate for a large variety of different substrates.

2. Methods

2.1. Waste media

To simulate a leachate of apple farm wastes, fermented apple juice (FAJ) was prepared by pressing Braeburn 65/75 apples; the juice was fermented in a sealed container at room temperature for at least 15 days. Flasks of 500 mL were filled and stored at 3 °C. Lees coming from the wine making process (WL) were obtained from a local wine grower. Yogurt waste (YW) was collected from the yogurt production line of a local dairy. Samples of both substrates were stored in 1.5 L plastic bottles at 3 °C or were deep-frozen for long-term storage.

When indicated, the media were diluted (1:20) with the buffer solution described below. pH and conductivity were systematically measured using a Schott Gerate pH-meter CG 882 and a Metrohm 660 conductimeter (Swiss). Total chemical oxygen demand (COD) was measured in anolytes and raw substrates following the standard method 8000 for 0–1500 mg COD/L (HACH France S.A.S.). Characteristics of the raw media and after dilution are reported in Table 1.

2.2. Inoculum sources and acclimation

Aerobic sludge was obtained from an urban wastewater treatment plant. Sludge was distributed in fully filled, closed 0.5 L bottles and acclimated to the different waste media by daily addition of 1 mL of raw substrate with a syringe placed in the screw cap. Substrate was added daily for 16 days and the whole acclimatizing period was one month. During the acclimating period, the sludge evolved rapidly to anaerobic conditions; for this reason we refer to anaerobic sludge in what follows.

The microbial flora was extracted from the garden compost by lixiviation. 10 mM NaCl solution was added to 1 L of compost to give a final volume of 3 L. This mixture was placed in 2 or 3 L Erlenmeyer flasks, stirred for 24 h and then filtered through felt cloth. The filtrate was placed in 0.5 L fully filled flasks and acclimated by addition of 1 mL waste media twice a week for 30 days; it was used one month after the last addition.

2.3. Electrochemical tests

The experimental unit was a cap-sealed three-electrode cell of 50 mL, or 500 mL when indicated, with graphite felt $(2 \times 2 \times 0.5 \text{ cm}, \text{RVG}, \text{ Carbone Loraine}, \text{ France})$ as the working electrode, an Ag/AgCl or an Standard Calomel Electrode (SCE) reference electrode and a platinum mesh $(2 \times 5 \text{ cm})$ as counter electrode.

Chronoamperometry was performed at different potentials and cyclic voltammetry was recorded in the range $-1.0\ \text{to}\ +1.0\ \text{V/SCE}$ at

 Table 1

 Characteristics of raw and diluted substrates.

10 mV/s in triplicate, using an SVP multichannel potentiostat (Bio-Logic Science Instruments; EC/Lab 2.0 software).

2.4. Microbial fuel cell studies

The two-chamber MFC was similar to the design already reported (Min et al., 2005). Two culture flasks (500 mL) were connected by a glass tube with a proton exchange membrane (3 cm diameter) in the middle (Ultrex, CMI-7000 Membranes International, Inc., USA). Graphite felt ($2 \times 5 \times 0.5$ cm or $2 \times 2 \times 0.5$ cm, RVG, Carbone Loraine) and platinum mesh (2×5 cm) were used as the anode and cathode respectively. Anolyte and catholyte had a volume of 500 mL. Four kinds of anolyte systems were used as follows:

- (a) diluted substrate without exogenous inoculum;
- (b) anaerobic sludge or compost leachate as inoculum with sodium acetate 2 mM as model substrate;
- (c) sludge inoculum fed with different diluted substrates;
- (d) compost leachate inoculum fed with different raw substrates.

The catholyte consisted of NaCl phosphate buffer (g/L): Na₂H-PO₄, 2.75; NaH₂PO₄, 3.67; NaCl, 0.584 g/L; pH 6.7, conductivity 5.2 mS/cm. This buffer solution was also used in substrate dilutions. Gas diffusers were placed in both compartments to inject nitrogen into the anodic compartment and air into the cathodic compartment when indicated. Nitrogen was supplied from a laboratory line and air was supplied by an aquarium pump (1.5-2 L/min). Anolyte and catholyte were stirred with magnetic barrels (350 rpm).

The electrical circuit was closed with an external resistance of 1000 Ω . The cell potential drop U_{cell} (V) was recorded with a multimeter (Integra series 2700, Keithley Instruments, Inc., USA) interfaced with a computer. Current density J (A/m²) was calculated as $J = U_{cell}/(R \cdot A_g)$, where R (Ω) is the external resistance and A_g (m²) is the anode projected surface area. The power density (W/m²) was measured as a function of current density by varying the external resistance from 100 to 56 324 Ω over a 3-min period.

2.5. Scanning electron microscopy (SEM)

SEM micrographs were taken with a Leo 435VP microscope (Germany) using the software SRV-32. Samples were fixed with 4% glutaraldehyde solution, post-fixed with 2% OsO_4 solution, gradually dehydrated with acetone 50%, 70%, 100% solutions, and finally coated with gold to be observed in SEM at 7.5 kV.

3. Results and discussion

3.1. Formation of electroactive biofilms under chronoamperometry with three different waste media (FAJ, WL and YW)

Chronoamperometry of three waste media: fermented apple juice (FAJ), wine lees (WL) and yogurt waste (YW) was performed at different potential values in the range +0.3 to +0.7 V/SCE with

Substrate	Conductivity ^a mS/cm	pH ^a	COD ^a mg/L	Dilution	COD ^b mg/L	Conductivity ^b mS/cm	pH^b
FAJ	4.20	4.01	52 514 ± 19 926	1:20	3501 ± 2510	4.77	6.14
WL	1.45	3.69	349 250 ± 39 527	1:20	10 843 ± 3904	5.12	6.06
YW	4.82	4.05	136 542 ± 37 983	1:20	8169 ± 2568	4.53	6.15

^a Raw substrate.

^b Diluted substrate.



Fig. 1. Electroactivity in raw YW. (A) Chronoamperometry at poised potentials +0.3, +0.5 and +0.7 V/SCE. (B) Cyclic voltammetry performed initially and after 20 days with electrode poised at +0.5 V/SCE.

graphite felt electrodes. Media were used as obtained, without dilution. As soon as electrodes were polarized, FAJ and WL gave significant current density, which rose with rising imposed potential and then dropped continuously with time to lower values around 50 mA/m² whatever the potential value. This behavior indicated the presence of a soluble compound that was initially oxidized and progressively consumed, but it did not indicate any formation of electrochemically active biofilm.

In contrast, the current generated in raw YW (Fig. 1A) at +0.3 and +0.5 V/SCE rose after 5 days lag-time up to value around 250 mA/m². Such a trend in current has commonly been identified as characteristic of the formation of an electroactive microbial bio-film on the electrode surface (Dulon et al., 2007). Biofilm formation was delayed at higher potential value (+0.7 V/SCE). As already reported, too high a potential value can be detrimental to biofilm formation or lower the reproducibility of the results (Parot et al., 2008a).

Voltammograms plotted at the beginning of the experiment (Fig 1 B) showed an oxidation wave that reached its maximal current value from around +0.7 V/SCE. Occurrence of this oxidation reaction can explain why chronoamperometry performed at +0.7 V/ SCE led to fast modification of the composition of the medium in the vicinity of the electrode in a way that was unfavorable to biofilm formation. The oxidation wave was no longer observed on the voltammograms plotted at the end of the 20 days' polarization. The oxidation reaction that appeared from around -0.3 V/SCE can be attributed to a biofilm-driven process. On the reduction side, the peak observed at -0.6 V/SCE on the initial curve was due to the reduction of dissolved oxygen trapped in the felt fibers, as was observed in control voltammograms (data not shown). The reduction wave that started near 0 V/SCE after the 20 days of polarization was due either to a biofilm-driven reduction process or to the reduction of components that formed during chronoamperometry.

It can be concluded that FAJ and WL did not contain an endogenous microflora able to support the development of electroactive biofilms, while YW presented promising capabilities that led to current densities of up to 250 mA/m^2 at +0.3 V/SCE.

3.2. Formation of electroactive biofilms under chronoamperometry with two inoculum sources (sludge and compost leachate)

Two sources of inoculum were used, anaerobic sludge and compost leachate. Actually, efficient bioanodes have been formed in garden compost (Parot et al., 2008a) but it has not yet been possible to use these biofilms out of the compost medium in which they were formed. For this reason, the bioanodes were formed in compost leachate here, i.e. in filtrate that was obtained by lixiviating garden compost with a saline solution (see Methods section). Chronoamperometry was performed in 500 mL compost leachate or in 50 mL anaerobic sludge suspension, both with 10 mM NaCl added. No acetate or other substrate was added.

Compost leachate tested at different potential values gave the three-phase current-time curves (lag-time, exponential increase, plateau) that indicated formation of an electroactive biofilm. Anaerobic sludge showed a faster current increase, almost linear from the beginning of the experiment. This behavior can be explained by the quite long acclimating period in closed flasks (see Methods section). Anaerobic sludge gave higher maximal current density, which was reached in a shorter period (110 mA/m² at +0.7 V/SCE, day 6) than for compost leachate (70 mA/m² at +0.6 V/Ag/AgCl, day 10). These results confirm the capacity of both media to be the source of efficient electroactive microbial species and suggest that biofilm microbial populations coming from wastewater may be more promising than those from compost.

3.3. Yogurt wastewater (YW) as inoculum and substrate in MFC

Following the results obtained with YW alone in Section 3.1, YW were tested for their capacity to form an MFC by themselves. The anode compartment was filled with YW diluted 1:20 v/v in buffer solution and the cathode compartment contained the buffer only. The electrical circuit was closed with a 1000 Ω external resistance. The power density increased with time to a maximal value of 1.9 mW/m² (49.4 mA/m²) at the 6th day, then fell slowly to 0.3 mW/m² (18.6 mA/m²) at the 13th day and kept this value for 5 days.

Polarization curves, which were recorded periodically during the test by changing the value of the external resistance, showed maximal power density of 3.9 mW/m² at current density of 22 mA/m² on the 8th day. This maximal power density remained small compared to the high current density (up to 250 mA/m²) observed under chronoamperometry (Section 3.1). The anode potential was measured during MFC operation vs. an Ag/AgCl reference electrode placed in the anode compartment. The polarization curves recorded by varying the value of the external resistance indicated that the open circuit potential was around -0.5 V and did not change significantly during MFC operation. The anode worked in the potential range from -0.35 to -0.25 V during MFC operation, i.e. very far from the potential values used for chronoamperometry tests (+0.3 to +0.7 V vs. SCE). The biofilm consequently remained in the MFC configuration far from its full capacity. Besides, biofilm formed under poised potential may develop different electrochemical behavior than biofilms formed in MFC conditions.

3.4. Sludge-MFC and compost-leachate MFC fed with acetate

In order to compare the electrochemical activity of the inoculum sources, MFCs were implemented with anaerobic sludge and compost leachate in the anode compartment. Anolytes were added with NaCl 10 mM to improve ionic conductivity. MFCs operated first for several days without any addition of supplementary substrate, with a $1000 \,\Omega$ external resistance. The current first increased, corresponding to biofilm formation, and then decreased back to zero, similarly to a cell discharge. When the current was back to around zero, 2 mM sodium acetate was added. In both cases, the current density increased very fast and returned close to zero after around 3 or 6 days for the compost-leachate MFC and the sludge-MFC respectively (Fig. 2). The sudden diminution of current observed at days 3 or 6 has commonly been noted in MFC supplied with acetate (Liu et al., 2005). The complete consumption of the substrate resulted in sharp current diminution after a stable plateau and not to a progressive current diminution that would follow the progressive depletion of acetate in the bulk. This means that the rate-limiting step does not linearly depend on the substrate concentration.

The sludge-MFC gave higher performance (*J* in the range from 120 to 210 mA/m², Coulombic efficiency 9.6%) than the compostleachate MFC (*J* from 100 to 160 mA/m², Coulombic efficiency 4.7%). The current densities obtained were in approximately the same ratio as the values obtained under chronoamperometry but had higher values (Section 3.2), which is consistent because acetate was added in the MFC experiments and not in the chronoamperometry tests.

3.5. Sludge-MFC fed with FAJ, WL and YW

Studies on MFC with complex substrates generally use diluted solutions. Heilmann and Logan (2006) diluted meat packing wastewater 1:4, Oh and Logan (2005) diluted cereal wastewater from 5950 to 595 mg COD/L, Zuo et al. (2006) diluted corn stover hydrolysates from 82 200 to 1000 mg COD/L. Complex industrial wastes are generally diluted to avoid undesirable planktonic metabolisms that may compete with the biofilm-catalyzed electricity generation and also to decrease fouling on electrode and membrane surfaces. The three waste media FAJ, WL and YW, were consequently tested here after 1:20 v/v dilution with phosphate buffer 50 mM containing 10 mM NaCl. This dilution ratio was chosen also because it brought to neutrality the dairy wastes, which generally represent most favorable environmental condition for microorganisms.

Three sludge-MFC systems were implemented with FAJ, WL or YW in the anode compartment. Anolytes were composed of 250 mL acclimated sludge (see Methods section) and 250 mL of



Fig. 2. Current density vs. time in sludge-MFC and compost-leachate MFC with sodium acetate 2 mM. ($R = 1000 \Omega$, $A_g = 10 \text{ cm}^2$). The curves start just after the addition of acetate.

diluted substrate. The resulting initial CODs in anolytes were 7600, 20 200 and 14 300 mg/L for FAJ, WL and YW respectively. MFCs operated at 1000 Ω without appreciable current generation for 16 days (<5 mW/m²). MFCs were kept for another month on open circuit for further acclimation of the inoculum sources to the substrate media and "spontaneous" COD removal. Thereafter, the electrical circuit was closed again through the 1000 Ω resistor (day 0 in Fig. 3), anolyte and catholyte were stirred at 350 rpm (from 8th day), and raw substrate was added sequentially at the 14th, 15th and 16th days. Neither nitrogen nor air flowed into the MFC.

Stirring improved MFC performance instantaneously with FAJ but not with YW. There was certainly no longer any substrate that could be metabolized by microorganisms in the YW-fed MFC. Power density increased with the first two additions of 1 mL raw substrate up to plateau values around 35 mW/m² with YW and 15 mW/m² with FAJ. WL never gave appreciable power density (always <0.25 mW/m²). The last addition of 2 mL raw substrate did not increase the power for FAJ and YW. Similar saturation profiles have been reported by numerous authors: electricity generation is dependent on fuel concentration until a plateau is reached. This saturation value can be attributed to a maximum rate of electron transfer reached by the metabolic mechanisms and/or by an increase in alternate electron acceptors that compete with the anode (Catal et al., 2008; Gil et al., 2003; Kim et al., 2007; Lee et al., 2002; Liu et al., 2005; Min and Logan, 2004; Zuo et al., 2006). At the end of the experiment, polarization curves were recorded by changing the external resistance in order to determine the maximal power density that could be reached by each MFC. The values reported in Table 2 confirm that anaerobic sludge is promising source of electrochemical biocatalysts to enlarge the variety of fuels for



Fig. 3. Effect of substrate addition in sludge-MFC fueled with fermented apple juice (FAJ) and yogurt waste (YW). (a) Stirring at 350 rpm. Substrate addition: (b) 1 mL, (c) 1 mL and (d) 2 mL. Dashed line at day 20 indicates change of the anolyte (replacing 450 mL with substrate).

Table 2

Maximal power density and correlated current density extracted from polarization curves in MFC inoculated with anaerobic sludge and fed with FAJ, WL and YW. Time corresponds to power density evolution depicted in Fig 3.

Time, days	FAJ		WL		YW	
	P _{max} , mW/m ²	J, mA/m ²	P _{max} , mW/m ²	J, mA/m ²	P _{max} , mW/m ²	J, mA/m ²
8	10.2	56.8	0.5	5.2	2	14.5
14	15.1	69.1	0.7	6.3	12.8	84.7
15	26.9	123.0	0.8	6.6	30.2	130.4
21	19.9	79.5	0.6	5.7	53.8	231.9
22	43.7	117.7	0.4	3.4	24.4	117.3
27	27.5	124.3	10^{-4}	1.8	4.6	28.7

MFC towards food-industry wastes, even though some substrates, such as WL, remain quite recalcitrant.

A second experimental campaign was carried out by removing 450 mL of the solution from the anodic compartment and replacing it by fresh diluted substrate (1:20 v/v). The power density diminished continuously from 34 to 3 mW/m² with YW, while the MFC with FAJ kept the same power density of $13-14 \text{ mW/m}^2$. The reduction of power density in the MFC fueled with YW can be explained by a rise in fermentable compounds that favored fermentative microbial metabolisms against electron transfer to the anode (Freguia et al., 2007). Competitive fermentation processes did not occur with FAJ, probably because the fermentation process had already been completed during the preparation of the substrate. The maximal power density (P_{max}) obtained during 1000 Ω operation and correlated current densities (*I*), measured by varying the value of the external resistance, confirmed the conclusions (Table 2): sludge-MFC fed with YW had the best performance and YW was optimally exploited at low concentration of substrate.

3.6. Compost-leachate MFC fed with FAJ and YW

The two substrates that gave the best performance in the sludge-MFC (Section 3.5), FAJ and YW, were tested in MFCs inoculated with leachate from garden compost. The anode compartments were filled with 500 mL previously acclimated compost leachate (see Methods section) and the external circuit was closed with a 1000 Ω resistance for 6 days in order to "discharge" the cell from the substrate contained in the inoculum. Then 1 mL of raw FAJ or YW was added to each cell. Fig. 4 gives the power density obtained, starting from the addition of the substrate.

With neither air nor nitrogen sparging, the first FAJ addition resulted in a power peak up to $P = 62 \text{ mW/m}^2$, which returned to zero in less than 1 day (Fig. 4A). The Coulombic efficiency corresponding to this peak was around 1% with respect to the COD contained in the added aliquots. This suggests that the components that contributed to current generation in FAJ were easily utilized by the compost microorganisms (immediate increase of power density) but at low concentration (low value of Coulombic efficiency). In contrast, the first addition of YW did not induce any significant power increase from the baseline, around 20 mW/m² (Fig. 4B).

The second substrate addition was simultaneous with air sparging in the cathode compartment (day 4). It resulted in a similar 1day peak with FAJ. The cathode reaction consequently did not affect this MFC. The second YW addition provoked an immediate increase of the power. In this case, air sparging in the cathode compartment significantly improved the behavior of the cell, indicating that the cathode introduced a limitation step in the YW-fed MFC. The power provided decreased from days 4 to 5 and it rose again at day 5 when nitrogen was bubbled into the anode compartment. This behavior clearly indicated that oxygen diffused though the membrane to the anode and disturbed the anode functioning, while nitrogen bubbling restored its full performance. The anode compartment likely contained facultative anaerobic microorganisms that were able to switch their metabolism from anode respiration to aerobic respiration. At the end of the experiment, the anode potential (Ea) varied towards negative values, indicating an improvement in anode efficiency.

In the FAJ-fed MFC, nitrogen sparging at day 5 did not have any effect on the power provided because compounds able to be metabolized had already been consumed at day 5. In contrast, the third FAJ addition (day 7) no longer resulted in a 1-day power peak but in stable power. The 2-day nitrogen sparging before the third FAJ addition likely modified the composition of the microbial population on the electrode surface, which became able to metabolize a larger range of organic compounds contained in the FAJ medium, confirming the beneficial effect of strict anaerobic conditions for both MFCs.

After continuous nitrogen (anode) and air (cathode) sparging, the power generation was stable in both MFCs. An ultimate substrate addition was made to verify the saturation value and determine the maximal power density provided in stationary conditions at 1000 Ω , which were 23 mW/m² for FAJ-fed MFC and 44 mW/m² for YW-fed MFC. The polarization curves (Fig. 5) gave maximal values around 64 mW/m² (at 253 mA/m²) and 92 mW/m² (at 404 mA/m²) respectively.

Strict control of simultaneous cathode aeration and anode deoxygenation are the main parameters that optimize MFC performance. These results obtained with similar compost-leachate MFCs also indicate that MFCs, even inoculated from the same source, can have different behaviors depending on the nature of the fuel. Such observations can explain, for instance, the discrepancies reported in the literature about the effect of aeration. It has often been observed that power density decreased as the air flow rate increased. Forcing air through a cathode tube has been shown to result in cell voltage reduction by 25% in an MFC inoculated and fed with wastewater (Liu et al., 2004). In a test using 0–200 mL/ min air flow rates, the power density of a similar MFC was reduced from 60 to 40 mW/m² (Min and Logan, 2004) but, conversely to these reports, Gil et al. (2003) showed a sharp current increase resulting from aeration of the cathode compartment at 100 mL/



Fig. 4. Effect of substrate addition, nitrogen (anode) and air (cathode) sparging on compost-leachate MFCs. (A) FAJ-fed MFC, (B) YW-fed MFC. Arrows indicate additions of 1 mL raw substrate. *R* = 1000 Ω, *A*_g = 10 cm².



Fig. 5. Power density recorded with various external resistances for compostinoculated MFC fed with fermented apple juice (FAJ) or yogurt waste (YW).

min in an MFC enriched over 3 years by feeding with starch processing wastewater.

3.7. SEM observations

Electrodes from sludge-MFC and compost-leachate MFC fed with YW were observed by SEM after 35 days of operation. The graphite fibers were poorly and non-uniformly colonized by bacteria. Dense microbial clusters with obvious formation of exopolymeric substances were rare, generally located only at the external electrode surface and not on the inner fibers. Fiber colonization was low with both inoculums similarly to that observed by other authors (Lee et al., 2003). Bacteria appeared to be predominantly *coccus* shaped as has also been mentioned by Liu and Logan (2004). Nevertheless, whatever the nature of the bioanodes, overall colonization appeared to be very poor. There is much room for improving MFC performance by promoting biofilm colonization onto anodes, as has been demonstrated in MFC with single or mixed cultures (Bond and Lovley, 2005; Ren et al., 2007).

4. Conclusions

Fermented apple juice, wine lees and yogurt waste were investigated here as possible fuels for MFC. Wine lees were definitely not suitable; yogurt waste was able to start an MFC with its endogenous microflora, while fermented apple juice was not. Nevertheless both FAJ and YW can be used in MFC using anaerobic sludge or compost leachate as inoculum source. Performance was drastically influenced by nitrogen/oxygen sparging in the anodic/cathodic compartments. Under optimized conditions, compost-leachateinoculated MFC fed with yogurt waste provided stable power density around 44 mW/m² (maximal value 92 mW/m² at 404 mA/m²). Using microflora coming from garden compost or from sludge revealed both efficient to exploit food-industry wastes as fuels. The low colonization of the graphite felt anode indicated that significant improvement can be expected by increasing microbial settlement on the electrode surface. Operating without addition of any buffer to the anode solution is another important direction to be investigated to progress towards economically-efficient set-ups.

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Appendix A. Supplementary data

Supplementary data associated with this article can be found, in the online version, at doi:10.1016/j.biortech.2009.11.076.

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