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1	Acceleration of the photocatalytic degradation of organics by <i>in-situ</i> removal of the products
2	of degradation
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11	Abstract

12 Photocatalytic degradation of sulphur containing organics in water was substantially 13 accelerated by the *in-situ* removal of sulphur oxide species formed by the dye degradation as 14 metal sulfate precipitates. The significant enhancement of the degradation of methylene blue 15 ($C_{16}H_{18}CIN_3S$) and acid yellow 42 ($C_{32}H_{24}N_8Na_2O_8S_2$), as examples of sulphur containing 16 organics, was achieved when the reactions were conducted in the presence of alkaline earth 17 cations (such as Ca^{2+} , Sr^{2+} , and Ba^{2+}).

18 Keywords: photocatalyst; titania; degradation; equilibrium shift

19 **1. Introduction**

Heterogeneous photocatalysis has been applied for the purification of the contaminated water by the oxidative decomposition of organic substances [1,2] and the concentration of heavy metals (As, Cd, and Cr) [3-8]. Designing structure, composition and morphology of photocatalysts has mainly been examined in order to improve the reaction efficiency [9-16]. 24 Alternatively, improved efficiency of photocatalytic reaction has been achieved by adjusting 25 reaction environments and conditions. The effects of pH of solution and the addition of 26 oxidizing agent (H₂O₂) have been investigated for the decomposition of various organic dyes 27 [17,18]. The decomposition of benzene to CO_2 was accelerated by the irradiation under 28 magnetic field [19], and that of an azo dye (acid orange 8) was modified by the irradiation 29 under sonication [20]. The yield and selectivity of the photocatalytic degradation of benzene 30 was affected by the presence of photo-inactive species such as clay minerals [21]. In order to 31 understand the mechanism of those positive effects as well as to improve the reaction efficiency 32 further, new examples on the effects of reaction environment to affect the photocatalytic 33 reactions are worth developing.

Here, we propose a new strategy to accelerate the decomposition of organics for wastewater treatment by chemical equilibrium shift, in which a product from the photocatalytic reaction system was removed. The idea in the present study is efficient removal of S oxide species (as sulfates) formed by the decomposition during the irradiation by capturing (*in-situ* removal) the S oxide species as precipitates of alkaline earth sulfates, which have very low solubility in water. Consequently, the dye photodegradation is expected to be promoted due to "equilibrium shift".

41 Sulphur containing organic compounds were selected as the target organic compounds 42 to be decomposed. The decomposition of sulphur containing azo dyes (e.g. Procion Red, 43 Reactive Brilliant Red, methyl orange) has been investigated in aqueous TiO₂ suspension; the 44 N oxide species as nitrates and S oxide species as sulfates were characterized as the products 45 in the solution after the dye decomposition [22-25]. It was reported that the amount of the 46 sulfate formed in the solution was lower than that expected from stoichiometry [26-32], which 47 indicated that some S oxide species were adsorbed on the surface of the photocatalyst. The adsorption of S oxide species may affect the photocatalytic reaction by blocking the catalyst's 48

surface for the substrate access, so that the capturing of the S oxide species from the reaction
system is expected to suppress the deactivation.

51 **2. Experimental methods**

52 2.1. Materials

53 TiO₂ (Evonik P25) was purchased from Evonik Industries. Strontium nitrate anhydrous, 54 calcium nitrate tetrahydrate, calcium acetate hydrate, methylene blue, acid yellow 42, and rhodamine 6G were purchased from Sigma-Aldrich Co., Ltd. Barium nitrate and magnesium 55 56 nitrate hexahydrate were purchased from Alfa Aesar Chemical Co., Ltd. Lithium nitrate was purchased from Honeywell FlukaTM. Sodium chloride and calcium chloride dihydrate was 57 58 purchased from Merck Ltd. Sodium nitrate solution was prepared by equivalent molar between 59 sodium hydroxide and nitric acid. Sodium hydroxide was purchased from Carlo Erba Reagents 60 S.r.l. Nitric acid was purchased from Chem-Lab NV. All chemicals were used without 61 purification.

62 2.2. Preparation of photocatalyst film

The photocatalyst film was prepared by casting an aqueous TiO₂ (P25) suspension onto borosilicate glass slides. The suspension was prepared by adding 0.1 g of P25 into 100 ml of DI water with magnetic stirring. Then, 0.8 ml of the P25 suspension was dropped on one side of the borosilicate glass slide and dried at room temperature. The suspension was naturally spread on the glass slide thanks to the wettability of the glass slide after the cleaning with 0.001 M of aqueous HCl solution for 1 day. The weight of P25 on the glass slide (ca. around 0.0008 g) was confirmed by measuring the weight increase after casting and drying.

70 2.3. Photocatalytic decomposition experiments

The photocatalytic decomposition of the S containing organic dyes (methylene blue and
 acid yellow 42) and the organic dye without S (rhodamine 6G) was carried out by putting the

73 film in an aqueous solution of the dyes and irradiated by a UV light (Eurosolar type 795 S 74 Solaria Facial Tanning Sun Face Tanner UV Type 3 as a light source) using four UV lamps 75 Philips CLEO 15 W. The film was immerged into the solution before UV irradiated for 1 h to 76 examine the adsorption of dyes on the TiO₂ surface. To evaluate the role of alkaline earth cations (Ca^{2+} , Sr^{2+} , and Ba^{2+}) on the decomposition of the dyes, the anhydrous strontium nitrate, 77 78 calcium nitrate tetrahydrate or barium nitrate was added into the dye solutions (initial concentration of the dye at 30 mg/L) as the source of Ca^{2+} , Sr^{2+} , and Ba^{2+} with fixing 79 80 concentration of alkaline earth cations at 0.5 M, respectively. The photocatalytic decomposition 81 of the dyes in the presence of sodium chloride, sodium nitrate, and lithium nitrate (as the source 82 of Na⁺ and Li⁺) was further examined to confirm the proposed mechanism under the identical 83 conditions (0.5 M of concentration of the inorganic salts). The effects of co-existing anions on 84 the decomposition of the dyes was examined by adding the 0.5 M of calcium chloride into the 85 solution. The pH of the solution before and after adding alkaline earth cations was measured by pH meter (HANNA, HI2004 edge). The irradiance was 0.01 W/cm², measured at distance 86 87 between the light source and the photocatalyst film by EIT Power UV Radiometer, 92380 GARCHES S/N0491029. The concentration of the dyes after the reaction was directly 88 determined by using JENWAY 7315 Spectrophotometer without any separation processes of 89 90 powder thanks to advantages of film application and the changes of the concentration of the 91 dyes (the absorbance at 664 nm for methylene blue, at 440 nm for acid yellow 42, and at 530 92 nm for rhodamine 6G) during irradiation were followed.

93 2.4. *Kinetic analyses*

Apparent pseudo first-order decomposition rate constant (k) was determined for the evaluation the photocatalytic decomposition of the dyes in the presence and the absence of alkaline earth cations in the solution. The rate constant (k = slope) is derived by the graph 97 relationship between $\ln(C_0/C_t)$ versus irradiation time, where C_0 is initial concentration of the 98 dyes and C_t is concentration of the dyes after the reaction at time t.

99 **3. Results and discussion**

100 3.1. Morphology and the advantages of the photocatalyst film

101 In the preliminary experiments using the titania photocatalyst (P25, Evonik, average 102 particle size of 22 nm from Aerosil Co., Ltd.) suspension, the state of the suspension was 103 significantly affected by the presence of metal salts (cause agglomeration and sedimentation, 104 as shown in Fig. S1). In order to avoid the effects of the difference (light scattering, diffusion 105 of reactant etc.) on the photocatalytic reaction, a film of photocatalyst was used in the present 106 study. P25 was employed as the photocatalyst due to its high efficiency for the photocatalytic 107 decomposition of organics and the stability of aqueous suspension. The aqueous suspension 108 was used to prepare homogeneous film on substrate by simple casting [33,34]. The photograph 109 of the P25 film and the SEM images are shown in Fig. 1. The film thickness was around 1.0 110 µm with the estimated porosity of 67% based on the calculation from the volume of solid 111 content (0.0008 g of P25 on the substrate and the density of P25, 3.9 g/cm³) and the estimated total volume of the film (6.25 cm^2 of the coating area and 1.0 μ m of the film thickness). The 112 113 porosity of random close packing of dense spherical particles is ~36% with the assumption for 114 monodispersed particles [35]. The higher porosity of the film (67%) was explained by the 115 sparse aggregation of P25 particles during drying, leading to interconnected pore between the 116 P25 particles as seen in the surface morphology of the film (Fig. 1c). Thanks to the thickness 117 $(1.0 \,\mu\text{m})$ and the high porosity (67%) of the film, the diffusion of organic molecules inside the 118 particle voids is expected to be easy and all P25 particles in the layer film are photocatalytically 119 active [36]. As the TiO₂ was immobilized on the substrate, alkaline earth sulfate formed after 120 the reaction was expected to be observed by naked eyes as precipitates. Moreover, it was

- 121 possible to determine the change of the concentration of the organics without the separation of
- 122 photocatalyst from the reaction mixture.



Fig. 1. (a) Photograph of the P25 immobilized on glass as the film. (b) Experimental set-up by
putting the P25 film into the solutions. (c) Surface and (d) cross-section images of the P25 film
from SEM observations.

- 127 3.2. Photocatalytic decomposition of sulphur-containing organics
- 128 Methylene blue ($C_{16}H_{18}CIN_3S$, abbreviated as MB) was selected as a sulphur containing 129 organic. The photocatalytic decomposition of MB was carried out in the absence or presence 130 of an alkaline cation, Na⁺ (adding NaCl), or of alkaline earth cations such as Ca²⁺, Sr²⁺, and

 Ba^{2+} (adding Ca(NO₃)₂, Sr(NO₃)₂, and Ba(NO₃)₂) in the solution, respectively (Fig. 2a). The 131 132 decomposition of MB was accelerated when the reaction was conducted in the presence of Ca^{2+} , Sr^{2+} , and Ba^{2+} (Fig. 2b). When the reaction was conducted in the presence of Na⁺ in the 133 solution, the decomposition rate was not improved and slightly retarded after the UV irradiation 134 135 for 2 h. The adsorption of MB on the P25 film has here to be discussed considering the surface 136 charge of titania. The pH of solution changed from 6.48 to 6.41, 5.86, 5.91, and 5.64 after the 137 addition of NaCl, Ca(NO₃)₂, Sr(NO₃)₂, Ba(NO₃)₂ in the solution, respectively. The isoelectric 138 point of P25 is reported to be around 6 [37]. Adsorption of MB on the P25 film was not observed for the reaction in Ca(NO₃)₂, Sr(NO₃)₂, and Ba(NO₃)₂ solutions. The adsorption of 139 140 humic acid on TiO₂ was before claimed as the main reason to promote the decomposition of humic acid in aqueous suspension containing Ca^{2+} [38]. The present phenomena are therefore 141 142 not explained by the effects of the MB adsorption.

The decomposition of MB in the presence of Mg(NO₃)₂, LiNO₃ and NaNO₃ was also examined. The concentration of MB decreased from 30 ppm to 27 ppm by the addition of Mg(NO₃)₂, LiNO₃, and NaNO₃ into the MB solution in the dark, suggesting that the adsorption of MB on P25 by the electrostatic interactions happened to some extent by the change of pH (from 6.48 to 7.32, 9.21, and 12.10 for Mg(NO₃)₂, LiNO₃ and NaNO₃ systems, respectively).



Fig. 2. Photocatalytic decomposition of (a) MB and (b) apparent pseudo first order plot of the MB decomposition by using the P25 film in the absence (black open circle) and the presence of Na⁺ (red triangle) and alkaline earth cations (Ca²⁺; blue full circle, Sr²⁺; blue square, and Ba²⁺; blue diamond) by adding NaCl, Ca(NO₃)₂, Sr(NO₃)₂, and Ba(NO₃)₂ in the solution, respectively.

148

154 An anionic dye, acid yellow 42 (C₃₂H₂₄N₈Na₂O₈S₂, abbreviated as AY42), was chosen as another example of sulphur containing organics to be examined for the photocatalytic 155 156 decomposition in the presence of metal salts. The photocatalytic decomposition of AY42 was successfully accelerated in the presence of Ca^{2+} , Sr^{2+} and Ba^{2+} , whereas no significant effects 157 on the decomposition of AY42 were observed by the addition of Mg^{2+} , Na^+ and Li^+ (Fig. 3a). 158 The decomposition of AY42 was also accelerated when CaCl₂ was added. Thus, the 159 decomposition rate of MB and AY42 was accelerated by the presence of Ca^{2+} , Sr^{2+} and Ba^{2+} , 160 and was not affected by the co-existing anions (NO₃⁻ or Cl⁻). The rate constants of the 161 decomposition of AY42 were 0.83, 1.17, and 1.14 h^{-1} for the reactions in the presence of Ba²⁺, 162 Ca^{2+} , and Sr^{2+} , respectively (Fig. 3b). The decomposition of AY42 was faster than that of MB 163

under the present conditions, which is thought to be due to the difference in the decompositionpathways of AY42 and MB [31].



166

Fig. 3. (a) Photocatalytic decomposition of AY42 and (b) apparent pseudo first order plot of
the AY42 decomposition by using the P25 film in the absence and the presence of inorganic
salts such as LiNO₃, NaNO₃, Mg(NO₃)₂, Ca(NO₃)₂, CaCl₂, Sr(NO₃)₂, and Ba(NO₃)₂.

170 3.3. Photocatalytic decomposition of organic without sulphur in the chemical structure

The decomposition of Rhodamine 6G ($C_{28}H_{31}ClN_2O_3$, abbreviated as R6G), which is a dye without sulphur in its structure, was examined in the presence of Ca^{2+} (by adding $CaCl_2$ or Ca(COOCH₃)₂ into the solution). It was reported that nitrate ions can interact with R6G to show the decrease of the concentration [39], so that calcium nitrate was not used to examine the decomposition of R6G in the present study. The changes in the concentration by the UV irradiation are shown in Fig. 4. The decomposition of R6G was not accelerated by the presence

- 177 of Ca^{2+} , supporting the important role of the *in-situ* removal of the sulphur containing species
- 178 on the photodegradation of sulphur containing organics.



179

Fig. 4. Photocatalytic decomposition of R6G in the absence and the presence of Ca²⁺ by adding
CaCl₂ or Ca(COOCH₃)₂ in the solution.

182 3.4. Proposed mechanism for the accelerated photodegradation of sulphur containing organics

183 Precipitate was clearly observed as the sediment after the decomposition of MB in the presence of Ca^{2+} (Fig. 5a), while the amount of the precipitate was too small to be collected for 184 its identification. The amount of CaSO₄ which may be formed considering the sulfate anions 185 186 resulting from MB oxidation, is estimated to 0.13 mg based on the stoichiometric calculation 187 assuming the complete decomposition of the MB in the starting solution. The presence of sulfate ions formed in the solution after the decomposition of MB in the presence of Ca^{2+} and 188 189 Na⁺ for 16 h (colorless solutions) was examined by ion chromatography, as shown in Fig. 5b. The concentration of sulfate ions in the solution containing Ca^{2+} and Na^{+} were 0.4 and to 6.4 190 191 mg.L⁻¹, respectively. Assuming the complete decomposition of MB, the estimated

192 concentration of sulfate ions in the solution is 9 mg.L⁻¹. It was thought that the sulfate ions 193 were removed from the solution by the precipitation with Ca^{2+} as CaSO₄.

The concentration of sulfate ions required for the formation of alkaline earth sulfates is 194 calculated to be 4.6, 0.05, and 2.1×10^{-5} mg.L⁻¹ for CaSO₄, SrSO₄ and BaSO₄, respectively. 195 196 from solubility equilibrium's equation using the initial concentration of alkaline earth cations (0.5 mol.L⁻¹) and the solubility product constant (K_{sp}) of 2.4×10⁻⁵, 2.51×10⁻⁶, and 1.07×10⁻¹⁰ 197 for CaSO₄, SrSO₄ and BaSO₄ [40,41], respectively. Those values were lower than the expected 198 199 concentration of sulfate ions in the present experiment after the complete decomposition of MB (ca. 9 mg.L⁻¹), meaning that the precipitation of CaSO₄, SrSO₄ and BaSO₄ was possible from 200 201 the concentration of the sulfate ions formed by the decomposition of MB.

The observed effect of the additives on the improved efficiency of the degradation of sulphur containing organics (AY42 and MB) is proposed to be a "chemical equilibrium shift", where the product (sulfate) was removed during the reactions by the alkaline earth cations (Ca^{2+} , Ba^{2+} , and Sr^{2+}) as sulfate precipitation. The presence of the precipitates and the loss of sulfate ions in the solution during the irradiation are the evidences of the removal of S oxide species as metal (Ca^{2+} , Ba^{2+} , and Sr^{2+}) sulfates.

In addition to the equilibrium shift concept, the deactivation of TiO_2 by the adsorption 208 209 of S oxide species was suppressed as another positive aspect of the present strategy. The film 210 was re-used to obtain the similar results for the MB decomposition in the absence and the presence of Ca²⁺ (Fig. S2). The efficient MB decomposition was found in the second and third 211 runs in the presence of Ca^{2+} , while the decomposition of MB was retarded when the P25 film 212 was re-used for the reaction in the absence of Ca^{2+} . These were consistent with the relatively 213 white color of the P25 film after the reactions in the presence of Ca^{2+} , Sr^{2+} , and Ba^{2+} , when 214 compared with the P25 film after the reactions without adding such ions as Ca^{2+} , Ba^{2+} , and 215 Sr^{2+} , or in the presence of Na⁺ (Fig. 5c). Therefore, the simultaneous prevention of S oxide 216

species adsorption on the catalysts and the efficient removal of product (as sulfates) areclarified as two possible mechanisms on the acceleration of MB decomposition.



Fig. 5. (a) Photograph of precipitated powder after the decomposition of MB in the presence
of Ca²⁺. (b) Characterization of sulfate ions remaining in the solution by Ion Chromatography.
(c) Photographs of the P25 film before and after the reaction.

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The effects of the added inorganic cations (such as K⁺, Na⁺, Ca²⁺ and Mg²⁺) [42-49] and anions (such as Cl⁻, SO₄²⁻ and NO₃⁻) [50-57] on the decomposition of various organics have been reported in the literature, where the reactions were conducted in aqueous TiO₂ suspensions. For the reactions in the presence of Ca²⁺, no significant acceleration was found on the decomposition of phenol [42,43], crystal violet [44], and glyphosate [45], while negative effects were observed for the decomposition of rhodamine B [46], nicosulfuron [47], and formic acid [48]. The adsorption of the added inorganic cation (Ca²⁺) and anions (SO₄²⁻) onto 230 the photocatalyst (TiO₂) was claimed as a reason of the inhibition of the decomposition of 231 nicosulfuron by blocking the accessible surface of TiO2. The sedimentation of P25 was observed in the aqueous suspension by the addition of Ca^{2+} salts (Fig. S1). The sedimentation 232 of P25 in the aqueous suspension is thought to affect the decomposition rate of MB by losing 233 234 the active surface for the reaction by the aggregation of P25 and also by increasing the diffusion 235 path from the solution to the active surface in the core of the formed aggregates as well as the 236 loss of incident light propagation. Thus, the decomposition efficiency in the aqueous 237 suspension of TiO₂ is difficult to be directly compared with the result achieved in the present 238 study by using P25 film as the photocatalysts. The application of P25 film in this work is a key 239 from the viewpoint of the product characterization.

240 4. Conclusions

The decomposition of the sulphur containing organics (MB and AY42) was 241 242 successfully accelerated for the first time by adding inorganic cations (alkaline earth cations) 243 to capture the S oxide species (as sulfates) formed during the decomposition. Since the 244 commonly used suspension of photocatalysts cause the aggregation during the reaction, the 245 film of P25 supported on borosilicate glass was used as the photocatalyst, which was a key of 246 the present success. The acceleration was explained by the two mechanisms, equilibrium shift and the suppressed poisoning (deactivation) of the catalyst's surface with sulfate. The present 247 concept (in-situ removal of a product) can be applied various photocatalytic reactions to be 248 249 accelerated.

250

CRediT authorship contribution statement

251 Siwada Deepracha: Methodology, Formal analysis, Investigation, Writing – Original 252 Draft, Writing - Review & Editing. André Ayral: Validation, Formal analysis, Resources,

Writing – Review & Editing. Makoto Ogawa: Conceptualization, Validation, Writing –
Review & Editing, Supervision.

255 Declaration of Competing Interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

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265 Appendix A. Supplementary data

266 Photocatalytic decomposition of MB by the P25 dispersed in the aqueous MB solution 267 with and without addition of $Ca(NO_3)_2$, photograph of the P25 dispersed in water with and 268 without addition of $Ca(NO_3)_2$, and reuse of the photocatalyst film for the photocatalytic 269 decomposition of MB

270 References

[1] L. Max, P. Pichat, Photocatalysis and water purification; from fundamentals to recent
applications, Wiley-VCH, Germany, 2013.

[2] K. Kabra, R. Chaudhary, R.L. Sawhney, Treatment of hazardous organic and inorganic
compounds through aqueous-phase photocatalysis: a review, Ind. Eng. Chem. Res. 43 (2004)
7683-7696.

276	[3] X. Guan, J. Du, X. Meng, Y. Sun, B. Sun, Q. Hu, Application of titanium dioxide in
277	arsenic removal from water: a review, J. Hazard. Mater. 215 (2012) 1-16.
278	[4] X. Bai, Y. Du, X. Hu, Y. He, C. He, E. Liu, J. Fan, Synergy removal of Cr (VI) and
279	organic pollutants over RP-MoS ₂ /rGO photocatalyst, Appl. Catal. B 239 (2018) 204-213.
280	[5] S. Ghosh, H. Remita, R.N. Basu, Visible-light-induced reduction of Cr (VI) by PDPB-
281	ZnO nanohybrids and its photo-electrochemical response, Appl. Catal. B 239 (2018) 362-372.
282	[6] R. Vinu, G. Madras, Kinetics of simultaneous photocatalytic degradation of phenolic
283	compounds and reduction of metal ions with nano-TiO ₂ , Environ. Sci. Technol. 42 (2012) 913-
284	919.
285	[7] X.J. Zhang, T.Y. Ma, Z.Y. Yuan, Titania-phosphonate hybrid porous materials:

preparation, photocatalytic activity and heavy metal ion adsorption, J. Mater. Chem. 18 (2008)
2003-2010.

[8] S. Huang, L. Gu, N. Zhu, K. Feng, H. Yuan, Z. Lou, Y. Li, A. Shan, Heavy metal recovery
from electroplating wastewater by synthesis of mixed-Fe₃O₄@SiO₂/metal oxide magnetite
photocatalysts, Green Chem. 16 (2008) 2696-2705.

[9] H. Yamashita, K. Mori, Y. Kuwahara, T. Kamegawa, M. Wen, P. Verma, M. Che, Singlesite and nano-confined photocatalysts designed in porous materials for environmental uses and
solar fuels, Chem. Soc. Rev. 47 (2018) 8072-8096.

[10] A. Kudo, Y. Miseki, Heterogeneous photocatalyst materials for water splitting, Chem.
Soc. Rev. 38 (2009) 253-278.

[11] K. Maeda, K. Domen, Solid solution of GaN and ZnO as a stable photocatalyst for
overall water splitting under visible light, Chem. Mater. 22 (2010) 612-623.

[12] T.W. Kim, S.J. Hwang, S.H. Jhung, J.S. Chang, H. Park, W. Choi, J.H. Choy,
Bifunctional heterogeneous catalysts for selective epoxidation and visible light driven
photolysis: nickel oxide-containing porous nanocomposite, Adv. Mater. 20 (2008) 539-542.

[13] X. Huang, Y. Chen, E. Walter, M. Zong, Y. Wang, X. Zhang, O. Qafoku, Z. Wang, K.M.
Rosso, Facet-specific photocatalytic degradation of organics by heterogeneous fenton
chemistry on hematite nanoparticles, Environ. Sci. Technol. 53 (2019) 10197-10207.

304 [14] D. Zhang, C. Lee, H. Javed, P. Yu, J.H. Kim, P.J. Alvarez, Easily recoverable,
305 micrometer-sized TiO₂ hierarchical spheres decorated with cyclodextrin for enhanced
306 photocatalytic degradation of organic micropollutants, Environ. Sci. Technol. 52 (2018) 12402307 12411.

308 [15] L. Guo, Y. Jing, B.P. Chaplin, Development and characterization of ultrafiltration TiO₂
309 magnéli phase reactive electrochemical membranes, Environ. Sci. Technol. 50 (2016) 1428310 1436.

[16] Z. Xiong, J. Ma, W.J. Ng, T.D. Waite, X.S. Zhao, Silver-modified mesoporous TiO₂
photocatalyst for water purification, Water Res. 45 (2011) 2095-2103.

[17] I.K. Konstantinou, T.A. Albanis, TiO₂-assisted photocatalytic degradation of azo dyes
in aqueous solution: kinetic and mechanistic investigations: a review, Appl. Catal. B 49
(2004) 1-14.

[18] Y. Ye, Y. Feng, H. Bruning, D. Yntema, H.H.M. Rijnaarts, Photocatalytic degradation
 of metoprolol by TiO₂ nanotube arrays and UV-LED: Effects of catalyst properties, operational
 parameters, commonly present water constituents, and photo-induced reactive species, Appl.

319 Catal. B 220 (2018) 171-181.

[19] W. Zhang, X. Wang, X. Fu, Magnetic field effect on photocatalytic degradation of
benzene over Pt/TiO₂, Chem. Commun. (2003) 2196-2197.

[20] E. Selli, Synergistic effects of sonolysis combined with photocatalysis in the degradation
of an azo dye, Phys. Chem. Chem. Phys. 4 (2002) 6123-6128.

324 [21] Y. Ide, M. Matsuoka, M. Ogawa, Controlled photocatalytic oxidation of benzene in

aqueous clay suspension, ChemCatChem 4 (2012) 628-630.

- 326 [22] F. Lin, Y. Zhang, L. Wang, Y. Zhang, D. Wang, M. Yang, J. Yang, B. Zhang, Z. Jiang, C.
- 327 Li, Highly efficient photocatalytic oxidation of sulfur-containing organic compounds and dyes
- on TiO_2 with dual cocatalysts Pt and RuO₂, Appl. Catal. B 127 (2012) 363-370.
- [23] C. Hu, C.Y. Jimmy, Z. Hao, P.K. Wong, Photocatalytic degradation of triazinecontaining azo dyes in aqueous TiO₂ suspensions, Appl. Catal. B 42 (2003) 47-55.
- 331 [24] V. Augugliaro, C. Baiocchi, A.B. Prevot, E. García-López, V. Loddo, S. Malato, G.
- Marcí, L. Palmisano, M. Pazzi, E. Pramauro, Azo-dyes photocatalytic degradation in aqueous
 suspension of TiO₂ under solar irradiation, Chemosphere 49 (2002) 1223-1230.
- 334 [25] M. Vautier, C. Guillard, J.M. Herrmann, Photocatalytic degradation of dyes in water:
- case study of indigo and of indigo carmine, J. Catal. 201 (2001) 46-59.
- 336 [26] F. He, U. Muliane, S. Weon, W. Choi, Substrate-specific mineralization and deactivation
- behaviors of TiO₂ as an air-cleaning photocatalyst, Appl. Catal. B (2020) 119145.
- 338 [27] N.M. Mahmoodi, M. Arami, N.Y. Limaee, K. Gharanjig, F.D. Ardejani, Decolorization
- and mineralization of textile dyes at solution bulk by heterogenous nanophotocatalysis using
- immobilized nanoparticles of titanium dioxide, Colloids Surf. A 290 (2006) 125-131.
- [28] N. González-García, J.A. Ayllón, X. Doménech, J. Peral, TiO₂ deactivation during the
 gas-phase photocatalytic oxidation of dimethyl sulfide, Appl. Catal. B 52 (2004) 69-77.
- 542 gas phase photoedalityte oxidation of dimetrify suffice, Appl. Catal. D 52 (2004) 05 77.
- 343 [29] C.H. Ao, S.C. Lee, S.C. Zou, C.L. Mak, Inhibition effect of SO₂ on NO_x and VOCs
- 344 during the photodegradation of synchronous indoor air pollutants at parts per billion (ppb) level
- 345 by TiO₂, Appl. Catal. B 49 (2004) 187-193.
- [30] M. Karkmaz, E. Puzenat, C. Guillard, J.M. Herrmann, Photocatalytic degradation of the
 alimentary azo dye amaranth: Mineralization of the azo group to nitrogen, Appl. Catal. B 51
 (2004) 183-194.

[31] H. Lachheb, E. Puzenat, A. Houas, M. Ksibi, E. Elaloui, C. Guillard, J.M. Herrmann,
Photocatalytic degradation of various types of dyes (alizarin S, crocein orange G, methyl red,
congo red, methylene blue) in water by UV-irradiated titania, Appl. Catal. B 39 (2002) 75-90.
[32] M. Stylidi, D.I. Kondarides, X.E. Verykios, Pathways of solar light-induced

photocatalytic degradation of azo dyes in aqueous TiO₂ suspension, Appl. Catal. B 40 (2003)
271-286.

[33] S. Deepracha, S. Bureekaew, M. Ogawa, Synergy effects of the complexation of a titania
and a smectite on the film formation and its photocatalyst' performance, Appl. Clay Sci. 169
(2019) 129-134.

[34] T. Goto, M. Ogawa, Visible light responsive photocatalytic flow reactor composed of
titania film photosensitized by metal complex-clay hybrid, ACS Appl. Mater. Interfaces 7
(2015) 12631-12634.

[35] F.A. Dullien, Porous media: fluid transport and pore structure, Academic press, Elsevier,
United States, 2012.

[36] D. Chen, F. Li, A.K. Ray, Effect of mass transfer and catalyst layer thickness on
photocatalytic reaction, AIChE J. 46 (2000) 1034-1045.

[37] M. Kosmulski, Isoelectric points and points of zero charge of metal (hydr) oxides: 50
years after Parks' review, Adv. Colloid Interface Sci. 238 (2016) 1-61.

367 [38] X.Z. Li, C.M. Fan, Y.P. Sun, Enhancement of photocatalytic oxidation of humic acid in

368 TiO₂ suspensions by increasing cations strength, Chemosphere 48 (2002) 453-460.

369 [39] L. Wang, B. Li, L. Zhang, L. Zhang, H. Zhao, Fabrication and characterization of a

370 fluorescent sensor based on Rh6G-functionalized silica nanoparticles for nitrite ion detection,

371 Sensor Actuat. B-Chem. 171 (2002) 946-953.

372 [40] W.I. Stephen, A. Ringbom, Complexation in analytical chemistry: Interscience

373 Publishers, John Wiley & Sons, New York, 1965.

- [41] B.S. Krumgalz, Temperature dependence of mineral solubility in water. Part 3. Alkaline
 and alkaline earth sulfates, J. Phys. Chem. Ref. Data 47 (2018) 023101.
- [42] N. Kashif, F. Ouyang, Parameter effect on heterogeneous photocatalyzed degradation
 of phenol in aqueous dispersion of TiO₂, Environ. Sci. Technol. 21 (2009) 527-533.
- 378 [43] V. Brezová, A. Blažková, E. Borošová, M. Čeppan, R. Fiala, The influence of dissolved
- 379 metal ions on the photocatalytic degradation of phenol in aqueous TiO₂ suspension, J. Mol.
- 380 Catal. A-Chem 98 (1995) 109-116.
- 381 [44] C. Sahoo, A. Gupta, A. Pal, Photocatalytic degradation of crystal violet (CI Basic Violet
- 382 3) on silver ion doped TiO₂, Dyes Pigm. 66 (2005) 189-196.
- [45] S. Chen, Y. Liu, Study on the photocatalytic degradation of glyphosate by TiO₂
 photocatalyst, Chemosphere 67 (2007) 1010-1017.
- [46] M. Makita, A. Harata, Photocatalytic decolorization of rhodamine B dye as a model of
 dissolved organic compounds: Influence of dissolved inorganic chloride salts in seawater of
 the Sea of Japan, Chem. Eng. Process 47 (2008) 859-863.
- 388 [47] A.M. Dugandžić, A.V. Tomašević, M.M. Radišić, N.Ž. Šekuljica, D.Ž. Mijin, S.D.
- Petrović, Effect of inorganic ions, photosensitizers and scavengers on the photocatalytic
 degradation of nicosulfuron, J. Photochem. Photobiol. A 336 (2017) 146-155.
- 391 [48] N. Negishi, Y. Miyazaki, S. Kato, Y. Yang, Effect of HCO₃⁻ concentration in
- 392 groundwater on TiO₂ photocatalytic water purification, Appl. Catal. B 242 (2019) 449-459.
- 393 [49] S. Sontakke, J. Modak, G. Madras, Effect of inorganic ions, H₂O₂ and pH on the
- 394 photocatalytic inactivation of Escherichia coli with silver impregnated combustion synthesized
- 395 TiO₂ catalyst, Appl. Catal. B 106 (2011) 453-459.
- 396 [50] J.C. Crittenden, Y. Zhang, D.W. Hand, D.L. Perram, E.G. Marchand, Solar detoxification
- 397 of fuel-contaminated groundwater using fixed-bed photocatalysts, Water Environ. Res. 68
- 398 (1996) 270-278.

- [51] D.C. Schmelling, K.A. Gray, P.V. Kamat, The influence of solution matrix on the
 photocatalytic degradation of TNT in TiO₂ slurries, Water Res. 31 (1997) 1439-1447.
- 401 [52] M.H. Habibi, A. Hassanzadeh, S. Mahdavi, The effect of operational parameters on the
 402 photocatalytic degradation of three textile azo dyes in aqueous TiO₂ suspension, J. Photochem.
 403 Photobiol. A 172 (2005) 89-96.
- 404 [53] A.G. Rincon, C. Pulgarin, Effect of pH, inorganic ions, organic matter and H₂O₂ on E.

405 coli K12 photocatalytic inactivation by TiO₂: implications in solar water disinfection, Appl.
406 Catal. B 51 (2004) 283-302.

407 [54] C. Guillard, H. Lachheb, A. Houas, M. Ksibi, E. Elaloui, J.M. Herrmann, Influence of
408 chemical structure of dyes, of pH and of inorganic salts on their photocatalytic degradation by
409 TiO₂ comparison of the efficiency of powder and supported TiO₂, J. Photochem. Photobiol. A
410 158 (2003) 27-36.

- [55] H.Y. Chen, O. Zahraa, M. Bouchy, Inhibition of the adsorption and photocatalytic
 degradation of an organic contaminant in an aqueous suspension of TiO₂ by inorganic ions, J.
 Photochem. Photobiol. A 108 (1997) 37-44.
- 414 [56] L. Wenhua, L. Hong, C. Sao'an, Z. Jianqing, C. Chunan, Kinetics of photocatalytic
 415 degradation of aniline in water over TiO₂ supported on porous nickel, J. Photochem. Photobiol.
 416 A 131 (2000) 125-132.
- 417 [57] M. Abdullah, G.K. Low, R.W.E. Matthews, Effects of common inorganic anions on rates
- 418 of photocatalytic oxidation of organic carbon over illuminated titanium dioxide, J. Phys. Chem.
- 419 94 (1990) 6820-6825.