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1 **Assessment of the PM_{2.5} oxidative potential in a coastal industrial city in Northern**
2 **France: relationships with chemical composition, local emissions and long range sources**

3

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5

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10

11 Abstract

12 The objective of this work was to relate PM_{2.5} Oxidative Potential (OP) data to PM
13 composition and PM local and distant source contributions. PM_{2.5} collected in Dunkerque, a
14 coastal industrial city in North of France, was extensively characterized for major and minor
15 chemical species. PM_{2.5} filters were extracted using a synthetic pulmonary fluid to achieve OP
16 estimation based on Ascorbic Acid (AA) and dithiothreitol (DTT) depletion assays. In order
17 to evidence relationships between OP values, chemical composition and local and distant
18 source contributions, correlation coefficient, Principal Component Analysis (PCA),
19 concentration roses and polar plots and concentration weighted trajectories were used.
20 Heterogeneous conclusions were drawn using the three first methods as the bivariate polar
21 plots lead to dismiss some of the correlation evidenced using correlation coefficient and PCA.
22 Both AA and DTT tests appeared complementary as they were not sensitive to the same
23 species/source contribution. **The bivariate polar plot representation of OP values versus wind**
24 **direction and wind speed revealed that PM_{2.5} concentration and combustion sources were**
25 **linked to OP-AA, whereas emissions from integrated steelworks, electric steelworks, heavy**

26 fuel oil combustion and traffic non-exhaust significantly contribute to OP-DTT. Sea-salts,
27 aged sea-salts, crustal, secondary sulfates and secondary nitrates sources were not found to
28 contribute to OP values. Constant weighted trajectories evidenced several source regions
29 responsible for high OP values with Belgium, Germany and Netherland at the leader position.
30 Contribution of inland regions appeared possibly related to the biomass and traffic related
31 combustion while heavy fuel oil combustion could be also involved in the contribution of
32 marine and coastal areas.

33

34 Keywords: PM_{2.5}; source contribution; oxidative potential; dithiothreitol (DTT) assay; acid
35 ascorbic (AA) assay, concentration weighted trajectories (CWT)

36

37 **Introduction**

38 Exposure to particulate matter (PM) is a major public health issue for all countries in
39 the world. Air quality modelling has revealed that 91% of the global population lives in areas
40 where annual average concentration of PM_{2.5} (particulate matter with an equivalent
41 aerodynamic diameter less than 2.5 µm) exceeds the corresponding WHO guideline limits
42 (10 µg·m⁻³) (WHO, 2018). Exposure to the outdoor fine particles has been identified as an
43 important risk factor of mortality with more than 4.2 million premature deaths in the World in
44 2016 (WHO, 2018). Burnett et al. (2018) applied a new global exposure mortality model to
45 study the association between PM_{2.5} concentration and non-accidental mortality considering
46 41 cohorts in 16 countries and estimated 8.9 million of PM_{2.5} related deaths in 2015 (Burnett
47 et al., 2018). It is consistent with the 8.8 million recently modelled by Lelieveld et al. (2020).
48 A similar approach applied to the 250 most populates cities in the world predicted a median
49 rate of PM_{2.5}-related mortality reaching 39 per 100,000 people, with differences in the cities
50 (13 to 125 deaths per 100,000 people) (Anenberg et al., 2019).

51 It is well established that particulate matter causes several diseases as lung cancers
52 (Raaschou-Nielsen et al., 2013; Sax et al., 2013), cardiovascular diseases (Massamba et al.,
53 2014), chronic obstructive pulmonary disease (COPD) (Li et al., 2016) and asthma (Bontinck
54 et al., 2020; Hendryx et al., 2019). Toxicological studies have shown that PM has the ability
55 to induce oxidative stress. This latter corresponds to an imbalance between the antioxidant
56 defence and the overproduction of reactive oxygen species (ROS) triggered by an exposure to
57 airborne PM. Oxidative stress appears as an early pathways potentially preceding severe
58 damages to cell components (lipids, proteins and DNA) (Borm et al., 2007; Nel, 2005). The
59 PM composition is a determinant of their ability to contribute to overproduce ROS, and both
60 organic compounds and some metals appeared to be associated with ROS production (Kelly
61 and Fussell, 2012).

62 Components that contribute mostly to PM mass were found to contribute the least to
63 the toxicity (Borm et al., 2007). For this reason, the PM atmospheric concentration is not the
64 only parameter to consider for assessing the impact of air pollution on human health. Indeed,
65 water soluble major ions as ammonium sulfate, ammonium nitrate and sea-salts are relatively
66 not correlated to the PM toxicity whereas trace elements as Pb, Cd, Ni, As and polycyclic
67 aromatic hydrocarbons as benzo[a]pyrene are known for their high toxicity (Lippmann, 2014)
68 and correspond to regulated species in Europe.

69 The assessment of toxicological effects of PM related to the oxidative stress is widely
70 investigated considering *in vitro* lung cell exposure (Badran et al., 2020; Gualtieri et al., 2017;
71 Jia et al., 2017). In parallel, acellular tests have been suggested over the last fifteen years to
72 evaluate the PM oxidative potential (OP), which is a measure of the ability of PM to deplete
73 certain antioxidant molecules in synthetic airway fluids (Ayres et al., 2008; Sauvain et al.,
74 2013). The dithiothreitol (DTT) simulates reductant species in cells like Nicotinamide adenine
75 dinucleotide or nicotinamide adenine dinucleotide phosphate. The DTT assay evaluates the
76 ability of PM redox active species to transfer electrons to dissolved molecular oxygen
77 (Kumagai et al., 2002; Li et al., 2003). In the absence of PM redox active species, DTT can
78 act as a reducer of dissolved oxygen leading to formation of H₂O₂. PM components can act as
79 catalyst to increase the reduction of oxygen species to ROS like superoxide anion and H₂O₂.
80 The latter contribute to the overall level of ROS that could be responsible for an oxidative
81 stress in cells. The ascorbic acid (AA) simulates antioxidants in the respiratory tract lining
82 fluid (RTLFL) and the AA assay measures the ability of PM to deplete antioxidants in the
83 RTLFL (Mudway et al., 2004). The application of DTT and AA assays in many studies showed
84 that they were correlated to both common and different PM constituents. The OP assessed
85 using DTT (OP-DTT) has been correlated not only with metals but also with oxygenated
86 organic compounds (Charrier and Anastasio, 2012; Cho et al., 2005; Crobeddu et al., 2017;

87 Verma et al., 2015). The OP derived from AA assay (OP-AA) has been found correlated
88 rather to metals (Fang et al., 2016; Pietrogrande et al., 2018). An OP assessment based on the
89 use of several assays, as AA and DTT in the present study, appears as an appropriate
90 approach for considering contributions of reactive species from several families of PM
91 components. Information on the relationship between PM components and OP values are
92 sometimes limited since the correlation is based on elements that are specific of a source but
93 not always responsible for the OP response (Gao et al., 2020a). In addition, numerous
94 elements and compounds are emitted by different sources and are not always considered at the
95 analysis step. Therefore, several authors have suggested the investigation of PM source
96 contribution to OP (Calas et al., 2019; Cesari et al., 2019; Yu et al., 2019) as a relevant
97 indicator to identify the sources with high potential toxic effects on human health.

98 In this context, the objective of this study was to relate OP data to PM composition and PM
99 source contribution in an industrial and urban area located in Northern France. The OP-DTT
100 and OP-AA assays were used and relationships between OP results, PM_{2.5} composition and
101 emission sources contributions were investigated by methods including correlation, principal
102 component analysis, bivariate polar plots, concentration roses and concentration weighted
103 trajectories.

104

105 1. Materials and methods

106 1.1. PM_{2.5} samples

107 1.1.1. Study area and sampling.

108 PM_{2.5} samples considered in this work were collected in Dunkerque (latitude:
109 51°2'10"N; longitude: 2°22'46"E) in winter and spring 2010-2011. Since 2010-2011, the
110 PM_{2.5} samples have been stored at -18°C in the dark, in hermetically sealed boxes. This
111 procedure ensures to prevent any volatilization or transformation of the organic compounds

112 over time. The detailed description of the sampling methodology and the study area is given
113 in Kfoury et al. (2016). Briefly, Dunkerque, 210,000 inhabitants with its suburbs, is an
114 industrialized city located in Northern France in the southern coast of the North-Sea. Due to
115 its location, this site is significantly influenced by common sources as sea spray, marine
116 traffic, domestic and road traffic emissions as well as more specific sources, notably industrial
117 emissions in the field of metallurgy, cement production and organic chemistry. Between the
118 industrial sources, an integrated steelworks (ISW) which is the biggest steel producer in
119 Europe, and an electric steel plant (ESP) are noticeable and located in the 260–290° wind
120 sector and east to the sampling site respectively (Figure S1). According to the European
121 pollutant release and transfer register (E-PRTR), cumulatively in 2010, all the industries
122 located in Dunkerque region emitted in the air 3,100 tons of total suspended particles
123 including 8,117 kg of Zn, 6,962 kg of Pb, 1,398 kg of Ni, 1,241 kg of Cu, 1,194 kg of Cr, 175
124 kg of Cd and 101 kg of As (E-PRTR, 2011). PM_{2.5} was collected on a semi-daily basis
125 (00:00–12:00 UTC and 12:00–24:00 UTC time period) using the high volume (30 m³·h⁻¹)
126 DA80 sampler (DA80, DIGITEL®, Switzerland). Two types of 150 mm diameter filters were
127 used: cellulose (grade 41, Whatman®, GE Healthcare Life Sciences, United Kingdom) for
128 major and trace elements analysis and OP determination, and high purity quartz microfiber
129 filters (grade QM-A, Whatman®, GE Healthcare Life Sciences, United Kingdom) for carbon
130 quantification.

131

132 1.2. PM_{2.5} composition and source contribution

133 PM_{2.5} samples were analyzed for their content in major and trace species as Ag, Al,
134 As, Ba, Bi, Ca, Cd, Co, Cr, Cu, Fe, K, Mg, Mn, Na, Ni, Pb, Rb, Sb, Sn, Sr, Ti, V, Zn, NH₄⁺,
135 Cl⁻, NO₃⁻, SO₄²⁻ and total carbon (TC) using Inductively Coupled Plasma-Mass spectrometry
136 (ICP-MS), Inductively Coupled Plasma Atomic Emission spectroscopy (ICP-AES), ion

137 chromatography and a CHNS/O elemental analyzer. The detailed protocols are described in
138 Kfoury et al. (2016) and Ledoux et al. (2017).

139 In a previous study, constrained weighted-non-negative matrix factorization (CW-
140 NMF) was applied to the whole PM_{2.5} composition dataset (173 PM_{2.5} samples) and led to the
141 identification of 11 contributing sources (Kfoury et al., 2016). The eight first ones can be
142 considered as sources typically encountered at an urban coastal site: sea salts, aged sea salts,
143 crustal, heavy fuel oil combustion, combustion 2 (mix between biomass and road traffic
144 combustion) and traffic non-exhaust (related to wear of the car mobile parts such as brakes,
145 tires, clutch ...) for which both a local and distant origin can be considered, and secondary
146 sulfates and secondary nitrates which are typically related to long range transport. Three other
147 source profiles were related to the local steelmaking activities located close to the sampling
148 site (less than 7 km): one for the ESP emissions and two for the ISW sources with a
149 distinction for the latter between the sintering stack (ISW-Sintering Stack) and the fugitive
150 emissions from the whole ISW site (ISW-Fugitive).

151 In this work, 57 out of the 173 PM_{2.5} samples were considered for studying their oxidative
152 potential. The sample selection was made based on PM_{2.5} concentrations and source
153 contributions. PM_{2.5} samples showing higher relative contribution for each of the eleven
154 sources were included in addition to the high (> 50 µg·m⁻³) and low (<10 µg·m⁻³) PM_{2.5}
155 concentration samples.

156

157 1.3. Oxidative potential measurements

158 OP was assessed by using two a-cellular methods based on the depletion rate of DTT
159 (OP-DTT) and AA (OP-AA) in the presence of the PM constituent. When performing these
160 tests, an important issue is first related to the choice of the leaching agent that will be used for
161 the PM extraction. The latter has to be representative of the physiological conditions

162 occurring in the PM target organs (Calas et al., 2017; Pietrogrande et al., 2019). As well, the
163 extraction procedure has also an important influence on the OP results and the review of the
164 literature showed large differences between the methods used (Bates et al., 2019). The fact
165 there is no standardized method for OP measurements leads to consider a compromise
166 between all the critical parameters such as solvent, extraction procedure and solid/liquid (S/L)
167 ratio (Hernandez-Pellon et al., 2018).

168

169 1.3.1. Sample preparation for OP tests

170 The Gamble's solution (pH=7.4) is one of the most used simulated lung fluids (SLF)
171 and represents the interstitial fluid in the deep lung. Therefore, it was chosen as a leaching
172 agent for the PM_{2.5} extraction. The Gamble's solution was daily freshly prepared according to
173 the specifications given in Colombo et al. (2008), using ultrapure water and mixing the stock
174 solutions following the order given in Table S1 (see supplementary materials) to avoid salt
175 precipitation. The citrate was used to simulate proteins that can be found in the lung lining
176 fluid and acetate instead of organic acids (Hedberg et al., 2010; Marques et al., 2011). All
177 chemicals were of analytical grades. The pH of the Gamble's solution was adjusted to 7.4 by
178 adding drop-wise HCl 10% (Wragg and Klinck, 2007) or NaOH 1M (Caboche et al., 2011).

179 For each of the 57 PM_{2.5} samples, about 1/10 of the cellulose filter, containing an
180 average of 1.40 mg of PM_{2.5}, was placed in a 15 mL glass tube with 14 mL of Gamble's
181 solution. The volume was chosen in accordance with real human exposure conditions with
182 respect to the ambient concentration of particles, the inhaled air volume and the pulmonary
183 fluid volume. In these conditions, the S/L ratio was in the 1/500 - 1/50000 range avoiding any
184 risk of saturation and competition between the soluble compounds (Caboche et al., 2011).

185 The extraction tubes were covered with aluminum foil and placed in a heated orbital shaker
186 for 24 h at 37°C, at a speed of 250 oscillations·min⁻¹. Solutions were then filtered on 0.45 µm

187 nylon filters, and the filtrates were placed in flasks, and stored at -20°C until OP analyses.
188 Blanks were considered by treating several field blank filters with the same protocol as
189 applied on PM_{2.5} filters.

190

191 1.3.2. Dithiothreitol OP test

192 Redox-active species in PM have the ability to oxidize DTT to DTT disulfide, and
193 catalyze the reduction of oxygen species to ROS like superoxide anion and H₂O₂. When
194 5,5-dithio-bis-(2-nitrobenzoic acid) (DNTB) is added, the remaining DTT reacts with DTNB,
195 generating DTT-disulfide and 2-nitro-5-thiobenzoic acid (TNB) (Figure S2). The latter is
196 quantified, by measuring its absorbance at 412 nm, which is also proportional to the
197 remaining quantity of DTT (Visentin et al., 2016).

198 The DTT protocol was optimized in a microplate version which makes easier the
199 consideration of several points (i.e. 7 points in this work) for evaluation of the DTT depletion.
200 DTT and DNTB stock solutions were prepared at a concentration of 0.4 mM and 1.5 mM,
201 respectively, in a phosphate buffer (pH=7.4). Since both solutions are sensitive to light and
202 temperature (Damodaran, 1985; Eyer et al., 2003), they were kept in amber glass vials in the
203 dark at -20°C until use. The OP-DTT measurement was carried out in 96 well black plates
204 with clear flat bottom (Coastar® 3631, Corning Incorporated Life Sciences, USA). 120 µL of
205 potassium phosphate buffer (pH=7.4), and 40 µL of the PM leachate (or blank or positive
206 control) were added in each well. The plate was then placed at 37°C for 10 min under
207 shaking. 25 µL of DTT (0.4 mM) was then added to initiate the oxidation reaction.

208 The remaining DTT concentration was estimated after 0, 5, 10, 20, 30, 45 and 60
209 minutes of reaction between PM leachate components and DTT, at 37°C. After each time of
210 reaction, 15 µL DTNB were then added to each well to convert remaining DTT to TNB. After
211 10 min at 37°C, the absorbance of TNB was measured at 412 nm (spectrophotometer

212 Multiskan Go, Thermo Fisher Scientific, Finland), and converted to DTT remaining quantity.
213 The DTT depletion rate was calculated using the slope of the curve (linear regression) of
214 remaining DTT versus time. 1.4 naphthoquinone (1.4 NQ), 100 μM , was used as a positive
215 control and obtained values ($1.23 \mu\text{M}\cdot\text{min}^{-1}$) were in the range of those obtained in the
216 literature.

217 For each $\text{PM}_{2.5}$ sample, OP-DTT values were obtained by correcting the $\text{PM}_{2.5}$ filter
218 slope values with the blank filter one. Samples were analyzed in triplicates and the given OP-
219 DTT values correspond to the arithmetic mean. The relative standard deviation was below 4.1
220 %. The OP-DTT values were finally expressed per unit of mass of PM (OP-DTT_m in
221 $\text{pmol}\cdot\text{min}^{-1}\cdot\mu\text{g}^{-1}$) and per unit of air volume (OP-DTT_v in $\text{nmol}\cdot\text{min}^{-1}\cdot\text{m}^{-3}$). The first unit can
222 be related to the intrinsic OP of 1 μg of emitted PM and will be used to compare OP of the
223 sources. The second one can be related to the intrinsic OP of 1 m^3 of the air people breathe
224 and is more adapted for studying the situation (wind direction and air mass origin...) in which
225 the impact on health could be high.

226

227 1.3.3. Ascorbic acid OP test

228 AA is present in the lung epithelial lining fluid (ELF), along with other antioxidants
229 like reduced glutathione (GSH), and uric acid (UA). Since AA has the same response,
230 whether in a mixture of these antioxidants or alone, it can be used as a simplified model of the
231 synthetic respiratory tract lining fluid (Zielinski et al., 1999). As for DTT, the redox active
232 species in PM have the ability to oxidize AA to dehydroascorbic acid (Figure S2), and
233 catalyze the formation of ROS by transferring an electron to oxygen molecules. AA stock
234 solution (1 mM) was prepared in ultrapure water, and stored at -20°C due to its instability at
235 room temperature (Bode et al., 1990). The OP-AA measurement was done in 96 well plates
236 with UV-transparent flat bottom (UV-Star® 655801, Greiner bio-one, Austria). 160 μL of PM

237 leachate was placed in each well of the plate, and the latter was heated to 37°C (5 min), before
238 adding 40 µL of AA solution and shaking for 1 min. Finally, the plate was placed in the
239 spectrophotometer (Multiskan Go, Thermo Fisher Scientific, Finland), at 37°C and the
240 absorbance was measured at 265 nm every 2 minutes for 2 h (Jung and Wells, 1998) with a
241 30 s shaking step before each measurement (Yang et al., 2014). OP-AA values were deduced
242 from the AA depletion curve versus time considering the slope of the linear part of the curve,
243 where AA depletion rate is maximal. Samples were analyzed in triplicates (RSD < 3.7 %),
244 blank filters were considered, 1.4 NQ was used as positive control (2.74 µM·min⁻¹) and OP-
245 AA values were expressed per unit of mass of PM and volume (OP-AA_m and OP-AA_v
246 respectively).

247

248

249 1.4. Data handling and processing

250 Relationships between OP results, PM_{2.5} composition and source contributions were
251 investigated by using several methods including correlation, principal component analysis
252 (PCA), bivariate polar plots, concentration roses, OP roses and concentration weighted
253 trajectories (CWT). The R software was used in addition to FactomineR (Lê et al., 2008) and
254 openair (Carslaw and Ropkins, 2012) packages to perform these calculations.

255

256 1.4.1. Correlation

257 Prior the determination of the correlation coefficient between the variables under
258 study, the Shapiro-Wilk test was performed to assess the normality of the data distribution. As
259 the normality was not verified ($p < 0.001$) for most of the variables, the Spearman correlation
260 coefficient was used. Two levels of significance were considered, $p < 0.05$ and $p < 0.001$,

261 allowing to conclude about the correlation with 95% and 99.9% of confidence respectively. In
262 addition, correlations were also investigated by using PCA after data normalization.

263

264 1.4.2. Bivariate polar plots and concentrations roses.

265 Bivariate polar plots and OP and concentration roses were used to investigate the
266 relationships between OP values and contribution of local sources.

267 Bivariate polar plots represent the variation of a quantitative parameter (concentration, OP
268 value, source contribution) versus the wind direction and wind speed. For each PM_{2.5} sample,
269 the parameter values obtained for the 12 h period were distributed in front of the 12
270 corresponding hourly wind speed and directions. The data set used to draw each polar plot
271 counts 57x12 lines. The OP and concentration roses, which give the average value in a
272 particular wind direction, were calculated using the same dataset after averaging the values at
273 45° intervals.

274

275 1.4.3. Concentration weighted trajectories

276 The concentration weighted trajectories (CWT) (Cheng et al., 2013; Dimitriou et al.,
277 2015) were used to investigate relationships between OP and distant sources. The CWT
278 methods assume that when an air mass is passing through a polluted area, the material can be
279 transported along the trajectory to the receptor site. By using air mass back-trajectories and
280 affecting values observed at the receptor site along the trajectories, it is possible, by the end
281 after averaging, to locate areas responsible for high values (Carslaw, 2019).

282 Air mass back-trajectories were calculated using the hybrid single particles Lagrangian
283 integrated trajectory (HYSPLIT) model (Rolph et al., 2017; Stein et al., 2015) considering the
284 Gridded Meteorological Data Archives (GDAS 0.5 degrees) and a height of 100 meters above
285 ground level at receptor site. For each PM_{2.5} sample, four 72 h backward trajectories were

286 assigned (one trajectory every 3 hours) and OP and source contribution values obtained for
287 the 12 hours period were merged with the 4 corresponding trajectories. The dataset used for
288 CWT calculation count 72x4x57 lines.

289

290 2. Results and discussion

291 2.1. PM_{2.5} composition et sources contribution

292 Table 1 regroups the average composition of the PM_{2.5} samples and Table 2
293 summarizes the chemical markers that were used for identifying the sources as well as their
294 average contributions (Kfoury et al., 2016). The PM_{2.5} average concentration of the selected
295 samples was 29.2 $\mu\text{g}\cdot\text{m}^{-3}$, which is similar to the average concentration typically encountered
296 during the winter-spring period in the studied area (Kfoury et al., 2016). Among the PM
297 components, NO_3^- , TC, SO_4^{2-} , NH_4^+ , Cl^- , and Na showed the highest average concentrations
298 ($> 300 \text{ ng}\cdot\text{m}^{-3}$) representing together about 67% of the PM_{2.5} mass. These elements appeared
299 characteristics of the secondary nitrate and sulfate source profiles as well as the combustion
300 and the marine sources (Table 2). NH_4^+ and Cl^- are also characteristic elements of the ISW
301 sintering stack emissions (Hleis et al., 2013). Fe, K, Ca, Mg and Al were found with
302 concentrations between 26 and 155 $\text{ng}\cdot\text{m}^{-3}$. These elements are known to belong to the
303 aluminosilicates category. Nevertheless, at Dunkerque urban-industrial site, they were shown
304 to be also associated with emissions from the ISW (Hleis et al., 2013) and notably dusts from
305 the sintering unit. Among the trace metals, Zn (67.6 $\text{ng}\cdot\text{m}^{-3}$), Pb (14.8 $\text{ng}\cdot\text{m}^{-3}$) and Mn
306 (13.7 $\text{ng}\cdot\text{m}^{-3}$) stand out and were mainly associated with the metallurgical activities occurring
307 in the ESP and the ISW (Table 2).

308

309 2.2. OP-AA and OP-DTT average values

310 The average mass- and volume-normalized OP-DTT and OP-AA values at Dunkerque
311 site are presented together with values published for other European and world sites in
312 Table 3, together with the leaching agents used for extracting PM samples. Ultrapure water,
313 ethanol, methanol, water, phosphate buffer and Gamble solution were used for leaching in
314 addition to ultrasonic treatment or shaking for various times. In one study only , the solution
315 was not filtered after extraction (Janssen et al., 2014). The heterogeneity of the methodologies
316 used could influence the obtained values and could make the comparison not reliable. As the
317 leaching agent is different, the response could also be different as published by Calas et al.
318 (2017) who compared responses in OP_DTT test on extracts obtained by using artificial
319 lysosomal fluid (ALF), ultrapure water, Gamble solution and Gamble solution implemented
320 with dipalmitoylphosphatidylcholine. Based on the analysis of 5 different PM samples,
321 including four real PM samples, it was shown that OP values observed using the ultrapure
322 water were always slightly higher than those obtained using the Gamble solution.
323 Nevertheless, focusing on the real PM samples, the difference was not exceeding 13% which
324 make reliable the comparison between values obtained using ultrapure water and Gamble
325 solution. In most of the comparison studies included in table 3, ultrapure water was used as
326 leaching agent. Solubility is also highly dependent of the pH of the leaching agent
327 (Hernandez-Pellon et al., 2018). As the phosphate buffer has the same pH value as the
328 Gamble solution (7.4), we can assume that OP values could be compared without severe bias.
329 However, Jedynska et al. (2017) used methanol as leaching agent and the leachate probably
330 concentrates more organic constituents than leachates obtained using Gamble, water or
331 phosphate buffer. In contrast to the other studies, Janssen et al. (2014) did not filter extracts
332 before testing OP which leads to the inclusion of the contributions of insoluble species in OP
333 values. Consequently, except for Jedynska et al. (2017), who used methanol and Janssen et al.

334 (2014), we can be confident with the reliability of the comparison between values from the
335 cited papers.

336 Focusing on values related to PM_{2.5} concentration, the average \pm standard deviation
337 OP-DTTv value, $0.36 \pm 0.24 \text{ nmol}\cdot\text{min}^{-1}\cdot\text{m}^{-3}$, appeared in accordance with values observed in
338 urban and urban background sites such as Paris (France), Atlanta (USA), Lecce (Italy),
339 Athens (Greece) (Chirizzi et al., 2017; Gao et al., 2020b; Jedynska et al., 2017;
340 Paraskevopoulou et al., 2019). On the other hand, OP-DTTv value is lower than the one
341 determined in Beijing and Hangzhou (China) or in Patalia (India) (Patel and Rastogi, 2018;
342 Wang et al., 2019; Yu et al., 2019). Except for the Netherlands, the corresponding PM_{2.5}
343 concentrations are also far higher ($> 113 \mu\text{g}\cdot\text{m}^{-3}$) than Dunkerque one. The average OP-
344 DTTm, $17.7 \pm 15.4 \text{ pmol}\cdot\text{min}^{-1}\cdot\mu\text{g}^{-1}$, OP-AAv, $1.12 \pm 1.15 \text{ nmol}\cdot\text{min}^{-1}\cdot\text{m}^{-3}$ and OP-AAm
345 values, $55 \pm 66 \text{ pmol}\cdot\text{min}^{-1}\cdot\mu\text{g}^{-1}$ were also comparable to the range published for European,
346 Asian and American sites.

347

348 2.3. Correlation of OP with chemical species and source contributions

349 In order to study the contribution of PM chemical components to the DTT and AA
350 oxidation activity, the correlation between OP-DTTv and OP-AAv values and the chemical
351 species was investigated based on the Spearman correlation coefficients (Table 4) and PCA
352 (Figure 1 and Table S2).

353

354 2.3.1. Spearman correlation

355 OP-DTTv and OP-AAv showed statistically significant correlation ($p < 0.05$) with
356 almost all the species under study (Table 4). However, they were not correlated with Ca and
357 Cl⁻, while Al, Co, V were only correlated with OP-DTTv, and Sr only with OP-AAv. Stronger
358 correlations were globally observed between the PM components and OP-DTTv compared to

359 OP-AAv. This latter observation is consistent with results from other research groups (Calas
360 et al., 2019; Fang et al., 2016). In particular, Ag, Al, As, Ba, Bi, Cd, Co, Cr, Cu, Fe, K, Mn,
361 Ni, Pb, Rb, Sb, Sn, Zn, NO_3^- and TC showed correlation coefficient higher than 0.6 with OP-
362 DTTv. Such a correlation level was observed only for As, Ba, Cu, Na, Pb, Sb, Sn and TC with
363 OP-AAv. A very strong correlation was found between OP-DTTv and Cu ($r = 0.85$), Cd
364 ($r = 0.82$) and Sn ($r = 0.82$).

365 The major PM components TC, SO_4^{2-} , NO_3^- and NH_4^+ appeared more correlated with
366 OP-DTTv than OP-AAv. This is consistent with the stronger correlation found between $\text{PM}_{2.5}$
367 concentrations and OP-DTTv in comparison with OP-AAv ($r = 0.67$ versus $r = 0.45$
368 respectively). Our results are in agreement with other observations showing that secondary
369 inorganic aerosols (SIA) components were correlated with OP-DTT (Ma et al., 2018;
370 Paraskevopoulou et al., 2019; Wang et al., 2019). On the contrary, Mg, Na and Sr showed
371 negative correlation coefficients with OP values. These elements are mainly related to the
372 marine influence at the study site. Finally a moderate correlation ($r = 0.46$, $p < 0.001$) was
373 found between OP-DTTv and OP-AAv which indicates that the two tests do not exactly
374 respond to the same species and prove the complementarity between both.

375 AA and DTT tests respond to the oxidative species solubilized in the gamble solution
376 after leaching of the collected PM. The leachate includes not only the PM components that
377 were identified and quantified, but also unanalyzed PM species. These latter, such as water
378 soluble organic carbon (WSOC) including humic-like substances (HULIS) or oxygenated
379 PAH, are an integral part of the leachate and also contribute to the OP values as well (Bates et
380 al., 2019; Gao et al., 2020a; Jiang et al., 2020; Lin and Yu, 2019; Ma et al., 2018). In this
381 study, the inorganic PM composition has been investigated and, based on this, the CW-NMF
382 model allowed to determine the source contributions. Consequently, studying the correlation

383 between OP values and source contributions allows indirectly to consider all the species
384 including the WSOC related to a specific source whether they have been analyzed or not.
385 Strong correlations ($r \geq 0.6$) were found between the "traffic non-exhaust", "secondary
386 nitrates", "ISW fugitive", "ISW sintering stack" and "electric steel plant" sources contribution
387 and OP-DTT values. The three industrial sources are known to emit metals with redox
388 properties (E-PRTR, 2011; Hleis et al., 2013; Kfoury et al., 2016) able to potentiate ROS
389 overproduction. "Secondary sulfates" and "heavy fuel oil combustion" were moderately
390 correlated ($0.4 \leq r < 0.6$) with only OP-DTTv. Considering the mass normalized data (Table
391 S3 in supplementary materials), OP-DTTm appeared strongly correlated with "ISW fugitive",
392 moderately correlated with "traffic non exhaust", "heavy fuel oil combustion", "crustal" and
393 "ISW sintering stack" while no correlation was found with secondary nitrates and only poor
394 correlation with secondary sulfates ($r=0.35$). Except for "aged sea-salts" ($r = -0.67$) and
395 "traffic non-exhaust" ($r = 0.66$), OP-AAv only showed, at best, moderate correlations with the
396 contribution of each of the 11 sources. Interestingly "sea-salts" and "aged sea-salts" sources
397 show negative correlations with OP-DTTv and OP-AAv values. However, it would be wrong
398 to conclude that emissions related to these sources are able to decrease the OP values. Indeed,
399 in the urban coastal area under study, high marine contributions are associated with high
400 speed winds and favorable condition of dispersion (Ledoux et al., 2006). Consequently, when
401 the marine origin is predominant, the $PM_{2.5}$ concentration tends to be relatively low (Figure
402 S3 and Figure S4) and metal concentration values as well, which could explain the low OP
403 values. Finally, no significant correlation was found with the "crustal" source suggesting that
404 this latter does not significantly contribute to the OP values.

405

406 2.3.2. Principal Component Analysis

407 Unlike Spearman correlation which tests the correlation between pairs of variables, the
408 PCA is a multivariate analysis. It also allows to evidence discrete relationships between
409 variables that will be regrouped as components. PCA was applied to the composition data of
410 the 57 PM_{2.5} samples for the 29 studied species concentration. The source contributions and
411 OP values were considered as supplementary variables.

412 The 5 first obtained axes explain 71.7% of the data variability. As OP-AAv and OP-
413 DTTv showed only significant correlation coefficients with the two first principal
414 components, the focus was only done on these latter (Figure 1 and Table S2). The variable
415 factor map of the two major axes which represent together 47.1% of the total variance is
416 presented in Figure 1c. Axis 1 is highly positively associated with metals (from high to low r
417 value with $r > 0.6$: Bi, K, Rb, Ag, Sn, As, Fe, Al, Ti, Pb, Mn, Ba, Co, TC and Ni) and
418 negatively associated with Na. So the first axis evidenced that several metals are correlated
419 and that when a concentration is high for one these elements, all the concentrations tends to
420 be also high for the others. From lower to higher values, axis 1 makes the distinction between
421 the typical sea components and the typical metals emitted by anthropogenic and industrial
422 activities. NH₄⁺, NO₃⁻, SO₄²⁻ and TC are negatively correlated with the second axis while Na,
423 Sr, Ca and Mg are positively correlated. Hence, the second axis makes the distinction between
424 the typical sea components (positive part of the axis) and the SIA (negative part of the axis).
425 In the region where the study takes place, SIA are mainly associated with winds blowing from
426 the East (continent) and long range transport while marine aerosol are mainly originates from
427 the North and western direction. When the marine contribution is high, the secondary aerosol
428 contribution tends to be low (Figure S3). This explains why SIA are placed against sea-salts
429 components on the second axis. NH₄⁺, NO₃⁻, SO₄²⁻ and TC contribute the most to the PM_{2.5}
430 mass, and a strong correlation ($r = -0.70$) was found between the PM_{2.5} concentration and the
431 second axis.

432 OP-DTTv was very strongly correlated ($r = 0.84$) with the first axis as a consequence
433 of the clear relationships between DTT test values and metal concentrations while OP-AAv
434 appeared only moderately anticorrelated ($r = -0.48$) with the second axis. Considering now
435 both dimensions but only 47.1% of the data variability, OP-AAv was related to the
436 "secondary nitrates", "secondary sulfates" and "combustion 2" (road traffic and biomass
437 combustion) sources. OP-AAv was also very strongly anticorrelated with "sea-salts" and
438 "aged sea-salts" sources. This result was not so well evidenced considering the whole data
439 variability by studying the correlation coefficient. Then for OP-DTTv, PCA findings were
440 mixed as it was only correlated with axis 1. Considering axes 1 and 2, only one clear
441 relationship appeared with the ISW sintering stack while several ones were evidenced from
442 the correlation coefficient study. A deepest investigation will be done hereafter.

443

444 2.4. OP values and relationship with local and distant sources

445 Concentration roses and bivariate polar plots were used to study the link between OP
446 observed values and local sources contribution while CWT was used to search for regional
447 influences.

448

449 2.4.1. Mass-normalized and volume-normalized OP

450 Roses representing OP value versus wind direction and bivariate polar plots
451 considering OP expressed in the two units (per m^3 or per $\mu g PM_{2.5}$) are given in Figure 2.
452 OP-AA and OP-DTT did not exhibit exactly the same behavior regarding the wind direction.
453 The maximum for mass-normalized OP-AAm and OP-DTTm was obtained when the wind
454 blew from the northwest (NW) wind sector, while the maximum for volume-normalized OP
455 was related to the southeast (SE) wind direction for OP-AAm, and SE, west (W) and
456 northeast (NE) direction for OP-DTTm. The W-NW and the SE sectors correspond to the

457 industrial park including the ISW and the urban sector, respectively, while in the northeast
458 direction is located the electric steel plant.

459 Especially when considering the OP-AA values, 1 μg of PM coming from the wind
460 sector including the industrial park showed the maximum intrinsic oxidative potential
461 compared to an equivalent mass of PM sampled from the others wind sectors. By normalizing
462 OP values to the air volume, this sector no longer exhibited the highest value. This is
463 important information derived from this study. Considering 1 m^3 of air that people breathe,
464 emissions from the industrial park do not appear as the most impacting on OP values. At our
465 study site, emissions from the SE sector corresponding locally to the urban sector or long
466 range transport influence (as it will be seen in section 2.4.3) appear to be the most
467 problematic, considering OP-AAv values. However, the OP-DTTv values showed not only a
468 maximum under the urban sector but also when the winds blew from the west direction
469 suggesting a high OP-DTTv potential for a source located in the industrial park.

470

471 2.4.2. OP relationships with the 11 identified CW-NMF sources

472 Figure 3 represents the volume-normalized source contributions bivariate polar plots
473 of the 11 identified $\text{PM}_{2.5}$ emission sources in addition to the one for $\text{PM}_{2.5}$ concentrations, for
474 comparison with the corresponding OP-DTTv and OP-AAv one. The use of bivariate polar
475 plots gives an additional strong criterion for validating the relationships between OP values
476 and PM source contribution evidenced with the Spearman correlation and PCA mathematical
477 tools. By using this, OP and source contribution values are distributed versus the wind
478 direction and wind speed. The geographical position of the sources as well the characteristics
479 of the emission (ie: resuspension or blowing back of chimney emission to ground level under
480 the effect of wind speed) are directly and indirectly considered when bivariate polar plots are
481 investigated (Carslaw et al., 2006; Ledoux et al., 2018). It makes possible to deepen

482 investigation and strengthen conclusions regarding relationships between OP values and
483 source contributions.

484 Clearly OP-AAv showed a pattern similar to that of PM_{2.5} and road traffic and biomass
485 combustion ("combustion 2"). HULIS could be the reason of the relationship between OP and
486 combustion sources as they are highly redox active species known to participate to the OP
487 values and mainly emitted by combustion processes (Dou et al., 2015; Ma et al., 2018). This
488 observation is also consistent with the one of Yu et al. (2019), who found that vehicle
489 emissions (traffic exhaust and non-exhaust as a whole) highly contribute to ROS production.
490 Based on the polar plot patterns, traffic non-exhaust could also participate to OP-AAv but the
491 absence of correlation between both would suggest to reject the idea. These first observations
492 are in agreement with the PCA findings (Figure 1). However, PCA was suggesting a
493 relationship between OP-AAv and the secondary inorganic aerosols for 47.1% of the data.
494 The bivariate polar plots tend to discredit this as the corresponding plots are quite different
495 (Figure 2 and Figure 3). Except with the PM_{2.5}, combustion 2 and possibly traffic non-exhaust
496 source, OP-AAv pattern do not show any similarity with other sources, which could be
497 contradictory with partial conclusion done from the correlation coefficient study. This tends
498 to demonstrate that investigations using only the correlation coefficient between sources
499 contribution (or species concentrations) and OP are not enough to conclude about
500 relationships.

501 Unlike OP-AAv, OP-DTTv bivariate polar plot showed similarities and compatibilities
502 with several sources. Between them, the ISW sintering stack seems to be responsible for the
503 spot observed in the west direction at a wind speed close to 5 m·s⁻¹. In the same direction, the
504 ISW fugitive and the heavy-fuel oil combustion source are also possible contributors. In
505 addition, in the east-northeast wind direction, the OP-DTTv concentration spot overlap the
506 one observed for the electric steel plant making emissions relate to this source another

507 OP-DDTv influencing factor. Relatively high values observed in the SE quarter could be the
508 fact of combustion 2 or traffic non-exhaust sources. In this case, observations matched quite
509 well those done from PCA study but also contradict again some conclusions done using the
510 correlation coefficient. For example strong correlation was found between secondary nitrates
511 and OP-DDTv ($r = 0.67$, $p < 0.001$), but their respective bivariate polar plots could challenge
512 it. Finally, our observation made possible to definitely dismiss the contribution of the sea-
513 salts, aged sea-salts, crustal and secondary sulfates to the OP values.

514 Up to our knowledge, this is the first time that bivariate polar plots of OP and sources
515 contributions were compared to study and validate the influence of source emission on OP
516 values. In most of the published works, only correlation coefficients are used to conclude on
517 relationships between OP and species /sources. This could be inadequate in light of our results
518 which clearly demonstrate the interest to use bivariate polar plots as additional information.

519

520 2.4.3. Regional influence

521 The computing of OP-AAV and OP-DDTv with the air-mass back trajectories using the CWT
522 method allowed to study the impact of regional emissions on OP values (Figure 4). Regions
523 contributing the most to OP-DDTv are inland, including the upper part of the western
524 continental Europe notably France, Belgium, Germany and Poland. For OP-AAV, the main
525 hotspot is located in Belgium and the center of Germany. Long range transport from these
526 source regions located east to the study site could also explain the high OP values locally
527 observed for eastern and southeastern wind on the bivariate polar plots. The North-Sea, the
528 English Channel and the coastal part of the Atlantic Ocean appears also as source region
529 especially for the OP-DDTv. This is in agreement with the relationships previously found
530 between OP-DDTv and heavy fuel oil combustion source and the CWT obtained for the latter
531 is also consistent with this. To be mentioned here that the English Channel and the North Sea

532 are among the most used shipping routes for marine transportation (EEA, 2013b; Ledoux et
533 al., 2018).

534 CWT results for secondary nitrates evidenced mainly Belgium and southern part of
535 Netherlands as source regions which is consistent with Pay et al. (2010) and Pay et al. (2012)
536 who identify these countries as important NO_x emitters. Secondary sulfates were originated
537 from Belgium, North of Germany and Poland who also showed exceedances of the SO₂ limit
538 value (EEA, 2012). In both cases, the similarity with the OP-AA_v and OP-DTT_v CWT is
539 poor.

540 The highest concentrations of PM_{2.5} were traced back to North of France, Belgium,
541 Luxembourg, Netherlands and Germany also. This is consistent with the potential source
542 contribution results obtained by Waked et al. (2018) and Pay et al. (2012) who evidenced
543 these countries as important source of PM₁₀ in relation with industrial and shipping related
544 emissions especially for Belgium and Netherlands. Combustion 2 and traffic non-exhaust
545 CWT patterns showed higher concentrations on the continent which is totally in agreement
546 with the definition of the sources. Source regions of biomass and traffic related combustion
547 traffic-non-exhaust emissions match quite well the OP-DTT_v and OP-AA_v CWT confirming
548 the relationships suggested at a local scale. Eastern Europe, especially Bulgaria, Czech
549 Republic and Poland, was also identified as an important source of PM in Europe (EEA,
550 2013a). Nevertheless, we were unable to observe their influence as there were no air masses
551 coming from these regions during the sampling period.

552

553 Conclusion

554 PM_{2.5} collected in a coastal industrial city in Northern France was responsible for OP
555 values in the same order of magnitude as those encountered in other world locations. PM_{2.5}
556 concentrations, some PM_{2.5} components and PM_{2.5} sources were found to be specifically

557 correlated with one of the two tests (AA and DTT). For one of the first time, bivariate polar
558 plots drawn for OP and source contributions were considered together to investigate their
559 relationships. Conclusions deduced from correlation coefficient studies and PCA were
560 challenged after using the OP and source contribution bivariate polar plots suggesting the
561 importance to use such a methodology. Ascorbic acid assay results appeared clearly linked to
562 the PM_{2.5} concentrations and combustion sources. Dithiothreitol assay was clearly related to
563 sources emitting metals, in our case ISW diffuse and sintering stack, electric steel plant
564 emissions, heavy fuel oil combustion and possibly other combustion sources. Sea-salts, aged
565 sea-salts, crustal, secondary sulfates and secondary nitrates sources do not seem to contribute
566 to OP values. Some high OP values source regions have also been identified (Germany,
567 Belgium and Netherlands) showing that long range transport is also a parameter to consider
568 when studying OP of particulate matter.

569

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List of figures:

Figure 1: Principal Component Analysis related to the 57 PM_{2.5} samples collected in Dunkerque area. Correlation coefficient between the five first axis and (a) PM_{2.5} components considered as active variables and, (b) source contribution and oxidative potential values (OP-AAv and OP-DTTv) considered as supplementary variables ; (c) Variables factor map for axis1 – axis2 (in black: active variables ; in blue: supplementary variables).

Figure 2: Volume and mass-normalized OP-DTT (respectively OP-DTTv and OP-DTTm) and OP-AA (respectively OP-AAv and OP-AAm) roses and polar plots at Dunkerque

Figure 3: Bivariate polar plots of the PM_{2.5} concentration and the contribution of 11 PM_{2.5} emission sources estimated by the CW-NMF analysis (in $\mu\text{g}\cdot\text{m}^{-3}$) at Dunkerque

Figure 4: CWT on OP-AAv, OP-DTTv, PM_{2.5} and sources contributions estimated by the CW-NMF analysis (in $\mu\text{g}\cdot\text{m}^{-3}$).

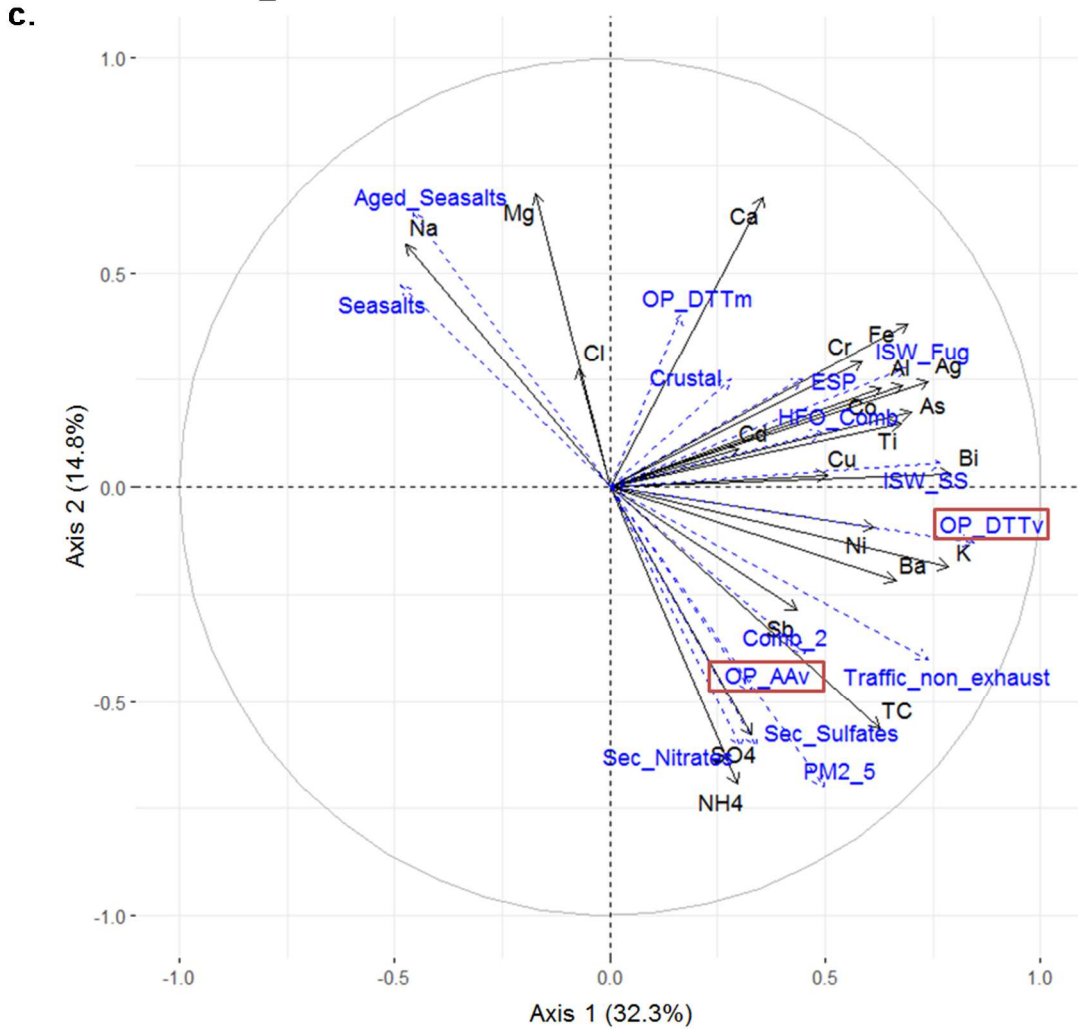
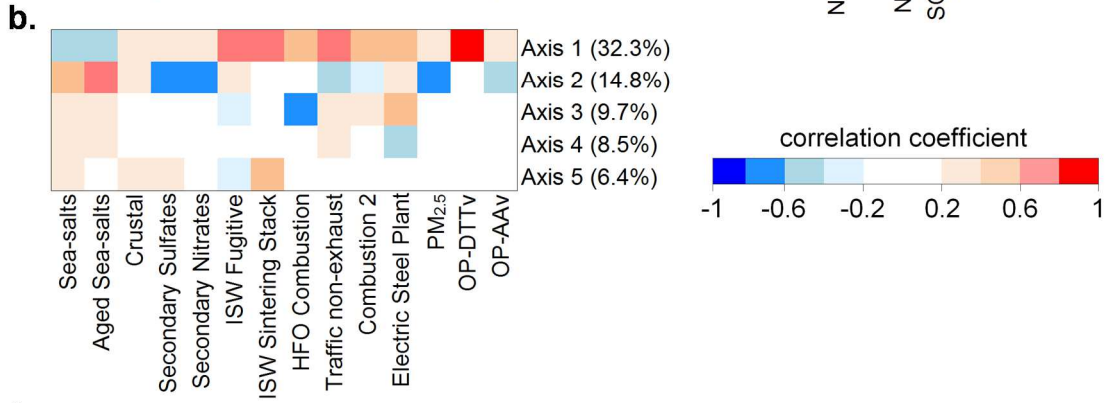
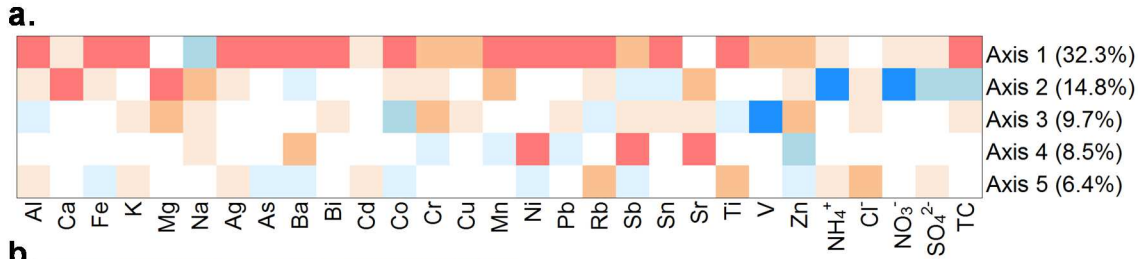
List of tables

Table 1: Average chemical composition (in $\text{ng}\cdot\text{m}^{-3}$) of $\text{PM}_{2.5}$ collected in Dunkerque and $\text{PM}_{2.5}$ concentration (in $\mu\text{g}\cdot\text{m}^{-3}$)

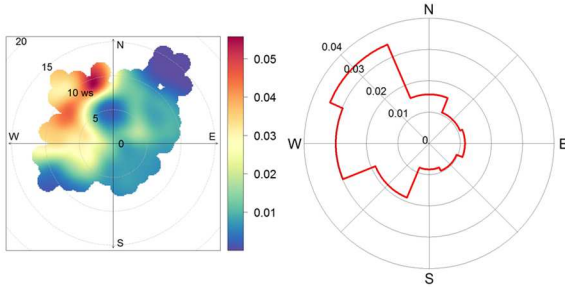
Table 2: Characteristics of the identified sources in Dunkerque (from Kfoury et al. 2016) and average contribution to the composition of the $\text{PM}_{2.5}$ selected samples

Table 3: OP-DDT and OP-AA values found at Dunkerque and comparison with values cited in the literature. (Leaching agent and conditions: G: Gamble solution, H_2O : ultrapure water, M: Methanol, E: Ethanol, PBS: Phosphate buffer (pH = 7.4), Shk: Shaking, US: ultrasonic treatment, No Filt: No filtration)

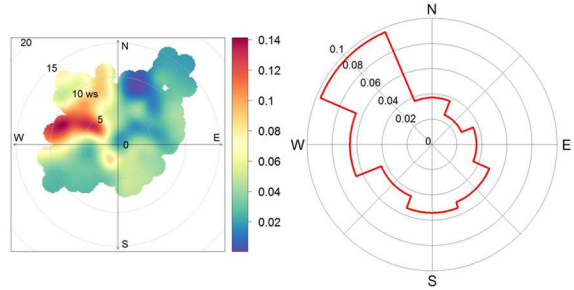
Table 4: Spearman correlation coefficient (r) between Oxidative Potential derived from AA and DTT depletion measurements (OP-AA_v and OP-DDT_v respectively) and $\text{PM}_{2.5}$ constituents, $\text{PM}_{2.5}$ concentrations and sources contributions – only correlation coefficients for which $p < 0.05$ are reported (* $p < 0.001$; bold: $r > 0.6$; italic: $r < 0.4$)



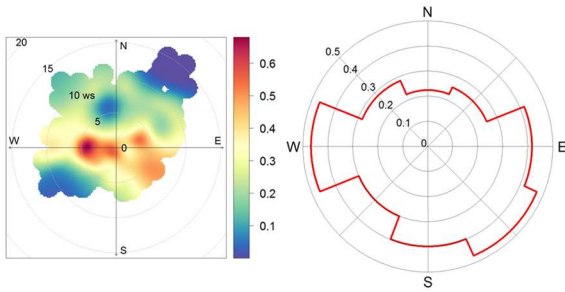
OP-DTTm



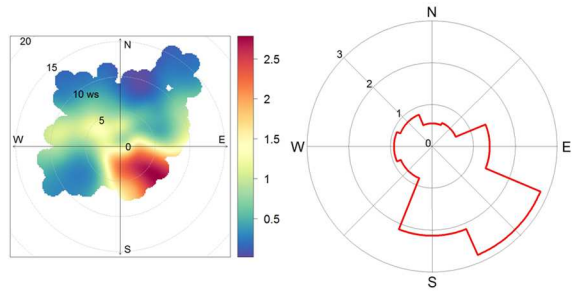
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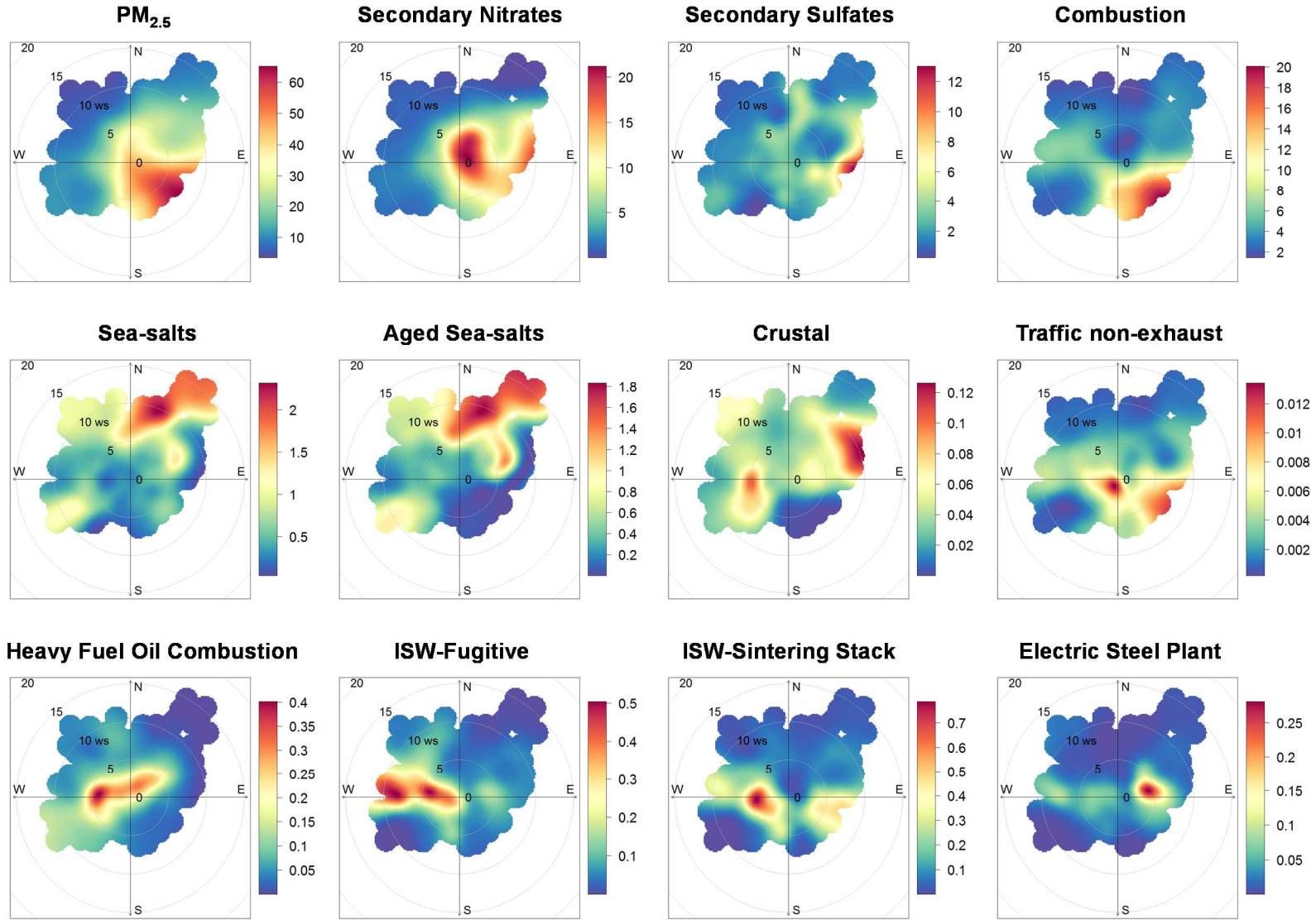


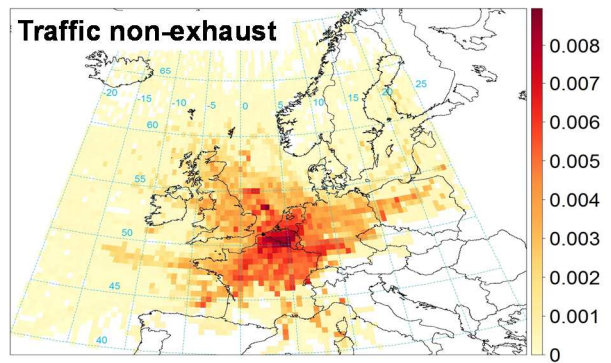
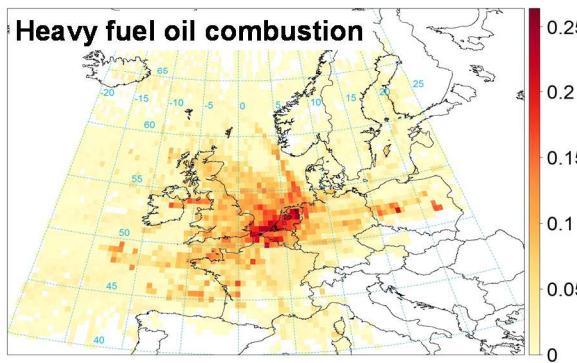
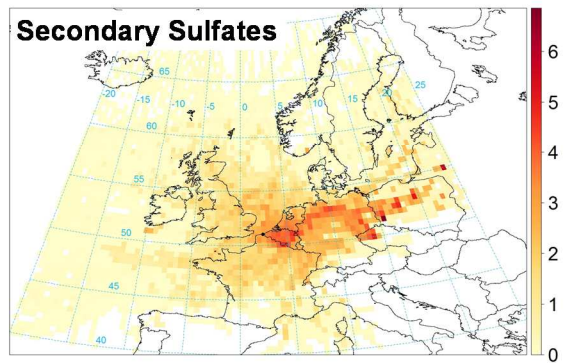
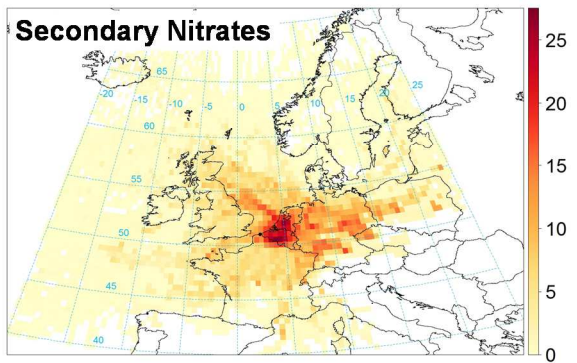
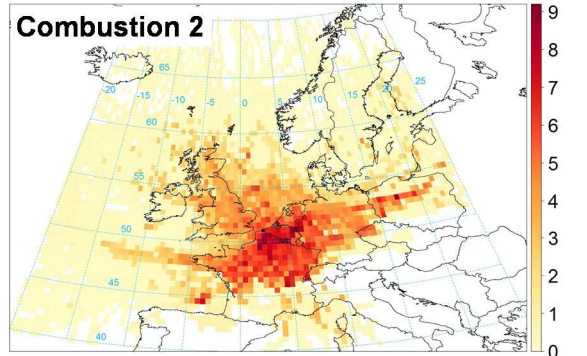
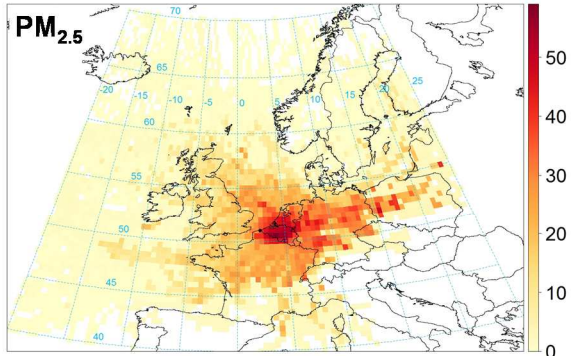
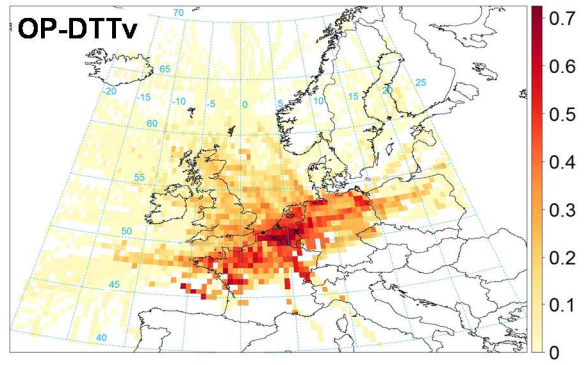
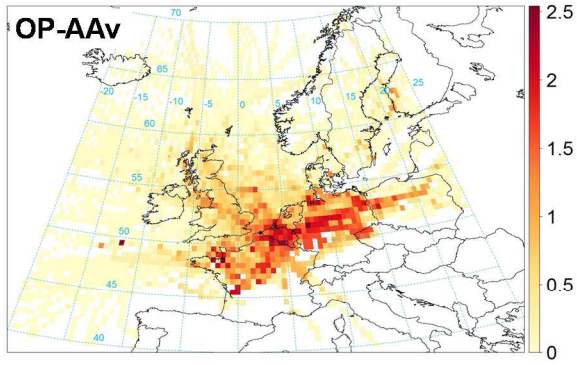
OP-DTTv



OP-AAv







Species	Mean	Median	Min	Max	SD	1st Qu	3rd Qu
Al	26.5	22.5	2.52	112.2	18.7	14.8	30.7
Ca	76.2	62.7	19.4	286	48.2	43.7	88.4
Fe	155	84.0	9.1	1542	231	37.1	195
K	122.7	91.1	29.2	358	83.0	62.0	156
Mg	61.9	39.3	4.35	305	64.1	16.6	72.1
Na	372	225	21.1	2225	419	105	404
Ag	0.07	0.04	0.00	0.42	0.08	0.01	0.09
As	1.09	0.78	0.00	9.45	1.38	0.31	1.45
Ba	0.88	0.74	0.19	5.01	0.67	0.56	1.08
Bi	0.15	0.11	0.00	0.62	0.14	0.04	0.26
Cd	0.46	0.23	0.01	7.98	1.08	0.08	0.44
Co	0.18	0.12	0.02	0.9	0.17	0.07	0.25
Cr	1.36	0.79	0.07	9.88	1.55	0.52	1.61
Cu	3.13	1.63	0.17	31.0	4.46	0.81	4.25
Mn	13.7	4.84	0.12	92.4	20.7	1.59	14.3
Ni	5.36	3.23	0.73	41.55	6.35	1.91	6.29
Pb	14.8	8.26	0.75	118	18.4	3.95	18.8
Rb	0.57	0.43	0.05	4.86	0.73	0.19	0.57
Sb	1.05	0.63	0.00	15.0	2.05	0.37	1.10
Sn	1.03	0.80	0.08	5.43	0.98	0.28	1.33
Sr	0.49	0.40	0.16	1.87	0.32	0.30	0.59
Ti	2.00	1.10	0.00	15.2	2.66	0.29	2.47
V	5.16	3.13	0.49	23.1	5.31	1.84	5.84
Zn	67.6	26.4	1.96	625	113	8.38	88.5
NH ₄ ⁺	2010	1380	172	4817	1301	1043	3128
Cl ⁻	692	576	39.35	1752	409	418	841
NO ₃ ⁻	8678	3619	331	49066	10683	1837	11545
SO ₄ ²⁻	2754	2008	127	9642	2022	1399	3453
TC	5151	3612	805	17547	4246	2008	6881
PM _{2.5} (µg/m ³)	29.2	18.8	3.1	90.8	24.4	11.30	44.30

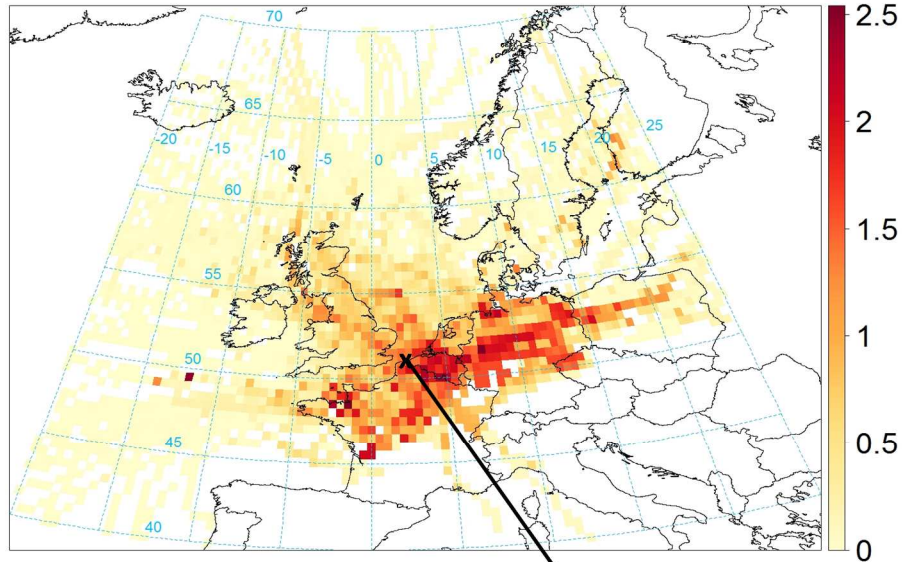
Source	Markers (by order of abundance)	Contribution (ng·m ⁻³)	
		Average ± SD	(min - max)
Sea-salts	Cl ⁻ , Na, Mg	699 ± 605	(0 - 3000)
Aged Sea-salts	SO ₄ ²⁻ , Na, NO ₃ ⁻ , Mg, Cl ⁻	566 ± 595	(0 - 2166)
Crustal	Ca, Al, Fe, Mg, K	56 ± 55	(0 - 257)
Secondary Sulfates	SO ₄ ²⁻ , NH ₄ ⁺	2645 ± 2313	(0 - 10985)
Secondary Nitrates	NO ₃ ⁻ , NH ₄ ⁺	10294 ± 13601	(0 - 60318)
Heavy Fuel Oil Combustion	TC, V, Ni	145 ± 163	(0 - 620)
Combustion 2	TC, NH ₄ ⁺ , NO ₃ ⁻ , K	5444 ± 4312	(0 - 19352)
Traffic non-exhaust	Zn, Fe, Sb, Sn, Cr, Al, Pb, Mn, Cu	4.7 ± 4.3	(0 - 25)
ISW-Fugitive	Fe, Ca, Zn, Al, Mn, Pb, As	150 ± 204	(0 - 1296)
ISW-Sintering Stack	Cl ⁻ , K, NH ₄ ⁺ , Fe, Pb, Rb, Cd	209 ± 271	(0 - 1382)
Electric Steel Plant	Zn, Fe, Pb, Mg, Pb, Cr, Ca, Mn, Cu	77 ± 142	(0 - 746)

Site	Type	Leaching agent	OP-DDTv (nmol·min ⁻¹ ·m ⁻³)	OP-DTTm (pmol·min ⁻¹ ·μg ⁻¹)	OP-AAv (nmol·min ⁻¹ ·m ⁻³)	OP-AAm (pmol·min ⁻¹ ·μg ⁻¹)	PM _{2.5} (μg·m ⁻³)	PM ₁₀ (μg·m ⁻³)	Reference
Dunkerque (FR)	urban industrial	G + Shk (24 h)	0.36	17.7	1.12	55	29.2		This study
Paris (FR)	urban	E + US (1 h)	0.23						(Jedynska et al., 2017)
Atlanta (USA)	urban	H ₂ O + US (30 min)	0.22	24					(Gao et al., 2020)
Hangzhou city (PRC)	urban	H ₂ O + US (30 min)	0.62	6.39			63.1		(Wang et al., 2019)
Beijing (PRC)	urban	H ₂ O + US (30 min)	12.6	130			113.8		(Yu et al., 2019)
Bologna (IT)	urban	PBS + US (15 min)	0.3-1.7		0.2-0.6				(Visentin et al., 2016)
Birmingham (USA)	urban	H ₂ O + US (30 min)			0.75				(Fang et al., 2016)
Grenoble (FR)	urban	G + Shk (2 h)	1.00-4.12		0.17-2.46			7.4-22.3	(Calas et al., 2019)
Nice (FR)	background urban	G + Shk (2 h)	1.81-3.00		0.31-1.91			13.1- 21.4	(Calas et al., 2019)
Talence (FR)	background urban	G + Shk (2 h)	1.08-2.97		0.21-1.59			8.7-26.2	(Calas et al., 2019)
Passy (FR)	background urban	G + Shk (2 h)	1.91-10.0		0.30-7.20			8.6-59.2	(Calas et al., 2019)
Sarno (IT)	background urban	H ₂ O + US (30 min)	0.19	11.67			16.3		(Cesari et al., 2019)
Lecce (IT)	background urban	H ₂ O + US (30 min)	0.29	18.4			16.8		(Chirizzi et al., 2017)
Trento (IT)	background urban	PBS + US (15 min)	0.58	50	0.7	50		19	(Pietrogrande et al., 2018)
Site in Netherlands	background urban	PBS + No Filt	1.4				17.3		(Janssen et al., 2014)
Athens (GR)	background urban	H ₂ O + US (30 min)	0.33	27.98					(Paraskevopoulou et al., 2019)
Salento (IT)	suburban	PBS + US (15 min)	0.19-0.29	10-11	0.09-0.21	5-8	20-26		(Perrone et al., 2019)
Patalia (IN)	semi-urban	M+H ₂ O + US (30 min)	3.8	27			150		(Patel and Rastogi, 2018)
Atlanta (USA)	side street	H ₂ O + US (30 min)	0.78	50			15.6		(Vreeland et al., 2017)
Belgrade (BU)	tunnel traffic	H ₂ O + Shk (90 min)	0.37	9.43					(Jovanovic et al., 2019)
Atlanta (USA)	highway	H ₂ O + US (30 min)	1.08	70			18.3		(Vreeland et al., 2017)
Ala (IT)	industrial (Zn)	PBS + US (15 min)	0.61	120	1.4	60		20.3	(Pietrogrande et al., 2018)

	OP-DTTv	OP-AAv		OP-DTTv	OP-AAv
<i>Species</i>			<i>Species</i>		
Al	0.63*		V	0.47*	
Ca			Zn	0.74*	0.49*
Fe	0.77*	0.51*	NH ₄ ⁺	0.50*	0.31
K	0.67*	0.55*	Cl ⁻		
Mg	-0.31	-0.54*	NO ₃ ⁻	0.68*	0.39
Na	-0.53*	-0.62*	SO ₄ ²⁻	0.56*	0.28
Ag	0.73*	0.52*	TC	0.72*	0.68*
As	0.75*	0.61*			
Ba	0.79*	0.69*	<i>Source contribution</i>		
Bi	0.79*	0.39	Sea-salts	-0.40	-0.39
Cd	0.82*	0.47*	Aged Sea-salts	-0.55*	-0.67*
Co	0.66*		Crustal		
Cr	0.70*	0.56*	Secondary Sulfates	0.58*	0.38
Cu	0.85*	0.64*	Secondary Nitrates	0.67*	0.34
Mn	0.72*	0.33	HFO Combustion	0.45*	
Ni	0.63*	0.28	Combustion 2	0.37	0.54*
Pb	0.79*	0.63*	Traffic non-exhaust	0.80*	0.66*
Rb	0.75*	0.54*	ISW-Fugitive	0.67*	0.46*
Sb	0.70*	0.60*	ISW-Sintering Stack	0.60*	0.53*
Sn	0.82*	0.61*	Electric Steel Plant	0.64*	0.43*
Sr		-0.45*			
Ti	0.67*	0.32	<i>Oxidative potential</i>		
			OP-DTTv		0.46*
PM _{2.5}	0.67*	0.45*	OP-AAv	0.46*	

PM_{2.5} Oxidative Potential

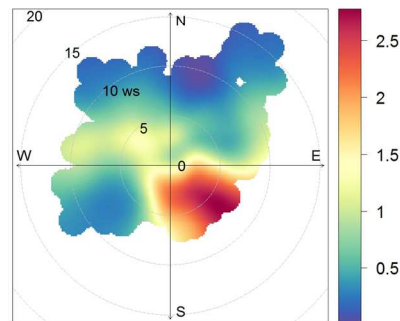
Ascorbid Acid assay



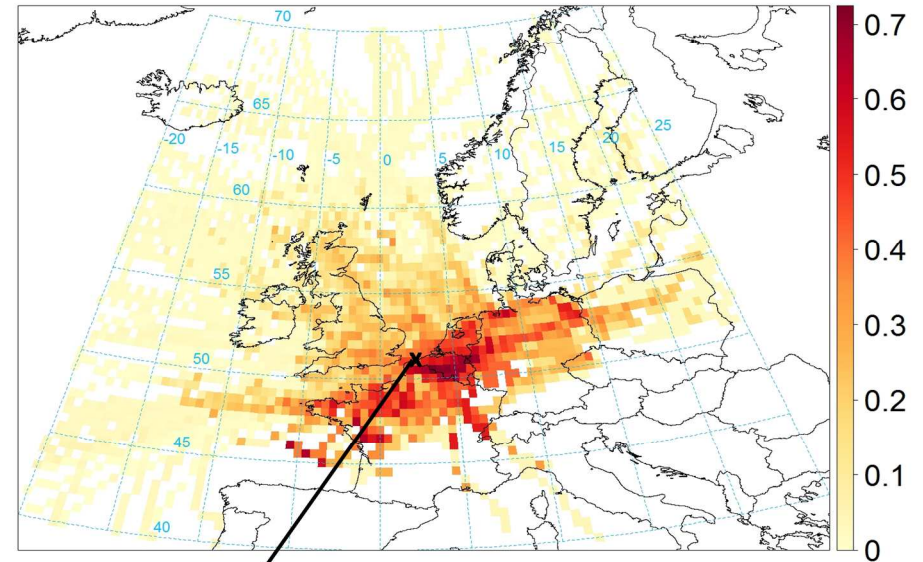
Related to:

Local and distant sources:
- Road traffic and biomass combustion

PM_{2.5} concentration levels



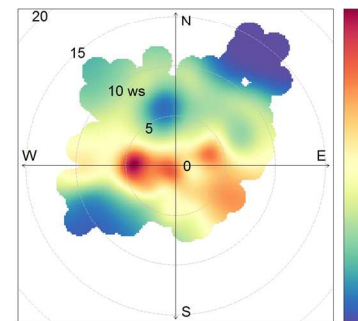
Dithiothreitol assay



Related to:

Local industrial sources :
- Integrated steelworks
- Electric Steel Plant

Local and distant sources:
- Heavy fuel oil combustion
- Traffic non-exhaust
- Combustion possibly



No contribution of sea-salts, aged sea-salts, crustal sources, secondary sulfates and secondary nitrates to OP values