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1 **COUPLING PHREEQC WITH ELECTRO-DIFFUSION TESTS FOR AN**
2 **ACCURATE DETERMINATION OF THE DIFFUSION PROPERTIES ON**
3 **CEMENTITIOUS MATERIALS**

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10
11
12 **Abstract**

13 An improved numerical method as regard with the widely used analytical approximation is
14 presented for determining the transport properties from electro-diffusion tests. Based on the
15 geochemical software PhreeqC, this study aims to better understand and quantify the part of
16 each phenomenon in the diffusion process: (i) the geometrical material properties (as porosity
17 and tortuosity); (ii) the chemical reactions between chloride ions and the cement solid phase;
18 (iii) electro-capillarity effects (such as electrical double layer effects). Tests have been
19 performed on CEM I hydrated cement pastes of different W/C ratio (0.35; 0.42 and 0.50) with
20 different chloride concentrations (500 and 50 mol/m³). The results revealed a good accuracy on
21 the chloride diffusion coefficients D_{E,Cl^-} given by the numerical model while having a weak set
22 of main adjustable parameters. The large amount of experimental data collected during the tests
23 increases the robustness of the model.

24
25 **Keywords:** Electro-Diffusion tests, Chloride, Cement materials, PhreeqC, Numerical model

29 1. Introduction

30 Many studies on chloride diffusion in cementitious materials are available in the civil
31 engineering literature. The aim of these investigations is to quantify, in terms of durability, the
32 degradation of reinforced concrete structures caused by chloride ions (marine or de-icing salt)
33 [1–6] or radionuclides migration [7–9]. Most of them are experimental works and use a flux
34 measure approach to determine the chloride diffusion coefficient [10–17]. To simplify, their
35 models almost consider the same ratio between the effective diffusion coefficient $D_{E,i}$ (m²/s)
36 for a ionic species i and its diffusion coefficient in free water $D_{W,i}$ (m²/s) [18,19]:

$$D_{E,i} = \emptyset G D_{W,i}. \quad (1)$$

37 Where $D_{E,i}$ depends on the diffusion coefficient in free water $D_{W,i}$, on the material porosity \emptyset ,
38 and another macroscopic geometrical parameter G . This last parameter represents the effect of
39 the tortuosity, of the constrictivity, etc. and remains difficult to measure at the sample's scale.

40 Because of the huge complexity of the microstructure and chemistry of cementitious materials,
41 most of the diffusive transport models in concrete are simplified. They only consider the
42 diffusion of one ionic species the same time (as chloride ions) [20–22] and simplified chemical
43 interactions as isotherms [23–27]. These isotherms help to simulate the chloride adsorption by
44 the cementitious phases but do not consider other chemical reactions as calcium absorption by
45 C-S-H gels (non-negligible for pH \geq 10.5) [28] or sodium retention by C-S-H [29]. As well,
46 chloride ions are bound by the cementitious material [24,30] but their adsorption is reduced by
47 the sulfate concentration [31]. Some researches also observed that the quantity of sulfate
48 absorbed by cementitious phases is more important than chloride one [32]. At last, the presence
49 of chloride ions in the pore solution causes the precipitation of phases such as chloroaluminates
50 [33].

51 More elaborated models hence consider the chloride interaction with the cement matrix
52 [20,34,35] and/or electrocapillary effects like the electrical double layer (EDL) [5,6,36,37].
53 Some authors developed rigorous macroscopic transport models with homogenized methods to
54 scale up the Nernst-Planck-Poisson-Boltzmann equation [38–43] from the microscopic scale
55 (i.e. pore scale) to the macroscopic scale of the material sample. The averaging homogenization
56 [44] is based on some simplifying assumptions and do not consider all the complexity of the
57 microstructure (e.g. Samson et al. model [45]). The periodic homogenization [46,47] is
58 mathematically more consistent and does not rely on supplementary simplifying assumptions
59 (e.g. for clay minerals [48] and newly for concrete [49–53]). However, this approach remains
60 formal and no convergence results have been obtained. More recently, homogenization by
61 double-scale methods enables to obtain the homogenized model and the convergence at the
62 same time [54–57]. Some results on properties of Nernst-Planck-Poisson and Nernst-Planck-
63 Poisson-Boltzmann system may also be found in [42,43]. Inspiring of these last methods, the
64 diffusion coefficients $D_{E,i}$ and $D_{W,i}$ can be related to a formation factor F_i [58] (instead of the
65 geometrical parameter G in eq. (1)) :

$$D_{E,i} = F_i D_{W,i}. \quad (2)$$

66 Where F_i , also named scale-up factor, is in general interpreted as the path available for the
67 diffusion of the ionic species i . Based on these analogies, a definition of F_i will be established
68 for the present work as follows:

$$F_i = \emptyset G g_i \quad (3)$$

69 F_i includes the porosity \emptyset , geometrical microscopic effects G (such as tortuosity and
70 constrictivity), and other microscopic effects g_i (such as the EDL).

71 However, all the mathematical models do not consider the ionic species activities and chemical
72 reactions, such as the homogenous reactions that occur in the solution (complexation, inter-ions
73 interaction) and the heterogeneous reactions with the material surface (sorption, solid phase

74 dissolution/precipitation) [59]. By diffusing into concrete, the ions modify the pore solution
75 chemical composition and disturb the local thermodynamic balance of the system. A series of
76 dissolution/precipitation reactions hence take place to restore the equilibrium state.

77 Other more elaborated multi-species transport models consider the diffusion of different species
78 coupled with chemical reactions [60,61] but do not allow the applying of an electrical field at
79 the material bounds to simulate accelerated diffusion test. Recently, the multi-species reactive
80 transport model PhreeqC was updated to simulate the application of an electrical potential all
81 over a chemical system [62]. This change has thus motivated this work which consists of
82 simulating electro-diffusion tests considering both multi-ionic diffusive transport and chemical
83 reactions.

84 Electro-diffusion tests were carried out on hydrated cement paste (HCP) samples with two
85 different chloride concentrations in upstream: 500 mol/m³ like in civil engineering applications,
86 simulating an average chloride concentration in seawater and 50 mol/m³ to tend the radionuclide
87 diffusive behaviour (see [63]). Section 2 introduces the experimental procedure of the material
88 characterisation and electro-diffusion and leaching tests as used in this work. The numerical
89 model on PhreeqC and its calibration are then presented in section 3. At last, the main results
90 obtained, in particular, the determination of effective diffusion coefficients are reported in
91 section 4.

92 **2. Experimental**

93 **2.1. Materials and characterization**

94 **Materials.** Experimental tests were carried out on CEM I hydrated cement pastes (CEM I
95 52.5N CE CPI2 NF called ‘CEM I white’) [NF EN 197-1 and NF P 15-318] with 0.35; 0.42 and
96 0.50 Water/Cement (W/C) ratio. Subsequently, these materials are noted CPI035, CPI042 and
97 CPI050 (for Cement Paste I). The specimens were stored in a basic solution with 83 mol/m³

98 **KOH and 25 mol/m³ NaOH** for 6 months at 20 °C and RH 100% to ensure the complete
 99 saturation of the material porosity. The saturation of samples was verified by immersion under
 100 vacuum in a desiccator for 24 hours. The first centimetres of each specimen from the surface
 101 was removed to avoid any edge effects and possible bleeding. Then, the specimens were cored
 102 and sawed in samples of 65 mm diameter and 10 mm thick.

103 **Hydrated cement paste characterizations.** Water porosities (Table 1) were measured in
 104 previous works [64]. The chemical composition of the pore solutions was measured
 105 equilibrating HCP powders in ultra-pure water with different Solid/Liquid ratio in $\frac{g_{\text{powder}}}{g_{\text{water}}}$
 106 (Table 1). All the concentrations were measured by ionic chromatography Metrohm 930
 107 Compact IC Flex with around 10% accuracy (depending on the ion studied and the dilution
 108 used for analyses).

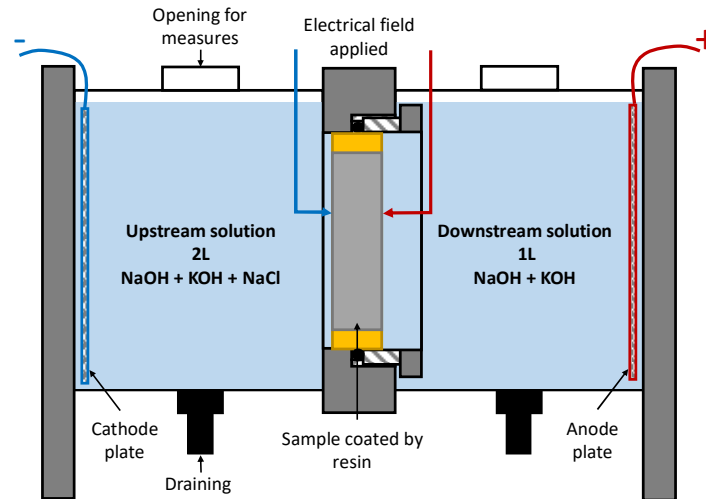
	CPI035	CPI042	CPI050
Water porosity	36%	40%	45%
Cl⁻ (mol/m³)	2.0	1.0	0.8
SO₄²⁻ (mol/m³)	0.21	0.22	0.18
Na⁺ (mol/m³)	387	327	245
K⁺ (mol/m³)	259	180	206

109 Table 1. Water porosity and pore solution composition of the three HCP [64]

110 2.2. Electro-diffusion tests

111 **Electro-diffusion test modified procedure** [64]. A saturated sample taken from the cure was
 112 resin-coated to let the diffusion be one-dimensional. It was placed between two cells (Figure 1):
 113 the upstream cell contains the tracer (chloride ions) in a basic solution composed of potassium
 114 hydroxide KOH and sodium hydroxide NaOH with concentrations given by the pore solution
 115 composition measurements (Table 1). The downstream cell contains only KOH and NaOH.
 116 During the test, the chloride diffusion through the material results from the concentration
 117 gradient between the two cells. To accelerate the process, a constant electrical field was applied

118 to the sample. The electrical field was fixed to respect a weak current value (below 50 mA). The
119 current density passing through the sample and the chloride concentration in downstream were
120 measured throughout the test. The ionic concentrations were then analyzed by ionic
121 chromatography. Temperature and solution pH were also punctually measured during the test.



122

123

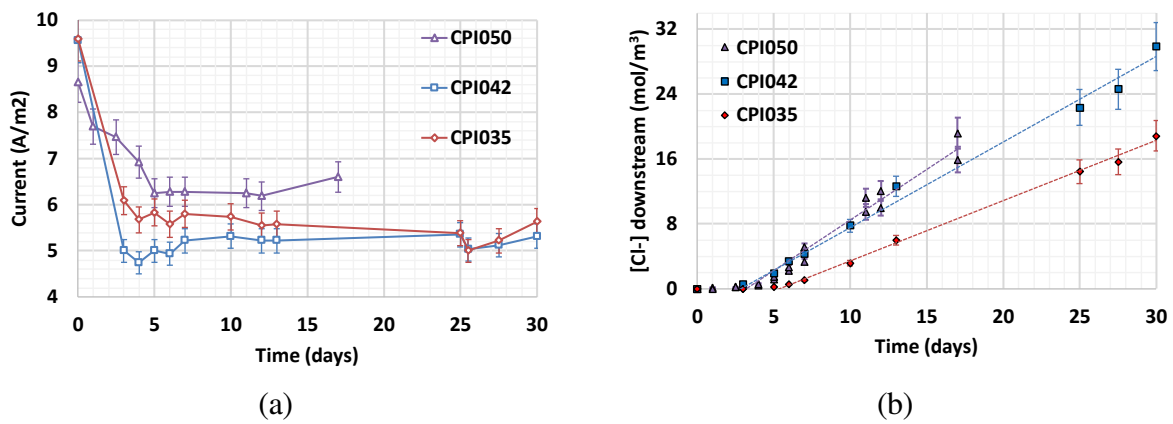
Figure 1. Schematic of electro-diffusion cells built [64]

124 **Electro-diffusion test results.** Tests were performed for the three HCP previously presented
125 CPI035; CPI042 and CPI050 with two different chloride concentrations: 500 mol/m³ and
126 50 mol/m³. The test temperature is around 20-25°C. Other initial conditions (volume solutions,
127 sample lengths and electrical field applied) are summarized in Table 2 and Table 3. The time
128 evolution of the current and chloride concentration downstream are presented in Figure 2 and
129 Figure 3.

Migration tests	Upstream/Downstream cell volume (L)	Sample Diameter x Length (mm)	Electrical field applied (V/m)
CPI050 1	2.17/1.17	63 x 11.3	14
CPI050 2	2.20/1.16	63 x 11.7	14
CPI042 1	2.10/1.21	63 x 11.8	12
CPI042 2	2.11/1.15	63 x 9.60	14
CPI035 1	2.12/1.15	63 x 10.3	21
CPI035 2	2.02/1.11	63 x 11.0	26

130

Table 2. Initial conditions summary for electro-diffusion tests with 500 mol/m³



131

Figure 2. Electro-diffusion tests with 500 mol/m³ chloride in upstream: Time evolution of (a)

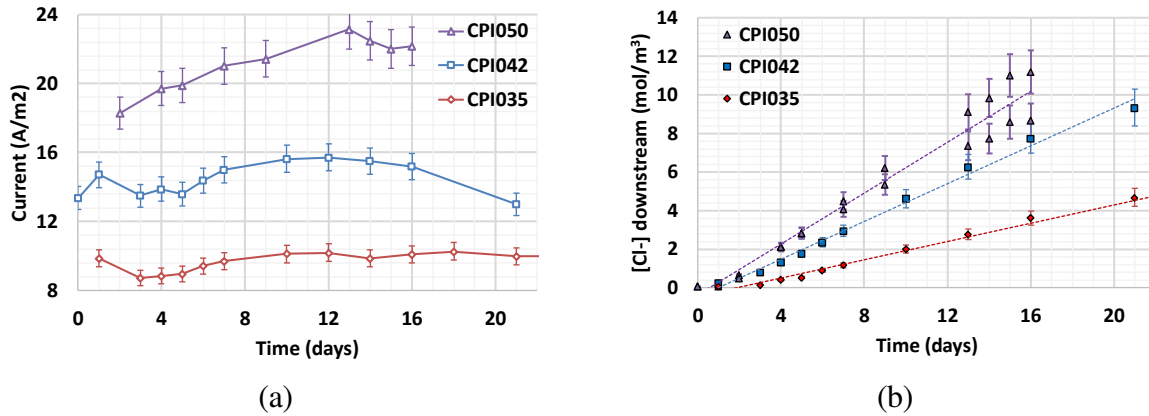
132

the current density and (b) the chloride concentration in downstream.

Migration tests	Upstream/Downstream cell volume (L)	Sample Diameter x Length (mm)	Electrical field applied (V/m)
CPI050 1	2.04/1.07	63 x 9.4	36
CPI050 2	1.89/1.02	63 x 9.5	37
CPI042 1	1.95/1.05	63 x 12.0	27
CPI042 2	1.99/1.07	63 x 12.0	27
CPI035 1	1.95/1.05	63 x 12.4	26
CPI035 2	1.85/0.90	63 x 12.8	27

133

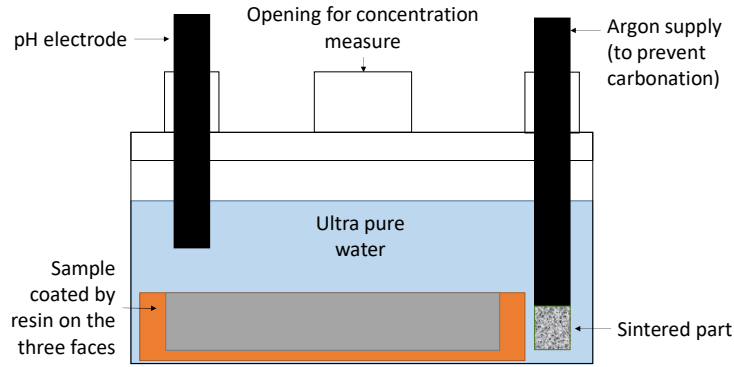
Table 3 – Initial conditions summary for electro-diffusion tests with 50 mol/m³



134 Figure 3. Electro-diffusion tests with 50 mol/m^3 chloride in upstream: Time evolution of (a) the
 135 current density and (b) the chloride concentration in downstream.

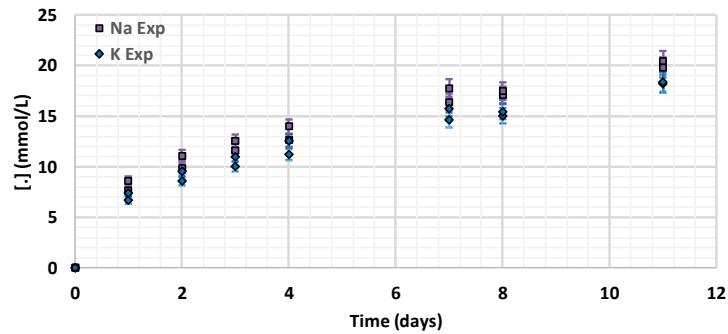
136 **2.3. Leaching tests**

137 **Semi-dynamic leaching test modified procedure.** The leaching tests are liquid-solid
 138 extraction of a solid sample by an aqueous solution. Tests are based on the NF X31-210 French
 139 standard. They were performed to quantify the diffusion of major ionic species in the pore
 140 solution i.e. the alkaline species Na^+ and K^+ . A HCP sample (same kind of electro-diffusion test
 141 sample) was coated on its circumference and on one face to obtain a one-direction ionic
 142 diffusion and simplify thereafter the numerical modelling (null flow condition at one end and
 143 constant flow at the only exposed surface). The HCP sample (around 33 ml) was put on a closed
 144 glass container and immersed in a volume of ultra-pure water (100 ml). This low volume of
 145 leaching solution allows observing small ionic concentrations with precision. However, it
 146 requires replacing each volume sampled (2.5 ml) by a new volume of ultra-pure water to
 147 conserve the sample immersed during all the tests (that qualifies the test as ‘semi-dynamic
 148 leaching’). The leachate and the air above were constantly saturated with Argon using a sintered
 149 bubbler to prevent the leachate carbonation (Figure 4). The leaching container has different
 150 orifices to regularly measure pH and sample the solution for alkali concentration analyses. The
 151 test usually lasts some weeks to one month.



152
153 Figure 4. Schematic view of the modified leaching test on a hydrated cement paste sample

154 **Leaching test results.** An example of a leaching test with a CPI050 hydrated cement paste
 155 sample is presented in Figure 5. The material was put in 100 ml of ultra-pure, the temperature
 156 was around 19.5°C and pH stabilized very quickly around 12.5. The evolutions of Na^+ and K^+
 157 concentrations in the leachate solution can be observed during the 11-day tests. The two
 158 alkaline species (Na^+ and K^+) have similar diffusion and concentration values.



159
160 Figure 5. Leaching test on CPI050: alkali concentration evolution with time in the leachate
 161 solution

162 3. Numerical part

163 3.1. Numerical model and assumptions

164 Phreeqc is a freeware developed in C/C++ by D. Parkhurst and C.A.J. Appelo and used in
 165 many common industrial and academic applications [65,66]. This software allows simulating
 166 classical 1D macroscopic transport equations coupled to reactive processes with a rich chemical

167 database (e.g. [67]). PhreeqC simulate, with a chemical reaction module, the dissolution and
168 precipitation of material solid phases, the oxidation-reduction reactions at the electrodes and the
169 balance between an aqueous phase and the atmosphere surrounding (pure phases, solid
170 solutions, exchange surfaces...). PhreeqC requires a consistent thermodynamic database
171 including in our case cementitious minerals and radionuclides. The numerical model is solved
172 using an explicit finite element scheme which is faster than an implicit one but can be unstable
173 if the initial conditions are not physically compatible or if the space step is too large (not
174 respecting Von Neumann criterion [68]).

175 Code improvements were recently made to consider the application of an electrical field applied
176 at the material boundaries to simulate an electro-diffusion test (version 3.3.12 Ampere). The
177 modifications were validated on electro-diffusion tests and the numerical solution was
178 compared with the Krabbenhoft analytic approach [62].

179 In this study, we consider that the ionic diffusion takes place under isothermal conditions; the
180 HCP are constantly saturated, and the samples have a sufficiently large thickness (1 cm) to
181 consider their microstructured equivalent to a representative one. Moreover, the diffusion is
182 assumed to be unidimensional (with sample coating).

183 **3.2. Numerical resolution**

184 The sample is spatially discretized in n cells where both chemical reactions and 1D diffusive
185 transport take place. The diffusive transport at the macroscopic scale follows the macroscopic
186 Nernst-Planck equation [40,41]:

$$\frac{\partial c_i}{\partial t} = -\frac{\partial J_i}{\partial x} - \frac{\partial q_i}{\partial t} \quad (4)$$

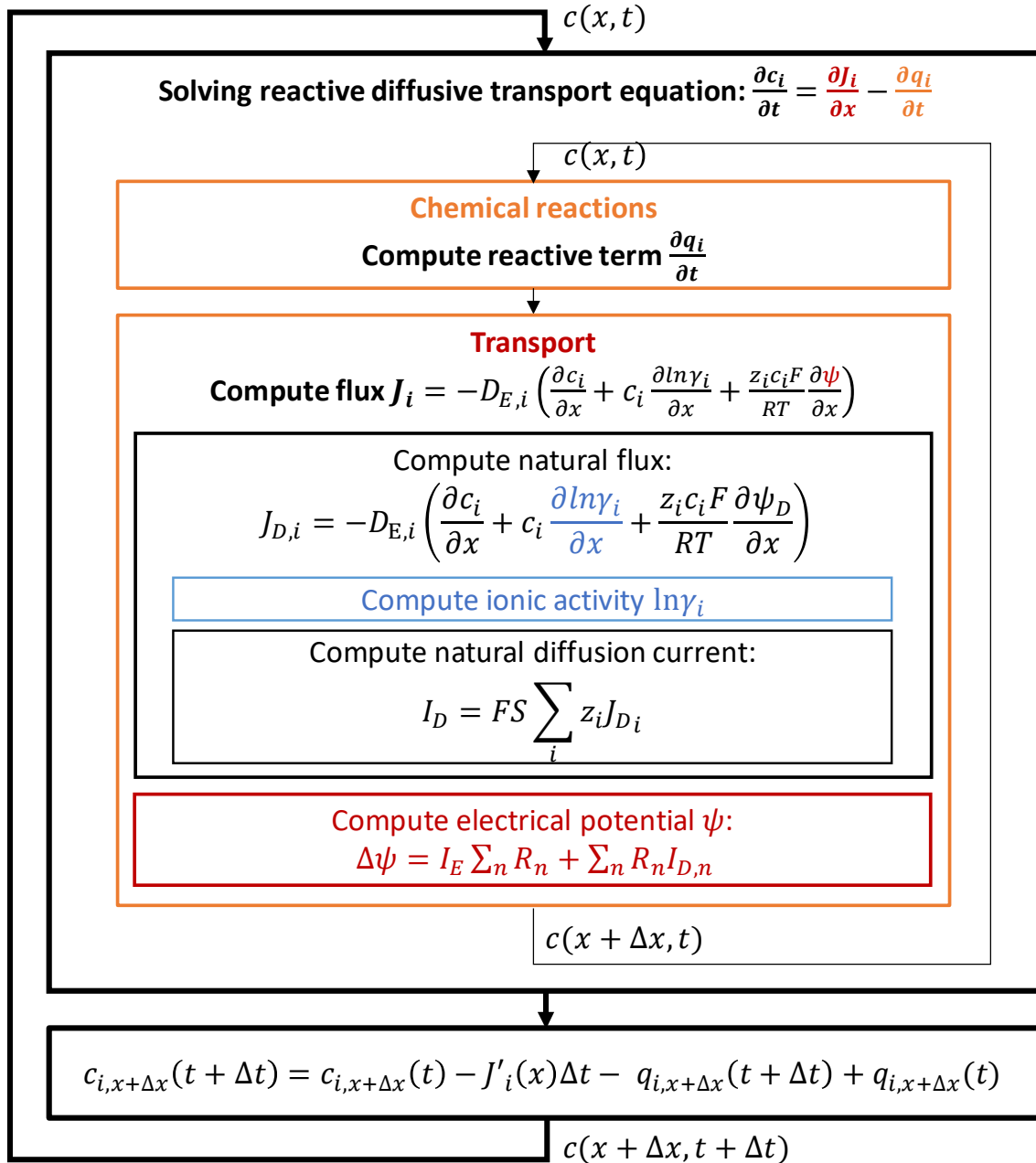
187 where J_i represents the ionic flux (mol/m²/s) defined by:

$$J_i = -D_{E,i} \left(\frac{\partial c_i}{\partial x} + c_i \frac{\partial \ln \gamma_i}{\partial x} + \frac{z_i c_i F}{RT} \frac{\partial \psi}{\partial x} \right). \quad (5)$$

188 Where γ_i denotes the ionic activity of ion i (-), ψ the induced potential (V), z_i the valence of
189 ion i (-), F the Faraday constant (C/mol), R the perfect gas constant (J/K/mol), T the
190 temperature (K). The chemical fixation q_i (mol/m³) is defined by the concentration $c_{s,i}$ (mol/kg)
191 of an ionic species bound to the surface of the material modelled [69]:

$$q_i = \frac{c_{s,i} \rho_b}{\phi_i} \quad (6)$$

192 where ρ_b is the density of the free area (kg/m³) and ϕ_i the accessible porosity (-). In the
193 numerical model proposed, the concentration $c_i(t + \Delta t)$ of a species i at time step $(t + \Delta t)$
194 depends on $c_i(t)$ respecting an explicit time scheme (Figure 6).



195

196

Figure 6 – Algorithm for solving electro-diffusion transport with chemical reactions

197

Chemical reactions. The computation of the reactive term $\frac{\partial q_i}{\partial t}$ is performed with a decoupled

198

PhreeqC chemical reaction module [70] which is not detailed in this paper. Chemical reactions

199

and their associated kinetics are solved with Range-Kutta method [71]. The chemical reaction

200

time step is automatically adapted to ensure the convergence of the computation algorithm of

201

$\frac{\partial q_i}{\partial t}$. In general, it is much more smaller than the time step used for solving the transport

202 equation, due to the different time scales involved. The iterative Newton-Raphson method [72]
 203 allows finding an optimal chemical state (it may also use an optimization method if chemical
 204 reactions are not linearly independent [73]).

205 **Natural diffusion equation.** The ionic activity $\ln\gamma_i$ is calculated using Truesdell-Jones
 206 equation for the majority ions [74], Debye-Huckel ones for the minority ions [75] or Davis ones
 207 by default when nothing else is specified [76]. The flux term $\frac{\partial J_i}{\partial x}$ is solved for a natural diffusion
 208 using the open current hypothesis considering that the electrical current due to natural diffusion
 209 is null [48,77]. $\frac{\partial J_i}{\partial x}$ is computed by a centered spatial scheme [78] and the current induced by
 210 natural diffusion in each cell j is hence obtained from:

$$I_{D,j} = FS \left(\sum_i z_i J_{D,i} \right)_j \quad (7)$$

211 where $J_{D,i}$ denotes the current due to natural diffusion only.

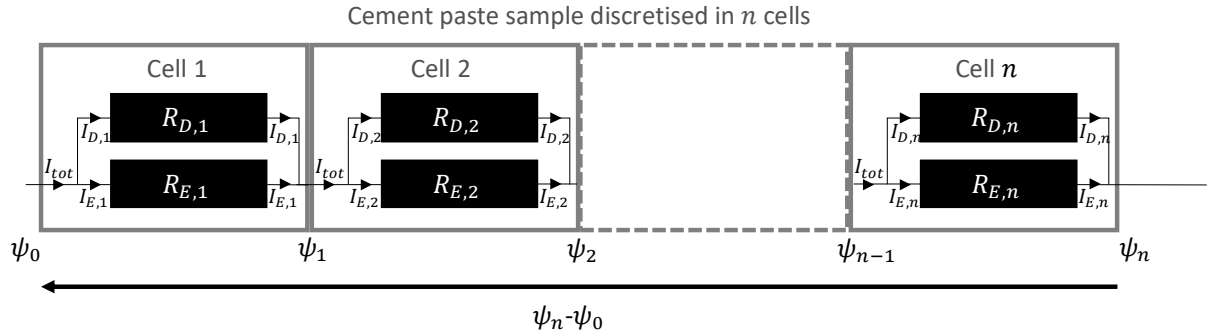
212 **Electro-diffusion equation.** Potential gradient $\nabla\psi_j$ at the bounds of a cell j (length Δx) is
 213 classically computed from:

$$\nabla\psi_j = \frac{(\psi_j - \psi_{j-1})}{\Delta x} \quad (8)$$

214 It is important to notice that ψ includes both the natural potential and the applied electrical
 215 potential ($\psi_n - \psi_0$) across the entire sample (Figure 7). The total current $I_{tot,j}$ flowing through
 216 a cell j hence comes both from the natural diffusion $I_{D,j}$ (calculated previously in (7)) and the
 217 application of an external electrical potential $I_{E,j}$. The total current I_{tot} is assumed to be the
 218 same for all cells [79]. This hypothesis considers that there is no accumulation of charges in the
 219 material (by an electrical analogy, no capacitor in the circuit):

$$I_{tot} = I_{D,1} + I_{E,1} = \dots = I_{D,n} + I_{E,n} \quad (9)$$

220 Each cell can, therefore, be represented as two resistances in parallel (defined by Ohm's law
 221 [80]): $R_{D,j} = (\psi_j - \psi_{j-1})/I_{D,j}$ and $R_{E,j} = (\psi_j - \psi_{j-1})/I_{E,j}$ which depends on the
 222 concentration of species in the pore solution of cell j and on their diffusion coefficients [81,82].



223
 224 Figure 7. Schematic representation of the material sample composed of n cells subjected to an
 225 applied electrical field

226 Assuming that the material is homogeneously saturated, the difference in electrical potential is
 227 evenly distributed in the cells. The potential difference across each cell $(\psi_j - \psi_{j-1})$ is
 228 calculated using $I_{E,j}$. For the first cell [80]:

$$I_{E,1} = \frac{\psi_1 - \psi_0}{R_1} = \frac{(\psi_n - \psi_0)/n}{R_1}. \quad (10)$$

229 According to Kirchhoff's law [83], we can calculate the current I_E in cell 2, 3... and by
 230 recurrence on the potential difference across j :

$$\psi_j - \psi_{j-1} = R_j(I_{tot} - I_{D,j}) \quad (11)$$

231 3.3. Test of numerical models

232 **Effective diffusion coefficient $D_{E,i}$.** In the expression of the ionic flux J_i (equation (5)), the
 233 effective diffusion coefficient $D_{E,i}$ depends on the formation factor $F_i = \phi P_{Current} P_{[i]}$ (see Eqs.
 234 (2) and (3)). Whereas the porosity ϕ is measured by water, the numerical parameters $P_{Current}$
 235 and $P_{[i]}$ have been introduced in PhreeqC to represent the influence of the sample
 236 microstructure geometry G and other microscopic effects g_i on a ionic species i :

- 237 - With the electro-diffusion tests, $P_{Current}$ and $P_{[Cl^-]}$ were both fitted with the evolutions of
238 the current and the chloride concentration in downstream. The current evolution
239 represents the ionic flux sum thus $P_{Current}$ represents the geometrical effects common
240 to all ions in the system. As for $P_{[Cl^-]}$, it takes into account only the specific effects that
241 influence the chloride transport (not other species).
- 242 - With the leaching tests, as $P_{Current}$ is determined by the electro-diffusion test on the same
243 type of material, the alkalis concentration curves are fitted with their specific numerical
244 parameters $P_{[Na^+]}$ and $P_{[K^+]}$.

245 **Chemical reactions.** PhreeqC uses 'phreeqc.dat' database validated and tested [84]. For this
246 work, the main components of cementitious materials were added through the routine **PHASES**:
247 portlandite, gypsum, calcite and CSH gels providing from the Thermochemie database
248 developed by the consortium Andra/Ondraf-Niras/RWM (<https://www.thermochemie-tdb.com/>) [85–87]; ettringite and monosulfoaluminate from CEMDATA07 [88]; Kuzel and
249 Friedel salts dissolution/precipitation [34] from Thermochemie. For each ionic species, the
250 diffusion coefficients in free water $D_{L,i}$ is informed (-dw) and, for this work, contains also the
251 numerical parameters $P_{[i]}$.

253 **Electro-diffusion test initial conditions.** The upstream solution is represented by a solution
254 named **SOLUTION 0** in PhreeqC (Figure 8) where its volume (-water in the **SOLUTION 0**
255 routine) and the average temperature measured during the test and initial concentrations are
256 informed. Then it is possible to compute pH according to the species present (using **charge**
257 routine) and to prevent non-physical compatibility because of uncertainties on pH and
258 concentration measures. The oxidation-reduction reactions at the electrodes are considered with
259 routine **EQUILIBRIUM_PHASES 0**.


```

SOLUTION 0 # Upstream solution
-water 2.12      # Volume (L)
-temperature 22.2 # Temperature en (°C)
-units mmol/kgw
pH 13.40 charge
Cl 50.95        # Initial concentration (mmol/L)
Na 69.79        # Initial concentration (mmol/L)
K 124.81       # Initial concentration (mmol/L)
END

EQUILIBRIUM_PHASES 0 # Electrode
  H2(g) 0 0      # Saturation index and gas molality (mol)
END

```

260

261 Figure 8 – Representation of the upstream cell and electrode in the PhreeqC code

262 The sample material is discretized into several cells (25 for the example of Figure 9). Each cell
263 contains both the initial pore solution composition (SOLUTION 1-25) and the solid phase
264 (EQUILIBRIUM_PHASES 1-25 and SOLID_SOLUTIONS 1-25). The main species in
265 solution were measured (see pore solution characterization in 2.1) while the minor species were
266 computed by PhreeqC equilibrating the solution with the solid phases (portlandite and
267 monosulfoaluminate for calcium and aluminium, respectively). The cementitious phases are
268 considered initially stable (their saturation index equals zero). Each pore sub-solution is
269 considered to be in contact with a sub-solid phase containing the major phases (C-S-H,
270 portlandite...). The range of initial molar content values of the cementitious phases respect
271 those observed for this kind of material. Then it is possible to slightly adjust the
272 monosulfoaluminate and ettringite molalities reacting with chloride ions. Reaction between
273 chloride and monosulfoaluminate AFm to form Friedel's salt or Kuzel's salt [34] can be
274 characterized by a solid solution (SOLID_SOLUTIONS 1-25) as defined below.
275 Dissolution/precipitation phenomenon or ionic exchange can hence be considered.

276

```
SOLUTION 1-25 # The 25 pore sub-solutions
-water 8.09e-4      # Sub-solutions volume (L) :  $V_{\text{Solution 1-25}} = \phi \frac{V_{\text{Sample}}}{25}$ 
-temperature 22.2  # Température (°C)
-units mmol/kgw
pH 13.5 charge
Na 23              # Initial concentration (mmol/L)
Ca 0.03 Portlandite # Initial concentration (mmol/L) adjusted by portlandite molality
K 52
S(6) 0.22
Al 0.18 Monosulfoaluminate
Cl 2.0
END
```

277

```
EQUILIBRIUM_PHASES 1-25 # The 25 sub solid phase of the material
CSH1.6 0 1e-3        # Saturation index and initial molality
CSH1.42 0 0          # Phase not initially present
CSH1.06 0 0
CSH0.7 0 0
Portlandite 0 1e-3
Ettringite 0 4e-5
Monosulfoaluminate 0 15e-5
END
```

278

```
SOLID_SOLUTIONS 1-25 # Solid solutions present in the material
cement
-comp Monosulfoaluminate 0 # Initial molality (mol)
-comp Friedel-salt 0
-comp Kuzel-salt 0
END
```

279 Figure 9 - Representation of the sample (pore solution, solid phase and solid solutions) in the

280 PhreeqC code

281 Similar to the upstream cell, a SOLUTION 26 routine represents the downstream cell. The

282 applied electrical potential applied ($\psi_{25} - \psi_0$), remaining constant during the electro-diffusion

283 test, is indicated by the function -potential (Figure 10).

```

EQUILIBRIUM_PHASES 26 # Electrode
O2(g) 0 0 # Saturation index and gas molality (mol)
END
SOLUTION 26 # Downstream solution
-water 1.1215 # Volume (L)
-temperature 22.2 # Temperature (°C)
-units mmol/kgw
pH 13.40 charge
Na 62.01 # Initial concentration (mmol/L)
K 156.59
-potential 13.65 # Constant electrical potential applied (V) ( $\psi_{25} - \psi_0$ )
END

```

284

285 Figure 10 - Representation of the downstream cell and electrode in the PhreeqC code

286 **Multi-species diffusive transport model.** The multi-species transport takes place in the 25
287 cells whose length (-lengths) is specified in the TRANSPORT routine (Figure 11). The material
288 boundary conditions (-boundary_conditions) consider the concentration evolution at upstream
289 and downstream (flux boundary also known as first type or Dirichlet boundary). The natural
290 transport is only due to diffusion (specified in the line -flow_direction) as the sample is
291 saturated (no advection). PhreeqC simulates a multi-species diffusion with the function -
292 multi_d. For this research need, it contains implicitly the numerical parameter $P_{Current}$ (Figure
293 11 in appendix).

```

TRANSPORT
-cells 25 # Number of cells
-lengths 0.432e-3 # Length of each cell (m)
-boundary_conditions flux flux # Flux conditions at the material bounds
-flow_direction diffusion_only # Only diffusion for transport
-multi_d true 1e-9 0.007 0.0 1 false # Multi-species diffusion with:
#  $D_{W,i}$  by default /  $P_{Current}$  / Not Used
-time_step 1 hour 5 # Time step: 1 hour divided in 5 sub time steps
-shifts 24 # Number of time step before the first sampling
END

```

294

295 Figure 11 – Multi-species diffusion transport in the PhreeqC code

296 The time step and the number of time steps (-**shifts**) are also given in routine **TRANSPORT**.

297 The simulation is divided into several sub-simulations corresponding to the electro-diffusion
298 test duration between each solution sampling. When the upstream and downstream cells are
299 sampled for concentration measurements, their volume decrease during the test: that is
300 considered in PhreeqC.

301 **Leaching test model specificity.** The leaching test model is similar to electro-diffusion one
302 except for the following points: The 100 ml leaching solution is represented in **SOLUTION 0**
303 routine by an ultra-pure water solution; the regular replacement of the solution sampled with
304 ultra-pure water is considered; a closed boundary condition represents the resin-coated surface;
305 a low electrical current (-**fix_current** in routine **TRANSPORT**) initiates the natural diffusion (to
306 have a non-null potential at first cells).

307

308 4. Coupling PhreeqC and experimental tests

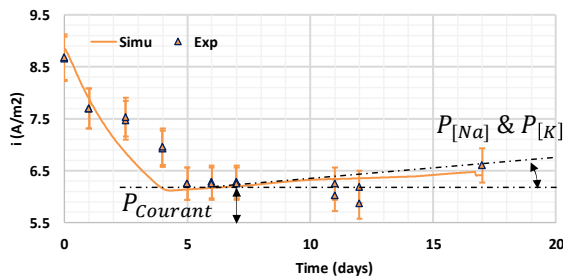
309 4.1. Calibration on HCP model CPI050 with 500 mol/m³ tests

310 **Step (a).** First, the global parameter $P_{Current}$ is adjusted using the current evolution with time.
311 During the transient regime, the current decreased due to the leaching of alkaline species from
312 the pore solution of the sample to the upstream cell. Then, the current became stable and
313 reached a value that can be adjusted from experiments of electro-diffusion tests with the
314 parameter $P_{Current}$ (Figure 12a₁). On the steady-state, the current tends to increase because
315 alkaline species in the downstream solution migrate through the material to upstream under the
316 effect of the electrical field applied. The slope of this curve is then adjusted by the numerical
317 parameters $P_{[Na^+]}$ and $P_{[K^+]}$. Moreover, as the concentration measurements during the electro-
318 diffusion tests do not allow obtaining a precise value for $P_{[Na^+]}$ and $P_{[K^+]}$ (because of the high
319 **Na⁺ and K⁺** concentrations used), these two parameters were determined with the PhreeqC
320 simulation of leaching tests on the same type of HCP (Figure 12a₂). As the leaching simulation
321 consider the replacement of the solution samples, it involves the curve discontinuities as
322 observed in Figure 12a₂.

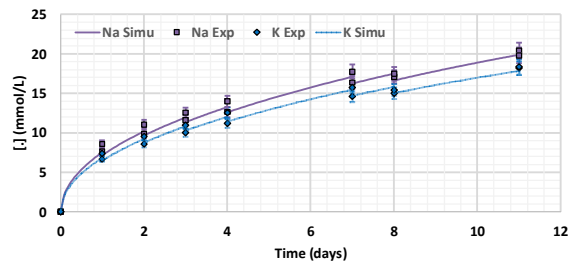
323 **Step (b).** The sulfate leaching from the HCP in the downstream cell during the electro-diffusion
324 test (as sulfate are not initially present upstream and downstream) is fitted to provide additional
325 information on sulfate-containing phases. Sulfate leaching is only influenced by the global
326 parameter $P_{Current}$ and the initial molar content of monosulfoaluminate (AFm) and ettringite
327 (AFt). The sulfate curve highlights two different slopes controlled by the AFm phase, and then
328 by the AFt phase (Figure 12b). As $P_{Current}$ has been adjusted previously with current, the initial
329 **content of AFm and AFt permit fitting the two slopes.** The ettringite saturation index was also
330 adjusted to force its precipitation that seems to be accelerated by the electrical field applied or
331 the high concentration of chloride used in the electro-diffusion test. **The consistency of the AFm**

332 content used in the code can be verified with the timelap in the chloride curve : the delay at the
 333 test beginning is mainly due to the adsorption of chloride ions by monosulfoaluminate (Figure
 334 12c).

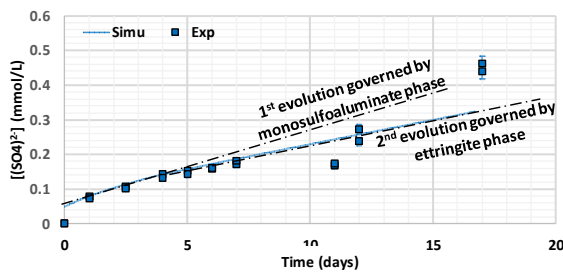
335 **Step (c).** Finally, the chloride parameter $P_{[Cl^-]}$ was adjusted to reproduce the steady state of
 336 chloride experimental curve (Figure 12c) and consider other effects (in particular charge effects
 337 or EDL) affecting the chloride diffusion that are not included in the global parameter $P_{Current}$.
 338 $P_{[Cl^-]}$ fitting does not influence significantly the current and sulfate behaviour. The correct fit of
 339 the 3 curves is checked one last time. Then, we deduce by different fitting, that the chloride
 340 diffusion coefficient can vary about $D_{E,Cl^-} = \emptyset P_{current} P_{[Cl^-]} D_{W,Cl} = (7.9-8.9) \times 10^{-12} \text{ m}^2/\text{s}$
 341 while respecting the current, sulfate and chloride curves.



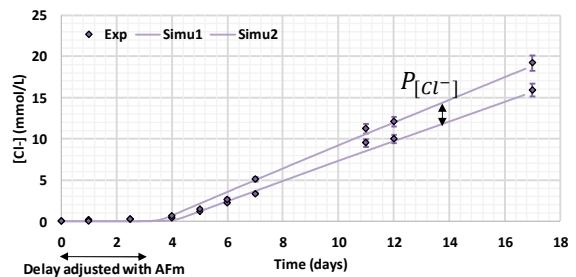
(a1) Current adjustment with $P_{Current}$, $P_{[Na^+]}$ and $P_{[K^+]}$



(a2) Determination of $P_{[Na^+]}$ and $P_{[K^+]}$ with a leaching test simulation on CPI050



(b) Sulfate behavior adjustment with ettringite molar content



(c) Chloride adjustment with AFm content and P_{Cl^-}

342 Figure 12. Calibration of an electro-diffusion test on CPI050 with 500 mol/m^3 of chlorides in
 343 upstream

344 4.2. Numerical model sensitivity

345 The numerical model sensitivity was evaluated on CPI050 with 50 mol/m³ of chloride in the
346 upstream cell (Figure 13) according to the main parameters influencing the chloride diffusion
347 coefficient (porosity, pore solution alkalinity, chloride timelap and the number of measured
348 points). The objective is to identify the code adjustment parameter presenting experimental
349 uncertainties and to quantify the consequences on the diffusion coefficient D_{E,Cl^-} PhreeqC.

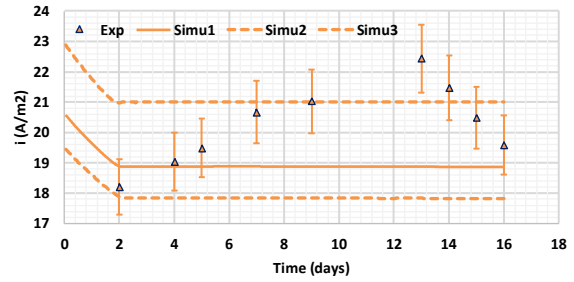
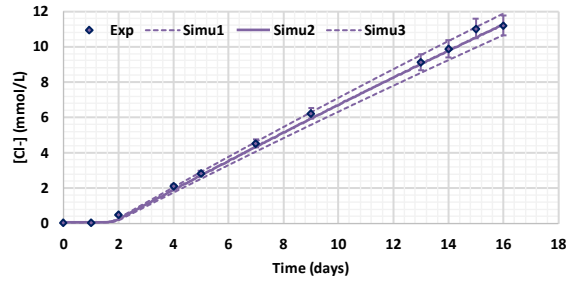
350 (a) The porosity is one of the determining parameters in the diffusion process. It is measured
351 with water at $\pm 1\%$ and considered in the sub-pore solution volumes of the PhreeqC model. If
352 there is an error of $\pm 5\%$ in the porosity informed in PhreeqC, there is no difference in the
353 sulfate concentration curve. However, the chloride and current curves (Figure 13a) vary
354 significantly since the porosity ϕ is included in the diffusion coefficient ($D_{E,i} =$
355 $\phi P_{Current} P_{[i]} D_{W,i}$). The numerical parameter $P_{Current}$ can be adjusted to counterbalance the
356 error of the porosity measurements without having a direct consequence on the value of $D_{E,i}$.
357 Indeed, in the code, the product $\phi P_{Current}$ remains constant. Some sensitivity studies in the
358 literature corroborate these results [89].

359 (b) The pore solution alkalinity is an indicator of the ionic species that have the most influence
360 on the current evolution. Hydroxide concentrations in the pore solution were measured by pH
361 meter at ± 0.2 . These uncertainties can induce important changes in chloride, sulfate and current
362 curves (Figure 13b). It could be possible to readjust the curves modifying $P_{Current}$ but it would
363 induce a non-negligible error of 13% on the diffusion coefficient D_{E,Cl^-} . The consistency of the
364 pore solution pH is rather checked corroborating the leaching test results which are carried out
365 on the same type of material (with the same numerical parameter $P_{Current}$). The leaching tests
366 permit to obtain a good approximation of the initial alkalis concentration in the pore solution
367 and of their diffusion coefficients, based on the evolution of their leaching. Thus, coupling the

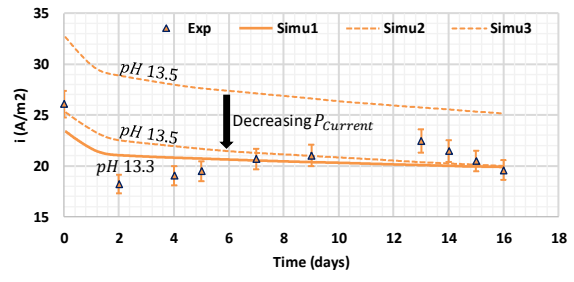
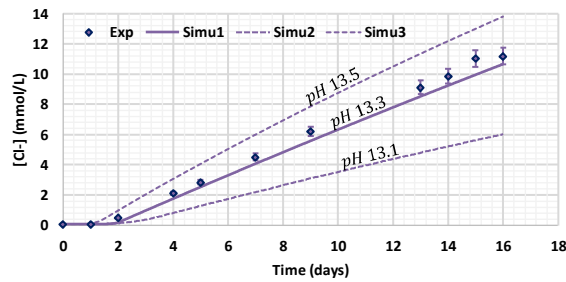
368 two numerical models (electro-diffusion and leaching tests) allows to cross the experimental
369 results and refine the measurements of pH and concentrations in the pore solution and the alkali
370 transport parameters ($P_{Current}$, P_{Na^+} , P_{K^+}).

371 (c) The chloride timelap in downstream depends not only on the geometry of the material but
372 also on chloride reactions with solid phases. Friedel and Kuzel's salts are assumed not to be
373 present initially, the only adjustable parameter that influence chloride timelap in downstream is
374 the initial monosulfoaluminate AFm molar content (Figure 13c). As it was not measured in this
375 work, the AFm molar content can be evaluated at $\pm 5 \text{ mol/m}^3$ based on literature data. However,
376 the numerical simulations allow increasing the accuracy to $\pm 1 \text{ mol/m}^3$ on the AFm molar
377 content by jointly fitting the curves of chloride and sulfate in downstream. The accuracy on the
378 AFm molar content thus mainly depends on the frequency of measurements at the test
379 beginning (during the unsteady state).

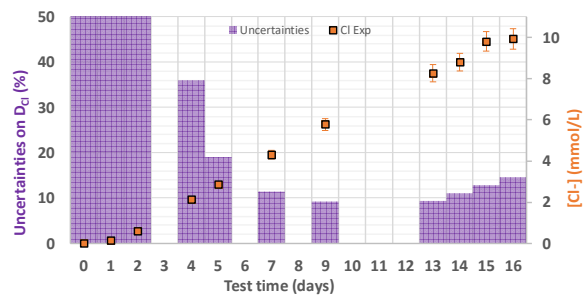
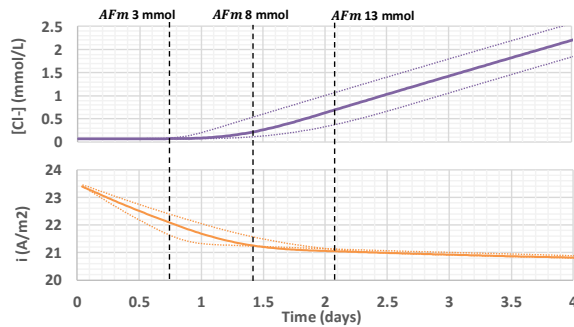
380 (d) For a high chloride concentration in upstream (500 mol/m^3), the steady-state regime was
381 reached very quickly then it was easy to reproduce the linear chloride curve and obtain an
382 accurate diffusion coefficient D_{E,Cl^-} . With a lower concentration such as 50 mol/m^3 , it can be
383 more difficult to fit the chloride curve if there are not enough measure points for chloride
384 concentration. The error on D_{E,Cl^-} according to the number of measuring points fitted in the
385 numerical model was evaluated by adjusting the two main parameters influencing chloride
386 evolution, i.e. AFm molar content for the chloride timelap and $P_{[Cl^-]}$ for curve slopes (Figure
387 13d). It reveals that it is necessary to have at least six non-null measured points on the chloride
388 curve. However, the uncertainties on D_{E,Cl^-} slightly increase after too long a time may be due
389 to the upstream solution carbonation or the material degradation caused by the electrical field
390 applied.



(a) Porosity influence on chloride and current curves. Porosity varying $\pm 5\%$



(b) pH influence on chloride and current curves. pH varying ± 0.2



(c) AFm influence on chloride and current curves (± 5 mmol)

(d) Number of measures influence on chloride curve

391 Figure 13. Numerical model sensitivity: examples on CPI050 electro-diffusion test with
 392 50 mol/m³ chloride in upstream

393 As a conclusion of the sensitivity study, the minimal uncertainties in the chloride diffusion
 394 coefficient are estimated at about 10% (for 7 measured points). Using a simplified method by
 395 measuring the flux in the steady-state regime [64], the sensitivity for the same case (CPI050
 396 with 50 mol/m³) is estimated at approximately 18%.

397 4.3. Adaptation of the CPI050 model to CPI042/CPI035

398 **CPI042/CPI035 models with 500 mol/m³ of chloride.** The previous numerical model CPI050
 399 was adapted for the two hydrated cement pastes CPI042 and CPI035. To this end, the numerical

400 parameters $P_{Current}$ and $P_{[Cl^-]}$ were slightly adjusted to well fit the current and chloride curves,
 401 and the saturation index for ettringite was adjusted to reproduce the sulfate curve. The chloride
 402 diffusion coefficient $D_{E,Cl^- PhreeqC}$ for the three HCP with 500 mol/m³ of chloride in upstream
 403 deduced from the simulations are presented in Table 4.

	CPI050	CPI042	CPI035
$D_{E,Cl^- PhreeqC}$ (10^{-12} m ² /s)	8.4 (±0.5)	6.8 (±0.4)	2.9 (±0.1)
$D_{E,Cl^- Simpl}$ (10^{-12} m ² /s) [64]	8.1 (±0.8)	4.7 (±0.6)	2.7 (±0.3)

404 Table 4 – Effective diffusion coefficient for chloride obtained with PhreeqC and from
 405 expression [64] on CPI050, CPI042 and CPI035 with 500 mol/m³ of chloride in upstream

406 The chloride diffusion coefficients obtained with PhreeqC are in the average of the
 407 experimental values for a CEM I HCP of the same ratio (between 0.35 and 0.5) [90]. We
 408 observe a difference in the chloride diffusive behaviour within the three materials which
 409 nevertheless have a closed microstructure. Indeed, W/C ratios, water and mercury porosities are
 410 not so far apart and the chloride diffusion coefficients obtained are of the same order of
 411 magnitude. Those results are coherent: chloride diffusion is slower when the HCP contains less
 412 water and less porosity to diffuse. These values are also coherent with those calculated by a
 413 simplified method measuring the flux $D_{E,Cl^- Simpl}$ [64] on the same electro-diffusion tests.
 414 Moreover, the values obtained with PhreeqC are two times more precise than the simplified
 415 expression $D_{E,Cl^- Simpl}$ commonly used (uncertainties of 5% and 12% resp.). This can be
 416 explained by more numerous experimental data used for the PhreeqC method by reproducing
 417 the current, sulfate and chloride behaviours.

418 The successive reversion between experiments and simulations allows refining the model and
 419 leads to a more realistic and reliable tool.

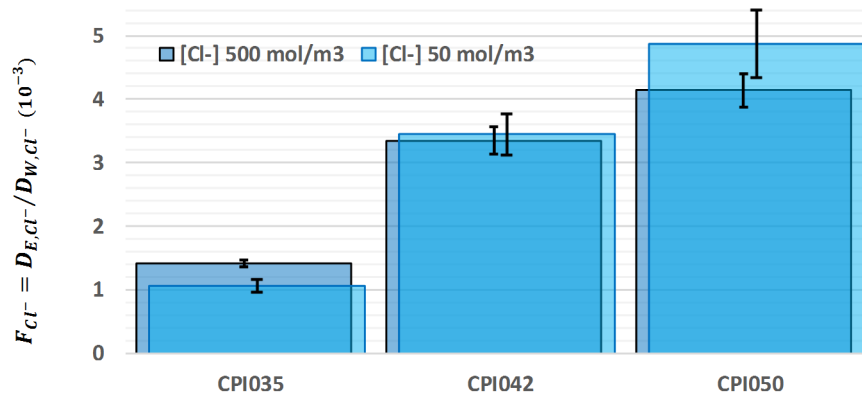
420 **CPI050/CPI042/CPI035 with 50 mol/m³ of chloride.** Using the same parameters $P_{Current}^{CPI050}$,
 421 $P_{Current}^{CPI042}$ and $P_{Current}^{CPI035}$ obtained from electro-diffusion tests with 500 mol/m³ of chlorides,

422 numerical current curves well fit the experiments. Indeed, the materials tested and their
423 geometries were similar. However, the ettringite dissolution in the numerical model was not
424 forced to fit the sulfate curve, as done in the three previous models. The AFt dissolution cannot
425 be provoked by the electrical potential as they were similar in 50 and 500 mol/m³ tests. We can
426 hence deduce that the acceleration of the ettringite dissolution observed in the 500 mol/m³ tests
427 should be caused by the high chloride concentration in upstream inducing a stronger ionic
428 strength than with only 50 mol/m³ of chloride in upstream. Then, for the three 50 mol/m³ tests,
429 the ettringite index saturation in the numerical models are set to zero. As well, AFt and AFm
430 phases have the same molar content as in 500 mol/m³ models; $P_{[Cl^-]}^{CPI050}$ for 50 mol/m³ test is the
431 same as for 500 mol/m³ test; and $P_{[Cl^-]}^{CPI042}$ and $P_{[Cl^-]}^{CPI035}$ are slightly adjusted to fit the chloride
432 curves. Thus, the chloride diffusion coefficients $D_{E,Cl^- PhreeqC}$ deduced from the simulations
433 are similar to those obtained from the 500 mol/m³ tests in view of uncertainties (Table 5). The
434 uncertainties on the chloride diffusion coefficient $D_{E,Cl^- PhreeqC}$ for 50 mol/m³ tests are slightly
435 greater than with 500 mol/m³ tests (uncertainties of 5% and 9% resp.). However, it is still
436 possible to observe differences in behaviour between the three HCP when reducing the chloride
437 concentration in upstream. As previously, the values calculated for the simplified diffusion
438 coefficient $D_{E,Cl^- simpl}$ are not accurate enough to observe any difference between the two
439 hydrated cement pastes CPI050 and CPI042 (uncertainties of around 31%). The numerical
440 model has the advantage to correct some experimental data uncertainties (such as the material
441 boundary conditions or temperature test fluctuation) as it is based on multiple experimental
442 curves and numerous measured points. The PhreeqC model is, therefore, more precise to use
443 with a lower concentration of chlorides.

	CPI050	CPI042	CPI035
$D_{E,Cl^-} \text{ PhreeqC}$ ($10^{-12} \text{ m}^2/\text{s}$)	9.9 (± 1.0)	7.0 (± 0.6)	2.2 (± 0.2)
$D_{E,Cl^-} \text{ Simpl}$ ($10^{-12} \text{ m}^2/\text{s}$) [64]	7.3 (± 1.3)	4.8 (± 1.4)	1.9 (± 0.9)

444 Table 5 – Effective diffusion coefficient of chloride obtained with PhreeqC and simplified
 445 expression on CPI050, CPI042 and CPI035 with 50 mol/m³ chloride in upstream

446 **Formation factors for the 500 and 50 mol/m³ tests.** F_{Cl^-} was calculated using a chloride
 447 diffusion coefficient in free water $D_{W,Cl^-} = 2.032 \times 10^{-9} \text{ m}^2/\text{s}$. For the HCP with the highest liquid
 448 volume (i.e. CPI050 and CPI042), accounting the uncertainties, F_{Cl^-} appeared similar
 449 regardless of the initial chloride concentration in upstream: $F_{Cl^-,500}^{CPI050} = F_{Cl^-,50}^{CPI050}$ and $F_{Cl^-,500}^{CPI042} =$
 450 $F_{Cl^-,50}^{CPI042}$ (Figure 14). However, we can observe a significant difference for CPI035 that seems
 451 not to be due to measurement uncertainties: $F_{Cl^-}^{CPI035}$ was smaller when the chloride ions were
 452 less concentrated upstream ($F_{Cl^-,50}^{CPI042} < F_{Cl^-,500}^{CPI042}$).



453
 454 Figure 14. Formation factors calculated for the three HCP with two different chloride
 455 concentrations in upstream

456 A lower porosity could involve microscopic phenomenon slowing down the chloride diffusion.
 457 This phenomenon could be mainly related to the electrocapillary effects such as the C-S-H
 458 surface charge [36]. Indeed, the influence of the electrical double layer (EDL), directly linked to
 459 the Debye length, is inversely proportional to the ionic concentration of the pore solution and

460 the pore size [91]. That could explain why EDL effects are negligible for the highest chloride
 461 concentration at 500 mol/m³ and for CPI050 and CPI042 which have high water volumes and
 462 porosities. Based on equation (3) and the observation in Figure 14, as the microstructure (\emptyset and
 463 G) is similar for a same type of material regardless the chloride concentration in upstream, the
 464 following results can be deduced:

$$\begin{aligned}
 g_{Cl^-,50}^{CPI050} &\cong g_{Cl^-,500}^{CPI050} \\
 g_{Cl^-,50}^{CPI042} &\cong g_{Cl^-,500}^{CPI042} \\
 g_{Cl^-,50}^{CPI035} &< g_{Cl^-,500}^{CPI035}
 \end{aligned} \tag{12}$$

465 Moreover, as no microscopic effects seems to be present for high chloride concentration and
 466 thus high ionic strength (i.e. $g_{Cl^-,500}^{CPI050} \cong g_{Cl^-,500}^{CPI042} \cong g_{Cl^-,500}^{CPI035} \cong 1$), it is possible to conclude
 467 from (12) that there is also no microscopic effects on CPI042 and CPI050 for lower chloride
 468 concentrations:

$$g_{Cl^-,50}^{CPI050} \cong g_{Cl^-,50}^{CPI042} \cong 1 \tag{13}$$

469 A different conclusion is nevertheless highlighted for the least porous HCP (CPI035) when
 470 chloride ions are less concentrated:

$$g_{Cl^-,50}^{CPI035} < 1. \tag{14}$$

471 That underlines the need to redefine the definition of the formation factor distinguishing
 472 macroscopic and microscopic geometrical effects and including expressly the EDL effects.

473 5. Conclusions

474 The numerical model developed on PhreeqC allowed simulating accurately electro-diffusion
 475 tests on cementitious materials considering both the diffusive transport and chemical reactions.
 476 Leaching test simulations completed the electro-diffusion tests one as they had a common
 477 physical-chemical basis and same numerical parameters for the sample (dimensions,
 478 microstructure complexity, cementitious phases, pore solution composition, etc.).

479 The sensitivity analyse revealed a weak set of main parameters in the numerical model
480 ($P_{Current}$ for global geometrical effects and $P_{[Cl^-]}$ for geometrical effects influencing only
481 chloride diffusion) and a good accuracy of the measurement of D_{E,Cl^-} ($\pm 10\%$). The AFm molar
482 content could also be a numerical parameter if it is not measured. $P_{Current}$ and $P_{[Cl^-]}$ are
483 equivalent to the tortuosity parameter that can be found in most transport models. The large
484 amount of experimental data, available for a given type of sample and used to calibrate the
485 numerical model, increases its robustness. Successive back and forth between experiments and
486 simulations make it possible to refine the modeling of physic-chemistry with PhreeqC.
487 First calibrated on hydrated cement pastes with electro-diffusion tests at 500 mol/m^3 of chloride
488 in upstream (civil engineering conditions), the numerical modelling allowed to accurately
489 compare chloride diffusive behaviour (uncertainties of 5% with the model compared to 9% with
490 a simplified equation). It also allowed separating the influence of physical and chemical effects
491 on the chloride transport.
492 Decreasing the chloride concentration in the upstream cell, uncertainties on the chloride
493 diffusion coefficients using a simplified equation became around 31%. This did not permit to
494 differentiate the materials studied. However, with the numerical model, a difference in
495 behaviour between the three HCP of different W/C ratio could be observe (and uncertainties
496 were only about 9% using the numerical model). With a lower chloride concentration in
497 upstream and an HCP with very low porosity, some microscopic effects slowing down the
498 chloride diffusion seemed to appear. Further studies are then required to verify this result.
499 Electro-diffusion tests and simulations have to be carried out with other tracers on HCP:
500 cationic species (such as Ca^{2+} and Li^+) to observe the influence of the ionic charge and neutral
501 species (such as HTO) which are not influenced by electrochemical effects.

502 **6. Acknowledgement**

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508

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