

Structure, surface and reactivity of activated carbon: From model soot to Bio Diesel soot

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1	Structure, surface and reactivity of activated carbon: From model
2	soot to Bio Diesel soot
3	
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22	
23	Abstract
24	The present work aims to investigate and compare the structure, surface and reactivity
25	of the activated Printex U carbon black and real Bio soot samples through BET,
26	Raman, HRTEM, XPS, DRIFTS and temperature-programmed oxidation (TPO). TPO
27	evidenced that the activated Printex U carbon black was more reactive than the
28	activated Biodiesel soot (B7 and B100) generated from a real engine under both O_2
29	and NO + O_2 . BET displayed an increased surface area and a decreased pore size for
30	the activated soot. Raman and HRTEM revealed that the activated soot showed

31 decreased structural order. The morphology features also confirmed different

combustion pathway for Printex U and B100. XPS illustrated that the activation decreased surface graphene carbon and increased surface oxygen content. DRIFTS indicated that real Bio soot had higher concentration of surface functional groups than Printex U model soot, and a more apparent increase in surface functional groups was observed for the activated Printex U soot. It was proved that the reactivity of the activated Printex U is more dependent on structural and surface properties compared with that of the activated Biodiesel soot samples.

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Keywords: Biodiesel soot; Nanostructure; Soot activation; Surface property; Soot
oxidation reactivity

42

43 1. Introduction

The high efficiency and durability of Diesel engines have motivated their wide 44 applications in passenger cars and heavy-duty trucks in the world. However, the 45 incomplete combustion of fuel or poor fuel/air (F/A) mixing in Diesel engine has 46 47 caused higher emission of fine soot particles for example PM_{2.5}, thereby leading to serious public health and environment pollution [1-4]. These fine soot particles as an 48 important fraction of aerosols present in the air around urban area with big influences 49 on the climate, biosphere and citizen health [5-9]. Diesel particulate matter (PM) 50 mainly consists of carbon components, which include elemental carbon (EC), organic 51 carbon (OC), and carbonate carbon (CC), the former two have an important impact on 52 human health [10]. Polycyclic aromatic hydrocarbons (PAHs) as a class of OC are 53 54 toxic and can readily condense onto PM under exhaust conditions. The presence of 55 PAHs in air has been proven to be a key factor that leads to an increase in number of cancer cases [11-13]. 56

Nowadays, the emissions of Diesel soot particles are limited all around world [14-17]. For example, the emissions of particulate matter from cars and heavy-duty vehicles are limited to 4.5 mg/km and 10 mg/kWh, respectively [14-16] and the national standard GB 20891-2014 for non-road Diesel engines requires that PM emission limits reach 0.025g/kWh in China [17]. Thus, the soot removal technology has become an important topic on improving environment level and protecting human

health. The use of Diesel particulate filter (DPF), which is placed downstream of 63 64 Diesel oxidation catalyst (DOC), has been proven to be an effective way to remove Diesel PM emissions [1,18-22]. Its function is to capture PM through a wall-flow 65 filtration mechanism. However, the saturated accumulation of PM in DPF can cause 66 higher backpressure that has many negative impacts on engine working, thus DPF has 67 to be regenerated by PM combustion [1]. The active or passive regeneration of DPF is 68 69 crucial in controlling soot emission. The soot particulates are oxidized periodically by O₂ at higher temperatures (> 550 °C) during active regeneration and also oxidized 70 continuously over the catalysts mainly by NO₂ during passive regeneration [7,23-27]. 71 The regeneration efficiency strongly depends on the reactivity of soot, the burning of 72 high-activity soot occurs at lower temperatures with small energy demand, for 73 example low fuel consumption [1,7]. Many studies also confirmed that the reactivity 74 75 of soot is related to soot nanostructure [28,29], soot morphology [28,30], fuel types [29,31,32], catalyst materials [23-27,33] and engine operation conditions [34,35]. 76

77 Carbon is the main composition (> 80 wt.%) of Diesel soot with a small amount of 78 soluble organic fraction (SOF) and ash [36-38]. The formation of Diesel soot particulates must experience nucleation, nanoparticles aggregation and surface growth 79 steps, accompanied by the evolution of their structure [38,39]. The primary carbon 80 particles are composed of ordered and amorphous carbons. The ordered carbon 81 surface shows a graphite-like structure with higher crystalline order and the 82 amorphous carbon presents lower degree of graphitization with many active edge sites 83 and defects [40-44]. The different structural features of soot particles would determine 84 85 their physicochemical properties and thereby their reactivity [41-46].

86 Boehman et al. [47] reported that the disordered carbon with more amorphous structure had more defects and thus exhibited higher reactivity. Many authors also 87 studied the relationship between soot structure and reactivity by Raman 88 microspectroscopy and high-resolution transmission electron microscopy (HRTEM) 89 [28,43,44,48]. It was found that the higher reactivity of soot was attributed to a higher 90 degree of disorder and a higher content of molecular carbon, which is because higher 91 amount of defects in carbon structure results in an increase in surface 92 functionalization [28]. It was also reported that the reactivity of soot would be 93

94 dependent on its primary particle size characterized by TEM [29,48]. On the other 95 side, some researchers found that the presence of oxygenated additives in fuels 96 enhanced the formation of oxygenated soot and thereby increased the oxidative 97 reactivity of soot [31,43,44,48-50]. This might be ascribed to higher content of 98 oxygen in soot particles produced from oxygenated fuel combustion [31,50,51]. 99 Therefore, the use of Biodiesel fuel as oxygenated fuel can apparently improve 100 non-catalytic and catalytic reactivity of soot [31,44,48,52].

101 Besides, the catalytic materials coated on the DPF wall also play an important role on determining the catalytic oxidation reactivity of Diesel soot since they are able to 102 decrease soot oxidation temperature. A large number of publications reveal a wide 103 study of catalyst materials used for DPF regeneration, which include noble metal 104 catalysts [23,24,53], perovskite-type catalysts [54,55], and ceria-based catalysts 105 [56-61]. Among them, the latter are more desirable catalyst materials for DPF 106 regeneration than the former two due to their lower cost and higher oxygen 107 108 storage/release capacity. It was reported that ceria-based mixed oxides can generate a great quantity of active oxygen species (e.g. O^- , O_2^- , O_2^{2-}), resulting from many 109 oxygen vacancies and high mobility of oxygen, and thus exhibit higher catalytic 110 activity for soot oxidation [26,57-64]. In addition, a higher oxidizing ability of NO 111 into NO₂ over ceria-based catalysts is also an advantage in promoting DPF passive 112 regeneration [27,58,68-70]. 113

In Knauer's work [28], the authors prepared the oxidized/activated Diesel soot 114 samples through Temperature-Programmed Oxidation (TPO) method in a temperature 115 range from 25 to 600 °C under O2 and studied the differences in nanostructure 116 117 between raw and oxidized soot samples. They found that the changes in chemical homogeneity and structural order of Diesel soot occurred after partial soot oxidation. 118 In real exhaust aftertreatment system, soot particles which are generated from Diesel 119 engine follow thermal gas-flow line and go through Diesel Oxidation Catalyst (DOC) 120 unit, and then are accumulated and regenerated in DPF [22,71]. In this process, Diesel 121 soot would be partially oxidized or activated before its complete combustion. It may 122 result in an impact of soot structure and even soot oxidation reactivity. However, this 123 point is hardly noticed in the previous studies. 124

Thus, the present work aims to investigate the influence of soot activation on 125 non-catalytic and catalytic reactivity, determined by Temperature-programmed 126 Oxidation (TPO) tests, and structural and surface properties, characterized by BET, 127 Raman spectra, HRTEM, Diffuse Reflectance Infrared Fourier Transform Spectra 128 (DRIFTS) and XPS. The real Biodiesel and Biodiesel-Diesel blend soot samples 129 generated on a real Diesel engine bench were used in this work. Printex U (Degussa) 130 as a model soot has been widely used for experimental work. So its activation was 131 132 also studied and compared with real Bio soot.

133 **2. Experimental**

134 2.1. Materials preparation

135 2.1.1. Soot samples preparation

The oxidized/activated soot samples were prepared by Temperature Programmed 136 Oxidation (TPO) method under 9% b.v. O₂/Ar (250 ml/min). A carbon black 137 (Printex-U) from Degussa as representative of model soot and Biodiesel derived soot 138 B7 and B100, which separately generated from the burning of 7% Biodiesel-Diesel 139 140 blends and 100% Biodiesel in a real Diesel engine [48], were used for this study. As shown in Fig. 1, when soot-TPO was performed till 500 °C (about 3-10% conversion) 141 and 600 °C (about 20-30% conversion), respectively, the activated soot samples were 142 taken out from the reactor and marked "soot-500" and "soot-600". The raw soot 143 samples were used as a reference. 144

145



146

147 Fig. 1. Soot conversion in TPO under 9% b.v. O₂/Ar at a heating rate of 10 °C/min.

148 2.1.2. Catalyst preparation

149 MnO_x -CeO₂ mixed oxides (CM) (Mn/Ce = 1/4 (mol/mol)) were prepared through a

150 co-precipitation method and used to evaluate the catalytic activity of soot. 151 Ce(NO₃)₃·6H₂O (solid, 95% pure) and Mn(NO₃)₂ (liquid, 50 wt.%) as the precursors 152 were firstly dissolved in deionized water, then the mixture solution was mixed with 153 the precipitant (pH > 9.0) consisting of NH₃·H₂O and (NH₄)₂CO₃ (chemical reagents, 154 Beijing). The obtained precipitates were filtered and dried by a spraying apparatus, 155 then calcined in static air at 500 °C for 3 h in a muffle furnace.

156

157 2.2. Soot characterization

The particles size distribution of soot was measured on FRITSCH laser particle sizer (ANALYSETTE 22 Nano Tec). Prior to each measurement, the aggregated soot sample was well dispersed in water by a close circulatory system and an integrated ultrasonic generator.

The special surface area of soot samples was measured on an automatic surface analyzer by N₂ adsorption/desorption method at -196 °C. Prior to tests, the samples were degassed at 200 °C for 2 h under vacuum.

Raman spectra of soot were obtained on an inVia Reflex spectrometer (Renishaw, London, England) with a YAG laser (532 nm). The output power of 0.9 mW was chosen in the scanning range of 800-3500 cm⁻¹. The spectrometer includes a grating with 600 grooves mm⁻¹ and a CCD detector with \times 50 magnification objective lens.

The morphology of soot was analyzed by high-resolution transmission electron microscopy (HR-TEM) (Tecnai G^2 F20) with an acceleration voltage of 200 kV. The soot powders were dispersed in ethanol and deposited over Cu grids with a holey carbon film.

XPS measurement was conducted on an AXIS Ultra DLD spectrometer (Kratos,
England) by using AlKα radiation as the excitation source (300 W). All binding
energies (B.E.) were referenced to carbon C 1s line at 284.6 eV.

DRIFTS (diffuse reflectance infrared Fourier transform spectra) of soot were measured by using a Nicolet 6700 spectrometer with a DTGS detector and a high temperature cell. Soot sample was diluted by KBr at a soot/KBr ratio of 1/40, then the mixture was placed in a reaction cell and pretreated at 100 °C in pure N₂ for 10 min. After that, the IR spectra were recorded.

181 2.3. Reactivity measurement

The reactivity of soot samples was measured by temperature-programmed 182 oxidation (TPO) under 9% b.v. O₂/Ar and 400 ppmv NO + 9% b.v. O₂/Ar. Prior to the 183 catalytic reactivity measurement, soot and the catalyst powder were mixed by a loose 184 contact at a weight ratio of 1/10 (2.0 mg/20 mg). To minimize the impact of hot spots, 185 the soot-catalyst mixture (22 mg) was diluted with 80 mg SiC pellets. The final 186 mixture was heated in a U-shaped quartz reactor at a ramp rate of 10 °C. The flow rate 187 was controlled to 250 ml/min (15 Nl/h). 2 mg soot and 80 mg SiC pellets were mixed 188 and then used for non-catalytic reactivity measurement under 9% b.v. O₂/Ar. 189 Concentrations of CO and CO2 (ppmv) were recorded by a Siemens Ultramat6 190 analyzer. In the obtained soot-TPO curves, T_i and T_m represent the ignition 191 temperature at which the CO_x (CO + CO₂) concentration reaches 100 ppm and the 192 temperature of maximal soot oxidation rate, respectively. 193

Based on the reaction $C(\text{soot}) + (1 + x)/2O_2 \rightarrow xCO_2 + (1-x)CO$, the specific reaction rate (V_{spec}) was normalized to catalyst quality (20 mg) or soot mass (2 mg) and thus calculated by the following equation:

197

198
$$V_{spec} = \frac{(X_{CO} + X_{CO2}) \times D}{3600 \times V_M \times m_{cat.}}$$
(a)

199 or

200
$$V_{spec} = \frac{(X_{co} + X_{co2}) \times D \times 10^3}{3600 \times V_M \times m_{ini.}}$$
 (b)

- 202 Where V_{spec} : the specific reaction rate in (a) μ mol_{soot}/s/g_{cat.} or (b) mmol_{soot}/s/g_{ini.},
- 203 X_{CO} and X_{CO2} : molar fractions of CO and CO₂ in ppmv,
- 204 *D*: flow rate (15 Nl/h),
- 205 V_M : molar volume (22.4 L/mol),
- 206 $m_{cat.}$: mass of catalyst (0.02 g),
- 207 $m_{ini.}$: mass of initial soot (0.002 g)
- 208 2.4. Isothermal experiment
- Isothermal experiments of soot oxidation were conducted at 350 °C under 9% b.v.
- 210 O₂/Ar and 400 ppm NO/9% b.v. O₂/Ar, respectively, to obtain the specific reaction

rate of soot. There is no apparent increase in soot conversion at 350 °C, thus soot oxidation reaction is in a relatively stable state. High space velocity and mixing SiC powder effectively inhibit the mass and heat transfer, and thereby soot oxidation reaction can be mostly controlled by the chemical kinetics [67,68]. After isothermal reaction for 1800 s, the "initial soot oxidation rate" at the starting of isothermal reaction and the "final soot oxidation rate" at the end were thus recorded.

- 217 **3. Results and discussion**
- 218 *3.1. Reactivity analysis*
- 219 *3.1.1. TPOs in O*₂



223 Fig. 2. Non-catalytic (a, c, e) and catalytic (b, d, f) reactivities of raw and oxidized soot samples

under 9% b.v. O₂/Ar. Reaction conditions: soot mass is 2.0 mg, catalyst mass is 20.0 mg, 80 mg

225	SiC was used for dilution, t	he flow rate is 250 ml/min, h	neating rate is 10 °C/min.
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Samplas	In	O_2	In NO	In NO + O_2		
Samples	$T_{\rm i}/^{\circ}{\rm C}$	$T_{\rm m}/^{\circ}{\rm C}$	$T_{\rm i}/^{\circ}{\rm C}$	$T_{\rm m}/^{\circ}{\rm C}$		
Printex U	535 ^a /483 ^b	674 ^a /644 ^b	412 ^b	609 ^b		
Printex U-500	540/508	672/662	382	580		
Printex U-600	484/458	600/560	338	490		
B7	511/379	676/662	370	556		
B7-500	534/479	681/673	378	553		
B7-600	540/483	672/664	374	543		
B100	490/414	662/644	339	565		
B100-500	527/480	667/662	404	590		
B100-600	528/477	647/646	382	561		

Table 1. T_i and T_m values of TPOs with different soot samples.

^aThe values were obtained without catalyst.

^bThe values were obtained with catalyst.

229 To explore the impact of thermal activation or partial oxidation of soot on its reactivity, the TPO experiments of raw and activated soot samples were conducted in 230 9% b.v. O₂/Ar. Fig. 2a shows the non-catalytic activities of the raw and activated 231 Printex U samples. By comparison, the activated soot Printex U-600 (about 20% of 232 233 conversion) presents an apparently increased reactivity, its T_i and T_m values visibly decrease to 484 and 600 °C from 535 and 674 °C of raw soot, respectively (Table 1). 234 235 However, the activated soot Printex U-500 (about 3% of conversion) has no apparent changes in reactivity compared to raw soot. Table 1 also lists their similar T_i and T_m 236 values. Fig. 2b presents the catalytic activities of Printex U samples. The same 237 phenomenon can be observed that the conversion temperatures of Printex U-600 238 239 visibly shift towards lower values compared with that of raw soot, for example, the T_i and $T_{\rm m}$ separately shift to 458 and 560 °C from 483 and 644 °C (Table 1). While the 240 TPO curve of Printex U-500 seems to be similar with that of raw soot and shows a 241 slight increase in T_i and T_m values. These results illustrate that the Printex U-600 is 242 more reactive than Printex U-500 and raw Printex U. 243

The Bio soot B7 and B100 were produced on a real engine bench using a medium duty Diesel engine. Their non-catalytic and catalytic activities were also tested by TPO before and after activation and the results are shown in Fig. 2c-f. It can be seen

that for the activated B7 and B100 samples, the low-temperature reactivity decreases 247 compared to raw soot. Table 1 also illustrates that the T_i values of the activated B7 248 and B100 samples are higher than those of raw soot during both non-catalytic and 249 catalytic oxidation. Besides, T_m values of B7-600 and B100-600 are similar with those 250 of raw soot and do not visibly shift to lower temperatures (except for non-catalytic 251 reactivity of B100-600). It is also observed that the $T_{\rm m}$ of B7-500 and B100-500 252 253 slightly shifts to higher temperatures compared to that of raw samples. Thus, the partial oxidation process of real Bio soot decreases the low-temperature reactivity (T_i) 254 but cannot apparently affect the high-temperature reactivity (T_m) . These results are 255 contrary to those of Printex U samples. 256



Fig. 3. The first and second TPO profiles of raw Printex U (a), B7 (b) and B100 (c) samples.
Experimental conditions: soot mass is 2.0 mg, SiC mass is 80 mg, without catalyst, the flow rate is
250 ml/min, heating rate is 10 °C/min.

To confirm the influence of the activation step on the reactivity of soot, firstly, the raw soot samples were partially oxidized under O_2/Ar through TPO step from 25 to 600 °C, then the activated soot was cooled down to 25 °C and the second TPO was

performed to obtain the total soot conversion profile. As shown in Fig. 3, for Printex 266 U soot (Fig. 3a), the soot oxidation temperatures for the second TPO curve visibly 267 shift to lower values relative to the first TPO, which reveals a promoting role of the 268 activation step on soot oxidation. For B7 and B100 samples (Fig. 3b, c), it can be 269 clearly observed that the second TPO profile presents a decreased reactivity (< $600 \,^{\circ}$ C) 270 compared with the first TPO. These results above are in good agreement with those 271 272 presented in Fig. 2.

3.1.2. TPOs in $NO + O_2$ 273



275

276 Fig. 4. Catalytic reactivities of Printex U (a), B7 (b) and B100 (c) samples under 400 ppm NO + 277 9% O₂/Ar. Reaction conditions: soot mass is 2.0 mg, catalyst mass is 20.0 mg, soot and catalyst were mixed by loose contact and then diluted by 80 mg SiC, the flow rate is 250 ml/min, heating 278 279 rate is 10 °C/min.

It was reported that the presence of NO can promote the oxidation of soot over 280 Mn-Ce catalysts, which is ascribed to two factors: firstly, NO is able to be oxidized 281 into NO₂ with higher oxidizing ability than O₂ over Mn-Ce mixed oxides [58,68]; 282 secondly, the adsorption/desorption of NO on Mn-Ce mixed oxides can increase 283 oxygen vacancies and desorption of active oxygen [69]. In the present work, TPO 284 profiles of the catalytic oxidation of Printex U and Bio soot samples under 400 ppm 285

NO + 9% b.v. O₂/Ar are presented in Fig. 4. An apparent promotion of NO on soot 286 activity can be observed by comparing the TPO profiles in Fig. 4 and Fig. 2. In the 287 case of Printex U, in comparison with raw soot, the oxidized samples especially for 288 Printex U-600 show an increased reactivity. It can be noted in Table 1 that the T_i and 289 $T_{\rm m}$ of Printex U-500 slightly shift to lower values but those of Printex U-600 display a 290 larger shift. In the previous TPO with catalyst under O₂ (Fig. 2b), Printex U-500 291 292 presents a similar TPO curve with the raw soot, and the T_i and T_m of Printex U-600 decrease by 25 and 84 °C, respectively, relative to those of raw Printex U. However, 293 in NO + O_2 (Fig. 4a), Printex U-500 exhibits a slight enhancement on soot oxidation 294 activity and the T_i and T_m of Printex U-600 separately decrease by 74 and 119 °C 295 compared to those of raw soot. The fact illustrates that the reactivity of the activated 296 Printex U soot is more easily promoted in NO + O_2 . In the case of B7 and B100, it can 297 be clearly observed that the low-temperature reactivity of activated soot samples 298 decreases at about 200-350 °C. The results are also confirmed in Fig. 2. For B7 soot, 299 300 the high-temperature reactivities of raw B7 and B7-500 samples are almost similar in view of the close T_m values (556 and 553 °C), and that of B7-600 shows a slight 301 promotion. For B100 soot, the T_m value of B100-500 shifts towards higher 302 temperature (590 °C) and that of B100-600 (561 °C) is similar compared with raw 303 soot (565 °C). 304







Fig. 5. Isothermal catalytic reaction at 350 °C in 9% O_2/Ar (a, c, e) and 400 ppm NO + 9% O_2/Ar (b, d, f). 2 mg soot and 20.0 mg catalyst were mixed by loose contact, 80 mg SiC pellets were used for dilution, the flow rate is 250 ml/min.

	9%	O ₂	400 ppm NO + 9% O ₂			
Soot sample	Initial oxidation rate	Final oxidation rate	Initial oxidation rate	Final oxidation rate		
	$(\mu mol_{soot}/s/g_{cat.})$	$(\mu mol_{soot}/s/g_{cat.})$	$(\mu mol_{soot}/s/g_{cat.})$	$(\mu mol_{soot}/s/g_{cat.})$		
Printex U	0.1791	0.0082	0.3291	0.3484		
Printex U-500	0.2156	0.0082	0.7614	0.5066		
Printex U-600	0.3922	0.0538	1.3092	1.3532		
B7	0.3928	0.0251	0.6409	0.3055		
B7-500	0.2011	0.0433	0.5450	0.3146		
B7-600	0.1524	0.0480	0.6774	0.4604		
B100	0.4612	0.0814	0.7631	0.3238		
B100-500	0.1492	0.0243	0.7312	0.4077		
B100-600	0.0884	0.0243	0.6855	0.5471		

Table 3. The results of isothermal reaction at 350 °C.

Fig. 5 and Table 3 show the catalytic reaction rates of raw and activated soot samples under isothermal conditions. As shown in Fig. 5a and Fig. 5b, the catalytic reaction rates of raw and activated Printex U samples follow the order Printex U-600 > Printex U-500 > Printex U at 350 °C in both O_2 and NO + O_2 . This reveals

that the activated Printex U shows higher reactivity for soot oxidation. Especially in the presence of NO, a more apparent increase in the reaction rate of the activated samples is observed in Fig. 5b and Table 3.

It is interesting to undernote that, not all soot samples can show the same 322 phenomenon with Printex U after activation. As shown in Fig. 5c-f and Table 3, for 323 real soot B7 and B100, the activated soot samples exhibit decreased oxidation rates in 324 O_2 but increased results in NO + O_2 . In Table 3, it can be observed that the initial 325 326 oxidation rates of the activated B7 and B100 samples visibly decrease in O₂. A slight increase in final oxidation rates of the activated B7 samples can be also seen, but the 327 oxidation rates are too small to illustrate the real case. In the case of the TPO results 328 in NO + O_2 , the initial oxidation rates of B7 and B100 samples are ranked as follow: 329 B7-600 (0.6774 μ mol_{soot}/s/g_{cat}) > B7 (0.6409 μ mol_{soot}/s/g_{cat}) > B7-500 (0.5450 330 and B100 (0.7631 μ mol_{soot}/s/g_{cat.}) > μ mol_{soot}/s/g_{cat.}) B100-500 (0.7312 331 μmol_{soot}/s/g_{cat.}) > B100-600 (0.6855 μmol_{soot}/s/g_{cat.}), respectively. However, their final 332 oxidation rates apparently increase after soot activation (Table 3). In previous TPO 333 334 results, it can be noted that the activated B7 and B100 samples show a decreased low-temperature catalytic reactivity in both O_2 and $NO + O_2$ mixture. The results in 335 isothermal reactions under O₂ (Fig. 5c, e) are in good agreement with this point, but 336 those in NO + O_2 (Fig. 5d, f) exhibit a contrary result. The fact illustrates that under 337 isothermal conditions the presence of NO is more favorable for the catalytic oxidation 338 of activated Bio soot. 339

Moreover, it can be noticed that for the isothermal reactions under O_2 the final 340 341 reaction rates sharply decline relative to the initial ones, but the presence of NO into 342 O_2 apparently inhibits the decrease of reaction rates and makes the final reaction rates remain higher values. The result would be attributed to the following reasons: (1) 343 MnO_x-CeO₂ mixed oxides have higher catalytic oxidation activity of NO into NO₂ 344 [57,58,69], the produced NO₂ with stronger oxidizing ability than O_2 can enhance the 345 soot oxidation rate; (2) the production of NO₂ over MnO_x-CeO₂ catalyst benefits C + 346 $NO_2 + O_2$ cooperative reactions [58,70]; (3) it was confirmed that NO 347 adsorption/desorption on MnO_x-CeO₂ enhances the production of active oxygen in the 348 oxides [69]. 349

In Seong's work [72], the authors analyzed the burning process of soot through 350 partially oxidizing Printex-U soot under O2-NO2 mixtures and found that the 351 oxidation-derived maturing process was much dependent on NO₂ content. Besides, 352 the presence of NO_2 during soot oxidation has been confirmed to have a good 353 promoting role on the oxidative reactivity of soot [23,24,58,64,72]. Thus, we suppose 354 that the activation of soot under different reaction gases would cause different 355 oxidation reactivities of the activated soot. In the present work, Printex U soot was 356 chosen and partially oxidized by TPO from 25 to 500 °C in 9% O₂/Ar, 400 ppm NO + 357 9% O₂/Ar and 400 ppm NO₂ + 9% O₂/Ar, respectively. Then these activated soot 358 samples were used for isothermal catalytic reactions under NO + O_2 at 350 °C to 359 compare their differences. As presented in Fig. S1 (see Supplementary Materials), the 360 isothermal reaction curves of the activated soot samples seem to have no more 361 apparent differences. When the reaction rate reaches a stabilized value, it can be 362 clearly seen that the activated soot under NO + O_2 shows a similar oxidation rate 363 (about 0.51 μ mol_{soot}/s/g_{cat.}) compared to that one under O₂, but under NO₂ + O₂ a 364 slightly increased oxidation rate (about 0.57 µmol_{soot}/s/g_{cat.}). This might be because 365 the presence of NO₂ affects the structure of soot particles and thereby the oxidation 366 reactivity during the activation process [34,72]. 367

- 368 3.2. Soot characterization
- *369 3.2.1. Laser granulometry*



Fig. 6. Particles aggregates size distribution of Printex U (a) and B100 (b) samples.



372

373 Fig. 7. Mean size of particles aggregates of different soot samples.

Fig. 6 presents the size distributions of particles aggregates of raw and activated 374 Printex U and B100 samples. It can be clearly seen that the activated soot samples 375 show different size distribution profiles compared to raw soot. For the two soot 376 377 samples Printex U and B100, it is found that the activation step leads to a decrease in the amount of smaller soot particles and the whole distribution profiles of activated 378 soot shift towards higher values. The same result can be also observed for the raw and 379 activated B7 soot samples (Fig. S2, see the Supplementary Materials). As shown in 380 Fig. 7, the mean size of soot particles aggregates increases after the activation, with 381 following the order: raw soot < soot-500 < soot-600. Especially for the real soot B7 382 and B100, the activated soot samples exhibit a larger increase in particles aggregates 383 sizes than Printex U. These results also illustrate that the soot particles with smaller 384 aggregates sizes would be firstly oxidized during soot oxidation. 385

386

387 *3.2.2. BET test*

Soot sample	$S_{\rm BFT} (m^2/g)^a$	Mean pore size <i>R</i>	Pore volume $V_{\rm p}$		
		(nm) ^a	(ml/g) ^a		
Printex U	97.8	18.94	0.90		
Printex U-500	111.8	19.07	1.05		
Printex U-600	321.3	1.21	1.30		
B100	124.1	39.01	1.26		
B100-500	154.9	1.64	0.92		
B100-600	217.1	1.21	1.29		
B7	231.6	25.55	2.73		
B7-600	296.2	1.21	1.91		

388 Table. 4. Textual properties of raw and activated soot samples

389 ^awhich was calculated by BJH method.

Table 4 lists the values of special surface area (S_{BET}), mean pore size (R) and pore 390 volume (V_p) of the raw and activated soot samples. The activation of raw soot 391 significantly increases the surface area and decreases the mean pore size. This 392 illustrates that the activation process results in a modification of textural property of 393 soot. For Printex U samples, Printex U-500 shows a slight increase in S_{BET} and R, 394 395 while Printex U-600 has a sharp increase in S_{BET} and a decrease in R. The pore volume (V_p) of Printex U also exhibits an increased trend after the activation. For real 396 Bio soot B7 and B100, a similar change in S_{BET} and R with Printex U can be observed 397 after activation, but one different point is that V_p values of B7-600 and B100-500 398 decrease compared to raw soot. The sharp decrease of mean pore size for the activated 399 soot indicates an apparent increase in the amount of smaller pores and a 400 disappearance of large pores. It is speculated that the initial soot oxidation may occur 401 in larger pores. 402

403



404 3.2.3. Raman analysis



407 Fig. 8. Raman spectra of raw and partially oxidized Printex U (a), B7 (b) and B100 (c) samples.



411 Fig. 9. Analysis of the Raman spectra and curve fits of raw Printex U (a), Printex U-500 (b) and

- 412 Printex U-600 (c).

414	Table 5.	Raman	results	of	soot	samp	les.

Sample	FWHM _{D1} /cm ⁻¹	FWHM _{D2} /cm ⁻¹	FWHM _G /cm ⁻¹	$I_{\rm D1}/I_{\rm G}$	$I_{\rm D2}/I_{\rm G}$	$I_{\rm D3}/I_{\rm G}$	$A_{\rm D1}/A_{\rm G}$	$A_{\rm D3}/A_{\rm G}$
Printex U	174	47	70	1.04	0.87	0.90	2.7	0.75
Printex U-500	182	57	81	1.04	0.97	0.98	3.1	1.09
Printex U-600	179	51	81	1.03	0.97	0.97	3.3	0.97
B7	153	42	70	1.01	0.30	0.42	2.3	0.39
B7-500	149	39	67	1.07	0.37	0.44	2.6	0.44
B7-600	151	47	71	1.10	0.42	0.43	2.6	0.42
B100	154	37	61	0.98	0.38	0.47	2.3	0.42
B100-500	145	31	63	1.10	0.41	0.51	2.4	0.51
B100-600	166	31	65	1.09	0.34	0.49	2.4	0.56

The Raman spectra of raw and activated soot samples were measured to analyze the 416 microstructural features and the results are presented in Figs. 8, 9 and Table 5. Fig. 9 417 presents the Raman analysis of Printex U samples, and that of B7 and B100 samples 418 can be also found in Fig. S3 and Fig. S4 (see Supplementary Materials). The Raman 419 spectra of these soot samples are characterized by two overlapping peaks, which 420 appear at around 1580 cm⁻¹ (G peak) and 1350 cm⁻¹ (D peak), respectively [28,38,73]. 421 422 The D peak is generally related to defect sites or disordered carbon and G peak 423 involves the ordered carbon. As shown in Fig. 8, in comparison with untreated raw soot, which has more separated D and G peaks, the Raman spectra for activated 424 425 samples show more strongly overlapping peaks. This might imply an increase in the content of molecular carbon in the activated soot [28]. Moreover, in Fig. 8, it is also 426 observed that raw soot samples exhibit broad Raman signals in the range of about 427 2300-3200 cm⁻¹, these should be ascribed to second-order bands due to (2*D) 428 overtones and (D + G) combinations of graphitic lattice vibration modes [38]. The 429 430 second-order signals become less pronounced after activation treatment for Printex U soot (Fig. 8a), but the visible decrease in the second-order signals cannot be seen for 431 the activated B7 and B100 soot (Fig. 8b and c). The result may illustrate that Printex 432 U soot becomes more disordered than B7 and B100 soot after activation. 433

Fig. 9 clearly illustrates the Raman spectrum of soot consists of four 434 Lorentzian-shaped bands (G, D1, D2, D4) at around 1580, 1350, 1620 and 1200 cm⁻¹, 435 respectively, and a Gaussian-shaped band D3 at about 1500 cm⁻¹ [28,73]. It has been 436 reported that D1 band is associated with disordered graphitic lattice and D3 band is 437 438 related to amorphous carbon or molecular carbon, thus D1 FWHM (full width at half maximum) and D3 band intensity as important parameters can provide some 439 information about chemical structure of graphite-like and amorphous carbon [28,29]. 440 Many earlier works have also reported the FWHM of G and D bands and the D/G 441 band intensity ratios (I_D/I_G) can correlate with degree of graphitization of 442 carbonaceous materials [74-79]. The authors thought that an increase in G and D band 443 FWHM and I_D/I_G values indicates a decrease in order. It was also published that the 444 ratios between D bands and G band areas (A_D/A_G) are related to the structural defects 445 in graphene layers [29]. Since it is unclear which of these parameters should be used 446

447 for Raman analysis, most of parameters are used here for comparison.

Table 5 lists spectroscopic parameters of different soot samples. For Printex U soot, 448 it is observed that the activated soot samples (Printex U-500 and -600) exhibit an 449 450 increased spectroscopic parameters including D and G band FWHM, I_D/I_G and A_D/A_G . This result illustrates that the activated soot has higher disorder and more structural 451 defects. Among these parameters, D1 band FWHM and A_{D1}/A_{G} values of the activated 452 soot samples increase, but their I_{D1}/I_G values are very similar with that of raw soot. 453 454 Here D1 FWHM should be more important parameter for analyzing carbon structure. The relative intensity of D3 band (I_{D3}/I_G) is associated with amorphous carbon or 455 molecular carbon. The increase of I_{D3}/I_G value of the oxidized soot samples indicates 456 an increase in molecular carbon. However, by comparison, the two activated Printex 457 U samples (Printex U-500 and -600) show similar spectroscopic parameters, which 458 seems to imply their similar structural features. 459

In the previous work [48], it has been proposed that the non-catalytic and catalytic 460 reactivities of soot are correlated to their structural properties. In the present work, the 461 previous TPOs results have confirmed that the activated Printex U samples show an 462 improved reactivity compared to raw sample, this well agrees with the Raman results. 463 Nonetheless, the similar Raman results of the two activated samples cannot explain 464 much higher reactivity of Printex U-600 than the other one. This may illustrate that 465 the reactivity of soot cannot be only ascribed to the structure but also to particle 466 morphology [48,80], ash content and composition [81,82] as well as oxygen content 467 [49]. 468

469 On the other hand, in the case of real soot B7 and B100, it can be seen in Table 5 470 that they have different structural features after activation compared with model soot Printex U. For B7 soot, as listed in Table 5, the activated soot B7-500 shows slightly 471 decreased FWHM values, which appears to imply an increase in order. However, the 472 I_D/I_G and A_D/A_G exhibit increased values, this seems to reveal an inverse result. It has 473 been reported that D band FWHM and I_D/I_G have no significant correlation with each 474 475 other [38]. Thus, it is difficult to well analyze the nanostructure of soot by Raman measurement and it may need more evidences to illustrate the nanostructural features. 476 For B7-600, D1 and G band FWHM values as well as I_{D3}/I_G ratio are very similar 477

with those of raw B7 sample. But other parameters especially for D2 band FWHM and I_{D2}/I_G exhibit an increased tend. It was reported that the I_{D2}/I_G ratio and the D2 band FWHM should be inversely proportional to the thickness of graphitic domains in soot [38,76]. Therefore, the increase in D2 band FWHM and I_{D2}/I_G might indicate the decrease of graphitic carbon.

For B100 soot, it can be seen in Table 5 that B100-500 sample shows apparently 483 decreased D1 and D2 bands FWHM but increased G band FWHM, I_D/I_G and A_D/A_G 484 485 values compared to raw soot. In general, the intensity of G band, D1 and G band FWHM as well as I_{D1}/I_G are related to the graphitic order. The decrease in D1 band 486 FWHM seems to reveal an increase in graphitic order. However, the G band FWHM 487 and I_{D1}/I_G increase, this inversely indicates a decrease in order. Additionally, I_{D3}/I_G as 488 important parameter is often used to evaluate the amorphous carbon content of soot. 489 The increase of I_{D3}/I_G implies an increase in amorphous carbon content. Plus the 490 increased tend of A_{D1}/A_G and A_{D3}/A_G , these results may illustrate an increase in 491 disorder for B100-500. In the case of B100-600, it can be observed in Table 5 that D1 492 493 and G band FWHM as well as A_{D3}/A_G exhibit higher values than both B100 and B100-500 samples, and I_{D1}/I_G , I_{D3}/I_G and A_{D1}/A_G show very close values with those of 494 B100-500. The result illustrates B100-600 has a lower order compared to the raw and 495 B100-500 samples. 496

497

498 *3.2.4. HRTEM*

499







502 Fig. 10. HRTEM images of Printex U (a, b), Printex U-500 (c) and Printex U-600 (d).



504



Fig. 11. HRTEM images of B100 (a, b), B100-500 (c) and B100-600 (d). 506

507 In order to better observe the structural and morphology changes of soot after 508 activation, the morphology and nanostructural properties of Printex U and B100 soot samples were characterized by HRTEM, and the results are shown in Fig. 10 and Fig. 509 11. In the case of Printex U soot, as shown in Fig. 10a, it exhibits turbostratic patterns 510 and an onion-like structure, the "tight" fringe layers concentrically surround nucleus 511

where the disordered carbon structure is observed. The raw Printex U also has a larger 512 number of parallel stacked graphene layers with a larger thickness (about 19.5 nm) 513 (Fig. 10b). This reveals a higher order for raw Printex U. Moreover, some edge 514 defective sites are also observed in Fig. 10a. After the soot being activated, a 515 decreased fringe-layer thickness can be clearly seen in Fig. 10c and d (about 11.5 nm 516 for Printex U-500 and about 7.3 nm for Printex U-600). The result illustrates that the 517 activation of Printex U decreases the degree of graphitization and Printex U-600 518 519 exhibits more apparent result. Besides, in comparison, it can be observed that the activated soot samples have an increased area of highly disordered carbon particles on 520 521 the center. For example, for Printex U-600, the fringe patterns on the center become less noticeable, this is due to the oxidation of less-ordered nuclei through diffused 522 oxygen into porous particles [83]. Thus, it is also confirmed that Printex U 523 experienced an internal burning-out process during partial oxidation under O₂ [34,72]. 524 The B100 Biodiesel soot exhibits different structural and morphology changes after 525 activation compared with Printex U. As shown in Fig. 11a and b, the raw soot displays 526 less obvious fringe layers and thinner carbon-layers thickness (about 5-10 nm) than 527 raw Printex U soot. A less chainlike orientation and looser fringe layers are also 528 observed in Fig. 11b. These results seem to illustrate that the real Bio soot has higher 529 disorder and lower degree of graphitization than Printex U, which is also found in the 530 previous work [48]. After partial oxidation, the activated samples B100-500 and 531 B100-600 all show unclear fringe layers, especially for the edge area (red arrow). This 532 might confirm that the activated samples become more disordered. In comparison in 533 534 Fig. 11c, the edge of the B100-500 soot particle shows vaguer fringe layers than the 535 area inside. Some void areas on the outer surface of primary particle are also found in Fig. 11d. Therefore, B100 soot tended to be oxidized through an external burning 536 process during the activation [72]. 537

Moreover, 50 soot particles were used for nanoparticles size analysis for each sample and nanoparticle-size distribution is presented in Fig. S5. With the soot being activated, the nanoparticles sizes of Printex U and B100 tend to decrease. Thus, the activated soot presents a smaller mean particle size than raw soot. This observation was also reported in Al-Ourashi's work [34].



544 *3.2.5. XPS analysis*

543

547 Fig. 12. XP spectra of C1s, O1s of the raw Printex U and B100 samples.

Complea	Pri	Printex U (at.%)			B100 (at.%)		
Samples	С	0	Graphene	С	0	Graphene	
Raw soot	90.3	9.7	59.4	80.9	19.1	49.1	
Soot-500	90.1	9.9	54.6	80.3	19.7	42.0	
Soot-600	89.2	10.8	45.1	78.0	22.0	37.7	

548 **Table 6.** XPS results of Printex U and B100 samples.

549

Fig. 12 presents the XP spectra of the raw Printex U and B100 samples in the C 1s 550 and O 1s energy regions, and those of the activated soot samples are shown in Fig. S6 551 and Fig. S7. It can be clearly seen in C 1s XPS that these soot samples mainly consist 552 of graphene, aliphatic carbon and molecular carbon (R-OH, C-O-C, -COOH, 553 -C(O)-O-C, -C=O, etc.) [84,85]. These molecular carbon species are also detected in 554 O 1s spectra. Table 6 lists the relative contents of surface elements of the detected 555 556 soot samples. It can be observed that the real Biodiesel soot shows higher oxygen content and lower content of graphene component than Printex U soot, this indicates 557

that Biodiesel soot has a higher disorder, which was also confirmed in the previous 558 work [48]. After the soot was activated at different temperatures, the activated soot 559 samples show different XPS results. For both Printex U and B100, the activated 560 samples all show an increased oxygen content and a decreased graphene content. By 561 comparison, soot-600 has a more apparent increase in oxygen content and a larger 562 decrease in graphene carbon than soot-500. Thus, it shows higher structural disorder, 563 this well agrees with the results of both Raman spectroscopy and HRTEM. The result 564 is well correlated with the reactivity of the activated Printex U soot but not with that 565 of the activated B100 soot (Table 1). 566

567



568 3.2.6. DRIFTS

569

570 Fig. 13. DRIFTS results of Printex U and B100 samples.

The DRIFT spectra of raw and activated Printex U and B100 soot samples were 571 investigated to obtain more information about oxygen functional groups on the carbon 572 surface. The obtained IR spectra are shown in Fig. 13. According to the DRIFTS 573 results and the previous reports [87-91], the bands at around 1205 and 1090 cm⁻¹ are 574 assigned to the stretching of C-OH and the bands at 1550 and 1420 cm⁻¹ are ascribed 575 to the symmetric and asymmetric stretching of COO in the aromatic carboxylate 576 species, as well as some aromatic groups are very likely to be found at around 1515 577 cm⁻¹. The stretching of C=O in lactones and carboxylic anhydrides can be observed at 578 1669 and 1745 cm⁻¹ as well as that in quinone and ceto-enol groups at 1610 cm⁻¹. The 579 weak band at around 1250 cm⁻¹ may be associated with C-O stretching. It was 580 reported that the formation of some surface oxygen complexes (SOCs) could be 581 confirmed at 1730-1750 cm⁻¹ (carboxylic anhydrides and lactones), 1595 and 1615 582

cm⁻¹ (quinone and ceto-enol groups) as well as 1255 cm⁻¹ (ethers and phenols) [86]. 583 Moreover, in another publication [30], it was reported that the band at 1726 cm⁻¹ is 584 ascribed to the vibration of C=O in aromatic ester groups and the band at 1595 cm⁻¹ 585 can be attributed to the stretching vibration of conjugated C=C bonds or the bending 586 mode of H₂O. The apparent absorption bands between 1726 and 1615 cm⁻¹ are 587 difficult to assign in details. The bending vibration of C-H is also found at about 1460, 588 1395 and 1390 cm⁻¹ [30,87]. The band at 1600 cm⁻¹ could be attributed to the 589 presence of H₂O and the bands at 1300-1000 cm⁻¹ are assigned to C-O vibrations [30]. 590 As shown in Fig. 13, the real Biodiesel soot B100 exhibits more surface oxygen 591 functional components than Printex U soot in view of much stronger absorption 592 signals, especially for the surface components (such as aromatic groups) appearing at 593 1600-1400 cm⁻¹. This result would give a better explanation for the XPS result that 594 B100 soot exhibits higher oxygen content and lower carbon content than Printex U. In 595 Fig. 13a, it can be seen that the IR result of Printex U-500 seems to be similar with 596 597 that of raw Printex U, but an apparent change can be observed for Printex U-600 with an increased absorbance signal. This illustrates that the contents of surface functional 598 groups, for example, aromatic groups (1550, 1515, 1460 and 1420 cm⁻¹), C-H (1460, 599 1395 and 1390 cm⁻¹) and C=O (1750-1726 cm⁻¹), visibly increase on the surface of 600 Printex U-600. It is also noted that the C=C components of Printex U-600 obviously 601 decrease compared to raw and Printex U-500 soot. The previous XPS results reveal 602 that Printex U-600 shows more apparent changes in oxygen and graphene contents 603 604 than Printex U-500, relative to raw soot. The IR results here would be a reasonable 605 explanation for this point. Besides, in Fig. 13b, Compared with raw B100, B100-600 exhibits a similar IR result and B100-500 has a slightly decreased IR signal. This 606 indicates that there is no an obvious change in surface oxygen functional species for 607 the raw and activated Biodiesel soot. 608

609

610 *3.3. Correlation of structural and surface properties to reactivity*

611 Many authors have analyzed the relationship between soot structure and oxidative 612 reactivity by using XRD, Raman and HRTEM characterizations and found that the 613 reactivity of soot depends on its structural properties, i.e. length scale, amorphous

carbon, the degree of graphitization, nanoparticle size and surface morphology 614 [29,38,40,47,73,90]. Wu et al. [86] studied the surface organic components (such as 615 aromatic carboxylate, carboxylic anhydrides and lactones) of soot during soot 616 combustion in the presence of catalysts through DRIFTS tests and summarized that 617 the catalytic reactivity of soot is related to the concentration of surface organic 618 components. Müller et al. [30] investigated the correlation of the non-catalytic 619 reactivity of different soot samples to surface functional groups characterized by 620 621 DRIFTS measurements and confirmed an important role of the functionalization on the onset of soot combustion. 622

Fig. 6 and Fig. 7 reveal that all the activated Printex U and Biodiesel soot samples 623 present an increase in particles aggregates size. The previous study has discussed the 624 relationship between particles aggregates size and soot reactivity, being found that 625 smaller particles aggregates are more reactive than larger ones [48]. The result in the 626 present work cannot illustrate this point. For example, the activated soot Printex 627 U-600 exhibits the best reactivity in both O_2 and $NO + O_2$, but its mean particles 628 629 aggregates size is the largest one. This might indicate that smaller soot particles aggregates have been oxidized during the activation. In the case of Biodiesel soot 630 samples, the activated soot shows a decreased low-temperature reactivity (Fig. 3), this 631 may be attributed to the loss of smaller particles aggregates. 632

It was reported that high surface area of soot could benefit to improve its oxidative 633 reactivity [85]. The previous BET results reveal that the activation of soot samples 634 increases their specific surface area. Thus, the activated Printex U soot samples, 635 636 especially for Printex U-600, show an apparent enhancement in both non-catalytic and 637 catalytic reactivity. However, this is not evident for Bio soot samples in correlating surface area to their reactivity. On the contrary, in the case of isothermal reactions at 638 350 °C, the activated Bio soot samples exhibit an increase in their "final oxidation 639 rate" under NO + O_2 (Table 3), which can be well correlated with their increased 640 surface area. 641

Raman spectroscopy is an important method to study nanostructural features of soot, D1 band FWHM and the relative intensity of D3 band (I_{D3}/I_G) were reported to be the main parameters to evaluate soot structure [28,29]. Thus, Fig. 14 presents the

correlation of the reactivities (T_m) of Printex U and B100 samples with structural 645 features. For Printex U samples, as shown in Fig. 14, the two activated soot samples 646 (Printex U-500 and -600) show similar and higher D1 band FWHM and I_{D3}/I_{G} ratios, 647 revealing their increased disorder structure. However, the non-catalytic and catalytic 648 reactivities of soot in O₂ follow the order: Printex U-600 > Printex U-500 \approx Printex U 649 and the catalytic reactivity in NO + O_2 follows the order: Printex U-600 > Printex 650 U-500 > Printex U. This result cannot be well correlated to Raman parameters. 651 Besides, in previous HRTEM results, it can be observed that the degree of 652 graphitization changes with the order: Printex U-600 < Printex U-500 < Printex U. 653 The content of graphene carbon on the surface of Printex U samples also follows this 654 order (Fig. 15). The results can well agree with the catalytic reactivity in NO + O_2 . On 655 the other hand, Fig. 15 shows that the order of surface oxygen content of Printex U 656 samples is Printex U-600 (10.8%) > Printex U-500 (9.9%) \approx Printex U (9.7%). The 657 previous DRIFTS results also reveal that Printex U and Printex U-500 have close 658 659 quantities of surface functional groups and Printex U-600 has an apparent increase in that compared to them. These conclusions are well correlated to the non-catalytic and 660 catalytic reactivities of soot in O₂. 661

In the case of B100 soot samples, as shown in Fig. 14 and Fig. 15, it can be known 662 that the non-catalytic reactivity in O_2 follows the order: B100 \approx B100-500 \leq 663 B100-600 and the catalytic reactivities in both O_2 and $NO + O_2$ follow the order: 664 B100-500 < B100 \approx B100-600. It is also observed in Fig. 14 that B100-500 shows 665 lower D1 FWHM value than B100 and B100-600, indicating an increase in graphitic 666 order, which seems to be correlated to the catalytic reactivity. The I_{D3}/I_G ratio of 667 668 B100-500 exhibits a slightly increased value than that of B100 and B100-600, this indicates higher content of amorphous carbon in B100-500 soot. However, that cannot 669 agree with the reactivity. In the previous discussion on Raman results, it was 670 mentioned that it might be difficult to analyze the nanostructure of soot by only using 671 Raman. The HRTEM results reveal that the activated soot samples show a higher 672 disorder than the raw B100. In Fig. 15, the XPS results also indicate that the contents 673 of surface graphene carbon and oxygen follow the order: B100 > B100-500 >674 B100-600. But these results cannot be well correlated to the reactivity of B100 675

samples. On the other side, it is noted in Fig. 13b that B100 and B100-600 show
similar DRIFT spectra and slightly higher concentration of surface functional groups
than B100-500, which is well correlated to their reactivities.





680

681 Fig. 14. Correlation of the non-catalytic and catalytic reactivities of Printex U and B100 with

682 Raman parameters including D1 band FWHM and I_{D3}/I_G .

683



684

Fig. 15. Correlation of the non-catalytic and catalytic reactivities of Printex U and B100 with surface oxygen and graphene contents obtained by XPS.

In fact, it is difficult to correlate structural and surface properties to the reactivity of 687 soot, because the factors that affect soot oxidative reactivity are complex, not only 688 related to the structure and surface property but also to soot oxidation reaction 689 conditions. For example, in the previous isothermal reactions at 350 °C under O₂, the 690 soot oxidation rates of raw and activated Printex U and B100 follow the order: Printex 691 U (0.0082 μ mol_{soot}/s/g_{cat.}) \approx Printex U-500 (0.0082 μ mol_{soot}/s/g_{cat.}) \leq Printex U-600 692 $(0.0538 \ \mu mol_{soot}/s/g_{cat.})$ and B100 $(0.0814 \ \mu mol_{soot}/s/g_{cat.}) > B100-500 \ (0.0243)$ 693 μ mol_{soot}/s/g_{cat.}) \approx B100-600 (0.0243 μ mol_{soot}/s/g_{cat.}), respectively. However, under NO 694

+ O_2 , their oxidation reaction rates all follow the order: raw soot < soot-500 < 695 soot-600. It is also noticed that the isothermal reaction rates under NO + O_2 can be 696 well correlated to HRTEM and XPS results. Moreover, some authors tried to link the 697 698 nanostructure characterized by Raman to soot oxidative reactivity and found there was no an apparent relationship between the reactivity and Raman results [48,92]. In 699 this work, similarly, the Raman results cannot be also well correlated to the reactivity 700 701 of soot. But some researchers also confirmed that the nanostructure analyzed by 702 Raman is one of the factors that determine the oxidation reactivity of soot [28,29,90]. It was also reported that the reactivity of soot is not only related to the nanostructure 703 but to soot composition [80,84,93], oxygen content [49] and surface functional groups 704 [30,91]. Thus, a co-influence of various factors should be considered in analyzing 705 soot oxidative reactivity. 706

Besides, in real applications of aftertreatment technology, the activation of Diesel 707 soot by partial oxidation with O₂ will take place during soot accumulation and DPF 708 709 regeneration. The activating process influencing physicochemical properties and the reactivity of soot should be studied, because this is very useful for understanding DPF 710 regeneration and optimizing aftertreatment technology. This study gives valuable 711 findings on the changes in the structure, surface and non-catalytic/catalytic reactivity 712 713 of Printex U model and real Bio soot before and after the activation. In comparison, the two types of soot show different changing trends. Many researchers often use 714 Printex U model to investigate the oxidative reactivity of Diesel soot, this might not 715 confirm the real case of the combustion of Diesel soot produced from vehicles' 716 717 engines. This would require to select more suitable model soot in experimental studies. 718 Furthermore, this work also reveals that the activation of soot is an important factor that affects DPF regeneration. Thus, the changes in soot structure and surface during 719 soot combustion should be also concerned in future study. 720

721

722 **4. Conclusions**

The present work studied and compared the structure, surface and reactivity of the activated Printex U model and real Bio soot samples through a series of characterizations and TPO measurements. In the case of model soot Printex U, it was

confirmed that the activated samples exhibited an apparently increased reactivity 726 especially under NO + O_2 . The result should be related to their higher surface area, 727 higher disorder structure, increased surface oxygen content and surface functional 728 729 groups, determined by BET, Raman, HRTEM, XPS and DRIFTS. On the other hand, after the activation of real Bio soot, TPO results revealed that there was no an obvious 730 enhancement in the reactivity and even the ignition (T_i) occurred at higher 731 732 temperature. This could not well correlated with the structure and surface evolving 733 towards higher disorder and oxygen content. However, the isothermal reactions under $NO + O_2$ had a good correlation with the structure and surface. Besides, HRTEM 734 indicated different oxidation routes of Printex U and Biodiesel soot. An important 735 finding in this work is to reveal the impact of the activation during soot accumulation 736 and DPF regeneration on the structure, surface and reactivity of Diesel soot, and 737 present the different trends of model and real Bio soot. It will provide some important 738 information in understanding and optimizing DPF regeneration. 739

740

741 **References**

- [1] Reşitoğlu IA, Altinişik K, Keskin A. The pollutant emissions from diesel-engine vehicles and
 exhaust aftertreatment systems, Clean Technol Environ Policy 2015;17:15-27.
- 744 [2] Twigg MV. Progress and future challenges in controlling automotive exhaust gas emissions.745 Appl Catal B 2007;70:2-15.
- [3] Giechaskiel B, Alföldy B, Drossinos Y. A metric for health effects studies of diesel exhaust
 particles. Aerosol Sci 2009;40:639-51.
- [4] Zhou HQ, Li XR, Zhao WH, Liu FS. Effects of separated swirl combustion chamber
 geometries on the combustion and emission characteristics of DI diesel engines. Fuel
 2019;253:488-500.
- [5] Carslaw KS, Boucher O, Spracklen DV, Mann GW, Rae JG, Woodward S, et al. A review of
 natural aerosol interactions and feedbacks within the earth system. Atmos Chem Phys
 2010;10:1701-37.
- [6] Mahowald N, ward DS, Kloster S, Flanner MG, Heald CL, Heavens NG, et al. Aerosol impacts
 on climate and biogeochemistry. Annu Rev Environ Resour 2011;36: 45-74.
- 756 [7] Niessner R. The many faces of soot: characterization of soot nanoparticles produced by

- r57 engines. Angew Chem Int Ed 2014;53:12366-79.
- 758 [8] Geng P, Yao C, Wang Q, Wei L, Liu J, Pan W, Han G. Effect of DMDF on the PM emission
- from a turbo-charged diesel engine with DDOC and DPOC. Appl Energy 2015;148:449-55.
- 760 [9] Liu ZG, Ottinger NA, Cremeens CM. Vanadium and tungsten release from V-based selective
- 761 catalytic reduction diesel aftertreatment. Atmos Environ 2015;104:154-61.
- 762 [10] Mei DQ, Zhu ZN, Mei CW, Chen Z, Yuan YN. Fractal morphology features and carbon
- component analysis of diesel particulates. Environ Sci Pollut Res 2019;26:14014-23.
- 764 [11] Roy R, Jan R, Gunjal G, Bhor R, Pai K, Satsangi PG. Particulate matter bound polycyclic
- aromatic hydrocarbons: Toxicity and health risk assessment of exposed inhabitants. AtmosEnviron 2019;210:47-57.
- 767 [12] Kumar SN, Verma P, Bastia B, Jain S. AK health risk assessment of polycyclic aromatic
 768 hydrocarbons: a review. J Pathol Toxicol 2014;1:16-30.
- 769 [13] Lin YC, Li YC, Shangdiar S, Chou FC, Sheu YT, Cheng PC. Assessment of PM_{2.5} and PAH
- content in PM_{2.5} emitted from mobile source gasoline-fueled vehicles in concomitant with the
 vehicle model and mileages. Chemosphere 2019;226:502-8.
- [14] European Parliament. Regulation (EC) No 715/2007 of the European Parliament and theCouncil of 20 June 2007 on type approval of motor vehicles with respect to emissions from light
- passenger and commercial vehicles (Euro 5 and Euro 6) and on access to vehicle repair and
 maintenance information, O.J. L171/1 (2007).
- [15] European Commission. Commission Regulation No 582/2011 of 25 May 2011 implementing
- and amending Regulation (EC) No 595/2009 of the European Parliament and of the Council with
- respect to emissions from heavy duty vehicles (Euro VI) and amending Annexes I and III to
- 779 Directive 2007/46/EC of the European Parliament and of the Council, O.J. L167/1 (2011).
- 780 [16] European Commission. Commission Regulation (EU) No 459/2012 of 29 May 2012
- amending Regulation (EC) No 715/2007 of the European Parliament and of the Council and
- 782 Commission Regulation (EC) No 692/2008 as regards emissions from light passenger and
- 783 commercial vehicles (Euro 6), O.J. L142/16 (2012).
- 784 [17] Zhong H, Tan J, Wang YL, Tian JL, Hu NT, Cheng J, Zhang XM. Effects of a Diesel
- particulate filter on emission characteristics of a China II non-road Diesel engine. Energy Fuels2017;31:9833-39.
- 787 [18] Bensaid S, Marchisio DL, Fino D. Numerical simulation of soot filtration and combustion

- within diesel particulate filters. Chem Eng Sci 2010;65:357-63.
- [19] Mokhri MA, Abdullah NR, Abdullah SA, Kasalong S, Mamat R. Soot filtration recent
 simulation analysis in diesel particulate filter (DPF). Procedia Eng 2012;41:1750-5.
- 791 [20] Caliskan H, Mori K. Environmental, enviroeconomic and enhanced thermodynamic analyses
- of a diesel engine with diesel oxidation catalyst (DOC) and diesel particulate filter (DPF) after
- 793 treatment systems. Energy 2017;128:128-44.
- [21] Lee J, Theis JR, Kyriakidou EA. Vehicle emissions trapping materials: Successes, challenges,
 and the path forward. Appl Catal B 2019;243:397-414.
- [22] Shao SS, Cheng JN, Zhang B, Ma NJ, Zhang DX, Wang YL, et al. Effects of a DOC plus
- 797 DPF system on emission characteristics of China Pi engineering vehicle diesel engine and
- 798 influence factors of trapping efficiency of PM for DOC plus DPF system. Energy Source Part A
- 799 2019;41:527-41.
- [23] Liu S, Wu XD, Weng D, Li M, Fan J. Sulfation of Pt/Al₂O₃ catalyst for soot oxidation: High
 utilization of NO₂ and oxidation of surface oxygenated complexes. Appl Catal B
 2013;138-139:199-211.
- 803 [24] Jeguirim M, Tschamber V, Ehrburger P. Catalytic effect of platinum on the kinetics of carbon
 804 oxidation by NO₂ and O₂. Appl Catal B 2007;76:235-40.
- [25] Leistner K, Nicolle A, Costa PD. Impact of the catalyst/soot ratio on diesel soot oxidation
 pathways. Energy Fuels 2012;26:6091-7.
- 807 [26] Zhang HL, Hu W, Zhou CX, Liu HR, Yao P, Wang JL, et al. A new understanding of
- 808 CeO₂-ZrO₂ catalysts calcinated at different temperatures: Reduction property and soot-O₂ reaction.
- 809 Appl Catal A 2018;563:204-15.
- 810 [27] Zhang HL, Hou ZY, Zhu Y, Wang JL, Chen YQ. Sulfur deactivation mechanism of
- 811 Pt/MnO_x -CeO₂ for soot oxidation: Surface property study. Appl Surf Sci 2017;396:560-5.
- 812 [28] Knauer M, Schuster ME, Su DS, Schlögl R, Niessner R, Ivleva NP. Soot structure and
- 813 reactivity analysis by Raman microspectroscopy, Temperature-Programmed Oxidation and
- High-resolution transmission electron microscopy. J Phys Chem A 2009;113:13871-80.
- 815 [29] Lapuerta M, Oliva F, Agudelo JR, Boehman AL. Effect of fuel on the soot nanostructure and
- consequences on loading and regeneration of diesel particulate filters. Combust Flame2012;159:844-53.
- 818 [30] Müller JO, Su DS, Jentoft RE, Kröhnert J, Jentoft FC, Schlögl R. Morphology-controlled

- reactivity of carbonaceous materials towards oxidation. Catal Today 2005;102-103:259-65.
- 820 [31] Abboud J, Schobing J, Legros G, Matynia A, Bonnety J, Tschamber V, et al. Impacts of
- 821 ester's carbon chain length and concentration on sooting propensities and soot oxidative reactivity:
- Application to Diesel and Biodiesel surrogates. Fuel 2018;222:588-98.
- 823 [32] Lamharess N, Starck L, Millet CN, Costa PD. Effect of biofuels on catalyzed diesel
- particulate filter regeneration. Top Catal 2013;56:462-6.
- [33] Hansen BB, Jensen AD, Jensen PA. Performance of diesel particulate filter catalysts in the
- presence of biodiesel ash species. Fuel 2013;106:234-40.
- [34] Al-Qurashi K, Boehman AL. Impact of exhaust gas recirculation (EGR) on the oxidative
 reactivity of diesel engine soot. Combust Flame 2008;155:675-95.
- 829 [35] Yehliu K, Armas O, Vander Wal RL, Boehman AL. Impact of engine operating modes and
- combustion phasing on the reactivity of diesel soot. Combust Flame 2013;160:682-91.
- [36] Hori S, Sato T, Narusawa K. Effects of diesel fuel composition on SOF and PAH exhaust
 emissions. JSAE Review 1997;18:255-61.
- [37] Kain J. Connecticut diesel emissions control technologies forum, Retrofit Overview. August
 17, 2005.
- [38] Sadezky A, Muckenhuber H, Grothe H, Niessner R, Pöschl U. Raman microspectroscopy of
- soot and related carbonaceous materials: spectral analysis and structural information. Carbon
 2005;43:1731-42.
- 838 [39] Setten B, Makkee M, Moulijn JA. Science and technology of catalytic diesel particulate
 839 filters. Catal. Rev. 2001;43:489-564.
- [40] Vander Wal RL, Tomasek AJ. Soot oxidation: dependence upon initial nanostructure.Combust Flame 2003;134:1-9.
- 842 [41] Alfe M, Apicella B, Barbella R, Rouzaud JN, Tregrossi A, Ciajolo A. Structure-property
- relationship in nanostructures of young and mature soot in premixed flames. P Combust Inst2009;32:697-704.
- [42] Ishiguro T, Suzuki N, Fujitani Y, Morimoto H. Microstructural changes of Diesel soot during
 oxidation. Combust Flame 1991;85:1-6.
- [43] Wang XC, Wang Y, Bai YQ, Wang P, Wang DX, Guo FN. Effects of 2,5-dimethylfuran
- 848 addition on morphology, nanostructure and oxidation reactivity of diesel exhaust particles. Fuel
- 849 2019;253:731-40.

- 850 [44] Schobing J, Tschamber V, Brillard A, Leyssens G, Iojoiu E, Lauga V. Impact of engine
- 851 operating cycle, biodiesel blends and fuel impurities on soot production and soot characteristics.
- 852 Combust Flame 2018;198:1-13.
- [45] Yehliu K, Vander Wal RL, Armas O, Boehman AL. Impact of fuel formulation on the
 nanostructure and reactivity of diesel soot. Combust Flame 2012;159:3597-606.
- [46] Barrientos EJ, Maricq MM, Boehman AL, Andersson JE. Impact of ester structures on the
- soot characteristics and soot oxidative reactivity of biodiesel. SAE Technical Paper 2015-01-1080(2015).
- [47] Boehman AL, Song J, Alam M. Impact of biodiesel blending on diesel soot and the
 regeneration of particulate filters. Energy Fuels 2005;19:1857-64.
- 860 [48] Zhang HL, Pereira O, Legros G, Iojoiu EE, Galvez ME, Chen YQ, et al. Structure-reactivity
- study of model and Biodiesel soot in model DPF regeneration conditions. Fuel 2019;239:373-86.
- 862 [49] Seong HJ, Boehman AL. Studies of soot oxidative reactivity using a diffusion flame burner.
- 863 Combust Flame 2012;159:1864-75.
- [50] Song J, Alam M, Boehman AL. Impact of alternative fuels on soot properties and DPF
 regeneration. Combust Sci Technol 2007;179:1991-2037.
- [51] Verma P, Pickering E, Jafari M, Guo Y, Stevanovic S, Fernando JFS, et al. Influence of
 fuel-oxygen content on morphology and nanostructure of soot particles. Combust Flame
 2019;205:206-19.
- [52] Corro G, Flores A, Pacheco-Aguirre F, Pal U, Banuelos F, Ramirez A, et al.. Biodiesel and
- fossil-fuel diesel soot oxidation activities of Ag/CeO₂ catalyst. Fuel 2019;250:17-26.
- [53] Wei Y, Zhao Z, Jin B, Yu X, Jiao J, Li K, et al.. Synthesis of AuPt alloy nanoparticles
- supported on 3D ordered macroporous oxide with enhanced catalytic performance for soot
 combustion. Catal Today 2015;251:103-13.
- [54] Dhal GC, Dey S, Mohan D, Prasad R. Study of Fe, Co, and Mn-based perovskite-type
- 875 catalysts for the simultaneous control of soot and NO_x from diesel engine exhaust. Mater
- 876 Discovery 2017;10:37-42.
- 877 [55] Mishra A, Prasad R. Preparation and application of perovskite catalysts for diesel soot
- emissions control: an overview. Catal Rev 2014;56:57-81.
- 879 [56] Bueno-López A, Krishna K, Makkee M, Moulijn JA. Enhanced soot oxidation by lattice
- 880 oxygen via La^{3+} -doped CeO₂. J Catal 2005;230:237-48.

- [57] Tikhomirov K, Kröcher O, Elsener M, Wokaun A. MnO_x-CeO₂ mixed oxides for the
 low-temperature oxidation of diesel soot. Appl Catal B 2006;64:72-78.
- [58] Zhang HL, Yuan SD, Wang JL, Gong MC, Chen YQ. Effects of contact model and NO_x on
- soot oxidation activity over Pt/MnO_x -CeO₂ and the reaction mechanisms. Chem Eng J 2017;327:1066-76.
- [59] Bassou B, Guilhaume N, Iojoiu EE, Farrusseng D, Lombaert K, Bianchi D, Mirodatos C.
- 887 High-throughput approach to the catalytic combustion of diesel soot II: Screening of oxide-based
- 888 catalysts. Catal Today 2011; 159: 138-143.
- [60] Zhu L, Yu J, Wang X. Oxidation treatment of diesel soot particulate on Ce_xZr_{1-x}O₂. J Hazard
 Mater 2007;140:205-10.
- [61] Yang ZZ, Hu W, Zhang N, Li Y, Liao YW. Facile synthesis of ceria-zirconia solid solutions
- 892 with cubic-tetragonal interfaces and their enhanced catalytic performance in diesel soot oxidation.
- 893 J Catal 2019; in press, DOI: 10.1016/j.jcat.2019.06.029.
- [62] Aneggi E, Leitenburg C, Trovarelli A. On the role of lattice/surface oxygen in ceria-zirconia
- catalysts for diesel soot combustion. Catal Today 2012;181:108-15.
- [63] Atribak I, Bueno-López A, García-García A. Role of yttrium loading in the physico-chemical
- properties and soot combustion activity of ceria and ceria-zirconia catalysts. J Mol Catal A: Chem
 2009;300:103-10.
- [64] Zhang HL, Wang JL, Zhang YH, Jiao Y, Ren CJ, Gong MC, et al. A study on H₂-TPR of
- 900 $Pt/Ce_{0.27}Zr_{0.73}O_2$ and $Pt/Ce_{0.27}Zr_{0.70}La_{0.03}O_x$ for soot oxidation. Appl Surf Sci 2016;377:48-55.
- 901 [65] Oi-Uchisawa J, Obuchi A, Ogata A, Enomoto R, Kushiyama S. Effect of feed gas
 902 composition on the rate of carbon oxidation with Pt/SiO₂ and the oxidation mechanism. Appl
 903 Catal B 1999;21:9-17.
- 904 [66] Yamazaki K, Kayama T, Dong F, Shinjoh H. A mechanism study on soot oxidation over
- 905 CeO₂-Ag catalyst with 'rice-ball' morphology. J Catal 2011;282:289-98.
- 906 [67] Liu S, Wu XD, Liu W, Chen WM, Ran R, Li M, et al. Soot oxidation over CeO₂ and Ag/CeO₂:
- 907 Factors determining the catalyst activity and stability during reaction. J Catal 2016;337:188-98.
- 908 [68] Lin F, Wu XD, Liu S, Weng D, Huang YY. Preparation of MnO_x-CeO_x-Al₂O₃ mixed oxides
- 909 for NO_x -assisted soot oxidation: Activity, structure and thermal stability, Chem Eng J 910 2013;226:105-12.
- 911 [69] Zhang HL, Zhou CX, Galvez ME, Costa PD, Chen YQ. MnO_x-CeO₂ mixed oxides as the

- 912 catalyst for NO-assisted soot oxidation: The key role of NO adsorption/desorption on catalytic
- **913** activity, Appl Surf Sci 2018;462:678-84.
- 914 [70] Zhang HL, Li SS, Lin QJ, Feng X, Chen YQ, Wang JL. Study on hydrothermal deactivation
- 915 of Pt/MnO_x-CeO₂ for NO_x-assisted soot oxidation: redox property, surface nitrates, and oxygen
- 916 vacancies, Environ Sci Pollut Res 2018;25:16061-70.
- 917 [71] Pajdowski P, Puchalka B. The process of diesel particulate filter regeneration under real
- 918 driving conditions. IOP Conf Ser: Earth Environ Sci 2019;214:012114.
- 919 [72] Seong H, Choi S. Oxidation-derived maturing process of soot, dependent on O₂-NO₂
- mixtures and temperatures. Carbon 2015;93:1068-76.
- 921 [73] Abboud J, Schobing J, Legros G, Bonnety J, Tschamber V, Brillard A, et al. Impacts of
- 922 oxygenated compounds concentration on sooting propensities and soot oxidative reactivity:
- 923 Application to Diesel and Biodiesel surrogates. Fuel 2017;193:241-53.
- 924 [74] Wang Y, Alsmeyer DC, McCreery RL. Raman spectroscopy of carbon materials: structural
 925 basis of observed spectra. Chem Mater 1990;2:557-63.
- 926 [75] Cuesta A, Dhamelincourt P, Laureyns J, Martinez-Alonso A, Tascon JMD. Raman
- microprobe studies on carbon materials. Carbon 1994;32:1523-32.
- 928 [76] Sze SK, Siddique N, Sloan JJ, Escribano R. Raman spectroscopic characterisation of
 929 carbonaceous aerosol. Atmos Environ 2001;35;561-8.
- 930 [77] Jawhari T, Roid A, Casado J. Raman spectroscopic characterisation of some commercially
- available carbon black materials. Carbon 1995;33:1561-5.
- 932 [78] Dippel B, Jander H, Heintzenberg J. NIR FT Raman spectroscopic study of flame soot. Phys
- 933 Chem Phys 1999;1:4707-12.
- 934 [79] Nakamura K, Fujitsuka M, Kitajima M. Disorder-induced line broadening in first-order
- Raman scattering from graphite. Phys Rev B 1990;41:12260-3.
- 936 [80] Lu T, Cheung CS, Huang Z. Size-resolved volatility, morphology, nanostructure, and
- 937 oxidation characteristics of diesel particulate. Energy Fuels 2012;26:6168-76.
- 938 [81] Bladt H, Ivleva NP, Niessner R. Internally mixed multicomponent soot: impact of different
- salts on soot structure and thermo-chemical properties. J Aerosol Sci 2014;70:26-35.
- 940 [82] Schmid J, Grob B, Niessner R, Ivleva NP. Multiwavelength Raman microspectroscopy for
- rapid prediction of soot oxidation reactivity. Anal Chem 2011;83:1173-9.
- 942 [83] Ishiguro T, Takatori Y, Akihama K. Microstructure of diesel soot particles probed by electron

- 943 microscopy: first observation of inner core and outer shell. Combust Flame 1997;108:231-4.
- [84] Valero-Romero MJ, García-Mateos FJ, Rodríguez-Mirasol J, Cordero T. Role of surface
 phosphorus complexes on the oxidation of porous carbons. Fuel Process Technol
 2017;157:116-26.
- 947 [85] Schobing J, Tschamber V, Brillard A, Leyssens G. Impact of biodiesel impurities on carbon
- 948 oxidation in passive regeneration conditions: Influence of the alkali metals. Appl Catal B949 2018;226:596-607.
- [86] Wu XD, Lin F, Xu HB, Weng D. Effects of adsorbed and gaseous NO_x species on catalytic
 oxidation of diesel soot with MnO_x-CeO₂ mixed oxides. Appl Catal B 2010;96:101-9.
- 952 [87] Escribano VS, López EF, Gallardo-Amores JM, Del C, Martínez H, Pistarino C, et al. A study
- 953 of a ceria-zirconia-supported manganese oxide catalyst for combustion of Diesel soot particles.

954 Combust Flame 2008;153:97-104.

- [88] Figueiredo JL, Pereira MFR, Freitas MMA, Órfão JJM. Modification of the surface chemistry
 of activated carbons. Carbon 1999;37:1379-89.
- 957 [89] Setiabudi A, Makkee M, Moulijn JA. The role of NO₂ and O₂ in the accelerated combustion
 958 of soot in diesel exhaust gases. Appl Catal B 2004;50:185-94.
- [90] Knauer M, Carrara M, Rothe D, Niessner R, Ivleva NP. Changes in structure and reactivity of
 soot during oxidation and gasification by oxygen, studied by microRaman spectroscopy and
 temperature programmed oxidation. Aerosol Sci Technol 2009;43:1-8.
- 962 [91] Tapia A, Salgado S, Martin P, Villanueva F, García-Contreras R, Cabanas B. Chemical
- 963 composition and heterogeneous reactivity of soot generated in the combustion of diesel and GTL
- 964 (Gas-to-Liquid) fuels and amorphous carbon Printex U with NO₂ and CF₃COOH gases. Atmos
 965 Environ 2018;177:214-21.
- 966 [92] Ye P, Sun CX, Lapuerta M, Agudelo J, Vander Wal R, Boehman AL, Toops TJ, Daw S.
 967 Impact of rail pressure and biodiesel fueling on the particulate morphology and soot
 968 nanostructures from a common-rail turbocharged direct injection diesel engine. Int J Engine Res
 969 2016;17:193–208.
- 970 [93] Bladt H, Schmid J, Kireeva ED, Popovicheva OB, Perseantseva NM, Timofeev MA, et al.
- 971 Impact of Fe content in laboratory-produced soot aerosol on its composition, structure, and
- 972 thermo-chemical properties. Aerosol Sci Technol 2012;46:1337-48.

