

# An in vitro model to evaluate the impact of environmental fine particles (PM0.3-2.5) on skin damage

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- 1 An *in vitro* model to evaluate the impact of environmental fine particles (PM<sub>0.3-2.5</sub>) on
- 2 skin damages
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#### 30 Abstract

31 Exposure to airborne particulate matter (PM) has significant effects on human health mainly leading to cardio-respiratory diseases. However very few data are available regarding the 32 impact of PM on the skin, so to better understand the impact of fine particle  $(PM_{0,3-2,5})$  on 33 both inflammatory response and epidermal structure, we exposed a reconstructed human 34 35 epidermis (RHE) to several doses of PM collected in Cotonou (Benin, West Africa). After 36 24h of exposure, inflammatory response, histological observations, and gene expression 37 related to oxidative stress, antioxidant defense and structural damages were determined. No PM-linked changes in tissue morphology or membrane integrity were observable. PM was 38 however cytotoxic in a dose dependent manner. An inflammatory response appeared as 39 40 shown by the increase in IL-1 $\alpha$  and IL-8 cytokine productions. PM also induced oxidative 41 stress, leading to an increase in 4-HNE immunostaining and to the up-regulation of HMOX1, 42 MT1G and MT1E. Finally, PM had a negative impact on fundamental skin functions such as 43 tissue anchorage, cell differentiation, cornification / skin desquamation and apoptosis. Our 44 data show that airborne fine particles have an adverse effect on skin integrity, most probably leading to accelerated ageing. 45

#### 46 Keywords

47 Airborne fine particles; *In vitro* skin model; Inflammatory response; Oxidative stress;
48 Premature skin ageing

#### 49 Abbreviations

4-HNE: 4-hydroxynonenal, BSA: bovine serum albumin, DAPI: 4',6-diamidino-2phenylindole dihydrochloride, PM: particulate matter, H&E: haematoxylin and eosin, MMP:
metalloproteinase, PBS: phosphate buffered saline, RHE: Reconstructed Human Epidermis,
RT-qPCR: real-time quantitative PCR

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#### 60 1. Introduction

61 Air pollution is a degradation of air quality by introduction of harmful substances 62 including gases, chemicals, biological contaminants, and particles. Air pollutants include particulate matter (PM), and are a major public health concern. PM is a heterogeneous and 63 64 complex mixture due to its various origins (natural: soil, forest fires or volcanoes, and anthropogenic: human activities, industry, traffic...). Emitting sources and atmospheric 65 conditions are key parameters contributing to the chemical and physical properties of ambient 66 particulate matter (Kim et al., 2016). The role of the aerodynamic diameter of PM in the 67 increase in the risks of chronic illness and premature death is widely documented. Much less 68 69 recognized however, is that the chemical constituents adsorbed on PM also have significant 70 effects on various diseases (Dergham et al., 2015; Kim et al., 2016; Song et al., 2013). A large number of studies have considered the effects of fine particles (PM<sub>2.5</sub>) on human health, the 71 72 majority being focused on respiratory and cardiovascular diseases. In the case of airway epithelium, for instance, the skin is in direct contact with airborne PM, and long-term 73 74 exposure to PM may induce or exacerbate pre-existing skin diseases such atopic dermatitis,

acne, psoriasis, or premature skin ageing (Mancebo and Wang, 2015; Drakaki et al., 2014; 75 76 Flohr and Mann, 2014; Kim et al., 2016). Different clinical studies have highlighted an increased level of rash after exposure to polluted air, a relation between traffic-associated 77 78 pollution and eczema in children, a modification of the skin integrity, and a contribution of traffic-based air pollution to facial lentigines (Huls et al., 2016; Krutmann et al., 2014; 79 80 Lefebvre et al., 2015; Vierkotter et al., 2010). At the cellular level, skin ageing modifies the 81 tissue structure with looseness of the epidermal-dermal junction, and of the dermis with the reduction of blood vessels and nerves (Puizina-Ivic, 2008). 82

Oxidative stress seems to be one of the underlying mechanisms that induces adverse 83 84 health effects and it is known to play a role in skin damage after exposure to some environmental factors like UV and pollutants (Carletto et al., 2000; Giacomoni and Rein, 85 2001; Krutmann et al., 2016; Lanuti and Kirsner, 2010; Makrantonaki et al., 2015; Valacchi et 86 87 al., 2012). A direct impact of the UV exposure has been shown on vitamin D production, loss of elastic fiber properties, and oxidative stress (Gilchrest, 2013; Kammeyer and Luiten, 2015; 88 89 Sherratt, 2013). Airborne pollutants such as nitrogen oxides, hydrocarbons, fine particles and 90 ozone, also have a deleterious impact on skin, particularly sensitized skins which react more strongly to stressors (Krutmann et al., 2014). Similarly, some studies have established a link 91 92 between tobacco smoke and skin alteration (Lecas et al., 2016; Morita et al., 2009; Urbanska et al., 2012a; Urbanska et al., 2012b). 93

Generation of reactive oxygen species, upregulation of pro-inflammatory factors by the
arylhydrocarbon receptor (AhR), and keratinocytic VEGF up-regulation have been
demonstrated as responsible mechanisms in the genesis of certain cutaneous pathologies (Bae
et al., 2015; Krutmann et al., 2014; Li et al., 2003; Vierkotter and Krutmann, 2012). In order
to better explain the underlying mechanisms, studies have been mainly performed on animals.
A 6h exposure to controlled PM<sub>2.5</sub> was shown to induce oxidative stress (*HMOX1* expression)

in neonatal and adult rat lungs and the stimulation of antioxidant enzymes (*SOD1* for
example) was observed for adults but not neonates, more susceptible to ultrafine particles
(Chan et al., 2013). Cytotoxicity and apoptosis of human keratinocytes were also associated
with combustion-generated organic nanoparticles (Pedata et al., 2012).

104 Mastrofrancesco et al. (2014) also reported that environmental pollution from diesel exhaust plays a critical role in the development of some autoimmune diseases 105 106 (Mastrofrancesco et al., 2014). The authors showed that rat scleroderma skin keratinocytes 107 and fibroblasts expressed cytokine genes (*IL-1a*, *IL-6*, *IL-8*, and *TNF-β*) at a higher level than control cells when treated with carbon-based nanoparticles collected from diesel exhaust. In 108 109 addition, mRNA gene expressions of some MMPs, Collagen types I and III, and VEGF were 110 significantly higher in untreated scleroderma skin cells than controls, with a dose dependent 111 stimulation of all MMPs (Mastrofrancesco et al., 2014).

Taking into account that oxidative stress and inflammation play a major role in air pollution mediated toxicity of the skin, the aim of the present work was to evaluate the impact of environmental fine particles ( $PM_{0.3-2.5}$ ), collected in a geographical area selected for its natural wealth in metallic elements, on skin tissue. For this purpose, we used a sensitive *in vitro* skin model (RHE) focusing on possible oxidative damage leading to morphological changes and modulation of protein expression.

118 2. Material and methods

#### 119 **2.1.** Chemicals

120 Culture media, maintenance medium for experiment days, and growth medium for non-121 experiment days, were purchased from SkinEthic<sup>TM</sup>, France. Penicillin-Streptomycin mixture 122 (penicillin: 100  $\mu$ g/mL; streptomycin: 100UI/mL) and Fungizon (Antibiotic-Antimycotic) 123 were from Gibco (Invitrogen, WI, USA), Phosphate-buffered saline (PBS) from Biosolve 124 Chimie (France), calcium chloride, goat serum, glycine, bovine serum albumin, glycerol, sodium azide, Eukitt<sup>®</sup>, sodium cacodylate, glutaraldehyde, hexamethyldisilazane, osmium
tetroxide, hydrochloric acid, and propylene oxide from Sigma-Aldrich (France). Acetic acid
and Triton X100 were purchased from Merck (Germany); paraffin from Labonord (France);
ethanol from VWR International (PA, USA). Xylene and paraformaldehyde (32%) were
provided by Euromedex (France); sodium chloride by Minisol (France); Mayer Haematoxylin
(RAL Diagnostics, France) and Eosin by Merck (Germany).

#### 131 2.2. In vitro skin model

A reconstructed human epidermis from normal human keratinocytes provided the RHE
model (SkinEthic<sup>™</sup>, France) (Rosdy and Clauss, 1990; Rosdy et al., 1993) that was used as
described previously (Lecas et al., 2016). For all experiments, the same batch (MF22-5) was
used. Signed informed consent and ethical approval were obtained by the supplier.

Upon reception RHEs were placed in an incubator at 37 °C, 5% CO<sub>2</sub> and 90% relative
humidity. Experiments were carried out 24h after RHE-acclimation (Poumay et al., 2004).

#### 138 2.3. Fine particles exposure ( $PM_{0.3-2.5}$ )

139 Detailed information about collection and physicochemical characteristics of the PM 140 collected in Cotonou city (Benin, West Africa) have been previously published (Table 1, Table S1) (Cachon et al., 2014). After collection, PM<sub>0.3-2.5</sub> were weighed and stored at -20°C 141 142 until experimentation. PM were suspended in the culture medium containing 1% fungizon and 143 sonicated at 50W during 10 min (Bath sonicator Branson CPXH, Dutscher, France) to avoid the presence of aggregates (Fig. S1). RHEs were exposed to three PM concentrations: 3, 15 144 and  $30 \,\mu\text{g/cm}^2$  as published by Cachon (Cachon et al. 2014). 30  $\mu\text{L}$  of suspension of the 145 146 desired PM concentration were deposited directly on the apical surface of RHE tissues. RHEs 147 were then placed on an orbital agitator for 10 mins to homogenize the PM suspension to the 148 apical surface and then incubated at 37°C, 5% CO<sub>2</sub> in a humidified incubator.

After 24h exposure to PM, culture media were collected to assess biological activity: LDH release, cytokine and MMP quantifications. RHEs were harvested to perform MTT assay, histological / immunohistochemical analysis and mRNA quantification by RT-qPCR.

Two controls were used: medium-control (named Control:  $30 \ \mu L$  of medium without PM deposited on apical surface) and incubator-control (without medium on apical side). As no difference was shown between these two controls, only the medium-control (Control) was introduced in the continuation of the present work.

#### 156 2.4. Functional integrity

157 Colorimetric MTT [3-(4,5-dimethylthiazol-2-yl)-2,5-diphenyltetrazolium bromide] assay 158 relies on the quantification of mitochondrial succinate dehydrogenase activity. Viable cells 159 convert a soluble tetrazolium salt to a formazan product having a maximum photometric 160 absorbance at a wavelength of 490 nm (Persoz et al., 2010). Absorbance was read with a 161 microtiter plate reader (Multiskan-EX, Thermo Scientific). Experiments were repeated at least 162 4 times in duplicate.

#### 163 2.5. Cellular viability: LDH release

Membrane integrity was determined by measuring the release of intracellular lactate dehydrogenase (LDH) into the basal culture medium (Lactacte deshydrogenase activity assay Kit, Sigma). LDH reduced NAD to NADH, which is detected by colorimetric at 450nm using a microtiter plate reader (Multiskan-EX, Thermo Scientific). LDH release is the reflection of the cellular viability and was expressed as an arbitrary unit. Experiments were repeated at least 4 times in duplicate.

#### 170 2.6. Proinflammatory mediator quantification

171 Two cytokines, interleukin 1 $\alpha$  (IL-1 $\alpha$  #KAC1191) and interleukin 8 (IL-8 #KHC0081), 172 were assessed using an ELISA assay following manufacturer's instructions (Novex<sup>TM</sup>, 173 France). Optical densities were determined using a microplate-reader (Multiskan-EX, Thermo

174 Scientific, USA) set to 450 nm. Calibration curves were established using Ascent Software<sup>®</sup> 175 (Thermo Scientific, MA, USA). Cytokine concentrations were expressed in pg/mL. The limits 176 of quantification were 3.9 and 15.6 pg/mL for IL-1 $\alpha$  and IL-8 respectively. Experiments were 177 repeated 4 times in duplicate.

#### 178 2.7. Matrix metalloproteinases assessment

179 Assessment of matrix metalloproteinases was performed using a multiplex panel (Human Magnetic Luminex Performance Assay Base Kit, MMP Panel [#LMPM000] R&D Systems, 180 181 France). Samples were incubated overnight with antibody-coated beads at 4°C under constant shaking. The following day, plates were washed twice and secondary antibody was added to 182 183 the samples for one-hour incubation at room temperature under constant shaking. Then, plates 184 were incubated for 30 minutes with streptavidin labelled with R-Phycoerythrin at room temperature under agitation. Washing process was repeated, before adding sheath fluid and 185 reading the results using Luminex<sup>®</sup> technology (Magpix<sup>®</sup>, Bio-Rad, France) after 5 minutes 186 187 shaking. Concentrations were calculated out of standard curves using recombinant proteins 188 and expressed in pg/mL. Experiments were repeated 2 times in triplicate.

#### 189 2.8. Histological and immunochemical assessments

Histological and immunohistochemical assessments were performed 24h after exposure.
RHEs were fixed in 4% paraformaldehyde for 4h and paraffin-embedded sections were sliced
for Mayer Haematoxylin and Eosin (H&E) and for immunohistochemistry.

193 Prior to Mayer H&E staining, RHE cross-sections (4  $\mu$ m) went through xylene 194 deparaffinization and rehydration with graded ethanol baths. Observations were made with an 195 inverted light microscope (Leica DM4000 B LED).

For immunohistochemistry, sections were blocked in 10% BSA/PBS before incubation
with anti-loricrin or anti-4-HNE primary antibodies (#ab24722 and #ab46545 respectively,
Abcam, UK). For both, an AlexaFluor®488-conjugated goat anti-mouse (#A-21121,

Invitrogen, USA) was used as a secondary antibody. Nuclei were stained with 4',6-diamidino-200 2-phenylindole dihydrochloride (DAPI, #32670 Sigma, France). The resulting 201 immunofluorescence was observed under a Nikon Eclipse 50i fluorescence microscope and 202 pictures were taken using Nikon NIS Elements software. Quantification of both loricrin and 203 4-HNE fluorescence signals of tissue sections was performed using Visilog software version 204 7.0. Experiments were repeated in 3 independent experiments.

#### 205 2.9. RNA isolation and RT-qPCR

206 Following the manufacturer's instructions, total RNA was extracted from tissues using the RNeasy<sup>®</sup> miniKit (Qiagen, Germany). RNA quality and quantity were checked with the 207 208 2100-Bioanalyzer (Agilent Technologies, France), then RNA was reverse transcribed using 209 the Transcriptor Universal cDNA Master (Roche Diagnostics, Switzerland). PCR primers were designed according to the Roche Universal Probe Library Assay Design Center 210 211 (www.universalprobelibrary.com). A LightCycler 480 system using LightCycler 480 Probes 212 Master (Roche, Germany) was used to perform quantitative RT-qPCR on 29-gene epidermal 213 function plates (Agilent RNA Nano chips #5067-1511, Roche Diagnostics, Switzerland). The  $2^{-\Delta\Delta Ct}$  method provided N-fold differential expression of mRNA gene expression and results 214 were obtained by calculating the fold change (FC) between two conditions, exposed versus 215 control, meaning the expression difference. mRNA expression was considered as increased 216 217 when  $FC \ge 1.5$ , while decreased when  $FC \le 0.7$ . GAPDH and RPL13 housekeeping genes 218 were used as internal controls. Results are the mean of one experiment assessed in duplicate.

219 2.10. Statistical analysis

Data are displayed as mean  $\pm$  standard deviation. All experiments were repeated at least in duplicate in two different experiments; the number of experiments conducted is indicated in the legend of each Figure. Results were analyzed by one-way ANOVA followed by a Tukey-Kramer post-test. Statistical analysis was performed using JMP v.10 (SAS division,
USA). p < 0.05 was considered significant.</li>

225 **3. Results** 

#### 226 3.1. Tissue morphology and cell functionality (Fig. 1-2)

227 The SkinEthic reconstituted human epidermis (RHE) model consists of the four 228 characteristic layers *stratum germinativum*, *stratum spinosum*, *stratum granulosum*, and 229 *stratum corneum* as shown in the Figure 1A.

A 24h exposure of RHEs to increasing PM concentrations (3, 15 and 30  $\mu$ g/cm<sup>2</sup>) did not 230 231 induce structural changes whatever the PM-level tested (Figs. 1B1, 1B2, 1B3): RHEs exposed 232 to PM were similar to the Control. This observation is confirmed by the absence of LDH 233 release certifying good membrane integrity (Fig. 2B). Nevertheless, MTT assay revealed a dose-dependent relation between the level of PM and cellular functionality, with a statistically 234 significant decrease after exposure of RHEs at 15 and 30  $\mu$ g/cm<sup>2</sup> compared to Control (Fig. 235 236 2A). This last result would seem to indicate that the mitochondrial succinate dehydrogenase 237 function is disturbed without leading to cell death as indicated by LDH and morphology data.

#### 238 3.2. Inflammatory response (Fig. 3)

To evaluate the inflammatory response of RHEs, the production of two cytokines, IL-1 $\alpha$ and IL-8, was evaluated after RHE-exposure to PM at 3, 15 and 30 µg/cm<sup>2</sup>. Compared to Control, a 24h-exposure to PM at 3 µg/cm<sup>2</sup> did not induce significant difference in the production levels of IL-1 $\alpha$  and IL-8 (Fig. 3). For higher doses of PM (15 and 30 µg/cm<sup>2</sup>), a significant dose-dependent increase in the production of the two cytokines tested was observed. For IL-8, the increase was as high as 52-fold the level detected in the Control after exposure of RHEs to 30 µg/cm<sup>2</sup> of PM.

*Remark:* These initial results concerning membrane integrity, functionality and inflammatory
responses, revealed significant cytokine production after RHE exposure at the highest dose of

PM  $(30 \ \mu g/cm^2)$  without membrane integrity disorder (LDH release) but with a loss of 50% of cellular functionality, probably due to the oxidative stress caused by metallic compounds adsorbed on the PM tested. In this context, the rest of the study has been focused on the exposure of RHE tissue to PM at 30  $\mu g/cm^2$ .

#### 252 3.3. Oxidative stress (Fig. 4)

Two steps have been taken to assess oxidative damage on RHE tissue: 4-HNE immunostaining with quantification (4-hydroxynonenal, one of the most important products of lipid peroxidation) and three oxidant gene expressions evaluated by RT-qPCR.

Immunostaining of 4-HNE, a widely used marker of oxidative stress, showed a substantial 256 increase in labelling, 24h after exposure of RHE tissue to PM at 30 µg/cm<sup>2</sup>, compared to 257 258 Control (Fig. 4A). The quantification of 4-HNE staining confirmed this significant increase by 4-fold, demonstrating that PM exposure does stimulate oxidative stress in our model (Fig. 259 4B). To investigate the underlying effects on the molecular and biological activities of RHE, 260 261 the activity of three genes involved in oxidative stress (HMOX1, MT1G and MT1E) was 262 studied by quantification of their mRNA. mRNA expression levels of the genes were stimulated by 1.6 to 3.0-fold respectively after exposure of RHEs to PM at 30  $\mu$ g/cm<sup>2</sup> (Fig. 263 4C). 264

#### 265 3.4. Differentiation and morphology (Fig. 5-6)

To evaluate the impact of PM on both extracellular matrix and tissue differentiation, two approaches were considered: MMP level quantification and loricrin immunostaining with quantification.

Immuno-quantification of matrix metalloproteinases panel showed that only two MMPs from the nine MMPs tested (MMP-1, -2, -3, -7, -8, -9, -10, -12, -13) were significantly modulated. Compared to Control, MMP-1 was induced by 1.5-fold and MMP-3 by 1.9-fold after exposure of RHEs to 30  $\mu$ g/cm<sup>2</sup> of PM (Fig. 5). In addition, immunostaining of loricrin,

a major protein component of terminally differentiated epidermal cells, showed a significant decrease in labelling after PM-exposure (Fig. 6A). This result is confirmed by the quantification of the intensity of loricrin labelling. Compared to the Control, a 1.9-fold loss of loricrin staining was observed 24h after PM-exposure (Fig. 6B).

#### 277 3.5. Other biological impacts of PM on RHEs (Fig. 7)

To investigate the underlying effects of PM at 30  $\mu$ g/cm<sup>2</sup> on the biological activity of 278 RHEs, 12 genes were selected as the most representative to evaluate cellular and molecular 279 280 modulation. Differentiation genes were either weakly stimulated (IVL) or weakly decreased (KRT1, KRT10) (Fig. 7). Expressions of cell-cell junction genes (EVPL (ENVOPLAKIN), 281 282 GJA1, CDH1, CD44), and anchorage genes (COL7A1, COL4A1) were suppressed after PM-283 exposure. Apoptotic gene expression was also studied (CASP3, CDKN2A, BIRC5), and results shown that upon PM exposure, BIRC5 expression was reduced while CDKN2A expression 284 285 was doubled and CASP3 was highly stimulated (6.3-fold) (Fig. 7).

#### 286 **4.** Discussion

The aim of our study was to better understand the impact of airborne fine particles (PM) on skin damage as suggested by other studies (Valacchi et al., 2012; Vierkotter and Krutmann, 2012).

290 In this study we used particles collected in Cotonou, an urban city of West Africa 291 (Benin). Our objective was to generally test environmental particles rich in such metallic 292 inorganic elements as metals, positively correlated with intracellular ROS production 293 (Dergham et al., 2015), and not to specifically test Cotonou's PM. Having said this, the 294 interest of particle samples tested in our study does lie in their origin: in African countries like 295 Cotonou the soils are rich in iron oxide ( $Fe_2O_3$ ) and aluminium oxide ( $Al_2O_3$ ) (Kikouama et al., 2009), with co-existing heavy metals at very high levels (Alford et al., 2010; Cheng, 296 297 2003). In comparison we can note that the level of metallic elements in PM collected in an urban city of France (Dunkerque) was 620 ng/m<sup>3</sup> versus 8,644 ng/m<sup>3</sup> in Cotonou (Dergham et
al., 2015).

The presence of such elements contributes to atmospheric pollution leading to various 300 301 biological dysregulations in Human, such as oxidative stress, this latter being at the origin of various pathologies (Cachon et al., 2014). The choice of the in vitro skin model was important 302 303 to our study. RHE tissue consists of an epidermis without dermis, obtained from keratinocytes 304 cultured at high cell density on a polycarbonate filter membrane with 0.4 µm pore diameter. 305 After 15 days of culture, keratinocytes formed a differentiated epidermis with histological features similar to those observed in vivo: a basal layer, spinous layer, granular layer, and a 306 307 stratum corneum (Poumay et al, 2004). This model was used to assess the irritating and/or corrosive properties of some chemical compounds (Kandárová et al, 2006). However, like all 308 309 in vitro models, RHE tissue has some limitations, principally the small experimentation 310 window (3 days) due to the model ageing quickly (Lecas et al, 2016). This led us to limit our 311 experiments to 24h of exposure. Nevertheless, the RHE model did allow the combined 312 implementation of 3D-reconstituted tissue and exposure at the air-liquid interface in a 313 controlled environment. Our work has thus highlighted the possibility of using this in vitro 3D-skin model to assess the importance of the chemical constituents of PM on the cellular 314 315 and molecular responses of exposed tissue.

Exposure to airborne particulate matter has toxic effects on human health as illustrated by drastic respiratory and skin diseases (Beelen et al., 2014; Drakaki et al., 2014; Raaschou-Nielsen et al., 2013). The skin can be impacted by several intrinsic and extrinsic phenomena (Makrantonaki et al., 2015). External factors such as UV and air pollutants including tobacco smoke can play an important role in skin ageing (Valacchi et al., 2012; Vierkotter and Krutmann, 2012). In order to take into account all levels (tissue, cell and molecular level) and have a better knowledge on the underlying mechanisms, an *in vitro* RHE skin model can be used to evaluate the toxicity level of extrinsic exposure such as tobacco smoke (Lecas et al.,
2016) and airborne pollutants (Magnani et al., 2016). The mode of action (MoA) hypothesis
of the impact of airborne pollution on the premature skin ageing process, using the RHE 3Dskin model exposed to fine particles, is illustrated in Figure 8.

327 While the PM did not induce obvious structural changes in RHE morphology and a loss 328 of membrane integrity, it did lead to a dose-dependent decrease in cellular functionality as shown by MTT assay, evaluating mitochondrial succinate dehydrogenase activity, a 329 330 mitochondrial enzyme. These results - showing a loss of cellular functionality with absence of both structural changes and tissue integrity injury - would appear contradictory, but is not. 331 332 Specifically, the complementary nature of the tests gives a more accurate view of the impact of a xenobiotic on a targeted tissue. The oxidizing potential of PM, due to the presence of 333 metal elements adsorbed on their surface, could explain this result and constitute the key 334 335 event 1 (KE1, Fig. 8).

336 To consolidate this result, we observed that PM was able to stimulate IL-1a and IL-8 337 production in a dose dependent manner, thus illustrating the inflammatory response of skin 338 tissue (AE1, Fig. 8). Oxidative stress was also induced, as illustrated by the increase of 4-HNE labelling in RHE (KE2, Fig. 8), and in accordance with the stimulation of HMOX1 339 expression. A similar induction of HMOX1, a sensitive marker for oxidative stress, was 340 341 observed upon exposure of a lung model to ultrafine particles (Chan et al., 2013; Li et al., 342 2003). In reaction, the tissue induces antioxidants as illustrated by previous works on the lung 343 (Curjuric et al., 2012; Huang et al., 2009) and, in our case, by the induction of MT1G and 344 MT1E expressions which corresponding proteins are known to be involved in heavy metal 345 binding (KE3, Fig. 8).

346 Considering tissue integrity, no morphological changes appeared after PM-exposure, and 347 increased levels of MMP-1 and MMP-3 production suggest an accelerated degradation of the

extracellular matrix (KE5, Fig. 8). The level of loricrin, a keratinocyte specific protein, was 348 349 also decreased (AE2, Fig. 8). Moreover, cell-cell junction, anchorage, and terminal 350 differentiation genes, except Involucrin (IVL), were drastically down-regulated on PM-351 exposure (AE3, Fig. 8). Involucrin (IVL) is known to play an important role in cornified 352 envelope development (Hanel et al., 2013). Inhibition of IVL was observed in several studies, 353 including when the *in vitro* RHE skin model was exposed to tobacco smoke, illustrating the 354 disorganization of skin barrier function (Lecas et al., 2016). However, stimulation of IVL was 355 also found after exposure of HaCaT cells and atopic dermatitis to ambient PM<sub>2.5</sub> leading to the hypothesis that cell death is the primary cause of skin alteration rather than protein 356 modification (Hirao et al., 2003; Li et al., 2017). KRT1 and KRT10 expressions were 357 358 decreased, revealing the skin barrier dysfunction (Pan et al., 2015). This tissue slackening was 359 also illustrated by the decrease in the expression of cell-cell junction (EVPL, GJA1, CDH1 360 and CD44) and anchorage (COL4A1 and COL7A1) genes. This is in line with the role of Gap 361 Junctions Intercellular Communications and Tight Junctions in the epithelial lining, which 362 modification might be implied in cancer development after PM<sub>10</sub> inhalation (Nemmar et al., 363 2013).

In order to better understand the underlying mechanism of toxicity on the RHE model, 364 apoptosis-linked genes (CASP3, CDKN2A and BIRC5 (SURVIVIN)) were also investigated. 365 366 BIRC5, known to block apoptosis and to regulate the cell cycle, was downregulated. 367 Moreover, both CASP3 and CDKN2A, pro-apoptotic genes, were upregulated in our study (6and 2-fold, respectively) (KE4, Fig. 8). Metal elements, being the main components of PM 368 369 from Cotonou, have been described as inductors of CASP3 in hepatic cells and macrophages 370 (Andreau et al., 2012). CDKN2A (p16) is known to have an ageing-dependent expression and 371 therefore to be increased in the elderly (Ressler et al., 2006).

In conclusion, using an *in vitro* 3D-model of human reconstituted epidermis, our results 372 373 showed that exposure to fine particles lead to an increase in some indicators such as oxidative 374 stress and inflammatory response, contributing to a tissue slackening following the 375 degradation of the extracellular matrix and cell-cell junctions. The chemical composition of 376 airborne particulate matter, and more particularly the metallic elements adsorbed on the 377 surface of the PM, could contribute to various MoA hypotheses leading to the premature skin 378 ageing and the loss of the protective role of the cutaneous barrier. The present work is original 379 and innovative, and provides new development opportunities to better understand the role of atmospheric particulate matter on skin damage. 380

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- 383 commercial, or not-for-profit sectors.

#### **384** Competing interests

R. Fitoussi, N. Blanchet and K. Vié are full-time employees of Clarins, a major company
specialized in the design, manufacturing and marketing of cosmetic products.

387

#### 388 **References**

- Alford, É. R., Pilon-Smits, E. A. H., Paschke, M. W., 2010. Metallophytes—a view from the
  rhizosphere. Plant and Soil 337, 33-50.
- Andreau, K., Leroux, M., Bouharrour, A., 2012. Health and cellular impacts of air pollutants:
  from cytoprotection to cytotoxicity. Biochem. Res. Int. 2012, 493894.
- Bae, O. N., Noh, M., Chun, Y. J., Jeong, T. C., 2015. Keratinocytic vascular endothelial
- 394 growth factor as a novel biomarker for pathological skin condition. Biomol. Ther. 23, 12-
- 395 18.

396	Beelen, R., Raaschou-Nielsen, O., Stafoggia, M., Andersen, Z.J., Weinmayr, G., Hoffmann,
397	B., Wolf, K., Samoli, E., Fischer, P., Nieuwenhuijsen, M., Vineis, P., Xun, W.W.,
398	Katsouyanni, K., Dimakopoulou, K., Oudin, A., Forsberg, B., Modig, L., Havulinna,
399	A.S., Lanki, T., Turunen, A., Oftedal, B., Nystad, W., Nafstad, P., De Faire, U., Pedersen,
400	N.L., Ostenson, C.G., Fratiglioni, L., Penell, J., Korek, M., Pershagen, G., Eriksen, K.T.,
401	Overvad, K., Ellermann, T., Eeftens, M., Peeters, P.H., Meliefste, K., Wang, M., Bueno-
402	de-Mesquita, B., Sugiri, D., Kramer, U., Heinrich, J., de Hoogh, K., Key, T., Peters, A.,
403	Hampel, R., Concin, H., Nagel, G., Ineichen, A., Schaffner, E., Probst-Hensch, N.,
404	Kunzli, N., Schindler, C., Schikowski, T., Adam, M., Phuleria, H., Vilier, A., Clavel-
405	Chapelon, F., Declercq, C., Grioni, S., Krogh, V., Tsai, M.Y., Ricceri, F., Sacerdote C.,
406	Galassi, C., Migliore, E., Ranzi, A., Cesaroni, G., Badaloni, C., Forastiere, F., Tamayo, I.,
407	Amiano, P., Dorronsoro, M., Katsoulis, M., Trichopoulou, A., Brunekreef, B., Hoek, G.,
408	2014. Effects of long-term exposure to air pollution on natural-cause mortality: an
409	analysis of 22 European cohorts within the multicentre ESCAPE project. Lancet (London,
410	England) 383, 785–795.
411	Cachon, B. F., Firmin, S., Verdin, A., Ayi-Fanou, L., Billet, S., Cazier, F., Martin, P. J.,
412	Aissi, F., Courcot, D., Sanni, A., Pirouz S., 2014. Proinflammatory effects and oxidative
413	stress within human bronchial epithelial cells exposed to atmospheric particulate matter
414	(PM(2.5) and PM(>2.5)) collected from Cotonou, Benin. Environ. Pollut. (Barking, Essex
415	: 1987) 185, 340-51.
416	Carletto, C., Nicolay, J. F., Courbebaisse, Y., 2000. Oxidative stress and cutaneous ageing:
417	the 'toxic second messengers' concept and an interesting family of products,
418	'pseudodipeptides'. Int. J. Cosmetic Sci. 22, 361-370.
419	Chan, J. K., Charrier, J. G., Kodani, S. D., Vogel, C. F., Kado, S. Y., Anderson, D. S.,

420 Anastasio, C., Van Winkle, L. S., 2013. Combustion-derived flame generated ultrafine

- soot generates reactive oxygen species and activates Nrf2 antioxidants differently in
  neonatal and adult rat lungs. Part. Fibre Toxicol. 10, 34.
- 423 Cheng, S., 2003. Heavy metal pollution in China: origin, pattern and control. Environ. Sci.
  424 Pollut. Res. Int. 10, 192-198.
- 425 Curjuric, I., Imboden, M., Nadif, R., Kumar, A., Schindler, C., Haun, M., Kronenberg, F.,
- 426 Kunzli, N., Phuleria, H., Postma, D. S., Russi, E.W., 2012. Different genes interact with
- 427 particulate matter and tobacco smoke exposure in affecting lung function decline in the428 general population. PloS one 7, e40175.
- 429 Dergham, M., Lepers, C., Verdin, A., Cazier, F., Billet, S., Courcot, D., Shirali, P., Garcon,
- 430 G., 2015. Temporal-spatial variations of the physicochemical characteristics of air
- pollution Particulate Matter (PM2.5-0.3) and toxicological effects in human bronchial
  epithelial cells (BEAS-2B). Environ. Res. 137, 256-267.
- 433 Drakaki, E., Dessinioti, C., Antoniou, C. V., 2014. Air pollution and the skin. Front. Environ.
  434 Sci. 2, 1-6.
- Flohr, C., Mann, J., 2014. New insights into the epidemiology of childhood atopic dermatitis.
  Allergy 69, 3-16.
- 437 Giacomoni, P. U., Rein, G., 2001. Factors of skin ageing share common mechanisms.
- 438 Biogerontology 2, 219-229.
- 439 Gilchrest, B. A., 2013. Photoaging. J. Invert. Dermatol. 133, E2-E6.
- Hanel, K. H., Cornelissen, C., Luscher, B., Baron, J. M., 2013. Cytokines and the skin barrier.
  Int. J. Mol. Sci. 14, 6720-6745.
- 442 Hirao, T., Terui, T., Takeuchi, I., Kobayashi, H., Okada, M., Takahashi, M., Tagami, H.,
- 443 2003. Ratio of immature cornified envelopes does not correlate with parakeratosis in
- 444 inflammatory skin disorders. Exp. Dermatol. 12, 591-601.

- Huang, Y. C., Li, Z., Carter, J. D., Soukup, J. M., Schwartz, D. A., Yang, I. V., 2009. Fine
  ambient particles induce oxidative stress and metal binding genes in human alveolar
  macrophages. Am. J. Respir. Cell. Mol. 41, 544-552.
- 448 Huls, A., Vierkotter, A., Gao, W., Kramer, U., Yang, Y., Ding, A., Stolz, S., Matsui, M., Kan,
- 449 H., Wang, S., Jin, L., Krutmann, J., Schikowski, T., 2016. Traffic-Related Air Pollution
- 450 Contributes to Development of Facial Lentigines: Further Epidemiological Evidence from
- 451 Caucasians and Asians. J. Investig. Dermatol. 136, 1053-1056.
- 452 Kandárová, H., Liebsc, M., Spielmann, H., Genschow, E., Schmidt, E., Traue, D., Guest, R.,
- 453 Whittingham, A., Warren, N., Gamer, A. O., Remmele, M., Kaufmann, T., Wittmer, E.,
- 454 De Wever, B., Martin Rosdy M., 2006. Assessment of the human epidermis model
- 455 SkinEthic RHE for in vitro skin corrosion testing of chemicals according to new OECD
- 456 TG 431. Toxicol. In Vitro 2, 547–559.
- 457 Kammeyer, A., Luiten, R. M., 2015. Oxidation events and skin aging. Ageing Res. Rev. 21,
  458 16-29.
- 459 Kikouama, J., Konan, K., Katty, A., Bonnet, J., Balde, L., Yagoubi, N., 2009.
- 460 Physicochemical characterization of edible clays and release of trace elements. Appl.
  461 Clay Sci. 43, 135-141.
- 462 Kim, K. E., Cho, D., Park, H. J., 2016. Air pollution and skin diseases: Adverse effects of
- 463 airborne particulate matter on various skin diseases. Life Sciences 152, 126-134.
- 464 Krutmann, J., Bouloc, A., Sore, G., Bernard, B. A., Passeron, T., 2017. The skin aging
- 465 exposome. J. Dermatol. Sci. 85, 152-161.
- 466 Krutmann, J., Liu, W., Li, L., Pan, X., Crawford, M., Sore, G., Seite, S., 2014. Pollution and
- 467 skin: from epidemiological and mechanistic studies to clinical implications. J. Dermatol.
- 468 Sci. 76, 163-168.

- Lanuti, E. L., Kirsner, R. S., 2010. Effects of pollution on skin aging. J. Investig. Dermatol.
  130, 2696.
- 471 Lecas, S., Boursier, E., Fitoussi, R., Vie, K., Momas, I., Seta, N., Achard, S., 2016. In vitro
  472 model adapted to the study of skin ageing induced by air pollution. Toxicol. Lett. 259, 60473 68.
- 474 Lefebvre, M. A., Pham, D. M., Boussouira, B., Bernard, D., Camus, C., Nguyen, Q. L., 2015.
- Evaluation of the impact of urban pollution on the quality of skin: a multicentre study in
  Mexico. Int. J. Cosmetic Sci. 37, 329-338.
- 477 Li, N., Sioutas, C., Cho, A., Schmitz, D., Misra, C., Sempf, J., Wang, M., Oberley, T.,
- Froines, J., Nel, A., 2003. Ultrafine particulate pollutants induce oxidative stress and
  mitochondrial damage. Environ. Health Perspect. 111, 455-460.
- Li, Q., Kang, Z., Jiang, S., Zhao, J., Yan, S., Xu, F., Xu, J., 2017. Effects of Ambient Fine
  Particles PM2.5 on Human HaCaT Cells. Int. J. Environ. Res. Public Health 14, 1-10.
- 482 Magnani, N. D., Muresan, X. M., Belmonte, G., Cervellati, F., Sticozzi, C., Pecorelli, A.,
- 483 Miracco, C., Marchini, T., Evelson, P., Valacchi, G., 2016. Skin Damage Mechanisms
- 484 Related to Airborne Particulate Matter Exposure. Toxicol. Sci. 149, 227-236.
- 485 Makrantonaki, E., Vogel, M., Scharffetter-Kochanek, K., Zouboulis, C. C., 2015. [Skin aging:
  486 Molecular understanding of extrinsic and intrinsic processes]. Hautarzt 66, 730-737.
- 487 Mancebo, S. E., Wang, S. Q., 2015. Recognizing the impact of ambient air pollution on skin
  488 health. J. Eur. Acad. Dermatol. Venereo., JEADV 29, 2326-2332.
- 489 Mastrofrancesco, A., Alfe, M., Rosato, E., Gargiulo, V., Beatrice, C., Di Blasio, G., Zhang,
- B., Su, D. S., Picardo, M., Fiorito, S., 2014. Proinflammatory effects of diesel exhaust
  nanoparticles on scleroderma skin cells. J. Immunol. Res. 2014, 138751.
- 492 Morita, A., Torii, K., Maeda, A., Yamaguchi, Y., 2009. Molecular basis of tobacco smoke-
- 493 induced premature skin aging. J. Investig. Dermatol. Symp. Proc. 14, 53-55.

- 494 Nemmar, A., Holme, J. A., Rosas, I., Schwarze, P. E., Alfaro-Moreno, E., 2013. Recent
- advances in particulate matter and nanoparticle toxicology: a review of the in vivo and in
  vitro studies. Biomed. Res. Int. 2013, 279371.
- 497 Pan, T. L., Wang, P. W., Aljuffali, I. A., Huang, C. T., Lee, C. W., Fang, J. Y., 2015. The
- 498 impact of urban particulate pollution on skin barrier function and the subsequent drug499 absorption. J. Dermatol. Sci 78, 51-60.
- 500 Pedata, P., Boccellino, M., La Porta, R., Napolitano, M., Minutolo, P., Sgro, L. A., Zei, F.,
- 501 Sannolo, N., Quagliuolo, L., 2012. Interaction between combustion-generated organic
- 502 nanoparticles and biological systems: in vitro study of cell toxicity and apoptosis in
- 503 human keratinocytes. Nanotoxicology 6, 338-352.
- Persoz, C., Achard, S., Leleu, C., Momas, I., Seta, N., 2010. An in vitro model to evaluate the
  inflammatory response after gaseous formaldehyde exposure of lung epithelial cells.
  Toxicol. Lett. 195, 99-105.
- 507 Poumay, Y., Dupont, F., Marcoux, S., Leclercq-Smekens, M., Herin, M., Coquette, A., 2004.
- A simple reconstructed human epidermis: preparation of the culture model and utilization
  in in vitro studies. Arch. Dermatol. Res. 296, 203-211.
- 510 Puizina-Ivic, N., 2008. Skin aging. Acta Dermatovenerol. Alp. Pannonica Adria. 17, 47-54.
- 511 Raaschou-Nielsen, O., Andersen, Z. J., Beelen, R., Samoli, E., Stafoggia, M., Weinmayr, G.,
- 512 Hoffmann, B., Fischer, P., Nieuwenhuijsen, M. J., Brunekreef, B., et al., 2013. Air
- 513 pollution and lung cancer incidence in 17 European cohorts: prospective analyses from
- 514 the European Study of Cohorts for Air Pollution Effects (ESCAPE). The Lancet.
- 515 Oncology 14, 813-822.
- 516 Ressler, S., Bartkova, J., Niederegger, H., Bartek, J., Scharffetter-Kochanek, K., Jansen-Dürr,
- 517 P., Wlaschek, M., 2006. p16 INK4A is a robust in vivo biomarker of cellular aging in
- 518 human skin. Aging Cell 5, 379-389.

- 519 Sherratt, M. J., 2013. Age-Related Tissue Stiffening: Cause and Effect. Adv. Wound Care 2,
  520 11-17.
- Song, S., Paek, D., Lee, K., Lee, Y. M., Lee, C., Park, C., Yu, S. D., 2011. Effects of ambient
  fine particles on pulmonary function in children with mild atopic dermatitis. Arch.
- 523 Environ. Occup. Health. 68, 228-34.
- 524 Urbanska, M., Nowak, G., Florek, E., 2012a. Cigarette smoking and its influence on skin
  525 aging. Przegl. Lek., 1111-1114.
- 526 Urbanska, M., Ratajczak, L., Witkowska-Nagiewicz, A., 2012b. [Analysis of knowledge
  527 about tobacco smoking influence on skin condition]. Przegl. Lek., 69, 1055-1059.
- 528 Valacchi, G., Sticozzi, C., Pecorelli, A., Cervellati, F., Cervellati, C., Maioli, E., 2012.
- 529 Cutaneous responses to environmental stressors. Ann. N. Y. Acad. Sci. 1271, 75-81.
- Vierkotter, A., Krutmann, J., 2012. Environmental influences on skin aging and ethnicspecific manifestations. Dermatoendocrinol. 4, 227-231.
- 532 Vierkotter, A., Schikowski, T., Ranft, U., Sugiri, D., Matsui, M., Kramer, U., Krutmann, J.,
- 2010. Airborne particle exposure and extrinsic skin aging. J. Invest. Dermatol. 130, 27192726.
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543	Figure	legends
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- 544 Fig. 1. Morphological observations of RHEs after PM-exposure.
- 545 RHE tissues were exposed to PM at 0 (Control), 3, 15, or 30 µg/cm<sup>2</sup> for 24h. After exposure
- 546 RHEs were fixed in 4% paraformaldehyde and cross-sections (4 µm) were stained with Mayer
- 547 Haematoxylin and Eosin (H&E) and observed using an inverted light microscope (Leica
- 548 DM4000B LED).
- 549 Scale bar represents  $25 \,\mu m$ ; 40x magnification. Representative pictures of n=3 independent 550 experiments.
- 551 (A) RHE Control. 1/Stratum germinativum; 2/Stratum spinosum; 3/Stratum granulosum;
- 552 4/Stratum corneum.
- 553 (B) RHE exposed at (B1)  $3 \mu g/cm^2$ ; (B2)  $15 \mu g/cm^2$ ; (B3)  $30 \mu g/cm^2$ .
- 554

#### 555 Fig. 2. Effects of PM on RHE functionality and membrane integrity.

- 556 RHE tissues were exposed to PM at 0 (Control), 3, 15, or  $30 \,\mu g/cm^2$  for 24h.
- 557 Experiments were repeated at least four times in duplicate.
- Results are presented as mean  $\pm$  SD ; \*\* p<0.01 ; \*\*\* p<0.001 compared to the Control.
- A. RHE functionality was assessed 24h after PM-exposure using MTT assay. Results are
- 560 expressed in absorbance (arbitrary units).
- 561 B. Membrane integrity of RHE was evaluated 24h after PM-exposure by intracellular lactate
- 562 dehydrogenase (LDH) release expressed in arbitrary unit per liter (U/L).
- 563 Maximal LDH activity was obtained after 5% SDS treatment (Positive control).

564

#### 565 Fig. 3. Inflammatory response of RHE after PM-exposure.

566 RHE tissues were exposed to PM at 0 (Control), 3, 15, or  $30 \,\mu g/cm^2$  for 24h.

- 567 After exposure cytokine secretions [IL-1 $\alpha$  (plain bar) and IL-8 (dashed bar)] were measured
- 568 by immuno-quantification (ELISA test) in the culture medium in the basolateral
- 569 compartment.
- 570 Data are representative of 4 experiments repeated in duplicate.
- 571 Results are presented as mean  $\pm$  SD; \*\*\* p<0.0001 compared to the Control.
- 572

#### 573 Fig. 4. Changes in 4-HNE expression level after PM-exposure.

- 574 RHE tissues were exposed to PM at 0 (Control) or  $30 \,\mu g/cm^2$  for 24h.
- 575 (A) After PM-exposure RHEs were fixed in 4% paraformaldehyde, embedded on paraffin and576 sliced.
- 577 Cross-sections (4 µm) were blocked in BSA/PBS buffer and stained using 4-HNE (protein
- 578 produced when lipid peroxidation is) and DAPI (nucleus).
- Scale bar represents 25 µm ; 200x magnification. Representative pictures of n=3 independent
  experiments.
- (B) Quantification of 4-HNE in comparison to Control using Visilog software version 7.0.
- 582 Data are representative of 3 independent experiments. Results are presented as mean  $\pm$  SD ;
- 583 \*\*\* p<0.0001 compared to the Control.
- 584 (C) After exposure, total RNA was extracted from tissues using the RNeasy® miniKit and
- 585 reverse transcribed using the Transcriptor Universal cDNA Master. The gene expressions
- 586 were normalized with GAPDH housekeeping gene.
- 587 Results were obtained by calculating the fold change (FC) between two conditions: exposed
- 588 *versus* Control. Increased oxidant gene expression was observed:  $FC \ge 1.5$ .
- 589 Results are the mean from one RT-qPCR experiments assessed in duplicate.
- 590
- 591 Fig. 5. Impact of PM exposure on MMP matrix metalloproteinase.

- 592 After PM-exposure of RHEs ( $30 \mu g/cm^2$ ), metalloproteinase secretion [MMP-1 (plain bar)
- and MMP-3 (dashed bar)] were measured by immuno-quantification (Luminex® technology)
- in the culture medium in the basolateral compartment.
- 595 Data are representative of 2 experiments repeated in triplicate.
- Results are presented as mean  $\pm$  SD; \*\*\* p<0.0001 compared to the Control.
- 597

#### 598 Fig. 6. Changes in loricrin expression level after PM exposure.

- 599 RHE tissues were exposed to PM at 0 (Control) or  $30 \,\mu g/cm^2$  for 24h.
- 600 (A) After PM-exposure RHEs were fixed in 4% paraformaldehyde and embedded on paraffin.
- 601 Cross-section of RHE (4 µm) stained using loricrin (major protein of differentiated epidermal
- 602 cells) and DAPI (nucleus).
- Scale bar represents 25 μm ; 200x magnification. Representative pictures of n=3 independent
   experiments.
- (B) Quantification of loricrin in comparison to control using Visilog software version 7.0.
- Data are representative of 3 independent experiments. Results are presented as mean  $\pm$  SD ;
- 607 \*\*\* p<0.0001 compared to the Control

609	Fig. 7. Changes in mRNA	A expression levels of gene	es involved in differentiation	on, cell to
			2	

- 610 cell junction, anchorage and apoptosis after exposure to  $30 \ \mu g/cm^2$  of PM.
- 611 RHE tissues were exposed at  $30 \,\mu\text{g/cm}^2$  of PM for 24hours. After exposure, total RNA was
- 612 extracted from tissues using the RNeasy® miniKit and reverse transcribed using the
- 613 Transcriptor Universal cDNA Master. The gene expressions were normalized with GAPDH
- 614 or RPL13 housekeeping gene.
- Results were obtained by calculating the fold change (FC) between two conditions: exposed *versus* Control).

617	mRNA expression was considered as increased when $FC \ge 1.5$ (dark grey) and decreased

618 when FC  $\leq$  0.7 (light grey).

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619 Results are the mean from one RT-qPCR experiments assessed in duplicate.
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620

#### 621 Fig. 8. MoA hypothesis for PM-induced premature skin ageing.

622 The mode of action (MoA) hypothesis of the airborne pollution on the premature skin ageing

623 process, using the RHE 3D-skin model exposed to fine particles, is subdivided into key

624 events, associative events and modulating factors. The initial key event is the increase of the

oxidative stress, associated with the modulation of the inflammatory response, which results

- 626 in an overexpression of various genes. Associative events include inhibition of the gap
- 627 junction communication and anchorage.
- 628

#### 629 Table 1. Sum of main metallic elements and organic components of fine particles (PM<sub>0.3</sub>.

- 630 **2.5**) from (Cotonou, Benin) (adapted from Cachon et al., 2014).
- 631 Chemicals are expressed as mass composition per mass ( $\mu g.g^{-1}$ ) and volume unit ( $ng/m^3$ ).

632 Sampling of fine particles was based on continuous collection using high volume samplers

633 ( $80 \text{ m}^3/\text{hr}$ ).

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#### 637 Supplementary data

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639 Figure S1. Microscopic observation of the homogeneous dispersion of PM prior RHE-

640 exposure.

- 641 Scale bar represents 100 μm; 20x magnification.
- 642 Representative pictures of n=3 preparations
- 643
- Table S1. Chemical composition of the fine particles (PM<sub>0.3-2.5</sub>) collected in (Cotonou,
- 645 **Benin**) (Cachon et al. 2014).
- 646 PM were collected using a high volume cascade impaction air sampler as described by Billet
- 647 (2007) and the physicochemical characterization was carried out by studying their size
- 648 distributions, specific surface areas, organic and inorganic chemicals composition and ionic
- 649 species as described by Cazier (2016). Chemicals are expressed as mass composition per
- 650 volume unit  $(ng.m^{-3})$ .
- Billet, S. et al., 2007. Ambient particulate matter (PM<sub>2.5</sub>): physicochemical characterization
- and metabolic activation of the organic fraction in human lung epithelial cells (A549).
- 653 Environ. Res. 105, 212–23.
- 654 Cazier, F. et al., 2016. Characterisation and seasonal variations of particles in the atmosphere
- of rural, urban and industrial areas: Organic compounds. J. Env Sci. S44, 45-56.
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## 660 Highlights

- Airborne particles induce skin damages
- Oxidative and inflammatory pathways modulation after skin exposure to PM
- Premature skin ageing after PM exposure

#### Fig. 1. Morphological observations of RHEs after PM-exposure.

RHE tissues were exposed to PM at 0 (Control), 3, 15, or 30  $\mu$ g/cm<sup>2</sup> for 24h. After exposure RHEs were fixed in 4% paraformaldehyde and cross-sections (4  $\mu$ m) were stained with Mayer Haematoxylin and Eosin (H&E) and observed using an inverted light microscope (Leica DM4000B LED). Scale bar represents 25  $\mu$ m ; 40x magnification.

Representative pictures of 3 independent experiments.

(A) RHE Control. 1/Stratum germinativum; 2/Stratum spinosum; 3/Stratum granulosum; 4/Stratum corneum.



(B) RHE exposed at (B1)  $3 \mu g/cm^2$ ; (B2)  $15 \mu g/cm^2$ ; (B3)  $30 \mu g/cm^2$ .

#### Fig. 2. Effects of PM on RHE functionality and membrane integrity.

RHE tissues were exposed to PM at 0 (Control), 3, 15, or 30 µg/cm<sup>2</sup> for 24h.

Experiments were repeated at least four times in duplicate.

Results are presented as mean  $\pm$  SD ; \*\* p<0.01 ; \*\*\* p<0.001 compared to the Control

(A) RHE functionality was assessed 24h after PM-exposure using MTT assay. Results are expressed in

absorbance (arbitrary units).

(B) Membrane integrity of RHE was evaluated 24h after PM-exposure by intracellular lactate dehydrogenase (LDH) release expressed in arbitrary unit per liter (U/L).

Maximal LDH activity was obtained after 5% SDS treatment (Positive control).



#### Fig. 3. Inflammatory response of RHE after PM-exposure.

RHE tissues were exposed to PM at 0 (Control), 3, 15, or 30 µg/cm<sup>2</sup> for 24h.

After exposure cytokine secretions [IL-1 $\alpha$  (plain bar) and IL-8 (dashed bar)] were measured by immuno-quantification (ELISA test) in the culture medium in the basolateral compartment.

Data are representative of 4 experiments repeated in duplicate.

Results are presented as mean  $\pm$  SD ; \*\*\* p<0.0001 compared to the Control



PM concentrations

#### Fig. 4. Changes in 4-HNE expression level after PM-exposure.

RHE tissues were exposed to PM at 0 (Control) or 30  $\mu$ g/cm<sup>2</sup> for 24h.

(A) After PM-exposure RHEs were fixed in 4% paraformaldehyde, embedded on paraffin and sliced. Cross-sections (4  $\mu$ m) were blocked in BSA/PBS buffer and stained using 4-HNE (protein produced when lipid peroxidation is) and DAPI (nucleus).

Scale bar represents 25  $\mu$ m ; 200x magnification. Representative pictures of 3 independent experiments.



(B) Quantification of 4-HNE in comparison to Control using Visilog software version 7.0.

Data are representative of 3 independent experiments. Results are presented as mean  $\pm$  SD ; \*\*\* p<0.0001 compared to the Control.



(C) After exposure, total RNA was extracted from tissues using the RNeasy® miniKit and reverse transcribed using the Transcriptor Universal cDNA Master. The gene expressions were normalized with GAPDH housekeeping gene.

Results were obtained by calculating the fold change (FC) between two conditions: exposed *versus* Control.

Increased oxidant gene expression was observed:  $FC \ge 1.5$ .

Results are the mean from one RTqPCR experiments assessed in duplicate.



#### Fig. 5. Impact of PM-exposure on MMP matrix metalloproteinase.

After PM-exposure of RHEs (30  $\mu$ g/cm<sup>2</sup>), metalloproteinase secretion [MMP-1 (plain bar) and MMP-3 (dashed bar)] were measured by immuno-quantification (Luminex® technology) in the culture medium in the basolateral compartment.

Data are representative of 2 experiments repeated in triplicate.

Results are presented as mean  $\pm$  SD ; \*\*\* p<0.001 compared to the Control.



#### Fig. 6. Changes in loricrin expression level after PM-exposure.

RHE tissues were exposed to PM at 0 (Control) or 30  $\mu$ g/cm<sup>2</sup> for 24h.

(A) After PM-exposure RHEs were fixed in 4% paraformaldehyde and embedded on paraffin. Crosssection of RHE (4 µm) stained using loricrin (major protein of differentiated epidermal cells) and DAPI (nucleus).

Scale bar represents 25  $\mu$ m ; 200x magnification. Representative pictures of 3 independent experiments.



(B) Quantification of loricrin in comparison to control using Visilog software version 7.0.

Data are representative of 3 independent experiments. Results are presented as mean  $\pm$  SD ;

\*\*\* p<0.0001 compared to the Control



# Fig.7. Changes in mRNA expression levels of genes involved in differentiation, cell to cell junction, anchorage and apoptosis after exposure to 30 μg/cm<sup>2</sup> of PM.

RHE tissues were exposed at 30  $\mu$ g/cm<sup>2</sup> of PM for 24h. After exposure, total RNA was extracted from tissues using the RNeasy® miniKit and reverse transcribed using the Transcriptor Universal cDNA Master. Gene expressions were normalized with GAPDH or RPL13 housekeeping gene.

Results were obtained by calculating the fold change (FC) between two conditions: exposed *versus* Control). mRNA expression was considered as increased when  $FC \ge 1.5$  (dark grey) and decreased when  $FC \le 0.7$  (light grey).

Results are the mean from one RTqPCR experiments assessed in duplicate.



#### Fig. 8. MoA hypothesis for PM-induced premature skin aging

The mode of action (MoA) hypothesis of airborne pollution on premature skin ageing process, using the RHE 3D-skin model exposed to fine particles, is subdivided into key events, associative events, and modulating factors. The initial key event is the increase in oxidative stress, associated with the modulation of inflammatory response, which results in an overexpression of various genes. Associative events include inhibition of the gap junction communication and anchorage.



### Table 1. Sum of main metallic elements and organic components of fine particles (PM<sub>0.3-2.5</sub>) from

(Cotonou, Benin) (adapted from Cachon et al., 2014).

Chemicals are expressed as mass composition per mass ( $\mu g.g^{-1}$ ) and volume unit ( $ng.m^{-3}$ ).

Samplings of fine particles were based on continuous collection using high volume samplers (80  $m^3/hr$ ).

	Chemical composition of fine particles PM <sub>0.3-2.5</sub> Cotonou, Benin, West Africa	
	μg.g <sup>-1</sup>	ng.m <sup>-3</sup>
Major and trace elements		
Al, Ba, Cr, Cu, Fe, Mn, Ni, Pb, Ti, V, Zn		
Sum of elements	63,723	8,644
Organic compounds		
Benzene, Toluene, o,p-Xylene, Naphtalene, Phenanthrene, Alkyl-benzene, Benzo[a]pyrene, Benzo[g,h,i] perylene, <i>etc</i>		
Sum of organic compounds	78	11