

Development and validation of an LC–MS/MS method for the simultaneous determination of bisphenol A and its chlorinated derivatives in adipose tissue

Nicolas Venisse, Guillaume Cambien, Julien Robin, Steeve Rouillon, Cédric Nadeau, Thomas Charles, Sylvie Rabouan, Virginie Migeot, Antoine Dupuis

▶ To cite this version:

Nicolas Venisse, Guillaume Cambien, Julien Robin, Steeve Rouillon, Cédric Nadeau, et al.. Development and validation of an LC–MS/MS method for the simultaneous determination of bisphenol A and its chlorinated derivatives in adipose tissue. Talanta, 2019, 204, pp.145 - 152. 10.1016/j.talanta.2019.05.103 . hal-03485822

HAL Id: hal-03485822 https://hal.science/hal-03485822

Submitted on 20 Dec 2021

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers. L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.



Distributed under a Creative Commons Attribution - NonCommercial 4.0 International License

Development and validation of an LC–MS/MS method for the simultaneous determination of Bisphenol A and its chlorinated derivatives in adipose tissue

Nicolas VENISSE^{a,b,*}, Guillaume CAMBIEN^{a,b}, Julien ROBIN^b, Steeve ROUILLON^{a,b}, Cédric
 NADEAU^c, Thomas CHARLES^d, Sylvie RABOUAN^{a,b,e}, Virginie MIGEOT^{a,b,e}, Antoine
 DUPUIS^{a,b,e}

- 7
- ⁸ ^aINSERM, University Hospital of Poitiers, University of Poitiers, CIC1402, HEDEX research
- 9 group, 86021 Poitiers Cedex, France.
- 10
- ¹¹ ^bBiology-Pharmacy-Public Health Department, University Hospital of Poitiers, 2 rue de la
- 12 Milétrie, 86021 Poitiers Cedex, France.
- ¹³ ^cDepartment of Gynecology and Obstetrics, University Hospital of Poitiers, 2 rue de la Milétrie,
- 14 86021 Poitiers Cedex, France.
- ^dDepartment of Urology, University Hospital of Poitiers, 2 rue de la Milétrie, 86021 Poitiers
 Cedex, France.
- ¹⁷ ^eFaculty of Medicine and Pharmacy, University of Poitiers, TSA 51115, 86073 Poitiers Cedex.
- 18 *Corresponding author.
- 19 Nicolas Venisse
- 20 INSERM, CIC1402
- 21 HEDEX research group
- 22 University Hospital of Poitiers
- 23 86021 Poitiers Cedex
- 24 France
- 25 Tel: 33 5 49 44 49 80
- 26 E-mail address: nicolas.venisse@chu-poitiers.fr
- 27
- 28 Declarations of interest: none.
- 29
- 30
- ---
- 31

32 Abstract

Bisphenol A (BPA) and its chlorinated derivatives (Clx-BPA) are environmental pollutants exhibiting endocrine-disrupting (ED) properties suspected to be involved in the pathogenesis of hormone-dependent cancers, such as breast and prostate cancers. Due to their lipophilic properties, they may accumulate in adipose tissue which could therefore be a suitable matrix to assess long-term exposure to these compounds and relationships with the tumorigenesis of these cancers.

An LC-MS/MS assay for the determination of BPA and Clx-BPA in adipose tissue samples was 39 developed and fully validated according to current bioanalytical validation guidelines. Ionization 40 41 was achieved using an electrospray source operating in the negative mode and quantification of target analytes was obtained in the multiple reaction monitoring mode. Both standard and 42 quality control (QC) samples were prepared in blank adipose tissue samples. Linearity was 43 44 demonstrated over the ranges 0.125 to 8.000 and 0.0125 to 0.8000 ng/mL for BPA and Clx-45 BPA, respectively. Accuracy and precision were demonstrated over the whole concentration range: intra and inter-day bias values were in the 85 - 114% range and imprecision of the 46 method did not exceed 14%. Lower limits of quantification were validated using QCs at 0.1250 47 and 0.0125 ng/mL for BPA and Clx-BPA, respectively. Internal standard-corrected matrix 48 49 effects were comparable in breast and prostate adipose tissues, demonstrating that this method could be used to reliably assay BPA and Clx-BPA in both tissues. The method was 50 sensitive enough to determine BPA and CIx-BPA in breast adipose tissue obtained from 51 52 women undergoing breast surgery, enabling identification of different patterns of exposure to these ED chemicals. 53

The method enables the reliable quantification of BPA and Clx-BPA in adipose tissue and could be used to assess long-term exposure to these compounds and potential associations with hormone-dependent cancers.

58 Keywords:

- 59 Adipose tissue; Bisphenol A and its chlorinated derivatives; Endocrine disruptors; Breast ;
- 60 Prostate; Hormone-dependent cancer.

62 **1. Introduction**

Bisphenol A (BPA) is a well-known ubiquiteous environmental pollutant exhibiting endocrine-63 disrupting properties [1]. Chlorinated derivatives of bisphenol A (Clx-BPA) are formed by 64 65 chlorination of bisphenol A present in water during water-disinfection process. As a result Clx-BPA have essentially been observed in various aqueous media such as tap water [2,3], 66 67 sewage sludge [4] and wastewater of paper recycling plant [5]. Clx-BPA have also been found in human biological matrices: urine [6-8], placental tissue [9], breast milk [10] and colostrum 68 [11]. Clx-BPA exhibit endocrine-disrupting properties: in vitro they show higher (up to 100 69 times) estrogenic activities than BPA [12]. Riu et al. have demonstrated that Clx-BPA are 70 71 human PPARy ligands and that activation of PPARy depends on the degree of halogenation [13]. In humans, Clx-BPA exposure has been associated with the occurrence of type 2 diabetes 72 73 [14] and obesity [15].

Due to direct activation of the estrogen receptor, BPA may have an impact on carcinogenesis in hormone-dependent cancers such as breast [16] and prostate cancers [17]. This is also the case for Clx-BPA [18]. The agonist effect of BPA and Clx-BPA on PPARγ receptors may also have an impact on tumoregenesis in hormone-dependent cancer [19,20] and they may exert indirect tumoregenesis activity through obesity and inflammation [21,22].

Assessing human exposure to environmental pollutants is a major component of health risk assessment [23]. For that purpose, several approaches have been proposed and determination of exposure can be either indirect or direct. However, direct measurement of environmental pollutants in human body fluids, commonly urine or blood, is now considered as the method of reference. This is the so-called Human Biomonitoring (HBM) that reflects the total body burden resulting from all routes of exposure [24].

As a means of relating BPA and Clx-BPA exposure to cancer pathogenesis, blood or urine concentrations are probably poor biomarkers of exposure due to short biological half-life, at least for BPA [25]. It has been demonstrated that persistent [26] as well as non-persistent pollutants [27,28] may accumulate in adipose tissue due to their lipophilic properties. These results suggest that concentration in adipose tissue may be a suitable biomarker in assessment of long-term exposure [28].

Not many methods have been reported for the determination of BPA in adipose tissue. One is based on gas chromatography-mass spectrometry (GC-MS) [29] while others are based on LC-MS/MS [27,28,30]. To our knowledge, no LC-MS/MS method has been reported for the simultaneous determination of BPA and Clx-BPA. Fernandez et al. have briefly reported on a GC-MS method [31].

Therefore the objectives of this study were to develop and validate an LC-MS/MS assay for simultaneous determination of BPA and Clx-BPA in adipose tissue and to apply the validated method to a set of real samples in view of demonstrating its suitability and relevance.

100 101

2. Materials and methods

2.1. Chemicals and reagents

Bisphenol A (BPA) and its deuterated analogue (BPA-d₁₆) were obtained from Sigma-Aldrich® 102 (St Louis, USA). Chlorinated derivatives of BPA (Clx-BPA) and the deuterated analogue of 103 dichlorobisphenol A (2,2'-DCBPA-d₁₂) were purchased from @rtMolecule® (Poitiers, France). 104 105 Acetonitrile, methanol and water were acquired from Carlo Erba® (Val de Reuil, France). They were of Ultra-High Performance Liquid Chromatography (UHPLC) grade and we demonstrated 106 that they were free of BPA and Clx-BPA. Materials in glass or polypropylene were used to 107 avoid any BPA contamination. MIX I (MgSO₄, NaCl, Na₂H citrate - 1.5 H₂O, Na₃ Citrate - 2 108 109 H₂O) and MIX VI (MgSO₄, Diamino PSA (Primary Secondary Amine), C₁₈ ec (Octadecyl modified silica phase, endcapped)) were purchased from Machery-Nagel® (Hœrdt, France). 110

111

2.2. Preparation of stock solutions, working solutions and spiking solutions

112 Analyte stock solutions were prepared in methanol at 10 μ g/mL for BPA and 1 μ g/mL for Clx-113 BPA. Deuterated internal standard (IS) stock solutions were at 10 μ g/mL for BPA-d₁₆ and 1 µg/mL for 2,2'-DCBPA-d₁₂. Stock solutions were stored at -20°C before use. Each day of 114 analysis, two working solutions were freshly prepared in methanol: one with analytes at 80 115 ng/mL and 8 ng/mL for BPA and Clx-BPA, respectively, and one at 40 ng/mL and 4 ng/mL for 116 BPA-d₁₆ and 2,2'-DCBPA-d₁₂, respectively. Working solutions were diluted with methanol to 117 obtain spiking solutions ranging from 1.25 ng/mL to 80.00 ng/mL and from 0.125 to 8.000 118 ng/mL ng/mL for BPA and Clx-BPA, respectively. These solutions were used to prepare 119 calibration standards in adipose tissue. 120

121 2.3. Sample collection

Samples of adipose tissue were obtained from patients during breast or prostate surgery at the university hospital of Poitiers, France. Patients who had given their informed consent after having been fully informed could be enrolled in this study. Samples were collected in glass tubes to avoid contamination with BPA. All samples were stored at -20 °C before use.

126 2.4. Sample extraction

127 Each sample of adipose tissue (100 mg) was homogenized in a clean glass tube with 500 µL of water using an Ultra-Turrax® (IKA® T18 Basic), and spiked with 50 µL of IS working 128 129 standard solution. Calibrant samples were prepared in adipose tissue by spiking tissue homogenate with 50µL of spiking solutions to obtain final concentrations ranging from 8.000 130 131 ng/mL to 0.125ng/mL and 0.8000 to 0.0125 ng/mL, for BPA and Clx-BPA, respectively. For blank and unknown samples, spiking solution was replaced by 50 µL of pure methanol. For 132 double-blank samples, spiking and IS solutions were replaced by 100 µL of pure methanol. 133 Then 500 µL of acetonitrile were added to homogenates. Samples were vortex-mixed for one 134 135 minute and 100 mg of MIX I were added. pH was adjusted to 5.5 with NaOH 1M, vortex-mixed again for one minute and centrifuged at 2330 × g for five minutes. The upper organic layer was 136 137 transferred into a clean glass tube, 100 mg of MIX VI were added, vortexed again for 30 seconds and centrifuged at 2330 × g for five minutes. The upper organic layer was transferred 138 139 into a clean glass-tube and evaporated to dryness at 37°C under a gentle nitrogen stream. Dry residues were dissolved in 200 µL of a water/methanol (70/30, v/v) solution, vortex-mixed and 140 transferred into glass vials for injection. Finally, 20 µL of extract were injected into the LC-141 142 MS/MS system.

143

Liquid chromatography

2.5.

A UHPLC Shimadzu® Nexera X2 (Kyoto, Japan) pump was equipped with a Waters® Acquity CSH C18 (3.5 μ m particle size, 2.1 x 100 mm, Milford, USA) column heated at 40°C. Chromatographic separation was achieved using a binary mobile phase (water (A)/methanol (B)) delivered in the gradient mode at a flow rate of 350 μ L.min⁻¹. Gradient elution started at 30% of B (0-0.5 minute); then linearly increased to 90% of B (0.5-7 min); then to 99% of B (7-7.5 min); maintained at 99% of B for 5 minutes (7.5-12.5 min); then decreased to 30% of B (12.5-13 min) and finally maintained at 30% of B for 2.5 minutes (13-15.5 min) for reequilibration of the column.

152

Mass spectrometry

Analyte quantification was performed with an API6500+ mass spectrometer from Sciex® 153 (Concord, Canada) using an electrospray ionization (ESI) interface, operating in negative 154 ionization mode. Ion source and MS/MS conditions were optimized with infusion of a 1 µg/mL 155 solution of the individual analytes at 10 µL/min. For each compound multiple reaction 156 monitoring mode (MRM) with two specific daughter fragment transitions for one parent 157 precursor ion was used. The first was used for quantification while the second was used for 158 confirmation. Dwell and cycle times were optimized using the Scheduled MRM® algorithm. 159 160 Data were acquired with Analyst® software and quantification was obtained with MultiQuant® 161 software.

162 2.7. Validation

2.6.

Bioanalytical method validation was performed according to EMA guidelines [32]. Selectivity, calibration curve linearity, accuracy, precision, lower limits of the method and matrix effect were determined.

To determine selectivity, for each analyte and deuterated standard we examined
chromatograms at their corresponding retention time in blank adipose tissue samples.
Acceptable selectivity was defined by the absence of any detectable peak.

169 Calibrant samples were prepared using a pool of adipose tissue obtained from donors. 170 Calibration curves, ranging from 0.125 to 8 (0.125, 0.25, 0.5, 1.0, 2.0, 4.0, 8.0) ng/mL, and 171 from 0.0125 to 0.8 (0.0125, 0.025, 0.05, 0.1, 0.2, 0.4, 0.8) ng/mL for BPA and Clx-BPA, 172 respectively, were constructed with 1/x weighting. For each analyte five sets of calibration 173 curves were examined. R squared, slope and y-intercept were determined. Linearity was 174 demonstrated when correlation coefficient was over 0.99. Inter- and intra-day accuracy and precision were determined using QCs prepared with blank
breast adipose tissue samples. QC concentrations were 0.125, 1.0, 4.0 ng/mL and 0.0125,
0.10, 0.40 ng/mL for BPA and Clx-BPA, respectively. Trueness values within the 85-115%
range were accepted for validation of accuracy; a maximal 80-120% range was accepted for
low QC. Precision was evaluated and expressed as coefficient of variations (CV%). Values
had to be lower or equal than 20% for low QC and 15% for medium and high QCs.

In each batch of analysis (for method validation and routine analysis), five blank samples were analyzed. The corresponding mean and standard deviation of the peak area were calculated for each target analyte. The limit of detection (LOD) was set to three times the standard deviation for each target analyte [8].

The limit of quantification (LOQ), defined as the lowest level of the calibration standards (0.125 ng/mL for BPA and 0.0125 ng/mL for Clx-BPA) was fully validated with low-level QCs prepared with breast adipose tissue spiked at this concentration [32].

Matrix effect (ME) and recovery (RE) were also determined over the whole concentration range according to the method described by Matuszewski et al. [33]. Briefly, the signals from the analysis of three sets of samples were compared: (A) extracted QCs prepared in breast adipose tissue, (B) blank breast adipose tissue samples spiked at the same concentrations after extraction and (C) corresponding neat standards prepared in mobile phase, all at three levels of concentration (low, medium and high). IS corrected matrix effect was also calculated [33].

A complementary set of experiments was conducted in order to demonstrate that this method could be applied to other types of adipose tissues for example prostate adipose tissue. Calibrant and QC samples were prepared using blank prostate adipose tissue. Matrix effect and recovery were assessed and compared to results obtained with breast adipose tissue. Linearity and accuracy were also assessed in prostate adipose tissue samples.

201

202 **3. Results**

203

3.1. LC-MS/MS conditions

Optimal conditions for the electrospray ionization source operating in the negative mode were determined as follows: ion spray voltage at -4500V, source temperature at 650°C, curtain gas pressure at 30.0 psi, collision gas set as high, ion source gas 1 pressure at 40.0 psi and ion source gas pressure 2 at 60.0 psi. Specific mass spectrometer parameters (declustering potential, entrance potential, collision energy and collision cell exit potential) for each quantification and ion transition pairs are reported in Table 1.

Chromatographic conditions allowed peak separation for each target analyte (Figure 1) andcorresponding retention times are reported in Table 1.

212

3.2. Method validation

213 No significant chromatographic peaks were observed at the retention times of Clx-BPA in blank breast adipose tissue samples (Figure 2), thereby demonstrating the specificity of the method. 214 A peak of BPA was present due to baseline contamination of human matrices with this 215 ubiquitous pollutant. Linearity of the method was demonstrated over the whole concentration 216 217 range with correlation coefficients >0.99 (Table 2). The high value of y-intercept found for 218 bisphenol A was due to basal occurrence of this compound in adipose tissue (Figure 2). The 219 accuracy and precision of the method were demonstrated over the whole range of concentration. Intra and inter-day bias values were in the range 85 – 114% in accordance with 220 221 bioanalytical validation guidelines. Imprecision of the method (both intra- and inter-day) did not exceed 14%. The limits of quantification, corresponding to the lowest standards of the 222 calibration curves, were validated at 0.1250 and 0.0125 ng/mL for BPA and Clx-BPA, 223 respectively. At the LOQ, good accuracies and precisions were demonstrated (Table 3) and 224 high signal-to-noise peaks were obtained (Figure 2). Matrix effects in breast tissue were low 225 to moderate since ion suppression and enhancement did not exceed 31 and 2%, respectively 226

(Table 4). They were satisfactorily corrected by internal standards (Table 4). Recoveries were
 moderate but, importantly, were consistent over the whole range of concentration (Table 4).

229 Slightly higher absolute matrix effects were measured in prostate adipose tissue but they were 230 also satisfactorily corrected by IS (Table 5). Recoveries of target analytes from prostate 231 adipose tissues were in the same range as those from breast adipose tissues (Table 5). 232 Linearity was also demonstrated using this matrix (Table 6).

233

3.3.

Application of the method

This method enables the determination of BPA and Clx-BPA in adipose tissue samples from women receiving breast surgery for breast lesions. BPA was systematically quantified in these samples with concentrations ranging from 0.238 to 1.745 ng/mL (1.19 - 8.73 ng/g) (Table 7). Exposure to Clx-BPA was variable among participants: overall detection frequency was 45%. Table 7 shows that some patients are exposed to several of these compounds (patient BE) whereas others are not (patient RS).

241 **4. Discussion**

For the first time, an LC-MS/MS method enabling the simultaneous determination of BPA and Clx-BPA in adipose tissue has been developed and validated according to well-established bioanalytical validation guidelines. Methods for assaying Bisphenol A in adipose tissue had previously been developed using LC-MS/MS [27,28,30] or GC-MS [29]. A GC-MS method for analysis of BPA and Clx-BPA initially developed for water and plasma samples has been succinctly described as a means of analysis of adipose tissue [31]

248 Several attempts were made to optimize extraction procedure. We started with a method similar to the one used for the analysis of Clx-BPA in milk samples [10] using either methanol 249 250 or acetonitrile as a precipitation reagent. This method enabled to obtain high recovery rates of 251 target analytes but also led to high matrix effects with adipose tissue. One key step to reduce 252 matrix effect was removal of lipid content from adipose tissue samples. The method proposed by Cariou et al. [34] was not chosen due to a labor-intensive protocol. We also avoided the 253 254 use of freezing for lipid removal, with freezing time ranging from 15 min to overnight [28,30], 255 to shorten experimentation time. Our final method involves a very simple sample preparation 256 step with QuEChERS reagents. While QuEChERS-based extraction has frequently been used in food analysis [35,36], to our knowledge this is the first time that it is considered as a means 257 of extracting environmental pollutants from human adipose tissue. Our extraction method 258 259 involves two simple and rapid steps: the first is a salting-out extraction step with acetonitrile and the second is a clean-up step using dispersive solid-phase extraction with a C18 sorbent 260 to remove lipids. Using our method, recoveries were moderate but did not impact the overall 261 performances. Indeed, aside from its simplicity and low cost, our method proves to be effective. 262 263 Separation of target analytes was achieved using a 3.5 µm particle size column. Although the HPLC system was suitable for UHPLC, this particle size was chosen for better column life time. 264

265 Our assay has been fully validated according to well-established bioanalytical validation 266 guidelines: linearity, repeatability and reproducibility have been demonstrated using calibration

standards and QCs prepared in adipose tissue, which is recommended by guidelines but not 267 always performed. Special emphasis was put on the study of matrix effect and overall matrix 268 269 effects were low for all target analytes in both tested matrices. In an ideal situation each target 270 analyte would have its own corresponding stable isotope labelled internal standard (SIL-IS) to correct matrix effect. However, to our knowledge only 2,2'-dichlorobisphenol A is commercially 271 272 available and other deuterated compounds should be custom synthetized which represents a 273 significant cost. Moreover, our data demonstrates that deuterated DCBPA enables to correct 274 matrix effect for other chlorine derivatives. Furthermore, IS-corrected matrix effects were comparable in breast and prostate adipose tissues, demonstrating that this method could 275 reliably assay BPA and Clx-BPA in both tissues. 276

BPA is ubiquitous in the environment, therefore specific precautions were implemented to avoid contamination with target analytes that may cause misleading results: tissue samples were collected and stored in glass vials, ultrapure solvents were used throughout the study and all reagents and material were tested for the absence of target analytes

Calibration standards and QCs were prepared using a pool of blank matrix as recommended by bioanalytical validation guidelines. However, target analytes may be present in blank matrices due to basal contamination of individuals since these compounds may be ubiquitously found in the environment and consequently in an individual's organisms. This was particularly the case for BPA. As a result, the mean signal corresponding to 5 blank samples determined for each series was systematically subtracted from signal of standard and QC samples to improve the accuracy of BPA in the low range.

Our LOQ for BPA was validated using QCs and was in the same range as the one obtained by Reeves et al. [30] (0.38 ng/g). No LOQ was defined in the paper by Artacho-Cordon et al. but the lowest QC tested for BPA was 2 ng/g [28]. No LOQ was validated for Clx-BPA in the paper by Fernandez et al. [31] but values defined as LOD were high (0.5 to 3.0 ng/mL) in comparison to our LOQ. The sensitivity of our assay was adequate to detect and quantify BPA and Clx-BPA in real samples. Bisphenol A was systematically quantified with BPA concentration in the 0.238-1.745 ng/mL range (1.19 - 8.73 ng/g). These concentrations were
of the same order of magnitude as those measured in adipose tissue originating in various
sources [27,29,31] but higher than those found by Reeves et al. in breast adipose tissue [30].
Our method enables to identify different patterns of exposure to Clx-BPA, with some patients
being exposed whereas others are not. Concentrations of Clx-BPA in our samples were lower
than those found by Fernandez et al. [31] who measured median concentrations equal to 3.14,
7.77 and 0.74 ng/g for CBPA, DCBPA and TCBPA , respectively.

301 As suggested by others [28], adipose tissue may be a suitable biological matrix for assessment of mid to long-term exposure to environmental pollutants, especially non-persistent 302 compounds presenting short half-lives. Collection of urine is undoubtly more convenient and 303 non-invasive but it may not reflect exposure over a prolonged period when half-life of target 304 305 compound is short [37]. This is the case for bisphenol A of which the half-life is about 5 hours whereas half-lives for Clx-BPA remain unknown. BPA and Clx-BPA are lipophilic compounds 306 with logP values ranging from 3.8 to 6.4 for BPA and TTCBPA, respectively [2], thereby 307 suggesting that they may accumulate in adipose tissue. 308

This method could also be used to assess the relationships between exposure to BPA and 309 Clx-BPA and local tumorigenesis effects of these endocrine-disrupting compounds. Indeed, 310 adipose tissue is important in tumor microenvironments and is now recognized as a major 311 factor in the development, growth and progression of cancer [38]. Adipose tissue inflammation 312 may be a key process in promotion of cancer [39] and it has been shown in ex vivo experiments 313 that low doses of BPA may promote secretion of inflammatory adipokines such as interleukin-314 6 and tumor necrosis factor a [40]. Obesity is also considered as a risk factor for breast adipose 315 tissue inflammation suggesting an indirect effect of obesogenic pollutants [15,41]. Moreover, 316 it has been shown in vitro that BPA and TTCBPA presented an affinity to the estrogen receptor 317 and expressed a proliferative potential in the human breast cancer cell line MCF-7 [18]. Finally, 318 319 as a biomarker of long-term exposure, this method could be applied in epidemiological studies.

321 5. **Conclusion**

In conclusion, this method enables the reliable determination of BPA and Clx-BPA in adipose tissue and could help to assess long-term exposure to these compounds as well as their possible associations with hormone-dependent cancers. For instance, this method will be applied to a study aiming to compare BPA and Clx-BPA exposure in women undergoing breast surgery for different breast lesion classes (benign, epithelial atypia and confirmed breast cancer).

328

330 Acknowledgments

- 331 We wish to thank Jeffrey Arsham for his highly helpful reading of our original text.
- 332 Contributors:
- 333 Conception or design of the work: NV, SR, CN, SR, VM, AD
- 334 Data collection: NV, GC, JR, CN, TC
- 335 Data analysis and interpretation: NV, GC, JR, AD
- 336 Drafting the article: NV, GC
- 337 Critical revision of the article: NV, AD, SR, VM, CN, TC
- 338 Final approval of the version to be published: NV; GC, JR, SR, CN, TC, SR, VM, AD
- 339

340 **Funding**:

341 Part of this research was funded by Roche Pharmaceuticals.

343 References

- L.N. Vandenberg, Low-dose effects of hormones and endocrine disruptors, Vitam.
 Horm. 94 (2014) 129–165. doi:10.1016/B978-0-12-800095-3.00005-5.
- M. Doumas, S. Rouillon, N. Venisse, C. Nadeau, P. Pierre Eugene, A. Farce, P.
 Chavatte, A. Dupuis, V. Migeot, P. Carato, Chlorinated and brominated bisphenol A
 derivatives: Synthesis, characterization and determination in water samples,
 Chemosphere. 213 (2018) 434–442. doi:10.1016/j.chemosphere.2018.09.061.
- [3] Z. Fan, J. Hu, W. An, M. Yang, Detection and occurrence of chlorinated byproducts of
 bisphenol a, nonylphenol, and estrogens in drinking water of china: comparison to the
 parent compounds, Environ. Sci. Technol. 47 (2013) 10841–10850.
- 353 doi:10.1021/es401504a.
- S. Song, M. Song, L. Zeng, T. Wang, R. Liu, T. Ruan, G. Jiang, Occurrence and profiles
 of bisphenol analogues in municipal sewage sludge in China, Environ. Pollut. 186
 (2014) 14–19. doi:10.1016/j.envpol.2013.11.023.
- H. Gallart-Ayala, E. Moyano, M.T. Galceran, On-line solid phase extraction fast liquid
 chromatography-tandem mass spectrometry for the analysis of bisphenol A and its
 chlorinated derivatives in water samples, J Chromatogr A. 1217 (2010) 3511–3518.
 doi:10.1016/j.chroma.2010.03.028.
- [6] C. Liao, K. Kannan, Determination of free and conjugated forms of bisphenol A in
 human urine and serum by liquid chromatography-tandem mass spectrometry, Environ.
 Sci. Technol. 46 (2012) 5003–5009. doi:10.1021/es300115a.
- N. Venisse, C. Grignon, B. Brunet, S. Thévenot, A. Bacle, V. Migeot, A. Dupuis,
 Reliable quantification of bisphenol A and its chlorinated derivatives in human urine
 using UPLC-MS/MS method, Talanta. 125 (2014) 284–292.
 doi:10.1016/j.talanta.2014.02.064.
- 368 [8] C. Grignon, N. Venisse, S. Rouillon, B. Brunet, A. Bacle, S. Thevenot, V. Migeot, A.
 369 Dupuis, Ultrasensitive determination of bisphenol A and its chlorinated derivatives in 370 urine using a high-throughput UPLC-MS/MS method, Anal Bioanal Chem. 408 (2016) 371 2255–2263. doi:10.1007/s00216-015-9288-8.
- I. Jiménez-Díaz, A. Zafra-Gómez, O. Ballesteros, N. Navea, A. Navalón, M.F.
 Fernández, N. Olea, J.L. Vílchez, Determination of Bisphenol A and its chlorinated
 derivatives in placental tissue samples by liquid chromatography-tandem mass
 spectrometry, J. Chromatogr. B Analyt. Technol. Biomed. Life Sci. 878 (2010) 3363–
 3369. doi:10.1016/j.jchromb.2010.10.021.
- A. Cariot, A. Dupuis, M. Albouy-Llaty, B. Legube, S. Rabouan, V. Migeot, Reliable
 quantification of bisphenol A and its chlorinated derivatives in human breast milk using
 UPLC-MS/MS method, Talanta. 100 (2012) 175–182.
- 380 doi:10.1016/j.talanta.2012.08.034.
- [11] V. Migeot, A. Dupuis, A. Cariot, M. Albouy-Llaty, F. Pierre, S. Rabouan, Bisphenol a and its chlorinated derivatives in human colostrum, Environ. Sci. Technol. 47 (2013) 13791–13797. doi:10.1021/es403071a.
- Y. Mutou, Y. Ibuki, Y. Terao, S. Kojima, R. Goto, Change of estrogenic activity and
 release of chloride ion in chlorinated bisphenol a after exposure to ultraviolet B, Biol.
 Pharm. Bull. 29 (2006) 2116–2119.
- [13] A. Riu, M. Grimaldi, A. le Maire, G. Bey, K. Phillips, A. Boulahtouf, E. Perdu, D. Zalko,
 W. Bourguet, P. Balaguer, Peroxisome proliferator-activated receptor γ is a target for
 halogenated analogs of bisphenol A, Environ. Health Perspect. 119 (2011) 1227–1232.
 doi:10.1289/ehp.1003328.
- [14] S.S. Andra, H. Kalyvas, X.D. Andrianou, P. Charisiadis, C.A. Christophi, K.C. Makris,
 Preliminary evidence of the association between monochlorinated bisphenol A
 exposure and type II diabetes mellitus: A pilot study, J Environ Sci Health A Tox Hazard
- 394 Subst Environ Eng. 50 (2015) 243–259. doi:10.1080/10934529.2015.981111.

- [15] S.S. Andra, K.C. Makris, Association between urinary levels of bisphenol A and its 395 396 monochlorinated derivative and obesity, J Environ Sci Health A Tox Hazard Subst Environ Eng. 50 (2015) 1169-1179. doi:10.1080/10934529.2015.1047674. 397
- [16] A. Shafei, M.M. Ramzy, A.I. Hegazy, A.K. Husseny, U.G. El-Hadary, M.M. Taha, A.A. 398 399 Mosa, The molecular mechanisms of action of the endocrine disrupting chemical 400 bisphenol A in the development of cancer, Gene. 647 (2018) 235-243. 401 doi:10.1016/j.gene.2018.01.016.
- [17] M. Di Donato, G. Cernera, P. Giovannelli, G. Galasso, A. Bilancio, A. Migliaccio, G. 402 Castoria, Recent advances on bisphenol-A and endocrine disruptor effects on human 403 404 prostate cancer, Mol. Cell. Endocrinol. 457 (2017) 35-42. 405
- doi:10.1016/j.mce.2017.02.045.
- [18] C.M. Olsen, E.T.M. Meussen-Elholm, M. Samuelsen, J.A. Holme, J.K. Hongslo, Effects 406 of the environmental oestrogens bisphenol A, tetrachlorobisphenol A, 407 408 tetrabromobisphenol A, 4-hydroxybiphenyl and 4,4'-dihydroxybiphenyl on oestrogen receptor binding, cell proliferation and regulation of oestrogen sensitive proteins in the 409 410 human breast cancer cell line MCF-7, Pharmacol. Toxicol. 92 (2003) 180-188.
- [19] M. Hoffmann, E. Fiedor, A. Ptak, Bisphenol A and its derivatives tetrabromobisphenol A 411 412 and tetrachlorobisphenol A induce apelin expression and secretion in ovarian cancer cells through a peroxisome proliferator-activated receptor gamma-dependent 413 mechanism, Toxicol. Lett. 269 (2017) 15–22. doi:10.1016/j.toxlet.2017.01.006. 414
- [20] S. Yousefnia, S. Momenzadeh, F. Seyed Forootan, K. Ghaedi, M.H. Nasr Esfahani, The 415 influence of peroxisome proliferator-activated receptor y (PPARy) ligands on cancer cell 416 tumorigenicity, Gene. 649 (2018) 14-22. doi:10.1016/j.gene.2018.01.018. 417
- [21] N.M. Iyengar, K.A. Brown, X.K. Zhou, A. Gucalp, K. Subbaramaiah, D.D. Giri, H. Zahid, 418 P. Bhardwaj, N.K. Wendel, D.J. Falcone, H. Wang, S. Williams, M. Pollak, M. Morrow, 419 420 C.A. Hudis, A.J. Dannenberg, Metabolic Obesity, Adipose Inflammation and Elevated Breast Aromatase in Women with Normal Body Mass Index, Cancer Prev Res (Phila). 421 10 (2017) 235-243. doi:10.1158/1940-6207.CAPR-16-0314. 422
- 423 [22] H. Song, J. Park, P.T.C. Bui, K. Choi, M.C. Gye, Y.-C. Hong, J.H. Kim, Y.J. Lee, Bisphenol A induces COX-2 through the mitogen-activated protein kinase pathway and 424 is associated with levels of inflammation-related markers in elderly populations. Environ. 425 Res. 158 (2017) 490-498. doi:10.1016/j.envres.2017.07.005. 426
- [23] M. Nieuwenhuijsen, D. Paustenbach, R. Duarte-Davidson, New developments in 427 428 exposure assessment: the impact on the practice of health risk assessment and epidemiological studies, Environ Int. 32 (2006) 996–1009. 429 doi:10.1016/j.envint.2006.06.015. 430
- [24] A. Zidek, K. Macey, L. MacKinnon, M. Patel, D. Poddalgoda, Y. Zhang, A review of 431 human biomonitoring data used in regulatory risk assessment under Canada's 432 Chemicals Management Program, Int J Hyg Environ Health. 220 (2017) 167–178. 433 doi:10.1016/j.ijheh.2016.10.007. 434
- [25] J.G. Teeguarden, N.C. Twaddle, M.I. Churchwell, X. Yang, J.W. Fisher, L.M. Seryak, 435 D.R. Doerge, 24-hour human urine and serum profiles of bisphenol A: Evidence against 436 sublingual absorption following ingestion in soup, Toxicol. Appl. Pharmacol. 288 (2015) 437 438 131-142. doi:10.1016/j.taap.2015.01.009.
- [26] J.E. Orban, J.S. Stanley, J.G. Schwemberger, J.C. Remmers, Dioxins and 439 440 dibenzofurans in adipose tissue of the general US population and selected 441 subpopulations, Am J Public Health. 84 (1994) 439-445.
- 442 [27] L. Wang, A.G. Asimakopoulos, K. Kannan, Accumulation of 19 environmental phenolic and xenobiotic heterocyclic aromatic compounds in human adipose tissue. Environ Int. 443 78 (2015) 45-50. doi:10.1016/j.envint.2015.02.015. 444
- [28] F. Artacho-Cordón, J.P. Arrebola, O. Nielsen, P. Hernández, N.E. Skakkebaek, M.F. 445 Fernández, A.M. Andersson, N. Olea, H. Frederiksen, Assumed non-persistent 446 environmental chemicals in human adipose tissue: matrix stability and correlation with 447 levels measured in urine and serum, Environ. Res. 156 (2017) 120-127. 448
- doi:10.1016/j.envres.2017.03.030. 449

- 450 [29] T. Geens, H. Neels, A. Covaci, Distribution of bisphenol-A, triclosan and n-nonylphenol
 451 in human adipose tissue, liver and brain, Chemosphere. 87 (2012) 796–802.
 452 doi:10.1016/j.chemosphere.2012.01.002.
- [30] K.W. Reeves, S. Schneider, J. Xue, K. Kannan, H. Mason, M. Johnson, G. MakariJudson, M.D. Santana, Bisphenol-A in breast adipose tissue of breast cancer cases and
 controls, Environ. Res. 167 (2018) 735–738. doi:10.1016/j.envres.2018.08.033.
- [31] M.F. Fernandez, J.P. Arrebola, J. Taoufiki, A. Navalón, O. Ballesteros, R. Pulgar, J.L.
 Vilchez, N. Olea, Bisphenol-A and chlorinated derivatives in adipose tissue of women, Reprod. Toxicol. 24 (2007) 259–264. doi:10.1016/j.reprotox.2007.06.007.
- 459 [32] European Medicines Agency Science Medicines Agency, Guidance on bioanalytical
 460 method validation, (2012) 23.
- 461 [33] P.J. Rudzki, E. Gniazdowska, K. Buś-Kwaśnik, Quantitative evaluation of the matrix
 462 effect in bioanalytical methods based on LC-MS: A comparison of two approaches, J
 463 Pharm Biomed Anal. 155 (2018) 314–319. doi:10.1016/j.jpba.2018.03.052.
- [34] R. Cariou, J.-P. Antignac, P. Marchand, A. Berrebi, D. Zalko, F. Andre, B. Le Bizec,
 New multiresidue analytical method dedicated to trace level measurement of
 brominated flame retardants in human biological matrices, J Chromatogr A. 1100 (2005)
 144–152. doi:10.1016/j.chroma.2005.09.040.
- 468 [35] C.V. Garcia, A. Gotah, Application of QuEChERS for Determining Xenobiotics in Foods
 469 of Animal Origin, J Anal Methods Chem. 2017 (2017) 2603067.
 470 doi:10.1155/2017/2603067.
- 471 [36] K. Madej, T.K. Kalenik, W. Piekoszewski, Sample preparation and determination of
 472 pesticides in fat-containing foods, Food Chem. 269 (2018) 527–541.
 473 doi:10.1016/j.foodchem.2018.07.007.
- [37] M. Casas, X. Basagaña, A.K. Sakhi, L.S. Haug, C. Philippat, B. Granum, C.B. 474 475 Manzano-Salgado, C. Brochot, F. Zeman, J. de Bont, S. Andrusaityte, L. Chatzi, D. Donaire-Gonzalez, L. Giorgis-Allemand, J.R. Gonzalez, E. Gracia-Lavedan, R. 476 Grazuleviciene, M. Kampouri, S. Lyon-Caen, P. Pañella, I. Petraviciene, O. Robinson, 477 478 J. Urquiza, M. Vafeiadi, C. Vernet, D. Waiblinger, J. Wright, C. Thomsen, R. Slama, M. Vrijheid, Variability of urinary concentrations of non-persistent chemicals in pregnant 479 480 women and school-aged children, Environ Int. 121 (2018) 561-573. doi:10.1016/j.envint.2018.09.046. 481
- [38] N.M. Iyengar, A. Gucalp, A.J. Dannenberg, C.A. Hudis, Obesity and Cancer
 Mechanisms: Tumor Microenvironment and Inflammation, J. Clin. Oncol. 34 (2016)
 484 4270–4276. doi:10.1200/JCO.2016.67.4283.
- [39] N.M. Iyengar, C.A. Hudis, A.J. Dannenberg, Obesity and cancer: local and systemic
 mechanisms, Annu. Rev. Med. 66 (2015) 297–309. doi:10.1146/annurev-med-050913022228.
- [40] M. Giulivo, M. Lopez de Alda, E. Capri, D. Barceló, Human exposure to endocrine
 disrupting compounds: Their role in reproductive systems, metabolic syndrome and
 breast cancer. A review, Environ. Res. 151 (2016) 251–264.
 doi:10.1016/j.envres.2016.07.011.
- 492 [41] S. Legeay, Ś. Faure, Is bisphenol A an environmental obesogen?, Fundam Clin
 493 Pharmacol. 31 (2017) 594–609. doi:10.1111/fcp.12300.
- 494

1 Figure captions:

2 Figure 1: Chromatogram of a blank adipose tissue sample. BPA: bisphenol A, MCBPA: monochlorobisphenol A; DCBPA: dichlorobisphenol A;

3 TCBPA: trichlorobisphenol A; TTCBPA: tetrachlorobisphenol A.

Figure 2: Chromatogram corresponding to the lower limit of quantification (LOQ): lowest standard sample prepared in adipose tissue sample
and spiked at 0.125 and 0.0125 ng/mL of BPA and Clx-BPA, respectively. BPA: bisphenol A, MCBPA: monochlorobisphenol A; DCBPA:
dichlorobisphenol A; TCBPA: trichlorobisphenol A; TTCBPA: tetrachlorobisphenol A.

Figure 3: Chromatogram of the adipose tissue sample from individual PN. BPA and TTCBPA were quantified while MCBPA, DCBPA and
 TCBPA were not detected (<LOD) (also see Table 5). BPA: bisphenol A, MCBPA: monochlorobisphenol A; DCBPA: dichlorobisphenol A;
 TCBPA: trichlorobisphenol A; TTCBPA: tetrachlorobisphenol A.

Figure 4: Chromatogram of the adipose tissue sample from individual LC. BPA and TTCBPA were quantified while MCBPA, DCBPA and TCBPA were detected but not quantified (<LOQ) (also see Table 5). BPA: bisphenol A, MCBPA: monochlorobisphenol A; DCBPA: dichlorobisphenol A; TCBPA: trichlorobisphenol A; TTCBPA: tetrachlorobisphenol A.

Figure 1: Chromatogram of a blank adipose tissue sample. BPA: bisphenol A, MCBPA: monochlorobisphenol A; DCBPA: dichlorobisphenol A;
 TCBPA: trichlorobisphenol A; TTCBPA: tetrachlorobisphenol A.



Figure 2: Chromatogram corresponding to the lower limit of quantification (LOQ): lowest standard sample prepared in adipose tissue sample and spiked at 0.125 and 0.0125 ng/mL of BPA and Clx-BPA, respectively. BPA: bisphenol A, MCBPA: monochlorobisphenol A; DCBPA: dichlorobisphenol A; TCBPA: trichlorobisphenol A; TTCBPA: tetrachlorobisphenol A.



- 24 25
- 26
- 27

Figure 3: Chromatogram of the adipose tissue sample from individual PN. BPA and TTCBPA were quantified while MCBPA, DCBPA and TCBPA were not detected (<LOD) (also see Table 5). BPA: bisphenol A, MCBPA: monochlorobisphenol A; DCBPA: dichlorobisphenol A; TCBPA: trichlorobisphenol A; TTCBPA: tetrachlorobisphenol A.



31 32

33

Figure 4: Chromatogram of the adipose tissue sample from individual LC. BPA and TTCBPA were quantified while MCBPA, DCBPA and TCBPA were detected but not quantified (<LOQ) (also see Table 5). BPA: bisphenol A, MCBPA: monochlorobisphenol A; DCBPA: dichlorobisphenol A; TCBPA: trichlorobisphenol A; TTCBPA: tetrachlorobisphenol A.



35

1 Tables

Table 1: Descriptive table of retention times, precursor ions, fragment ions and MS/MS
optimization conditions (declustering potential, DP; entrance potential, EP; collision energy,
CE; collision cell exit potential, CXP) for all analytes. For CE and CXP, first and second
values are for quantification and confirmation ions, respectively

Analyte	Retention time (min)	Precursor ion (m/z)	Quantification ion (m/z)	Confirmation ion (m/z)	DP (V)	EP (V)	CE (V)	CXP (V)
BPA	5.62	227.0	211.8	132.8	-65	-10	-24 /-34	-19 /-11
BPA-d16	5.55	241.1	222.9	141.9	-90	-10	-28 /-34	-25 /-15
MCBPA	6.20	260.9	181.7	209.7	-120	-10	-38 /-30	-19 /-21
DCBPA	6.72	294.9	215.8	279.8	-75	-10	-40 /-30	-24 /-13
TCBPA	7.20	329.0	249.7	277.7	-85	-10	-42 /-34	-27 /-31
TTCBPA	7.64	364.8	313.7	285.6	-110	-10	-36 /-44	-33 /-29
DCBPA- d12	6.65	307.0	224.9	252.8	-120	-10	-42 /-34	-23 /-25

6

Table 2: Mean (+/- SD) slope, correlation coefficient (r²) and y-intercept of calibration curves
 prepared with breast adipose tissue (n=5).

Compound	BPA	MCBPA	DCBPA	ТСВРА	ТТСВРА
Slope (mean ± SD)	1.082 ± 0.148	1.514 ± 0.496	3.574 ± 1.148	3.321 ± 0.900	3.011 ± 1.203
Correlation coefficient r ² (mean ± SD)	0.999 ± 0.000	0.997 ± 0.003	0.997 ± 0.003	0.997 ± 0.003	0.997 ± 0.002
Y-intercept (mean ± SD)	0.487 ± 0.058	0.007 ± 0.004	0.003 ± 0.005	0.009 ± 0.011	0.023 ± 0.025

- Table 3: Intra and inter-day accuracy and precision for each analyte in breast adipose tissue at each level of QC (n=5)

Compound	Concentration of	Accuracy	y (%)	Precision (%)	
compound	QC (ng/mL)	Intra-day	Inter-day	Intra-day	Inter-day
BPA	0.125 (LOQ)	110	95	8	3
	1.0	106	101	6	4
	4.0	109	104	4	6
MCBPA	0.0125 (LOQ)	114	112	3	14
	0.10	97	102	10	8
	0.40	96	104	2	5
DCBPA	0.0125 (LOQ)	99	104	4	13
	0.10	96	102	5	6
	0.40	102	104	3	5
ТСВРА	0.0125 (LOQ)	99	101	4	7
	0.10	100	102	8	4
	0.40	101	102	3	5
ТТСВРА	0.0125 (LOQ)	85	109	8	7
	0.10	101	103	8	7
	0.40	106	108	6	13

16 Table 4: Determination of matrix effect, IS-corrected matrix effect and recovery in breast

¹⁷ adipose tissue at each level of QC (n = 3)

Compound	Concentration of QC (ng/mL)	Matrix effect (%)	IS-corrected matrix effect (%)	Recovery (%)
BPA	0.125	98	111	50
	1.0	77	83	22
	4.0	81	98	20
MCBPA	0.0125	80	100	19
	0.10	77	85	14
	0.40	69	83	15
DCBPA	0.0125	80	99	18
	0.10	94	98	15
	0.40	81	98	21
ТСВРА	0.0125	92	114	17
	0.10	100	104	16
	0.40	97	116	20
ТТСВРА	0.0125	100	126	17
	0.10	102	106	15
	0.40	94	112	17

18

20 Table 5: Determination of matrix effect, IS-corrected matrix effect and recovery in prostate

²¹ adipose tissue at each level of QC (n = 3)

Compound	Concentration of QC (ng/mL)	Matrix effect (%)	IS-corrected matrix effect (%)	Recovery (%)
BPA	0.125	61	79	40
	1.0	66	82	18
	4.0	72	86	18
MCBPA	0.0125	67	87	15
	0.10	62	85	14
	0.40	67	90	16
DCBPA	0.0125	74	95	14
	0.10	67	92	15
	0.40	72	97	16
ТСВРА	0.0125	68	88	19
	0.10	91	123	10
	0.40	66	87	15
ТТСВРА	0.0125	94	121	14
	0.10	93	113	9
	0.40	63	87	15

22

Table 6: Slope, correlation coefficient (r²) and y-intercept of calibration curves prepared with prostate adipose tissue.

Compound	BPA	MCBPA	DCBPA	ТСВРА	ТТСВРА
Slope	1.41	2.22	4.77	6.38	3.15
Correlation coefficient r ²	0.994	0.992	0.996	0.990	0.992
Y-intercept	0.480	0.002	0.004	-0.057	- 0.023

Table 7: BPA and Clx-BPA concentrations measured in 5 breast adipose tissue samples
obtained from women undergoing breast surgery. <LOQ: compound detected but not
quantified; <LOD: compound not detected.

	Endocrine disruptor concentrations (ng/mL)					
Patient ID	BPA	МСВРА	DCBPA	ТСВРА	ТТСВРА	
RS	1.433	<lod< td=""><td><lod< td=""><td><lod< td=""><td><lod< td=""></lod<></td></lod<></td></lod<></td></lod<>	<lod< td=""><td><lod< td=""><td><lod< td=""></lod<></td></lod<></td></lod<>	<lod< td=""><td><lod< td=""></lod<></td></lod<>	<lod< td=""></lod<>	
PN	0.788	<lod< td=""><td><lod< td=""><td><lod< td=""><td>0,022</td></lod<></td></lod<></td></lod<>	<lod< td=""><td><lod< td=""><td>0,022</td></lod<></td></lod<>	<lod< td=""><td>0,022</td></lod<>	0,022	
PC	1.510	0.046	<lod< td=""><td><lod< td=""><td><lod< td=""></lod<></td></lod<></td></lod<>	<lod< td=""><td><lod< td=""></lod<></td></lod<>	<lod< td=""></lod<>	
LC	0.238	<loq< td=""><td><loq< td=""><td><loq< td=""><td>0.026</td></loq<></td></loq<></td></loq<>	<loq< td=""><td><loq< td=""><td>0.026</td></loq<></td></loq<>	<loq< td=""><td>0.026</td></loq<>	0.026	
BE	1.745	0.013	<lod< td=""><td>0.013</td><td>0.014</td></lod<>	0.013	0.014	

