



**HAL**  
open science

## Designing local air pollution policies focusing on mobility and heating to avoid a targeted number of pollution-related deaths: Forward and backward approaches combining air pollution modeling, health impact assessment and cost-benefit analysis

Hélène Bouscasse, Stephan Gabet, Glen Kerneis, Ariane Provent, Camille Rieux, Nabil Ben Salem, Harry Dupont, Florence Troude, Sandrine Mathy, Rémy Slama

### ► To cite this version:

Hélène Bouscasse, Stephan Gabet, Glen Kerneis, Ariane Provent, Camille Rieux, et al.. Designing local air pollution policies focusing on mobility and heating to avoid a targeted number of pollution-related deaths: Forward and backward approaches combining air pollution modeling, health impact assessment and cost-benefit analysis. *Environment International*, 2022, 159, 10.1016/j.envint.2021.107030 . hal-03470856

**HAL Id: hal-03470856**

**<https://hal.science/hal-03470856v1>**

Submitted on 5 Jan 2024

**HAL** is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.



Distributed under a Creative Commons Attribution - NonCommercial - NoDerivatives 4.0 International License

## **Title**

Designing local air pollution policies focusing on mobility and heating to avoid a targeted number of pollution-related deaths: forward and backward approaches combining air pollution modeling, health impact assessment and cost-benefit analysis

## **Author names, academic degrees, and affiliations**

Hélène Bouscasse<sup>\*1</sup>, PhD, Stephan Gabet<sup>\*2</sup>, PhD, PharmD, Glen Kerneis<sup>3</sup>, MSc, Ariane Provent<sup>4</sup>, Msc, Camille Rieux<sup>4</sup>, PhD, Nabil Ben Salem<sup>4</sup>, PhD, Harry Dupont<sup>4</sup>, Florence Troude<sup>4</sup>, PhD, Sandrine Mathy<sup>3</sup>, PhD, Rémy Slama<sup>2</sup>, PhD.

<sup>1</sup> CESAER, Agrosup Dijon, INRAE, Bourgogne Franche-Comté Univ., Dijon, France;

<sup>2</sup> Univ. Grenoble Alpes, Inserm, CNRS, Team of Environmental Epidemiology applied to Reproduction and Respiratory Health, IAB, 38000 Grenoble, France;

<sup>3</sup> Univ. Grenoble Alpes, CNRS, INRAE, Grenoble INP, GAEL, Grenoble, France;

<sup>4</sup> Atmo Auvergne-Rhône-Alpes, Grenoble, France.

\* Co-first authorship.

## **Co-corresponding authors**

Sandrine Mathy

GAEL - CS 40700

38058 Grenoble CEDEX 9, France

E-mail: [sandrine.mathy@univ-grenoble-alpes.fr](mailto:sandrine.mathy@univ-grenoble-alpes.fr)

Rémy Slama

IAB, Team of Environmental Epidemiology

Inserm, CNRS and University Grenoble-Alpes

38700 La Tronche, France

E-mail: Remy.slama@univ-grenoble-alpes.fr

## **Acknowledgments**

We thank Atmo Auvergne-Rhône-Alpes, and in particular Marie-Blanche Personnaz, Steve Micallef, and Cécile Fontagné, for developing MECANO chain-tool and for providing air pollution and population density data, the Epidemiology center on causes of death (CépiDC, Inserm) for providing mortality data, and the Urban planning agency of the Grenoble region (AURG) for providing population structure and traffic-related data. We also thank Grenoble Alpes Metropole for its support.

## **Funding**

The present study is part of the QAMECS and MobilAir research projects, funded by the French Environment and Energy Management Agency (ADEME) and by the IDEX Program of Grenoble-Alpes University ComUE in the framework of the “Investing for the future” Program of the French National Research Agency (ANR-15-IDEX-02). It was initiated as part of the “Breathable city in 5 years” Program of the French Ministry of Ecological and Solidarity Transition for Grenoble conurbation.

## **Author contributions**

**Hélène Bouscasse:** Conceptualization, Formal analysis, Writing - Original Draft; **Stephan Gabet:** Conceptualization, Formal analysis, Writing - Original Draft; **Glen Kerneis:** Formal analysis, Writing - Original Draft; **Ariane Provent:** Methodology; **Camille Rieux:** Methodology; **Nabil Ben Salem:** Methodology; **Harry Dupont:** Methodology; **Florence**

**Troude:** Methodology, Writing -Original Draft; **Sandrine Mathy:** Conceptualization, Writing  
- Review & Editing; **Rémy Slama:** Conceptualization, Writing - Review & Editing.

## **Abstract**

### **Context**

Policies aiming at decreasing air pollutants (e.g., fine particulate matter, PM<sub>2.5</sub>) are often designed without targeting an explicit health benefit nor carrying out cost-benefit analyses.

### **Methods**

We developed a transdisciplinary backward and forward approach at the conurbation level: from health objectives set by local decision-makers, we estimated which reductions in PM<sub>2.5</sub> exposures and emissions would allow to reach them, and identified urban policies leading to these reductions (backward approach). We finally conducted health impact and cost-benefit analyses of these policies (forward approach). The policies were related to the most emitting sectors in the considered area (Grenoble, France), wood heating and transport sectors. The forward approach also considered the health impact and co-benefits of these policies related to changes in physical activity and CO<sub>2</sub> emissions.

### **Findings**

Decision-makers set three health targets, corresponding to decreases by 33% to 67% in PM<sub>2.5</sub>-attributable mortality in 2030, compared to 2016. A decrease by 42% in PM<sub>2.5</sub> exposure (from 13.9 µg/m<sup>3</sup>) was required to reach the decrease by 67% in PM<sub>2.5</sub>-attributable mortality. For each Euro invested, the total benefit was about 30€ for policies focusing on wood heating, and 1 to 68€ for traffic policies. Acting on a single sector was not enough to attain a 67% decrease in PM<sub>2.5</sub>-attributable mortality. This target could be achieved by replacing all inefficient wood heating equipment by low-emission pellet stoves and reducing by 36% the traffic of private motorized vehicles. This would require to increase the share of active modes (walking, biking...), inducing increases in physical activity and additional health benefits beyond the initial target. Annual net benefits were between €484 and €629 per capita for policies with report on active modes, compared to between €162 and €270 without.

### **Conclusions**

Urban policies strongly reducing air pollution-attributable mortality can be identified by our approach. Such policies can be cost-efficient.

## **Keywords**

Dispersion model; economic analysis; fine particulate matter (PM<sub>2.5</sub>); health impact assessment; transportation modal shift.

## **Policy points**

- A 33% reduction in PM<sub>2.5</sub>-attributable mortality by 2030 (compared to 2016) is achievable in Grenoble through simple public policies. It would lead to a benefit/cost ratio equal to 39 and generate an annual per capita net benefit of 162€.
- Reaching the most ambitious health target (-67%) would require to replace all inefficient wood heating facilities and to replace a significant share of private cars travels by active transportation modes. It would lead to a benefit/cost ratio between 4.2 and 6 and generate an annual per capita net benefit between 484€ and 629€.
- A large part of the benefits related to air pollution control policies are not directly linked to air quality increases but to the increase in physical activity due to the modal shift from cars to active modes of transport.

## **Abbreviations**

$\Delta_{NAC}$ : difference in the number of attributable cases

LEZ: Low Emission Zones

PM: particulate matter

PM<sub>2.5</sub>: fine particulate matter

RR: risk ratio

VoT: Value of Time

WHO: World Health Organization

## **Number of Tables/Illustrations**

5 tables and 7 figures (in color).

## **Supporting information**

5 tables and 1 figure.

## 1 Introduction

Atmospheric pollution has major and proven effects on mortality and morbidity, mainly through cardiovascular and respiratory health.<sup>1-3</sup> With nine out of ten people in urban areas worldwide in 2019 being exposed to levels of fine particulate matter (PM<sub>2.5</sub>) from outdoor air pollution above the 2005 World Health Organization (WHO) guideline value (annual average PM<sub>2.5</sub> concentration of 10 µg/m<sup>3</sup>),<sup>4</sup> and even more above the current World Health Organization (WHO) guideline value of 5 µg/m<sup>3</sup>, air pollution is a major public health issue. Exposure to ambient PM<sub>2.5</sub> annually causes an estimated 4.2 million (Uncertainty Interval, UI, 3.7, 4.8) deaths worldwide<sup>5</sup> (based on 2015 and 2018 exposure levels, and a counterfactual level of 2.4-5.9 µg/m<sup>3</sup> corresponding to the theoretical minimum risk exposure level) and 417,000 annual death cases in Europe<sup>6</sup> (based on 2015 and 2018 exposure levels, and a counterfactual level of 0 µg/m<sup>3</sup>); in France, about 40,000 people would prematurely die each year due to exposure to ambient anthropogenic PM<sub>2.5</sub> exposure (based on 2016-2019 exposure levels and a counterfactual level of 5 µg/m<sup>3</sup>).<sup>7</sup> The health impacts of air pollution cost an estimated \$1.7 trillion in OECD countries in 2010, while specifically in France the annual direct and indirect economic costs amount to €100 billion.<sup>8</sup>

Air pollution thus represents one of the biggest controllable environmental risk to health. Public policies are one family of options to reduce the health and economic impacts of air pollution. These policies should target the main sources of atmospheric pollution. Such public policies can combine different instruments: legal instruments (e.g., norms on emissions for vehicles or other sources or on concentrations of specific pollutants, low emission zones), price instruments (e.g., economic incentives to increase the share of low emitting equipment, of non-motorized mobility, of public transports, which emit less per capita, to improve insulation of homes), technological solutions (e.g., development of low emission devices to replace polluting ones), urban planning (e.g., transportation infrastructures), and public awareness. Cities in Europe and



elsewhere have undertaken measures to limit air pollution emissions, especially from transportation and heating sources. In cities where ambitious programs were implemented, environmental evaluations have documented decreases. For example, ban of coal sales in Dublin in 1990 led to the reductions in black smoke air concentrations by 70% and in non-trauma death rates by 5.7% (95% CI, 4, 7), when comparing concentration levels over the 72 months before and after the measure came into force.<sup>9</sup> Decreases in PM<sub>10</sub> by as much as 50% were reported in Tokyo between 2001 and 2010,<sup>10</sup> and between 5% and 13% in Germany following the adoption of low emission zones.<sup>11,12</sup> Until now, the adoption of such public policies did not allow to quickly alleviate the above-mentioned societal burden, in particular in France.<sup>13,14</sup>

One possible reason is that policies to reduce pollution are often designed without explicit consideration of a targeted health impact. In some cases, they rely on an *ex ante* environmental evaluation ignoring any health consequence. Such an evaluation is even seldom done a posteriori: a review of LEZs in Europe<sup>15</sup> showed that only in rare cases was the environmental evaluation supplemented with a health impact assessment (HIA).<sup>16</sup> In contrast, in the USA, the Clean air act requires the Environmental Protection Agency to periodically estimate the impact of the Clean air act on public health, economy, and environment.

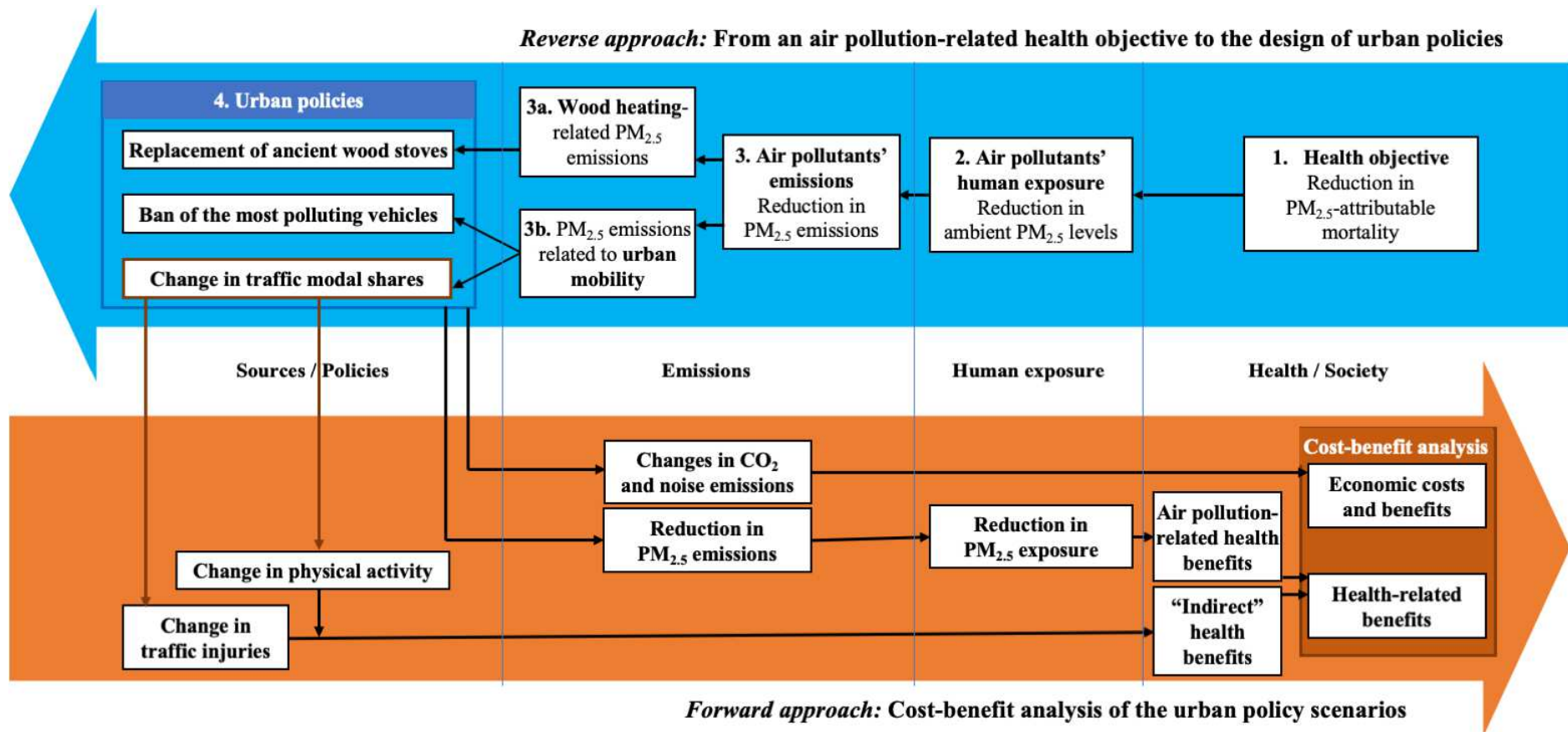
Starting from a target health impact to appropriately dimension urban policies would be a logical and important change of approach for many areas of the world.

Scientifically, this raises the challenge to develop a reverse or backward approach consisting of starting from a target formulated in terms of improvement of health (e.g., reducing by 50% the PM<sub>2.5</sub>-related mortality), and subsequently identifying urban policies allowing to reach such a health target. Specific health co-benefits can arise from policies aiming primarily to reduce air pollution, such as those related to changes in physical activity linked with the promotion of mode shift from cars to e.g., bicycles, and their assessment would be required to

fully appraise the health improvements than can be expected from such policies.<sup>17,18</sup> Traffic is also a major source of noise, whose exposure has been associated to heart disease and stress<sup>19,20</sup> as well as decreases in well-being<sup>21</sup>, as well as a source of greenhouse gas; changes in these factors, which also influence health,<sup>22,23</sup> need to be considered.

If the cost of the targeted policies is estimated, cost-benefit analyses can in addition be conducted. In the USA, analyses of benefits and costs are planned by the Clean Air Act law, and the analyses conducted so far indicate that the benefits exceed costs by a factor of 3 to 90, with a central estimate equal to 30.<sup>24</sup> No such figure is available in France. Making them available in the French and European contexts, where atmospheric pollution limits are much higher than in the USA (with a regulatory limit of 25  $\mu\text{g}/\text{m}^3$  for  $\text{PM}_{2.5}$  yearly mean concentration in Europe, compared to 12  $\mu\text{g}/\text{m}^3$  in the USA), would be very relevant for citizens and decision makers. Such estimates should also be provided at the city level, to consider accurately specific local sources and because air pollution mitigation policies are often implemented at this level.

We therefore developed a backward and forward approach at the city level, starting from predetermined health objectives and identifying the corresponding change in the population's exposure to air pollution (focusing on  $\text{PM}_{2.5}$ ), then the corresponding  $\text{PM}_{2.5}$  emission reductions, and finally defining urban policies compatible with these health objectives (Figure 1, “reverse approach”). We then carried out a cost-benefit analysis of these policies, including climate change and various health co-benefits (Figure 1, forward cost-benefit analysis). The present study was led in Grenoble urban area, a French middle size city with good characterization of air pollution impacts,<sup>25</sup> as well as good knowledge on the main air pollution sources of  $\text{PM}_{2.5}$ , i.e., wood heating and road traffic.



**Figure 1.** Overview of the study, showing the reverse approach from an air pollution-related health improvement objective to the identification of the corresponding urban policies as well as the cost-benefice analysis of the designed urban policy scenarios.

## 1 **2 Material and methods**

### 2 **2.1. A reverse approach to develop scenarios for achieving specific air pollution-related** 3 **health objectives**

4 The main steps of our backward and forward approaches are described in Figure 1. The details  
5 are provided below.

#### 6 2.1.1 From health targets to changes in air pollution exposure

7 Grenoble conurbation belongs to Auvergne-Rhône-Alpes region, which is in the South-eastern  
8 part of France. Grenoble conurbation has a surface of 541 km<sup>2</sup>, includes 49 municipalities, and,  
9 with 444,000 inhabitants in 2014, is the 11<sup>th</sup> French largest conurbation in terms of population.<sup>26</sup>  
10 In Grenoble, emissions from wood heating account for an annual average of 63% of PM<sub>2.5</sub>  
11 emissions,<sup>27</sup> a proportion higher during cold winter days. These emissions come from  
12 inefficient equipment such as open fireplaces and stoves with low energy efficiency. This is  
13 why Grenoble conurbation has been offering a subsidy for the purchase of efficient wood stoves  
14 for the replacement of these installations since 2016. To obtain the wood stove replacement  
15 premium, households have to own a fairly old device, destroy it and install a new one with an  
16 efficiency label. Road transport accounts for 17% of yearly PM<sub>2.5</sub> emissions.<sup>27</sup> In 2017, the  
17 Grenoble conurbation set up a low emission zone (LEZ) for the goods transport vehicles; only  
18 vehicles without a “CritAir” label (the least efficient ones) were concerned. This restriction,  
19 which only applied to Grenoble center when it was introduced, has been extended since 2019  
20 to 10 voluntary municipalities bordering Grenoble, and since 2020 to 27 municipalities among  
21 the 49 included in the Grenoble conurbation.

22 Based on exposure levels from 2015-2017, we previously estimated that the population  
23 density-weighted long-term average exposure to anthropogenic PM<sub>2.5</sub> in Grenoble conurbation  
24 corresponds to 145 (95% Confidence Interval, CI, 90–199) attributable premature death cases

25 each year.<sup>25</sup> We also previously showed that achieving a decrease by 33%, 50%, and 67% of  
26 the premature mortality attributable to anthropogenic PM<sub>2.5</sub> exposure by 2030, compared to the  
27 2015-2017 average (baseline), would require to reach the yearly average PM<sub>2.5</sub> exposure of 11.0,  
28 9.6, and 8.0 µg/m<sup>3</sup>, respectively, starting from the baseline population-weighted yearly level of  
29 13.9 µg/m<sup>3</sup>.<sup>25</sup> These health objectives were defined by the elected officials of Grenoble  
30 metropolitan area.

31 The study area population's yearly average exposure to PM<sub>2.5</sub> was assessed coupling the  
32 modeled PM<sub>2.5</sub> concentrations with information on population density, as previously  
33 described.<sup>25</sup> Atmospheric pollutant concentrations were modeled at a 10-meter spatial  
34 resolution combining two models: CHIMERE, a mesoscale chemistry-transport model,<sup>28</sup> and  
35 SIRANE, a proximity scale air pollutant-dispersion model.<sup>29</sup> Population density by sex and  
36 five-year age classes was available for 2014 at the same spatial resolution (10-meter), based on  
37 data from the National Institute of Statistics and Economic Studies<sup>26</sup> and the National Institute  
38 of Geographic and Forestry Information.<sup>30</sup> All assumptions made throughout this study are  
39 listed in Table S1 (*see supplement*).

40

#### 41 2.1.2 From air pollution exposure changes to PM emissions reductions

42 Many studies use forward-looking emission reduction modelling, in which each scenario is  
43 evaluated independently. The MECANO approach that we developed is an alternative way of  
44 carrying out this assessment using the capacity of models to perform source tracking. We used  
45 a model nesting to track the mass of pollutants emitted into the atmosphere with the  
46 Comprehensive Air Quality Model with Extensions (CAMx) and SIRANE models. These two  
47 models allow to trace emissions at the regional level and identify their local sources (*see*  
48 *supplement*, Figure S1).

49 CAMx model (Ramboll Environ, Arlington, Virginia, USA) is a Eulerian chemistry-  
50 transport model able to simulate atmospheric concentrations of particulate and gaseous  
51 pollutants. The regional domain was bounded to the Rhône-Alpes Region using a 3x3-km grid  
52 resolution. Boundary conditions on this domain were given by CHIMERE modelling at the  
53 27x27-km scale over Europe. Meteorological fields come from a simulation with the Weather  
54 Research and Forecast (WRF) model (NCAR, Boulder, Colorado, USA) configured on 3 nested  
55 domains of 27x27, 9x9 and 3x3 km. Emissions within the Rhône Alpes Region were provided  
56 by Atmo Auvergne-Rhône-Alpes inventory. Outside this area, emissions were retrieved from  
57 the EMEP database.<sup>31</sup>

58 CAMx and SIRANE were both used in a source apportionment approach for PM<sub>2.5</sub>. The  
59 contributions of emission sources from different sectors of activity and geographical areas were  
60 identified and followed during the simulation. The flagged emissions sources were grouped in  
61 CAMx according to five activities (industry, road traffic, residential heating, agriculture, and  
62 others). An interface was developed to apply coefficients to the different areas and activity  
63 sectors. In parallel, SIRANE was launched with traffic emissions only. Given the better spatial  
64 resolution of SIRANE model (to the road scale) compared to CAMx (1-km grid), traffic sector-  
65 related emissions from CAMx were not considered and were substituted by SIRANE model's  
66 estimates.

67 We then combined the models' estimates, assuming linearity between the emissions of all  
68 pollutants related to a given activity (traffic or urban heating) and the concentrations followed  
69 in the PSAT module of CAMX. This approach allowed to drastically reduce the calculation  
70 time of each scenario, from almost one week to less than one hour. This approach therefore  
71 made it possible to quickly screen various reduction scenarios. The limitations of this approach  
72 and its possible domain of validity are discussed in supplementary material (Box S1). To check  
73 the most ambitious scenario (67% decrease in PM<sub>2.5</sub>-related mortality), we made a forward

74 estimation of this emission scenario using CAMX. The population-weighted difference  
75 between this estimate and the backward estimate from MECANO approach was  $0.01\mu\text{g}/\text{m}^3$ .

76

### 77 2.1.3 From emission reduction to urban policy scenarios

78 The definition of urban policies to achieve health objectives focused on traffic and urban  
79 heating, the two main local sources of  $\text{PM}_{2.5}$ . The traffic scenarios explored different hypotheses  
80 of modal shift towards modes alternative to the private car in order to reach the reduction of  
81 vehicle.km consistent with the health objectives.

82 To identify the potential modal shift in the Grenoble conurbation, we relied on the Grenoble-  
83 area Household Travel Survey (2010), which describes all trips (travel time, mode of transport,  
84 purpose...) performed by a representative sample of the conurbation inhabitants. For each trip  
85 of the household, we enriched the database with data on travel distances and times by car  
86 (calculated with *Odomatrix*, from Hilal et al. (2018)<sup>32</sup>) and urban public transit (calculated with  
87 the Grenoble multimodal model). We first defined, as a reference scenario, the transport modal  
88 share of the main transport modes in 2010 (Table 1). We then developed three orientations of  
89 modal shift (later called sub-scenarios) from motorized vehicles (cars, taxis, motorized two-  
90 wheelers, vans, tractors...): one with an emphasis on public transport, one with an emphasis on  
91 active modes (walk and bicycle), and a third with an emphasis on active modes including also  
92 electrically-assisted cycling (E-bikes). The two latter sub-scenarios also considered carpooling,  
93 a mode of transport favored by local mobility policies, for which a shift was considered possible  
94 depending on the number of single drivers making the same trip (same origin and destination).

95 We considered the behavior of people using alternative modes to the car as a basis for  
96 describing which modes could be used by current car users, based on the time or distance of the  
97 trip and loop (i.e., trip series from home-to-home); from this we derived a potential mode shift.  
98 We considered the home-to-home loop because a trip that could be replaced by an active mode

99 when considered alone could not be done by the same mode of transportation when regarded  
100 as part of a loop (for instance, a 1-km trip to bring children to school followed by a 20-km trip  
101 to go from the school to work). For the potential of modal shift to public transportation, we  
102 considered the ratio of public transport travel time and car travel time, for the trip and for the  
103 loop, while for the potential of modal shift to the active modes (walking and cycling), we  
104 considered the distances of the trip and the home-to-home loop.

105 For the first sub-scenario, thresholds concerning time and distance of the trip and the  
106 loop were set at the median of the observations in the Household Travel Survey. From this, we  
107 then deduced a potential reduction in terms of motorized vehicle-km (cf. 3.1.2).

108 The thresholds under the two other sub-scenarios were chosen to achieve the same level  
109 of reduction in motorized vehicle.km as in the first sub-scenario, while remaining compatible  
110 with people's practices (in terms of time spent in transportation and distance to be traveled on  
111 foot or by bicycle) and the conurbation layout.

112 Table 1 displays the motivation for each modal shift sub-scenario, as well as the time  
113 and distance thresholds used. When a trip was potentially transferable to more than one mode  
114 of transport, the mode report hierarchy was as follows: i) walking; ii) cycling; iii) E-biking; iv)  
115 public transportation; v) carpooling.

116 Lastly, for each mode of transport, we estimated the change in the number of travelled  
117 kilometers between the reference scenario (2010) and each modal shift sub-scenario. On the  
118 basis of the Grenoble-area Household Travel Survey, we also assessed the travelers' age  
119 distribution under the reference and each sub-scenario, for each transport mode.

120 Analyses were performed using R software version 3.6.3 (R Foundation for Statistical  
121 Computing, Vienna, Austria).



Sub-scenario	Reference	“Public transport” sub-scenario	“Active modes” sub-scenario	“E-bike” sub-scenario
<b>Description</b>	Observations of the Household travel survey (2010). Transport modal share of the main transport modes (in % of passenger.km): -57% are drivers of cars and of two wheelers, -12% are passengers in cars and two wheelers, -22% corresponds to public transport, -7% for walk -2% for bicycles.	Starting from the reference scenario, this scenario implements a modal shift from motorized vehicles for all trips or loops whose distances or ratios of travel time are below the median of the observations concerning the mode towards which the shift is done.	Favors the active modes to reduce the mode share of public transport. Introduces carpooling.	Introduces E-biking in addition to carpooling. The modal share of public transport is close to the level it has in the reference scenario.
<b>Walking (distance)</b>		< 0.7 km (trip) < 2 km (loop)	< 2 km (trip); corresponds to the stated objective of the local urban mobility plan < 5 km (loop)	< 2 km (trip); corresponds to the stated objective of the local urban mobility plan < 5 km (loop)
<b>Cycling (distance)</b>		< 1.8 km (trip) < 5.1 km (loop)	< 5 km (trip); corresponds to the 9 <sup>th</sup> decile currently observed in trips performed by bike < 12.5 km (loop)	< 3 km (trip) < 7.5 km (loop)
<b>Threshold values of trips entailing a shift from the car</b>				
<b>E-biking (distance)</b>				< 7.8 km (trip) < 19.5 km (loop)
<b>Public transportation (ratio with car travel time)</b>		< 1.7 × car travel time (trip and loop)	< 1.5 × car travel time (trip and loop)	< 1 × car travel time (trip and loop)
<b>Carpooling</b>		Not considered	Drivers may become passengers if their origin-destination trip was made in a driver’s car at least 400 times a day; we then randomly allocated 4 out of every 10 trips in a driver’s car to a passenger car.	Drivers may become passengers if their origin-destination trip was made in a driver’s car at least 400 times a day; we then randomly allocated 4 out of every 10 trips in a driver’s car to a passenger car.
<b>Motorized vehicle.km</b>	3.18 M. vehicle.km	2.05 M. vehicle.km	2.05 M. vehicle.km	2.07 M. vehicle.km
<b>Motorized person.km</b>	3.85 M person.km	2.44 M. person.km	2.44 M. person.km	2.45 M. person.km

**Table 1.** Definition of the modal shift sub-scenarios of *Traffic\_3* scenario.

Loop threshold distances are equal to about 2.5 times trip threshold distances, which is approximately the number of trips in a loop. M: million

## 2.2 Health and economic evaluation of urban policies

### 2.2.1 Health impact assessment

Several long-term health effects likely to be linked to PM<sub>2.5</sub> exposure, and also to physical activity for most of them, were considered: all-cause non-accidental mortality (ICD10: A00-R99), lung cancer incidence (ICD10: C33-34), ischemic heart disease incidence (ICD10: I20-25), and cerebrovascular disease incidence (ICD10: I60-69). All-cause non-accidental mortality data were obtained from the Epidemiological center on causes of death (CépiDC, Inserm) for people aged 30 years and older, at the municipality scale. Lung cancer incidence in the study area was estimated using regional data from the French National Cancer Institute (data 2007-2016).<sup>33</sup> Numbers of new beneficiaries from the “long duration diseases” database of the French health insurance system (ALD<sub>30</sub>)<sup>34</sup> were used as incidence data for ischemic heart disease and cerebrovascular disease.

From the literature, we selected concentration-response functions between PM<sub>2.5</sub> exposure or physical activity and each health outcome; robust risk-ratios, such as those from meta-analyses, were preferred (Table 2).

The expected “direct” health benefits related to the reduction of air pollution were estimated for each health outcome and each policy plan as the difference in the number of attributable cases ( $\Delta_{NAC}$ ) between the baseline (2015-2017 period) and a counterfactual situation under which the policy is assumed to be implemented. Differences in attributable cases  $\Delta_{NAC_{i,j}}$  were first estimated at each 10-m air pollutant model grid point before being added to estimate the  $\Delta_{NAC}$  for the whole area.<sup>25</sup> We also expressed health gains in terms of number of life-years in people aged 30 years and older. A cessation lag before obtaining complete health benefits was applied: 30% of the total health gain the first year after implementation, 50% distributed over years two to five, and 20% through the following fifteen years.<sup>35</sup>

We also evaluated “indirect” health benefits such as those subsequent to the increase in physical activity under scenarios supporting active mobility, through the difference of the average metabolic equivalent hours (MET.hours) per week and inhabitant between the situation with the scenario and that without implementation of the scenario (at the same period). Increase in walking in link with shift to public transport was also considered. To that end, traveled kilometers were firstly converted in MET.hours considering an average speed of 4.8, 15.1 and 18.3 km.h<sup>-1</sup> and 3.5, 5.8 and 4.5 METs for walking, cycling, and E-bikes respectively.<sup>36,37</sup> Like for health benefits related to air pollution decreases, we assessed for each modal report scenario the number of preventable death cases, and the corresponding number of years of life saved, as well as the number of preventable cardiovascular disease cases.

“Indirect” effects also included those related to traffic injuries. Total number of incidents was obtained at the municipality scale by sex and transportation mode for the 2008-2012 period, from the National inter-ministerial road safety observatory.<sup>38</sup> For each transportation mode, the injury risk (fatal, serious, or light) per km traveled was estimated as the ratio of the annual average number of incidents to the total number of kilometers traveled in the study area in 2010 (*see supplement*, Table S2). The number of traffic injuries expected by mode under each sub-scenario was then estimated applying these injury risks. The increase in injuries in walkers in relation with shift to public transport was also considered (public transport users have to reach their bus/tram on foot). Concerning fatal injuries only, the associated life-year loss was estimated by comparing, by sex and mode, the average age at death among victims under the considered scenario to the life expectancy at age 40 (i.e., the average age of Grenoble conurbation commuters), which was equal in France to 40.6 years in men and 46.2 years in women in 2014.<sup>39</sup>

Health event	Study source	Meta-analytical relative risk (95% CI)	Unit
<b>Effects of PM<sub>2.5</sub> exposure</b>			
Non-accidental mortality	World Health Organization (2014) <sup>40</sup>	1.066 (1.040 – 1.093)	For a 10 µg/m <sup>3</sup> increase
Lung cancer incidence	Hamra et al. (2014) <sup>41</sup>	1.09 (1.04 – 1.14)	For a 10 µg/m <sup>3</sup> increase
Ischemic heart disease incidence	Cesaroni et al. (2014) <sup>42</sup>	1.28 (0.96 – 1.69)	For a 10 µg/m <sup>3</sup> increase
Cerebrovascular disease incidence	Stafoggia et al. (2014) <sup>43</sup>	1.42 (0.77 – 2.62)	For a 10 µg/m <sup>3</sup> increase
<b>Effects of physical activity</b>			
Non-accidental mortality (for walking)	Kelly et al. (2014) <sup>17</sup>	0.89 (0.83 – 0.96)	For 11.25 MET.hour.week <sup>-1</sup> of walking
Non-accidental mortality (for cycling)	Kelly et al. (2014) <sup>17</sup>	0.90 (0.87 – 0.94)	For 11.25 MET.hour.week <sup>-1</sup> of cycling
Ischemic heart disease incidence	Kyu et al. (2016) <sup>18</sup>	0.860 (0.796 – 0.926)	For 10.0 MET.hour.week <sup>-1</sup> of activity
Cerebrovascular disease incidence	Kyu et al. (2016) <sup>18</sup>	0.862 (0.780 – 0.954)	For 10.0 MET.hour.week <sup>-1</sup> of activity

MET.hours: metabolic equivalent hours.

**Table 2.** Concentration-response functions used to estimate the long-term effects of air pollution exposure to fine particulate matter (PM<sub>2.5</sub>) and physical activity on health. A relative risk below one indicates a protective effect.

Analyses were performed using Stata version 15.1 (Stata Corp., College Station, TX, USA) while geographical data were handled with QGIS version 3.4 (OSGeo Foundation, Beaverton, OR, USA).

### 2.2.2 Cost-benefit analysis

The perspective of the cost-benefit analysis is societal at the local level, so that we did not consider costs to the general economy, such as possible lost incomes in the car manufacturing industry or gains for public transport or bike manufacturers, for example. In order to consider all health gains and lag effects, the cost-benefit analysis covered the period 2017-2045. All amounts were expressed in Euros 2021. We considered the time profiles of the costs and

benefits of the urban policies over a period of 29 years from 2017 to 2045. We assumed a discount rate equal to 4% for both costs and benefits.<sup>44</sup>

#### *2.2.2.1 Health benefits induced by pollution reduction*

For each scenario, we estimated the avoided costs associated with all-cause non-accidental mortality, lung cancer incidence, and cerebrovascular disease incidence due to pollution alleviation. We considered direct and indirect tangible costs as well as intangible costs. Direct costs refer to medical costs, indirect costs to loss of productivity due to sick leave and intangible costs to grief and loss of quality of life. The assessment methods for all-cause non-accidental mortality and lung cancer incidence relied on *Aphekom project* methodology and values.<sup>45</sup> These intangible costs, based on a literature review of contingent valuations, corresponded to €103,592 per year of life saved (69,054-138,120; the extreme values corresponding to variations of 33% around the central value).<sup>45</sup> For cerebrovascular disease incidence, we estimated direct medical costs by considering the differences between patients admitted in nursing home or not<sup>46</sup> and we assumed a seven-year life expectancy;<sup>47</sup> unit health costs are given in Table S3 (*see supplement*). Due to a lack of reliable data on costs associated with ischemic heart disease incidence, no economic valuation was provided for this health outcome.

#### *2.2.2.2 Other externalities specific to the transport sector*

In addition to the health benefits related to pollution exposure reduction, we considered five indirect health impacts specific to the transport sector. Firstly, we considered the cost related to the change in the frequency of road traffic accidents. Concerning fatal injuries, intangible costs of corresponding life years were the same as above, i.e., €103,592, whereas costs of serious and light injuries were respectively equal to 12.5% and 0.5% of the value of a statistical life, which is 3.458 M€,<sup>48</sup> per case. Secondly, we monetized the years of life gained from the additional physical activity generated by the shift to active modes of transport, using the same value of 103,592€ per year of life saved. Thirdly, the composition of the vehicle fleet and modal share

can also affect the exposure of the population to noise pollution in the vicinity of traffic. We relied on an average cost of noise per vehicle.km from a published study,<sup>49</sup> and we applied a noise pollution cost 50% lower for electric cars.<sup>50</sup> Fourth, we considered non-health benefits related to greenhouse emission decrease. The fleet regulation and the growth of active mobility decrease greenhouse gas (GHG) emissions due to road traffic. GHG emission reductions were monetarized following the recently computed global social cost of carbon equal to US\$417 per ton of CO<sub>2</sub>.<sup>51</sup> Lastly, we considered benefits related to the value of time spent in transport. The global journey time variation was assessed, assuming an average speed based on the literature for the active modes<sup>17,36,37</sup> and on local transport data for motorized modes. We then applied perceived Values of Time (VoT) per mode of transport to convert these changes in journey duration into Euros.<sup>52-55</sup>

#### *2.2.2.3. Public costs and private expenditures specific to mobility-related policies*

We also estimated the costs engaged by the public actors under each scenario. The implementation of a LEZ in itself does not require significant investment, contrarily to the development of infrastructures to facilitate cycling and the development of public transport. In the following, we rely on the estimated budget for the realization of the mandatory urban mobility plan,<sup>56</sup> provided by the Grenoble conurbation to define its transport policy. The budgets required into the estimation for each scenario were thus deduced by multiplying the urban mobility plan costs associated to each mode of transport by the ratios of vehicle.km of each mode in all scenarios, compared to that of the urban mobility plan. A portion of the costs remained constant through all scenarios because of their low sensitivity to changes in infrastructure plans.

The change in mobility behavior induced by the traffic policies also affects household expenditures for mobility. We estimated private total costs integrating the investment cost and the cost of use of each transport mode per vehicle.km to estimate the impact of the modal report

on the cost for the users concerned. The total cost per kilometer were respectively from €0.32 to €0.51 for cars (depending on the type of motorization), €0.15 for public transport, and from €0.12 to €0.22 for bikes (depending on the share of E-bikes and the use intensity).

#### *2.2.2.4 Public costs and private expenditures specific to policies in the heating sector*

Public costs related to the wood stove replacement premium include subsidies and operating costs (communication and animation costs supported by the conurbation to convince inhabitants to replace their wood burning facilities by less polluting ones). The subsidies are the premium itself (which amounts to €1600 for most households and €2000 for the most deprived ones) and national grants aimed at promoting the energy transition. Operating costs include communication and animation costs supported by the conurbation. Private costs are the expenses incurred by households for installing a new heating system after deduction of public subsidies, which are more important for the most deprived people. We also considered that household saved €300 on energy bills each year due to more efficient heating devices.

## **3 Results**

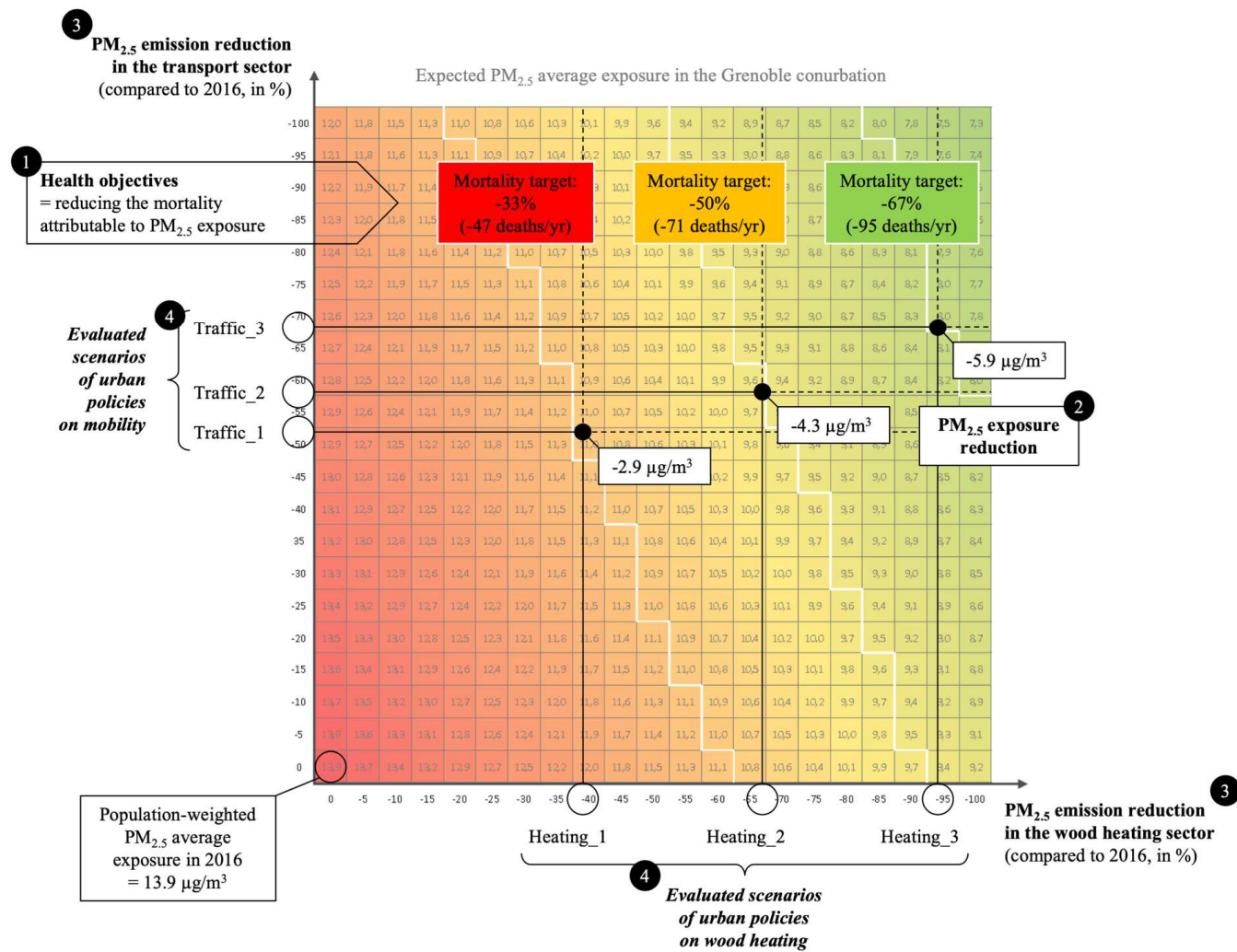
### **3.1 Reverse approach: identification of policies allowing to reach the health targets**

#### 3.1.1 Scenarios combining actions on wood heating and traffic

As already mentioned, based on exposure levels from 2015-2017, we previously estimated that the population density-weighted long-term average exposure to anthropogenic PM<sub>2.5</sub> in Grenoble conurbation corresponds to 145 (95% CI, 90, 199) attributable premature death cases each year.<sup>25</sup> Our previous study<sup>25</sup> identified the air pollution decrease required to reach the three health targets we selected, namely -33%, -50%, and -67% of this PM<sub>2.5</sub>-attributable mortality. Figure 2 shows these air pollution decreases and allows identifying combinations of emission reduction targets in each sector expected to achieve each of the three health objectives. A large

number of combinations allow to reach these targets, and for practical reasons, we decided to consider three global scenarios (one for each target) each combining one policy in the transport sector and one policy in the wood heating sector; these policies are detailed in Table 3. The most ambitious health target (-67% of the PM<sub>2.5</sub>-attributable mortality) could be achieved in 2030 by replacing all inefficient wood heating equipment by pellet stoves and by reducing by 36% the traffic of private motorized vehicles.





**Figure 2.** Expected PM<sub>2.5</sub> annual average exposure in the Grenoble conurbation population with regard to the combination of urban policies in the wood heating sector and in the transport sector (in µg/m<sup>3</sup>).

Following our reverse approach: ❶ starting from health objectives seeking to reduce the PM<sub>2.5</sub>-attributable mortality; ❷ then assessing the reduction in PM<sub>2.5</sub> exposure needed to reach these health objectives (symbolized by zigzagging white lines); ❸ assessing the corresponding reduction in PM<sub>2.5</sub> emissions in the wood heating and the transport sectors; ❹ and designing compatible urban policy scenarios that will be evaluated through a cost-benefit analysis.

**Table 3.** Description of the heating- and traffic-related policy scenarios, and their combinations, aiming at reducing fine particulate matter (PM<sub>2.5</sub>) exposure and improving public health.

<b>Health target</b>	<b>-33% of the mortality attributable to PM<sub>2.5</sub> (-47 deaths/year)</b>		<b>-50% of the mortality attributable to PM<sub>2.5</sub> (-71 deaths/year)</b>		<b>-67% of the mortality attributable to PM<sub>2.5</sub> (-95 deaths/year)</b>	
Corresponding decrease in PM <sub>2.5</sub> exposure	-2.9 µg/m <sup>3</sup>		-4.3 µg/m <sup>3</sup>		-5.9 µg/m <sup>3</sup>	
<b>Combination of policy scenarios allowing to reach the health target</b>	<b>Heating_1</b>	<b>Traffic_1</b>	<b>Heating_2</b>	<b>Traffic_2</b>	<b>Heating_3</b>	<b>Traffic_3</b>
Description of the policy scenarios	Replacement of one-third of non-efficient wood stove appliances over the 2016-2020 period, as planned by the wood stove replacement premium plan. <sup>57</sup>	The already planned LEZ avoiding goods transport vehicles labeled <i>CritAir 2</i> or above.	Once the wood stove replacement premium target has been reached in 2020 ( <i>Heating_1</i> ), replacement of all inefficient wood heating equipment between 2021 and 2030 through a subsidy by efficient wood stoves.	Extension of the LEZ to all vehicles (including passenger cars) labeled <i>CritAir 2</i> or above, without modal shift compared to the modal distribution observed in 2010 (Grenoble-area Household Travel Survey); in this scenario, only the technological lever is considered (i.e., the share of cars in the overall mobility remains unchanged compared to the reference year).	Once the wood stove replacement premium target has been reached in 2020 ( <i>Heating_1</i> ), additional subsidy to replace <i>all</i> inefficient wood heating equipment between 2021 and 2030 by pellet stoves, which currently are the least polluting wood heating facilities.	In addition to <i>Traffic_2</i> , we assumed a modal shift, corresponding to a decrease by 36% in the number of motorized vehicle.km compared to 2016; the <i>Traffic_3</i> measure was broken down into 3 sub-scenarios, which differed in terms of the modes (public transport, actives modes, car-sharing) towards which the decrease in the vehicle.km of motorized vehicles were dispatched.
Decrease in PM <sub>2.5</sub> emissions in the considered sector	-39%	-52%	-67%	-58%	-94%	-68%
Corresponding decrease in PM <sub>2.5</sub> exposure	-1.8 µg/m <sup>3</sup>	-1.0 µg/m <sup>3</sup>	-3.1 µg/m <sup>3</sup>	-1.1 µg/m <sup>3</sup>	-4.4 µg/m <sup>3</sup>	-1.3 µg/m <sup>3</sup>

LEZ: Low emission zone.

### 3.1.2 Sub-scenarios implementing shifts in mode share in the -67% scenario

We selected three sub-scenarios of modal shift under the scenario *Traffic\_3*, which led to a 36% reduction in the number of vehicle.km travelled using motorized vehicles compared to the baseline situation (2010). We estimated that a 36% reduction in the number of vehicle.km travelled in motorized vehicles was required to reach the PM<sub>2.5</sub> exposure decrease due to traffic changes identified for the scenario allowing the 67% decrease in PM<sub>2.5</sub> mortality.

In the first sub-scenario (Figure 3), the modal shift from car to alternative modes is then possible if the distances travelled (for active modes) or the time ratios (car/public transport) are below the median of these same indicators observed for current users of these alternative modes (median threshold). In this sub-scenario, named *Public Transport* due to the strong modal shift to public transport, the modal share of motorized modes dropped to 44% (i.e., -36% compared to the baseline of 69%), whereas public transport and cycling modal shares doubled (from 22% to 45% and from 2.3% to 4.5%, respectively).

The second sub-scenario was named *Active modes* due to the strong increase of the modal share of walking and cycling while the third sub-scenario was named *E-Bikes* due to the introduction of this electric mode. In the *Active modes* and *E-bikes* sub-scenarios (Figure 3), the shares of cycling (including E-Bikes) reached 12% and 23%, respectively. The walking share rose to 9.6%, against 6.6% in the baseline and 6.7% in the *Public Transport* sub-scenario. The public transport modal shares were lower in these scenarios than in the first one and reached respectively 34.4% and 23.1%.



**Figure 3.** Transport mode share under the reference situation (2010) and the 3 traffic sub-scenarios of the scenario *Traffic\_3* (expressed in percentages of passenger.kilometers travelled).

### 3.2 Forward approach: cost-benefit analysis of the identified urban policies

#### 3.2.1 Health impact assessment

On average, 2,601 people aged 30 years and more non-accidentally died each year in Grenoble urban area under the baseline situation. There were 219 new lung cancer annually diagnosed, 568 new ischemic heart disease cases, and 325 new cerebrovascular disease cases.

Table 4 shows the number of death and disease cases which could be prevented yearly under each policy, because of changes in  $PM_{2.5}$  exposure, physical activity, or traffic injuries, while Table 5 presents the benefits of the policy combinations and Table S4 (*see supplement*) details the expected changes in mortality and morbidity for the distinct modal shift sub-scenarios. Reaching the -33%  $PM_{2.5}$ -attributable mortality objective by combination of the *Heating\_1* and *Traffic\_1* policies prevented, additionally to the targeted decrease of 47 (95% CI, 29, 65) death cases a year, 5 (95% CI, 2, 8) incident lung cancer cases, 31 (95% CI, 0, 82) strokes, and 39 (95% CI, 0, 81) incident ischemic heart disease cases each year. Combination of *Heating\_2* and *Traffic\_2* policies (50% decrease in  $PM_{2.5}$ -attributable mortality) avoided 70

(95% CI, 43, 97) death cases, and also 8 (95% CI, 4, 12) incident lung cancer cases, 46 (95% CI, 0, 118) strokes, and 57 (95% CI, 0, 118) incident ischemic heart disease cases each year. Lastly, *Heating\_3* plus *Traffic\_3* policies led to “direct” health gains equal to 93 (95% CI, 58, 130) death cases, 10 (95% CI, 5, 16) incident lung cancer cases, 60 (95% CI, 0, 150) strokes, and 76 (95% CI, 0, 154) incident ischemic heart disease cases annually prevented. In addition, under the third transport modal shift sub-scenario (“E-bikes”), the “indirect” health benefits (related to the increase in physical activity and the changes in fatal traffic injuries) corresponded to 179 (95% CI, 85, 257) death cases (192% of the number of deaths directly avoided), 2 (95% CI, 1, 4) incident lung cancer cases, 37 (95% CI, 12, 62) strokes, and 66 (95% CI, 34, 99) incident ischemic heart disease cases prevented each year.

	Wood heating-related policies			Road traffic-related policies				
	Heating_1	Heating_2	Heating_3	Traffic_1	Traffic_2	Traffic_3		
						“Public transport” sub-scenario <sup>b</sup>	“Active modes” sub-scenario <sup>b</sup>	“E-bike+” sub-scenario <sup>b</sup>
<b>Change in:</b> <sup>a</sup>								
Premature deaths	-30 (-42, -19)	-52 (-72, -32)	-72 (-100, -45)	-16 (-23, -10)	-18 (-25, -11)	-79 (-117, -38) -58 (-87, -24)	-179 (-260, -86) -158 (-230, -73)	-201 (-287, -98) -179 (-257, -85)
Corresponding years of life	+958 (587, 1335)	+1647 (1009, 2297)	+2308 (1412, 3221)	+516 (316, 718)	+575 (353, 801)	+2602 (1235, 3860) +1928 (821, 2920)	+6096 (2851, 9027) +5421 (2437, 8088)	+6890 (3280, 10087) +6216 (2866, 9147)
Lung cancer incidence	-3 (-5, -2)	-6 (-9, -3)	-8 (-12, -4)	-2 (-3, -1)	-2 (-3, -1)	-2 (-4, -1)	-2 (-4, -1)	-2 (-4, -1)
Stroke incidence	-20 (-53, 0)	-34 (-85, 0)	-46 (-112, 0)	-11 (-30, 0)	-12 (-33, 0)	-25 (-56, -3) -11 (-18, -3)	-46 (-90, -10) -31 (-52, -10)	-52 (-100, -12) -37 (-62, -12)
Ischemic heart disease incidence	-25 (-52, 0)	-42 (-86, 0)	-58 (-117, 0)	-14 (-29, 0)	-15 (-32, 0)	-37 (-66, -10) -19 (-29, -10)	-73 (-120, -29) -55 (-83, -29)	-84 (-136, -34) -66 (-99, -34)
<b>Cumulated and discounted net benefits (in M€ over the whole period 2017-2045)</b>	1300	1902	2545	1466	2228	197	4399	6329
<b>Benefit/cost ratio (over the period 2017-2045)</b>	26.8	26.7	35.4	67.0	19.8	1.1	3.1	4.7

<sup>a</sup> Estimate and 95% confidence interval (95% CI) of the number of cases annually prevented, and corresponding gain (estimate, 95% CI) in years of life for prevented death cases.

<sup>b</sup> In black: Total change, combining changes related to the decrease in air pollutant exposure, the increase in physical activity, and, if relevant, the changes in fatal traffic injuries; in italic grey: change related to modal shift only (i.e., related to the increase in physical activity and, if relevant, the changes in fatal traffic injuries; *see supplement*, Table S4).

**Table 4.** Changes in air pollutant exposure, mortality and morbidity for each scenario of urban policy related to wood heating and road traffic.

## 3.2.2 Economic analysis

### 3.2.2.1 Heating-related measures

All the measures aimed at replacing wood heating equipment led to a positive net annual balance (benefits - costs) from the very first years, rapidly generating each year between 60 M€ (for *Heating\_1*) and 140 M€ (for *Heating\_3*) in net benefits (Figure 4). Almost all of this balance was made up of the health benefits linked to the reduction of atmospheric pollution. When expressed as a benefit/cost ratio of values averaged over the whole 2017-2045 period, for each Euro invested by the public policy, the benefits generated were higher than 25€ (Table 4, *see supplement*, Table S5).

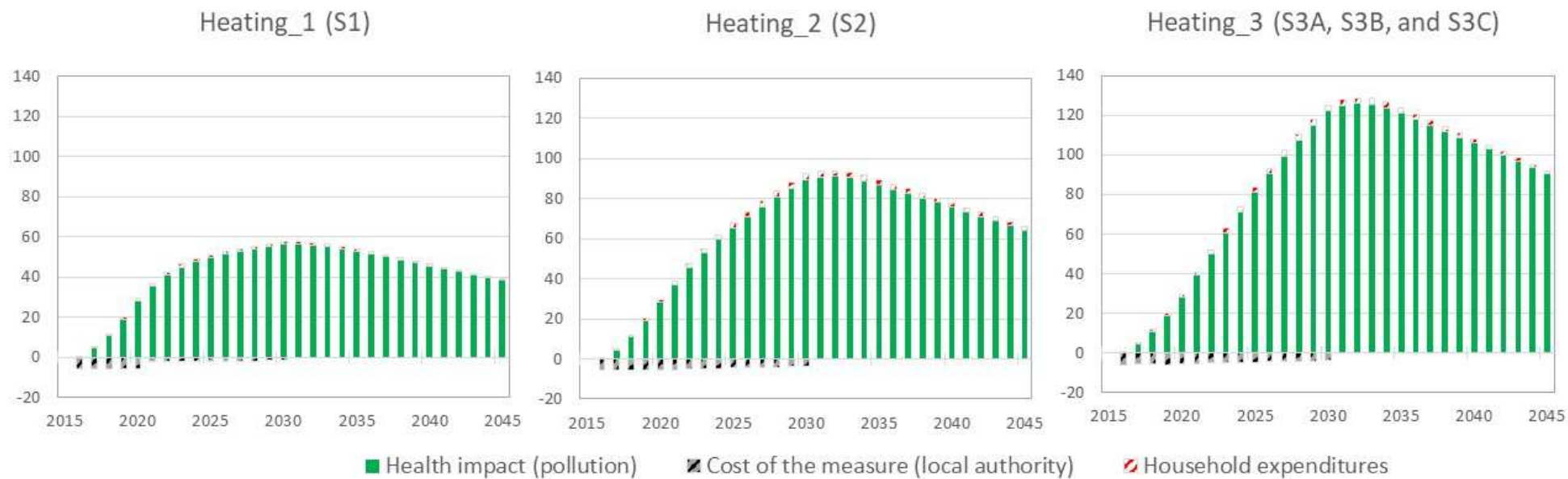
### 3.2.2.2 Traffic-related measures

For traffic policies (**Figure 5**), *Traffic\_1* and *Traffic\_2* led to annual net benefits from 50 M€ to 100 M€ as early as 2025, comparable to the heating-related measures. *Traffic\_3* sub-scenario focusing on the development of public transport resulted in an initially negative balance sheet, with a net cost around €200 million per year on average until 2030. Thereafter, costs and benefits were broadly balanced until the end of the period. In contrast, the other *Traffic\_3* sub-scenarios based on the development of active modes and E-bikes led to a positive net annual balance almost from the beginning of the period, and above 200 M€ after 2030. When expressed as a ratio, one Euro invested in traffic-related measures led to benefits of 19.6 for *Traffic\_2*, and respectively 1.1, 3.1 and 4.7 for *Traffic\_3* alternatives focusing on public transport, active modes and E-bikes (Table 4, *see supplement*, Table S5).

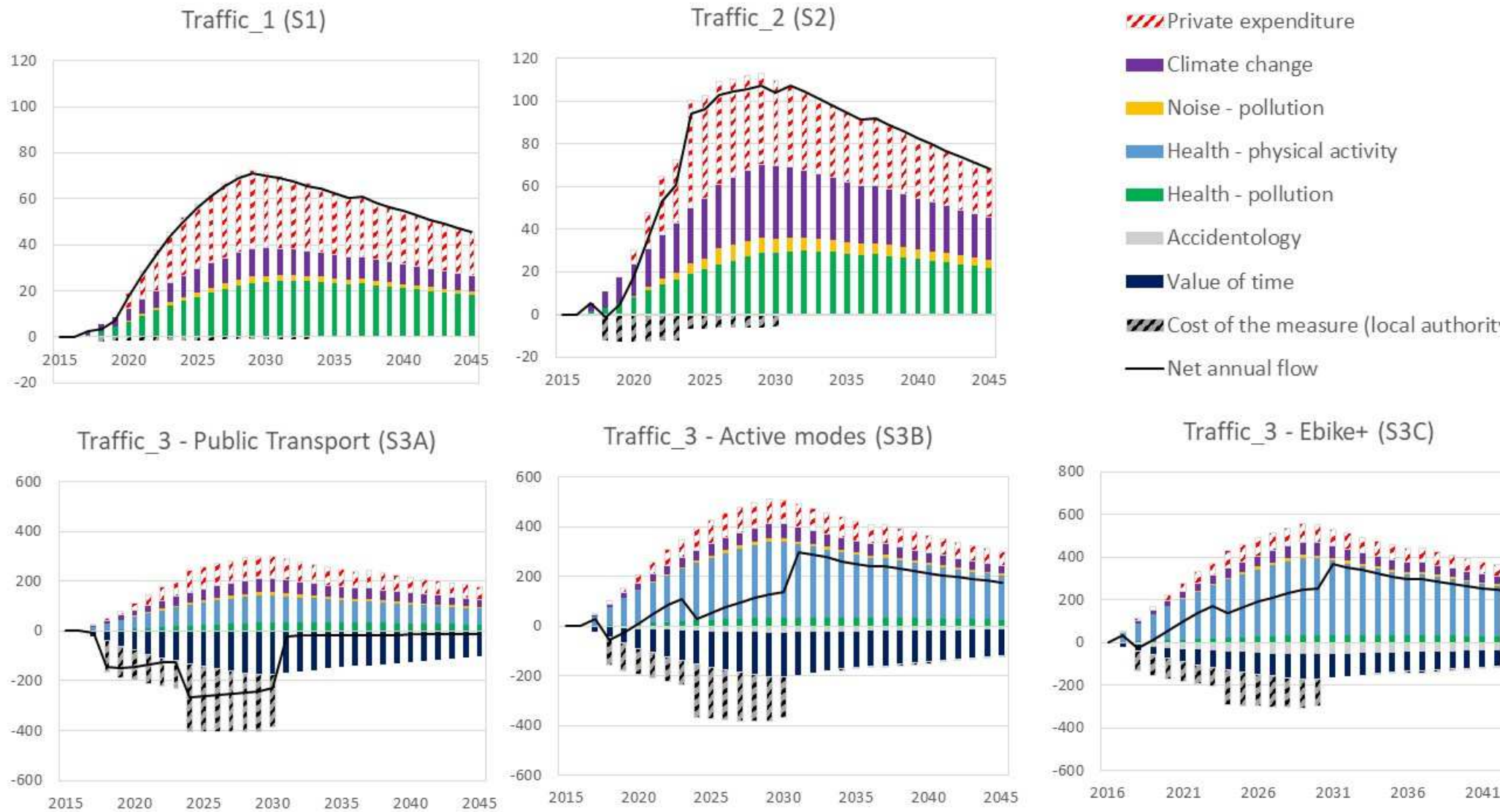
In terms of structure of benefits, the traffic-related benefits mainly consisted of health benefits related to pollution reduction, reduction of the impact on climate change and savings by households on the investment and operating costs of cars. The time profile and structure of the cost-benefit analysis of *Traffic\_3* differed between *Traffic\_1* and *Traffic\_2* in three ways. First, the implementation of modal shifts in *Traffic\_3* required substantial investments in

infrastructure, contrarily to the first two traffic scenarios. Second, health benefits related to increased physical activity represented more than half of the benefits generated by each *Traffic\_3* alternative; these benefits linked to physical activity corresponded to about ten times the value of the health benefits linked to the pollution reduction *per se*. Third, the losses related to the value of time and the increase in travel time due to modal shifts represented a significant net cost.





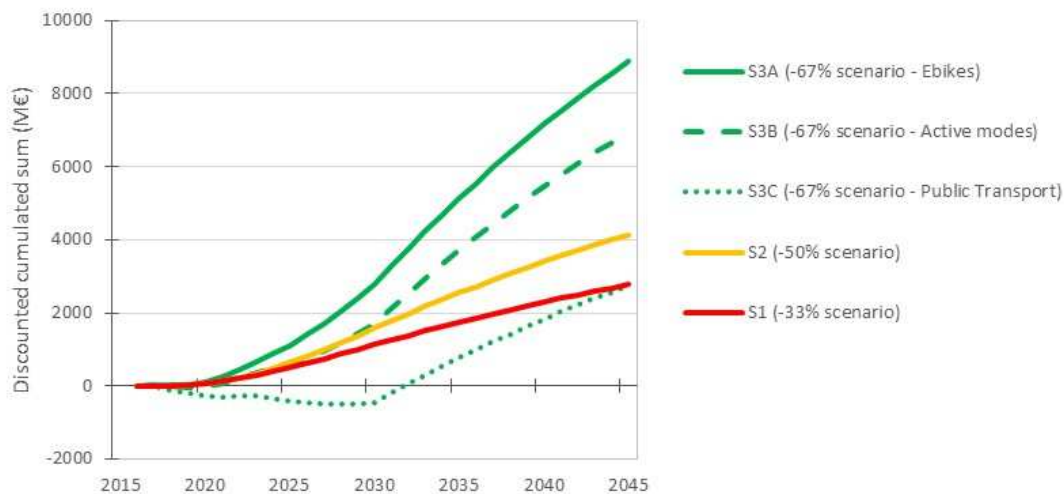
**Figure 4.** Annual flow of costs and benefits generated by wood heating-related measures (Million €).



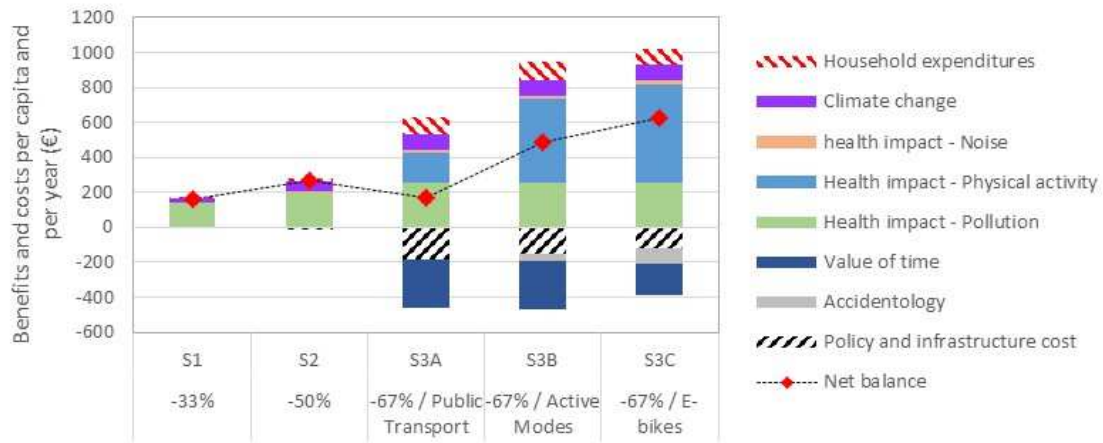
**Figure 5.** Annual flow of costs and benefits generated by measures in the transport sector (Million €).

### 3.2.2.3 Cost-benefit analyses of scenarios combining heating and traffic-related measures

The discounted sum of costs and benefits of the various combinations of wood heating and mobility measures targeting a 67% decrease in PM<sub>2.5</sub>-related mortality is presented in **Figure 6**. Over the whole period, the two sub-scenarios favoring modal shifts towards active modes or electric bicycles were the most interesting economically, generating annual net benefits equal to €6.7 billion (i.e. €468 per capita each year) and €8.6 billion (i.e. €615 per capita each year) respectively over the whole period, i.e., 2 to 4 times more than the -33% and -50% scenarios (Table 5 and **Figure 7**). All scenarios presented a positive cumulative balance almost from the first years, except for the sub-scenario based on public transport, which required a very high public investment and generated fewer health benefits related to physical activity than the two other *Traffic\_3* alternatives.



**Figure 6.** Time profile of the discounted sum of benefits and costs for each of the five considered scenarios (Million €, 2017-2045 period).



**Figure 7.** Result of the cost-benefit analysis per year and per capita for each of the five considered scenarios (S1: -33% in PM<sub>2.5</sub>-related mortality, S2: -50%; S3: -67%; see Table 3 and Figure 3 for the definition of scenarios).

The overall gain is shown by the red diamond shape while surfaces indicate the structure of costs and benefits.

Health target	-33% of the mortality attributable to PM <sub>2.5</sub> (-47 deaths/year)	-50% of the mortality attributable to PM <sub>2.5</sub> (-71 deaths/year)	-67% of the mortality attributable to PM <sub>2.5</sub> (-95 deaths/year)		
	Heating_1 + Traffic_1	Heating_2 + Traffic_2	Heating_3 + Traffic_3 “Public Transport”	Heating_3 + Traffic_3 “Actives modes”	Heating_3 + Traffic_3 “E-bikes+”
<b>Change in:</b> <sup>a</sup>					
Premature deaths	-47 (-65, -29)	-70 (-97, -43)	-151 (-217, -83)	-251 (-360, -131)	-273 (-387, -143)
Lung cancer incidence	-5 (-8, -2)	-8 (-12, -4)	-10 (-16, -5)	-10 (-16, -5)	-10 (-16, -5)
Stroke incidence	-31 (-82, 0)	-46 (-118, 0)	-71 (-168, -3)	-92 (-202, -10)	-98 (-212, -12)
Ischemic heart disease incidence	-39 (-81, 0)	57 (-118, 0)	-95 (-183, -10)	-131 (-237, -29)	-142 (-253, -34)
<b>Cumulated and discounted net benefits (in M€ over the whole period 2016-2045)</b>	2,767	4,130	2,742	6,944	8,875
<b>Net benefit per year and per capita (over the period 2016-2045)</b>	162	270	168	484	629
<b>Benefit/cost ratio (over the period 2016-2045)</b>	39.3	22.5	2.1	4.2	6.0

<sup>a</sup> Estimate and 95% confidence interval (95% CI) of the number of cases annually prevented, and corresponding gain (estimate, 95% CI) in years of life for prevented death cases.

**Table 5.** Changes in mortality and morbidity for each combination of policies related to wood heating and road traffic, allowing to reach the pre-defined health targets.

## **4 Discussion**

We used a backward modelling approach to define policies to achieve air pollution mortality reduction targets set by local decision-makers, and then used a forward modelling approach to assess their health and economic impacts, including indirect benefits not considered in the backward approach. The most ambitious health target (-67% of the PM<sub>2.5</sub>-attributable mortality) could be achieved in 2030 by replacing all inefficient wood heating equipment by pellet stoves and by reducing by 36% the traffic of private motorized vehicles. Assuming no change in distance travelled, such a reduction in motorized traffic would require identical increases in active modes share (walking, biking...); these would also induce, thanks to the increase in physical activity, additional health benefits beyond the initial target. Wood heating system replacement and strategies maximizing active mobility by other means than the development of public transport were the most cost-effective policies. At the scale of this urban area of 444,000 inhabitants, over a 25-year period, reaching the -67% PM<sub>2.5</sub>-attributable mortality target by promoting active modes would generate at the conurbation level cumulated net benefits of around €7-9 billion in 2045 whereas the -33% PM<sub>2.5</sub>-attributable mortality target, reachable with policies already planned, would lead to a net benefit equal to €2.7 billion.

### **4.1. Policy recommendations**

Policies to be implemented for wood heating are easily identified. Subsidizing the most performing appliances can be particularly effective from both a health point of view (since they drastically reduce emissions related to wood heating compared to old appliances) and from an economic point of view, since the leverage effect is very important (a benefit of around 30€ for each euro invested by public authorities). In practice, the implement. Banning the most polluting appliances (in particular open fireplaces) is worth considering as, in the considered

area, these were responsible for 63% of PM<sub>2.5</sub> emissions in 2016. When it comes to the transport sector, the health impact of PM<sub>2.5</sub> emission reductions was less important because of lower emission levels, but the modal shifts considered produced indirect health benefits adding to those directly related to pollution reduction, particularly through the increase in physical activity, the decrease in greenhouse gas emissions, and the reduction of noise, representing respectively 91%, 15%, and 3% of net benefits in the -67% scenario including a modal shift toward E-bikes (S3C). The deployment of cycling appeared particularly interesting for several reasons: first, because of its co-benefits, but also because the investments in infrastructure and facilities needed to promote cycling (bike lines, facilities that make cycling enjoyable, bicycle parking, showers in the workplace) can be easily implemented through light changes in the existing urban infrastructures. In the future, electric bicycles are expected to be in great demand as an alternative for medium-distance journeys made today by private car. The electric bicycle pushes back the limits in terms of distance of the modal shift from the private car and can be adopted by people reluctant to ride a muscle-powered bicycle.

One challenge is the development of ambitious programs to encourage the practice of conventional and electric bicycles. Beyond economic incentives (subsidies for bicycle purchase, reimbursement by employers) and dedicated facilities, it is also important that policies and programs target the psychological levers (perception, intention, altruism, social norms...) allowing to promote cycling share.<sup>58</sup>

## **4.2 Methodological strengths and limitations**

### **4.2.1 Mobility behavioral changes**

We made several assumptions worth being discussed. Concerning the sub-scenarios of transportation mode shift, we considered the behavior of people using alternative modes to the car as a basis for describing which modes could be used by current car users. This hypothesis is restrictive because mode choice is multi-factorial and other factors come into play, such as

age, gender or number of children. However, we considered that only car trips with distance or travel time characteristics less than the median of the distances and travel times achieved in the alternative modes could shift to those modes. The potential shift could actually be much greater if we considered the characteristics of a larger share or even of all cyclists and other users of alternative modes to the car. Even though some trips considered transferable to alternative modes are not transferable because of characteristics such as age or reason for travel, other trips that we did not consider transferable because we have limited ourselves to the median threshold are in fact transferable. The objective here was not to give one single good answer but a range of possibilities to achieve a substantial reduction (-36%) in motorized vehicle.km, anchoring the alternatives on one or other of the modes of transport (active modes, E-bikes).

The public transport scenario led to a strong modal shift to public transport, which could not be absorbed by current or even possibly future infrastructure, due to financial and land-use constraints, according to the local stakeholders. This sub-scenario, purely based on a statistical description of the Household Travel Survey (2010), also did not include the fast-developing E-bikes, nor carpooling. To address these limitations, the two alternative scenarios were constructed and evaluated, in particular, one interesting feature of the E-bikes sub-scenario is the stability in the public transport modal share compared to the reference, which helps limiting infrastructure costs.

Finally, it would be useful to complement the definition of alternative modal distribution scenarios induced by the implementation of a LEZ with an impact assessment on the utilization rate of transport infrastructures.

#### 4.2.2 Validity of health impact assessment

The validity of health impact assessment studies strongly depends on input parameters, namely exposure data, concentration-response functions, and health outcomes data. Concerning the



exposure data, we used the fine-scale (10-m grid) air pollutant-dispersion SIRANE model, based on a network street approach allowing to evaluate air pollutant concentrations inside streets in urban canopy. In addition to the primary validation of the SIRANE model,<sup>59</sup> a validation of model estimates is routinely conducted with Delta Tool.<sup>60</sup> The very fine spatial resolution is a major strength, as relying on less fine models can lead to underestimating of health impacts.<sup>61</sup> We developed a specific approach (MECANO) allowing to estimate the impact of various reductions of emissions in the traffic and residential heating sectors on the PM<sub>2.5</sub> population-weighted exposure, as illustrated Figure 2. This approach allowed to quickly provide a set of combinations of emission reductions in each sector entailing a given change in PM<sub>2.5</sub> exposure, and thus likely to reach the targeted health improvement, or beyond. We have checked its validity for the most ambitious scenario considered by running the dispersion model with the emission reductions identified by MECANO, and obtained a very low difference (0.01 µg/m<sup>3</sup>), not raising concern regarding the validity of MECANO approach.

We exclusively relied on meta-analytical risks, which reflect the whole evidence of the literature and are likely to be more robust and less biased than relative risks from individual studies. We assumed these concentration-response functions to be linear, which seemed to be reasonable for air pollution effects within the exposure range in this study.<sup>62</sup> Woodcock et al.<sup>63</sup> showed that the slope of the relationship of physical activity to mortality flattened for long durations of physical activity; with regard to the increases in physical activity expected under our mobility scenarios, we used the relative risks estimated in starting ranges of MET.hours. In the absence of specific dose-response functions regarding E-bikes, we applied the relative risks established for cycling while using E-bikes-specific MET and speed.

Another strength of our health impact assessment consists in the reliability of the health data sources, namely national institutions for public health and for road safety as well as the national health insurance system. Although non-accidental mortality rates and traffic injury

cases could be obtained at the municipality scale, lung cancer and cardiovascular diseases incidence data were available at the local scale (the *department*), which might give more weight to random spatial fluctuations than for the formers. Lastly, we focused on the main health outcomes related to air pollution exposure; given the existence of other health conditions possibly associated with air pollution (e.g., breast cancer, mental disorders, diabetes mellitus...),<sup>64-66</sup> the total health benefit and related economic savings are probably underestimated.

#### 4.2.3 Assumptions related to the cost-benefit analysis

Several assumptions should be subjected to a sensitivity analysis: the intangible costs, in particular the value of human life, whose values vary considerably in the literature,<sup>67</sup> the value of time (VoT) specific to each mode of transport, the external costs for noise, climate change (since the social value of carbon is the subject of lively academic debates<sup>68</sup>) and the discount rate suitable to deal with environmental issues, whose choice has a significant impact on the valuation of future benefits.<sup>69</sup>

The promotion and deployment of alternatives to the car will profoundly change the context in which the determinants of mobility choices will apply. The values of a certain number of behavioral and economic parameters used in this study will evolve under the effect of a change in context. The aggregate impact of modal shifts on the valuation of time spent in transport is a key determinant of the results of the cost-benefit analysis. The values taken here are based on literature and on the current valuation of this time as it is perceived today with urban transport infrastructures that are still largely oriented towards the private car and not very well adapted to cycling. It is likely that the adaptation of infrastructures and the development of bicycle paths truly separated from the roadways reserved for cars and of bicycle highways will change the perception of the time spent cycling, and thus decrease its valuation. Therefore,

the perceived value of time can vary significantly between individuals, depending on their relationship to active modes of transport. Furthermore, in the sub-scenarios in which the modal shift toward active modes was highest, we can reasonably assume that the last people to adopt these modes are those who enjoy it least, and thus who have a higher value of time when traveling by active modes. However, we also assumed a learning effect of active modes of transport and a reduction in car traffic, and thereby a growing pleasure, which will reduce the perceived value of time spent cycling. The two effects could compensate each other and lead to average time values for cycling not very far from the current values found in the literature.

#### **4.3 Involvement of local decision-makers in the definition of health objectives used to define urban policies**

The participatory approach developed as part of the MobilAir project has made it possible to involve the elected representatives of the local urban authority in the project, with the aim of encouraging the appropriation of the study results by decision-makers, and the implementation of the necessary urban policies. In particular, local decision-makers defined the health objectives. This was done as part of a process of exchange between the research consortium and the services and elected officials of the local authority. The smooth running of the project and its success required interactions with decision-makers to identify and discuss the relevance of sectoral measures on an urban scale on road traffic or wood heating. This took the form of a committee including high-level decision-makers that met once a year, in addition with more frequent meetings of the scientists with lower-level technical staff. During the first decision-makers' committee, we presented the health impact of air pollution in the urban area, the general objectives of the project as well as the first results on the contribution of the main sectors emitting PM and their geographical origin. Following the first committee, the decision-makers asked the research consortium to study in detail three health objectives (the -33%, -50% and -

67% reductions in PM<sub>2.5</sub>-related mortality). On this basis, an exploration of sectoral measures in relation to these health objectives was carried out and presented to the second committee of decision-makers. Following this, the elected representatives asked to explore and evaluate the scenarios of urban policies presented here.

#### **4.4 Interactions between climate change and air pollution**

Interactions between climate change and PM-related air pollution are complex and numerous.<sup>70</sup> Synergies and antagonisms exist between policies to reduce pollution and policies to reduce greenhouse gases. It is important that policies implemented to reduce PM-related pollution do not lead to increases in greenhouse gases, and vice-versa. The case of wood heating, using a renewable energy but emitting fine particulate matter if old stoves are used, provides an illustration of the related challenges.

### **5 Conclusion**

This inter-disciplinary work between epidemiologists, air pollution modelers and economists aimed at defining and evaluating scenarios of urban policies to significantly reduce air pollution and its impacts. To our knowledge, such a reverse and forward approach has not previously been carried out at the city or national level.

We showed that the most ambitious scenarios in terms of targeted health improvement were also the most interesting from an economic point of view, particularly if the urban policies lead to strong increases in active mobility, the health benefits thereof proving to be much greater than those directly induced by the reduction of air pollution. The policies allowing to achieve these objectives were realistic, relying on behavioral mobility changes that are also realistic.

Policies simultaneously relying on the full replacement of highly polluting wood-heating facilities and shifts from private car to bike mobility for short trips are likely to allow benefits of 4.2 to 6 Euro for each Euro invested and generate net annual benefits per person of €484 to €629.

## 6 References

1. Hamra GB, Guha N, Cohen A, et al. Outdoor particulate matter exposure and lung cancer: a systematic review and meta-analysis. *Environ Health Perspect.* 2014;122(9):906-911. doi:10.1289/ehp/1408092
2. Pope CA, Burnett RT, Thurston GD, et al. Cardiovascular Mortality and Long-Term Exposure to Particulate Air Pollution: Epidemiological Evidence of General Pathophysiological Pathways of Disease. *Circulation.* 2004;109(1):71-77. doi:10.1161/01.CIR.0000108927.80044.7F
3. World Health Organization. *WHO Expert Meeting: Methods and Tools for Assessing the Health Risks of Air Pollution Local, National and International Level – Meeting Report.* WHO Regional Office for Europe; 2014:112. Accessed March 24, 2015. <http://www.euro.who.int/en/health-topics/environment-and-health/air-quality/publications/2014/who-expert-meeting-methods-and-tools-for-assessing-the-health-risks-of-air-pollution-at-local,-national-and-international-level>
4. World Health Organization. *World Health Statistics 2019: Monitoring Health for the SDGs : Sustainable Development Goals.* World Health Organization; 2019.
5. Cohen AJ, Brauer M, Burnett R, et al. Estimates and 25-year trends of the global burden of disease attributable to ambient air pollution: an analysis of data from the Global Burden of Diseases Study 2015. *The Lancet.* 2017;389(10082):1907-1918. doi:10.1016/S0140-6736(17)30505-6
6. European Environment Agency. *Air Quality in Europe: 2020 Report.* Publications Office; 2020. Accessed April 16, 2021. <https://data.europa.eu/doi/10.2800/786656>
7. Medina S, Adélaïde L, Wagner V, et al. *Impact de la pollution de l'air ambiant sur la mortalité en France métropolitaine. Réduction en lien avec le confinement du printemps 2020 et nouvelles données sur le poids total pour la période 2016-2019.* Santé Publique France; 2021:63 pages. Accessed May 19, 2021. <https://www.santepubliquefrance.fr/determinants-de-sante/pollution-et-sante/air/documents/enquetes-etudes/impact-de-pollution-de-l-air-ambiant-sur-la-mortalite-en-france-metropolitaine.-reduction-en-lien-avec-le-confinement-du-printemps-2020-et-nouvelle>
8. French Senate. *Commission d'enquête sur le coût économique et financier de la pollution de l'air.* French Senate; 2015:306 pages. Accessed April 19, 2020. <http://www.senat.fr/rap/r14-610-1/r14-610-11.pdf>
9. Clancy L, Goodman P, Sinclair H, Dockery DW. Effect of air-pollution control on death rates in Dublin, Ireland: an intervention study. *The Lancet.* 2002;360(9341):1210-1214. doi:10.1016/S0140-6736(02)11281-5
10. Hara K, Homma J, Tamura K, Inoue M, Karita K, Yano E. Decreasing trends of suspended particulate matter and PM2.5 concentrations in Tokyo, 1990-2010. *J Air Waste Manag Assoc* 1995. 2013;63(6):737-748.
11. Cyrus J, Peters A, Soentgen J, Wichmann H-E. Low emission zones reduce PM10 mass concentrations and diesel soot in German cities. *J Air Waste Manag Assoc.* 2014;64(4):481-487. doi:10.1080/10962247.2013.868380

12. Fensterer V, Küchenhoff H, Maier V, et al. Evaluation of the impact of low emission zone and heavy traffic ban in Munich (Germany) on the reduction of PM<sub>10</sub> in ambient air. *Int J Environ Res Public Health*. 2014;11(5):5094-5112. doi:10.3390/ijerph110505094
13. Cour des comptes. *Les politiques de lutte contre la pollution de l'air - Enquête demandée par la Commission des finances du Sénat.*; 2020:202 pages. [https://www.ccomptes.fr/system/files/2020-09/20200923-58-2-rapport-politiques-lutte-contre-pollution-air\\_0.pdf](https://www.ccomptes.fr/system/files/2020-09/20200923-58-2-rapport-politiques-lutte-contre-pollution-air_0.pdf)
14. European Court of Justice. *Manquement d'État – Environnement – Directive 2008/50/CE – Qualité de l'air ambiant – Article 13, paragraphe 1, et annexe XI – Dépassement systématique et persistant des valeurs limites pour le dioxyde d'azote (NO<sub>2</sub>) dans certaines zones et agglomérations françaises – Article 23, paragraphe 1 – Annexe XV – Période de dépassement "la plus courte possible" – Mesures appropriées.*, 15 pages (2019). Accessed October 23, 2020. <https://www.dalloz-actualite.fr/flash/france-condamnee-par-cjue-pour-pollution-de-l-air-par-dioxyde-d-azote#.X5LbZy8it61>
15. ADEME, Pouponneau M, RINCENT AIR, Forestier B, Cape F. *Les Zones à Faibles Émissions (Low Emission Zones) à Travers l'Europe : Déploiement, Retours d'expériences, Évaluation d'impacts et Efficacité Du Système.* ADEME, Rincenc Air, AJBD; 2018:112 pages. Accessed August 3, 2018. <https://www.ademe.fr/zones-a-faibles-emissions-low-emission-zones-lez-a-travers-leurope>
16. Cesaroni G, Boogaard H, Jonkers S, et al. Health benefits of traffic-related air pollution reduction in different socioeconomic groups: the effect of low-emission zoning in Rome. *Occup Environ Med*. 2012;69(2):133-139. doi:10.1136/oem.2010.063750
17. Kelly P, Kahlmeier S, Götschi T, et al. Systematic review and meta-analysis of reduction in all-cause mortality from walking and cycling and shape of dose response relationship. *Int J Behav Nutr Phys Act*. 2014;11(1):1-15. doi:10.1186/s12966-014-0132-x
18. Kyu HH, Bachman VF, Alexander LT, et al. Physical activity and risk of breast cancer, colon cancer, diabetes, ischemic heart disease, and ischemic stroke events: systematic review and dose-response meta-analysis for the Global Burden of Disease Study 2013. *BMJ*. 2016;354:i3857-i3866. doi:10.1136/bmj.i3857
19. Belojevic G, Jakovljevic B, Stojanov V, Paunovic K, Ilic J. Urban road-traffic noise and blood pressure and heart rate in preschool children. *Environ Int*. 2008;34(2):226-231. doi:10.1016/j.envint.2007.08.003
20. Bodin T, Albin M, Ardö J, Stroh E, Östergren P-O, Björk J. Road traffic noise and hypertension: results from a cross-sectional public health survey in southern Sweden. *Environ Health*. 2009;8(1):38. doi:10.1186/1476-069X-8-38
21. Gidlöf-Gunnarsson A, Öhrström E. Noise and well-being in urban residential environments: The potential role of perceived availability to nearby green areas. *Landsc Urban Plan*. 2007;83(2-3):115-126. doi:10.1016/j.landurbplan.2007.03.003
22. Woodward A, Smith KR, Campbell-Lendrum D, et al. Climate change and health: on the latest IPCC report. *The Lancet*. 2014;383(9924):1185-1189. doi:10.1016/S0140-6736(14)60576-6
23. Kalsch H, Hennig F, Moebus S, et al. Are air pollution and traffic noise independently associated with atherosclerosis: the Heinz Nixdorf Recall Study. *Eur Heart J*. 2014;35(13):853-860. doi:10.1093/eurheartj/eh426

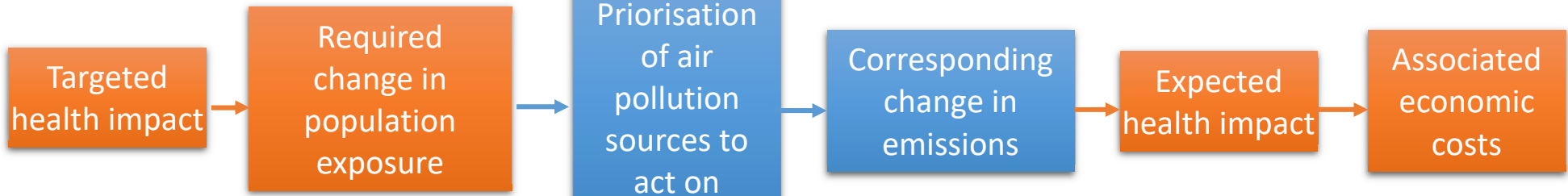
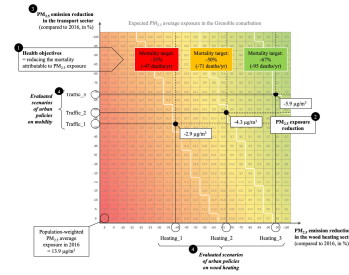
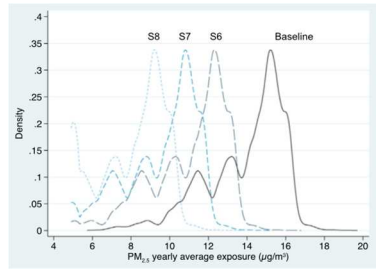
24. US EPA. *The Benefits and Costs of the Clean Air Act from 1990 to 2020 - Final Report - Rev. A*. US Environmental Protection Agency; 2011:238 pages.
25. Morelli X, Gabet S, Rieux C, Bouscasse H, Mathy S, Slama R. Which decreases in air pollution should be targeted to bring health and economic benefits and improve environmental justice? *Environ Int*. 2019;129:538-550. doi:10.1016/j.envint.2019.04.077
26. Insee. Population en 2014. Published October 24, 2017. Accessed June 4, 2018. <https://www.insee.fr/fr/statistiques/3137409>
27. Atmo AuRA. *Bilan Qualité de l'Air 2017 – Isère*. Atmo Auvergne-Rhône-Alpes; 2018:29 pages. <https://www.atmo-auvergnerhonealpes.fr/article/bilan-2017-isere>
28. Menut L, Bessagnet B, Khvorostyanov D, et al. CHIMERE 2013: a model for regional atmospheric composition modelling. *Geosci Model Dev*. 2013;6(4):981-1028. doi:10.5194/gmd-6-981-2013
29. Soulhac L, Salizzoni P, Cierco F-X, Perkins R. The model SIRANE for atmospheric urban pollutant dispersion; part I, presentation of the model. *Atmos Environ*. 2011;45(39):7379-7395. doi:10.1016/j.atmosenv.2011.07.008
30. IGN. BDTOPO®. Published 2017. Accessed November 24, 2017. <http://professionnels.ign.fr/bdtopo>
31. Simpson D, Benedictow A, Berge H, et al. The EMEP MSC-W chemical transport model – technical description. *Atmospheric Chem Phys*. 2012;12(16):7825-7865. doi:10.5194/acp-12-7825-2012
32. Hilal M, Fugerey-Scarbel A, Gojard S. *Odomatrix : Mesurer des temps de trajet pour améliorer les conditions de vie.*; 2018:17 pages. <https://hal.archives-ouvertes.fr/hal-02068383>
33. INCa. Les Données du cancer - Épidémiologie - Incidence régionale et départementale 2007-2016. Published November 18, 2019. Accessed April 20, 2020. <http://lesdonnees.e-cancer.fr/Themes/epidemiologie/Incidence-regionale-et-departementale>
34. French health insurance system. Données statistiques - Affection de longue durée (ALD) - Incidence. Published December 18, 2018. Accessed May 13, 2019. <https://www.ameli.fr/l-assurance-maladie/statistiques-et-publications/donnees-statistiques/affection-de-longue-duree-ald/incidence/incidence-des-ald-en-2017.php>
35. US EPA. *Uncertainty Analyses to Support the Second Section 812 Benefit-Cost Analysis of the Clean Air Act*. US Environmental Protection Agency; 2010:200 pages.
36. Ainsworth BE, Haskell WL, Herrmann SD, et al. 2011 Compendium of Physical Activities: a second update of codes and MET values. *Med Sci Sports Exerc*. 2011;43(8):1575-1581. doi:10.1249/MSS.0b013e31821ece12
37. Berntsen S, Malnes L, Langåker A, Bere E. Physical activity when riding an electric assisted bicycle. *Int J Behav Nutr Phys Act*. 2017;14(1):1-7. doi:10.1186/s12966-017-0513-z
38. ONISR. Base de données accidents corporels de la circulation. Published June 2018. Accessed May 13, 2019. <https://www.data.gouv.fr/fr/datasets/base-de-donnees-accidents-corporels-de-la-circulation>



39. Insee. Espérance de vie à divers âges - Données annuelles de 1994 à 2019. Published January 14, 2020. Accessed April 7, 2020. <https://www.insee.fr/fr/statistiques/2416631#tableau-figure1>
40. World Health Organization. *WHO Expert Meeting: Methods and Tools for Assessing the Health Risks of Air Pollution Local, National and International Level – Meeting Report*. WHO Regional Office for Europe; 2014:112. Accessed March 24, 2015. <http://www.euro.who.int/en/health-topics/environment-and-health/air-quality/publications/2014/who-expert-meeting-methods-and-tools-for-assessing-the-health-risks-of-air-pollution-at-local,-national-and-international-level>
41. Hamra GB, Guha N, Cohen A, et al. Outdoor particulate matter exposure and lung cancer: a systematic review and meta-analysis. *Environ Health Perspect*. 2014;122(9):906-911. doi:10.1289/ehp/1408092
42. Cesaroni G, Forastiere F, Stafoggia M, et al. Long term exposure to ambient air pollution and incidence of acute coronary events: prospective cohort study and meta-analysis in 11 European cohorts from the ESCAPE Project. *BMJ*. 2014;348(jan21 3):f7412-f7412. doi:10.1136/bmj.f7412
43. Stafoggia M, Cesaroni G, Peters A, et al. Long-Term Exposure to Ambient Air Pollution and Incidence of Cerebrovascular Events: Results from 11 European Cohorts within the ESCAPE Project. *Environ Health Perspect*. 2014;122(9):919-925. doi:10.1289/ehp.1307301
44. Haute Autorité de Santé. *Choix Méthodologiques Pour l'évaluation Économique à La HAS*. HAS.; 2011.
45. Chanel O. *Aphekom – Guidelines on Monetary Cost Calculations Related to Air-Pollution Health Impacts – Deliverable D6*. National Center for Scientific Research (CNRS), GREQAM and IDEP; 2011:45 pages.
46. Chevalier J, Delaitre O, Hammès F, de Pouvourville G. Cost-effectiveness of dabigatran versus vitamin K antagonists for the prevention of stroke in patients with atrial fibrillation: A French payer perspective. *Arch Cardiovasc Dis*. 2014;107(6-7):381-390. doi:10.1016/j.acvd.2014.04.009
47. de Pouvourville G. Coût de la prise en charge des accidents vasculaires cérébraux en France. *Arch Cardiovasc Dis Suppl*. 2016;8(2):161-168. doi:10.1016/S1878-6480(16)30330-5
48. ONISR. *La sécurité routière en France - Bilan de l'accidentalité de l'année 2017*. ONISR; 2018.
49. Quinet E. *L'évaluation socioéconomique des investissements publics - Tome 1 - Rapport final*. Commissariat général à la stratégie et à la prospective; 2013:354 pages. <http://www.strategie.gouv.fr>
50. CGDD. *Analyse coûts bénéfiques des véhicules électriques - Les voitures*. General Commission for Sustainable Development - French Ministry of Ecological Transition; 2017:97 pages. <http://temis.documentation.developpement-durable.gouv.fr/document.html?id=Temis-0086746>
51. Ricke K, Drouet L, Caldeira K, Tavoni M. Country-level social cost of carbon. *Nat Clim Change*. 2018;8(10):895-900. doi:10.1038/s41558-018-0282-y

52. Bouscasse H, de Lapparent M. Perceived comfort and values of travel time savings in the Rhône-Alpes Region. *Transp Res Part Policy Pract.* 2019;124:370-387. doi:10.1016/j.tra.2019.04.006
53. Litman T. Valuing Transit Service Quality Improvements. *J Public Transp.* 2008;11(2):43-63. doi:10.5038/2375-0901.11.2.3
54. Papon F. Le vélo, un mode de déplacement super économique. Published online 2013:7.
55. Litman T. *Guide to Calculating Mobility Management Benefits.* Victoria Transport Policy Institute; 2019:49 pages. Accessed January 12, 2021. <https://vtpi.org/tmben.pdf>
56. SMTC. *Agglomération grenobloise - Plan de Déplacements Urbains 2030.*; 2018:334 pages. <http://www.smtc-grenoble.org/le-pdu-horizon-2030>
57. Grenoble-Alpes Métropole. *Délibération du Conseil métropolitain - Séance du 28 septembre 2018 - Objet : Environnement, Air, Climat et Biodiversité - Prime air bois - Bilan et évolution du dispositif d'aide au renouvellement des appareils de chauffage au bois non performants.*; 2018:6 pages. Accessed January 12, 2021. <https://www.grenoblealpesmetropole.fr/18-deliberations.htm>
58. Bamberg S. Chapter 9 Understanding and Promoting Bicycle Use – Insights from Psychological Research. In: Parkin J, ed. *Transport and Sustainability.* Emerald Group Publishing Limited; 2012:219-246. doi:10.1108/S2044-9941(2012)0000001011
59. Soulhac L, Salizzoni P, Mejean P, Didier D, Rios I. The model SIRANE for atmospheric urban pollutant dispersion; PART II, validation of the model on a real case study. *Atmos Environ.* 2012;49:320-337. doi:10.1016/j.atmosenv.2011.11.031
60. EU's Joint Research Centre.  $\Delta$  DELTA. Published June 18, 2019. Accessed June 18, 2019. <https://aqm.jrc.ec.europa.eu/index.aspx>
61. Morelli X, Rieux C, Cyrus J, Forsberg B, Slama R. Air pollution, health and social deprivation: A fine-scale risk assessment. *Environ Res.* 2016;147:59-70. doi:10.1016/j.envres.2016.01.030
62. Burnett R, Chen H, Szyszkowicz M, et al. Global estimates of mortality associated with long-term exposure to outdoor fine particulate matter. *Proc Natl Acad Sci.* 2018;115(38):9592-9597. doi:10.1073/pnas.1803222115
63. Woodcock J, Franco OH, Orsini N, Roberts I. Non-vigorous physical activity and all-cause mortality: systematic review and meta-analysis of cohort studies. *Int J Epidemiol.* 2011;40(1):121-138. doi:10.1093/ije/dyq104
64. White AJ, Bradshaw PT, Hamra GB. Air Pollution and Breast Cancer: a Review. *Curr Epidemiol Rep.* 2018;5(2):92-100. doi:10.1007/s40471-018-0143-2
65. Buoli M, Grassi S, Caldiroli A, et al. Is there a link between air pollution and mental disorders? *Environ Int.* 2018;118:154-168. doi:10.1016/j.envint.2018.05.044
66. Eze IC, Hemkens LG, Bucher HC, et al. Association between Ambient Air Pollution and Diabetes Mellitus in Europe and North America: Systematic Review and Meta-Analysis. *Environ Health Perspect.* 2015;123(5):381-389. doi:10.1289/ehp.1307823
67. Chanel O, Medina S, Pascal M. Évaluation économique de la mortalité liée à la pollution atmosphérique en France. *J Health Care Manag Econ.* 2020;38(2):77-92.

68. Pindyck RS. The social cost of carbon revisited. *J Environ Econ Manag*. 2019;94:140-160. doi:10.1016/j.jeem.2019.02.003
69. Guesnerie R, Stern N. Introduction to the Special Issue on Managing Climate Change. *J Public Econ Theory*. 2012;14(2):189-196. doi:10.1111/j.1467-9779.2011.01538.x
70. von Schneidemesser E, Monks PS, Allan JD, et al. Chemistry and the Linkages between Air Quality and Climate Change. *Chem Rev*. 2015;115(10):3856-3897. doi:10.1021/acs.chemrev.5b00089



Decision makers

Scientists

Decision makers

Scientists