

In vitro assessment of polychlorinated biphenyl bioaccessibility in meat: Influence of fat content, cooking level and consumer age on consumer uptake

Christelle Planche, Jérémy Ratel, Frédéric Mercier, Cheng Zhang, Magaly Angénieux, Patrick Blinet, Philippe Marchand, Gaud Dervilly, Isabelle Albert,

Jessica Tressou, et al.

▶ To cite this version:

Christelle Planche, Jérémy Ratel, Frédéric Mercier, Cheng Zhang, Magaly Angénieux, et al.. In vitro assessment of polychlorinated biphenyl bioaccessibility in meat: Influence of fat content, cooking level and consumer age on consumer uptake. Food Chemistry, 2022, 374, pp.131623. 10.1016/j.foodchem.2021.131623. hal-03441539

HAL Id: hal-03441539 https://hal.science/hal-03441539

Submitted on 9 May 2023

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers. L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

1	In vitro assessment of polychlorinated biphenyl bioaccessibility in meat:
2	Influence of fat content, cooking level and consumer age on consumer
3	uptake
4	
5	
6	Christelle Planche ^{a,b,c} , Jérémy Ratel ^a , Frédéric Mercier ^a , Cheng Zhang ^a , Magaly Angénieux ^a ,
7	Patrick Blinet ^a , Philippe Marchand ^d , Gaud Dervilly ^d , Isabelle Albert ^e , Jessica Tressou ^e ,
8	Laurent Debrauwer ^{b,c} , Erwan Engel ^{a*}
9	
10	
11	^a MASS group, UR QuaPA, INRAE, F-63122 Saint-Genès-Champanelle, France
12	^b Toxalim, Université de Toulouse, INRAE, ENVT, INP-EI-Purpan, Université de Toulouse 3
13	Paul Sabatier, F-31027 Toulouse, France
14	^c Axiom Platform, UMR Toxalim, MetaToul-MetaboHUB, National Infrastructure of
15	Metabolomics and Fluxomics, F-31027
16	^d LABERCA, Oniris, INRAE, Nantes, France
17	^e MIA, UMR INRAE, AgroParisTech, Université Paris Saclay, Paris 518, Paris, France.
18	
19	
20	* Corresponding Author:
21	Erwan Engel: Email: erwan.engel@ inrae.fr Tel: +33 4 73624589
22	
~~	
23	

24 Abstract

In a risk assessment perspective, this work aims to assess the bioaccessibility of PCBs in 25 meat. A standardised in vitro static digestion protocol was set up and coupled with extraction, 26 27 clean-up and GC×GC-ToF/MS multianalyte method to monitor the fate of PCBs in meat during digestion. Starting with spiked meat, PCB bioaccessibility in 11% fat medium-cooked 28 meat varied in adults from 20.6% to 30.5% according to congeners. PCB bioaccessibility 29 increased to 44.2-50.1% in 5% fat meat and decreased to 6.2-9.1% and to 14.6-19.4% in 30 digestion conditions mimicking infants and elderly, respectively. Intense cooking also 31 decreased PCB bioaccessibility to 18.0-26.7%. Bioaccessibility data obtained with spiked 32 33 meat were validated with measurements carried out in incurred meat samples. Finally, mean uptake distributions are obtained from a modular Bayesian approach. These distributions 34 feature a lower mode when the fat content is higher, the meat is well-done cooked, and the 35 36 consumers are older.

37

38 Highlights

• An *in vitro* digestion protocol was set up to assess PCB bioaccessibility in meat

• The mean PCB bioaccessibility was 26% in 11% fat ground beef

41 • PCB bioaccessibility and mean uptakes increased when the fat level decreased in meat

42 • PCB bioaccessibility and mean uptakes decreased when meat is well-done cooked

43 • PCB bioaccessibility and mean uptakes were lower in elderlies than in adults

44

45 Keywords

Bioaccessibility; Polychlorinated biphenyls; Meat; *In vitro* digestion; Comprehensive twodimensional gas chromatography-time-of-flight mass spectrometry (GC×GC-ToF/MS);
Dietary uptake

50

Many micropollutants may be found in food, and in particular in products of animal 51 52 origin (Dervilly-Pinel et al., 2017). Among these micropollutants are polychlorinated biphenyls (PCBs), persistent organic pollutants (POPs) massively used as insulators until the 53 1980s. PCBs accumulate in animal tissues during growth, and may end up in meat products 54 intended for human consumption. The consumption of food of animal origin is considered as 55 one of the main sources of human exposure to PCBs (Tressou et al., 2017). While chemical 56 risk is generally assessed from the total concentration of micropollutants in fresh food, the 57 58 bioaccessible fraction of the micropollutants present in food (i.e. the fraction liable to cross the intestinal barrier and to induce toxic effects) can be considerably lower than the total 59 amount (Engel et al., 2015). This is due to both the changes occurring during food processing 60 61 (Planche et al., 2017), and the physiological processes of the digestion modifying the amounts of contaminants truly available for absorption in the systemic circulation. Therefore, the 62 63 bioaccessibility of micropollutants must be considered to make an accurate assessment of 64 their impact on human health (Tressou et al., 2017).

To study the bioaccessibility of micropollutants such as PCBs, in vitro digestion models 65 are largely used because they offer the advantage of being faster, better controlled and less 66 67 costly than in vivo methods, while not being restricted by ethical considerations. However, marked variations are observed among the different in vitro digestion models described in the 68 literature (incubation time, pH, constituents and concentrations of digestive and intestinal 69 70 solutions, etc.), thus yielding widely disparate results. For example, Oomen et al. (2003) studied pollutant bioaccessibility in soils with an *in vitro* digestion model and they observed 71 different results according to the type of bile salts used, with for example a bioaccessibility of 72 73 lead 3-5.5 times greater when chicken bile was used compared to pig and ox biles. More

recently, He *et al.* (2018) observed a near-doubling of triphenyl phosphate (TPP)
bioaccessibility when the intestinal digestion time increased from 4h to 8h.

To allow comparing results from different laboratories, a network of scientists proposed 76 77 a standardised international protocol to mimic as closely as possible the physiological conditions of digestion (Minekus et al., 2014). Minekus et al. (2014) have setup a consensus 78 in vitro protocol consisting in three steps, namely oral, gastric and intestinal digestion. This 79 protocol has been developed specifically for food matrices, which may explain differences 80 observed in terms of ratios of matrix to digestive fluids or even in terms of sample stirring rate 81 when compared to other protocols such as the protocol developed by Oomen et al. (2003) 82 83 dedicated to soil samples. The consensus protocol developed by Minekus et al. (2014) has already been used to assess the bioaccessibility of food contaminants such as polybrominated 84 diphenyl ethers (Cruz et al., 2020) or pesticides (Milinčić et al., 2020), but to the best of our 85 86 knowledge, it has never been adapted to the study of PCBs in food.

To assess the bioaccessibility of PCBs in food after digestion, appropriate analytical methods are required for the monitoring and quantification of contaminants. Multianalyte methods are of particular interest because they allow the simultaneous monitoring of many substances in a single run. Several studies have shown the usefulness of comprehensive twodimensional gas phase chromatography (GC×GC) in tandem with time-of-flight mass spectrometry (ToF-MS) to analyse a great number of different PCB congeners (Planche *et al.*, 2015).

Various factors are liable to influence the bioaccessibility of contaminants in meat, among which the composition of the raw material, and in particular its fat content (Xing *et al.*, 2008; Yu *et al.*, 2010; Shen *et al.*, 2016). Food processing steps such as cooking could also cause variations in bioaccessibility values (He *et al.*, 2010; Shen *et al.*, 2016). Lastly, agerelated physiological variations in the human digestion process exist, in particular regarding 99 concentrations of digestive enzymes: the age of consumers could therefore also affect
100 bioaccessibility values (Dupont *et al.*, 2010; Levi *et al.*, 2014).

With the ultimate aim of making a better assessment of risks due to chemical 101 102 contamination of food, we have determined the bioaccessibility of PCBs in meat. For this purpose, the first part of this paper was focused on the set up of an *in vitro* digestion protocol 103 coupled with extraction, clean-up and GC×GC-ToF/MS multianalyte analysis to monitor the 104 fate of the 18 most relevant PCBs in meat during digestion. In a second part, this method was 105 used to assess the influence of fat level, cooking intensity and the age of consumers on the 106 bioaccessibility of PCBs in meat as well as on the uptake to these micropollutants due to beef 107 108 consumption. The validity of these results obtained with intentionally contaminated (spiked) 109 meat is discussed in the light of measurements carried out on naturally contaminated (incurred) samples. 110

111

112 **2.** Materials and Methods

113

2.1. Chemicals and standards

114

Hexane, dichloromethane, acetone and toluene were organic trace analysis grade 115 solvents (Sigma-Aldrich, Saint-Quentin Fallavier, France). For in vitro digestion, constituents 116 of simulated salivary fluid, simulated gastric fluid and simulated intestinal fluid (KCl, 117 KH₂PO₄, NaHCO₃, NaCl, MgCl₂(H₂O)₆, (NH₄)₂CO₃), together with HCl, NaOH, CaCl₂, 118 pepsin from porcine gastric mucosa (P6887), pancreatin from porcine pancreas (P7545) and 119 porcine bile extract (B8631) were from Sigma-Aldrich. For PCB extraction, trichloroacetic 120 121 acid and activated aluminium oxide (acidic, Brockmann I) were from Sigma-Aldrich. Diatomaceous earth was obtained from Thermo Fisher Scientific (Waltham, MA, USA). PCB 122 reference standards including 12 dioxin-like PCBs (3.3'.4.4'-Tetrachlorobiphenyl, 3.4.4'.5-123

124	Tetrachlorobiphenyl, 2.3.3'.4.4'-Pentachlorobiphenyl, 2.3.4.4'.5-Pentachlorobiphenyl,
125	2.3'.4.4'.5-Pentachlorobiphenyl, 2'.3.4.4'.5-Pentachlorobiphenyl, 3.3'.4.4'.5-
126	Pentachlorobiphenyl, 2.3.3'.4.4'.5-Hexachlorobiphenyl, 2.3.3'.4.4'.5'-Hexachlorobiphenyl,
127	2.3'.4.4'.5.5'-Hexachlorobiphenyl, 3.3'.4.4'.5.5'-Hexachlorobiphenyl, 2.3.3'.4.4'.5.5'-
128	Heptachlorobiphenyl) and 6 non-dioxin-like PCBs (2.4.4'-Trichlorobiphenyl, 2.2'.5.5'-
129	Tetrachlorobiphenyl, 2.2'.4.5.5'-Pentachlorobiphenyl, 2.2'.3.4.4'.5'-Hexachlorobiphenyl,
130	2.2'.4.4'.5.5'-Hexachlorobiphenyl, 2.2'.3.4.4'.5.5'-Heptachlorobiphenyl) were from
131	AccuStandard Europe (Niederbipp, Switzerland). Internal standards were used during in vitro
132	digestion and PCB extraction for the accurate quantification of target compounds: 3'-F-PCB-
133	28, 3-F-PCB-52, 3'-F-PCB-81, and 5'-F-PCB-156 (Chiron, Trondheim, Norway), ¹³ C-labeled
134	PCB-111 and ¹³ C-labeled PCB-194 (Wellington laboratories, Guelph, ON, Canada).
135	
136	2.2. Meat samples
137	
138	Two types of meat samples were used: spiked and incurred samples. Spiked meat was
139	prepared with ground beef samples from the same blend of muscles (<15% fat) purchased
140	from a French supplier. The exact fat level in these samples was measured at 11%. Aliquots
141	weighing 125 g were stored at -80 °C before use. Matrix blanks of these samples were made
142	before spiking. Incurred beef meat samples were obtained under the French project SOMEAT
143	(Contract No. ANR-12-ALID-0004. Safety of Organic Meat, available at www.so-meat.fr),
144	and were ground before use.
145	
146	2.3. Spiking and cooking
147	2.3.1. Sample spiking
148	

7

Ground beef was spiked according to Planche *et al.* (2015), combining contaminant addition to ground meat with matrix homogenisation. Briefly, ground beef (120 g) was immersed in dichloromethane (DCM) containing the 18 most relevant PCBs in meat (Sirot *et al.*, 2012), the mixture evaporated down under a hood, and the residue homogenised for 2 min in a blender. A spiking concentration of 20 ng.g⁻¹ of fresh meat was selected for each congener to give PCB concentrations in ready-to-run samples within the range of linearity of GC×GC-ToF/MS for these compounds.

- 156
- 157

2.3.2. Cooking method

158

159 To study the effect of cooking on the bioaccessibility of PCBs in meat, circular small ground beef patties weighing 26 g (2.5 cm thick) were then shaped to copy commercial ground 160 161 beef patties. These ground beef patties were cooked in a stainless-steel frying pan (diameter 17 cm) on a temperature-controlled induction hob (Bosch Electroménager, Saint-Ouen, France) 162 163 according to Planche et al. (2017). Briefly, aluminium foil (11 µm thickness) was placed on the 164 bottom of the frying pan to recover juices released during meat cooking. Three different cooking conditions were used to simulate rare (50 °C at the core), medium (70 °C at the core 165 according to WHO recommendations for ground meat) and well-done (85 °C at the core) meat 166 167 (n = 3 for each cooking condition). These cooking conditions were obtained by 7 min heating (turned once) at 160 °C at the bottom of the pan, 14 min heating (turned three times) at 200 °C 168 at the bottom of the pan, and 14 min heating (turned three times) at 250 °C at the bottom of the 169 170 pan, respectively. Temperatures at the core of the meat and at the bottom of the pan were measured with thermocouples (RS Components, Beauvais, France). Before in vitro digestion, 171 the cooked meat was minced in a blender to simulate mastication. 172

173

2.3.3. Determination of fat content

175

Fat content was determined according to Blanchet-Letrouvé *et al.* (2014) with slight modifications. Raw and cooked meat were first freeze-dried. Aliquots (1 g) of the resulting powder were then extracted by accelerated solvent extraction (ASE) using a Dionex ASE 350 extractor (Sunnyvale, CA, USA) with 22 mL stainless-steel extraction cells. Toluene-acetone (70:30) was used as extraction solvent at a temperature of 120 °C and pressure of 1500 psi with three extraction cycles per sample. The extracts obtained were then evaporated down under a hood and weighed to determine fat content.

- 183
- 184

2.4. In vitro digestion

185

186 In vitro adult digestions (n = 3 for each condition) were performed according to Minekus et al. (2014). Firstly, to simulate the oral phase, 5 g of raw or cooked meat was 187 188 mixed with 3.5 mL of simulated salivary fluid (SSF), 25 µL of 0.3 M CaCl₂ and 1.475 mL of 189 ultrapure water for 2 min at 150 rpm in a water bath at 37 °C. A gastric phase was then carried out by adding 7.5 mL of simulated gastric fluid (SGF), 1.6 mL of porcine pepsin 190 solution (25 000 U mL⁻¹) and 5 µL of 0.3 M CaCl₂. The pH was adjusted to 3.0 with 1 M 191 192 HCl, and the mixture was shaken in darkness for 2 h at 150 rpm in a water bath at 37 °C. To simulate the intestinal step, 11 mL of simulated intestinal fluid (SIF), 2.5 mL of 160 mM bile 193 salts, 5 mL of pancreatin solution (800 U mL⁻¹) and 40 µL of 0.3 M CaCl₂ were added to 194 195 gastric chyme. The pH was adjusted to 7.0 with 1 M NaOH, and the mixture was shaken in darkness for 2 h at 150 rpm in a water bath at 37 °C. The digesta obtained were centrifuged 196 197 for 15 min at $10,000 \times g$. After filtration of the supernatant through a 1.2 µm glass fiber prefilter and a 0.45 µm nylon filter (Phenomenex, Torrance, CA), the filters were rinsed with 198

hexane. The filtrate contained the bioaccessible fraction of PCBs, and the pellet and filterrinse contained the non-bioaccessible fraction of PCBs (Figure 1).

Following Dupont *et al.* (2010), *in vitro* infant digestion was simulated by reducing the pepsin concentration by a factor of 8, the bile salt concentration by a factor of 4 and the pancreatin concentration by a factor of 10 compared with *in vitro* adult digestion.

Following Levi & Lesmes (2014), the digestion of elderly persons (age 75 years) was simulated *in vitro* by reducing the pepsin concentration by a factor of 1.3, the bile salt concentration by a factor of 1.5 and the pancreatin concentration by a factor of 2.2 compared with *in vitro* adult digestion.

208 The meat digestion protocol was validated by measuring the amount of peptides produced by proteolysis during meat digestion. Absorbance of the digesta at 280 nm was 209 measured after the gastric step according to Gatellier et al. (2009) (UVIKON 923, double 210 211 beam UV/VIS spectrophotometer). Before this measurement, trichloroacetic acid (TCA) (50% by volume) was added to the gastric digesta (TCA/digesta, v/v 1:2) to precipitate undigested 212 proteins at 37 °C for 15 min (Gatellier et al., 2009). The mixture was centrifugated at 4000 213 214 rpm at 4 °C for 15 minutes. The absorbance at 280 nm (n = 3) was then measured on the supernatant, and compared with that of the simulated gastric fluid (SGF), the SGF + raw meat 215 mixture and the SGF + pepsin mixture at the same concentrations as those used in the gastric 216 217 step of the digestion protocol. The quantity of proteolytic peptides could not be measured for the intestinal digestion step because of the absorption of the pancreatin and bile salts at 280 218 219 nm.

220

221

2.5. Multianalyte analysis of PCBs

2.5.1. Extraction

222

223

PCB extraction procedure is shown in Figure 1. To extract bioaccessible PCBs, the 224 filtrate obtained after in vitro digestion was first mixed with a trichloroacetic acid (TCA) 225 solution (0.5 g.mL⁻¹) in a volume ratio of 2:1 (filtrate: TCA solution) for protein precipitation. 226 After 30 min at 37 °C, the mixture was centrifuged (15 min, $10,000 \times g$) and a liquid-liquid 227 extraction of the resulting supernatant was performed with 70 mL of solvent and repeated 228 three times. For this step, the PCB recovery rates after an extraction by hexane and DCM 229 were compared, these two solvents having already been used for the extraction of 230 organochlorine contaminants in bioaccessible fractions (Xing et al., 2008; Wang et al., 2011). 231 For that, a mixture of PCBs was spiked at 2.5 µg.mL⁻¹ in 40 mL of ultrapure water (Millipore, 232 Bedford, MA, USA) and a liquid-liquid extraction was performed with 40 mL of hexane 233 (repeated three times) or with 40 mL of DCM (repeated three times) (n = 3 for each solvent). 234 The recovered organic solvent was then evaporated to dryness (Rocket, Genevac Ltd.), and 235 100 µL of hexane was added for PCB analysis by GC×GC-ToF/MS. 236

For the bioaccessible fraction, after the liquid-liquid extraction with the most efficient 237 solvent, the organic phase containing PCBs was evaporated to ~1 mL. A 34 mL stainless-steel 238 extraction cell was then prepared to extract bioaccessible PCBs by ASE. Briefly, 10 g of 239 acidic alumina was placed at the bottom of the cell. Filter papers were placed underneath and 240 on the top of the alumina layer. The cell was then filled with the pellet obtained after TCA 241 protein precipitation and centrifugation, dispersed in 5 g of diatomaceous earth. The 242 evaporated (~1 mL) solvent phase obtained after liquid-liquid extraction was added at the top 243 of the cell. 244

Same ASE extraction was performed to extract non-bioaccessible PCBs contained in the pellet obtained at the end of the *in vitro* digestion. The hexane filter rinse was added at the top of the cell.

248	ASE was then carried out according to Planche et al. (2015) using a Dionex ASE 350
249	extractor (Sunnyvale, CA, USA). Hexane was used as extraction solvent at a temperature of
250	100 °C and pressure of 1500 psi. ASE extraction included heating (5 min), static time (5 min)
251	and purging (90 s) with two extraction cycles per sample. The extract (approximately 40 mL
252	per extraction cell) was evaporated down (Rocket, Genevac Ltd.) using toluene as a keeper to
253	minimise losses during the evaporation step; 4.5 mL of DCM was then added. For cleaning
254	extracts, gel permeation chromatography (GPC) (Gilson, Middleton, WI, USA) was carried
255	out on an S-X3 Bio-Beads column (Bio-Rad, Philadelphia, USA) using DCM as eluting
256	solvent at a flow rate of 5 mL min ⁻¹ . The fraction obtained was evaporated to dryness
257	(Rocket, Genevac Ltd.), then 100 μ L of hexane were added before analysis. All the samples
258	were spiked with internal standards at 100 ng mL ⁻¹ at the different steps.

- 259
- 260

$2.5.2. GC \times GC - ToF/MS$

261

262 Samples were analysed with a time-of-flight mass spectrometer (Pegasus 4D, Leco) 263 coupled to a two-dimensional gas chromatograph (6890, Agilent Technologies) equipped with a dual stage jet cryogenic modulator (licensed from Zoex) according to Planche et al. with 264 slight modifications (Planche et al., 2015). A Rtx-Dioxin2 1D column (60 m × 0.25 mm × 265 0.25 µm) (Restek, Bellefonte, PA, USA) was connected by a deactivated ultimate union 266 (Agilent Technologies, Santa Clara, CA) to a BPX-50 2D column (2 m \times 0.1 mm \times 0.1 μ m) 267 (SGE, Austin, TX, USA). A splitless injection of 1 µl of sample extract was performed 268 through a CTC CombiPal autosampler (CTC Analytics, Zwingen, Switzerland) with a 269 split/splitless inert liner (Restek, Sky® 4.0 mm ID liner) set at 280 °C. Ultra-pure grade 270 helium (purity 99.9995%) was used as carrier gas with a constant flow rate of 1.5 mL min⁻¹. 271 Purge time was set to 60 s with a flow rate of 50 mL min⁻¹. The primary oven temperature 272

was initially set at 90 °C for 1 min, ramped to 200 °C at 20 °C min⁻¹, and then to 300 °C at 2 273 °C min⁻¹ for 10 min. The secondary oven temperature was set at 5 °C higher than the primary 274 oven temperature. The modulator temperature was set at 15 °C higher than the primary oven 275 temperature, and the modulation period was 5 s, with 1.20 s and 1.30 s for the hot and cold 276 pulses, respectively. The transfer line temperature was set at 280 °C. The mass spectrometer 277 was operated with an electron ionisation source (ionisation energy of 70 eV), a detector 278 voltage of 1800 V and a data acquisition rate of 100 spectra s⁻¹. The run time for each sample 279 was 66.5 min. Analytical blank samples of pure solvent were run to check the absence of 280 targeted analytes. GC×GC data were processed using the ChromaTOF software (Leco, 281 version 4.50.8.0). 282

For incurred samples, in which the concentration of contaminants is unknown, PCBs were analyzed by the French National Reference Laboratory (LABERCA, Nantes, France) according to Berge *et al.* (2011), in order not to be restricted by the sensitivity of the GC×GC-ToF/MS method.

- 287
- 288 **2.6. Data processing**

289

Bioaccessibility of PCBs was calculated according to Kang *et al.* (2013):

291 Bioaccessibility (%) = $\frac{\text{Bioaccessible PCBs}}{\text{Bioaccessible PCBs} + \text{Non-bioaccessible PCBs}} \times 100$, where

bioaccessible PCBs and non-bioaccessible PCBs are the amounts of PCBs found after *in vitro* digestion in the bioaccessible and non-bioaccessible fractions, respectively. Data were processed using the Statistica Software version 12 (Dell Software, Paris, France). Principal component analyses (PCA) were performed on the bioaccessibility data of PCBs in meat after *in vitro* digestion to visualise the structure of the data. To determine whether the cooking process, the meat fat content or the physiological differences due to the age of consumers had

any impact on the bioaccessibility of PCBs in meat, a one-way analysis of variance (ANOVA; 298 p < 0.05) was performed on bioaccessibility data from GC×GC-ToF/MS. A Newman-Keuls 299 mean comparison test was then performed on the resulting dataset. 300

- 301
- 302

2.7. Determination of the mean uptake distribution according to fat level, cooking mode and consumer age 303

304

Using the modular Bayesian approach presented in Tressou et al. (2017) and integrating 305 initial contamination data obtained after meat production for both organic and conventional 306 307 production, data reflecting the effect of cooking, data on the levels and frequency of consumption of conventional and organic meat, and data on the effect of digestion presented 308 in the current paper, we propose to compare the exposure after digestion (uptake) according to 309 310 the fat level, the cooking mode, and the consumer age. The bioaccessibility step is the last one of our modular approach and the angle used in Tressou et al. (2017) was to look at the effect 311 312 of the cooking mode for adults eating 15% fat beef meat. Due to a lack of information on the 313 percentage of fat of the meat consumed and the inability of the model to deal with consumers eating both 5% and 15% fat beef meat, all meat was considered as 15% fat. In section 3.2, we 314 look at the uptake considering all the beef meat is 5% fat meat, without changing any other 315 316 assumptions (the study population is adults, the cooking mode is medium as recommended by WHO). 317

In section 3.3, we show the effect of cooking as in Tressou et al. (2017), but using the 318 available data on individuals' body weight. In section 3.4, we consider the effect of age on 3 319 populations: children less than 6 years old, adults (18-64 years old), and seniors (65 years old 320 and older). For each population, consumption data was extracted from the different databases 321 322 (SoMeat data, INCA2 data, KANTAR data) described in Tressou et al. (2017), to differentiate as much as possible their behaviors. Based on the different data sets, we extract the percentage of meat consumer, the mean quantity consumed and individual body weight, and the proportion of organic meat consumption.

326 For children less than 6 years old, the detailed individual meat consumptions and body weights of 163 children aged 3 to 5 were obtained from the INCA2 data, information about 327 organic meat consumption of 780 households with at least a child less than 6 years old from 328 KANTAR data, and information from 94 respondents belonging to households with children 329 aged 0 to 9 years old from the SOMEAT project survey (http://www.so-meat.fr/). For adults, 330 the detailed individual meat consumptions and body weights of 2276 adults aged 18 to 64 331 332 were obtained from the INCA2 data, information about organic meat consumption of 3381 households from KANTAR data with one the reference adults aged 18 to 64, and information 333 from 360 respondents aged 18 to 64 from the SOMEAT project survey. For seniors, the 334 335 detailed individual meat consumptions and body weights of 348 adults aged 65 to 79 were obtained from the INCA2 data, information about organic meat consumption of 1474 336 337 households from KANTAR data with one the reference adults aged 65 or more, and 338 information from 26 respondents aged 65 and more from the SOMEAT project survey.

339

- **340 3. Results and Discussion**
- 341

3.1. Analysis of PCBs in meat digesta

342

To assess the bioaccessibility of the 18 most relevant PCBs in meat (Sirot *et al.*, 2012), the standard *in vitro* static digestion protocol described by Minekus *et al.* (2014) was implemented starting with 5 g of meat spiked with PCBs (Figure 1). The protocol was validated by verifying the increase in the amount of peptides produced by proteolysis after the gastric step of digestion according to Gatellier *et al.* (2009). The mean absorbance of the 348 gastric digesta was 1.56 (OD units) against nil for the simulated gastric fluid (SGF) solution, 349 0.49 for the SGF + meat mixture and 0.48 for the SGF + pepsin mixture. The absorbance of 350 the gastric digesta was thus significantly greater than the other absorbances measured 351 (p < 0.05), confirming the proteolysis in the course of the gastric digestion of the meat. After 352 considering dilution steps, the mean absorbance (1.18 OD units) was in the same order of 353 magnitude as that found by Gatellier *et al.* (2009) in their gastric raw meat digesta (0.98 OD 354 units), thus validating our *in vitro* digestion set-up.

A protocol for the extraction and analysis of PCBs in the bioaccessible fraction 355 (filtrate) and non-bioaccessible fraction (pellet and filter rinse) was then developed (Figure 1). 356 357 For the liquid/liquid extraction, a comparison was made between hexane (Xing et al., 2008) and dichloromethane (DCM) (Wang et al., 2011) as the extraction solvent. Results showed a 358 mean recovery rate for PCBs in the bioaccessible fraction of 98% using hexane, with values 359 360 ranging from 56 \pm 0.7% for PCB 28 to 112 \pm 19% for PCB 156 (Table S1. Supplementary material). With DCM, the mean recovery rate was 90%, with values ranging from $71 \pm 4.9\%$ 361 362 for PCB 101 to $118 \pm 6.5\%$ for PCB 180 (Table S1. Supplementary material). Except for PCB 363 28 with hexane as the extraction solvent, all the recovery rates lay in the classically accepted range of 70–130% according to the EPA Method 8000C (2003), with RSD \leq 10% for most of 364 the compounds. Of the 18 PCBs targeted in this study, 14 showed higher recovery rates, and 365 366 closer to 100%, with hexane than with DCM. Hexane was therefore subsequently used in this study for the liquid/liquid extraction of PCBs in the bioaccessible fraction. 367

After ASE extraction of PCBs and GPC defatting of the extracts obtained respectively for bioaccessible and non bioaccessible fractions (Figure 1), GC×GC-ToF/MS analysis of PCBs was carried out according to Planche *et al.* (2015). The visual separation of the 18 PCBs (Figure 2) was validated based on resolution factors (R_s) with $R_s = \Delta r_t/w_b$, where r_t is the retention time and w_b the average peak width at the base (Table S2. Supplementary material). For the 18 PCBs, as $R_{s,1D} \ge 0.6$ in the first separation dimension or $R_{s,2D} \ge 0.4$ in the second separation dimension, all peaks were considered resolved.

Lastly, to verify the accuracy of our protocol, the recovery rates for PCBs contained in the raw meat were determined after completion of the procedure (*in vitro* digestion, extraction and analysis by GC×GC-ToF/MS). The PCBs contained in the meat before digestion were thus quantified (n = 3) and compared with the sum of the PCBs found in the bioaccessible and non-bioaccessible fractions (n = 3) after digestion. The mean of the resulting recovery rates was 95 ± 16%, confirming the efficiency and robustness of the protocol.

381

382

3.2. Impact of meat fat content on bioaccessibility of PCBs

383

To determine the impact of meat fat content on the bioaccessibility of PCBs, two 384 different types of meat were purchased: 15% fat ground beef which is the type of beef meat 385 386 most consumed by the French population (51 g per week) (Mota et al., 2021) and 5% fat ground beef as an example of low fat meat (Cardona et al., 2020). After verification, the exact 387 fat level in the first type of meat was 11% which corresponds to a model of high fat meat 388 (Cardona et al., 2020). It is important to note that no significant difference was observed 389 between 11% and 5% fat meat in terms of lipid relative composition (Table S3. 390 391 Supplementary material). Bioaccessibility results of the 18 most relevant PCBs in meat are 392 reported in Table 1. For meat with 11% fat content, PCB bioaccessibility was between 20.6% and $30.5\% \pm 1.4-6.7\%$ against $44.2-50.1\% \pm 3.0-12.5\%$ for meat with 5% fat content. As 393 shown in Figure S1. (Supplementary material), fat content is thus a key determinant in the 394 bioaccessibility of PCBs. Compared to the results previously obtained with meat (Planche et 395 al., 2015), the low relative standard deviations (3.0%-12.5% for 5% fat meat; 1.4%-6.7% for 396 11% fat meat) point out the poor bioacessibility variability between PCB congeners. Note that 397 for PCB 126, which displays the highest toxic equivalent factor (WHO-TEF = 0.1), the 398

bioaccessibility was $24.0 \pm 0.6\%$ for meat with 11% fat content and $44.7 \pm 3.4\%$ for meat with 5% fat content, confirming that bioaccessibility is a key factor to be considered in risk analysis.

402 These values are of the same order of magnitude as those obtained by Shen et al. (2016), who reported a mean bioaccessibility of PCBs in beef samples (4.8% fat) of 49.0 \pm 403 3.3% after boiling (5 min at 100 °C). By contrast, they differ with those of Xing et al. (2008), 404 who found a mean bioaccessibility of PCBs of 3% in raw fish samples (3-15% fat). Besides 405 the fact that Xing et al. worked on raw food matrices whereas we used cooked ones, there 406 were important disparities between the in vitro digestion protocols used, in particular 407 408 regarding the sample incubation times with the digestive enzymes used in the different steps of the protocol (1 h in Xing et al. against 2 h in our study for the gastric digestion step, and 6 409 h against 2 h in the intestinal digestion step). These differences in protocols thus rule out any 410 411 objective comparison of the data obtained, this is the reason why the standardised international protocol setup by Minekus et al. (2014) was implemented in the present study. 412 413 Xing et al. (2008) and Shen et al. (2016), like us, found a significantly greater PCB 414 bioaccessibility (p < 0.05) when the food fat content is lower. Xing *et al.* (2008) measured a mean PCB bioaccessibility of 3% in fish samples (3-15% fat), whereas bioaccessibility was 415 25% in spinach and salad samples (0.1% fat on average). In our study, we found the relation 416 417 between the lipid levels of cooked meat and bioaccessibility of PCBs to be inversely proportional: when the percentage of fat was divided by 1.9, the bioaccessibility was 418 multiplied by 1.8. As PCBs are lipophilic compounds (Log Kow = 4.09-8.18 according to 419 420 Hawker & Connell (1988)), their digestion starts with their release from the food bolus and their dissolution in the fat phase of the meal, followed by emulsification into lipid droplets of 421 gastric emulsion (Xavier & Mercadante, 2019). Then, they are transferred to mixed micelles, 422 423 composed by products of lipid hydrolysis, phospho-lipids, cholesterol, and bile salts and only

the fraction of PCBs incorporated into mixed micelles is bioaccessible (Xavier & Mercadante, 424 2019). For high fat meat, the solubilization capacity of the mixed micelles may be limited by 425 the concentration of available digestive enzymes and emulsifiers (e.g. pancreatin and bile 426 salts) (Tan et al., 2020). Thus, we can suggest that there is a fat level threshold above which 427 lipophilic compounds such as PCBs are not fully released from the lipid phase due to the fact 428 that there were proportionally less mixed micelles available to solubilize it (Tan et al., 2020). 429 Considering that there is no significant difference between 11% and 5% fat meat in terms of 430 lipid relative composition (Table S3. Supplementary material), this may explain why an 431 inversely proportional relation is observed between lipid levels and PCB bioaccessibility. This 432 433 trend was also observed by Xia et al. (2017) for lipophilic micronutrients like the hydrophobic β-carotene, with significantly lower bioaccessibility in a simulated high-fat diet than in a 434 simulated low-fat diet. A significant fraction of the lipid phase was not fully digested and, 435 436 therefore, a fraction of the β -carotene was not released into the intestinal fluids, remaining in the nondigested lipid phase (Xia et al., 2017). Considering that PCB concentrations are 437 438 positively correlated to the fat content of food (Carlson et al., 2005), our results suggest that 439 PCBs in fat-rich food are less bioaccessible, which may therefore mitigate the hazard related to their presence in such food. 440

Figure 3a shows the intake distribution of exposure obtained according to the fat content after digestion. The reduced bioaccessibility obtained with 11% fat meat leads to reduced uptakes compared to 5% fat meat.

- 444
- 445

3.3. Impact of cooking intensity on bioaccessibility of PCBs

446

447 To determine the impact of cooking intensity on PCB bioaccessibility in meat, *in vitro* 448 digestions were carried out on ground beef containing 11% fat, raw and cooked at three

different intensities: rare (50 °C at the core), medium, (70 °C at the core according to WHO 449 recommendation), and well-done (85 °C at the core) (n = 3 for each condition). Table S4. 450 (Supplementary material) shows that PCB bioaccessibility was $23.3-35.1\% \pm 6.0-13.9\%$ for 451 452 raw meat, $20.5-28.2\% \pm 1.0-8.9\%$ for rare-cooked meat, $20.6-30.5\% \pm 1.4-6.7\%$ for mediumcooked meat (WHO recommendations), and $18.0-26.7\% \pm 2.3-11.6\%$ for well-done meat. As 453 mentioned above, the low coefficients of variation point out the poor bioaccessibility 454 variability between PCB congeners. No significant difference was observed between raw, 455 rare- and WHO-cooked meat, whereas the bioaccessibility measured on well-done meat was 456 significantly lower (p < 0.05). For PCB 126, the bioaccessibility decreased from $25.2 \pm 10.0\%$ 457 458 for raw meat to $18.0 \pm 8.1\%$ after intense cooking.

Although they did not assess the impact of cooking intensity on PCB bioaccessibility, 459 Shen et al. (2016) showed that cooking modes need to be considered in bioaccessibility 460 461 studies: for example, the PCB bioaccessibility in meat was significantly lower (p < 0.05) after boiling (5 min at 100 °C) than after frying in cooking oil (5 min at 200-300 °C). Various 462 463 studies conducted on trace elements (Cu, Cd, Zn, etc.) have also shown that the cooking of 464 fish and other seafood generally lowers bioaccessibility, irrespective of the cooking method used: frying, grilling, boiling or steaming (He et al., 2010). The lower bioaccessibility 465 observed by these authors may be explained by the change in meat fat content induced by 466 467 cooking (Xing et al., 2008). In the present study, whereas raw meat contained 11% fat, cooked meat contained respectively 13%, 15% and 16% fat after rare, WHO and intense 468 cooking, respectively. As discussed above, the moderate decrease in bioaccessibility could 469 470 thus be at least partly linked to the increase in meat fat content. Another possible explanation for the reduced bioaccessibility after intense cooking is the protein denaturing phenomenon 471 that occurs during cooking. This process leads to tissue shrinking, making them harder and 472 more compact, and therefore hindering digestive enzymes activity (Kulp et al., 2003; He et 473

al., 2010). Indeed, pepsin activity on myofibrillary proteins has been shown to be reduced by
58% in cooked meat (45 min cooking at 100 °C) compared with raw meat (Santé-Lhoutellier *et al.*, 2008). Cooking may also cause the formation of disulphide-bonded proteins, making
the proteins less digestible (He *et al.*, 2010). All these processes are especially significant
under extreme cooking conditions, and may explain the decrease in bioaccessibility only in
the case of intense cooking.

Figure 3b presents the intake distribution of exposure obtained according to the cooking intensity after digestion. The figure shows that the reduced bioaccessibility obtained with well-done meat compared to rare or WHO-cooked meat leads to reduced uptakes.

- 483
- 484

3.4. Impact of consumer age on bioaccessibility of PCBs

485

486 To determine the impact of physiological conditions related to the age of the consumer on PCB bioaccessibility in meat (11% fat, WHO cooking), in vitro digestions were carried 487 488 out, simulating the digestion of an infant (n = 3) according to Dupont *et al.* (2010), and that of 489 an elderly person (n = 3) following Levi & Lesmes (2014). These data were then compared with those previously obtained for adult digestion simulation (Minekus et al., 2014). Table 2 490 shows that the mean PCB bioaccessibility was $7.7 \pm 0.8\%$ for the infant model and $17.1 \pm$ 491 492 1.5% for the elderly model, against $25.9 \pm 2.3\%$ for the adult model, with again, a poor variability between PCB congeners. As shown in Figure S1. (Supplementary material), 493 bioaccessibility thus varies significantly (p < 0.05) according to the consumer age. 494

These variations are consistent with concentrations of pepsin, bile salts and pancreatin, which were 8, 4 and 10 times lower in the infant model, and 1.3, 1.5 and 2.2 times lower in the elderly model compared to the adult model (Dupont *et al.*, 2010; Levi & Lesmes, 2014) confirming that the lower the concentrations of these enzymes, the lower the PCB

bioaccessibility values. Accordingly, Jadán-Piedra et al. (2016) observed an increase of Hg 499 bioaccessibility in fish with the concentration of pepsin, whose activity facilitated the 500 cleavage of protein-contaminant bonds. Likewise, the bioaccessibility of heterocyclic 501 502 aromatic amines was found to increase with the concentration of lipase containing pancreatin (Kulp et al., 2003). Furthermore, the present study also shows that the concentration of bile 503 salts is a key determinant of bioaccessibility: the concentration of bile salts being 4 times 504 lower in the infant model than in the adult model, PCB bioaccessibility in meat was divided 505 by 3.4. Similarly, in the elderly model, both the concentration of bile salts and the measured 506 bioaccessibility were 1.5 times lower than in the adult model. These results are consistent 507 508 with those reported by Yu et al. (2011), who studied the bioaccessibility of PBDEs in dust. These authors observed that when the concentration of bile salts was multiplied by 2.2, the 509 bioaccessibility of PBDEs rose 2.8-fold. During digestion, bile salts may facilitate the 510 emulsification of fats containing lipophilic substances such as PBDEs or PCBs, thereby 511 inducing higher bioaccessibility values (Yu et al., 2011). 512

513 Figure 3c shows the intake distribution of exposure obtained according to the 514 consumer age after digestion. Comparing the uptakes of the three subpopulations, Figure 3c seems to indicate that the reduced bioaccessibility of children and elderly leads to reduced 515 uptakes for children and elderly. Children uptakes remain more variable because of large 516 517 differences that exist between children in terms of meat consumption. Moreover, children uptakes can exceed those of adults as their exposure before digestion is higher than adults and 518 senior exposures (Figure S2. Supplementary material). This is mainly due to the fact that the 519 520 quantity of meat consumed by children (mean 109 g/week) (Figure S3. Supplementary material) is relatively high compared to their body weight (mean 19 kg) while the quantities 521 of meat consumed by adults and elderly (means 191 g/week and 145 g/week, respectively) 522

523 (Figure S3. Supplementary material) are smaller compared to their body weights (mean 70524 kg).

To conclude, the highest PCB bioaccessibility values may be observed in adults, after 525 a consumption of raw, rare or WHO-cooked meat or with 5% fat meat. In contrast, the lowest 526 PCB bioaccessibility values may be obtained for the elderly, after a consumption of well-done 527 meat or with 11% fat. For children, the exposure after digestion varies widely because, as 528 discussed above, there are large differences between individuals in terms of meat 529 consumption. However, we can note that for children, undercooked ground beef is considered 530 a high-risk product due to the haemolytic uremic syndrome (HUS) associated with Shiga 531 532 toxin-producing Escherichia coli (STEC)-contaminated beef (Brusa et al., 2020). Following the recommendations issued to limit this risk, bioaccessibility values obtained after medium 533 or intense cooking should therefore cover the most common situations observed for children. 534 535 In contrast, beef is most frequently eaten rare or medium cooked by adults (Loukiadis et al., 2017), which corresponds to the highest bioaccessibility values obtained in this study even if 536 537 the impact of cooking on PCB bioaccessibility is moderate. These cooking modes are also the 538 most commonly observed with the elderly because individuals with mastication problems may avoid well-done meat (Gil-Montoya et al., 2015). Considering the meat fat level, the results 539 obtained with high fat ground beef (which corresponds to the lowest PCB bioaccessibility 540 541 values) correspond to the most frequent situation observed for adults (Cardona et al., 2020). For the elderly, it should be recommended to consume low fat meat in order to limit the risk 542 of cardiovascular diseases (Bronzato & Durante, 2017) but it is important to note that this 543 544 recommendation leads to higher PCB bioaccessibility values.

545

546 **3.5.** Comparison of results with naturally contaminated samples

547

All the results presented above were achieved with intentionally contaminated (spiked) 548 samples in order to control the matrix composition and its PCB load (181.8 ng/g lipid weight 549 for 11% fat meat and 400 ng/g lipid weight for 5% fat meat). However, PCB bioaccessibility 550 551 determinations were also carried out on naturally contaminated (incurred) meat samples containing PCBs bioaccumulated at ultra-trace concentration during animal breeding in order 552 to assess the scope of the conclusions drawn from spiked model samples. For this purpose, in 553 *vitro* digestions were carried out, after medium cooking of beef meat samples (n = 4) which 554 concentrations in PCBs have been determined applying sensitive and selective GC-HRMS 555 (BE) ISO17025 accredited method. After mincing and blending, the four raw samples were 556 557 analysed, showing concentrations ranging from 3.4 to 44.6 ng/g lipid weight for non-dioxinlike PCBs and from 1575 to 4032 pg/g lipid weight for dioxin-like PCBs. As reported in 558 Table 3, mean PCB bioaccessibility was $45.0 \pm 8.2\%$ for sample 1 (2.9% fat), $44.0 \pm 9.0\%$ for 559 560 sample 2 (5.0% fat), $49.4 \pm 10.9\%$ for sample 3 (5.3% fat) and $31.9 \pm 13.6\%$ for sample 4 (9.5% fat) after a medium cooking. The mean values of bioaccessibility obtained for samples 561 562 2 and 3, both presenting fat contents close to 5%, were of the same order of magnitude as 563 those obtained, after medium cooking, on spiked 5% fat-meat used in our study (47.5 \pm 1.7%). In addition, no significant difference was observed between the bioaccessibility values 564 measured from meat containing 2.9% fat and 5-5.3% fat. This suggests the occurrence of a fat 565 566 content threshold below which the PCB bioaccessibility remains stable. This trend was already described for polybrominated diphenyl ethers (PBDEs) by Yu et al. (2010), who 567 observed that PBDE bioaccessibility was significantly correlated with the fat content of food 568 569 only above 1.8% of their fresh weight. To confirm this trend and validate this hypothesis for PCB bioaccessibility, measures would have to be carried out on a greater number of low-fat 570 samples. 571

Data presented in Table 3 also indicate that for a higher fat content (e.g. sample 4, 572 9.5% fat), the values of bioaccessibility were significantly lower (p < 0.05) than for the other 573 samples, in line with our results with spiked meat, thus confirming the inversely proportional 574 575 relation between PCB bioaccessibility and meat fat content. Interestingly, unlike the spiked samples, the incurred samples were found to present bioaccessibility values that varied 576 according to the PCB congeners: bioaccessibility was higher for the least chlorinated 577 congeners and lower for the most chlorinated, with bioaccessibility values for example 578 significantly higher (p < 0.05) for PCB 28 (trichlorobiphenyl) than for PCBs 180 and 189 579 (heptachlorobiphenyls). This suggests that the relationship that exists in incurred samples 580 581 between PCBs and matrix components cannot be identically reproduced in spiked samples. Xing et al. (2008), Kang et al. (2013) and Wang et al. (2013) also observed that the 582 bioaccessibility of PCBs in incurred samples was generally higher for the least chlorinated 583 584 congeners. The most chlorinated PCBs display the highest Log Kow values (Wang et al., 2013), so that these congeners may be more readily retained in fat, and as a consequence 585 586 harder to solubilise in the bioaccessible fraction (Kang et al., 2013).

587

588 4. Conclusion

589

590 Coupled with GC×GC-ToF/MS analysis, the standardised *in vitro* static digestion protocol 591 implemented in this study enabled to assess the bioaccessibility of PCBs in meat. For ground 592 beef meat with 11% fat, only a quarter of PCBs contained in this matrix is liable to cross the 593 intestinal barrier and induce toxic effects. These data confirm that bioaccessibility is a key 594 factor to be considered in risk analysis. The fat content of food, the consumer age and to a 595 lesser extent, the cooking intensity are factors that cause bioaccessibility to vary. Indeed, 596 PCBs are less bioaccessible when the fat content increases, after an intense cooking or even in

the infant and the elderly, thus reducing the exposure to PCBs due to meat consumption in 597 these situations. More extensive investigations will be necessary to elucidate in details the 598 mechanisms responsible for these results. Moreover, in order to study more precisely the main 599 600 factors influencing contaminant bioaccessibility, in vitro dynamic digestion models could be used to simulate digestion as realistically as possible. It would also be interesting to determine 601 how PCBs interact with the gut microbiota and whether these interactions are relevant for 602 human health. Work is underway to address this challenging issue (Defois et al., 2018). 603 Finally, our work highlights the great relevance of using a consensus digestion protocol for 604 comparisons between different studies. The wide use of such protocols would allow the 605 606 creation of a consistent database with the ultimate goal of improving chemical risk assessment procedures considering accurate bioaccessibility data. In this perspective, simplified high 607 throughput digestion protocols were developed, based on the protocol setup in the present 608 609 study, in order to obtain data for epidemiological studies on the health impact of food (Wedekind et al., 2020). 610

611

612 **Conflict of interest**

613 The authors declare that there is no conflict of interest.

614

615 Acknowledgements

This study was supported by the French National Research Agency, project SOMEAT, Contract No. ANR-12-ALID-0004. Safety of Organic Meat. Available at http://www.someat.fr. We thank Didier Dupont (STLO-INRA), Olivia Ménard (STLO-INRA) and Claire Dufour (SQPOV-INRA) for their advice regarding *in vitro* digestion protocols, and JeanPaul Rigaudiere and Frédéric Capel (UNH-INRA) for the lipid composition determination ofmeat samples.

622

623 **References**

Berge, P., Ratel, J., Fournier, A., Jondreville, C., Feidt, C., Roudaut, B., Le Bizec, B., &
Engel, E. (2011). Use of volatile compound metabolic signatures in poultry liver to backtrace dietary exposure to rapidly metabolized xenobiotics. *Environmental Science* & *Technology*, 45(15), 6584-6591.

628

Blanchet-Letrouvé, I., Zalouk-Vergnoux, A., Vénisseau, A., Couderc, M., Le Bizec, B., Elie,
P., Herrenknecht, C., Mouneyrac, C., & Poirier, L. (2014). Dioxin-like, non-dioxin like
PCB and PCDD/F contamination in European eel (Anguilla anguilla) from the Loire
estuarine continuum: spatial and biological variabilities. *Science of the Total Environment,*472, 562-571.

634

Bronzato, S., & Durante, A. (2017). A contemporary review of the relationship between red
meat consumption and cardiovascular risk. *International journal of preventive medicine*,
8:40.

638

Cardona, M., Gorriz, A., Barat, J. M., & Fernández-Segovia, I. (2020). Perception of fat and
other quality parameters in minced and burger meat from Spanish consumer studies. *Meat science, 166*, 108138.

642

643	Carlson, D. L., & Hites, R. A. (2005). Polychlorinated biphenyls in salmon and salmon feed:
644	global differences and bioaccumulation. Environmental Science & Technology, 39(19),
645	7389-7395.
646	
647	Cruz, R., Mendes, E., Maulvault, A. L., Marques, A., Casal, S., & Cunha, S. C. (2020).
648	Bioaccessibility of polybrominated diphenyl ethers and their methoxylated metabolites in
649	cooked seafood after using a multi-compartment in vitro digestion model. Chemosphere,
650	252, 126462.
651	
652	Brusa, V., Costa, M., Padola, N. L., Etcheverría, A., Sampedro, F., Fernandez, P. S., Leotta,
653	G. A., & Signorini, M. L. (2020). Quantitative risk assessment of haemolytic uremic
654	syndrome associated with beef consumption in Argentina. PLoS One, 15(11), e0242317.
655	
656	Defois, C., Ratel, J., Garrait, G., Denis, S., Le Goff, O., Talvas, J., Mosoni, P., Engel, E., &
657	Peyret, P. (2018). Food chemicals disrupt human gut microbiota activity and impact
658	intestinal homeostasis as revealed by in vitro systems. Scientific reports, $\delta(1)$, 1-12.
659	
660	Dervilly-Pinel, G., Guérin, T., Minvielle, B., Travel, A., Normand, J., Bourin, M., Royer, E.,
661	Dubreil, E., Mompelat, S., Hommet, F., Nicolas, M., Hort, V., Inthavong, C., Saint-Hilaire,
662	M., Chafey, C., Parinet, J., Cariou, R., Marchand, P., Le Bizec, B., Verdon, E., & Engel, E.
663	
	(2017). Micropollutants and chemical residues in organic and conventional meat. Food
664	(2017). Micropollutants and chemical residues in organic and conventional meat. <i>Food chemistry</i> , 232, 218-228.
664 665	(2017). Micropollutants and chemical residues in organic and conventional meat. <i>Food chemistry</i> , 232, 218-228.
664 665 666	 (2017). Micropollutants and chemical residues in organic and conventional meat. <i>Food chemistry</i>, 232, 218-228. Dupont, D., Mandalari, G., Molle, D., Jardin, J., Léonil, J., Faulks, R. M., Wickham, M. S. J.,
664 665 666 667	 (2017). Micropollutants and chemical residues in organic and conventional meat. <i>Food chemistry, 232</i>, 218-228. Dupont, D., Mandalari, G., Molle, D., Jardin, J., Léonil, J., Faulks, R. M., Wickham, M. S. J., Mills, E. N. C., & Mackie, A. R. (2010). Comparative resistance of food proteins to adult

668	and infant in vitro digestion models. Molecular Nutrition & Food Research, 54(6), 767-
669	780.
670	
671	Engel, E., Ratel, J., Bouhlel, J., Planche, C., & Meurillon, M. (2015). Novel approaches to
672	improving the chemical safety of the meat chain towards toxicants. Meat Science, 109, 75-
673	85.
674	
675	Gatellier, P., & Santé-Lhoutellier, V. (2009). Digestion study of proteins from cooked meat
676	using an enzymatic microreactor. Meat Science, 81(2), 405-409.
677	
678	Gil-Montoya, J. A., de Mello, A. L. F., Barrios, R., Gonzalez-Moles, M. A., & Bravo, M.
679	(2015). Oral health in the elderly patient and its impact on general well-being: a
680	nonsystematic review. Clinical interventions in aging, 10, 461.
681	
682	Hawker, D. W., & Connell, D. W. (1988). Octanol-water partition coefficients of
683	polychlorinated biphenyl congeners. Environmental Science & Technology, 22(4), 382-
684	387.
685	
686	He, M., Ke, C. H., & Wang, W. X. (2010). Effects of cooking and subcellular distribution on
687	the bioaccessibility of trace elements in two marine fish species. Journal of Agricultural
688	and Food Chemistry, 58(6), 3517-3523.
689	
690	He, R. W., Li, Y. Z., Xiang, P., Li, C., Cui, X. Y., & Ma, L. Q. (2018). Impact of particle size
691	on distribution and human exposure of flame retardants in indoor dust. Environmental
692	research, 162, 166-172.

69	3
----	---

- Jadán-Piedra, C., Clemente, M. J., Devesa, V., & Vélez, D. (2016). Influence of Physiological
 Gastrointestinal Parameters on the Bioaccessibility of Mercury and Selenium from
 Swordfish. *Journal of Agricultural and Food Chemistry*, 64(3), 690-698.
- 697
- Kang, Y., Yin, Y., Man, Y., Li, L., Zhang, Q., Zeng, L., & Wong, M. H. (2013).
 Bioaccessibility of polychlorinated biphenyls in workplace dust and its implication for risk
 assessment. *Chemosphere*, *93(6)*, 924-930.
- 701
- Kulp, K. S., Fortson, S. L., Knize, M. G., & Felton, J. S. (2003). An in vitro model system to
 predict the bioaccessibility of heterocyclic amines from a cooked meat matrix. *Food and Chemical Toxicology*, *41(12)*, 1701-1710.
- 705
- Levi, C. S., & Lesmes, U. (2014). Bi-compartmental elderly or adult dynamic digestion
 models applied to interrogate protein digestibility. *Food & Function*, *5(10)*, 2402-2409.
- 708
- Loukiadis, E., Bièche-Terrier, C., Malayrat, C., Ferré, F., Cartier, P., & Augustin, J. C.
 (2017). Distribution of Escherichia coli O157: H7 in ground beef: Assessing the clustering
 intensity for an industrial-scale grinder and a low and localized initial contamination. *International journal of food microbiology*, 250, 75-81.
- 713
- Milinčić, D. D., Vojinović, U. D., Kostić, A. Ž., Pešić, M. B., Trifunović, B. D. Š., Brkić, D.
 V., Stević M. Z., Kojić M. O., & Stanisavljević, N. S. (2020). In vitro assessment of
 pesticide residues bioaccessibility in conventionally grown blueberries as affected by
 complex food matrix. *Chemosphere, 252*, 126568.
- 718

719	Minekus, M., Alminger, M., Alvito, P., Ballance, S., Bohn, T., Bourlieu, C., Carrière, F.,
720	Boutrou, R., Corredig, M., Dupont, D., Dufour, C., Egger, L., Golding, M., Karakaya, S.,
721	Kirkhus, B., Le Feunteun, S., Lesmes, U., Macierzanka, A., Mackie, A., Marze, S.,
722	McClements, D. J., Ménard, O., Recio, I., Santos, C. N., Singh, R. P., Vegarud, G. E.,
723	Wickham, M. S. J., Weitschies, W., & Brodkorb, A. (2014). A standardised static in vitro
724	digestion method suitable for food-an international consensus. Food & Function, 5(6),
725	1113-1124.

726

Mota, J. D. O., Guillou, S., Pierre, F., & Membré, J. M. (2021). Public health risk-benefit
assessment of red meat in France: Current consumption and alternative scenarios. *Food and Chemical Toxicology*, *149*, 111994.

730

Oomen, A. G., Rompelberg, C. J. M., Bruil, M. A., Dobbe, C. J. G., Pereboom, D. P. K. H., &
Sips, A. J. A. M. (2003). Development of an *in vitro* digestion model for estimating the
bioaccessibility of soil contaminants. *Archives of environmental contamination and toxicology*, 44(3), 281-287.

735

Planche, C., Ratel, J., Mercier, F., Blinet, P., Debrauwer, L., & Engel, E. (2015). Assessment
of comprehensive two-dimensional gas chromatography-time-of-flight mass spectrometry
based methods for investigating 206 dioxin-like micropollutants in animal-derived food
matrices. *Journal of Chromatography A*, 1392, 74-81.

740

Planche, C., Ratel, J., Blinet, P., Mercier, F., Angénieux, M., Chafey, C., Zinck, J., Marchond,
N., Chevolleau, S., Marchand, P., Dervilly-Pinel, G., Guérin, T., Debrauwer, L., & Engel,
E. (2017). Effects of pan cooking on micropollutants in meat. *Food Chemistry*, 232, 395404.

74

746	Santé-Lhoutellier, V., Astruc, T., Marinova, P., Greve, E., & Gatellier, P. (2008). Effect of
747	meat cooking on physicochemical state and in vitro digestibility of myofibrillar
748	proteins. Journal of Agricultural and Food Chemistry, 56(4), 1488-1494.
749	
750	Shen, H., Starr, J., Han, J., Zhang, L., Lu, D., Guan, R., Xu, X., Wang, X., Li, J., LI, W.,
751	Zhang, Y., & Wu, Y. (2016). The bioaccessibility of polychlorinated biphenyls (PCBs) and
752	polychlorinated dibenzo-p-dioxins/furans (PCDD/Fs) in cooked plant and animal origin
753	foods. Environment International, 94, 33-42.
754	
755	Sirot, V., Tard, A., Venisseau, A., Brosseaud, A., Marchand, P., Le Bizec, B., & Leblanc, J.
756	C. (2012). Dietary exposure to polychlorinated dibenzo-p-dioxins, polychlorinated
757	dibenzofurans and polychlorinated biphenyls of the French population: results of the
758	second French Total Diet Study. Chemosphere, 88(4), 492-500.
759	
760	Tan, Y., Zhang, Z., Zhou, H., Xiao, H., & McClements, D. J. (2020). Factors impacting lipid
761	digestion and β -carotene bioaccessibility assessed by standardized gastrointestinal model
762	(INFOGEST): oil droplet concentration. Food & Function, 11(8), 7126-7137.
763	
764	Tressou, J., Ben Abdallah, N., Planche, C., Dervilly-Pinel, G., Sans, P., Engel, E., & Albert, I.
765	(2017). Exposure assessment for dioxin-like PCBs intake from organic and conventional
766	meat integrating cooking and digestion effects. Food and Chemical Toxicology, 110, 251-
767	261.
768	

769	Wang, H. S., Zhao, Y. G., Man, Y. B., Wong, C. K., & Wong, M. H. (2011). Oral
770	bioaccessibility and human risk assessment of organochlorine pesticides (OCPs) via fish
771	consumption, using an in vitro gastrointestinal model. Food Chemistry, 127(4), 1673-1679.
772	
773	Wang, W., Huang, M. J., Zheng, J. S., Cheung, K. C., & Wong, M. H. (2013). Exposure
774	assessment and distribution of polychlorinated biphenyls (PCBs) contained in indoor and
775	outdoor dusts and the impacts of particle size and bioaccessibility. Science of the Total
776	Environment, 463, 1201-1209.
777	
778	Wedekind, R., Keski-Rahkonen, P., Robinot, N., Mercier, F., Engel, E., Huybrechts, I., &
779	Scalbert, A. (2020). Metabolic Signatures of 10 Processed and Non-processed Meat
780	Products after In Vitro Digestion. Metabolites, 10(7), 272.
781	
782	Xavier, A. A. O., & Mercadante, A. Z. (2019). The bioaccessibility of carotenoids impacts the
783	design of functional foods. Current opinion in food science, 26, 1-8.
784	
785	Xia, Z., McClements, D. J., & Xiao, H. (2017). Influence of Lipid Content in a Corn Oil
786	Preparation on the Bioaccessibility of β -Carotene: A Comparison of Low-Fat and High-Fat
787	Samples. Journal of Food Science, 82(2), 373-379.
788	
789	Xing, G. H., Yang, Y., Chan, J. K. Y., Tao, S., & Wong, M. H. (2008). Bioaccessibility of
790	polychlorinated biphenyls in different foods using an in vitro digestion
791	method. Environmental Pollution, 156(3), 1218-1226.

- 793 Yu, Y. X., Li, J. L., Zhang, X. Y., Yu, Z. Q., Van de Wiele, T., Han, S. Y., Wu, M. H., Sheng,
- G. Y., & Fu, J. M. (2010). Assessment of the bioaccessibility of polybrominated diphenyl
 ethers in foods and the correlations of the bioaccessibility with nutrient contents. *Journal of Agricultural and Food Chemistry*, *58(1)*, 301-308.
- 797
- Yu, Y., Pang, Y., Zhang, X., Li, C., Yu, Z., & Fu, J. (2011). Optimization of an in vitro
- method to measure the bioaccessibility of polybrominated diphenyl ethers in dust using
- response surface methodology. *Journal of Environmental Sciences*, *23(10)*, 1738-1746.

Highlights

- An in vitro digestion protocol was set up to assess PCB bioaccessibility in meat
- The mean PCB bioaccessibility was 26% in 11% fat ground beef
- PCB bioaccessibility and mean uptakes increased when the fat level decreased in meat
- PCB bioaccessibility and mean uptakes decreased when meat is well-done cooked
- PCB bioaccessibility and mean uptakes were lower in elderlies than in adults

= 3) after medium cooking. WHO-TEF are toxic equivalency factors defined by World Health Organization to assess the toxicity of dioxin-like PCBs. ^{a,b}: The letters represent the groups given by the Newman-Keuls mean comparison test when compared for each compound, based on the Table 1. Bioaccessibility (BA) (%) and relative standard deviation (RSD) (%) of spiked PCBs in 5% fat meat (n = 3) and 11% fat meat (nbioaccessibility obtained for the different fat contents. Two groups with distinct letters can be considered as significantly different (p < 0.05).

			5% fat m	eat	11% fat n	neat
Compound	Congener	TEF	BA	RSD	BA	RSD
		1 1 1 1	(0)	(%)	(%)	(%)
2,4,4'-Trichlorobiphenyl	28		49.5^{b}	9.6	24.9^{a}	1.4
2,2',5,5'-Tetrachlorobiphenyl	52		48.7 ^b	12.4	25.9^{a}	2.2
3,3',4,4'-Tetrachlorobiphenyl	LL	0.0001	44.4 ^b	9.8	20.6^{a}	6.1
3,4,4',5-Tetrachlorobiphenyl	81	0.0003	45.9 ^b	7.8	23.0^{a}	4.3
2,2',4,5,5'-Pentachlorobiphenyl	101		48.1^{b}	4.7	26.9^{a}	3.5
2,3,3',4,4'-Pentachlorobiphenyl	105	0.00003	47.5 ^b	12.5	26.4^{a}	2.7
2,3,4,4',5-Pentachlorobiphenyl	114	0.00003	47.6^{b}	8.6	25.2 ^a	5.0
2,3',4,4',5-Pentachlorobiphenyl	118	0.00003	48.4^{b}	9.1	24.9^{a}	6.1
2',3,4,4',5-Pentachlorobiphenyl	123	0.00003	48.7 ^b	8.7	25.8^{a}	5.1
3,3',4,4',5-Pentachlorobiphenyl	126	0.1	44.7 ^b	7.7	24.0^{a}	2.7
2,2',3,4,4',5'-Hexachlorobiphenyl	138		48.4^{b}	8.0	27.5 ^a	6.7
2,2',4,4',5,5'-Hexachlorobiphenyl	153		50.1^{b}	8.3	27.9^{a}	6.1
2,3,3',4,4',5-Hexachlorobiphenyl	156	0.00003	46.8^{b}	7.6	26.7^{a}	6.2
2,3,3',4,4',5'-Hexachlorobiphenyl	157	0.00003	46.5^{b}	6.0	23.8^{a}	4.7
2,3',4,4',5,5'-Hexachlorobiphenyl	167	0.00003	48.2^{b}	9.3	28.0^{a}	2.3
3,3',4,4',5,5'-Hexachlorobiphenyl	169	0.03	44.2 ^b	3.0	25.9 ^a	3.0
2,2',3,4,4',5,5'-Heptachlorobiphenyl	180		49.4^{b}	4.8	30.5 ^a	4.7
2,3,3',4,4',5,5'-Heptachlorobiphenyl	189	0.00003	47.3^{b}	4.8	28.5 ^a	3.1

simulating the digestion of an infant (n=3), an adult (n=3) or an elderly person (n=3). ^{a,b,c}: The letters represent the groups given by the Table 2. Bioaccessibility (BA) (%) and relative standard deviation (RSD) (%) of spiked PCBs in medium-cooked meat (11% of fat) after Newman-Keuls mean comparison test when compared for each compound, based on the bioaccessibility obtained for the different consumer ages. Two groups with distinct letters can be considered as significantly different (p < 0.05)

			Infa	ant	Adi	ult	Elde	srly
Compound	Congener	WHO-	BA	RSD	BA	RSD	BA	RSD
		I TT	(%)	(%)	(%)	(%)	(%)	(%)
2.4.4'-Trichlorobiphenyl	28		7.0 ^a	14.1	24.9°	1.4	15.0 ^b	5.0
2.2'.5.5'-Tetrachlorobiphenyl	52		6.4^{a}	9.6	25.9 °	2.2	15.5 ^b	5.0
3.3'.4.4'-Tetrachlorobiphenyl	TT	0.0001	7.2 ^a	9.7	20.6°	6.1	14.6^{b}	9.5
3.4.4'.5-Tetrachlorobiphenyl	81	0.0003	6.8^{a}	4.6	23.0°	4.3	15.0^{b}	5.1
2.2'.4.5.5'-Pentachlorobiphenyl	101		8.0 ^a	10.2	26.9°	3.5	17.9 ^b	4.6
2.3.3'.4.4'-Pentachlorobiphenyl	105	0.00003	7.7 а	8.3	26.4°	2.7	$16.8^{\rm b}$	5.8
2.3.4.4'.5-Pentachlorobiphenyl	114	0.00003	7.5 ^a	8.1	25.2 °	5.0	$16.7^{\rm b}$	5.8
2.3'.4.4'.5-Pentachlorobiphenyl	118	0.00003	7.3 ^a	6.5	24.9 °	6.1	$16.3^{\rm b}$	6.3
2'.3.4.4'.5-Pentachlorobiphenyl	123	0.00003	8.1 ^a	7.8	25.8 °	5.1	17.7 ^b	4.0
3.3'.4.4'.5-Pentachlorobiphenyl	126	0.1	7.2 ^a	12.9	24.0°	2.7	15.3 ^b	5.5
2.2'.3.4.4'.5'-Hexachlorobiphenyl	138		8.6 ^a	8.7	27.5 °	6.7	$18.4^{\rm b}$	4.2
2.2'.4.4'.5.5'-Hexachlorobiphenyl	153		8.8 ^a	9.5	27.9 °	6.1	18.6^{b}	4.8
2.3.3'.4.4'.5-Hexachlorobiphenyl	156	0.00003	7.4 ^a	9.7	26.7°	6.2	$18.1^{\rm b}$	3.9
2.3.3'.4.4'.5'-Hexachlorobiphenyl	157	0.00003	8.2 ^a	6.7	23.8°	4.7	18.5 ^b	3.2
2.3'.4.4'.5.5'-Hexachlorobiphenyl	167	0.00003	8.4 ^a	6.6	28.0°	2.3	$18.1^{\rm b}$	5.8
3.3'.4.4'.5.5'-Hexachlorobiphenyl	169	0.03	6.2 ^a	14.3	25.9°	3.0	17.0 ^b	5.3
2.2'.3.4.4'.5.5'-Heptachlorobiphenyl	180		9.1 ^a	7.8	30.5 °	4.7	$19.4^{\rm b}$	3.7
2.3.3'.4.4'.5.5'-Hentachlorohinhenvl	189	0.00003	8.0^{a}	6.3	28.5 °		19.1 ^b	4.5

РСВ	Sample 1	Sample 2	Sample 3	Sample 4
congener	2.9% of fat	5.0% of fat	5.3% of fat	9.5% of fat
	10.3 ng PCBs/g fat	48.6 ng PCBs/g fat	8.7 ng PCBs/g fat	5.0 ng PCBs/g fat
PCB 28	61.6	63.9	76.8	70.8
PCB 52	56.4	51.4	65.8	48.7
PCB 77	44.9	46.5	48.8	35.8
PCB 81	46.2	45.1	49.3	38.5
PCB 101	61.2	59.8	65.9	45.7
PCB 105	40.9	39.8	39.7	23.9
PCB 114	44.7	38.3	54.5	37.1
PCB 118	45.0	34.8	47.9	22.3
PCB123	48.9	54.8	57.2	35.3
PCB 126	43.6	47.1	42.5	30.9
PCB 138	31.8	34.5	41.2	20.4
PCB 153	37.9	36.6	39.7	19.9
PCB 156	40.1	38.0	42.9	25.0
PCB 157	39.2	38.3	42.6	22.4
PCB 167	44.8	35.2	45.4	23.0
PCB 169	38.2	48.2	41.3	27.0
PCB 180	39.1	37.0	38.8	16.0
PCB 189	35.9	37.4	41.4	20.5

 Table 3. Bioaccessibility (%) of PCBs in four medium-cooked naturally contaminated samples after simulating adult digestion.



Fig 1: Experimental procedure showing sample preparation, *in vitro* digestion and PCB extraction and analysis.



Fig 2: GC×GC-ToF/MS contour plot with the 18 most relevant PCB congeners to monitor in meat. X-axis and y-axis represent the retention of molecules on the first and the second chromatographic column, respectively. The column set consisted of a Rtx-Dioxin2 (60 m \times 0.25 mm \times 0.25 μ m) as the first dimension column and a BPX-50 (2 m × 0.1 mm × 0.1 μ m) as the second dimension column according to Planche *et al.* (2015).



Fig 3: Mean uptakes according to meat fat level (a), cooking intensity (b) or consumer age (c). 3a): red color is 5% fat, blue color is 11% fat; adult population, meat medium cooked. 3b): from the lightest to the darkest: rare, medium, well-done; 11% fat, adult population. 3c): pink color for children, blue color for adults, and brown color for seniors - 11% fat, meat medium cooked.

Table S1. Recovery rate (%) of PCBs in the bioaccessible fraction after a liquid/liquid

extraction, with hexane or dichloromethane (DCM) as extraction solvent.

Compound name	PCB congener	WHO-TEF (Toxic equivalency factor)	Recovery in the bioacce depending on solv	v rate (%) ssible fraction the extraction vent
)	Hexane	DCM
2,4,4'-Trichlorobiphenyl	28		56 ± 0.7	75 ± 12
2,2',5,5'-Tetrachlorobiphenyl	52		72 ± 3.3	72 ± 8.0
2,2',4,5,5'-Pentachlorobiphenyl	101		93 ± 3.1	71 ± 4.9
3,4,4',5-Tetrachlorobiphenyl	81	0,0003	97 ± 13.0	79 ± 8.6
3,3',4,4'-Tetrachlorobiphenyl	77	0,0001	98 ± 8.7	83 ± 8.9
2',3,4,4',5-Pentachlorobiphenyl	123	0,00003	102 ± 13.0	82 ± 4.0
2,3',4,4',5-Pentachlorobiphenyl	118	0,00003	104 ± 16.0	83 ± 5.2
2,3,4,4',5-Pentachlorobiphenyl	114	0,00003	107 ± 9.7	83 ± 5.6
2,2',4,4',5,5'-Hexachlorobiphenyl	153		93 ± 8.0	89 ± 4.8
2,3,3',4,4'-Pentachlorobiphenyl	105	0,00003	104 ± 4.4	86 ± 3.9
2,2',3,4,4',5'-Hexachlorobiphenyl	138		98 ± 6.5	95 ± 1.3
3,3',4,4',5-Pentachlorobiphenyl	126	0,1	105 ± 13.0	85 ± 2.0
2,3',4,4',5,5'-Hexachlorobiphenyl	167	0,00003	106 ± 15.0	95 ± 5.7
2,3,3',4,4',5-Hexachlorobiphenyl	156	0,00003	112 ± 19.0	103 ± 7.9
2,3,3',4,4',5'-Hexachlorobiphenyl	157	0,00003	107 ± 10.0	101 ± 6.4
2,2',3,4,4',5,5'-Heptachlorobiphenyl	180		103 ± 20.0	118 ± 6.5
3,3',4,4',5,5'-Hexachlorobiphenyl	169	0,03	107 ± 8.9	105 ± 6.4
2,3,3',4,4',5,5'-Heptachlorobiphenyl	189	0,00003	109 ± 8.9	110 ± 6.2

Table S2. Elution order of the 18 PCBs targeted in meat using GC×GC-ToF/MS with a

Rtx-Dioxin2/BPX-50 column set. ${}^{1}r_{t}$ (s) and ${}^{2}r_{t}$ (s) are the retention time in seconds on

respectively the first and the second dimension.

Compound name	PCB congener	WHO-TEF (Toxic equivalency factor)	$^{l}r_{t}(\mathbf{s})$	$^{2}r_{t}(\mathbf{s})$
2,4,4'-Trichlorobiphenyl	28		1580	3,2
2,2',5,5'-Tetrachlorobiphenyl	52		1610	3,4
2,2',4,5,5'-Pentachlorobiphenyl	101		2085	3,6
3,4,4',5-Tetrachlorobiphenyl	81	0,0003	2265	3,7
3,3',4,4'-Tetrachlorobiphenyl	77	0,0001	2320	3,8
2',3,4,4',5-Pentachlorobiphenyl	123	0,00003	2385	3,7
2,3',4,4',5-Pentachlorobiphenyl	118	0,00003	2415	3,7
2,3,4,4',5-Pentachlorobiphenyl	114	0,00003	2465	3,9
2,2',4,4',5,5'-Hexachlorobiphenyl	153		2490	3,7
2,3,3',4,4'-Pentachlorobiphenyl	105	0,00003	2545	4,0
2,2',3,4,4',5'-Hexachlorobiphenyl	138		2630	3,9
3,3',4,4',5-Pentachlorobiphenyl	126	0,1	2730	3,9
2,3',4,4',5,5'-Hexachlorobiphenyl	167	0,00003	2795	3,8
2,3,3',4,4',5-Hexachlorobiphenyl	156	0,00003	2915	4,0
2,3,3',4,4',5'-Hexachlorobiphenyl	157	0,00003	2930	4,1
2,2',3,4,4',5,5'-Heptachlorobiphenyl	180		2960	4,0
3,3',4,4',5,5'-Hexachlorobiphenyl	169	0,03	3125	3,9
2,3,3',4,4',5,5'-Heptachlorobiphenyl	189	0,00003	3280	4,1

Table S3. Lipid relative composition (n=3) of 11% and 5% fat meat (g/100 g fatty acids). Total lipids were extracted from skeletal muscle as described by Floch *et al.* (1) and the organic phase was evaporated under nitrogen. Analysis of fatty acid methyl-esters (FAMEs) was performed according to Pinel *et al.* (2) on a gas chromatograph (Thermo Electron Corporation, Waltham, MA, USA) coupled with a flame ionization detector using a select FAME (Agilent Technologies, Les Ulis, France) column (0.25 mm inner diameter, 100 m., 0.25 µm film thickness) and helium as the carrier gas (2.6 bar, constant pressure, inlet temperature of 250 °C).

	5% fat meat	11% fat meat
SFA	45 ± 0.25	49.0 ± 0.17
MUFA	44 ± 0.69	45.0 ± 0.35
PUFA	6.5 ± 0.37	3.0 ± 0.03
LA and ALA	4.7 ± 0.19	3.2 ± 0.04

SFA: saturated fatty acids; MUFA: monounsaturated fatty acids; PUFA: polyunsaturated fatty acids; LA: linoleic acid; ALA: alpha-linolenic acid.

(1) Folch, J.; Lees, M.; Sloane Stanley, G.H. A simple method for the isolation and purification of total lipides from animal tissues. *J Biol Chem* 1957, 226, 497-509.

(2) Pinel, A., Rigaudière, J. P., Jouve, C., & Capel, F. (2018). Modulation of insulin resistance and the adipocyte-skeletal muscle cell cross-talk by LCn-3PUFA. *International journal of molecular sciences*, *19(9)*, 2778.



Fig S1. Unsupervised discrimination of meat samples according to the bioaccessibility (%) of 18 targeted PCB. Meat samples differed according to level of meat fat (5 or 11%), intensity of meat cooking (raw, rare, medium and well done) before digestion, and age of consumer (child, adult, elderly)

different intensities (rare (n = 3), medium (n = 3) and well-done (n = 3)). ^{a,b,c}: The letters represent the groups given by the Newman-Keuls mean comparison test when compared for each compound, based on the bioaccessibility obtained for the different cooking intensities. Two groups Table S4. Bioaccessibility (BA) (%) and relative standard deviation (RSD) (%) of spiked PCBs in raw meat (n = 3) and in meat cooked at with distinct letters can be considered as significantly different (p < 0.05)

		OHM	Raw r	neat	Rare (sooked	Medium	cooked	Well don	le cooked
Compound	Congener	TEF -	μΛ	USA	n n	DSD -	R A	un DSD	BA BA	DOD
		1711								
			(0/)	(0/)	(0/)	(0/)	(0/)	(0/)	(0/)	(0/)
2,4,4'-Trichlorobiphenyl	28		26.1^{a}	13.5	24.5 ^a	7.4	24.9^{a}	1.4	22.6^{a}	5.1
2,2',5,5'-Tetrachlorobiphenyl	52		27.4^{a}	11.8	26.9^{a}	8.9	25.9^{a}	2.2	24.1^{a}	2.3
3,3',4,4'-Tetrachlorobiphenyl	LL	0.0001	24.1^{b}	12.7	20.5^{ab}	4.7	20.6^{ab}	6.1	18.3^{a}	6.7
3,4,4',5-Tetrachlorobiphenyl	81	0.0003	24.6°	8.1	21.4 ^b	2.0	23.0^{bc}	4.3	19.0^{a}	5.1
2,2',4,5,5'-Pentachlorobiphenyl	101		28.9^{a}	7.5	27.3^{a}	6.8	26.9^{a}	3.5	26.7^{a}	4.2
2,3,3',4,4'-Pentachlorobiphenyl	105	0.00003	26.1^{b}	7.1	24.2 ^b	5.4	26.4^{b}	2.7	20.5^{a}	7.5
2,3,4,4',5-Pentachlorobiphenyl	114	0.00003	25.8^{a}	10.1	24.2 ^a	4.5	25.2 ^a	5.0	22.6^{a}	6.5
2,3',4,4',5-Pentachlorobiphenyl	118	0.00003	25.5^{a}	9.3	24.3 ^a	4.3	24.9^{a}	6.1	21.5^{a}	5.9
2',3,4,4',5-Pentachlorobiphenyl	123	0.00003	27.0^{a}	8.5	26.4^{a}	6.8	25.8^{a}	5.1	25.8^{a}	2.5
3,3',4,4',5-Pentachlorobiphenyl	126	0.1	25.2 ^b	10.0	22.1 ^b	4.3	24.0^{b}	2.7	18.0^{a}	8.1
2,2',3,4,4',5'-Hexachlorobiphenyl	138		27.4^{a}	8.4	26.7^{a}	7.5	27.5 ^a	6.7	24.6^{a}	5.0
2,2',4,4',5,5'-Hexachlorobiphenyl	153		28.2^{a}	8.7	28.2^{a}	6.6	27.9^{a}	6.1	26.3^{a}	7.2
2,3,3',4,4',5-Hexachlorobiphenyl	156	0.00003	27.7^{b}	8.7	25.7^{b}	1.0	26.7^{b}	6.2	21.9^{a}	6.2
2,3,3',4,4',5'-Hexachlorobiphenyl	157	0.00003	27.5 ^b	10.9	25.4 ^{ab}	2.4	23.8^{ab}	4.7	21.9^{a}	5.5
2,3',4,4',5,5'-Hexachlorobiphenyl	167	0.00003	28.0^{b}	6.0	25.9^{ab}	6.0	28.0^{b}	2.3	23.3^{a}	11.6
3,3',4,4',5,5'-Hexachlorobiphenyl	169	0.03	23.3^{ab}	9.1	23.7^{ab}	3.8	25.9^{b}	3.0	21.2 ^a	5.7
2,2',3,4,4',5,5'-Heptachlorobiphenyl	180		29.0^{ab}	11.1	27.9^{ab}	4.3	30.5^{b}	4.7	24.4^{a}	7.2
2,3,3',4,4',5,5'-Heptachlorobiphenyl	189	0.00003	35.1°	13.9	27.9^{b}	4.1	28.5^{b}	3.1	21.7^{a}	7.5



Fig S2. Mean intakes according to consumer age.



Fig S3. Mean meat consumption (g/week) according to consumer age.