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Prenatal exposure to paraquat and nanoscaled TiO₂ aerosols alters the gene expression of the developing brain

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2 Abstract

3 Nanopesticides are innovative pesticides involving engineered nanomaterials in their formulation to increase the efficiency of plant protection products, while mitigating their 4 5 environmental impact. Despite the predicted growth of the nanopesticide use, no data is 6 available on their inhalation toxicity and the potential cocktail effects between their components. In particular, the neurodevelopmental toxicity caused by prenatal exposures 7 8 might have long lasting consequences. In the present study, we repeatedly exposed gestating 9 mice in a whole-body exposure chamber to three aerosols, involving the paraquat herbicide, nanoscaled titanium dioxide particles (nTiO₂), or a mixture of both. Particle number 10 11 concentrations and total mass concentrations were followed to enable a metrological followup of the exposure sessions. Based on the aerosols characteristics, the alveolar deposited dose 12

in mice was then estimated. RNA-seq was used to highlight dysregulations in the striatum of 13 pups in response to the *in utero* exposure. Modifications in gene expression were identified at 14 post-natal day 14, which might reflect neurodevelopmental alterations in this key brain area. 15 The data suggest an alteration in the mitochondrial function following paraquat exposure, 16 17 which is reminiscent of the pathological process leading to Parkinson disease. Markers of 18 different cell lineages were dysregulated, showing effects, which were not limited to dopaminergic neurons. Exposure to the nTiO₂ aerosol modulated the regulation of cytokines 19 20 and neurotransmitters pathways, perhaps reflecting a minor neuroinflammation. No synergy 21 was found between paraquat and nTiO2. Instead, the neurodevelopmental effects were 22 surprisingly lower than the one measured for each substance separately.

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25 Introduction

In response to the growing worldwide demand for food, nanotechnology applications in 26 27 agriculture elicit a particular interest, taking the form of innovative nanofertilizers and nanopesticides (NPe). This term mostly refers to either pesticides with nano-sized active 28 29 substances or mixtures in which engineered nanomaterials (ENM) are mixed with conventional active ingredients to control their transport, release, or degradation (Sun et al., 30 31 2019). ENMs are defined as intentionally produced products with any external dimension or an internal surface structure at the nanoscale, i.e., in the 1 - 100 nm size range 32 (ISO/TS 80004-1:2015). Overall, nanopesticides may represent important technological 33 innovations to address the sanitary and environmental drawbacks of pesticides, as they could 34 35 decrease the required tonnages of active substances. More than 1000 patents linked with these 36 technologies have already been granted in the past years (Kah et al., 2019) and a rapid increase of NPe use is expected (Gilbertson et al., 2020). 37

38 Several authors already emphasized the knowledge gap surrounding the sanitary impact of 39 these emerging products (Kah et al., 2018; Sun et al., 2019; Agathokleous et al., 2020) and health agencies highlighted the necessity to assess the toxicological consequences of a 40 deliberate release of ENM in the environment (USDA, 2015). ENMs can cross the placenta 41 42 (Bongaerts et al., 2020; Guillard et al., 2020) and the brain-blood barrier (Bencsik et al., 2018; Cena and Jativa, 2018; Boyes and van Thriel, 2020) and might thus alter fetal brain 43 development. Subcutaneous administration of nTiO₂ into gestating mice notably increases 44 45 dopamine levels in the cortex and striatum of the offspring (Takahashi et al., 2010; Umezawa et al., 2012). Prenatal exposure to nTiO₂ also has long term consequences in rats, enhancing 46 depressive-like behaviors at adult stage (Cui et al., 2014). 47

48 ENMs contained in nanopesticides might also exert an adverse effect by potentiating the 49 toxicity of pesticides. Although the current regulation enforces to assess the toxicity of both

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50 ENMs and pesticides, the novel properties arising from their combination is not carefully 51 addressed in current test guidelines (Kookana et al., 2014). However, ENMs can modify the 52 absorption, biodistribution and reactivity of associated chemicals (Naasz et al., 2018). 53 Synergistic hepatotoxicity of a mixture of paraquat (PQ) and SiO₂ nanoparticles has also been 54 reported in mice (Nishimori et al., 2009).

We address here the possible occurrence of synergistic effects induced by the prenatal 55 exposure to aerosols involving a mixture of an active substance and an ENM. As a proof of 56 principle, we associated PQ with TiO₂ P25 (nTiO₂), which is both an ENM suitable for 57 agricultural applications (Wang et al., 2016) and a reference toxicological nanomaterial 58 extensively studied in the literature (Shi et al., 2013; Zhang et al., 2015). We prioritized the 59 60 developmental period because of its high sensitivity to environmental factors (Barouki et al., 61 2012) and the neurodevelopment because any adverse effect at this stage could have irreversible consequences. 62

PQ was selected for its known neurotoxicity. Exposures of gestating mice (20 mg/kg/day) 63 64 provokes irreversible neurobehavioral and cognitive deficits in the offspring (Ait-Bali et al., 2016). Human occupational exposure to PQ is also suspected to increase the occurrence of 65 Parkinson Disease (Kamel et al., 2007; Tanner et al., 2011; Wang et al., 2011; Moisan et al., 66 2015; Brouwer et al., 2017). Although the association of PQ and P25 nTiO₂ has not been 67 reported in the fields, it might be suitable for agriculture, as the nTiO₂ photocatalytic activity 68 can degrade PQ under UV illumination (Florencio et al., 2004; Nicosia et al., 2021). Their 69 70 association could reduce the half-life of PQ in soils, which otherwise can reach several years 71 (Sartori and Vidrio, 2018).

We chose to expose mice by inhalation, as it is a major route of occupational exposure to pesticides. Depending on the spraying conditions and the volatility of the pesticide formulation, 10-75% of the applied tonnages are unintentionally disseminated in the air (Pimentel, 2014). As a result, low concentrations of pesticide residues can be detected in air samples of both rural and urban sites (Schummer et al., 2010; Coscollo et al., 2013). In this study, we used an ad hoc aerosol exposure chamber for rodents, which was previously characterized using a metrological protocol, to generate PQ, $nTiO_2$ and PQ + $nTiO_2$ aerosols (Hamdaoui et al., 2021).

80 Material and methods

81 Animals and experimental design

82 Experimentations were conducted according to the European regulation concerning the protection of animals dedicated to scientific purposes (European Directive 2010/63/EU). The 83 protocol was approved by the local Ethics Committee (CECCAPP) and the French Ministry 84 85 for Research and Higher Education (CECCAPP_ENS_2019_015). Two-month old primiparous C57Bl6/J mice (gestation day 4, GD4, purchased from Charles Rivers) were 86 randomly assigned to hosting cages (2 females/cage) and housed in standard environmental 87 88 conditions (humidity of 55 \pm 10%; temperature 22 \pm 2°C; 12/12 h light/dark cycle) at the Plateau de Biologie Expérimentale de la Souris (SFR BioSciences Gerland - Lyon Sud, 89 France Accreditation N° D54-547-10). Animals had free access to water and a supplemented 90 91 diet for pregnant mice (Safe 105 rodent pellets).

92 Groups of gestating mice were exposed to aerosols of either PQ (0.1 mg/m³), nTiO₂ (10 93 mg/m^3) or PQ + nTiO₂ (10 mg/m³). These target concentrations were chosen to match the 94 values of occupational exposures reported in the literature, in $nTiO_2$ production sites (5) 95 mg/m³; particle size in the range of 15 - 710.5 nm) (Lee et al., 2011) and in PQ treated fields (0.12 mg/m³) (Morshed et al., 2010). One additional control group was kept in the same 96 97 conditions during the duration of the exposure sessions, but without aerosol exposure. Each group of exposure was composed of 12 gestating mice exposed at the same time. Gestating 98 mice were repeatedly exposed from GD6 to GD18 for 1.5 hour, 6 days/ week, for a total of 11 99

sessions of exposure, including an individual body weight follow-up. Male and female pups
received a lethal intraperitoneal injection (6 mL/kg) at post-natal day 14 (PND14) of a
mixture of Ketamine (33 mg/mL) and Xylazine (6.7 mg/mL). Blood was collected on ice in
heparin tubes, then centrifuged (15 min at 2,500 g) and the supernatants were stored at -80°C.
Organs were weighted and sampled, then stored at -80 °C for later analysis. The striatum was
dissected and snap frozen in liquid nitrogen.

106 Aerosol exposures

107 The aerosols and the exposure chamber used in this study have been previously characterized (Hamdaoui et al., 2021). Briefly, the aerosol exposure chamber was composed of a rodent 108 109 cage (Tecniplast GR900, internal volume of 19.8 L), used as a whole-body exposure system, 110 coupled with a nebulizer (model 3076, TSI Inc.) to produce aerosols from colloidal 111 suspensions. It was operated to generate three different aerosols i.e., nTiO₂ alone, PQ alone 112 and $nTiO_2$ mixed with PQ (PQ + $nTiO_2$). Aerosols were produced from daily prepared colloidal suspensions with PQ dichloride hydrate (Sigma Aldrich reference 856177) in 113 ultrapure water (MilliQ, Millipore, 18.2 MQ.cm resistivity) and/or bulk powders of nTiO₂ 114 (Sigma Aldrich reference718467). All colloidal suspensions were constantly stirred during the 115 nebulization process. The generated nTiO₂ aerosols were composed of nanostructured 116 agglomerates and aggregates of nTiO2 nano-objects. The stability of PQ concentration in 117 118 suspension with $nTiO_2$ was verified, to ensure that our experimental conditions did not cause its photocatalytic degradation during aerosol generations. The metrological characterization of 119 120 this experimental set-up was previously described in terms of particle number size 121 distribution, mass and number concentrations, spatial and temporal stability (Hamdaoui et al., 2021). The proper functioning of the device and the stability of aerosols were monitored 122 123 during each animal exposures in real-time, by using a condensation particle counter (model 3007, TSI Inc.) to measure the particle number concentrations inside the exposure chamber. 124

In addition, the mass concentrations of aerosols were also monitored by aerosol filtration using a 47 mm filter-holder (Pallflex® filters, EmfabTMTX40HI20-205 WW type) followed by gravimetric measurements. The chamber average concentrations did not deviate from the mean by more than \pm 20%, in accordance to the test guidelines for inhalation toxicology (OECD, TG 412).

130 Gestating mice deposited particle mass estimation

The deposition fractions of aerosols in gestating mice and human lungs were calculated using 131 the Multiple Path Particle Dosimetry Model (MPPD version 3.04). Several estimations were 132 necessary to determine the relation between mice model and human exposure by using the 133 134 alveolar surface as weighting factor (Jarabek et al., 2005). The characteristics of aerosol 135 previously described (Hamdaoui et al., 2021) were used to estimate the alveolar deposition 136 fraction for each aerosol. According to the recommended exposure limit of the National Institute for Occupational Safety and Health (NIOSH) for ultrafine TiO₂ and PQ, 300 µg/m³ 137 was used for $nTiO_2$ and 100 μ g/m³ for PQ for the calculation in human. The target 138 concentrations used in this study (10 mg for $nTiO_2$ and 100 μ g/m³ for PO) were chosen for 139 mice, with a minute ventilation of 22.2 mL/min (Tankersley et al., 1994)and an alveolar 140 surface of 0.05 m² (Stone et al., 1992). A scenario of occupational exposure was chosen, 141 142 involving a woman working for 8 months during pregnancy, with a nasal minute ventilation of 19 L/min, corresponding to workers achieving light duties (Cassee and de Winter-Sorkina, 143 2003) and a human alveolar surface of 102 m² (Stone et al., 1992). We assumed that this 144 145 worker would be exposed 8 h per day, 5 days per week, 20 days per month using the Yeh/Schum symmetric lung model without clearance. 146

147 Plasmatic markers of inflammation

148 Systemic inflammation is known to play an important role in neurologic and 149 neurodevelopmental impairments and inflammation is typically measured by quantifying

circulating cytokines. Pups plasma were thawed and pooled by litters ($n\geq 5/litter$). Fifty 150 microliters were assayed using an ELISA multiplexed kit (MCYTOMAG-70K-26, Bio-151 Techne) to measure the concentrations of 8 cytokines (MCP-1, Cxcl1, IL-1 beta, IL-4, IL-6, 152 153 IL-17, TNF-alpha) and the matrix metalloproteinase MMP12 on a Luminex platform (Luminex MAGPIX, ThermoFisher). These markers were selected for their implication in 154 neurodevelopment (MCP-1, IL-1 beta, IL-6 and TNF-alpha) (Nist and Pickler, 2019), their 155 role as pro-inflammatory factors (IL-4, IL-10, IL-17 and Cxcl1), or in extracellular matrix 156 157 remodeling (MMP12) following tissues insults. The gene *Mmp12* has already been described as an early marker of the lung insults in response to low concentrations of PQ in mice (Tomita 158 159 et al., 2007).

160 *RT-qPCR*

Total RNA was extracted using RNeasy Mini kit (Qiagen) from frozen striatum (n≥8/group). 161 On column DNase digestion was used to remove contaminating DNA. RNA concentrations 162 were measured with a Nanodrop spectrophotometer (ThermoScientific) and 1 µg of each 163 RNA sample was reverse transcribed in murine leukemia virus reverse transcriptase 164 (Promega) and random DNA hexamer primers. Quantitative PCR was performed according to 165 a standard protocol, using the Biorad iQ SYBRGreen kit and the Biorad CFX96 thermocycler. 166 167 For each pair of primers (Table S1) a standard curve was established to verify PCR efficiency (90%–110%) and quantification was performed using the $2^{-\Delta\Delta Ct}$ method (Livak and 168 Schmittgen, 2001). We verified that the Gapdh and Rplp0 housekeeping genes which were 169 170 used as internal controls do not vary significantly in our samples according to RNA-seq (Figure S1). Differences among experimental groups were analyzed with the Kruskal-Wallis 171 test followed by Dunn's post hoc test to compare mRNA relative mean levels using Prism 172 (GraphPad 8.3 Software, Inc, San Diego CA). 173

174 *RNA-seq analysis*

Striatum RNA was extracted using the Macherey-Nagel NucleoSpin RNA II kit. cDNA 175 176 libraries were prepared using the total RNA SENSE kit (Lexogen, Vienna Austria) and analyzed on an Nextseq500 sequencer (Illumina). Raw data of single-end sequencing 177 178 (GSE169579) were converted to count tables using htseq-count (Galaxy Version 0.6.1galaxy3) (Anders et al., 2015). Differential gene expression analysis was performed 179 between exposure and control groups with Deseq2 (Galaxy Version 2.1.8.3, one-factor, 180 181 parametric fit type, automatic outlier replacement and filtering, p-adjusted values <0.05; 182 minimum read number >10) (Love et al., 2015). RNASeq data are accessible through GEO Series accession number: 183 GSE143933

184 (www.ncbi.nlm.nih.gov/geo/query/acc.cgi?acc=GSE169579#).

185 Gene Set Enrichment Analysis (GSEA)

Gene Set Enrichment Analysis (https://www.gsea-msigdb.org/gsea/index.jsp) was performed 186 187 as previously described (Mooney and Wilmot, 2015)Statistical significance was tested by 188 empirical permutation followed by multiple hypotheses correction using Kolmogorov-Smirnov statistics (Subramanian et al., 2005; Tamayo et al., 2016). Several gene set databases 189 were used: hallmark, C2 curated (KEGG, REACTOME) and C5 ontology (biological process) 190 191 (https://www.gsea-msigdb.org/gsea/msigdb). A p-value cut-off of p <0.01 and an FDR <0.05 192 were used to select the significant gene sets, and only a limited number was selected to avoid 193 pathway redundancy. Gene sets were filtered by size, for too specific sets (<15 genes) or too general sets (>300 genes) to compare the control group versus each exposed group, using 194 195 1000 random permutations of the gene sets for resampling. A network analysis was performed 196 to highlight dysregulated pathways using the Cytoscape software (version 3.8.2) (Cline et al., 197 2007) with the Enrichment Map plug-in (Merico et al., 2010). The selected gene sets with a False Discovery Ratio <0.05 (FDR), were represented with edges linking gene sets showing 198 199 an overlap >30% and then clustered in function of their common biological functions.

9

201 **Results**

202 Experimental design

We exposed 3 groups of gestating mice from the inbred C57Bl6/J strain, which has a welldefined genetic background, to aerosols containing low concentration of either PQ (0.1 mg/m³), nTiO₂ (10 mg/m³) or PQ + nTiO₂ (10 mg/m³). These concentrations were selected for their relevance to human exposure (see methods for estimations). Aerosols exposure sessions were started at GD5, before the beginning of neurulation, to ensure a continuous exposure during the entire neurodevelopment. The last session was at GD18, i.e., 1.5 day before birth, to limit maternal stress during the perinatal period.

210 Pups were kept with their mother until PND14, a stage at which the main neurodevelopmental 211 processes are completed. We then used an unbiased transcriptome analysis of gene expression 212 of the post-natal striatum to highlight possible dysregulations caused by in utero exposures, 213 and a possible cocktail effect. Following a previous study (Gollamudi et al., 2012), we selected this brain area for several reasons. It first represents a relatively homogenous and 214 215 well delimitated brain area, where 95% of the neurons are GABAergic medium spiny 216 neurons, known to be highly sensitive to moderate neurodevelopmental alterations (Richard et al., 2020). In addition, striatum neurons receive dopaminergic innervation from subtantia 217 218 *nigra*, one of the main sites of the PQ neurotoxicity. Thus, we expected that the striatum gene 219 expression could represent a sensitive endpoint to early neurodevelopmental insults.

220 Aerosol exposures and dosimetry assessment

The particle number concentrations assessed in real-time within the exposure chamber were stable over the duration of exposures for the three aerosols (Figure 1) and the results of particle number and mass concentrations are reported in Table 1. The average number concentrations did not deviate from the mean by more than 9% and the mass concentrations
matched the desired target concentrations with a maximum relative standard deviation of 8%.

226 The alveolar deposition fraction in gestating mice and human beings were calculated using the MPPD model to assess the alveolar deposited mass in each species (Table S2). For nTiO₂ and 227 PQ, they were respectively 384 μ g/m² and 229 μ g/m² in human beings, and 298 μ g/m² and 9 228 $\mu g/m^2$ in mice. Therefore, mice exposures during gestation resulted in an alveolar particle 229 deposition below what a pregnant worker could theoretically be exposed to, if working during 230 231 eight months at levels of exposure corresponding to the occupational exposure limit. The deposition fraction of the aerosol of $PQ + nTiO_2$ was estimated in a similar manner. Due to 232 similar aerosol characteristics in terms of size distribution, no significant difference was found 233 234 in its alveolar deposition in comparison with nTiO₂ alone in human being or mice. The 235 mixture resulted in interactions between PQ and nTiO₂ particles in the aerosol phase, but the MMPD modelling cannot deconvolute the deposited mass attributable to each substance 236 237 separately. Therefore, we were not able to decipher the deposition fraction in each substance (PQ versus $nTiO_2$) in the aerosol of PQ + $nTiO_2$. 238

239 General health endpoints of exposed females and litters

240 The treated mice did not display any obvious sign of discomfort. Statistical analysis did not evidence any significant effect of aerosol exposure on gestating mice weight gain, offspring 241 242 survival, and litter body weights at different post-natal stages (Supplementary data, Figure S2). In addition, prenatal aerosol exposures did not have any incidence on the characteristics 243 of litters (number, sex ratio) or organ weight at PND14 (Table S3). Only two markers were 244 245 detected in the plasma of pups (MMP12 and Cxcl1) (Figure 2). Cxcl1 chemokine was found to be slightly but significantly increased in the PQ group of exposure in comparison to the 246 control group, which is an indication of a slight systemic inflammation in this group. 247

248 Assessing neurodevelopmental toxicity by transcriptome analysis

RNA was extracted from the striatum of pups at PND14 (Table S4). We first used RT-qPCR 249 to address the possible accidental presence of neighboring neural tissues in the striatum 250 samples. The expression of Drd2 and Adora2a, whose expression is restricted to striatum, and 251 252 *Fezf1* which is selectively expressed in the neighboring cortex, was homogeneous. Sixteen striatum samples (4/group, 50% males and 50% females) coming from sixteen different litters 253 were selected for RNA-seq analysis. Deseq2 was then used to identify differentially expressed 254 genes (DEGs) in each group of exposure in comparison to the control group. The Principal 255 256 Component Analysis (PCA) of the total normalized counts (Figure 3A) highlighted different profiles for the different groups. The first principal component, explaining 37 % of the 257 variation mainly reflected the groups exposure. The second principal component, explaining 258 26 % of the variation, mainly reflected the distinction between females and males. However, 259 intragroup comparison between males and females did not evidence different response to 260 261 aerosol exposures. Because of a hypothalamic contamination during dissection, one outlier coming from the PQ group was detected thanks to Deseq2 and excluded from the analysis. 262

Both PQ and nTiO₂ aerosols had a clear influence on gene expression. The number of 263 differentially expressed genes (DEGs) was 2038 for PQ, 1249 for nTiO₂, and only 287 for PQ 264 265 + nTiO₂. Up-regulation was more frequent that down-regulation (Figure 3B) as up-regulated genes represented 59%, 89% and 85% of the total number of DEGs for PO, $nTiO_2$ and PO + 266 nTiO₂, and the fold-changes exceeded 1.5 for 120, 61 and 36 DEGs respectively. The 267 overlapping DEGs between the three groups of exposure reached 170 (Figure 3C), suggesting 268 that some similarities between the mode of action of PQ and nTiO₂. A smaller subset of 18 269 DEGs were upregulated in the three groups of exposure, with a fold-change superior to 1.5. In 270 agreement with the previous global analysis, the combined exposure to $PQ + nTiO_2$ had 271 globally a lesser influence on these 18 genes (Figure 3D). 272

RT-qPCR was used to confirm some of the RNA-seq observations on a larger number of 273 274 animals for 8 genes selected for their known role in neurodevelopment (Figure 4). These genes were markers of neuronal differentiation (Nefl, Nefh) (Hogberg et al., 2009) reactive 275 276 astrogliosis resulting from neurological insults (*Gfap*) (Eng and Ghirnikar, 1994) myelination (Fa2h, Mobp) (Royland et al., 2008; Potter et al., 2011) GABAergic neurons maturation 277 (Kcnc2) (Fukumoto et al., 2018) and microglial activation (Chga) (Ciesielski-Treska and 278 279 Aunis, 2000). Bdnf encodes a neutrophin, which is essential for the neuron survival, growth 280 and differentiation, whose modification has been associated with mental disorders (Ikegame et al., 2013). The expression of Nefl and Nefh, which encode filament proteins expressed 281 282 respectively during early and late neuronal maturation, were both found dysregulated in the PQ group, while only *Nefl* was found dysregulated in the PQ + $nTiO_2$ group. Interestingly, the 283 284 expression of the Chga gene colocalizes with activated microglia in neurodegenerative areas 285 (Ciesielski-Treska et al., 1998), was overexpressed in all exposure groups with the same amplitude. Moreover, Kcnc2 which is highly expressed in the medium spiny neurons of 286 287 striatum was clearly upregulated in the PQ and PQ + $nTiO_2$ groups. The expressions of *Gfap* 288 and myelin-forming markers (Fa2h and Mobp) were upregulated solely in the PQ group. Overall, the dysregulation of these genes, with the possible exception of Bdnf, was more 289 290 pronounced in the PQ group, supporting the observations coming from RNA-seq data.

291 Interpretation of transcriptome changes - Gene Set Enrichment Analysis

Gene Set Enrichment Analysis (GSEA) can identify differential expression of gene sets, when the expression difference between genes is minimal and cannot be evidence by classical methods on a single gene basis. GSEA was used to interpret the broad but often modest changes in gene expression revealed by RNA-seq, to identify relevant biological pathways. Overall, 77 gene sets were found dysregulated in the striatum of at least one group of mice (Supplementary Tables S5 to S7). 298 PQ exposure resulted in the up-regulation of 7 gene sets involved in gliogenesis and axon 299 ensheathment, and 4 gene sets associated with neuron projection and axon guidance (Figure 5A). This suggested an alteration in the timing of differentiation for both oligodendrocytes 300 301 and the main neuronal cell type, the medium spiny neurons. Among the five identified pathways which were down-regulated, two were likely to reflect a possible direct effect and 302 persisting of PQ in the post-natal brain, following fetal contamination, as they are known to 303 304 be involved in the etiology of the PD. These are related to redox homeostasis (1 gene set) and 305 mitochondrial electron transport (8 genes sets), which may be linked to PQ oxidant properties. The subset of genes accounting for the enrichment of this pathway, the leading-edge genes 306 307 according to GSEA results, was analyzed using ClueGO (version 2.5.7) (Bindea et al., 2009) from the Cytoscape package in order to visualize the deregulated pathways in a 308 functionally grouped gene network. A focus presenting the 53 leading edge genes of 309 310 Parkinson disease gene set (KEGG, total of 110 genes) is presented in Figure 5B, showing 311 that 32 of them were specifically related to the oxidative phosphorylation pathway. In 312 addition, the gene set associated with the cell redox homeostasis (total of 52 genes) 313 encompassed 24 leading edge genes. Within this set, 18 genes were related to the oxidoreductase activity, involving for example the Glutathione reductase (Gsr), which is 314 necessary to maintain high levels of reduced glutathione in the cytosol (Yang et al., 2006). 315

The consequences of gestational exposure to $nTiO_2$ were also highlighted (Figure 5A), with only one down-regulated pathway (centromere complex assembly, 2 gene sets), and 4 upregulated pathways. These pathways were related to gliogenesis, axon ensheathment and guidance (5 gene sets). This is an indication that $nTiO_2$ also may have adverse effects on oligodendrocytes differentiation and neuronal maturation. The identification of the ion homeostasis pathway (5 gene sets) also suggested an alteration of the neuronal electrophysiology. The appearance of the pathway "Chemokines production and B-Cell 323

324

proliferation" pointed out a potential low grade neuroinflammation. Finally, nTiO₂ exposure also altered the "regulation of neurotransmitters" pathway (11 gene sets).

The gestational exposure to the combination of $PQ + nTiO_2$ had a visible influence on only 3 325 pathways (Figure 5A) which is consistent with its decreased influence on the gene expression. 326 These pathways are: "regulation of neurotransmitters" (12 gene sets), "neurons and dendrites 327 extension" (2 gene sets) and "voltage potassium channels" (4 gene sets). As the regulation of 328 neurotransmitters was commonly found upregulated only in the $nTiO_2$ and PQ + $nTiO_2$ 329 groups, the dysregulation of genes involved in neurotransmitters release is probably caused by 330 nTiO₂, independently of PQ. The first 50 leading edge genes accounting for the enrichment of 331 this pathway were represented using ClueGO (Figure 5B). These common genes accounting 332 333 for the enrichment of this pathway and their associated fold-changes were represented on a 334 radar plot for the $nTiO_2$ and PQ + $nTiO_2$ groups, showing a similar influence of both exposures (Figure S4). This demonstrated that $PQ + nTiO_2$ did not have an additive effect on 335 336 this pathway, which is dysregulated by both nTiO₂ and the mixture. Interestingly, two genes related to the glutamate neurotransmission were highlighted: the Glutamate receptor (Grina3) 337 and the Metabotropic glutamate receptor 4 (Grm4). α-synuclein was also found in the leading 338 339 edge set of this pathway and different trends were observed, as a slight decrease was found after exposure to PQ (Fold-change=0.80, adjusted p-value=0.01) and nTiO₂ (Fold-340 change=0.82, adjusted p-value=0.02), but not to PQ + nTiO₂. 341

342

343 **Discussion**

344 *Effect of maternal exposure to aerosols*

We present novel evidence showing that prenatal exposures to aerosols of PQ and/or $nTiO_2$ at low concentrations, which are theoretically considered safe, have a significant effect on the 347 gene expression in the developing brain . The calculation of the deposition fractions for each 348 aerosol using dosimetry modeling allowed an interspecies extrapolation. We could identify 349 some consequences of the *in utero* exposure by analyzing the gene expression in the post-350 natal striatum. In this context, transcriptome analysis proved to be an unbiased, powerful and 351 sensitive method to detect neurodevelopmental alterations.

352 Paraquat neurodevelopmental toxicity

Despite a very low deposition of PQ in maternal lungs, clear changes in gene expression were 353 observed in the striatum of pups after birth, irrespective of the gender, which reflect 354 alterations in different cell lineages. The observed changes suggest moderate defects in 355 356 neuronal maturation and oligodendrocytes differentiation. They are also indicative of 357 inflammation, with astrogliosis and microglial cells activation. Therefore, the influence of PQ 358 in the developing striatum is not limited to dopaminergic neurons and we propose that it 359 combines direct and indirect effects on neural cells. Based on previous toxicological knowledge and our observations, PQ may directly alter the mitochondrial function and the 360 redox status in neurons, during fetal and early post-natal life, as it does in adult brain to 361 induce Parkinson disease (Jomova et al., 2010). Some of the highlighted changes in gene 362 expression are possibly involved in this process: Glutathione reductase (encoded by the Gsr 363 gene) is necessary to maintain high levels of reduced glutathione in the cytosol (Yang et al., 364 2006), Selenoprotein (product of the *Selenot* gene) protects dopaminergic neurons against 365 oxidative stress in mouse models of Parkinson disease (Boukhzar et al., 2016), Parkin RBR 366 367 E3 ubiquitin protein ligase (*Prkn*), and α -synuclein (*Snca*) maintain the dopamine neuronal homeostasis, whose dysfunction is related to the onset of Parkinson disease (Kuroda et al., 368 2012; Oczkowska et al., 2013). These initial mitochondrial alterations may indirectly 369 influence the maturation of the neuronal networks, the post-natal myelination process, and 370 eventually favor neuroinflammation. Taken together, these observations suggest that PQ alters 371

mitochondrial functions in the striatum of exposed pups in a way that is reminiscent of thepathological process leading to Parkinson disease.

374 This interpretation is consistent with an analysis of the striatum transcriptome performed after acute exposure of adult mice to PQ (Gollamudi et al., 2012). This also caused 375 dysregulations of the signaling pathways related to Parkinson disease. Exposure of gestating 376 mice to a high dose of PQ via the oral route (Ait-Bali et al., 2016) also induced various 377 neurological and behavioral alterations in the offspring. In a recent study, adult C57Bl/6J 378 379 mice were exposed to PQ aerosols (Anderson et al., 2021), leading to a significant PQ burden in different brain regions, including the striatum, associated with a male-specific deficit in 380 olfaction, which is part of the prodromal features of Parkinson disease. 381

382 The amplitude of the observed changes is modest but does not rule out unforeseen latter consequences, as early alterations can exacerbate the sensitivity to additional stressors at adult 383 384 stage (Barouki et al., 2012). Previous analysis also showed that early exposure to PQ + Zn in rats affects the oxidative stress, inflammation, cell death, dopamine metabolism and storage 385 regulating machineries (Mittra et al., 2020). Therefore, re-exposure of adult rats caused more 386 pronounced changes in all the observed endpoints of toxicity, compared with rat exposed only 387 during adulthood, thereby showing that early insults emphasized the toxicity of secondary 388 389 exposures. Similar conclusions were drawn for the early exposure to PQ combined with maneb (Colle et al., 2020). 390

391 *nTiO*₂ neurodevelopmental toxicity

The neurotoxicity of $nTiO_2$ was previously addressed at higher concentrations than those used here. Subcutaneous injection of large quantities of $nTiO_2$ to gestating mice induces a mitochondrial respiration dysfunction in the pups brain (Hathaway et al., 2017) altering the gene expression in the post-natal striatum of pups (Umezawa et al., 2012) and having long term consequences on their cognitive functions (Engler-Chiurazzi et al., 2016). Prenatal 397 exposure to nTiO₂ aerosols also results in maternal and fetal microvascular dysfunction 398 (Stapleton et al., 2015) and affects the production of neurotransmitters (Shimizu et al., 2009). Our results confirm that gestational exposure to nTiO₂ has adverse effect on striatum 399 400 maturation, at a concentration, which is much lower than those used in previous studies. 401 Changes in gene expression suggest a modulation of neuronal maturation, neurotransmitter production, post-natal myelination, and possibly a low-grade neuro-inflammation. The 402 403 convergence of our conclusions with previous studies suggests that the observed changes are 404 neither transient nor adaptative but rather reflects moderate adverse effects on the brain maturation. The possibility that exposure to aerosols containing realistic concentrations of 405 406 nTiO₂ could compromise the neurodevelopment should thus be a matter of concern when considering their extensive use. 407

408 Interaction between nTiO₂ and PQ exposure.

409 We did not observe a synergistic nor an additive effect when combining PQ and nTiO₂. Instead, their association in aerosol phase seemed to lessen the effect of each substance. 410 RNA-seq and RT-qPCR showed that the number of differentially expressed genes and the 411 412 amplitude of the changes in gene expression were globally lower compared to pups exposed 413 to either PQ or nTiO₂ solely. Moreover, unlike the PQ group, the concentration of the Cxcl1 414 inflammation marker was not increased in the pups plasma. The reduction of the mixture 415 effect is not due to the PO degradation by nTiO₂, as verified previously (Hamdaoui et al., 2021) but might reflect interactions between PQ and nTiO₂ particles. The aerosols of nTiO₂ 416 417 and PQ + $nTiO_2$ presented very similar particle size distributions, and the mixture displayed 418 number concentrations that did not correspond to the sum of the reported concentrations for 419 PQ and nTiO₂ taken separately. These elements suggest that mixing these substances prior to 420 the aerosol generation does not produce two distinct populations of particles, but rather a composite population of particles associating PQ and nTiO₂. As PQ can be adsorbed on nTiO₂ 421

(Vohra and Tanaka, 2003; Florencio et al., 2004; Phuinthiang and 422 in suspension Kajitvichyanukul, 2019) these interactions are likely to occur during the nebulization process 423 and in the aerosol phase. We hypothesize that this sorption changes the deposition fraction, 424 biodistribution and particle toxicokinetics in a way which remains to be addressed. While PO 425 is suspected to cross the placental barrier (Ingebrigtsen et al., 1984) sorption to nTiO₂ might 426 prevent this translocation. It has been shown previously that nTiO₂ caused a retention of the 427 Cd²⁺ ions, thereby decreasing their bioavailability and preventing ecotoxicological effects. 428 429 Eventually, $nTiO_2$ might have the same effect on the PQ toxic forms (PQ²⁺/PQ⁺), by decreasing its bioavailability and therefore its toxicity. *Limitations of the present study* 430

We generated ultrafine PQ particles that do not strictly correspond to the primary size of 431 sprayed pesticide particles in agriculture, which are rather coarse particles (>10 µm) mostly 432 433 considered as not inhalable (Chester and Ward, 1984). Nevertheless, according to its long environmental persistence and the wind erosion of soil particles containing PQ, occupational 434 435 or residential populations can be exposed to inhalable PQ particles. This is in line with a size range of 0.1 - 1 µm measured for airborne residue of some pesticides (Coscollo et al., 2013). 436 Secondary exposure to pesticide residues therefore encompass particles that may be deposited 437 438 in the airways after their atmospheric transport.

We studied the consequences of gestational exposures to PQ and nTiO₂ aerosols at a single time-point in the striatum, which does not allow to address the reversibility of the observed changes. Also, in the absence of complementary histological, biochemical and behavioral analysis, the interpretation of these changes remains hypothetical. Finally, further work would be needed to clarify the mechanisms by which the combination of PQ and nTiO₂ seems to lower their inhalation effects, and whether this conclusion would also apply to other cocktails of substances potentially suitable for nanopesticide formulation.

446 **Conclusions**

This study demonstrates the pertinence to use an aerosol exposure chamber and RNA-seq 447 analysis to better evaluate the unknown effects of low and repeated exposures to 448 nanopesticide aerosols on neurodevelopment. Considering the growing demand for these 449 450 innovative products and the lack of precise guidelines surrounding their health risk 451 assessment, these results should reinforce awareness about their future sanitary impact. Taken together, these data also suggest that the risk associated with the exposure of aerosols 452 containing PQ or nTiO₂ during critical stages, like pregnancy and early development, may 453 454 have been previously underestimated.

455

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464

465 Authors contribution.

466 QH conceived the aerosol exposure device, performed the experiments, the data analysis, the 467 tissue collection and wrote the manuscript, YZ performed data analysis, SR, DA, RG, SM and 468 KG participated to mouse breeding and tissue collection. FGL conceived the aerosol exposure 469 device. AB conceived and supervised the experiments, FF supervised the project and wrote 470 the manuscript. All authors reviewed and edited the manuscript.

471

472 Figure legends

- 473 Figure 1. Real time measurement of aerosol particle number concentration in the chamber
 474 during animal exposures.
- 475 Data are presented as means \pm standard deviation (SD) for each aerosol.

476 Figure 2. Plasmatic Markers of inflammation in pups (PND14) exposed in utero.

Two out of 8 assessed cytokines were detectable by ELISA. No changes were detected in MMP12 expression in any of the exposure groups. The Cxcl1 expression is similar between control and pups exposed to $nTiO_2$ alone or associated to PQ alone. PQ exposures caused a significant increase in Cxcl1 in comparison to the control group (n \geq 5/litter, * p-value <0.05), thereby suggesting a low grade inflammation. Comparisons between groups were performed with the Kruskal-Wallis test (nonparametric test), followed by multiple mean comparisons with Dunn's test. Data presented as means ± SEM.

484 Figure 3. Transcriptome analysis of pups striatum. (A) PCA of the gene expression in 15 485 pups. (B) Volcano plots summarizing the comparisons between the groups of exposure versus the control group. (C) Venn diagram showing the overlap between DEGs for the 3 exposed 486 groups (adjusted p-value <0.05; no threshold on fold-change), a main part of genes is 487 488 associated with the PQ exposure only, $nTiO_2$ shared 68% of its DEGs with PQ, PQ + $nTiO_2$ 489 shared 81% with PQ and 65% with nTiO₂. (E) Fold-changes of the 18 DEGs up-regulated in 490 the 3 groups with a fold-change >1.5; note that the response to the PQ + $nTiO_2$ is always 491 equal or lower than the response to the single substances.

492 *Figure 4. RT-qPCR analysis of gene expression in pups striatum at PND14.*

493 Comparisons between groups ($N \ge 8$ for each group, 50% females and 50% males coming 494 from at least 4 different litters) were performed with the Kruskal-Wallis test, followed by 495 multiple mean comparisons with Dunn's test. * p<0.05 **p<0.01. Data presented as means ± 496 SEM.

497

498 Figure 5. Gene Set Enrichment Analysis of striatum gene expression profiles.

(A) Three MSigDB genes collections were analyzed: Gene Ontology (GO) Biological Process 499 ontology and two Canonical Pathways gene sets derived from the KEGG pathway database 500 501 and from the Reactome pathway database. Each node represents a gene set significantly 502 enriched in comparison with controls. The gene sets corresponding to each node, can be found in supplementary data (Figure S3). Nodes are clustered in pathways based on their shared 503 504 genes and their common biological functions. Edges are filtered with an overlap cutoff: the 505 minimum edge between two nodes depict that 35% of one gene set is common with the other 506 (B) ClueGO analysis of the leading-edge gene set associated with KEGG Parkinson disease in 507 the PQ group, represented to highlight the main biological function associated with those, notably the oxidative phosphorylation. (C) ClueGO analysis of the leading-edge gene set 508 509 associated with the GO regulation of neurotransmitter levels in the nTiO₂ and PQ+nTiO₂ 510 groups.

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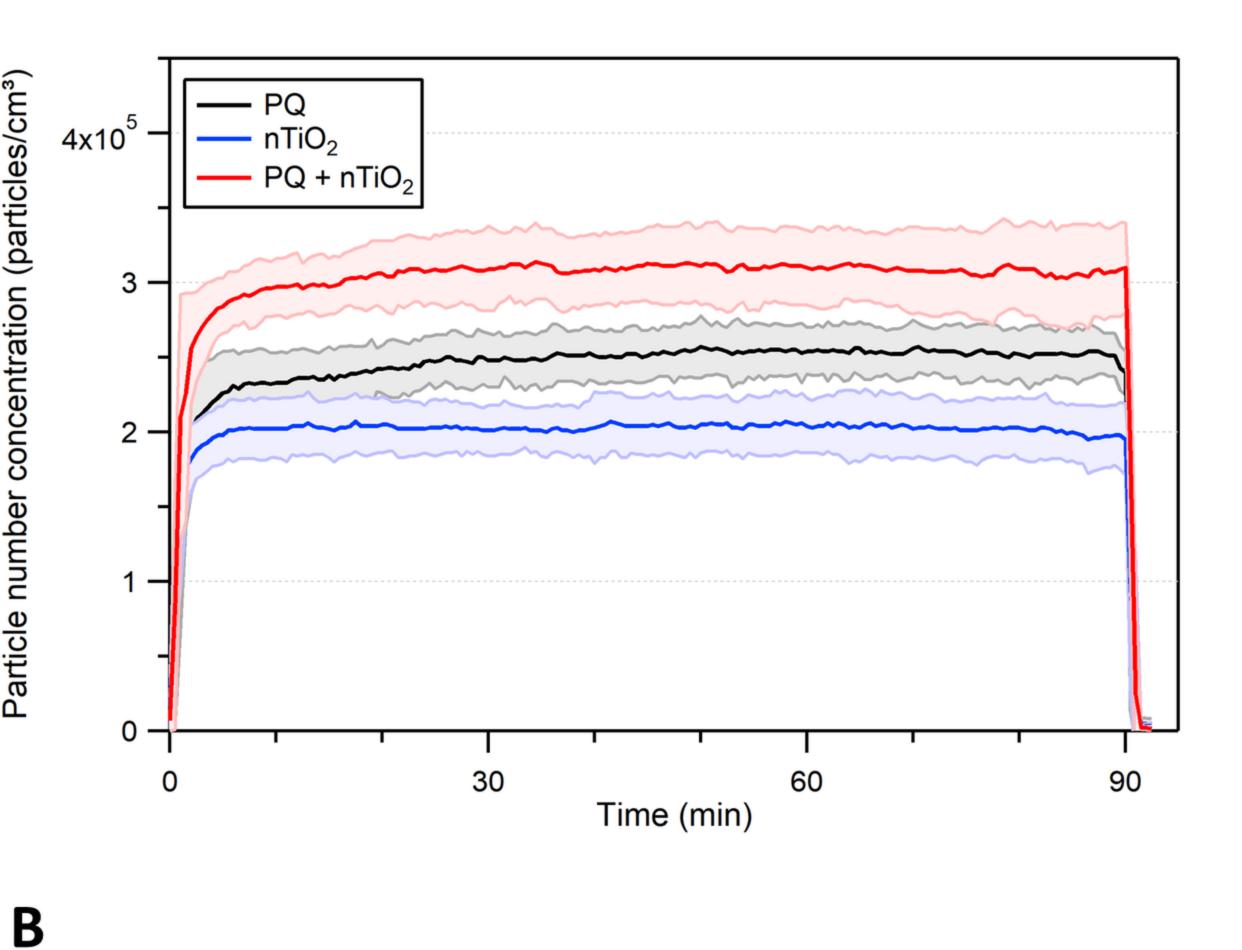
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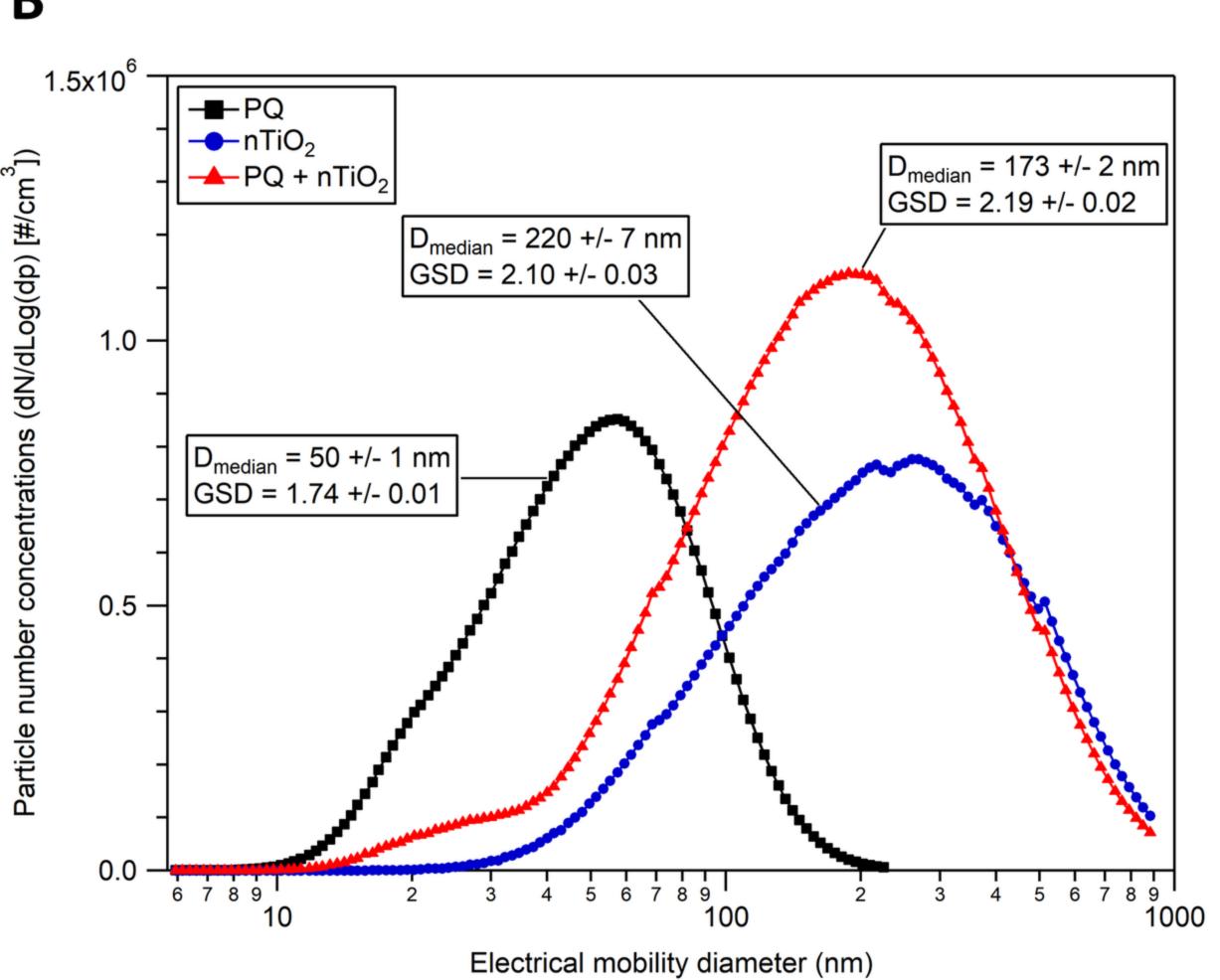
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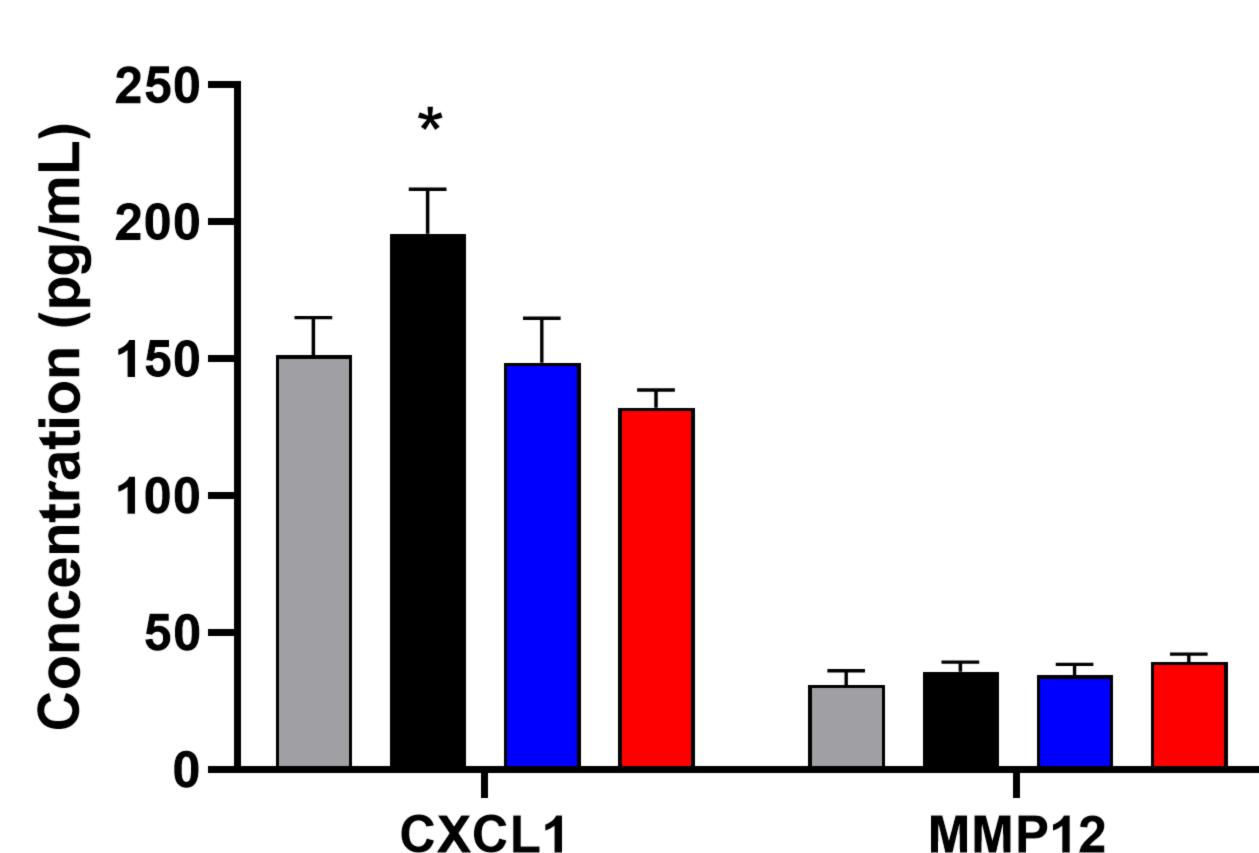
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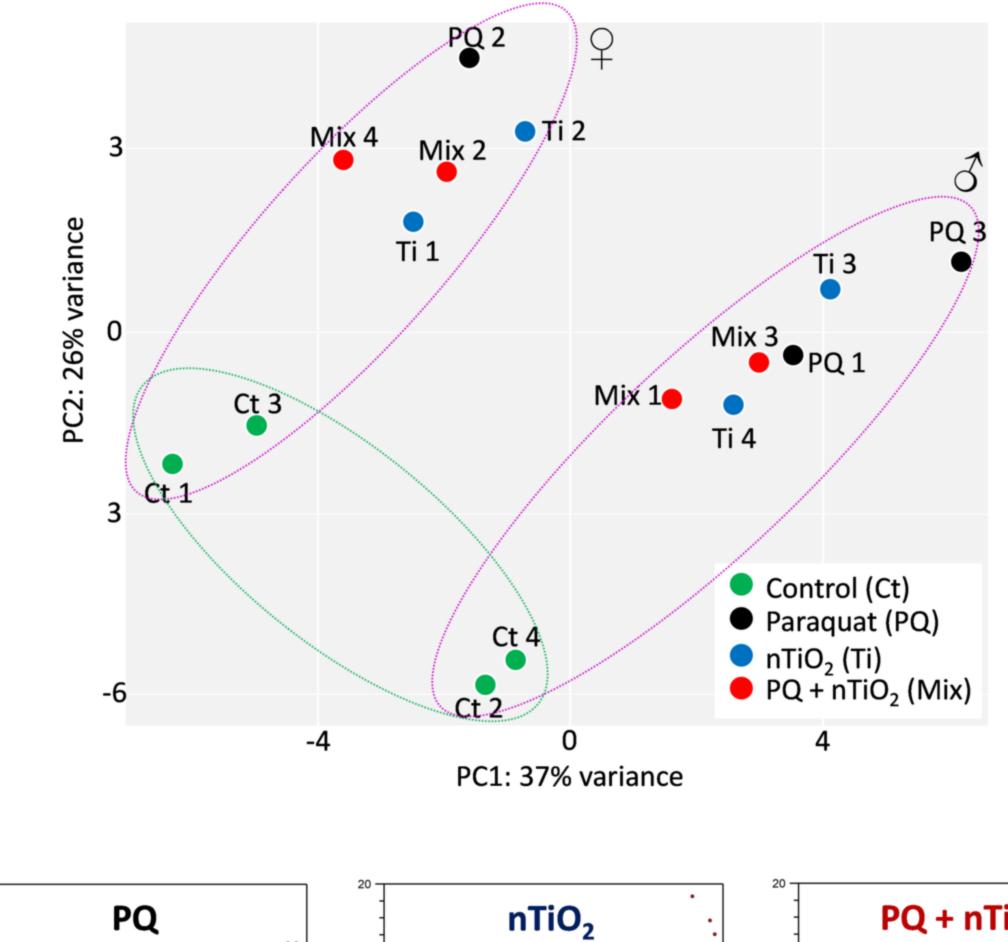
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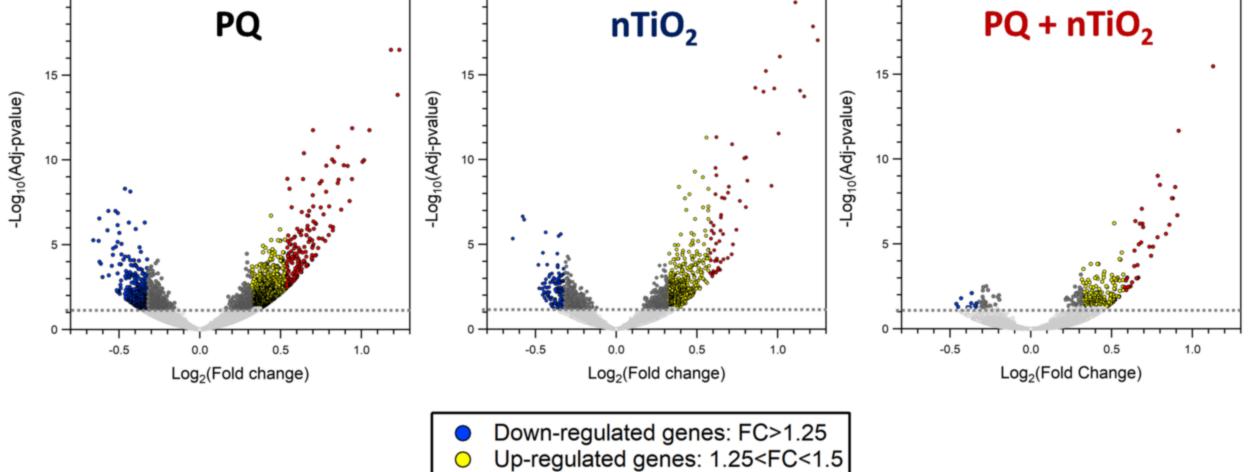




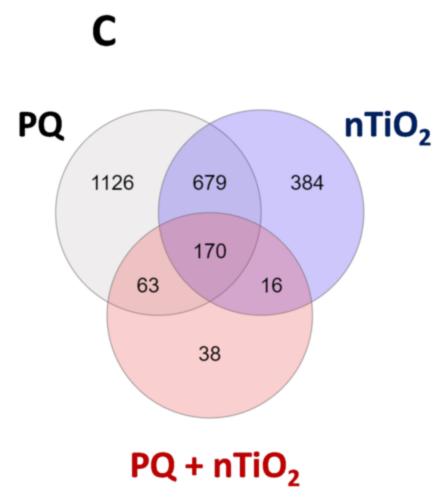


Control PQ nTiO₂ PQ + nTiO₂





Up-regulated genes: FC>1.50

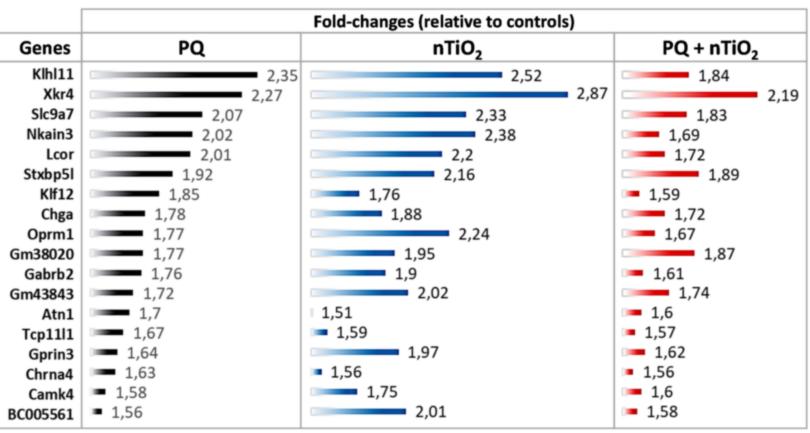


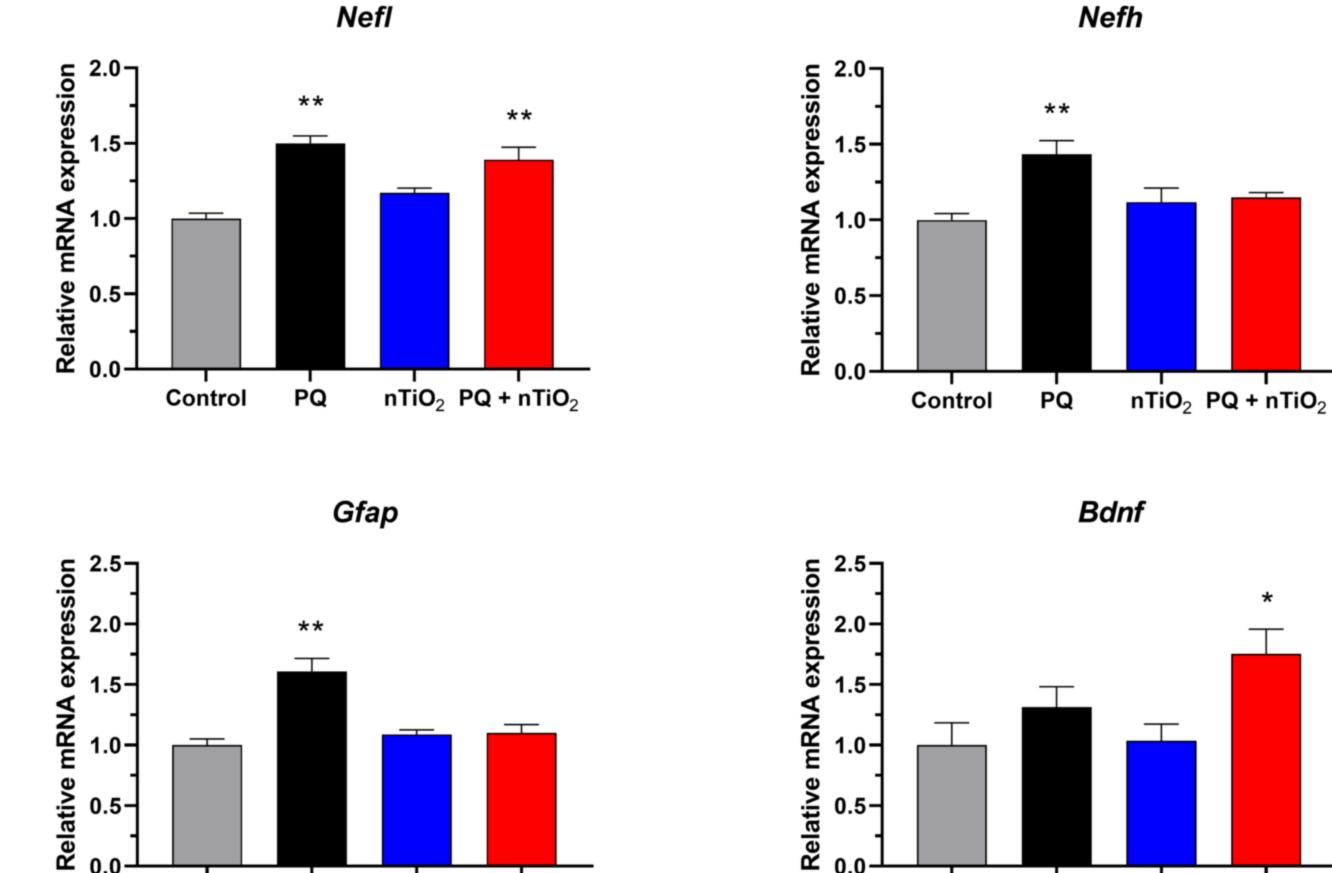
Α

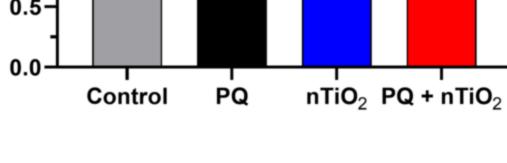
В

20

D





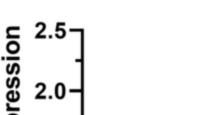


Fa2h



PQ

Control







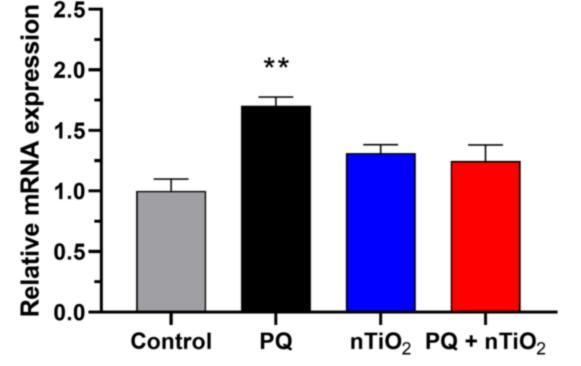
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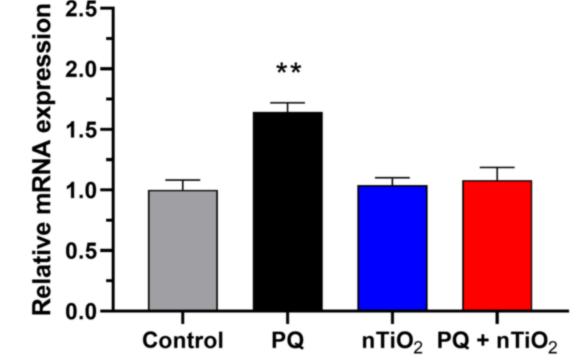
Mobp

*

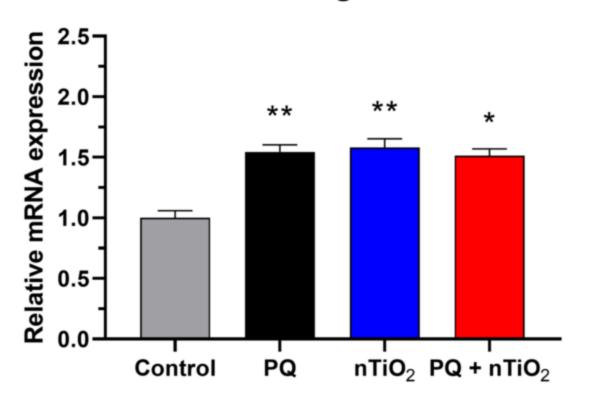
nTiO₂ PQ + nTiO₂

**

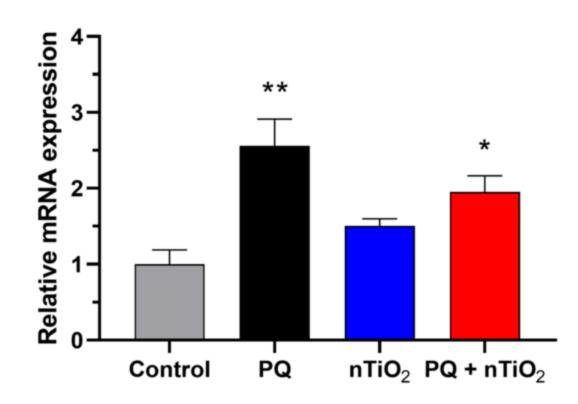


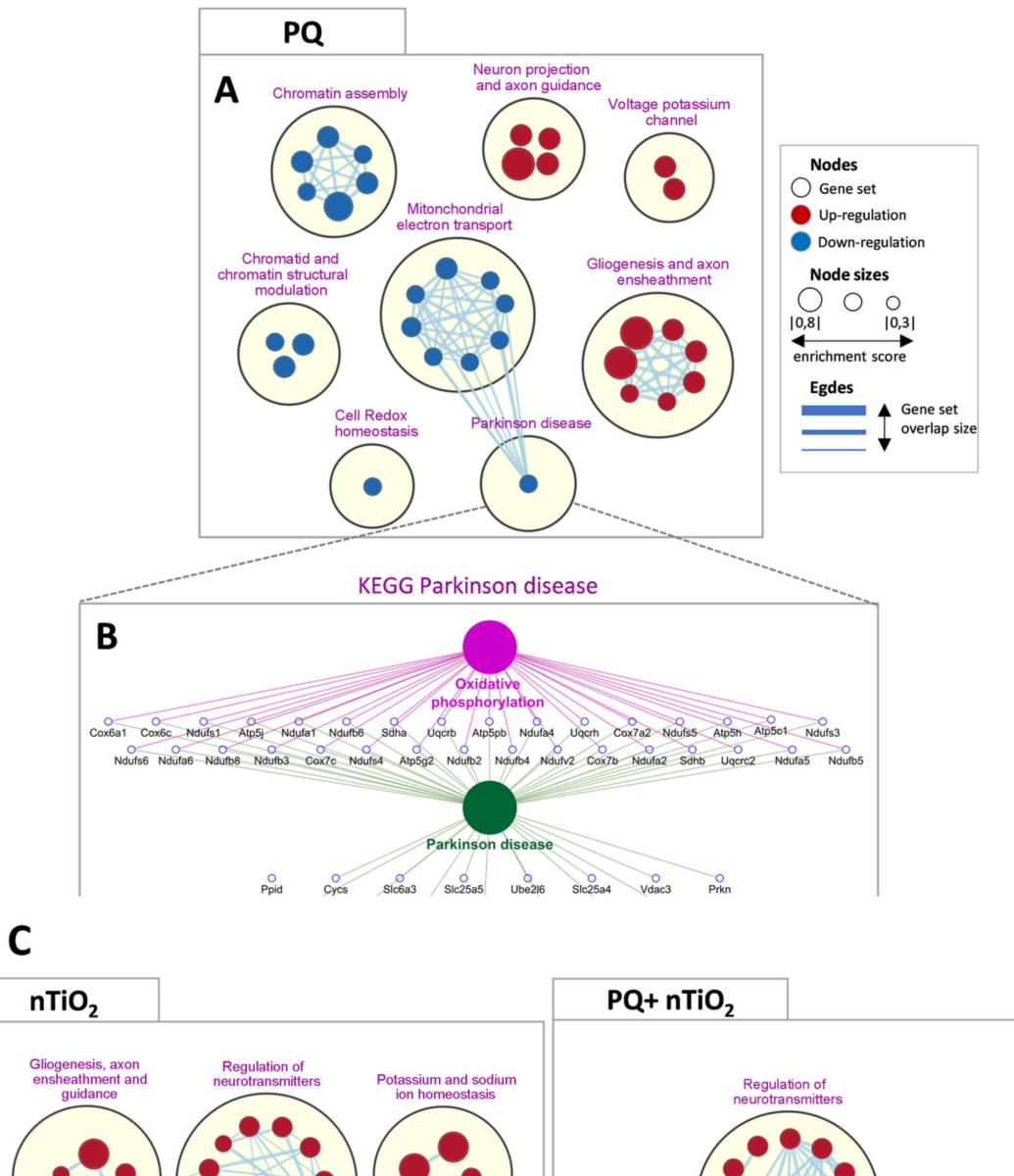


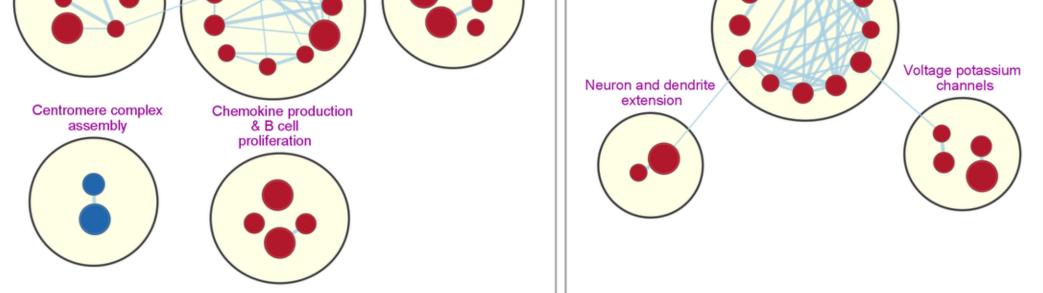
Chga



Kcnc2







GO regulation of neurotransmitter levels

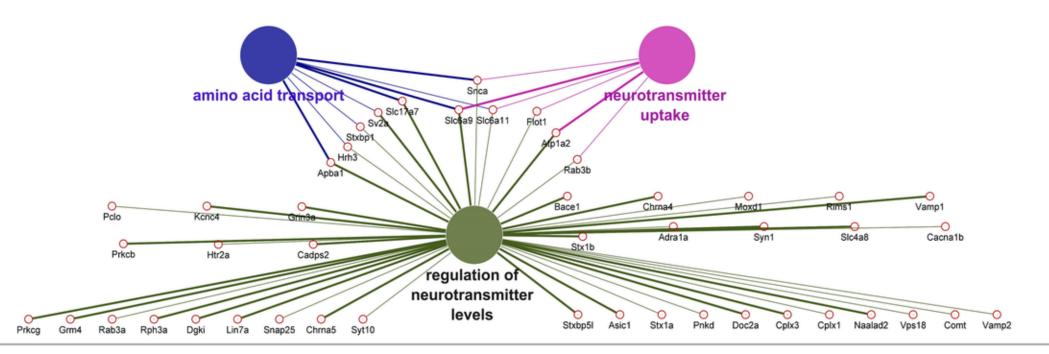


Table 1. Aerosol characteristics

Aerosol	PQ	nTiO ₂	PQ + nTiO ₂
Mass concentration (mg/m ³)	$*[98.7 \pm 3.2] \times 10^{-3}$	10.1 ± 0.6	9.9 ± 0.8
Number concentration [$\times 10^3$ particles/cm ³]	249 ± 17	202 ± 17	308 ± 27
*Particle median diameter (nm)	50 ± 1	220 ± 7	173 ± 2
*Geometric standard deviation	1.74 ± 0.01	2.10 ± 0.03	2.29 ± 0.02

 $\frac{1}{1000} Means \pm SD are shown. * values which were not were measured during the animal exposure sessions but during the prior metrological characterization of the exposure device.$