



HAL
open science

Investigation of Fe isotope systematics for the complete sequence of natural and metallurgical processes of Ni lateritic ores: Implications for environmental source tracing

Gildas Ratié, Jérémie Garnier, Lucieth Cruz Vieira, Daniel Araújo, Michael Komárek, Franck Poitrasson, Cécile Quantin

► To cite this version:

Gildas Ratié, Jérémie Garnier, Lucieth Cruz Vieira, Daniel Araújo, Michael Komárek, et al.. Investigation of Fe isotope systematics for the complete sequence of natural and metallurgical processes of Ni lateritic ores: Implications for environmental source tracing. *Applied Geochemistry*, 2021, 127, pp.104930. 10.1016/j.apgeochem.2021.104930 . hal-03382874

HAL Id: hal-03382874

<https://hal.science/hal-03382874>

Submitted on 18 Oct 2021

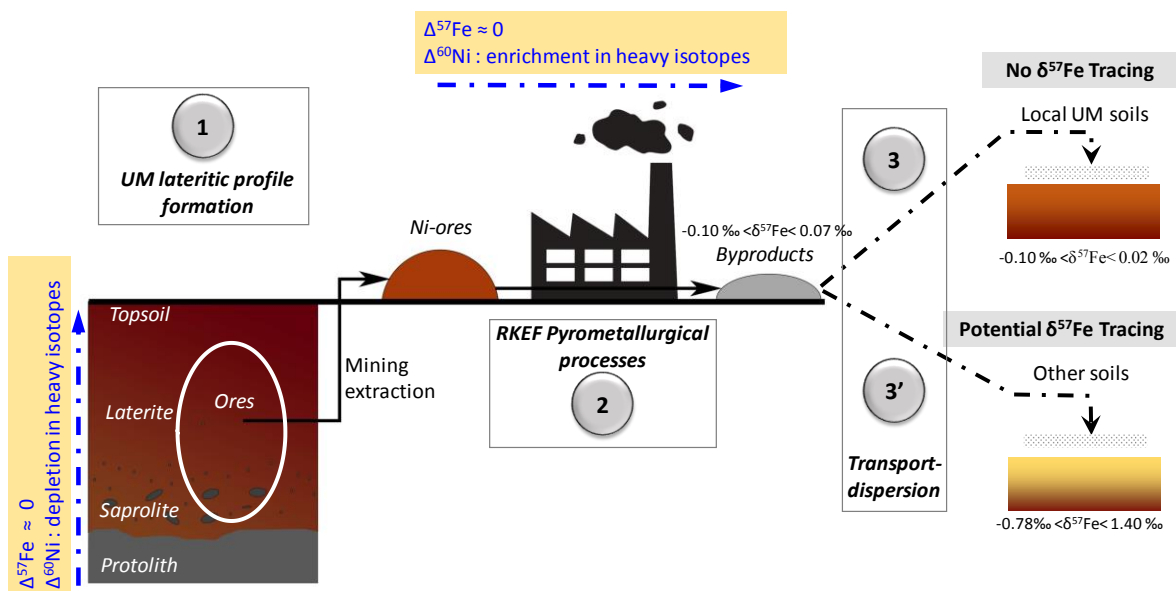
HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

Investigation of Fe isotope systematics for the complete sequence of natural and metallurgical processes of Ni lateritic ores: Implications for environmental source tracing

*Gildas Ratié¹, Jérémie Garnier², Lucieth Cruz Vieira², Daniel F. Araújo³, Michael Komárek¹,
Franck Poitrasson⁴, Cécile Quantin⁵*

Graphical abstract



**Investigation of Fe isotope systematics for the complete sequence
of natural and metallurgical processes of Ni lateritic ores:
Implications for environmental source tracing**

*Gildas Ratié¹, Jérémie Garnier², Lucieth Cruz Vieira², Daniel F. Araújo³, Michael Komárek¹,
Franck Poitrasson⁴, Cécile Quantin⁵*

Highlights

- Fe mobility does not induce isotopic fractionation during chemical weathering.
- There is no evidence of Fe isotope fractionation during the RKEF smelting process.
- Smelting slags are stamped with $\delta^{57}\text{Fe}$ values from Ni laterite ores.
- The use of Fe isotopes as an environmental tracer in lateritic soils is limited.
- Fe isotopes may be potential tracer of mining activities in non-lateritic soils.

1
2
3
4 1 **Investigation of Fe isotope systematics for the complete sequence of**
5
6 2 **natural and metallurgical processes of Ni lateritic ores: Implications**
7
8
9 3 **for environmental source tracing**
10
11
12 4

13
14 5 *Gildas Ratié^{1*}, Jérémie Garnier², Lucieth Cruz Vieira², Daniel F. Araújo³, Michael Komárek¹, Franck*
15 6 *Poitrasson⁴, Cécile Quantin⁵*

17 7 *¹Department of Environmental Geosciences, Faculty of Environmental Sciences, Czech University of Life*
18 8 *Sciences Prague, Kamýcká 129, 165 00, Prague - Suchdol, Czech Republic*

21 9 *²UnB, IG/GMP-ICC Centro, Campus Universitário Darcy Ribeiro, 70910-970, Brasilia-DF,*
22 10 *Brazil/Laboratoire Mixte International “Observatoire des Changements Environnementaux” (LMI OCE),*
23 11 *Institut de Recherche pour le Développement/University of Brasilia, Campus Darcy Ribeiro, Brasilia, Brazil*

26 12 *³Laboratoire de Biogéochimie des Contaminants Métalliques, Ifremer, Centre Atlantique, F44311 Nantes*
27 13 *Cedex 3, France*

30 14 *⁴Laboratoire Géosciences Environnement Toulouse, UMR 5563 CNRS-UPS-IRD-CNES, 14-16, avenue*
31 15 *Edouard Belin, 31400, Toulouse, France.*

33 16 *⁵UMR 8148 GEOPS, Université Paris-Saclay – CNRS, 91405 Cedex, France.*

35 17 **corresponding author: ratie@fzp.czu.cz*

38 18 **Keywords**

- 41 19
- 42 20 • metal cycle
 - 43 21
 - 44 22 • chemical weathering
 - 45 23
 - 46 24 • smelting
 - 47 25
 - 48 26 • byproducts
 - 49 27
- 50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

24 **Abstract**

25 Metal isotopes are versatile pollutant source trackers, but biogeochemical processes can overprint or alter
26 the original source isotopic fingerprint and thus hinder contamination tracing. Here, we explore Fe isotope
27 systematics for the complete range of natural and metallurgical processes related to Ni lateritic ores from
28 Barro Alto, Brazil, to assess its potential as tracer in polluted lateritic soil contexts developed in an
29 ultramafic system.

30 The homogeneous $\delta^{57}\text{Fe}$ values from protolith to soil confirmed that no significant Fe isotopic variation
31 occurred during the formation of the deep lateritic profile. In addition, no Fe isotopic fractionation was
32 found during the smelting process. Although the $\delta^{57}\text{Fe}$ values resulting from mining activities fall within the
33 range of terrestrial sample signatures, the conservation of the $\delta^{57}\text{Fe}$ values from the ores to the byproducts
34 is an advantage for tracing anthropogenic sources when (i) the pyrometallurgical plant uses feeding material
35 with Fe ores imported from other geological formations exhibiting different $\delta^{57}\text{Fe}$ values and/or (ii) the
36 byproducts are transported or dispersed to other locations with different $\delta^{57}\text{Fe}$ signatures in the topsoil.

37 **1. Introduction**

38 The increasing global demand for metals has led to intense mining activities and their consequent
39 remobilization and enrichment in surface compartments, notably soils (Nriagu and Pacyna, 1988; Rauch
40 and Pacyna, 2009). Of particular concern is the soil contamination associated with metal extraction from
41 saprolite ores such as the extraction of Ni, where its concentration of approximately 3 wt% implies high
42 economic value (Butt and Cluzel, 2013). However, in recent decades, limonitic ore refining has increased
43 as a result of the application of modern technologies (allowing better yield recovery). In that ore, Fe contents
44 can reach 40 wt% and Ni can occur in concentrations of up to 1 wt% in Fe oxides (primarily hematite and
45 goethite) (Manceau et al., 2000; Quantin et al., 2002; Dublet et al., 2012 and 2015, Ratié et al., 2018).
46 Smelting processes such as the rotary kiln-electric furnace (RKEF) process are used to extract Ni from
47 mixtures of saprolite and limonite ores, yielding crude ferronickel (FeNi) composed of approximately 70%

1
2
3
4 48 Fe and 30% Ni (Crundwell et al., 2011). In the feeding material, Fe is primarily present as Fe(III) in Fe
5
6 49 oxides and Fe(II)/Fe(III) in the saprolite (e.g., smectite-type, serpentine, pyroxene, spinel, olivine, etc.). The
7
8 50 byproducts of the ore refining process, i.e., fly ash and slag wastes, are Si-, Fe- and Mg-rich materials.

10 51 Generally, the pyrometallurgical wastes are either stored in the surrounding environment in settling
12 52 ponds (fly ash), dumped (slags) or partially reprocessed for metal recovery (fly ash). Such disposal sites are
13 53 susceptible to rainfall leaching and wind remobilization that can lead to the contamination of the superficial
14 54 environment and pose major risks to public health (Ettler et al., 2018). The release of metals in soils by
15 55 leaching is highly time-dependent (Barna et al., 2004; Bril et al., 2008; Seignez et al., 2008; Ettler and Johan,
16 56 2014) and increases when slag disposal sites are flooded and/or occurs in water-saturated environments
17 57 (Ganne et al., 2006; Navarro et al., 2008; Houben et al., 2013). Thus, understanding the dynamics of metals
18 58 at contaminated sites, i.e., their sources, pathways and sinks is of the highest priority to develop effective
19 59 environmental management and monitoring programs.

20 60 To that purpose, the use of metal isotopic signatures can be useful in the identification and
21 61 quantification of contaminant sources and for understanding how biogeochemical processes affect
22 62 contaminant transport (Bullen, 2014; Wiederhold, 2015). The primary challenge to successfully applying
23 63 isotopes as environmental tracers is to identify isotopic signatures that are distinctive between anthropogenic
24 64 and natural materials and to deconvolve the original isotopic signal from subsequent isotopic fractionations
25 65 induced by biogeochemical processes. In the case of stable isotopes of Zn and Cd, industrial or metallurgical
26 66 fractionation during ore refining results in manufactured products and byproducts that are isotopically
27 67 distinct from their natural sources (e.g., Mattielli et al., 2006; Kavner et al., 2008; Sivry et al., 2008; Sonke
28 68 et al., 2008; Shiel et al., 2010; Chrastný et al., 2016; Klein and Rose, 2020). In contrast, Cu and Ni show
29 69 little or no stable isotope fractionation during ore refining by smelting due to their high boiling points
30 70 (Bigalke et al., 2010; Ratié et al., 2016). As a consequence, the manufactured metals, slags and other
31 71 metallurgical byproducts have an isotopic signature similar to that of the ore concentrates. In soil pollution
32 72 contexts, overlaps between ore and natural background isotope compositions may compromise source

1
2
3
4 73 tracking (Ratié et al., 2016; Šillerová et al., 2017). To overcome this drawback, the coupling of two stable
5
6 74 metal isotope systems has been used to enhance source discrimination and deconvolution of the different
7
8 75 biogeochemical processes involved in acid-mine mining (Borrok et al., 2009), coastal systems (Araújo et
9
10 76 al., 2019a, b) and urban atmospheres (Souto-oliveira et al., 2017, 2018).

11
12
13 77 In this work, we explore Fe isotope systematics in lateritic soils from an ultramafic system and the
14
15 78 associated Ni ores refined in a pyrometallurgical system in Barro Alto, Brazil. Previously, an analogous
16
17 79 study was conducted to investigate Ni isotopes, which demonstrated a low level of Ni isotope fractionation
18
19 80 during ore refining that did not allow the use of Ni isotopes as tracers of contamination (Ratié et al., 2016).
20
21 81 Here, we attempt to gain new insights by using Fe isotopes which have never been explored in this context
22
23 82 despite their potential.

24
25
26 83 Iron isotopes demonstrate special features of fractionation, both abiotically and biotically induced
27
28 84 in natural and anthropic materials, that can be useful in our case study (Dauphas et al., 2017; Wu et al.,
29
30 85 2019). As demonstrated by Poitrasson et al. (2008) for chemical weathering in Cameroon, the lateritization
31
32 86 process, which occurs over several million years, result in almost no $\delta^{57}\text{Fe}$ variation. This feature was
33
34 87 subsequently confirmed on other laterites from China and the Philippines, the latter being developed on
35
36 88 peridotites (Liu et al., 2014; Li et al., 2017). In contrast, modern soil studies from both temperate and tropical
37
38 89 areas and even Paleoproterozoic laterites showed much greater Fe isotopic variation (Fante and Depaolo,
39
40 90 2004; Emmanuel et al., 2005; Thompson et al., 2007; Wiederhold et al., 2007; Yamaguchi et al., 2007;
41
42 91 Fekiacova et al., 2013; Akerman et al., 2014). A key driving factor was the separation of two iron pools
43
44 92 having different iron redox states, and therefore contrasted Fe isotope signatures (Wu et al., 2019). Iron
45
46 93 isotope systematics was also successfully used to fingerprint anthropogenic and natural sources in river
47
48 94 sediments (Chen et al., 2014) in an alpine watershed impacted by acid mine drainage (Borrok et al., 2009;
49
50 95 Herbert Jr and Schippers, 2008). Iron isotopes were also used to trace anthropogenic combustion through
51
52 96 the collection of aerosols from sources in the Sahara, North America, Europe (Flament et al., 2008; Conway
53
54 97 et al., 2019) and Japan (Kurisu et al., 2016). These studies suggest that anthropogenic Fe signatures

1
2
3
4 98 originating from metallurgical, industrial and urban activities can display significant differences in Fe
5
6 99 isotopes relative to natural sources. However, the understanding of potential fractionation of Fe in ore
7
8
9 100 smelting/refining remains unclear. Only one study has been conducted to date, which examines Fe
10
11 101 production by an ancient Galo-Roman bloomery process. The study of a major Roman site of Fe production
12
13 102 known as “Les Martyrs” (Montagne Noire Massif, SW France) showed no significant Fe isotope
14
15 103 fractionation from the Roman production of iron bars (Milot et al., 2016). Thus, the present study aims to
16
17 104 explore (1) the Fe isotope fractionation associated with Ni-rich laterite ore formation, (2) the Fe isotope
18
19 105 fractionation associated with Ni laterite ore smelting and refining during the RKEF processing, and (3) the
20
21 106 potential of Fe isotopes to trace the environmental impact of FeNi production.
22
23
24

26 107 **2. Materials and methods**

29 108 **2.1. Ore deposit and mining contexts**

31
32 109 The RKEF process for the production of FeNi was first developed in 1953-1954 and was applied
33
34 110 commercially to the treatment of garnieritic ores in New Caledonia. Later, it was adopted by FeNi producers
35
36 111 for Ni ore deposits across the globe: the Dominican Republic, Colombia, Venezuela, Indonesia, Japan, etc.
37
38 112 (Warner et al., 2006). In recent years, at least three major new FeNi smelters have been constructed and are
39
40 113 in operation: Barro Alto and Onça Puma in Brazil and Koniambo in New Caledonia (Oxley et al., 2016).
41
42 114 The Ni deposits of Barro Alto, located in the midwestern region of Goiás (in Central Brazil), constitute a
43
44 115 large Ni reserve that is exploited by the Anglo American company using open pits.
45
46

47
48 116 The metallurgical plant at Barro Alto uses the RKEF process to produce FeNi from a nominal 2.4
49
50 117 Mt/y of ore. Its production has increased nearly 2-fold since 2011 to 43 kt of total Ni output in 2018 (Anglo
51
52 118 American PLC Annual Report, 2012 and 2018). The deposit, with the ore reserves estimated in 2018 at 52
53
54 119 Mt containing 586 kt of Ni (Anglo American PLC Annual Report, 2018) is in the Barro Alto mafic-
55
56 120 ultramafic complex that is part of the Pre-Cambrian shield. This ultramafic complex is composed of
57
58
59
60
61
62
63
64
65

1
2
3
4 121 serpentinized dunites, pyroxenites and gabbros (Ferreira Filho et al., 2010). The mineralization corresponds
5
6 122 to the surficial weathered portions of the serpentinites (Butt and Cluzel, 2013).

7
8
9 123 Four main steps are involved in FeNi production (Crundwell et al., 2011): drying of the ore before
10
11 124 its introduction into the rotating kiln; calcination with coal, oil or other organic products within the kiln;
12
13 125 reduction in an electric furnace and refining of the molten FeNi in another electric furnace (Fig. 1). These
14
15 126 processes generate enormous quantities of byproducts (Dalvi et al., 2004; Warner et al., 2006) containing
16
17 127 significant amounts of metals (Ni, Co, Cr, Mn, Fe, etc.) (Ettler et al., 2016).

18
19
20 128 The fly ash (F) generated contains large amounts of Fe and Ni and is recovered by electrostatic
21
22 129 filters. The collected fly ash is then recycled into the calcination kiln (Fig. 1). The smelting slags (SS) are
23
24 130 composed of high temperature silicates, amorphous glass as well as inclusions of small FeNi metallic
25
26 131 particles (Ettler et al., 2016). They are dumped and stored near the plant. The molten FeNi is then refined
27
28 132 through a two-step process that produces two types of refining slags: black refining slag (BRS) and white
29
30 133 refining slag (WRS) after the removal of P and S, respectively. The FeNi is produced in the form of small
31
32 134 ingots or water-granulated “beans”.

33 34 35 36 37 135 **2.2. Samples**

38
39
40 136 The list of samples is detailed in Table 1. The sampling for Fe isotope determinations included
41
42 137 natural samples from soils and lateritic profiles in the Barro Alto ultramafic region and materials used and
43
44 138 produced during the RKEF processes. The latter included the Ni ores employed as feeding material (n=2),
45
46 139 the smelting slags (SS, n=2), white and black refining slags (WRS and BRS, n=2) and the final manufactured
47
48 140 FeNi ingot (n=1). As previously mentioned, the fly ash is reinjected in the calcination step.

49
50
51 141 A 28 m deep lateritic profile drilled by the Anglo American company were obtained at intervals of
52
53 142 1 m, labeled “RC”, was used for this study (Ratié et al., 2018). As the overburden (0-3 m) was removed to
54
55 143 facilitate drilling by the mining company, the core/profile starts at the depth of 3 m. To complete the profile,
56
57 144 a soil in the vicinity was collected at three different depths: 0-10 cm (BAS1 0-10) or topsoil, 10-30 cm
58
59 145 (BAS1 10-30) and 30-80 cm (BAS1 30-80) (Ratié et al., 2015). Five lateritic samples of the RC profile (RC

1
2
3
4 146 0-1, RC 6-7, RC 16-17, RC 24-25, and RC 27-28) were selected for Fe isotope characterization. RC 0-1 was
5
6 147 defined as the top of the lateritic profile. RC 6-7 was the part of the lateritic profile dominated by Fe oxides.
7
8 148 RC 16-17 was the smectitic horizon exhibiting high Ni content, and RC 24-25 was a characteristic saprolitic
9
10 149 sample rich in serpentine and exhibiting a relatively low Fe content. RC 27-28 was the deepest sample,
11
12 150 mainly composed of primary minerals (olivine) and was considered the protolith. The sample selection
13
14 151 strategy involved the sampling of a weathering gradient of the ultramafic parent rock.
15
16
17
18

19 152 **2.3. Sample preparation and Fe chemical separation**

20
21 153 All the samples were homogenized and finely crushed, and approximately 100 mg of the samples
22
23 154 was aliquoted to Savillex vessels. The samples were then digested on a hot plate using a multiple-step acid
24
25 155 procedure with HF, HNO₃, and HCl. First, an acid mixture of 5 mL of concentrated HF and 1.5 mL of
26
27 156 HClO₄ at 180°C was added until evaporation was complete. Subsequently, a mixture of concentrated HCl-
28
29 157 HNO₃ (3.75 mL and 1.25 mL, respectively) at 150°C was added and evaporated to dryness. Finally, the
30
31 158 samples were dissolved in an acid medium of 6 M HCl and split into aliquots for elemental and isotopic
32
33 159 determinations. For this step, the sample solution aliquots were processed through chromatographic columns
34
35 160 for chemical separation prior to isotope analysis. The iron was purified using Bio Rad AG1 X4 (200–400
36
37 161 mesh) anionic resin loaded into thermo-retractable Teflon columns for exchange chromatography in an HCl
38
39 162 medium as described by Poitrasson et al. (2004). Blank levels of the chemical procedure reached ~4 ng of
40
41 163 Fe, which is negligible for the sample preparation process. All of the reagents were of analytical grade or
42
43 164 bidistilled and the sample preparation for isotope analysis was conducted in the clean laboratories of GEOPS
44
45 165 (Université Paris Saclay, France).
46
47
48
49
50

51 166 **2.4. Iron isotope composition measurements**

52
53
54 167 Iron isotope measurements were performed at the GET laboratory (Toulouse, France) using the
55
56 168 Observatoire Midi-Pyrénées ICP facility in high or medium mass resolution mode on a Thermo Electron
57
58
59
60
61
62
63
64
65

1
2
3
4 169 Neptune MC ICP MS. The Fe isotopic ratios were determined following the procedure detailed by
5
6 170 Poitrasson and Freydier (2005).

7
8
9 171 This method involved a mass bias correction using a combination of the “standard-sample
10
11 172 bracketing” approach using IRMM-14 as the Fe standard and Ni doping of the purified Fe samples. This
12
13 173 approach accurately corrected for mass bias deviations due to residual matrix effects. The Fe isotope
14
15 174 compositions were expressed in the delta notation relative to the European reference material IRMM-14 as
16
17
18 175 follows:

19
20
21
22 176
$$\delta^{57}\text{Fe} = \left(\frac{\left(\frac{^{57}\text{Fe}}{^{54}\text{Fe}} \right)_{\text{sample}}}{\left(\frac{^{57}\text{Fe}}{^{54}\text{Fe}} \right)_{\text{IRMM-14}}} - 1 \right) \times 1000 \text{ (Eq. 1).}$$

23
24

25
26 177 The GET in-house hematite standard from Milhas (Pyrénées, France) was measured every 6
27
28 178 samples. The long-term external reproducibility of the method was estimated from replicate analyses of this
29
30 179 standard in every session. In this work, the mean $\delta^{57}\text{Fe}$ value of individual measurements for hematite was
31
32 180 $0.762 \pm 0.083 \text{ ‰}$ (2 SD, n=21) in the GET laboratory, whereas data pooled in groups of 3 (which is the
33
34 181 minimum number of times each sample should normally be analyzed) yielded a $\delta^{57}\text{Fe} = 0.764 \pm 0.057 \text{ ‰}$
35
36 182 (2 SD, n=7). These values are consistent with those from previous measurements conducted for over three
37
38 183 years in the same analytical sequences and performed in various laboratories (Poitrasson et al., 2014).
39
40 184 Variation in the $\delta^{57}\text{Fe}$ reported for the samples in this study is expressed as two standard errors (2 SE) of
41
42
43 185 the mean (n=3).
44
45
46

47 186 **3. Results and discussion**

48 49 50 187 **3.1. Bulk compositions**

51
52
53 188 The mineralogy and chemical composition of the entire set of samples are discussed in detail in
54
55 189 Ratié et al. (2015, 2016) and Ettler et al. (2016). The weathered material derived from the ultramafic rocks
56
57
58 190 is strongly depleted in Mg and enriched in Fe from the base to the top of weathering profile. The mineralogy
59
60 191 of the weathered profile changes from the base (RC 27-28) to the top (BAS1 0-10, BAS1 10-30, BAS1 30-

1
2
3
4 192 80) from the dominance of primary minerals (serpentine, chlorite, amphibole, olivine and traces of quartz)
5
6 193 to secondary minerals such as goethite and hematite with some preserved primary minerals such as chromite.
7

8
9 194 Based on the overall sampling from Ratié et al. (2016), the industrial plant feeding material, i.e., the
10
11 195 ore, exhibits high Fe and Mg contents of 118-178 g kg⁻¹ and 81.4-110 g kg⁻¹, respectively, whereas nickel
12
13 196 content ranges from 16.9 to 23.2 g kg⁻¹. Iron and Mg contents of the smelting slags (SS) range from 68.8 to
14
15 197 142.7 g kg⁻¹ and 153 to 188 g kg⁻¹, respectively, and the Ni concentration is relatively low (≤ 2 g kg⁻¹).
16

17
18 198 According to a previous study of these smelting wastes (Ettler et al., 2016), more than 95% of the total Fe
19
20 199 occurs as Fe(II) in the smelting slag, whereas 80% of the total Fe in the reinjected fly ash is present as
21
22 200 Fe(III). The refined slags (WRS and BRS) are richer in Fe (71.2-179 g kg⁻¹) compared with smelting slags.
23
24 201 Ferronickel is composed of roughly two-thirds Fe (66-69 wt%) and one-third Ni (31-34 wt%).
25

26
27 202 The Anglo American plant uses 2.4 Mt/y of Ni ore to produce 41,000 t/y of Ni as FeNi (Moore,
28
29 203 2012 and personal communications). The quantity of Ni introduced in the process, as calculated using an
30
31 204 ore Ni content of 1.96 ± 0.23 wt%, is $47,000 \pm 5,400$ t/y. This led to a production yield of nearly $88 \pm 10\%$
32
33 205 Ni for the 2016 production (Anglo American PLC Annual Report, 2017). For Fe, given the mean Fe content
34
35 206 in ore of 15.1 wt% and a production of 41,000 t of FeNi, 362,400 t of Fe were processed and 82,000 t of Fe
36
37 207 were produced as FeNi with almost 80% of the initial Fe remaining in the waste. The difference between
38
39 208 the incoming Fe/Ni and FeNi production corresponds to the residual Fe/Ni in the different waste materials.
40
41
42
43

44 209 **3.2. Iron isotope compositions**

45 46 47 210 **3.2.1. Ultramafic rocks weathering**

48
49 211 Based on mantle-derived and crustal igneous rocks, the bulk silicate Earth shows a homogenous Fe
50
51 212 isotopic signature of approximately $\delta^{57}\text{Fe} = 0.10 \pm 0.03$ ‰ (Poitrasson and Freydier, 2005; Poitrasson, 2006;
52
53 213 Johnson and Beard, 2006). In Barro Alto, the deepest sample from the profile (RC 27-28), which contains
54
55 214 the typical mineral assembly of serpentinized ultramafic rocks, was determined to be the least weathered
56
57 215 sample and thus it was considered representative of the protolith material (Ratié et al., 2018). The base of
58
59 216 the weathering profile (0.08 ± 0.20 ‰) is consistent with the bulk silicate Earth value (Poitrasson, 2006).
60
61
62
63
64
65

1
2
3
4 217 The $\delta^{57}\text{Fe}$ values of the weathered materials range from $-0.10 \pm 0.07 \text{ ‰}$ (BAS1 0-10) to 0.07 ± 0.05
5
6 218 ‰ (RC 0-1) and fall within the range of values reported in the literature for soils (Wu et al., 2019). Given
7
8 219 the level of analytical variation (2 SE), these results show no significant isotopic differences in $\delta^{57}\text{Fe}$ values
9
10 220 between the protolith and the weathered materials, similarly to other lateritic profiles elsewhere. (Cameroon,
11
12 221 Poitrasson et al., 2008) (China, Liu et al., 2014; Philippines, Li et al., 2017).

15 222 The gain and loss of Fe during chemical weathering can be evaluated by the calculation of the
16
17 223 enrichment factor “ τ_{Fe} ” (Table 1). A negative value for τ_{Fe} reflects a true loss in Fe from the weathered
18
19 224 material compared with the protolith, and a positive value indicates a gain in Fe. If τ_{Fe} is 0, Fe is considered
20
21 225 immobile during weathering with respect to the regolith. The entire Barro Alto profile displays τ_{Fe} values
22
23 226 ranging from -0.10 to 0.22, which suggests that Fe shows little mobility from all of the layers of the profile
24
25 227 (Table 1, Fig. 2). However, a caveat is that this inference does not consider possible soil density changes
26
27 228 that were not measured in this study. The topography of the complex is characterized by a succession of
28
29 229 hills and valleys with altitudes ranging from 750 m to 1100 m dominating the large plain (De Oliveira et al.,
30
31 230 1992). As a consequence, the weathering conditions occurring on the complex are considered as well
32
33 231 drained. In the tropical condition, from the base to the top of the profile, olivine and serpentine are replaced
34
35 232 by Fe-oxides and Mg silicates through a series of transitional phyllosilicates (Colin et al. 1990; Butt and
36
37 233 Cluzel, 2013). In addition, the bulk Fe isotopic composition remains homogeneous along the lateritic profile
38
39 234 (-0.10 ‰ to 0.08 ‰), indicating that the $\delta^{57}\text{Fe}$ values were not significantly altered by the loss or gain of Fe
40
41 235 during chemical weathering (Fig 2). These features agree with the oxidative conditions along the lateritic
42
43 236 profile and the high rate of lixiviation. Therefore, the formation of secondary Fe-bearing phases plays a
44
45 237 minor role in fractionating Fe isotope during ultramafic rock weathering. This result is (i) consistent with
46
47 238 those obtained from weathering profiles in Cameroon (Poitrasson et al., 2008), China (Liu et al., 2014) and
48
49 239 the Philippines (Li et al., 2017) under tropical conditions
50
51
52
53
54
55

56 240 In contrast, in the case of Ni, weathering was associated with isotopic fractionation as part of the Ni
57
58 241 was leached, leading to a weathering profile that was depleted in heavy Ni isotopes. This depletion of heavy
59
60
61
62
63
64
65

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

242 Ni isotopes was interpreted as the preferential sorption and incorporation of light Ni isotopes into Fe oxides
243 (Wasylenki et al., 2015) and phyllosilicates (type 2:1) (Ratié et al., 2018) in addition to Ni isotopic
244 fractionation during the first stage of weathering, i.e., during mineral dissolution (Ratié et al., 2015, 2018).

245 **3.2.2. Metallurgical production**

246 The Fe isotopic composition of the metallurgical samples ranges from -0.10 ± 0.09 ‰ (smelting
247 slags) to 0.07 ± 0.12 ‰ (FeNi). The Fe isotope compositions are analytically indistinguishable from the
248 feeding materials to the final FeNi product (Fig. 3). Although the process yields for Fe are very low (22%),
249 the isotope composition is homogeneous in all byproducts. Given the high-temperature natural processes
250 that occur in the Earth's core and the differentiation in an early silicate magma ocean, it seems logical that
251 there is not significant Fe isotope fractionation in the FeNi alloy and the ultramafic silicate melt (Poitrasson
252 et al., 2009). Within the 2-7.7 GPa pressures, the chemical and Fe isotope equilibrium was reached at
253 2,000°C within 100 s ($\Delta^{57}\text{Fe}_{\text{metal-silicate glass}} = 0.047 \pm 0.063$ ‰). The high temperature conditions found in the
254 electric furnace at 1,600°C could induce a similar rapid equilibrium and hence inhibit detectable Fe isotopic
255 fractionation between FeNi and the feeding material.

256 In the furnace, metal isotope fractionation is dependent on the relative isotope mass difference, the
257 viscosity of the alloy, the mass of the matrix atoms and the temperature range (Ott, 1969; Lodding et al.,
258 1970; Ginoza and March, 1985). In the RKEF processes, the smelting temperature (1,600°C) is very close
259 to the Fe fusion point (1,538°C), and the homogeneous $\delta^{57}\text{Fe}$ value in the metallurgical wastes argues for
260 an absence of the thermal gradients responsible for possible metal stable isotope fractionation. In fact, in
261 modern enhanced industrial processes such as RKEF, the Fe distribution is homogeneous at the molten scale
262 during the different steps. This inference is supported by experiments reproducing the ancient bloomery
263 process at 1,300°C, which shows no significant Fe isotopic heterogeneity within the Fe metal products,
264 although they did not go beyond the pasty state at such low temperatures (Milot et al., 2016).

3.2.3. Implications for environmental tracing

Our study shows that the anthropogenic $\delta^{57}\text{Fe}$ values fall within the range of the raw material signatures in Barro Alto that include the Earth's mean Fe isotopic composition, in a manner similar to that of the Ni isotope systematics previously investigated (Fig. 3). Such an outcome hinders the use of Fe isotopes as an environmental tracer in the context of soils impacted by metallurgical activity if the raw material comes from the same locality, which is the case at Barro Alto.

However, the conservation of the $\delta^{57}\text{Fe}$ values from the feeding material to the metallurgical wastes can be advantageous to trace anthropogenic sources in cases (i) where pyrometallurgical plants use feeding material imported from another deposit exhibiting $\delta^{57}\text{Fe}$ values that are distinct from local environment and when (ii) the metallurgical byproducts deposited in open-air undergo redox reactions triggered by changes in the biogeochemical conditions of the surrounding environment. A compilation of published $\delta^{57}\text{Fe}$ of ores (Milot et al., 2016, 2018) have noted that the iron isotopic signature exhibits a wide range of values from -2.8 ‰ to 2.4 ‰ for different mineral deposits (sedimentary, hydrothermal, skarn and supergene deposits (e.g., Graham et al., 2004; Markl et al., 2006; Johnson et al., 2008; Fabre et al., 2011; Wang et al., 2011; Cheng et al., 2015; Pi et al., 2015; Wawryk and Foden, 2015; Texeira et al., 2017). Moreover, modern soils from temperate areas (e.g., Fantle and Depaolo, 2004; Emmanuel et al., 2005; Wiederhold et al., 2007; Fekiacova et al., 2013, Wu et al., 2019) and wet tropical soils involving Fe redox cycling (Thompson et al., 2007; Akerman et al., 2014) yield $\delta^{57}\text{Fe}$ ranges of -0.9 ‰ to 1.4 ‰. This range represents significant variations relative to the natural $\delta^{57}\text{Fe}$ values in Barro Alto soils, which range from -0.10 ‰ to 0.02 ‰ only. This therefore permits the use Fe isotopes to trace contamination when raw materials from other locations are employed, as long as they are not regular lateritic soils. Therefore, the use of a different feeding material for the industrial plant at the Barro Alto site would allow tracing of the anthropogenic input to the local environment. In this case, Fe isotopes could be a better tracer than Ni isotopes that showed isotope fractionation during the smelting process (Ratié et al., 2016).

Finally, a review published by Warner et al. (2006) has shown similar concentration results during different steps of the Ni RKEF smelting processes throughout the world (the Dominican Republic,

1
2
3
4 291 Colombia, Venezuela, Brazil, Japan, New Caledonia, Indonesia, Ukraine, Macedonia and Greece). The
5
6 292 feeding material from Barro Alto exhibits means concentration values of 2 wt% of Ni and 15.3 wt% of Fe
7
8 293 ($n = 13$, Ratié et al., 2016), whereas the global average is 1.9 ± 0.5 wt% and 17 ± 5 wt%, respectively. The
9
10 294 total average Fe content in slag material is 10 wt% at Barro Alto, whereas in the global mean value is $15 \pm$
11
12 295 10 wt%. The feeding material composition and the main wastes are therefore similar for the Barro Alto
13
14 296 smelter and the global laterite Ni smelters. Moreover, the calcination, smelting and refining temperatures
15
16 297 used are similar for the Barro Alto plant (850°C , $1,600^\circ\text{C}$ and $1,550^\circ\text{C}$, respectively) and the other cited
17
18 298 RKEF smelters ($880 \pm 120^\circ\text{C}$, $n=13$; $1,570 \pm 35^\circ\text{C}$, $n = 12$; $1,440 \pm 120^\circ\text{C}$, $n = 10$, respectively). This
19
20 299 comparison suggests that the behavior of the Fe isotope system during pyrometallurgical processing at the
21
22 300 Barro Alto plant is likely applicable to other plants elsewhere in the world.
23
24
25
26
27

28 301 **4. Conclusions**

29
30
31 302 For the first time, this study shows the Fe isotope composition for the complete series of natural and
32
33 303 anthropogenic processes in the Barro Alto ultramafic complex. No significant Fe isotope variations were
34
35 304 identified in either the pedogenesis of lateritic soils or the pyrometallurgical processes of Ni ore refining.
36
37 305 The $\delta^{57}\text{Fe}$ value of the protolith fell within the range of the Fe isotope composition of the bulk silicate Earth,
38
39 306 which is estimated at approximately 0.1 ‰ (Poitrasson, 2006), and is similar to deep lateritic profiles studied
40
41 307 elsewhere. In the Ni laterite RKEF smelting process at Barro Alto, the rapid equilibrium of Fe isotopes
42
43 308 between the different phases composing the FeNi ore melt at $1,600^\circ\text{C}$ in the electric furnace results in
44
45 309 undetectable Fe isotopic fractionation. This mechanism is supported by laboratory experiments under
46
47 310 controlled conditions with high pressure and temperature equilibration between FeNi alloys and ultramafic
48
49 311 silicate melts (Poitrasson et al., 2009).
50
51
52
53

54 312 As a consequence, the $\delta^{57}\text{Fe}$ values obtained from both pedogenesis (protolith to topsoil) and
55
56 313 pyrometallurgical samples are homogeneous and may challenge the discrimination of anthropic and natural
57
58 314 sources. Further studies should verify the potential fractionation induced by postdepositional processes as
59
60
61
62
63
64
65

1
2
3
4 315 demonstrated for Ni (Ratié et al., 2016) and Zn (Yin et al., 2018). Nonetheless, the conservation of the $\delta^{57}\text{Fe}$
5
6 316 values from the ores to the byproducts is an advantage for tracing anthropogenic sources when (i) the
7
8 317 pyrometallurgical plant uses feeding material with Fe ores imported from other geological unities exhibiting
9
10 318 different $\delta^{57}\text{Fe}$ values and/or (ii) the byproducts are transported or dispersed to other locations with different
11
12 319 $\delta^{57}\text{Fe}$ values in the topsoil. In these cases, Fe isotopes could be a more suitable environmental tracer of
13
14 320 anthropogenic sources than Ni isotopes.
15
16
17
18

19 321 **Acknowledgments**

22 322 We express our sincere thanks to the Anglo American staff at the Barro Alto mine (Goiás, Brazil).
23
24 323 The authors acknowledge Gael Monvoisin for maintaining the GEOPS clean laboratory facilities and J.
25
26 324 Chmeleff from the GET laboratory (Toulouse, France) for maintaining the MC ICP MS in good working
27
28 325 order. The analytical work in Toulouse along with travel by JG and LCV was partly funded by the LMI-
29
30 326 OCE and the 7th European Community Framework Program (NIDYFICS, n°318123) and the CLIM-
31
32 327 AMAZON laboratory (INCO-LAB grant n°295091). The analytical work in Brazil and France was partly
33
34 328 funded by the LMI-OCE and CNPq 445423/2014-5, 400029/2015-4 and 420697/2018-7. Gildas
35
36 329 acknowledges the Faculty of Environmental Sciences, Czech University of Life Sciences, Prague (Grant n°
37
38 330 42900/1312/3166), the Czech Science Foundation 18-07585S, the French Ministry of National Education
39
40 331 and Research (G. Ratié PhD grant), CAPES and the CNPq for support during his PhD. Jérémie Garnier was
41
42 332 supported by CNPq grant 302722/2018-1. Daniel F. Araújo thanks the POLLUSOLS project framework
43
44 333 and Ifremer (Centre Atlantique) for the financial support.
45
46
47
48
49

50 334 **References**

51
52
53 335 Anglo American PLC Annual Report, 2012, 2012.
54
55 336 Anglo American PLC Annual Report, 2017, 2017.
56
57 337 Anglo American PLC Annual Report, 2018, 2018.
58 338 Akerman, A., Poitrasson, F., Oliva, P., Audry, S., Prunier, J., Braun, J.J., 2014. Isotopic fingerprint of Fe
59 339 cycling in an equatorial soil-plant-water system: The Nsimi watershed, South Cameroon.
60 340 Chemical Geology, 385: 104-116.
61
62
63
64
65

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

341 Araújo, D.F., Ponzevera, E., Briant, N., Knoery, J., Sireau, T., Mojtahid, M., Metzger, E., Brach-Papa, C.,
342 2019a. Assessment of the metal contamination evolution in the Loire estuary using Cu and Zn
343 stable isotopes and geochemical data in sediments. *Mar. Pollut. Bull.* 143, 12–23.

344 Araújo, D. F., Ponzevera, E., Briant, N., Knoery, J., Bruzac, S., Sireau, T., Brach-Papa, C., 2019b. Copper,
345 zinc and lead isotope signatures of sediments from a mediterranean coastal bay impacted by
346 naval activities and urban sources. *Applied Geochemistry*, 111, 104440.

347 Barna, R., Moszkowicz, P., Gervais, C., 2004. Leaching assessment of road materials containing primary
348 lead and zinc slags. *Waste Manag.* 24, 945–955.

349 Beard, B.L., Handler, R.M., Scherer, M.M., Wu, L., Czaja, A.D., Heimann, A., Johnson, C.M., 2010. Iron
350 isotope fractionation between aqueous ferrous iron and goethite. *Earth Planet. Sci. Lett.* 295,
351 241–250.

352 Bigalke, M., Weyer, S., Kobza, J., & Wilcke, W., 2010. Stable Cu and Zn isotope ratios as tracers of sources
353 and transport of Cu and Zn in contaminated soil. *Geochimica et Cosmochimica Acta*, 74(23),
354 6801-6813.

355 Borrok, D.M., Wanty, R.B., Ridley, W.I., Lamothe, P.J., Kimball, B.A., Verplanck, P.L., Runkel, R.L.,
356 2009. Application of iron and zinc isotopes to track the sources and mechanisms of metal loading
357 in a mountain watershed. *Appl. Geochem.* 24, 1270–1277.
358 <https://doi.org/10.1016/j.apgeochem.2009.03.010>

359 Bril, H., Zainoun, K., Puziewicz, J., Courtin-Nomade, A., Vanaecker, M., Bollinger, J.-C., 2008. Secondary
360 phases from the alteration of a pile of zinc-smelting slag as indicators of environmental
361 conditions: An example from Swietochłowice, Upper Silesia, Poland. *Can. Mineral.* 46, 1235–
362 1248.

363 Bullen, T.D., 2014. 7.10 - Metal Stable Isotopes in Weathering and Hydrology, in: Holland, H.D., Turekian,
364 K.K. (Eds.), *Treatise on Geochemistry (Second Edition)*. Elsevier, Oxford, pp. 329–359.

365 Butt, C.R., Cluzel, D., 2013. Nickel laterite ore deposits: weathered serpentinites. *Elements* 9, 123–128.

366 Chen, J. B., Busigny, V., Gaillardet, J., Louvat, P., & Wang, Y. N. (2014). Iron isotopes in the Seine River
367 (France): Natural versus anthropogenic sources. *Geochimica et Cosmochimica Acta*, 128, 128-
368 143.

369 Cheng, Y.B., Mao, J.W., Zhu, X.K., Wang, Y., 2015. Iron isotope fractionation during supergene weathering
370 process and its application to constrain ore genesis in Gaosong deposit, Gejiu district, SW China.
371 *Gondwana Research*, 27(3): 1283-1291.

372 Chrastný, V., Čadková, E., Vaněk, A., Teper, L., Cabala, J., Komárek, M., 2015. Cadmium isotope
373 fractionation within the soil profile complicates source identification in relation to Pb–Zn mining
374 and smelting processes. *Chem. Geol.* 405, 1–9. <https://doi.org/10.1016/j.chemgeo.2015.04.002>

375 Colin, F., Nahon, D., Trescases, J. J., & Melfi, A. J. (1990). Lateritic weathering of pyroxenites at
376 Niquelandia, Goiás, Brazil; the supergene behavior of nickel. *Economic Geology*, 85(5), 1010-
377 1023.

378 Conway, T.M., Hamilton, D.S., Shelley, R.U., Aguilar-Islas, A.M., Landing, W.M., Mahowald, N.M., John,
379 S.G., 2019. Tracing and constraining anthropogenic aerosol iron fluxes to the North Atlantic
380 Ocean using iron isotopes. *Nat. Commun.* 10, 1–10.

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

381 Crundwell, F.K., Moats, M.S., Ramachandran, V., Robinson, T.G., Davenport, W.G., 2011. Chapter 7 -
382 Refining Molten Ferronickel, in: Davenport, F.K.C.S.M.R.G.R.G. (Ed.), *Extractive Metallurgy*
383 of Nickel, Cobalt and Platinum Group Metals. Elsevier, Oxford, pp. 85–93.

384 Dalvi, A.D., Bacon, W.G., Osborne, R.C., 2004. Past and future of nickel laterite projects.

385 Dauphas, N., John, S.G., Rouxel, O., 2017. Iron isotope systematics. *Rev. Mineral. Geochem.* 82, 415–510.

386 De Oliveira, S. B., Trescases, J. J., & Melfi, A. J. (1992). Lateritic nickel deposits of Brazil. *Mineralium*
387 *deposita*, 27(2), 137-146.

388 Dublet, G., Juillot, F., Morin, G., Fritsch, E., Fandeur, D., Brown, G.E., 2015. Goethite aging explains Ni
389 depletion in upper units of ultramafic lateritic ores from New Caledonia. *Geochim. Cosmochim.*
390 *Acta* 160, 1–15. <https://doi.org/10.1016/j.gca.2015.03.015>

391 Dublet, G., Juillot, F., Morin, G., Fritsch, E., Fandeur, D., Ona-Nguema, G., Brown, G.E., 2012. Ni
392 speciation in a New Caledonian lateritic regolith: A quantitative X-ray absorption spectroscopy
393 investigation. *Geochim. Cosmochim. Acta* 95, 119–133.

394 Emmanuel, S., Erel, Y., Matthews, A., Teutsch, N., 2005. A preliminary mixing model for Fe isotopes in
395 soils. *Chem. Geol.* 222, 23–34.

396 Ettler, V., Johan, Z., 2014. 12 years of leaching of contaminants from Pb smelter slags:
397 Geochemical/mineralogical controls and slag recycling potential. *Appl. Geochem.* 40, 97–103.

398 Ettler, V., Kvapil, J., Šebek, O., Johan, Z., Mihaljevič, M., Ratié, G., Garnier, J., Quantin, C., 2016. Leaching
399 behaviour of slag and fly ash from laterite nickel ore smelting (Niquelândia, Brazil). *Appl.*
400 *Geochem.* 64, 118–127.

401 Ettler, V., Polák, L., Mihaljevič, M., Ratié, G., Garnier, J., Quantin, C., 2018. Oral bioaccessibility of
402 inorganic contaminants in waste dusts generated by laterite Ni ore smelting. *Environ. Geochem.*
403 *Health* 40, 1699–1712.

404 Fabre, S., Nedelec, A., Poitrasson, F., Strauss, H., Thomazo, C., Nogueira, A., 2011. Iron and sulphur
405 isotopes from the Carajas mining province (Para, Brazil): Implications for the oxidation of the
406 ocean and the atmosphere across the Archaean-Proterozoic transition. *Chemical Geology*, 289(1-
407 2): 124-139.

408 Fantle, M.S., DePaolo, D.J., 2004. Iron isotopic fractionation during continental weathering. *Earth Planet.*
409 *Sci. Lett.* 228, 547–562.

410 Fekiacova, Z., Pichat, S., Cornu, S., Balesdent, J., 2013. Inferences from the vertical distribution of Fe
411 isotopic compositions on pedogenetic processes in soils. *Geoderma*, 209: 110-118.

412 Ferreira Filho, C.F., Pimentel, M.M., de Araujo, S.M., Laux, J.H., 2010. Layered intrusions and volcanic
413 sequences in Central Brazil: geological and geochronological constraints for Mesoproterozoic
414 (1.25 Ga) and Neoproterozoic (0.79 Ga) igneous associations. *Precambrian Res.* 183, 617–634.

415 Flament, P., Mattielli, N., Aimo, L., Choel, M., Deboudt, K., de Jong, J., Rimetz-Planchon, J., Weis, D.,
416 2008. Iron isotopic fractionation in industrial emissions and urban aerosols. *Chemosphere* 73,
417 1793–1798. <https://doi.org/10.1016/j.chemosphere.2008.08.042>

418 Gall, L., Williams, H.M., Halliday, A.N., Kerr, A.C., 2017. Nickel isotopic composition of the mantle.
419 *Geochim. Cosmochim. Acta* 199, 196–209. <https://doi.org/10.1016/j.gca.2016.11.016>

420 Ganne, P., Cappuyns, V., Vervoort, A., Buvé, L., Swennen, R., 2006. Leachability of heavy metals and
421 arsenic from slags of metal extraction industry at Angleur (eastern Belgium). *Sci. Total Environ.*
422 356, 69–85. <https://doi.org/10.1016/j.scitotenv.2005.03.022>

- 1
2
3
4 423 Ginoza, M., March, N.H., 1985. Isotope Effects in Liquid Metals. *Phys. Chem. Liq.* 15, 75–111.
5 424 <https://doi.org/10.1080/00319108508078471>
6
7 425 Graham, S., Pearson, N., Jackson, S., Griffin, W., O'Reilly, S.Y., 2004. Tracing Cu and Fe from source to
8 426 porphyry: in situ determination of Cu and Fe isotope ratios in sulfides from the Gratsberg Cu-
9 427 Au deposit. *Chemical Geology*, 207: 147-169.
10
11 428 Herbert Jr, R.B., Schippers, A., 2008. Iron isotope fractionation by biogeochemical processes in mine
12 429 tailings. *Environ. Sci. Technol.* 42, 1117–1122.
13
14 430 Houben, D., Couder, E., Sonnet, P., 2013. Leachability of cadmium, lead, and zinc in a long-term
15 431 spontaneously revegetated slag heap: implications for phytostabilization. *J. Soils Sediments* 13,
16 432 543–554.
17
18 433 Johnson, C.M., Beard, B.L., 2006. Fe isotopes: an emerging technique for understanding modern and
19 434 ancient biogeochemical cycles. *GSA TODAY* 16, 4.
20
21 435 Johnson, C. M., Beard, B. L., & Roden, E. E., 2008. The iron isotope fingerprints of redox and
22 436 biogeochemical cycling in modern and ancient Earth. *Annu. Rev. Earth Planet. Sci.*, 36, 457-
23 437 493.
24
25 438 Kavner, A., John, S.G., Sass, S., Boyle, E.A., 2008. Redox-driven stable isotope fractionation in transition
26 439 metals: Application to Zn electroplating. *Geochim. Cosmochim. Acta* 72, 1731–1741.
27
28 440 Klein, S., & Rose, T. (2020). Evaluating copper isotope fractionation in the metallurgical operational chain:
29 441 An experimental approach. *Archaeometry*.
30
31 442 Kurisu, M., Takahashi, Y., Iizuka, T., Uematsu, M., 2016. Very low isotope ratio of iron in fine aerosols
32 443 related to its contribution to the surface ocean. *J. Geophys. Res. Atmospheres* 121, 11,119-
33 444 11,136.
34
35 445 Li, M., He, Y.S., Kang, J.T., Yang, X.Y., He, Z.W., Yu, H.M., Huang, F., 2017. Why was iron lost without
36 446 significant isotope fractionation during the lateritic process in tropical environments? *Geoderma*,
37 447 290: 1-9.
38
39 448 Liu, S.A., Teng, F.Z., Li, S.G., Wei, G.J., Ma, J.L., Li, D.D., 2014. Copper and iron isotope fractionation
40 449 during weathering and pedogenesis: Insights from saprolite profiles. *Geochimica Et*
41 450 *Cosmochimica Acta*, 146: 59-75.
42
43 451 Lodding, A., Mundy, J.N., Ott, A., 1970. Isotope Inter-Diffusion and Self-Diffusion in Solid Lithium Metal.
44 452 *Phys. Status Solidi B* 38, 559–569. <https://doi.org/10.1002/pssb.19700380206>
45
46 453 Manceau, A., Schlegel, M.L., Musso, M., Sole, V.A., Gauthier, C., Petit, P.E., Trolard, F., 2000. Crystal
47 454 chemistry of trace elements in natural and synthetic goethite. *Geochim. Cosmochim. Acta* 64,
48 455 3643–3661.
49
50 456 Markl, G., von Blanckenburg, F., Wagner, T., 2006. Iron isotope fractionation during hydrothermal ore
51 457 deposition and alteration. *Geochimica Et Cosmochimica Acta*, 70(12): 3011-3030.
52
53 458 Mattielli, N., Rimetz, J., Petit, J., Perdrix, E., Deboudt, K., Flament, P., Weis, D., 2006. Zn Cu isotopic
54 459 study and speciation of airborne metal particles within a 5-km zone of a lead/zinc smelter.
55 460 *Geochim. Cosmochim. Acta Suppl.* 70, A401–A401.
56
57 461 Milot, J., Poitrasson, F., Baron, S., Coustures, M.-P., 2016. Iron isotopes as a potential tool for ancient iron
58 462 metals tracing. *J. Archaeol. Sci.* 76, 9–20.
59
60 463 Milot, J., Baron, S., Poitrasson, F., 2018. Potential use of Fe isotopes for ancient non-ferrous metals tracing
61 464 through the example of a lead-silver production site (Imiter miner, Anti-Atlas, Morocco). *Journal*
62 465 *of Archaeological Science*, 98: 22-33.
63
64
65

- 1
2
3
4 466 Moore, P. (2012). Anglo's new nickel. *International Mining*, March, 20-24.
- 5
6 467 Navarro, A., Cardellach, E., Mendoza, J.L., Corbella, M., Domènech, L.M., 2008. Metal mobilization from
7 468 base-metal smelting slag dumps in Sierra Almagrera (Almería, Spain). *Appl. Geochem.* 23, 895–
8 469 913. <https://doi.org/10.1016/j.apgeochem.2007.07.012>
- 9
10 470 Nriagu, J., Pacyna, J., 1988. Quantitative assesment of worldwide contamination of air, water and soils by
11 471 trace-metals. *Nature* 333, 134–139. <https://doi.org/10.1038/333134a0>
- 12 472 Ott, A., 1969. Isotope Separation by Thermal Diffusion in Liquid Metal. *Science* 164, 297–297.
13 473 <https://doi.org/10.1126/science.164.3877.297>
- 14
15 474 Oxley, A., Smith, M.E., Caceres, O., 2016. Why heap leach nickel laterites? *Miner. Eng.* 88, 53–60.
- 16 475 Pi, Q.H., Zhong, R.C., Hu, R.Z., 2015. Tracing the ore-formation history of the shear-zone-controlled
17 476 Huogeqi Cu-Pb-Zn deposit in Inner Mongolia, northern China, using H, O, S, and Fe isotopes.
18 477 *Ore Geology Reviews*, 71: 263-272.
- 19
20 478 Platzner, I. T., 1997. *Modern isotope ratio mass spectrometry*. Wiley, Chichester, p 514.
- 21 479 Poitrasson, F., 2006. On the iron isotope homogeneity level of the continental crust. *Chem. Geol.* 235, 195–
22 480 200. <https://doi.org/10.1016/j.chemgeo.2006.06.010>
- 23
24 481 Poitrasson, F., Freyrier, R., 2005. Heavy iron isotope composition of granites determined by high resolution
25 482 MC-ICP-MS. *Chem. Geol.* 222, 132–147. <https://doi.org/10.1016/j.chemgeo.2005.07.005>
- 26
27 483 Poitrasson, F., Roskosz, M., Corgne, A., 2009. No iron isotope fractionation between molten alloys and
28 484 silicate melt to 2000 °C and 7.7 GPa: Experimental evidence and implications for planetary
29 485 differentiation and accretion. *Earth Planet. Sci. Lett.* 278, 376–385.
30 486 <https://doi.org/10.1016/j.epsl.2008.12.025>
- 31
32 487 Poitrasson, F., Vieira, L.C., Seyler, P., dos Santos Pinheiro, G.M., Mulholland, D.S., Bonnet, M.-P.,
33 488 Martinez, J.-M., Lima, B.A., Boaventura, G.R., Chmeleff, J., 2014. Iron isotope composition of
34 489 the bulk waters and sediments from the Amazon River Basin. *Chem. Geol.* 377, 1–11.
- 35
36 490 Poitrasson, F., Viers, J., Martin, F., Braun, J.-J., 2008. Limited iron isotope variations in recent lateritic soils
37 491 from Nsimi, Cameroon: implications for the global Fe geochemical cycle. *Chem. Geol.* 253, 54–
38 492 63.
- 39
40 493 Quantin, C., Becquer, T., Rouiller, J.H., Berthelin, J., 2002. Redistribution of metals in a New Caledonia
41 494 ferralsol after microbial weathering. *Soil Sci. Soc. Am. J.* 66, 1797–1804.
- 42
43 495 Ratié, G., Garnier, J., Calmels, D., Vantelon, D., Guimarães, E., Monvoisin, G., Nouet, J., Ponzevera, E.,
44 496 Quantin, C., 2018. Nickel distribution and isotopic fractionation in a Brazilian lateritic regolith:
45 497 Coupling Ni isotopes and Ni K-edge XANES. *Geochim. Cosmochim. Acta* 230, 137–154.
- 46
47 498 Ratié, G., Jouvin, D., Garnier, J., Rouxel, O., Miska, S., Guimarães, E., Cruz Vieira, L., Sivry, Y., Zelano,
48 499 I., Montarges-Pelletier, E., Thil, F., Quantin, C., 2015. Nickel isotope fractionation during
49 500 tropical weathering of ultramafic rocks. *Chem. Geol.* 402, 68–76.
50 501 <https://doi.org/10.1016/j.chemgeo.2015.02.039>
- 51
52 502 Ratié, G., Quantin, C., Jouvin, D., Calmels, D., Ettler, V., Sivry, Y., Vieira, L.C., Ponzevera, E., Garnier,
53 503 J., 2016. Nickel isotope fractionation during laterite Ni ore smelting and refining: Implications
54 504 for tracing the sources of Ni in smelter-affected soils. *Appl. Geochem.* 64, 136–145.
- 55
56 505 Rauch, J.N., Pacyna, J.M., 2009. Earth's global Ag, Al, Cr, Cu, Fe, Ni, Pb, and Zn cycles. *Glob.*
57 506 *Biogeochem. Cycles* 23, GB2001. <https://doi.org/10.1029/2008GB003376>
- 58
59
60
61
62
63
64
65

- 1
2
3
4 507 Seignez, N., Gauthier, A., Bulteel, D., Damidot, D., Potdevin, J.-L., 2008. Leaching of lead metallurgical
5 508 slags and pollutant mobility far from equilibrium conditions. *Appl. Geochem.* 23, 3699–3711.
6 509 <https://doi.org/10.1016/j.apgeochem.2008.09.009>
- 8 510 Shiel, A.E., Weis, D., Orians, K.J., 2010. Evaluation of zinc, cadmium and lead isotope fractionation during
9 511 smelting and refining. *Sci. Total Environ.* 408, 2357–2368.
10 512 <https://doi.org/10.1016/j.scitotenv.2010.02.016>
- 12 513 Šillerová, H., Chrastný, V., Vítková, M., Francová, A., Jehlička, J., Gutsch, M.R., Kocourková, J., Aspholm,
13 514 P.E., Nilsson, L.O., Berglen, T.F., 2017. Stable isotope tracing of Ni and Cu pollution in North-
14 515 East Norway: Potentials and drawbacks. *Environ. Pollut.* 228, 149–157.
- 16 516 Sivry, Y., Riotte, J., Sonke, J.E., Audry, S., Schäfer, J., Viers, J., Blanc, G., Freydier, R., Dupré, B., 2008.
17 517 Zn isotopes as tracers of anthropogenic pollution from Zn-ore smelters The Riou Mort–Lot River
18 518 system. *Chem. Geol.* 255, 295–304.
- 20 519 Sonke, J.E., Sivry, Y., Viers, J., Freydier, R., Dejonghe, L., Andre, L., Aggarwal, J.K., Fontan, F., Dupre,
21 520 B., 2008. Historical variations in the isotopic composition of atmospheric zinc deposition from
22 521 a zinc smelter. *Chem. Geol.* 252, 145–157. <https://doi.org/10.1016/j.chemgeo.2008.02.006>
- 24 522 Souto-Oliveira, C.E., Babinski, M., Araújo, D.F., Andrade, M.F., 2018. Multi-isotopic fingerprints (Pb, Zn,
25 523 Cu) applied for urban aerosol source apportionment and discrimination. *Sci. Total Environ.* 626,
26 524 1350–1366.
- 28 525 Souto-Oliveira, C.E., Babinski, M., Araújo, D.F., Weiss, D.J., Ruiz, I.R., 2019. Multi-isotope approach of
29 526 Pb, Cu and Zn in urban aerosols and anthropogenic sources improves tracing of the atmospheric
30 527 pollutant sources in megacities. *Atmos. Environ.* 198, 427–437.
- 32 528 Teixeira, N.L., Caxito, F.A., Rosière, C.A., Pecoits, E., Vieira, L.C., Frei, R., Sial, A.N., Poitrasson, F.,
33 529 2017. Trace elements and isotope geochemistry (C, O, Fe, Cr) of the Cauê iron formation,
34 530 Quadrilatero Ferrífero, Brazil: Evidence for widespread microbial dissimilatory iron reduction
35 531 at the Archean/Proterozoic transition. *Precambrian Research*, 298: 39-55.
- 37 532 Thompson, A., Ruiz, J., Chadwick, O.A., Titus, M., Chorover, J., 2007. Rayleigh fractionation of iron
38 533 isotopes during pedogenesis along a climate sequence of Hawaiian basalt. *Chem. Geol.* 238, 72–
39 534 83.
- 41 535 Wang, Y., Zhu, X.K., Mao, J.W., Li, Z.H., Cheng, Y.B., 2011. Iron isotope fractionation during skarn-type
42 536 metallogeny: A case study of Xinqiao Cu-S-Fe-Au deposit in the Middle-Lower Yangtze valley.
43 537 *Ore Geology Reviews*, 43(1): 194-202.
- 45 538 Warner, A.E.M., Diaz, C.M., Dalvi, A.D., Mackey, P.J., Tarasov, A.V., 2006. JOM world nonferrous
46 539 smelter survey, part III: Nickel: Laterite. *Jom* 58, 11–20.
- 48 540 Wasylenki, L. E., Howe, H. D., Spivak-Birndorf, L. J., & Bish, D. L. (2015). Ni isotope fractionation during
49 541 sorption to ferrihydrite: implications for Ni in banded iron formations. *Chemical Geology*, 400,
50 542 56-64.
- 52 543 Wawryk, C.M., Foden, J.D., 2015. Fe-isotope fractionation in magmatic-hydrothermal mineral deposits: A
53 544 case study from the Renison Sn-W deposit, Tasmania. *Geochimica et Cosmochimica Acta*, 150:
54 545 285-298.
- 56 546 Wiederhold, J.G., 2015. Metal Stable Isotope Signatures as Tracers in Environmental Geochemistry.
57 547 *Environ. Sci. Technol.* 49, 2606–2624. <https://doi.org/10.1021/es504683e>

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

548 Wiederhold, J.G., Teutsch, N., Kraemer, S.M., Halliday, A.N., Kretzschmar, R., 2007. Iron isotope
549 fractionation in oxic soils by mineral weathering and podzolization. *Geochim. Cosmochim. Acta*
550 71, 5821–5833.

551 Wu, B., Amelung, W., Xing, Y., Bol, R., & Berns, A. E., 2019. Iron cycling and isotope fractionation in
552 terrestrial ecosystems. *Earth-science reviews*, 190, 323-352.

553 Yamaguchi, K.E., Johnson, C.M., Beard, B.L., Beukes, N.J., Gutzmer, J., Ohmoto, H., 2007. Isotopic
554 evidence for iron mobilization during Paleoproterozoic lateritization of the Hekpoort paleosol
555 profile from Gaborone, Botswana. *Earth Planet. Sci. Lett.* 256, 577–587.

556 Yin, N.-H., van Hullebusch, E.D., Benedetti, M., Lens, P.N.L., Sivry, Y., 2018. Zn isotopes fractionation
557 during slags’ weathering: One source of contamination, multiple isotopic signatures.
558 *Chemosphere* 195, 483–490. <https://doi.org/10.1016/j.chemosphere.2017.11.184>.

Table captions

Table 1: List of natural and pyrometallurgical samples and their Fe and Zr contents. $\tau_{\text{Fe}} = ([\text{Fe}]/[\text{Zr}]_{\text{sample}} / [\text{Fe}]/[\text{Zr}]_{\text{protolith}} - 1)$, where the sample RC 27-28 represents the protolith. The Fe isotope composition and two standard errors are calculated from the 3 analyses using Student's t-corrected factor (Platzner, 1997). $\delta\text{Fe}/\text{amu}$ is the deviation of the Fe isotope composition of a sample relative to the standard and normalized to a mass difference of 1 atomic mass unit (amu).

Sample name	Depth	[Fe]	[Zr]	τ_{Fe}	$\delta^{57}\text{Fe}$	2 SE	$\delta\text{Fe}/\text{amu}$
		g kg^{-1}	mg kg^{-1}		‰		
By products							
Ore 1		165			0.00 ± 0.18		0.00
Ore 5		118			0.03 ± 0.20		0.01
SS7		106			-0.10 ± 0.09		-0.03
SS8		124			0.05 ± 0.14		0.02
WRS		71			0.07 ± 0.08		0.02
BRS		179			0.05 ± 0.11		0.02
FeNi2		689			0.07 ± 0.12		0.02
Soil samples							
BAS1 0-10	0-10 cm	271	14	-0.10	-0.10 ± 0.07		-0.03
BAS1 10-30	10-30 cm	261	13	-0.09	0.02 ± 0.15		0.01
BAS1 30-80	30-80 cm	255	13	-0.08	0.00 ± 0.13		0.00
Lateritic samples							
RC0-1	3-4 m	256	11	0.08	0.07 ± 0.05		0.02
RC6-7	9-10 m	369	14	0.22	-0.06 ± 0.11		-0.02
RC16-17	19-20 m	85	4	0.01	0.02 ± 0.19		0.01
RC24-25	27-28 m	67	3	0.05	0.02 ± 0.34		0.01
RC27-28	30-31 m	49	2	0.00	0.08 ± 0.20		0.03

Figure captions

Figure 1: Schematic view of the FeNi smelting and refining processes modified from Ettler et al. (2016) showing the average Ni and Fe contents in ore (n= 7), F (n=10), SS (n=8), WRS (n=1), BRS (n=1) and FeNi (n=2) (Ratié et al., 2016).

Figure 2: Iron isotopic values ($\delta^{57}\text{Fe}$ in ‰) vs. the τ_{Fe} normalized by Zr (Poitrasson et al., 2008; our study), Th (Liu et al., 2014) and Ti (Li et al., 2017). The vertical black line represents the border between the gain and loss of Fe. The blue band represents the $\delta^{57}\text{Fe}$ value of the bulk silicate Earth (Poitrasson et al., 2006). Error bars represent 2 SD (Liu et al., 2014; Li et al., 2017) and 2 SE (Poitrasson et al., 2008; our study).

Figure 3: $\delta\text{Ni}/\text{amu}$ and $\delta\text{Fe}/\text{amu}$ values for lateritic profile, ultramafic (UM) soils and products from mining (feed material) and smelting (smelting slags, refining slags, and FeNi) activities. $\delta\text{Ni}/\text{amu}$ values were calculated based on the $\delta^{60/58}\text{Ni}$ values in Ratié et al. (2015, 2016, 2018). The green band represents the $\delta\text{Ni}/\text{amu}$ value of the bulk silicate Earth (Gall et al., 2017) and the blue band represents the $\delta\text{Fe}/\text{amu}$ value of the bulk silicate Earth (Poitrasson, 2006). The dotted line separates anthropogenic samples from geogenic samples.

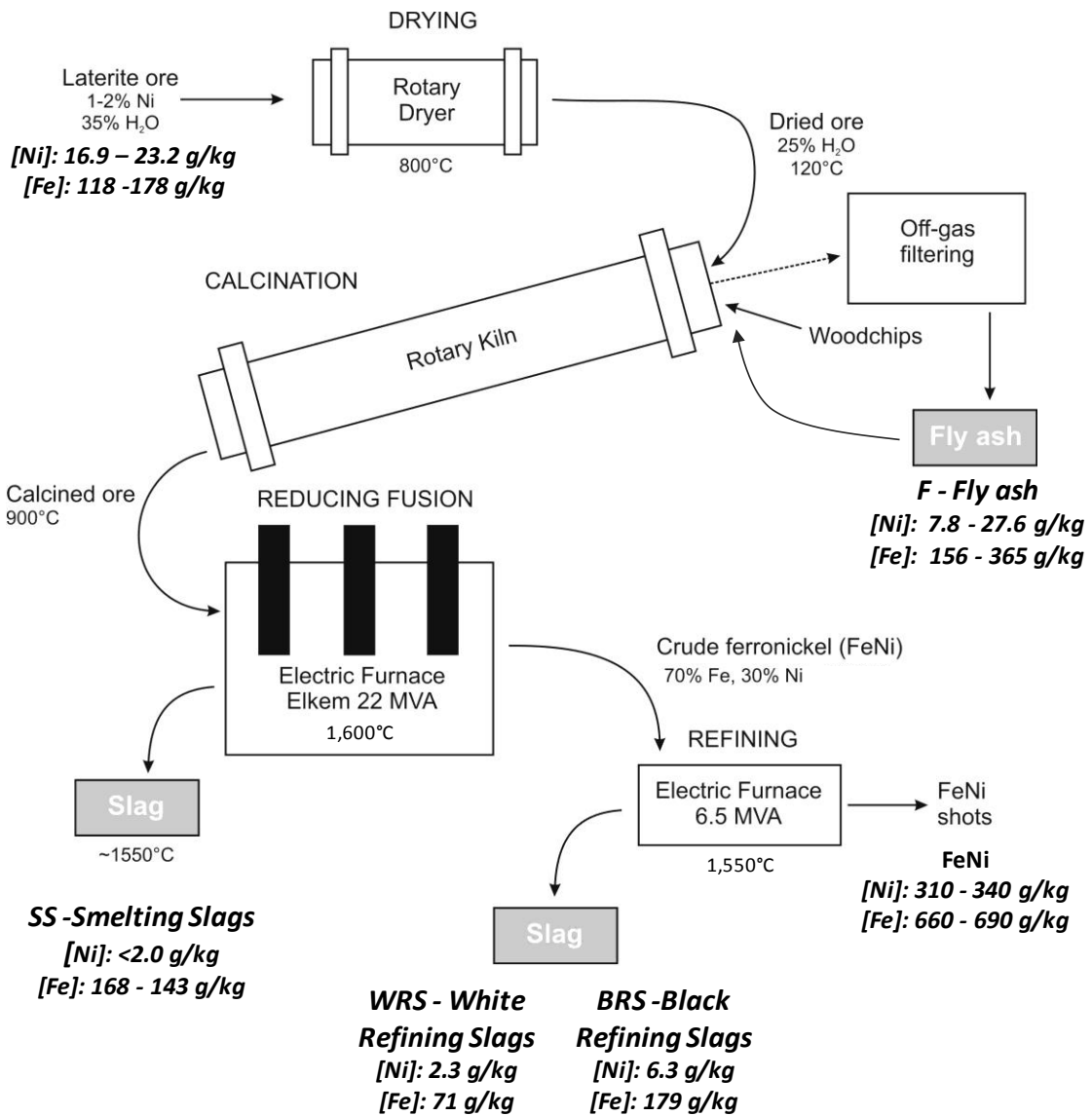


Figure 1

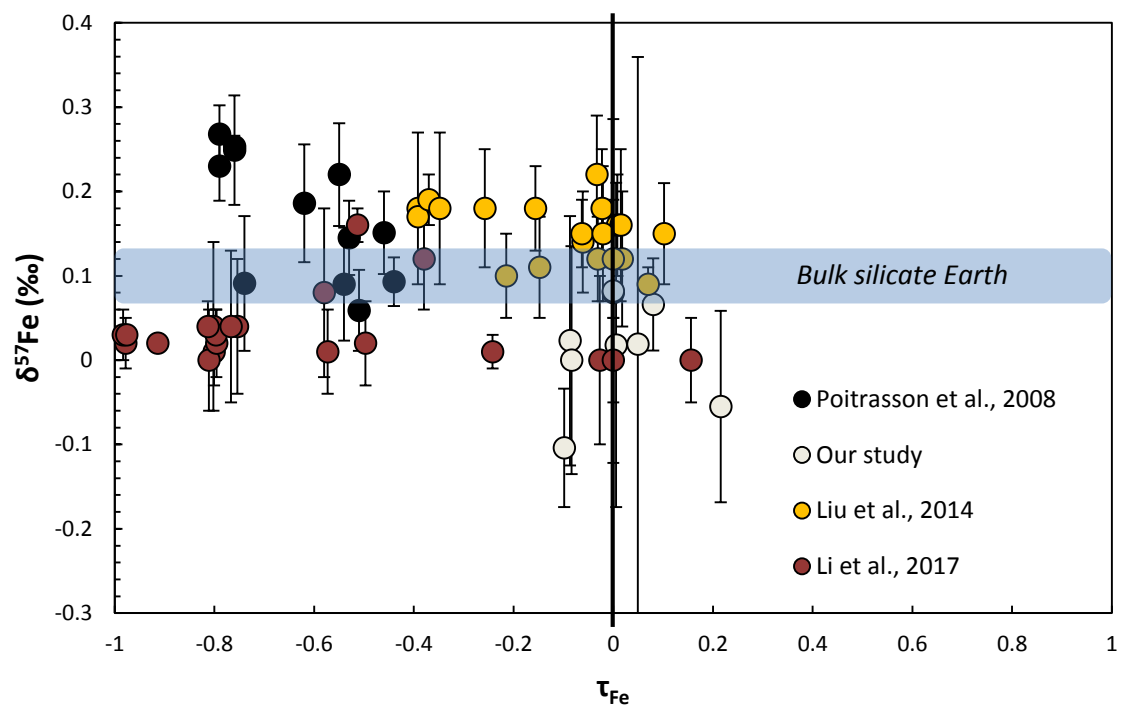


Figure 2

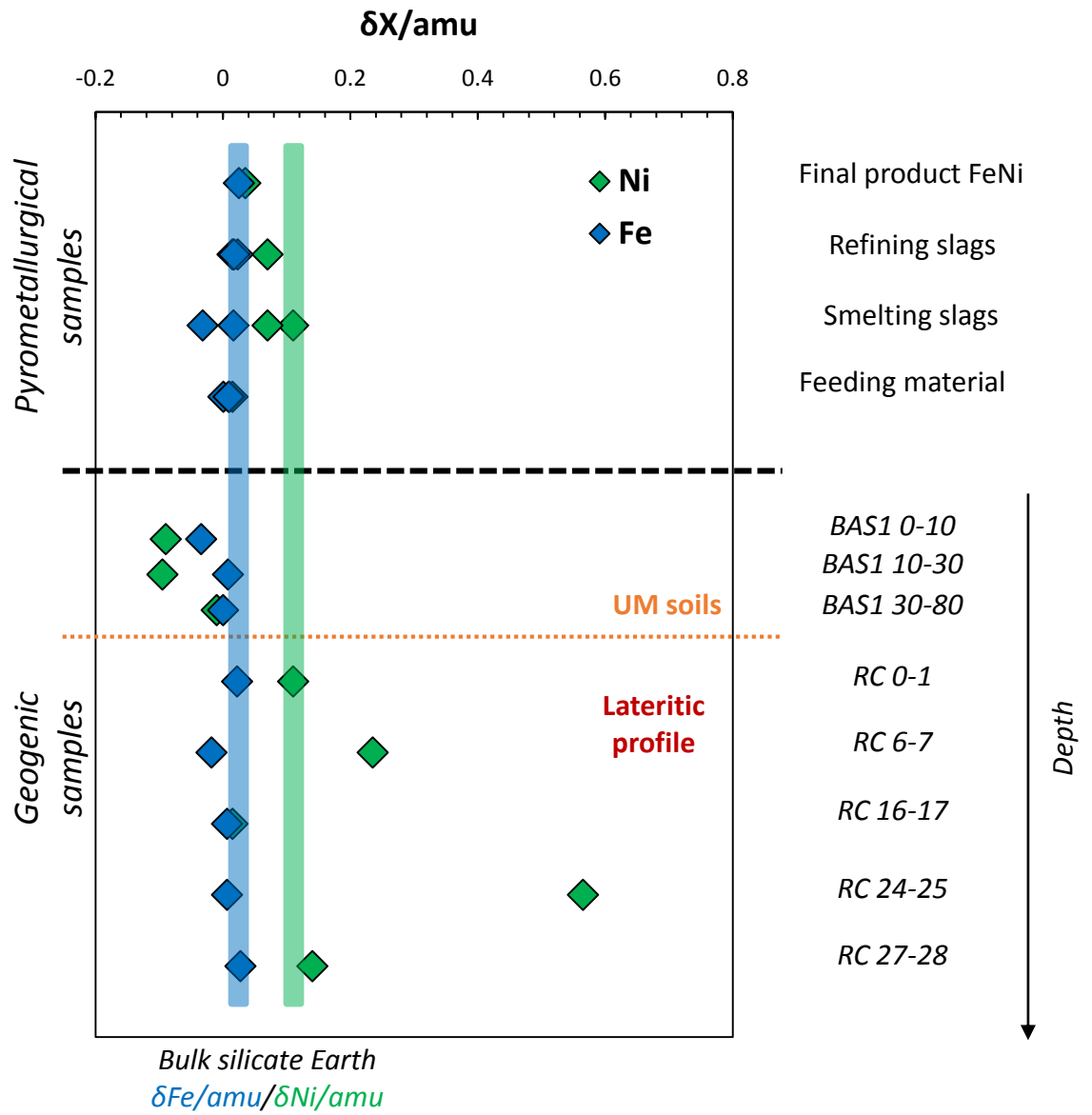


Figure 3