Supplementary information

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3	Plants affect the dissipation and leaching of anilide pesticides in soil
4	mesocosms: insights from compound-specific isotope analysis (CSIA)
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48 A. Physico-chemical properties of acetochlor, alachlor, *S*-metolachlor, butachlor and metalaxyl.

	Acetochlor	Alachlor	S-metolachlor	Butachlor	Metalaxyl
Chemical structure					
Molecular formula ^a	$C_{14}H_{20}CINO_2$	$C_{14}H_{20}CINO_2$	$C_{15}H_{22}CINO_2$	C ₁₇ H ₂₆ ClNO ₂	$C_{15}H_{21}NO_4$
Number of carbon atoms of alkoxy chain	2	1	1	4	1
Molar mass $[g \cdot mol^{-1}]^a$	269.77	269.77	283.79	311.9	279.33
Solubility (water) $[mg \cdot L^{-1}]^{a}$	282	240	480	20	8400
Bulk density $[g \cdot mL^{-1}]^a$	1.12	1.13	1.12	1.08	1.2
Log K _{ow} ^a	4.14	3.09	3.05	4.5	1.75
$K_{0c} [mL \cdot g^{-1}]^a$	156	335	185	700	162
Vapor pressure (25 °C) [mPa] ^a	2.2×10^{-2}	2.9	3.7	0.24	0.75
$H_{cc} 20 \text{ °C} [dimensionless]^a$	2.1×10^{-3}	3.2×10 ⁻³	8.98×10^{-7}	1.54×10^{-6}	1.6×10 ⁻⁵
DT ₅₀ - soil (Typical) [d] ^a	14	14	15	56	36

49 ^a Values from IUPAC <u>https://iupac.org/</u>

50 **B.** Characteristics of vineyard and forest soils

51 *1. Soil analysis*

Water content was measured by drying soil aliquots at 105 °C (ISO 11465). Soil 52 physicochemical parameters were determined following standard analytical procedure 53 on dried soil sieved at 2-mm. pH was measured in a 1:5 v/v soil/water (pHw) and in 1:5 54 55 v/v soil/0.1N KCl solution (pHk) (ISO 10390). Cation exchange capacity was measured by the cobalt-hexamine chloride extraction method (ISO 23470). Total carbon and total 56 organic carbon (TOC) were measured by combustion on a CHN elementary analyzer 57 (ISO 10694). Prior to organic carbon measurement, the vineyard soil was decarbonized 58 using an acid fumigation. The carbonate content of the forest soil was negligible and 59 60 decarbonization was not necessary. The difference between total carbon and organic carbon was considered to quantify carbonate content. Particle size distribution was 61 measured continuously using a laser granulometer (Beckmann Coulter LS230) and 62 63 converted into three classes of texture content following the WRB: <2 µm clay, 2-50 µm silt, 50-2000 µm sand. Major and minor elements were measured following the 64 alkaline soil fusion method (ISO 14869-2:2002) by ICP-AES and ICP-MS. 65 Mineralogical analysis was performed using X-Ray diffraction (D8 Advance Eco 66 67 Diffractometer).

	Soil type	Vineyard soil (V)	Forest soil (F)
	Origin	Rouffach (Alsace,	Strengbach (Alsace,
	Oligin	France)	France)
	Soil sieving	< 2 mm	< 2 mm
	Treatment	Air-dried	Air-dried
	Clay (< 2µm) [%]	24.4	6.8
	Silt (2 - 50 µm) [%]	61.1	31.6
	Sand (50 - 2000 µm) [%]	14.5	61.6
	Soil texture (USDA system)	Silt Loam	Sandy Loam
	Total carbon (TC) [%]	4.0	4.5
re	Total nitrogen (TN) [%]	0.3	0.2
xtu	Organic matter (OM; loss on	1 /	5.2
Te	ignition) [%]	1.4	5.5
	pH (water)	8.0	3.4
	pH (KCl)	7.3	2.6
	Cation exchange capacity (CEC)	19.0	14.2
	$[\text{cmol}^+\text{kg}^{-1}]$	16.0	14.5
	CaCO ₃ [%]	2.9	-
e		Calcite	Smectite (2:1)
ullin sitio		Albite	Vermiculite (2:1)
sta		Chlorite	Illite-mica (2:1)
)ry m]		Illite-mica (2:1)	Kaolinite (1:1)
0 5		Kaolinite (1:1)	Montmorillonite (2:1)
	Water content [%]	18.2	46.9
	Organic carbon (OC) [%]	1.1	4.5
	Bulk density (at field) [g cm ⁻³]	1.4	1.0
liu]	SiO ₂	52.7	61.3
y st	Al_2O_3	7.7	14.9
dr	MgO	1.4	0.6
% [S]	CaO	15.1	0.0
ts [AF	Fe_2O_3	3.2	2.6
CP.	MnO	0.1	0.0
(IC	TiO ₂	0.5	0.0
r el	Na ₂ O	0.8	0.6
ajo	K ₂ O	1.8	4.3
Μ	P_2O_5	0.4	0.2
ts ES)	Sr	204	71
nen -AJ	Ba	401	230
CP CP	Cr	74	29
se e	Ni	30	2
rac m]	Sc	7	5
T [pp	V	58	29

	Soil type	Vineyard soil (V)	Forest soil (F)		
	Origin	Rouffach (Alsace, France)	Strengbach (Alsace, France)		
	Soil size	< 2 mm	< 2 mm		
	Treatment	Air-dried	Air-dried		
	Со	9	2		
	Cu	75	10		
	Zn	23	10		
	As	9	9		
	Rb	102	411		
	Y	27	10		
	Zr	268	116		
	Nb	11	23		
	Mo	0.4	0.4		
	Cd	1.3	0.5		
$\widehat{\mathbf{G}}$	Sn	5.2	17.5		
SM-T	Sb	1.4	3.1		
	Cs	10.4	37.3		
(IC	La	29.9	25.1		
n]	Ce	62.1	51.1		
Idd	Pr	7.2	5.8		
S []	Nd	28.2	21.0		
ent	Sm	6.0	3.9		
ŝ	Eu	1.1	0.4		
ele	Gd	4.7	2.5		
ace	Tb	0.8	0.4		
ľr:	Dy	4.5	1.9		
L .	Ho	1.0	0.4		
	Er	2.5	1.0		
	Tm	0.2	0.2		
	Yb	2.5	1.0		
	Lu	0.4	0.2		
	Hf	7.5	3.8		
	Та	0.9	3.9		
	W	8.0	21.0		
	Pb	28.3	39.5		
	Тh	11.0	11.3		
	U	2.6	3.3		

73 C. Water holding capacity (WHC)

WHC was determined by saturating soil and pressurizing them at -33kPa for 24h (Rai et al., 2017). WHC for V was $34 \pm 3.5\%$ and $48 \pm 0.8\%$ for F. To homogenize water content throughout the experiments, the final water content was set at 50% (*W/W*) in all mesocosms. This corresponds to 300 mL per dry weight soil per mesocosm.

78

79 D. Spiking mesocosms

A stock solution of the pesticides mix (acetochlor, alachlor, S-metolachlor, butachlor, 80 81 metalaxyl) at 1 g/L (200 mg each compound in 200 mL) was prepared in 82 dichloromethane (DCM). An aliquot of pesticide stock solution in DCM was dissolved in 100 mL of ultra-pure water (obtained from a MilliQ device) to reach the target soil 83 concentration (25 mg/kg) of the pesticide mix. The solution was stirred several hours 84 until total evaporation of DCM to achieve a homogeneous pesticides contaminated water 85 solution. The aqueous pesticide solution was used to adjust the soil water content (field 86 capacity) of each mesocosms. 100 mL of aqueous pesticide solution (150 mg/L) were 87 added per spiked mesocosms. Additionally, 200 mL of ultra-pure water were used to 88 89 rinse the glass-beakers and the remaining amount of pesticides was transferred by capillarity from the bottom of the soil mesocosms (final amount of 180 mg of 90 pesticides). 91

92

93 E. Preparation and application of artificial rainfalls

The artificial rainwater was prepared according to the ERMCA408 reference material (ERM certification report, 2010) with sterile ultra-pure water and targeted a pH of 6.3 ± 0.6 . Artificial rainfalls were applied by pumping rainwater using a five-holes tube

- 97 peristaltic pump. Rainfall characteristics are shown in Table S1. Rainfall covered the
- 98 whole surface of each mesocosms (201 cm^2).

Pattern	Duration	Intensity	Rate	Volume
description	(min)	(mm/h)	(mL/min)	(mL)
Mid-lasting pattern with intermediate intensity	12	62	25	300

Table S1. Characteristics of rainfalls applied on days 45, 60 and 75 across the experiment.

100

101 F. Samples collection and preparation

Bulk soil: Bulk soil samples consisted in mesocosms soil from unplanted
mesocosms. In planted mesocosms, bulk soil consisted in the easily detached soil
fraction after manually shaking plants.

105 Rhizosphere: The rhizosphere fraction was obtained from planted mesocosms. An 106 adapted protocol from Barillot et al., (2013) was used to collect rhizosphere soil attached to the roots. Briefly, roots and rhizosphere soil were shaken vigorously in ultra-107 108 pure water. Ultra-pure water was replaced until water remained clean (at least three 109 times), so that rhizosphere soil is detached from the roots. Remaining rhizosphere was obtained as a suspension, centrifuged (4000 rpm at 4°C for 20 min) to separate solid 110 from liquid phase and pooled all aliquots. Concentration of pesticides was determined in 111 112 both phases and results were pooled for the whole rhizosphere. In bulk soil and rhizosphere samples, determination of pH and CEC was done following methods 113 114 explained in SI-B1.

Roots: Roots were ultrasonicated for 30 min and collected using a 1 mm-sieve.
They were rinsed again with ultra-pure water and when clean, they were dried (60 °C, 1
week). Roots were considered clean when adhering soil particles were removed. The

presence of soil on roots was checked with binoculars. When particles were found onroots, a brush was used to discard them.

Aerial part: One sample per planted mesocosms was obtained at day 75 by cutting
the green part from the stem basis. Samples were washed separately with ultrapure
water.

Leachates: Leachates were collected after each rainfall (L1 on day 45, L2 on day 123 124 60 and L3 on day 75) in a collector below the mesocosms. Each mesocosms was left for 125 12h in the climate chamber to allow complete rain leaching. Pesticides and chemical characterization (pH, DOC, major cations and anions and minor elements) were 126 127 determined in the leachates separately for each mesocosm. Pesticides were determined by GC-MS and major and minor elements were determined by ICP-AES, and ICP-MS 128 129 and ionic chromatography (ICS5000- Thermo-Fisher). pH determination was carried out using a calibrated pH-electrode. 130

131

132 G. Pesticides extraction

133 1. Extraction from soils/rhizosphere

134 Soil extraction and cleaning procedure were adapted from methods described by Ivdra 135 et al. (2014). Soil samples (around 5 g dry soil) were put into a 15 mL centrifuge tubes, mixed with 3 mL of the extraction solvent (DCM:Pentane 3:1) and vortexed 5 s 136 followed by 5 min into ultrasonic bath (Branson 5510, 40 kHz) to homogenize it. In 137 138 order to increase extraction efficiency, Millipore ultra-pure water was added to reach 50% (W/W) when necessary. Samples were then vortexed for 1 min and centrifuged for 139 140 5 min at 5000 rpm. The supernatant was transferred to amber glass vials and the extraction procedure was repeated twice without further addition of water. The resulting 141

supernatants from the three extraction cycles were combined and concentrated under a gentle nitrogen stream at room temperature to the last drop and adjusted to 1 mL of acetonitrile. The extracts were dried over anhydrous MgSO₄, and cleaned up using 12.5 mg PSA and 75 mg MgSO₄. After vortexing the mixture for 30 s and centrifuging at 5000 rpm for 90 s, the supernatant was transferred to clean GC vials and stored in the dark at -20°C until quantification and carbon isotopic analysis.

148 2. Extraction from leachates

Extraction cartridges were washed successively with 5 mL ethanol, rinsed with ACN (5 mL), and conditioned with deionized water (10 mL). Up to 6 samples were loaded on parallel cartridges which was then dried under nitrogen flux for 10 min. The pesticides were eluted by 5 mL of EtOAc followed by 5 mL of ACN. The extracts were subsequently concentrated under nitrogen flux to 1 droplet, and adjusted to 1 mL with of ACN for quantification and carbon isotopic composition analyses. Validity of the method was previously developed by Elsayed et al., (2014).

156

157 H. Pesticides quantification by GC-MS

158 All pesticides were analyzed by gas chromatography (GC, Trace 1300, Thermo Fisher Scientific) coupled with a mass spectrometer (MS, ISQ[™], Thermo Fisher Scientific). 159 Metolachlor- d_{11} was automatically added to each sample as an internal standard at a 160 constant concentration using a TriPlus RSH[™] system (Thermo Fisher Scientific). 161 162 Samples (1.5 µL volume) and internal standard (1 µL at 1000 µg/L) were injected into a 163 split/splitless injector operated in split mode with a split flow at 6.0 mL/min and held at 280°C. Separation was performed on a TG-5MS column (30 m \times 0.25 mm ID, 0.25 μ m 164 film thickness), with helium as carrier gas at a 1.5 mL/min flow rate. The GC oven 165

program was held at 50°C for 1 min, ramped to 160°C at 30°C/min, then to 220°C at 4 °C/min, and finally, to 300°C at 30 °C/min held for 1 min. The MS transfer line and source were heated at 320°C. Each sample was measured in triplicate. Characteristics of GC-MS measurements for all pesticides are provided below (Table S2).

- 170 Table S2 Environmental limits of detection (LOD), limits of quantification (LOQ), and
- accuracy of GC-MS measurements of ACE, ALA, SM, BUT and MTY in samples.

	GC-MS					
Compound	LOD [µg/L]	LOQ [µg/L]	Accuracy [%]			
Acetochlor	7.5	22.7	100.0 ± 7.5			
Alachlor	7.6	23.0	98.5 ± 6.6			
S-Metolachlor	6.5	19.6	98.7 ± 5.5			
Butachlor	5.7	17.4	98.6 ± 5.4			
Metalaxyl	7.1	21.4	99.7 ± 7.8			

172

173 Recoveries of the extraction method were evaluated by spiking a stock solution of all 174 pesticides in soil samples in a concentration ranged from 0.5 to 150 μ g/g. Pesticides 175 extraction was performed one day after spiking while samples were storage in the fridge 176 (24h, 4 °C). Recoveries are provided in Table S3.

177

178 **Table S3** Recovery rate of ACE, ALA, SM, BUT and MTY from soil samples.

	Vineyard soil ^a	Forest soil ^a
	Recovery [%]	Recovery [%]
ACE	59.5 ± 2.2	49.8 ± 3.5
ALA	59.5 ± 2.0	47.9 ± 3.0
SM	63.9 ± 1.9	49.2 ± 2.9
BUT	39.0 ± 1.7	30.8 ± 2.2
MTY	67.8 ± 2.6	57.8 ± 2.9

^a Means of measurements (n = 3) of 12 soil extraction experiments with a concentration rage from 0.5 to 150 µg/g of

the pesticide mix.

181 I. EA-IRMS and GC-C-IRMS methods for pesticides analysis

182 The carbon/nitrogen isotope compositions were determined using a GC-C-IRMS system consisting of a TRACE[™] Ultra Gas Chromatograph (ThermoFisher Scientific) coupled via a 183 184 GC IsoLink/Conflow IV interface to an isotope ratio mass spectrometer (DeltaV plus, 185 ThermoFisher Scientific). The oxidation furnace of the interface consisted of a Ni/Cu/Pt reactor and was set to a temperature of 1000°C. A TG-5MS column (60 m × 0.25 mm ID, 186 187 $0.25 \,\mu\text{m}$ film thickness) was used for chromatographic separation, with helium as the carrier 188 gas at a flow rate of 1.5 mL min-1. For carbon analysis, the column was held at 50 °C for 1 min, heated at a rate of 15°C·min-1 to 150 °C, then up to 250 °C at 2°C·min-1 and then 189 190 heated at 20°C·min-1 to 300 °C and held for 3 min. For nitrogen analysis, the column was 191 held at 50°C for 1 min, heated at a rate of 20 °C·min-1 to 150 °C, then up to 270 °C at 192 10°C·min-1 and then heated at 30°C·min-1 to 300 °C and held for 3 min. Samples and 193 standards (0.5 to 3.5 µL volume) were injected into a split/splitless injector operated in 194 splitless mode and held at 250°C.

195 EA-IRMS and GC-C-IRMS instrumental parameters for pesticides analysis EA-IRMS GC-C-IRMS

		ICIVID.		10						
		ab)		0			MDL			Linearity ranges
punodu	vs. VPDB] ^a external SD)	vs. VPDB] RMS (our la SD)	vs. VPDB] ^b = 12) \$D)	vs. VPDB] ⁽ SD)	Amo inje co	unt of C cted on lumn	Amo comp inje	unt of bound cted	Ampl m/z 44 at MDL	Amount of compound injected
CO	$\delta^{13}C$ [% ¹ (n = 2) (5	$\delta^{13}C$ [‰ (n = 2) EA-I (5	δ ¹³ C [‰ τ (n : (2	Δδ ¹³ C [‰ (S	[ng]	[nmol]	[pmol] on column	[mg/L] in solution	[mV]	[mg/L] in solution
ACE	-29.3 (<0.1)	-28.6 (<0.1)	-29.6 (0.2)	-0.4 (0.1)	17.2	1.5	106	9.5	1264	10 - 200
ALA	-29.4 (<0.1)	n.d.	-29.6 (0.3)	-0.3 (0.1)	17.0	1.4	103	9.3	767	10 - 200
SM	-30.9 (<0.1)	-30.2 (0.2)	-30.6 (0.6)	-0.6 (0.1)	18.8	1.6	95	9.8	1044	10 - 200
BUT	-26.8 (<0.1)	-26.5 (<0.1)	-27.0 (0.4)	-0.8 (0.1)	15.7	1.3	76	7.2	1322	10 - 200
MTY	-31.5 (<0.1)	-31.3 (<0.1)	-31.9 (0.3)	-0.4 (0.2)	16.1	1.4	93	9.6	874	10 - 200

- ^a For comparison, the same standards were also measured at OEA laboratories. UK. on an EA-IRMS (EuroVector.
- 196 197 198 Milan. Italy) coupled via a ConFlo III (open split. Thermo Fisher Scientific. Bremen. Germany) to a MAT 253 isotope ratio mass spectrometer (Thermo Fisher Scientific)
- 199 200 ^b Reported δ^{13} C errors correspond to the standard deviation (1 σ) calculated from triplicate measurements of duplicate tests.
- 201 ^c The error associated with $\Delta \delta^{13}$ C values was calculated via error propagation based on ± 1 SD of mean δ^{13} C value.
- $\Delta \delta^{13}C$ values were determined as the difference between the measured $\delta^{13}C$ value of each compound standard obtained by EA-IRMS and the mean $\delta^{13}C$ value of the corresponding standards, obtained by replicate injections in 202
- 203 204 GC-C-IRMS systems under optimal conditions.
- 205 SD: Standard deviation. MDL: method detection limit. N.d.: not detected.

206 J. Carbon stable isotope fractionation ($\Delta \delta^{13}$ C) for soil extraction procedure

The effect of the pesticide extraction procedure on δ^{13} C values of ACE, ALA, SM, BUT 207 208 and MTY was evaluated by spiking into ultra-pure water a DCM stock solution of ACE, ALA, SM, BUT and MTY standards with known isotope signature to determine the 209 offset generated by the extraction procedure. ACE, ALA, SM, BUT and MTY 210 211 concentrations in the soil of ranged from 0.5 up to 150 mg/kg. Solutions were stirred for 212 1 h until evaporation of the solvent to achieve a homogeneous, pesticide-contaminated water solution. Each soil (about 5 g) was then spiked with a pesticide mixture. Water 213 content of soil samples was adjusted to 50% (W/W) for both vineyard and forest soils. 214 Samples were stirred vigorously for 1 h for homogenization, and incubated one day in 215 the dark at 4°C prior to extraction. Each condition was tested in duplicate. Extraction 216 was performed using the procedures described above (section G.1). 217

The impact of the extraction of ACE, ALA, SM, BUT and MTY from soil matrices on efficiency and isotope fractionation varied among compounds and soil types. The study confirmed the absence of significant fractionation effects for all pesticides ($\Delta\delta^{13}C <$ 1‰), even at low concentrations outside the previously established linearity range (Table S4).

Table S4 Carbon isotopic fractionation ($\Delta \delta^{13}$ C) associated with the extraction of ACE, ALA, SM, BUT and MTY for the vineyard and the forest soils. The offset generated by the soil extraction did not significantly change regardless the initial amount of pesticides.

Desticida	Vineyard ^a	Forest ^a
resticite	$\Delta\delta^{13}C^{b}$ [‰ vs	$SVPDB] \pm SD$
Acetochlor	-0.1 ± 0	0 ± 0.1
Alachlor	-0.2 ± 0	-0.1 \pm 0.1
S-metolachlor	-0.7 \pm 0.1	-0.5 \pm 0

Butachlor	-1	\pm	0.1	-0.8	\pm	0.1
Metalaxyl	0.1	±	0.2	-0.2	±	0.2

^a Means of measurements (n > 3) of 12 individual soil extraction experiments with different concentrations of ACE, ALA, SM, BUT and MTY.

^b The error given for the $\Delta\delta^{13}$ C values was calculated *via* error propagation based on ± 1 SD of mean ¹³C values from *n* = 3 measurements for each sample. $\Delta\delta^{13}$ C values were determined as the difference between the measured δ^{13} C value of a given compound in the extract and the mean δ^{13} C value of the corresponding standard obtained by replicate

232 injections in the GC-C-IRMS.

233

234 K. Nitrogen stable isotope fractionation ($\Delta \delta^{15}N$) for leachate extraction procedure

235 with SPE of S-metolachlor (SM)

Pesticides were extracted from leachates following a protocol adapted from US EPA
method 525.2 and described by Elsayed et al. (2014). Briefly, a SPE (solid phase
extraction) procedure was followed, using SolEx C18 cartridges (1 g, Dionex®, CA,
USA) on an AutroTrace 280 SPE system (Dionex®, CA, USA).

240

241 EA-IRMS and GC-C-IRMS instrumental parameters for N-CSIA of SM

242

	EA-IRMS		GC-C-IR	MS		$\Delta \delta^{15} N$ associated with SPE extraction
pune	R] ^a al	IR] ^b		MDL		Δδ ¹⁵ N [‰ vs. AIR] (SD)
mpc	s. Al tern)	s. A	Amount of	Amount of	Ampl m/z	
Ĉ	% vs (SD)	‰ v (SD)	N injected	compound injected	28 at MDL	
	$\delta_{15}^{15}N$ [15] $\Lambda_{21}^{15}N$ [15] $\Lambda_{21}^{15}N$	[ng]	[mg/L] in solution	[mV]		
SM	0.5 (<0.1)	0.1 (0.7)	17	113	150	0.0 (0.5)

^a For comparison, the same standards were also measured using an EA-IRMS ((Flash EA IsoLinkTMCN IRMS, Thermo Fisher Scientific, Bremen, Germany).

^b The error given for $\Delta\delta^{15}$ N values was calculated via error propagation based on ± 1 SD of mean ¹⁵N values. $\Delta\delta^{15}$ N values were determined as the difference between the measured δ^{15} N value of a given compound in the extract and the mean δ^{15} N value of the corresponding standard obtained by replicate injections in the GC-C-IRMS systems optimal conditions.

249 SD: Standard deviation. MDL: method detection limit. N.d.: not detected.

250 L. Water mass budget

251 Overall water budget of the mesocosms throughout the experiment.

	Soil		Initial soil water content	Rainfall per application ^a	L45	L60	L75	ETP
Total water volume leached [mL]			300	300	217.4 ± 19.5	235.0 ± 21.1	228.9 ± 19.4	5.7 ± 5.6
Rainfall volume leached out [%]		planted			72.5 ± 6.5	78.3 ± 7.0	76.3 ± 6.5	
Rainfall volume retained in mesocosms [%]	yard				35.1 ± 3.0	35.5 ± 6.3	34.4 ± 3.0	
	ine		300	300	240.4 ± 4.7	254.4 ± 1.6	240.9 ± 8.7	
	>	unplanted			80.1 ± 1.6	84.8 ± 0.5	80.3 ± 2.9	
					30.3 ± 1.5	29.3 ± 1.2	30.3 ± 2.3	
			300	300	115.6 ± 13.4	115.3 ± 28.2	76.6 ± 9.3	10.6 ± 5.0
		planted			38.5 ± 4.5	38.4 ± 9.4	25.5 ± 3.2	
	est				61.4 ± 3.0	68.4 ± 5.1	64.1 ± 0.7	
	for		300	300	101.4 ± 23.9	132.3 ± 43.3	74.0 ± 6.7	
		unplanted			33.8 ± 8.0	44.1 ± 14.4	24.7 ± 1.2	
					58.6 ± 1.9	61.1 ± 8.0	58.2 ± 1.1	

252 L45 (n=3), 60 (n=2) and L75 (n=2) represent leachates collected at those times. ETP: average potential evapotranspiration throughout the incubation.

^a Three rainfalls applied. Uncertainty corresponds to mean standard deviation (SD).

254 M. Pesticide mass budget

Pesticides loads in leachates (t45, 60 and 75) and mass budget in soil and rhizosphere (t1, 45 and 75) throughout the experiment.

0				ACE	ALA			SM		BUT	MTY		
Soil type		Incubation [days]	Leachate pesticide loads [µg]	Pesticide in soil/rhizosph ere [µg]	Leachate pesticide loads [µg]	Pesticide in soil/rhizospher e [µg]	Leachate pesticide loads [µg]	Pesticide in soil/rhizosphere [µg]	Leachate pesticide loads [µg]	Pesticide in soil/rhizosphere [µg]	Leachate pesticide loads [µg]	Pesticide in soil/rhizosphere [µg]	
Vineyard		1	-	14945±144	-	14492±160	-	14945±353	-	10237±248	-	14953±565	
	Planted	45	195±13	483±3/8.3±2. 1	295±7	641±3/0	362±41	900±18/10.8±3.0	505±6	485±0/3.3±4.7	979±47	2470±69/11±7	
	Tianca	60	$1.2{\pm}1.0$	-	2.9±0.4	-	31±21	-	1.8±0.2	-	156±11	-	
		75	0.7±0.5	120±0/3.9±1. 5	0.5±0.0	167±6/3.5±5.0	21±5	190±29/12.1±7.7	12±1	126±0/2.0±2.2	118±25	433±187/13±8	
		1	-	14946±505	-	14492 ± 405	-	14946±544	-	10237.3±76.6	-	14953±751	
	Unplanted	45	326±124	496±5	387±101	658±3	813±135	1520±79	597±24	478.7±1.1	1719±243	1626±74	
		60	57±17	-	58±18	-	184.5±5.9	-	6.2±5.4	-	161.8±6.5	-	
_		75	20±1	114±11	13±9	161±18	168±154	493±42	21±14	194.8±57.4	133±1	269±4	
		1	-	14946±128	-	14953±698	-	14946±493	-	12264.7±154.1	-	14954±779	
		45	102±7	2887±300/37 .8±1.0	204±92	3450±38/52±0	210±92	7349±829/63±0	91±8	3518±364/42±1	420±143	11380±984/100±2	
	rianteu	60	0.6 ± 0.0	-	1.0±0.1	-	4.6±0.7	-	1.7±0.2	-	14±7	-	
Forest		75	0.4±0	544±278/3.9 ±0.8	0.9±0.2	660±367/3.5±1 .5	1.9±0.7	2368±1330/12±3	21±5	356±269/4.3±1.4	15±6	6228±1185/14±17	
		1	-	14946±68	-	14954±63	-	14946±55	-	12265±722	-	14954±192	
	Unplanted	45	109±28	2191±164	161±35	2395±192	167±38	3726±632	127±48	1746±153	361±101	6355±1291	
	Unplaitteu	60	6.2 ± 6.0	-	7.7±7.0	-	12±6	-	2.8±0.7	-	50±31	-	
		75	0.9±0.2	1047±319	1.4±0.2	1350±498	3.0±0.5	3666±675	15±9	549±393	20±7	6179±20	

- Not applicable

258	N. Dissipation kinetics calculations								
259	Single First-Order Rate Model (SFO) was applied to the dataset:								
260	$C_t = C_0 e^{-k \cdot t}$ Equation S1								
261	The linear form of the single first-order equation is:								
262	$\ln \frac{c_t}{c_0} = -k \cdot t \qquad \qquad \text{Equation S2}$								
263	Where, C_t = concentration at time t, C_0 = initial concentration, k = rate constant of decline 1/days, t = time								
264	The half-life or time needed for the degradation of 50% of initial pesticide concentration (DT_{50} or $T_{1/2}$, days) was estimated from equation								
265	$DT_{50} = T_{1/2} = \frac{\ln\left[\frac{C_0}{\frac{C_0}{2}}\right]}{k} = \frac{0.693}{k}$ Equation S3								

266

Table S5. Summary of k and $T_{1/2}$ values obtained for acetochlor, alachlor, *S*-metolachlor, butachlor and metalaxyl in soil mesocosms

-	Soil		k (days ⁻¹)						T _{1/2} (days)					
	5011	-	ACE	ALA	SM	BUT	MTY	ACE	ALA	SM	BUT	MTY		
-	Vineyard	Planted	0.069	0.054 ± 0.018	0.058	0.176 ± 0.179	0.042	10.1	14.0 ± 5.4	12.0	7.0 ± 4.5	16.6		
		Unplanted	0.069	0.054 ± 0.017	0.047	0.171 ± 0.183	0.050	10.1	14.0 ± 5.4	14.8	10.7 ± 2.1	13.8		
-	Forest	Planted	0.037	0.033	0.017	0.092 ± 0.095	0.013	19.5	21.8	51.5	20.5 ± 9.1	54.3	_	
		Unplanted	0.035	0.032	0.022	0.095 ± 0.092	0.013	19.8	21.5	33.1	16.6 ± 0.8	56.9		
68	Uncertainty	not		given for	ACE,	ALA,	SM,	МЛ	ΓY,	when	n	<	- :	

269 **O. Dual isotope plots and Lambda (Λ) values**

270 Dual isotope plot (δ^{15} N vs. δ^{13} C) of data for SM in the vineyard planted and unplanted 271 mesocosms.



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